# Chapter 2: Assessment of the Pacific Cod Stock in the Eastern Bering Sea 

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## EXECUTIVE SUMMARY

## Summary of Changes in Assessment Inputs

Relative to the November edition of last year's BSAI SAFE report, the following substantive changes have been made in the EBS Pacific cod stock assessment.

## Changes in the Input Data

1) Catch data for 1991-2015 were updated, and preliminary catch data for 2016 were incorporated.
2) Commercial fishery size composition data for 2015 were updated, and preliminary size composition data from the 2016 commercial fisheries were incorporated.
3) Size composition data from the 2016 EBS shelf bottom trawl survey were incorporated.
4) The numeric abundance estimate from the 2016 EBS shelf bottom trawl survey was incorporated (the 2016 estimate of 640 million fish was down about $35 \%$ from the 2015 estimate).
5) Age composition data from the 2015 EBS shelf bottom trawl survey were incorporated.
6) Mean length at age data from the 2015 EBS shelf bottom trawl survey were incorporated.
7) Seasonal catch per unit effort (CPUE) data for the trawl, longline, and pot fisheries from 2015 were updated, and preliminary CPUE data for the trawl, longline, and pot fisheries from 2016 were incorporated.
8) The time series (1997-2015, odd-numbered years only) of relative population numbers and size composition from the NMFS longline survey were incorporated into one model.

## Changes in the Assessment Methodology

Many changes have been made or considered in the stock assessment model since the 2015 assessment (Thompson 2015). Six models were presented in this year's preliminary assessment (Appendix 2.1), as requested in May and June by the Joint Team Subcomittee on Pacific Cod Models and the SSC. After reviewing the preliminary assessment, the BSAI Plan Team and SSC requested that two models from the preliminary assessment (one of which is the base model that has been used for setting harvest specifications since the 2011 assessment) and four new models be presented in the final assessment. The assessment author recommends using one of the new models (Model 16.6) to set harvest specifications for 2017 and 2018.

## Summary of Results

The principal results of the present assessment, based on the author's new recommended model, are listed in the table below (biomass and catch figures are in units of t ) and compared with the corresponding quantities from last year's assessment as specified by the SSC:

| Quantity | As estimated or specified last year for: |  | As estimated or recommended this year for: |  |
| :---: | :---: | :---: | :---: | :---: |
|  | 2016 | 2017 | 2017* | 2018* |
| $M$ (natural mortality rate) | 0.34 | 0.34 | 0.36 | 0.36 |
| Tier | 3a | 3a | 3a | 3a |
| Projected total (age $0+$ ) biomass (t) | 1,830,000 | 1,780,000 | 1,260,000 | 1,110,000 |
| Projected female spawning biomass (t) | 466,000 | 530,000 | 327,000 | 340,000 |
| B100\% | 806,000 | 806,000 | 620,000 | 620,000 |
| $\mathrm{B}_{40 \%}$ | 323,000 | 323,000 | 248,000 | 248,000 |
| $B_{35 \%}$ | 282,000 | 282,000 | 217,000 | 217,000 |
| $F_{\text {OFL }}$ | 0.35 | 0.35 | 0.38 | 0.38 |
| $\operatorname{maxF}_{\text {ABC }}$ | 0.30 | 0.30 | 0.31 | 0.31 |
| $F_{\text {ABC }}$ | 0.22 | 0.22 | 0.31 | 0.31 |
| OFL (t) | 390,000 | 412,000 | 284,000 | 302,000 |
| maxABC (t) | 332,000 | 329,000 | 239,000 | 255,000 |
| ABC (t) | 255,000 | 255,000 | 239,000 | 255,000 |
| Status | As determined last year for: |  | As determined this year for: |  |
|  | 2014 | 2015 | 2015 | 2016 |
| Overfishing | No | n/a | No | n/a |
| Overfished | n/a | No | n/a | No |
| Approaching overfished | n/a | No | n/a | No |

*Projections are based on assumed catches of $255,000 \mathrm{t}, 203,000 \mathrm{t}$, and 212,000 t in 2016, 2017, and 2018, respectively.

## Responses to SSC and Plan Team Comments on Assessments in General

Four comments on assessments in general were addressed in the preliminary assessment (Appendix 2.1). In the interest of efficiency, they are not repeated in this section. The SSC made three additional comments on assessments in general after the preliminary assessment was completed (note that numbering of comments here is continuous with numbering of comments in the preliminary assessment; note also that SSC comments directed to the Plan Teams rather than the assessment authors are not included here):

SSC14 (10/16 minutes): "The SSC reminds groundfish and crab stock assessment authors to follow their respective guidelines for SAFE preparation." Close attention was paid to the SAFE chapter guidelines as this assessment was being prepared.

SSC15 (10/16 minutes):"The SSC found the model numbering in the Eastern Bering Sea (EBS) Pacific cod model extremely helpful and looks forward to having more standardized model numbering across all stock assessment documents." This assessment continues to use the model numbering convention adopted in last year's final assessment and this year's preliminary assessment.

SSC16 (10/16 minutes): "The SSC requests that stock assessment authors bookmark their assessment documents and commends those that have already adopted this practice." This assessment is fully bookmarked.

## Responses to SSC and Plan Team Comments Specific to this Assessment

Eleven comments specific to this assessment, some of which contained multiple parts, were addressed in the preliminary assessment (Appendix 2.1). In the interest of efficiency, they are not repeated in this section, with three exceptions: comments SSC7, SSC8, and SSC12. These three comments, along with BSAI Plan Team (BPT) and SSC comments that were developed following completion of the preliminary assessment, are shown below.

SSC7 (12/15 minutes): "While the model selection criteria proposed by the author are reasonable, we note that these criteria do not take into account the model fit itself. Model fit and retrospective performance should be more strongly considered in the selection of a final model for specifications." Model fit and retrospective performance are considered in selection of the final model (see "Choice of Final Model," under "Model Evaluation" in the "Results" section).

SSC8 (12/15 minutes): "Although the SSC has repeatedly stressed the need to incrementally evaluate model changes, the SSC did not intend this to imply an automatic preference for the status quo model (as implied by the authors criterion \#1) if alternatives with better performance are available." The status quo model was not given automatic preference in this assessment.

SSC12 (6/16 minutes): "The SSC encourages the author to conduct a retrospective analysis across historically used models in addition to the standard retrospective analysis using the current model." In addition to the standard comparison of the spawning biomass and age 0 recruitment time series from the current assessment and last year's assessment, this assessment includes a retrospective analysis of the spawning biomass time series from all assessments since 2006 (Figure 2.15).

BPT1 (9/16 minutes): "The Team recommends bringing forward as many of the following six models, listed in prioritized order, as time permits, but Models 11.5 and 16.1 at a minimum:
A. Model 11.5
B. Model 16.1
C. Model 16.1 without empirical weight-at-age
D. Model 16.1 without empirical weight-at-age and including NMFS LL survey
E. Model 16.1 with time-varying survey selectivity
F. Model 16.1 with time-varying fishery selectivity"

All six of the Team's recommended models are included in this assessment. The "placeholder" names for the last four models in the above list (C, D, E, and F) have been replaced by the "final" model names 16.6, 16.7, 16.8, and 16.9. See also comment SSC17.

SSC17 (10/16 minutes): "The SSC agrees with the Plan Team recommendation to focus on model 16.1 for this assessment cycle and explore additional modifications as time allows. If time is available, we agree with the Plan Team that examining the incremental effects of empirical weight-at-age data and NMFS longline survey data in the model are reasonable next steps." All of the Team's recommended models are presented in this assessment, including Model 16.1 and models that examine the incremental effects of empirical weight-at-age data (Model 16.6) and NMFS longline survey data (Model 16.7). See also comment BPT1.

SSC18 (10/16 minutes):"The observed discrepancies among different models in these assessments are a good - if perhaps extreme - example of the model uncertainty that pervades most assessments. This
uncertainty is largely ignored once a model is approved for specifications. We encourage the authors and Plan Teams to consider approaches such as multi-model inference to account for at least some of the structural uncertainty. We recommend that a working group be formed to address such approaches." The procedure used to select a final model for this assessment includes a model-averaging aspect (see "Choice of Final Model," under "Model Evaluation" in the "Results" section).

SSC19 (10/16 minutes):"Regarding the mid-year model vetting process, the SSC re-iterates its recommendation from June to continue for now. The process has proven useful for the industry as an avenue to provide formal input and for the author to prioritize the range of model options to consider." Planning for next year's assessment will include continuation of the mid-year model vetting process.

SSC20 (10/16 minutes): "With regard to data weighting, the SSC recommends that the authors consider computing effective sample sizes based on the number of hauls that were sampled for lengths and weights, rather than the number of individual fish." Because none of the SSC's requested models included computation of effective sample sizes on the basis of the number of sampled hauls, this recommendation will be forwarded to the Joint Team Subcommittee on Pacific Cod Models for consideration at next year's meeting (see comment SSC19).

SSC21 (10/16 minutes): "The SSC notes that, in spite of the concerns over dome-shaped survey selectivity in the survey, there are many potential mechanisms relating to the availability of larger fish to the survey gear that could result in these patterns, regardless of the efficiency of the trawl gear to capture large fish in its path. For example, in the Bering Sea the patterns could be due to larger Pacific cod being distributed in deeper waters or in the northern Bering Sea at the time of the survey. The northern Bering Sea survey planned for 2017 should provide additional information on the latter possibility." Data from the 2017 trawl survey of the northern Bering Sea will be examined when they become available.

SSC22 (10/16 minutes): "Although there is genetic evidence for stock structuring within the Pacific cod population among regions, the uncertainty in model scale for all three regions seems to suggest that some sharing of information among the three assessments might be helpful. Over the long term, authors could consider whether a joint assessment recognizing the population structuring, but simultaneously estimating key population parameters (e.g., natural mortality, catchability or others) might lend more stability and consistency of assumptions for this species." This recommendation will be forwarded to the Joint Team Subcommittee on Pacific Cod Models for consideration at next year's meeting (see comment SSC19).

## INTRODUCTION

## General

Pacific cod (Gadus macrocephalus) is a transoceanic species, occurring at depths from shoreline to 500 m . The southern limit of the species' distribution is about $34^{\circ} \mathrm{N}$ latitude, with a northern limit of about $65^{\circ} \mathrm{N}$ latitude (Lauth 2011). Pacific cod is distributed widely over the eastern Bering Sea (EBS) as well as in the Aleutian Islands (AI) area. Tagging studies (e.g., Shimada and Kimura 1994) have demonstrated significant migration both within and between the EBS, AI, and Gulf of Alaska (GOA). However, recent research indicates the existence of discrete stocks in the EBS and AI (Canino et al. 2005, Cunningham et al. 2009, Canino et al. 2010, Spies 2012). Although the resource in the combined EBS and AI (BSAI) region had been managed as a single unit from 1977 through 2013, separate harvest specifications have been set for the two areas since the 2014 season.

Pacific cod is not known to exhibit any special life history characteristics that would require it to be assessed or managed differently from other groundfish stocks in the EBS.

## Review of Life History

Pacific cod eggs are demersal and adhesive. Eggs hatch in about 15 to 20 days. Spawning takes place in the sublittoral-bathyal zone ( 40 to 290 m ) near bottom. Eggs sink to the bottom after fertilization and are somewhat adhesive. Optimal temperature for incubation is $3^{\circ}$ to $6^{\circ} \mathrm{C}$, optimal salinity is 13 to 23 parts per thousand (ppt), and optimal oxygen concentration is from 2 to 3 ppm to saturation. Little is known about the optimal substrate type for egg incubation.

Little is known about the distribution of Pacific cod larvae, which undergo metamorphosis at about 25 to 35 mm . Larvae are epipelagic, occurring primarily in the upper 45 m of the water column shortly after hatching, moving downward in the water column as they grow.

Juveniles occur mostly over the inner continental shelf at depths of 60 to 150 m . Adults occur in depths from the shoreline to 500 m , although occurrence in depths greater than 300 m is fairly rare. Preferred substrate is soft sediment, from mud and clay to sand. Average depth of occurrence tends to vary directly with age for at least the first few years of life. Neidetcher et al. (2014) have identified spawning locations throughout the Bering Sea and Aleutian Islands.

It is conceivable that mortality rates, both fishing and natural, may vary with age in Pacific cod. In particular, very young fish likely have higher natural mortality rates than older fish (note that this may not be particularly important from the perspective of single-species stock assessment, so long as these higher natural mortality rates do not occur at ages or sizes that are present in substantial numbers in the data). For example, Leslie matrix analysis of a Pacific cod stock occurring off Korea estimated the instantaneous natural mortality rate of 0 -year-olds at $2.49 \%$ per day (Jung et al. 2009). This may be compared to a mean estimate for age 0 Atlantic cod (Gadus morhua) in Newfoundland of 4.17\% per day, with a $95 \%$ confidence interval ranging from about $3.31 \%$ to $5.03 \%$ (Robert Gregory, DFO, pers. commun.); and age 0 Greenland cod (Gadus ogac) of $2.12 \%$ per day, with a $95 \%$ confidence interval ranging from about $1.56 \%$ to $2.68 \%$ (Robert Gregory and Corey Morris, DFO, pers. commun.).

Although little is known about the likelihood of age-dependent natural mortality in adult Pacific cod, it has been suggested that Atlantic cod may exhibit increasing natural mortality with age (Greer-Walker 1970).

At least one study (Ueda et al. 2006) indicates that age 2 Pacific cod may congregate more, relative to age 1 Pacific cod, in areas where trawling efficiency is reduced (e.g., areas of rough substrate), causing their selectivity to decrease. Also, Atlantic cod have been shown to dive in response to a passing vessel (Ona and Godø 1990, Handegard and Tjøstheim 2005), which may complicate attempts to estimate catchability $(Q)$ or selectivity. It is not known whether Pacific cod exhibit a similar response.

As noted above, Pacific cod are known to undertake seasonal migrations, the timing and duration of which may be variable (Savin 2008).

## FISHERY

## Description of the Directed Fishery

During the early 1960s, a Japanese longline fishery harvested EBS Pacific cod for the frozen fish market. Beginning in 1964, the Japanese trawl fishery for walleye pollock (Theragra chalcogramma) expanded and cod became an important bycatch species and an occasional target species when high concentrations were detected during pollock operations. By the time that the Magnuson Fishery Conservation and Management Act went into effect in 1977, foreign catches of Pacific cod had consistently been in the

30,000-70,000 t range for a full decade. In 1981, a U.S. domestic trawl fishery and several joint venture fisheries began operations in the EBS. The foreign and joint venture sectors dominated catches through 1988, but by 1989 the domestic sector was dominant and by 1991 the foreign and joint venture sectors had been displaced entirely.

Presently, the Pacific cod stock is exploited by a multiple-gear fishery, including trawl, longline, pot, and jig components (although catches by jig gear are very small in comparison to the other three main gear types, with an average annual catch of less than 200 t since 1992). The breakdown of catch by gear during the most recent complete five-year period (2011-2015) is as follows: longline gear accounted for an average of $54 \%$ of the catch, trawl gear accounted for an average of $31 \%$, and pot gear accounted for an average of $14 \%$ (percentages do not sum to 100 due to rounding).

In the EBS, Pacific cod are caught throughout much of the continental shelf, with NMFS statistical areas $509,513,517,519$, and 521 each accounting for at least $5 \%$ of the average catch over the most recent 5year period (2011-2015).

Catches of Pacific cod taken in the EBS for the periods 1964-1980, 1981-1990, and 1991-2016 are shown in Tables 2.1a, 2.1b, and 2.1c, respectively. The catches in Tables 2.1a and 2.1b are broken down by fleet sector (foreign, joint venture, domestic annual processing). The catches in Table 2.1b are also broken down by gear to the extent possible. The catches in Table 2.1c are broken down by gear.

Appendix 2.2 contains an economic performance report on the BSAI Pacific cod fishery.

## Effort and CPUE

Figures 2.1 and 2.2 show, subject to confidentiality restrictions, the approximate locations in which hauls or sets sampled during 2015 and 2016 contained Pacific cod. To create these figures, the areas managed under the FMP were divided into $20 \mathrm{~km} \times 20 \mathrm{~km}$ squares. For each gear type, a square is shaded if hauls/sets containing Pacific cod from more than two distinct vessels were sampled in it during the respective gear/season/year (Figure 2.1) or gear/year (Figure 2.2). Figure 2.1 shows locations of sampled EBS hauls/sets containing Pacific cod for trawl, longline, and pot gear, for the January-April, May-July, and August-December seasons. Figure 2.2 shows locations of sampled EBS hauls/sets for the same gear types, but aggregated across seasons. More squares are shaded in Figure 2.2 than in Figure 2.1 because aggregating across seasons increases the number of squares that satisfy the confidentiality constraint.

Various gear-specific time series of fishery catch per unit effort (CPUE) are plotted in Figure 2.3. Based on linear regressions over the last 10 years (i.e., beginning in 2006), the slopes for 11 out of the 14 plots are positive and 3 are negative. However, only 3 of the positive slopes and 2 of the negative slopes are statistically significant at the $95 \%$ level.

## Discards

The catches shown in Tables 2.1b and 2.1c include estimated discards. Discards of Pacific cod in the EBS Pacific cod fisheries are shown for each year 1991-2016 in Table 2.2. Amendment 49, which mandated increased retention and utilization of Pacific cod, was implemented in 1998. From 1991-1997, discard rates in the Pacific cod fishery averaged about 4.9\%. Since then, they have averaged about 1.4\%.

## Management History

The history of acceptable biological catch (ABC), overfishing level (OFL), and total allowable catch (TAC) levels is summarized and compared with the time series of aggregate (i.e., all-gear, combined area)
commercial catches in Table 2.3. Note that, prior to 2014, this time series pertains to the combined BSAI region, so the catch time series differs from that shown in Table 2.1, which pertains to the EBS only.

From 1980 through 2015, TAC averaged about $84 \%$ of ABC (ABC was not specified prior to 1980), and from 1980 through 2015, commercial catch averaged about $92 \%$ of TAC. In 10 of these 36 years, TAC equaled ABC exactly, and in 8 of these 36 years, catch exceeded TAC (by an average of $3 \%$ ). However, three of those overages occurred in 2007, 2008, and 2010, when TAC was reduced by $3 \%$ to account for a small, State-managed fishery inside State of Alaska waters within the AI subarea (similar reductions have been made in all years since 2006); thus, while the combined Federal and State catch exceeded the Federal TAC in 2007, 2008, and 2010 by about 2\% or less, the overall target catch (Federal TAC plus State GHL) was not exceeded.

Total catch has been less than OFL in every year since 1993.
Changes in ABC over time are typically attributable to three factors: 1) changes in resource abundance, 2) changes in management strategy, and 3) changes in the stock assessment model. Assessments conducted prior to 1985 consisted of simple projections of current survey numbers at age. In 1985, the assessment was expanded to consider all survey numbers at age from 1979-1985. From 1985-1991, the assessment was conducted using an ad hoc separable age-structured model. In 1992, the assessment was conducted using the Stock Synthesis modeling software (Methot 1986, 1990) with age-based data. All assessments from 1993 through 2003 continued to use the Stock Synthesis modeling software, but with length-based data. Age data based on a revised ageing protocol were added to the model in the 2004 assessment. At about that time, a major upgrade in the Stock Synthesis architecture resulted in a substantially new product, at that time labeled "SS2" (Methot 2005). The assessment was migrated to SS2 in 2005. Changes to model structure were made annually through 2011, but the base model has remained constant since then (see Appendix 2.3). A note on software nomenclature: The label "SS2" was dropped in 2008. Since then, the program has been known simply as "Stock Synthesis" or "SS," with several versions typically produced each year, each given an alpha-numeric label.

Beginning with the 2014 fishery, the Board of Fisheries for the State of Alaska has established guideline harvest levels (GHLs) in State waters between 164 and 167 degrees west longitude in the EBS subarea (these have supplemented GHLs that had been set aside for the Aleutian Islands subarea since 2006). The table below shows the formulas that have been used to set the State GHL for the EBS (including the formula anticipated for setting the 2017 GHL):

| Year | Formula |
| :--- | :--- |
| 2014 | $0.03 \times($ EBS ABC + AI ABC $)$ |
| 2015 | $0.03 \times($ EBS ABC + AI ABC $)$ |
| 2016 | $0.064 \times$ EBS ABC |
| 2017 | $0.064 \times$ EBS ABC |

Table 2.4 lists all implemented amendments to the BSAI Groundfish FMP that reference Pacific cod explicitly. The final rule implementing Amendment 113, which deals with the fishery for Pacific cod in the Aleutian Islands, has not been published as of this writing. The proposed rule is available at https://alaskafisheries.noaa.gov/sites/default/files/81fr50444.pdf.

## DATA

This section describes data used in the current stock assessment models. It does not attempt to summarize all available data pertaining to Pacific cod in the EBS.

The following table summarizes the sources, types, and years of data included in the data file for at least one of the stock assessment models:

| Source | Type | Years |
| :--- | :--- | :--- |
| Fishery | Catch biomass | $1977-2016$ |
| Fishery | Catch size composition | $1977-2016$ |
| Fishery | Catch per unit effort | $1991-2016$ |
| EBS shelf bottom trawl survey | Numerical abundance | $1982-2016$ |
| EBS shelf bottom trawl survey | Size composition | $1982-2016$ |
| EBS shelf bottom trawl survey | Age composition | $1994-2015$ |
| EBS shelf bottom trawl survey | Mean size at age | $1994-2015$ |
| NMFS longline survey | Relative population number | $1997-2015$ (odd years only) |
| NMFS longline survey | Size composition | $1997-2015$ (odd years only) |

## Fishery

## Catch Biomass

Catches taken in the EBS for the period 1977-2016 are shown for the three main gear types in Table 2.5. Table 2.5 makes use of two different types of season: catch seasons and selectivity seasons. The catch seasons are defined as January-February, March-April, May-July, August-October, and NovemberDecember. Three selectivity seasons are defined by combining catch seasons 1 and 2 into selectivity season 1, equating catch season 3 with selectivity season 2 , and combining catch seasons 4 and 5 into selectivity season 3. The catch seasons were the result of a statistical analysis described in the 2010 assessment (Thompson et al. 2010), and the selectivity seasons were chosen to correspond as closely as possible to the traditional seasons used in assessments prior to 2010 (given the revised catch seasons).

In years for which estimates of the distribution by gear or period were not available, proxies based on other years' distributions were used to create Table 2.5. Catches for the years 1977-1980 may or may not include discards.

The 2014 assessment included an evaluation of 12 methods for projecting year-end catch for the last year in the time series (Thompson 2014). It turned out that the best estimator was simply to set the current year's catch during seasons 4-5 equal to the previous year's catch during those same seasons (up to the TAC for the current year). In Table 2.5, catches for the August-October and November-December seasons of 2016 were estimated by this method, except that ABC was used as the upper limit rather than TAC, due to the $6.4 \%$ GHL in the State-managed fishery. The other catches shown in Table 2.5 consist of "official" data from the NMFS Alaska Region. However, other removals of Pacific cod are known to have occurred over the years, including removals due to subsistence fishing, scientific research, and fisheries managed under other FMPs. Estimates of such other removals are shown in Appendix 2.4.

## Catch Size Composition

Fishery size compositions are presently available, by gear, for at least one gear type in every year from 1977 through the first part of 2016. Beginning with the 2010 assessment (Thompson et al. 2010), size composition data have been based on 1-cm bins ranging from 4 to 120 cm . Because displaying these data would add a large number of pages to the present document, they are not shown here but are available at: http://www.afsc.noaa.gov/REFM/Docs/2016/EBS_Pcod_fishery_sizecomp_data.xlsx.

## Catch Per Unit Effort

Fishery catch per unit effort data are available by gear and season for the years 1991-2016 and are shown in Table 2.6. Units are $\mathrm{kg} /$ minute for trawl gear, $\mathrm{kg} / \mathrm{hook}$ for longline gear, and $\mathrm{kg} /$ pot for pot gear; data for 2016 are partial. The "sigma" values shown in the tables are intended only to give an idea of the relative variability of the respective point estimates, and are not actually used in any of the analyses presented here.

## Weight at Age

Four years of mean-weight-at-age data are available from the fishery. Most of these come from the longline fishery. The weight-at-age estimation procedure involves a two-stage bootstrap resampling of the data (James Ianelli, AFSC, pers. commun.). Observed tows were first selected with replacement, followed by re-sampling actual lengths and age specimens given those set of tows. Catch (in biomass) within each stratum is also used to compute the relative weights of length and age samples. For each bootstrap sample, the methods outlined in Kimura (1990) were used to compute estimates. The resulting data (in kg ) are shown below (no data available for ages 0 or 1 ):

| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 0.001 | 1.420 | 2.006 | 2.938 | 3.785 | 5.022 | 6.666 | 7.146 | 8.507 | 10.004 | 5.224 |
| 2009 | 0.524 | 1.482 | 2.139 | 3.092 | 3.981 | 5.259 | 5.535 | 8.927 | 8.715 | 7.876 | 7.993 |
| 2010 | 0.787 | 1.635 | 2.340 | 3.046 | 3.961 | 5.377 | 5.921 | 5.518 | 11.946 | 3.825 | 4.142 |
| 2011 | 0.001 | 1.278 | 2.210 | 3.244 | 4.256 | 5.637 | 7.529 | 6.177 | 3.018 | 4.445 | 3.537 |
| Ave: | 0.655 | 1.454 | 2.174 | 3.080 | 3.996 | 5.324 | 6.413 | 6.942 | 8.046 | 6.537 | 5.224 |

## Survey

## EBS Shelf Bottom Trawl Survey

## Population Indices

Strata 1-6 of the EBS shelf bottom trawl survey have been sampled annually since 1982, and comprise the standard survey area used in this assessment. Beginning in 1987, strata 8 and 9, located to the northwest of the standard survey area, have also been sampled annually. Although strata 8 and 9 do contain Pacific cod, the biomass contained in those strata is typically a small fraction of that contained in the overall survey area (i.e., strata 1-6 plus strata 8-9), averaging less than $3 \%$ over the time series. Rather than estimate separate catchability and selectivity parameters for the pre-1987 (strata 1-6) and post-1986 (strata 1-6 plus strata 8-9) portions of the time series, the assessment models for EBS Pacific cod have always used data from strata 1-6 only.

Estimates of total abundance (both in biomass and numbers of fish) obtained from the trawl survey are shown in Table 2.7, together with their respective standard errors. Upper and lower $95 \%$ confidence intervals are also shown for the biomass estimates. Survey results indicate that biomass remained relatively constant from 1982 through 1988. The highest biomass ever observed by the survey was the 1994 estimate of $1,368,120 \mathrm{t}$. Following the high observation in 1994, the survey biomass estimate declined steadily through 1998. The survey biomass estimates remained in the 596,000-619,000 t range from 2002 through 2005. However, the survey biomass estimates dropped after 2005, producing an alltime low in 2007 and again in 2008. Estimated biomass more than doubled between 2009 and 2010, then remained relatively stable for the next three years, followed by another large increase (36\%) in 2014,
which was sustained through 2015. The 2016 estimate of $944,621 \mathrm{t}$ represented a $14 \%$ drop relative to 2015, although it is still $20 \%$ above average for the time series.

Numerical abundance has shown more variability than biomass, with a mean squared relative inter-annual change of 0.31 for biomass and 0.36 for abundance. The estimates from 2009-2015 were uniformly above average; the estimate for each of those years was at least $7 \%$ above average for the time series, with a mean of $26 \%$ above average. The estimate for 2016, however, was $4 \%$ below average.

## Size and Age Composition

The size compositions from the EBS shelf bottom trawl survey for the years 1982-2016 are shown in Table 2.8 (actual numbers of fish measured are shown in column 2 on the first page). The data are shown according to the $1-\mathrm{cm}$ bins described above for fishery size composition data. Rows in Table 2.8 sum to the actual number of fish measured in each year (subject to slight rounding error).

Age compositions from the 1994-2015 surveys are available. The age compositions and actual sample sizes are shown in Table 2.9.

## Mean Length at Age and Weight at Age

Mean-length-at-age data are available for all of the years in which age compositions are available. These are shown, along with sample sizes, in Table 2.10.

Mean-weight-at-age data are also available for all of the years in which age compositions are available. These are shown in Table 2.11.

NMFS Longline Survey
Relative Population Number
Table 2.12 shows NMFS longline survey estimates of relative population number for 1997-2015 (oddnumbered years only), together with their respective standard errors and upper and lower $95 \%$ confidence intervals. The time series reached a high of 204,250 in 1997 and a low of 95,553 in 2009. The most recent estimate of 157,996 (in 2015) was $4 \%$ above the average for the time series.

## Size Composition

The size compositions from the NMFS longline survey for the years 1997-2015 are shown in Table 2.9b (actual numbers of fish measured are shown in column 2 in the upper panel). These data are shown according to the $1-\mathrm{cm}$ bins described above for fishery size composition data. Rows in Table 2.13 sum to the actual number of fish measured in each year (subject to slight rounding error).

## ANALYTIC APPROACH

## General Model Structure

Although Pacific cod in the EBS and AI were managed on a BSAI-wide basis through 2013, the stock assessment model has always been configured for the EBS stock only. Since 1992, the assessment model has always been developed under some version of the SS modeling framework (technical details given in Methot and Wetzel 2013; see especially Appendix A to that paper). Beginning with the 2005 assessment, the EBS Pacific cod models have all used versions of SS based on the ADMB software package (Fournier
et al. 2012). A history of previous model structures, including details of the model used to set harvest specifications for this year, is given in Appendix 2.3.

Version 3.24u (compiled on 08/29/14) of SS was used to run the models in this assessment. The relevant user manual can be obtained at https://www.st.nmfs.noaa.gov/Assets/science program/SS User Manual 3.24s.pdf .

## Description of Alternative Models

As in the final 2015 assessment and this year's preliminary assessment, model numbering follows the protocol given by Option A in the SAFE chapter guidelines. The goal of this protocol is to make it easy to distinguish between major and minor changes in models and to identify the years in which major model changes were introduced. Names of models constituting major changes get linked to the year that they are introduced (e.g., Model 11.5 is one of at least five models introduced in 2011 that constituted a major change from the then-current base model), while names of models constituting minor changes get linked to the model that they modify (e.g., Model 11.5 a would refer to a model that constituted a minor change from Model 11.5). Names of all final models adopted since the first application of an ADMB-based version of SS (in 2005) were translated according to the current naming convention in Table 2.11 of the 2015 assessment (Thompson 2015).

This year's preliminary assessment included Model 11.5, which has been the accepted model since 2011, and five new models (Models 16.1-16.5). Per request of the BSAI Plan Team and SSC (see comments BPT1 and SSC19), six models are presented in this final assessment: Models 11.5 and Model 16.1 from the preliminary assessment, and four variants of Model 16.1 (the paragraphs following this list describe the features of these models in greater detail):

- Model 16.6: Model 16.1 without empirical weight at age
- Model 16.7: Model 16.1 without empirical weight at age and including the NMFS LL survey
- Model 16.8: Model 16.1 with time-varying survey selectivity
- Model 16.9: Model 16.1 with time-varying fishery selectivity

Detailed descriptions of Models 11.5 and 16.1 are given in Appendix 2.1, and a comparison of their key features is provided in Table 2.14. Basically, Model 11.5 is a multi-gear, multi-season model using a time-variant, potentially domed (6-parameter double normal) selectivity function, with the natural mortality rate $(M)$ and the trawl survey catchability coefficient $(Q)$ fixed outside the model (the latter based on the results of Nichol et al. (2007)); while Model 16.1 is single-gear, single-season model using a time-invariant, asymptotic (2-parameter logistic) selectivity function, with $M$ and $Q$ estimated inside the model.

Empirical weight at age was first explored for the EBS Pacific cod stock in this year's preliminary assessment. Some key similarities and differences between the models without empirical weight at age (Models 11.5, 16.6, and 16.7) and those with empirical weight at age (Models 16.1, 16.8, and 16.9) are as follow: All six models estimate (internally) a time-invariant relationship between mean length and age, which is used for fitting the size composition data, among other things. Models without empirical weight at age use externally estimated parameters describing a weight-at-length relationship (seasonally varying but constant across years in the case of Model 11.5, annually varying in the cases of Models 16.6 and 16.7) in combination with the internally estimated length-at-age relationship to compute weight at age. Models with empirical weight at age bypass the link between weight at age and length at age, and instead use externally estimated, time-varying schedules of weight at age directly.

In Model 16.7, logistic selectivity was assumed for the NMFS longline survey, just as for fishery and trawl survey selectivity.

Time-varying selectivity in Models 16.8 and 16.9 was implemented in the form of annual deviations from a base selectivity function. The "sigma" parameters governing the extent to which selectivity devs can vary from zero (specified as inputs to the model, not estimated internally) in Models 16.8 and 16.9 were set at large values to maximize those models' ability to fit the data, essentially treating each dev as an unconstrained parameter. Values of the sigma parameters were increased across several trial runs of each model until the resulting estimate of 2016 spawning biomass did not change (to 3 significant digits) with further increases (see "Parameter Estimates" in the "Results" section).

As in previous assessments, development of the final versions of all models included calculation of the Hessian matrix and a requirement that all models pass a "jitter" test of 50 runs. In the event that a jitter run produced a better value for the objective function than the base run, then:

1. The model was re-run starting from the final parameter file from the best jitter run.
2. The resulting new control file, with the parameter estimates from the best jitter run incorporated as starting values, became the new base run.
3. The entire process (starting with a new set of jitter runs) was repeated until no jitter run produced a better value for the objective function than the most recent base run.

The preliminary assessment described a change in the method used for the jitter analysis, involving an attempt to standardize the bounds within which individual parameters were "jittered." Specifically, once a model was ready to be subjected to the jitter test, the bounds for each parameter in the model were adjusted to match the $99.9 \%$ confidence interval (based on the normal approximation obtained by inverting the Hessian matrix). A jitter rate (equal to half the standard deviation of the logit-scale distribution from which "jittered" parameter values are drawn) was set at 1.0 for all models. Standardizing the jittering process in this manner may not explore parameter space as thoroughly as in previous assessments; however, it should make the jitter rate more interpretable, and show the extent to which the identified minimum (local or otherwise) is well behaved.

Except for $d e v$ vectors in all models, all parameters were estimated with uniform prior distributions. All devs were assumed to be additive.

## Parameters Estimated Outside the Assessment Model

## Natural Mortality

A value of 0.34 has been used for the natural mortality rate $M$ in all BSAI Pacific cod stock assessments since 2007. This value was based on Equation 7 of Jensen (1996) and an age at maturity of 4.9 years (Stark 2007). In response to a request from the SSC, the 2008 assessment included a discussion of alternative values and a justification for the value chosen (Thompson et al. 2008). However, it should be emphasized that, even if Jensen's Equation 7 is exactly right, variability in the estimate of the age at maturity implies that the point of estimate of 0.34 is accompanied by some level of uncertainty. Using the variance for the age at $50 \%$ maturity published by Stark (0.0663), the $95 \%$ confidence interval for $M$ extends from about 0.30 to 0.38 .

The value of 0.34 adopted in 2007 replaced the value of 0.37 that had been used in all BSAI Pacific cod stock assessments from 1993 through 2006.

For historical completeness, some other published estimates of $M$ for Pacific cod are shown below:

| Area | Author | Year | Value |
| :--- | :--- | :--- | :--- |
| Eastern Bering Sea | Low | 1974 | $0.30-0.45$ |
|  | Wespestad et al. | 1982 | 0.70 |
|  | Bakkala and Wespestad | 1985 | 0.45 |
|  | Thompson and Shimada | 1990 | 0.29 |
|  | Thompson and Methot | 1993 | 0.37 |
| Gulf of Alaska | Thompson and Zenger | 1993 | 0.27 |
|  | Thompson and Zenger | 1995 | 0.50 |
| British Columbia | Ketchen | 1964 | $0.83-0.99$ |
|  | Fournier | 1983 | 0.65 |

Model 11.5 in this assessment fixes $M$ at the value of 0.34 used since 2007. Models in the 16.x series estimate $M$ internally.

## Variability in Estimated Age

Variability in estimated age in SS is based on the standard deviation of estimated age between "reader" and "tester" age determinations. Weighted least squares regression has been used in the past several assessments to estimate a proportional relationship between standard deviation and age. The regression was recomputed this year over ages 1 through 13, yielding an estimated slope of 0.0852 (i.e, the standard deviation of estimated age was modeled as $0.0852 \times$ age) and a weighted $R^{2}$ of 0.98 . This regression corresponds to a standard deviation at age 1 of 0.085 and a standard deviation at age 20 of 1.705. These parameters were used for the models in the present assessment.

## Weight at Length

Long-term base values along with annual and seasonal deviations of the parameters governing the weight-at-length schedule were estimated in the 2012 assessment using the method described in Annex 2.1.2 of Thompson and Lauth (2012), based on fishery data collected from 1974 through 2011. The same method was used this year to update all weight-length parameters using fishery data through 2015.

Using the functional form weight $=\alpha \times$ length $^{\beta}$, where weight is measured in kg and length is measured in cm , the long-term base values for the parameters were estimated as $\alpha=5.64006 \mathrm{E}-06$ and $\beta=3.183145$.

Seasonal additive log-scale offsets from the base parameter values (used in Model 11.5 only) were reestimated in this year's preliminary assessment, resulting in the following values:

| Season: | Jan-Feb | Mar-Apr | May-Jul | Aug-Oct | Nov-Dec |
| :--- | ---: | ---: | ---: | ---: | ---: |
| $\alpha:$ | $-2.277 \mathrm{E}-02$ | $2.893 \mathrm{E}-03$ | $1.913 \mathrm{E}-02$ | $2.261 \mathrm{E}-03$ | $-1.416 \mathrm{E}-02$ |
| $\beta:$ | $5.219 \mathrm{E}-03$ | $-6.735 \mathrm{E}-04$ | $-4.500 \mathrm{E}-03$ | $-5.263 \mathrm{E}-04$ | $3.262 \mathrm{E}-03$ |

Models in the 16.x series allow for inter-annual, but not intra-annual, variability in weight-length parameters. Values of annual additive offsets from the base $\alpha$ and $\beta$ values are shown in Table 2.15.

## Weight at Age

Begin-year mean weight at age in the population was estimated outside the assessment model by linearly interpolating the survey mean weights at age shown in Table 2.11. The results of the interpolation are shown in Table 2.16. These values were used in Models 16.1, 16.8, and 16.9. Weight at age in Models $11.5,16.6$, and 16.7 was determined by the internally estimated length-at-age parameters and the externally estimated weight-at-length parameters described immediately above.

## Maturity

A detailed history and evaluation of parameter values used to describe the maturity schedule for BSAI Pacific cod was presented in the 2005 assessment (Thompson and Dorn 2005). A length-based maturity schedule was used for many years. The parameter values used for the length-based maturity schedule in the 2005 and 2006 assessments were set on the basis of a study by Stark (2007) at the following values: length at $50 \%$ maturity $=58 \mathrm{~cm}$ and slope of linearized logistic equation $=-0.132$. However, in 2007, changes in SS allowed for use of either a length-based or an age-based maturity schedule. Beginning with the 2007 assessment, the accepted model has used an age-based schedule with intercept $=4.88$ years and slope $=-0.965$ (Stark 2007). The use of an age-based rather than a length-based schedule follows a recommendation from the maturity study's author (James Stark, Alaska Fisheries Science Center, pers. commun.). The age-based parameters were retained for the models in the present assessment.

## Stock-Recruitment "Steepness"

Following the standard Tier 3 approach, both models assume that there is no relationship between stock and recruitment, so the "steepness" parameter is set at 1.0 in each.

## Parameters Estimated Inside the Assessment Model

A total of 190 parameters were estimated inside SS for Model 11.5. These include:

1. all three von Bertalanffy growth parameters
2. standard deviation of length at ages 1 and 20
3. mean ageing bias at ages 1 and 20
4. log mean recruitment since the 1976-1977 regime shift
5. offset for log-scale mean recruitment before the 1976-1977 regime shift
6. devs for log-scale initial (i.e., 1977) abundance at ages 1 through 3
7. annual log-scale recruitment devs
8. initial (equilibrium) fishing mortality for the Jan-Apr trawl fishery
9. base values for all trawl survey selectivity parameters
10. gear-, season-, and-block-specific selectivity parameters for nine fisheries
11. annual devs for the ascending_width parameter of the trawl survey selectivity function

Parameter counts for models in the 16.x series were as follow:

| Model 16.1 | Model 16.6 | Model 16.7 | Model 16.8 | Model 16.9 |
| ---: | ---: | ---: | ---: | ---: |
| 77 | 77 | 80 | 109 | 157 |

Parameters in items \#1-9 in the above list were also estimated by all models in the 16.x series, except that the initial fishing mortality rate pertained to the entire fishery, not just the Jan-Apr trawl fishery, and the trawl survey selectivity function involved only two parameters rather than five (the single exception occurred in Model 16.8, where the parameter representing the difference in the ages of $95 \%$ and $50 \%$ selection was fixed at a value of 0.01 -see "Parameter Estimates" in the "Results" section). In addition, the following parameters were also estimated by all models in the 16.x series:

1. natural mortality rate
2. Richards growth parameter
3. log-scale recruitment standard deviation $\left(\sigma_{R}\right)$
4. devs for log-scale initial (i.e., 1977) abundance at ages 4 through 20
5. log catchability for the trawl survey
6. base values for both fishery selectivity parameters

Model-specific parameters in the 16.x series were as follow:

- $\quad \log$ catchability for the NMFS longline survey (Model 16.7)
- values for both parameters of the NMFS longline survey selectivity function (Model 16.7)
- annual devs for the $A 50 \%$ parameter of the trawl survey selectivity function (Model 16.8)
- annual devs for both parameters of the fishery selectivity function (Model 16.8)

In all models, uniform prior distributions were used for all parameters, except that dev vectors were constrained by input standard deviations, which are somewhat analogous to a joint prior distribution.

For all parameters estimated within individual SS runs, the estimator used was the mode of the logarithm of the joint posterior distribution, which was in turn calculated as the sum of the logarithms of the parameter-specific prior distributions and the logarithm of the likelihood function.

In addition to the above, the full set fishing mortality rates were also estimated internally (year-, season-, and gear-specific for Model 11.5; only year-specific for Models in the 16.x series), but not in the same sense as the above parameters. The fishing mortality rates are determined (almost) exactly as functions of other model parameters, because SS assumes that the input total catch data are true values rather than estimates, so the fishing mortality rates can be computed algebraically given the other parameter values and the input catch data. An option does exist in SS for treating the fishing mortality rates as full parameters, but previous explorations have indicated that adding these parameters has almost no effect on other model output (Methot and Wetzell 2013).

## Objective Function Components

All models in this assessment include likelihood components for catch, initial (equilibrium) catch, trawl survey relative abundance, fishery and trawl survey size composition, survey age composition, recruitment, "softbounds" (equivalent to an extremely weak prior distribution used to keep parameters from hitting bounds), and parameter deviations. In addition, Model 11.5 includes an objective function component called "F ballpark," which acts like a weak prior distribution on fishing mortality in a userspecified year.

In SS, emphasis factors are specified to determine which likelihood components receive the greatest attention during the parameter estimation process. As in previous assessments, all likelihood components were given an emphasis of 1.0 here.

## Use of Size Composition Data in Parameter Estimation

Size composition data are assumed to be drawn from a multinomial distribution specific to a particular year, gear, and season within the year (Model 11.5) or just year (Models 16.x). In the parameter estimation process, SS weights a given size composition observation according to the emphasis associated with the respective likelihood component and the sample size specified for the multinomial distribution from which the data are assumed to be drawn. In developing the model upon which SS was originally based, Fournier and Archibald (1982) suggested truncating the multinomial sample size at a value of 400 in order to compensate for contingencies which cause the sampling process to depart from the process that gives rise to the multinomial distribution. For many years, the Pacific cod assessments assumed a multinomial sample size equal to the square root of the true length sample size, rather than the true length sample size itself. Given the true length sample sizes observed in the EBS Pacific cod data, this
procedure tended to give values somewhat below 400 while still providing SS with usable information regarding the appropriate effort to devote to fitting individual length samples.

Although the "square root rule" for specifying multinomial sample sizes gave reasonable values, the rule itself was largely ad hoc. In an attempt to move toward a more statistically based specification, the 2007 assessment used the harmonic means from a bootstrap analysis of the available fishery length data from 1990-2006 (Thompson et al. 2007). The harmonic means were smaller than the actual sample sizes, but still ranged well into the thousands. A multinomial sample size in the thousands would likely overemphasize the size composition data. As a compromise, the harmonic means were rescaled proportionally in the 2007 assessment so that the average value (across all samples) was 300 . However, the question then remained of what to do about years not covered by the bootstrap analysis (2007 and pre1990) and what to do about the survey samples. The solution adopted in the 2007 assessment was based on an observed consistency in the ratios between the harmonic means (the raw harmonic means, not the rescaled harmonic means) and the actual sample sizes: Whenever the actual sample size exceeded about 400 fish, for the years prior to 1999 the ratio was very consistently close to 0.16 , and for the years after 1998 the ratio was very consistently close to 0.34 .

This consistency was used to specify the missing values as follows: For fishery data, records with actual sample sizes less than 400 were omitted. Then, the sample sizes for fishery length compositions from years prior to 1999 were tentatively set at $16 \%$ of the actual sample size, and the sample sizes for fishery length compositions from 2007 were tentatively set at $34 \%$ of the actual sample size. For the pre-1982 trawl survey, length compositions were tentatively set at $16 \%$ of an assumed sample size of 10,000 . For the post-1981 trawl survey length compositions, sample sizes were tentatively set at $34 \%$ of the actual sample size. Then, with sample sizes for fishery length compositions from 1990-2007 tentatively set at their bootstrap harmonic means (not rescaled), all sample sizes were adjusted proportionally so that the average was 300 .

The same procedure was used in the 2008 and 2009 assessments. For the 2010 assessment, however, this procedure had to be modified somewhat, because the bootstrap values for the 1990-2006 size composition data did not match the new bin and seasonal structures. To be as consistent as possible with the approach used to set sample sizes in the 2008 and 2009 assessments, the 2010 and 2011 assessments set sample sizes by applying the $16 / 34 \%$ rule for all size composition records with actual sample sizes greater than 400 (not just those lying outside the set of 1990-2006 fishery data), then rescaling proportionally to achieve an average sample size of 300. The same procedure was used for the 2012-2015 assessments, except the pre-1982 trawl survey data were no longer used. Model 11.5 in this year's assessment uses the same procedure as the 2012-2015 assessments. Models $16 . x$ use a similar procedure, except that the input sample sizes for the fishery and trawl survey (and NMFS longline survey, in the case of Model 16.7) are scaled so that the average is 300 for each, rather than 300 for all size composition data combined. The full sets of input sample sizes are shown in Table 2.17.

## Use of Age Composition Data in Parameter Estimation

Like the size composition data, the age composition data are assumed to be drawn from a multinomial distribution specific to a particular gear, year, and season within the year. Age composition data are input in the same way for all of the models. Input sample sizes for the multinomial distributions were computed by scaling the actual number of otoliths read in each year (Table 2.9, column 2) proportionally such that the average of the input sample sizes was equal to 300 , giving the following:

| Year: | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| N: | 204 | 163 | 203 | 205 | 181 | 246 | 246 | 263 | 248 | 361 | 284 |
| Year: | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| $\mathrm{~N}:$ | 365 | 371 | 412 | 346 | 403 | 369 | 358 | 372 | 405 | 349 | 244 |

## Use of Fishery CPUE and Survey Relative Abundance Data in Parameter Estimation

Fishery CPUE data are included in the Model 11.5 for comparative purposes only, and are not included at all in Models 16.x. Their respective catchabilities (in Model 11.5) are estimated analytically, not statistically.

For the surveys, each year's survey abundance estimate is assumed to be drawn from a lognormal distribution specific to that year. The model's estimate of survey abundance in a given year serves as the geometric mean for that year's lognormal distribution, and the ratio of the survey abundance estimate's standard error to the survey abundance estimate itself serves as the distribution's coefficient of variation, which is then transformed into the "sigma" parameter for the lognormal distribution.

## Use of Recruitment Deviation "Data" in Parameter Estimation

The likelihood component for recruitment is different from traditional likelihoods because it does not involve "data" in the same sense that traditional likelihoods do. Instead, the log-scale recruitment dev plays the role of the datum in a normal distribution with mean zero and specified (or estimated) standard deviation; but, of course, the devs are parameters, not data.

## RESULTS

## Model Evaluation

The two models used in this assessment are described under "Model Structure" above.
Goodness of Fit, Parameter Estimates, and Derived Quantities
Goodness of Fit
Table 2.18 shows the objective function value for each data component and sub-component in each model. The first part of the table shows negative log-likelihoods for the aggregate data components. The second and third parts of the table break down the size composition and survey abundance index components into fleet-specific values. Models 16.1, 16.7, and 16.8 use the same data sets (including the empirical weight-at-age data), so their objective function values are directly comparable. However, each data set used in the other three models is unique to the respective model, so the objective function values from Models 11.5, 16.6, and 16.7 are not comparable to any other model. Table 2.18 also shows parameter counts for the models, including separate counts for true parameters and constrained devs.

Table 2.19 provides alternative measures of how well the model fits the fishery CPUE (Model 11.5 only) and survey relative abundance data. The column labeled "oave" shows the average of the log-scale standard deviations provided as part of the data. The four right-hand columns show root mean squared errors (RMSE; values closer to бave are better), mean normalized residuals (MNR; values closer to zero are better), standard deviations of normalized residuals (SDNR; values closer to unity are better), and correlations between observed and estimated values (values to unity are better). The first 9 rows of Table 2.19 pertain to the fishery CPUE data. Although Model 11.5 does not actually attempt to fit these data
(only the survey CPUE are used), of the 9 correlations with fishery CPUE, all but one are positive. The next 6 lines pertain to the trawl survey index data, which all of the models try to fit. Except for Model 11.5, all of the models give MNRs fairly close to zero for the trawl survey data. However, all of the models give RMSEs quite a bit higher than oave and SDNRs quite a bit higher than unity for the trawl survey data. The last row pertains to the NMFS longline survey (Model 16.7 only).

Figure 2.4a shows the models’ fits to the trawl survey abundance data. The proportions of years in which each model's estimate falls within the respective 95\% confidence interval are shown below:

Model 11.5 $\quad$ Model 16.1 $\quad$ Model 16.6 $\quad$ Model 16.7 | Model 16.8 | Model 16.9 |  |
| ---: | ---: | ---: |
| 0.80 | 0.77 | 0.74 |

Figure 2.4b shows Model 16.7's fit to the NMFS longline survey abundance data. Model 16.7’s estimates fall within the respective $95 \%$ confidence interval $80 \%$ of the time.

Table 2.20 shows how output "effective" sample sizes ("Neff," McAllister and Ianelli 1997) compare to input sample sizes ("Ninp") for the size composition data. Three sets of ratios are provided for each fleet: 1) the arithmetic mean ("A") of the Neff/Ninp ratio, 2) the ratio of arithmetic mean Neff to arithmetic mean Ninp, and 3) the ratio of harmonic mean ("H") Neff to arithmetic mean Ninp. All of the models give ratios greater (usually much greater) than or equal to unity for all cases for all three measures, except for the Aug-Dec longline fishery in Model 11.5, where the ratio with the harmonic mean in the numerator is only 0.91 . In the case of the survey(s), the ratio with the harmonic mean in the numerator ranges is either equal to or slightly greater than unity for all models (range: 1.00 to 1.15).

Table 2.21 provides a similar analysis for the age composition data, except that the rows in the main part of this table correspond to individual records rather than fisheries or surveys (all age composition data in the models come from the trawl survey). The bottom two rows in the table show the ratios of the means (using the arithmetic mean as the numerator in the next-to-last row and the harmonic mean in the last row). For Models 11.5, 16.1, and 16.6, the ratio with the harmonic mean in the numerator is quite a bit less than unity. For Model 16.1, even the ratio with the arithmetic mean in the numerator is quite a bit less than unity.

The models' fits to the age composition data are shown for each year in Figure 2.5 and aggregated across years in Figure 2.6. Because of the large number of size composition records ( $\mathrm{n}=459$ for Model 11.5, $\mathrm{n}=75$ for all models in the 16.x series except Model 16.7, $\mathrm{n}=85$ for Model 16.7), figures showing the models' fits to each record are not included in this document, but are available at: http://www.afsc.noaa.gov/REFM/Docs/2016/EBS_Pcod_sizecomp_fits.xlsx. Time-aggregated size composition fits are shown in Figure 2.7.

Estimates of mean size at ages 1 through 3 (at the time of the survey) from the model are compared to the long-term average survey size composition (through 50 cm ) in Figure 2.8. All of the models tend to match the modes, within one or two cm (the estimates from the models in the $16 . \mathrm{x}$ series tend to be so close to each other that only the minima and maxima for those models are shown). Model 11.5’s fits to the mean-size-at-age data are shown in Figure 2.9 (recall that this model does not actually attempt to fit these data, and the models in the 16.x series do not include these data at all).

## Parameter Estimates

Table 2.22 displays all of the parameters (except fishing mortality rates, because these are functions of other parameters) estimated internally in the model, along with the standard deviations of those estimates. Table 2.22 consists of the following parts:

- Table 2.22a shows scalar parameters for all models
- Table 2.22b shows devs for the initial (1977) age composition for all models
- Table 2.22c shows annual log-scale recruitment devs for all models
o These are plotted in Figure 2.10
- Table 2.22d shows fishery selectivity parameters for Model 11.5
- Table 2.22e shows survey selectivity parameters (including annual devs) for Model 11.5
- Table 2.22 f shows main selectivity parameters for models in the $16 . \mathrm{x}$ series
- Table 2.22g shows annual selectivity devs for Models 16.8 and 16.9
o The values in this part of the table may be difficult to interpret, because the estimates are logit-transformed within the bounds ( $-10,10$ ), with very large input sigma values ( $\sigma=10$ for A50\% in Model 16.8, $\sigma=100$ for both parameters in Model 16.9)

As noted under "Parameters Estimated Inside the Assessment Model" in the "Analytic Approach" section, the parameter representing the difference in the ages of $95 \%$ and $50 \%$ selection for the trawl survey in Model 16.8 was fixed at a value of 0.01 . Fixing the value of this parameter was necessary because otherwise it tended to approach zero and the value of the A50\% parameter tended to approach unity, meaning that even very small devs produced large changes in selectivity at age 1 , and rounding errors appeared to be a problem.

The log-scale trawl survey catchability estimates shown in Table 2.22a imply the following values of $Q$ on the back-transformed (natural) scale:

| Model 11.5 |  | Model 16.1 |  | Model 16.6 |  | Model 16.7 |  | Model 16.8 |  | Model 16.9 |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Est. | CV | Est. | CV | Est. | CV | Est. | CV | Est. | CV | Est. | CV |
| 0.77 | n/a | 0.61 | 0.062 | 0.88 | 0.065 | 1.03 | 0.056 | 0.66 | 0.056 | 0.61 | 0.061 |

Table 2.23 shows estimates of fishing mortality. Table 2.23a shows fishing mortality by year in all models, and Table 2.23b shows full-selection seasonal fishing mortality rates for each gear, season, and year in Model 11.5 only.

## Derived Quantities

Figure 2.11a shows the time series of female spawning biomass relative to $B_{100 \%}$ as estimated by each model, and Figure 2.11b shows the time series of total biomass as estimated by each model, along with the time series of observed survey biomass. Average (across years) ratios of total biomass (as estimated by the models) to survey biomass (as specified in the data) are shown below:

## $\begin{array}{rrrrrr}\text { Model 11.5 } & \text { Model 16.1 } & \text { Model 16.6 } & \text { Model 16.7 } & \text { Model 16.8 } & \text { Model } 16.9 \\ 1.74 & 1.32 & 1.24 & 1.05 & 1.21 & 1.36\end{array}$

Figure 2.12a shows survey selectivity as estimated by the models (base values for Models 11.5 and 16.8, as those models exhibit time-varying survey selectivity). Model 11.5 allows survey selectivity to be dome-shaped, while all models in the 16.x series force it to be asymptotic. Figure 2.12b shows how trawl survey selectivity varies over time in Models 11.5 and 16.8.

Figure 2.13a shows gear-, season-, and block-specific fishery selectivity as estimated by Model 11.5. In general, selectivities that are not forced to be asymptotic tend to show decreasing selectivity at large size in Model 11.5. Figure 2.13b shows fishery selectivity as estimated by the models in the $16 . x$ series (base values for Model 16.9; note that Model 11.5 does not include base values for fishery selectivity, because
each block-specific fishery selectivity parameter is estimated independently). Figure 2.13c shows how fishery selectivity varies over time in Model 16.9.

Table 2.24 contains selected output from the standard projection model, based on SS parameter estimates from the two models, along with the probability that the maximum permissible ABC in each of the next two years will exceed the corresponding true-but-unknown OFL and the probability that the stock will fall below $B_{20 \%}$ in each of the next five years (probabilities are given by SS rather than the standard projection model). Note that some of the quantities in Table 2.24 are conditional on catches estimated under Scenario 2 ("author's F") in the "Harvest Recommendations" section.

## Choice of Final Model

## Retrospective Analysis

The SSC has recommended that retrospective performance be considered in the selection of a final model (comment SSC7). Retrospective analyses for all of the models are shown in Figure 2.14. Values of $\rho$ (Mohn 1999, equation corrected in the 2013 Retrospective Working Group report) are shown below:

| Model: | 11.5 | 16.1 | 16.6 | 16.7 | 16.8 | 16.9 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| $\rho:$ | 0.475 | 0.194 | 0.147 | 0.144 | 0.094 | 0.250 |

Model 16.8 has the lowest value of $\rho(0.094)$, and Model 11.5 has by far the highest ( 0.475 ).
Although any amount of retrospective bias is undesirable, eliminating such bias entirely is typically an extremely difficult task, which raises the question of how much retrospective bias is acceptable. HurtadoFerro et al. (2015) suggest that, for a stock with a natural mortality rate of 0.2 , values of $\rho$ higher than 0.20 or lower than -0.15 should be "cause for concern," while for a stock with a natural mortality rate of 0.4 , values of $\rho$ higher than 0.30 or lower than -0.22 should be "cause for concern." Interpolating the upper limits (because all values of $\rho$ in the above table are positive) of this rule of thumb gives the relationship $\rho \max =0.1+0.5 M$. The values of $M$ estimated by the models (or assumed, in the case of Model 11.5) imply the following values of $\rho$ max:

| Model: | 11.5 | 16.1 | 16.6 | 16.7 | 16.8 | 16.9 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| M: | 0.340 | 0.378 | 0.363 | 0.344 | 0.375 | 0.376 |
| مmax: | 0.270 | 0.289 | 0.282 | 0.272 | 0.288 | 0.288 |

Model 11.5 is the only model where $\rho$ exceeds the $\rho$ max associated with the model's value of $M$.

## Other Considerations Regarding Individual Models

All of the models give good fits to the size composition data, but only models $16.7,16.8$, and 16.9 give good fits to the age composition data (based on the ratio of the harmonic mean effective sample size to the arithmetic mean input sample size), and none of the models gives a particularly good fit to the trawl survey abundance data (insofar as they all give RMSEs at least $49 \%$ higher than oave and SDNRs of at least 1.47).

Based on AIC, Model 16.9 would be strongly preferred over 16.1, and Model 16.8 would be strongly preferred over either Model 16.1 or Model 16.9. Use of AIC to make comparisons involving any of the other models is not meaningful, due to the fact that they use different data sets.

Only two of the models (11.5 and 16.9) allow time-varying fishery selectivity. Lack of time-varying fishery selectivity in the other four models may be problematic, given that the various gear types likely have different selectivity schedules and the proportions of the catch taken by the various gear types has changed considerably over time. For example, over the period 1991-2016, the proportions of the catch taken by trawl, longline, and pot gear have ranged from 0.24-0.62, 0.37-0.66, and 0.02-0.14, respectively.

Similarly, only two of the models (11.5 and 16.8) allow time-varying survey selectivity, and none of the models allow time-varying survey catchability. Lack of time-varying survey selectivity or catchability may be problematic, given that none of the models gives an acceptable fit to the trawl survey index.

Also on the subject of selectivity, none of the models in the 16.x series allows for the possibility of domeshaped selectivity for either the trawl survey or the fishery (or the NMFS longline survey, in the case of Model 16.7), whereas the BSAI Plan Team and SSC have recently supported allowing for this possibility, at least for the trawl survey:

BPT (9/15): "The model estimates of lower survey selectivity at larger sizes/ages result from the subsequent commercial catches of larger fish that must have been present at the time of the survey but were not caught in the survey in proportion to their abundance, so dome-shaped survey selectivity seems inescapable."

SSC (10/16): "The SSC notes that, in spite of the concerns over dome-shaped selectivity in the survey, there are many potential mechanisms relating to the availability of larger fish to the survey gear that could result in these patterns, regardless of the efficiency of the trawl gear to capture large fish in its path."

Models 16.1, 16.8, and 16.9 fix the time series of weight at age at externally estimated values, rather than using internally estimated length at age (time-invariant for all models) and externally estimated weight at length (time-invariant in the case of Model 11.5, time-varying in the cases of Models 16.6 and 16.7) to determine the time series of weight at age. Assuming that the estimates are accurate, the main advantage of using externally estimated weight at age is that this method integrates any changes in the length-at-age and weight-at-length relationships without having to estimate them inside the model. Disadvantages (in the context of the present assessment) include the following:

1. No smoothing was applied to the estimates, even though they exhibit a fair amount of variability, at least some of which seem implausible. For example, $10 \%$ of the within-cohort changes in weight from ages $a$ to $a+1$ are negative.
2. Age data exist for only 18 of the 35 years in the survey time series and only 4 of the 39 years in the fishery time series. Long-term averages were used for all years with no age data.
3. The fishery age data come primarily from the longline fishery, and may not be representative of the overall fishery.
4. Because the trawl survey takes place in summer, beginning-of-year population weights at age were calculated by averaging mid-year weight(age,year) and mid-year weight(age-1,year-1), implying that weight at age changes linearly within each one-year interval.
5. Consistent with the last several assessments, all of the models in this year's assessment estimate a positive ageing bias (Table 2.22a), a finding which was recently confirmed by Kastelle et al. (2017) on the basis of stable isotope analysis, meaning that the empirical weights at age are likely biased downward.

It may be advisable to examine more statistically sophisticated approaches for estimating weight at age outside of the assessment model, such as those that have been explored for the EBS walleye pollock assessment (Ianelli, this volume).

Also on the subject of statistical sophistication, it may be noted that considerable effort has been expended in the last five years toward developing alternatives to Model 11.5 that incorporate state-of-theart statistical methodology. In contrast, all of the models in the 16.x series eschew these developments in favor of ad hoc definitions of "reasonable" fits to composition data and "non-strange" selectivity patterns.

## Model Averaging Considerations

Although none of the models included in this assessment contains all of the features or exhibits all of the performances that might be desired, it is still necessary to choose a final model, and model averaging considerations may provide some guidance for doing so.

In the context of the EBS Pacific cod assessment models, the SSC's first reference to use of model averaging came in December, 2008:
"Consider the strengths and weaknesses of model averaging as an alternative to model selection...."

The above request resulted in the inclusion of a discussion of model averaging in the 2009 assessment (Thompson et al. 2009, p. 401-403). At the time of the 2009 assessment, the practice was to include, to the extent possible, every model that was requested by anyone. One of the concerns expressed in the discussion was that the resulting set of models might be biased, in which case a model averaging approach could easily lead to a worse estimate than simply choosing the best single model. The discussion concluded with the statement, "Therefore, even though model averaging appears to have considerable potential in principle, the approach should not be implemented for the Pacific cod assessments until outstanding issues, such as a protocol for choosing a representative set of models, have been resolved." However, given that the set of models included in the present assessment was the result of a formal, scientific vetting process, the concern about possible bias should be lessened somewhat, even though the resulting set of models is still, in all likelihood, significantly non-random.

Individual members of the SSC have advocated a model averaging approach for the EBS Pacific cod assessment at various times during the last few years, with the SSC as a whole making the following recommendation at this year's October meeting (comment SSC18):
> "The observed discrepancies among different models in these assessments are a good - if perhaps extreme - example of the model uncertainty that pervades most assessments. This uncertainty is largely ignored once a model is approved for specifications. We encourage the authors and Plan Teams to consider approaches such as multi-model inference to account for at least some of the structural uncertainty."

Another potential difficulty with a model averaging approach involves reconciling such an approach with the management framework described in the FMP. The SSC acknowledged this potential difficulty at this year's June meeting:
"The time may be right for a workshop ... on how to select and weight models for ensemble modeling and how to use an ensemble approach with our current harvest control rules" (emphasis added).

As an appropriate method for using a full model averaging approach in the context of the current management framework has yet to be determined, a possible short-term compromise would be to choose the single model that gives a 2017 maximum permissible ABC closest to the average across all models. This implies an equal weighting of models, which is a departure from traditional model averaging
technique (e.g., Buckland et al. 1997, Hoeting et al. 1999, Burnham and Anderson 2004). However, Stewart and Martell (2015) argued that developing a statistically rigorous method for weighting assessment models with different data sets and different likelihood functions is an extremely challenging task, and that equal weighting may prove to be a reasonable way forward for the time being, particularly if the models in the ensemble have been chosen carefully. The average 2017 maximum permissible ABC across all models is $246,500 \mathrm{t}$. If it is determined that Model 11.5 is no longer credible, the average across all models in the 16.x series is 228,200 $t$. In either case, the single model whose 2017 maximum permissible ABC comes closest to the average is Model 16.6 (2017 maximum permissible ABC = 239,000 t).

## Final Model: Conclusion

Given that each of the models has something to commend it but each also leaves something to be desired, and that a full model averaging approach does not seem possible at this time, it is reasonable to choose Model 16.6 as the final model, because its 2017 maximum permissible ABC comes closest to the average across all models.

In addition to the within-model retrospective analyses shown in Figure 2.14, Figure 2.15 provides a retrospective look at how the estimated spawning biomass time series has changed between assessments since 2006 (i.e., considering changes in both data and model), as requested by the SSC (see comment SSC12). Note that major model changes occurred in 2007, 2008, 2010, 2011, and 2016; and a minor model change occurred in 2009.

## Final Parameter Estimates and Associated Schedules

As noted previously, estimates of all statistically estimated parameters (except fishing mortality rates) are shown for all models in Table 2.22. Estimates of annual fishing mortality rates are shown for all models in Table 2.23a.

Schedules of begin-year length at age, mid-year length at age, and selectivity at age (both fishery and trawl survey) from Model 16.6 are shown in Table 2.25.

Schedules of time-varying weight at age from Model 16.6 are shown in Table 2.26.

## Time Series Results

## Definitions

The biomass estimates presented here will be defined in three ways: 1) age $0+$ biomass, consisting of the biomass of all fish aged 0 years or greater in January of a given year; 2) age $3+$ biomass, consisting of the biomass of all fish aged 3 years or greater in January of a given year; and 3) spawning biomass, consisting of the biomass of all spawning females in a given year. The recruitment estimates presented here will be defined as numbers of age 0 fish in a given year. To supplement the full-selection fishing mortality rates already shown in Table 2.23, an alternative "effective" fishing mortality rate will be provided here, defined for each age and time as $-\ln \left(N_{a+1, t+1} / N_{a, t}\right)-M$, where $N=$ number of fish, $a=$ age measured in years, $t=$ time measured in years, and $M=$ instantaneous natural mortality rate. In addition, the ratio of full-selection fishing mortality to $F_{35 \%}$ will be provided.

## Biomass

Table 2.27 shows the time series of age $0+$, age $3+$, and female spawning biomass since 1977 as estimated last year and this year (projections through 2017 are also shown for this year's assessment). The estimated spawning biomass time series are accompanied by their respective standard deviations.

The estimated time series of age $0+$ and female spawning biomass are shown, together with the observed time series of trawl survey biomass, in Figure 2.16. Confidence intervals are shown for estimates of female spawning biomass and for the trawl survey biomass estimates.

## Recruitment and Numbers at Age

Table 2.28 shows the time series of age 0 recruitment (1000s of fish) for the years since 1977 as estimated last year and this year. Both estimated time series are accompanied by their respective standard deviations.

For the time series as a whole, the largest year class appears to have been the 2008 cohort, and the year classes since 2008 include top three year classes of all time (2008, 2011, and 2013). The set of year classes comprising the top ten is the same this year as last year, except that the 1992 cohort has been added to the list and the 1979 cohort has been removed.

Recruitment estimates for the entire time series (1977-2015) are shown in Figure 2.17, along with their respective $95 \%$ confidence intervals.

The coefficient of autocorrelation for the recruitment time series is -0.11 .

To date, it has not been possible to estimate a reliable stock-recruitment relationship for this stock. A possible relationship between recruitment and an environmental index is discussed in the "Ecosystem Considerations" section, under "Ecosystem Effects on the Stock."

The estimated time series of numbers at age is shown in Table 2.29.

## Fishing Mortality

Table 2.30 shows "effective" fishing mortality by age and year for ages 1-19 and years since 1977.
Figure 2.18 plots the estimated/projected trajectory of relative fishing mortality $\left(F / F_{35 \%}\right)$ and relative female spawning biomass ( $B / B_{35 \%}$ ) from 1977 through 2018 based on full-selection fishing mortality, overlaid with the current harvest control rules. Projected values for 2016 and 2017 are from Scenario 2 under "Harvest Recommendations," below. It should be noted that, except for the projection years, these trajectories based on SS output, which may not match the estimates obtained by the standard projection program exactly.

## Harvest Recommendations

The results presented in this section are based on Model 16.6. Because the structure of this model differs substantively from Model 11.5 (the model accepted for the last five years by the SSC), a set of parallel results for the items in this section, based on Model 11.5, is provided in Appendix 2.5.

## Amendment 56 Reference Points

Amendment 56 to the BSAI Groundfish Fishery Management Plan (FMP) defines the "overfishing level" (OFL), the fishing mortality rate used to set OFL ( $F_{O F L}$ ), the maximum permissible ABC, and the fishing mortality rate used to set the maximum permissible ABC . The fishing mortality rate used to set ABC ( $F_{A B C}$ ) may be less than this maximum permissible level, but not greater. Because reliable estimates of reference points related to maximum sustainable yield (MSY) are currently not available but reliable estimates of reference points related to spawning per recruit are available, Pacific cod in the EBS have generally been managed under Tier 3 of Amendment 56. Tier 3 uses the following reference points: $B_{40 \%}$, equal to $40 \%$ of the equilibrium spawning biomass that would be obtained in the absence of fishing; $F_{35 \%}$, equal to the fishing mortality rate that reduces the equilibrium level of spawning per recruit to $35 \%$ of the level that would be obtained in the absence of fishing; and $F_{40 \%}$, equal to the fishing mortality rate that reduces the equilibrium level of spawning per recruit to $40 \%$ of the level that would be obtained in the absence of fishing. The following formulae apply under Tier 3:

$$
\begin{aligned}
& \text { 3a) Stock status: } B / B_{40 \%}>1 \\
& F_{O F L}=F_{35 \%} \\
& F_{A B C} \leq F_{40 \%} \\
& \text { 3b) Stock status: } 0.05<B / B_{40 \%} \leq 1 \\
& F_{O F L}=F_{35 \%} \times\left(B / B_{40 \%}-0.05\right) \times 1 / 0.95 \\
& F_{A B C} \leq F_{40 \%} \times\left(B / B_{40 \%}-0.05\right) \times 1 / 0.95 \\
& \text { 3c) Stock status: } B / B_{40 \%} \leq 0.05 \\
& F_{O F L}=0 \\
& F_{A B C}=0
\end{aligned}
$$

Model 16.6's estimates $F_{35 \%}$ and $F_{40 \%}$ are 0.38 and 0.31 , respectively.
Model 16.6's estimates of $B_{100 \%}, B_{40 \%}$, and $B_{35 \%}$ are $620,000 \mathrm{t}, 248,000 \mathrm{t}$, and $217,000 \mathrm{t}$, respectively.
Specification of OFL and Maximum Permissible ABC
Given the assumptions of Scenario 2 (below), female spawning biomass for 2017 and 2018 is estimated by Model 16.6 to be well above the $B_{40 \%}$ value of $248,000 t$, thereby placing Pacific cod in sub-tier "a" of Tier 3 for both 2017 and 2018. Given this, Model 16.6 estimates OFL, maximum permissible ABC, and the associated fishing mortality rates for 2017 and 2018 as follows:

| Year | Overfishing Level | Maximum Permissible ABC |
| ---: | ---: | ---: |
| 2017 | OFL $=284,000 \mathrm{t}$ | $\operatorname{maxABC}=239,000 \mathrm{t}$ |
| 2018 | OFL $=302,000 \mathrm{t}$ | $\max A B C=255,000 \mathrm{t}$ |
| 2017 | $F O F L=0.38$ | $\operatorname{maxFABC}=0.31$ |
| 2018 | $F O F L=0.38$ | $\operatorname{maxFABC}=0.31$ |

The age 0+ biomass projections for 2017 and 2018 from Model 16.6 (using SS rather than the standard projection model) are 1,260,000 t and $1,110,000 \mathrm{t}$.

For comparison, the age 3+ biomass projections for 2017 and 2018 from Model 16.6 (again using SS) are $1,230,000 \mathrm{t}$ and $1,060,000 \mathrm{t}$.

## Standard Harvest Scenarios, Projection Methodology, and Projection Results

A standard set of projections is required for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with an estimated vector of numbers at age for January 1, 2017. This requires an appropriate estimate of total catch for 2016. Because each year's stock assessment is finalized before complete (i.e., year-long) catch data are available for that year, it is necessary to extrapolate the available catch data through the end of the year. Year-end catch for 2016 was estimated to equal the ABC, at a value of $255,000 \mathrm{t}$, using the method described under "Catch Biomass" in the "Data" section.

In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Except for the first projection year under Scenario 2 (see paragraph below), total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

For predicting future catches under Scenario 2, the 2014 assessment (Thompson 2014) described development of the following estimator for future total catch as a function of future ABC: For $A B C \geq 148,000 t$, catch $=59,200 t+0.6 \times A B C$; for $A B C<148,000 t$, catch $=A B C$. This estimator was used again in the present assessment, giving a catch of 202,655 t for 2017.

Five of the seven standard scenarios are sometimes used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TACs for 2017 and 2018, are as follow ("max $F_{A B C}$ " refers to the maximum permissible value of $F_{A B C}$ under Amendment 56):

Scenario 1: In all future years, $F$ is set equal to $\max F_{A B C}$. (Rationale: Historically, TAC has been constrained by ABC, so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, $F$ is set equal to a constant fraction ("author's $F$ ") of $\max F_{A B C}$, where this fraction is equal to the ratio of the $F_{A B C}$ value for 2017 recommended in the assessment to the $\max F_{A B C}$ for 2017, and where catches for 2017 and 2018 are estimated at their most likely values given the 2017 and 2018 maximum permissible ABCs under this scenario. (Rationale: When $F_{A B C}$ is set at a value below max $F_{A B C}$, it is often set at the value recommended in the stock assessment; also, catch tends not to equal ABC exactly.)

Scenario 3: In all future years, $F$ is set equal to the 2011-2015 average F. (Rationale: For some stocks, TAC can be well below ABC, and recent average $F$ may provide a better indicator of $F_{\text {TAC }}$ than $F_{A B C}$.)

Scenario 4: In all future years, the upper bound on $F_{A B C}$ is set at $F_{60 \%}$. (Rationale: This scenario provides a likely lower bound on $F_{A B C}$ that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 5: In all future years, $F$ is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether a stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follow (for Tier 3 stocks, the MSY level is defined as $B_{35 \%}$ ):

Scenario 6: In all future years, $F$ is set equal to $F_{O F L}$. (Rationale: This scenario determines whether a stock is overfished. If the stock is 1) above its MSY level in 2016 or 2 ) above $1 / 2$ of its MSY level in 2016 and expected to be above its MSY level in 2026 under this scenario, then the stock is not overfished.)

Scenario 7: In 2017 and 2018, $F$ is set equal to $\max F_{A B C}$, and in all subsequent years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is 1) above its MSY level in 2018 or 2) above $1 / 2$ of its MSY level in 2018 and expected to be above its MSY level in 2028 under this scenario, then the stock is not approaching an overfished condition.)

Projections corresponding to the standard scenarios are shown for Model 16.6 in Tables 2.31-2.37.
In addition to the seven standard harvest scenarios, Amendments $48 / 48$ to the BSAI and GOA Groundfish Fishery Management Plans require projections of the likely OFL two years into the future. While Scenario 6 gives the best estimate of OFL for 2017, it does not provide the best estimate of OFL for 2017, because the mean 2017 catch under Scenario 6 is predicated on the 2017 catch being equal to the 2017 OFL, whereas the actual 2017 catch will likely be less than the 2017 OFL. Tables 2.24 and 2.32 contain the appropriate one- and two-year ahead projections for both ABC and OFL under Model 16.6.

## ABC Recommendation

Since 2005, the SSC has set ABC at the maximum permissible level every year with the exceptions of the 2007, 2014, and 2015 assessment cycles, when, in each case, the SSC held the ABCs for the next two years constant at the then-current level. Specifications for 2006-2011 were set under Tier 3b, and specifications for 2012-2017 were set under Tier 3a.

In the present assessment, spawning biomass is estimated to be well above $B_{40 \%}$, and is projected to remain so for at least the next couple of years. This high biomass is largely the result of the 2006, 2008, 2011, and 2013 year classes, whose strengths have now been confirmed by multiple surveys.

The two concerns that resulted in the decisions during the 2014 and 2015 assessment cycles to keep ABC constant at the 2014 level no longer remain. The first of these was doubt over reliability the sharply declining right-hand limb of the trawl survey selectivity function as estimated by Model 11.5 and the low value of catchability ( 0.77 ) assumed in that model. However, Model 16.6 assumes asymptotic trawl survey selectivity, and catchability is freely estimated at a value of 0.88 . The second was the large and positive retrospective bias in Model 11.5's estimates of current-year spawning biomass. However, the retrospective bias exhibited by Model 16.6, while still positive, is much lower than that of Model 11.5 ( 0.147 versus 0.475 ), and is within the range cited by Hurtado-Ferro et al. (2015) as not being cause for concern.

Because the previous concerns no longer remain, it seems appropriate to set ABC for 2017 and 2018 at the maximum permissible levels (Scenario 2) of 239,000 t and 255,000 t .

## Area Allocation of Harvests

No recommendations are made regarding area allocation of harvests.

## Status Determination

Under the MSFCMA, the Secretary of Commerce is required to report on the status of each U.S. fishery with respect to overfishing. This report involves the answers to three questions: 1) Is the stock being subjected to overfishing? 2) Is the stock currently overfished? 3) Is the stock approaching an overfished condition?

Is the stock being subjected to overfishing? The official catch estimate for the most recent complete year (2015) is $232,832 \mathrm{t}$. This is less than the 2015 OFL of $346,000 \mathrm{t}$. Therefore, the EBS Pacific cod stock is not being subjected to overfishing.

Harvest Scenarios \#6 and \#7 are intended to permit determination of the status of a stock with respect to its minimum stock size threshold (MSST). Any stock that is below its MSST is defined to be overfished. Any stock that is expected to fall below its MSST in the next two years is defined to be approaching an overfished condition. Harvest Scenarios \#6 and \#7 are used in these determinations as follows:

Is the stock currently overfished? This depends on the stock's estimated spawning biomass in 2016:
a. If spawning biomass for 2016 is estimated to be below $1 / 2 B_{35 \%}$, the stock is below its MSST.
b. If spawning biomass for 2016 is estimated to be above $B_{35 \%}$, the stock is above its MSST.
c. If spawning biomass for 2016 is estimated to be above $1 / 2 B_{35 \%}$ but below $B_{35 \%}$, the stock's status relative to MSST is determined by referring to harvest Scenario \#6 (Table 2.36). If the mean spawning biomass for 2026 is below $B_{35 \%}$, the stock is below its MSST.
Otherwise, the stock is above its MSST.
Is the stock approaching an overfished condition? This is determined by referring to harvest Scenario \#7 (Table 2.37):
a. If the mean spawning biomass for 2018 is below $1 / 2 B_{35 \%}$, the stock is approaching an overfished condition.
b. If the mean spawning biomass for 2018 is above $B_{35 \%}$, the stock is not approaching an overfished condition.
c. If the mean spawning biomass for 2018 is above $1 / 2 B_{35 \%}$ but below $B_{35 \%}$, the determination depends on the mean spawning biomass for 2028. If the mean spawning biomass for 2028 is below $B_{35 \%}$, the stock is approaching an overfished condition. Otherwise, the stock is not approaching an overfished condition.

Based on the above criteria and Tables 2.36 and 2.37, the stock is not overfished and is not approaching an overfished condition.

## ECOSYSTEM CONSIDERATIONS

## Ecosystem Effects on the Stock

A primary ecosystem phenomenon affecting the Pacific cod stock seems to be the occurrence of periodic "regime shifts," in which central tendencies of key variables in the physical environment change on a scale spanning several years to a few decades (Zador, 2011). One well-documented example of such a regime shift occurred in 1977, and shifts occurring in 1989 and 1999 have also been suggested (e.g., Hare and Mantua 2000). In the present assessment, an attempt was made to estimate the change in mean recruitment of EBS Pacific cod associated with the 1977 regime shift. According to the assessment model, pre-1977 mean recruitment was only about $31 \%$ of post-1976 mean recruitment. Establishing a link between environment and recruitment within a particular regime is more difficult. In the 2004 assessment (Thompson and Dorn 2004), for example, the correlations between age 1 recruits spawned since 1977 and monthly values of the Pacific Decadal Oscillation (Mantua et al. 1997) were computed and found to be very weak.

In the 2012 assessment, annual log-scale recruitment devs estimated by the assessment model were regressed against each of several environmental indices summarized by Zador (2011). The highest univariate correlation was obtained for the spring-summer North Pacific Index (NPI), which was developed by Trenberth and Hurrell (1994). The NPI is the area-weighted sea level pressure over the region $30^{\circ} \mathrm{N}-65^{\circ} \mathrm{N}, 160^{\circ} \mathrm{E}-140^{\circ} \mathrm{W}$. Further investigations were conducted with monthly NPI data from the Climate Analysis Section of the National Center for Atmospheric Research. The best univariate model obtained in the 2012 analysis was a linear regression of recruitment devs from 1977-2011 against the October-December average NPI (from the same year). Vestfals et al. (2014) have also noted a positive correlation between Pacific cod recruitment and the NPI, although not the October-December average NPI in particular.

In each assessment since 2012, the regression analysis has been updated. This year's regression resulted in a correlation of $0.55\left(R^{2}=0.30\right)$. The time series, regression line, and $95 \%$ confidence interval from this year’s regression are shown in the upper panel of Figure 2.19. According to this regression, the probability of the 2015 year class being higher than the median for the time series is $51 \%$. However, the datum for 2015 (magenta diamond in the upper panel) falls quite a bit below the predicted value from the regression. This marks the first time in the last 11 years (cohorts) that the sign of the dev estimated by the assessment model differs from the sign predicted by the regression (although the dev predicted by the regression is extremely close to zero: 0.014).

In each assessment since 2013, the main regression analysis has been accompanied by a cross-validation analysis involving creation of 100,000 "training" data sets, each one obtained by randomly sub-sampling $50 \%$ of the data without replacement. A regression was performed on each of the training sets, and then the performance of each regression was computed against the corresponding "test" (i.e., non-training) data set. When the NPI was not included as an explanatory variable (i.e., only the intercept of the regression was estimated), the RMSE (computed across all 100,000 test data sets) was 0.68 , but when the NPI was included as an explanatory variable, the RMSE was reduced to 0.59 . The distribution of slope parameter estimates from the cross-validation is shown in the middle panel of Figure 2.19. Note that the entire distribution is well above zero, indicating that the observed correlation is very unlikely to be entirely spurious. Two years, 1990 and 2002 (yellow and green diamonds in the upper panel), turned out to be far more influential than any other year in determining the magnitude of the estimated slope, and both of these influences were negative (lower panel of Figure 2.19). In other words, the positive slope is not due to the influence of outliers; if anything, the outliers are making the relationship appear less strong than would be the case without them.

The prey and predators of Pacific cod have been described or reviewed by Albers and Anderson (1985), Livingston (1989, 1991), Lang et al. (2003), Westrheim (1996), and Yang (2004). The composition of Pacific cod prey varies to some extent by time and area. In terms of percent occurrence, some of the most important items in the diet of Pacific cod in the BSAI and GOA have been polychaetes, amphipods, and crangonid shrimp. In terms of numbers of individual organisms consumed, some of the most important dietary items have been euphausids, miscellaneous fishes, and amphipods. In terms of weight of organisms consumed, some of the most important dietary items have been walleye pollock, fishery offal, yellowfin sole, and crustaceans. Small Pacific cod feed mostly on invertebrates, while large Pacific cod are mainly piscivorous. Predators of Pacific cod include Pacific cod, halibut, salmon shark, northern fur seals, Steller sea lions, harbor porpoises, various whale species, and tufted puffin. Major trends in the most important prey or predator species could be expected to affect the dynamics of Pacific cod to some extent.

## Fishery Effects on the Ecosystem

Potentially, fisheries for Pacific cod can have effects on other species in the ecosystem through a variety of mechanisms, for example by relieving predation pressure on shared prey species (i.e., species which serve as prey for both Pacific cod and other species), by reducing prey availability for predators of Pacific cod, by altering habitat, by imposing bycatch mortality, or by "ghost fishing" caused by lost fishing gear.

## Incidental Catch Taken in the Pacific Cod Fisheries

Incidental catches taken in the Pacific cod fisheries, expressed as proportions of total incidental EBS catches (i.e., across all targets) for the respective species, are summarized in Tables 2.38-2.41. Catches for 2016 in each of these tables are incomplete. Table 2.38 shows incidental catch of FMP species taken from 1991-2016 by each of the three main gear types. Table 2.39 shows incidental catch of certain species of squid and members of the former "other species" complex taken from 2003-2016, aggregated across gear types. Table 2.40 shows incidental catch of prohibited species taken from 1991-2016, aggregated across gear types. Note that all entries for 2003 are marked " $n / \mathrm{a}$ " in Table 2.40, due to an error in the database that was discovered too late to be corrected in time for this assessment. Table 2.41 shows incidental catch of non-target species groups taken from 2004-2016, aggregated across gear types (Table 2.41 starts in 2004 rather than 2003 for the same reason that the entries for 2003 are marked " $\mathrm{n} / \mathrm{a}$ " in Table 2.40).

## Steller Sea Lions

Sinclair and Zeppelin (2002) showed that Pacific cod was one of the four most important prey items of Steller sea lions in terms of frequency of occurrence averaged over years, seasons, and sites, and was especially important in winter. Pitcher (1981) and Calkins (1998) also showed Pacific cod to be an important winter prey item in the GOA and BSAI, respectively. Furthermore, the size ranges of Pacific cod harvested by the fisheries and consumed by Steller sea lions overlap, and the fishery has operated to some extent in the same areas used by Steller sea lion as foraging grounds (Livingston (ed.), 2002).

One of the main research emphases of the AFSC Fisheries Interaction Team (now disbanded) was to determine the effectiveness of management measures designed to mitigate the impacts of the Pacific cod fisheries (among others) on Steller sea lions. A study conducted in 2002-2005 using pot fishing gear demonstrated that the local concentration of cod in the Unimak Pass area is very dynamic, so that fishery removals did not create a measurable decline in fish abundance (Conners and Munro 2008). A preliminary tagging study in 2003-2004 showed some cod remaining in the vicinity of the release area in the southeast Bering Sea for several months, while other fish moved distances of 150 km or more northnorthwest along the shelf, some within a matter of two weeks (Rand et al. 2015).

## Seabirds

The following is a summary of information provided by Livingston (ed., 2002): In both the BSAI and GOA, the northern fulmar (Fulmarus glacialis) comprises the majority of seabird bycatch, which occurs primarily in the longline fisheries, including the hook and line fishery for Pacific cod. Shearwater (Puffinus spp.) distribution overlaps with the Pacific cod longline fishery in the Bering Sea, and with trawl fisheries in general in both the Bering Sea and GOA. Black-footed albatross (Phoebastria nigripes) is taken in much greater numbers in the GOA longline fisheries than the Bering Sea longline fisheries, but is not taken in the trawl fisheries. The distribution of Laysan albatross (Phoebastria immutabilis) appears to overlap with the longline fisheries in the central and western Aleutians. The distribution of short-tailed albatross (Phoebastria albatrus) also overlaps with the Pacific cod longline fishery along the Aleutian chain, although the majority of the bycatch has taken place along the northern portion of the Bering Sea shelf edge (in contrast, only two takes have been recorded in the GOA). Some success has been obtained in devising measures to mitigate fishery-seabird interactions. For example, on vessels larger than 60 ft . LOA, paired streamer lines of specified performance and material standards have been found to reduce seabird incidental take significantly.

## Fishery Usage of Habitat

The following is a summary of information provided by Livingston (ed., 2002): The longline and trawl fisheries for Pacific cod each comprise an important component of the combined fisheries associated with the respective gear type in each of the three major management regions (BS, AI, and GOA). Looking at each gear type in each region as a whole (i.e., aggregating across all target species) during the period 1998-2001, the total number of observed hauls/sets was as follows:

| Gear | BS | AI | GOA |
| :--- | :--- | :--- | :--- |
| Trawl | 240,347 | 43,585 | 68,436 |
| Longline | 65,286 | 13,462 | 7,139 |

In the BS, both longline and trawl effort was concentrated north of False Pass (Unimak Island) and along the shelf edge represented by the boundary of areas 513 , 517 (in addition, longline effort was concentrated along the shelf edge represented by the boundary of areas 521-533). In the AI, both longline and trawl effort were dispersed over a wide area along the shelf edge. The catcher vessel longline fishery in the AI occurred primarily over mud bottoms. Longline catcher-processors in the AI tended to fish more over rocky bottoms. In the GOA, fishing effort was also dispersed over a wide area along the shelf, though pockets of trawl effort were located near Chirikof, Cape Barnabus, Cape Chiniak and Marmot Flats. The GOA longline fishery for Pacific cod generally took place over gravel, cobble, mud, sand, and rocky bottoms, in depths of 25 fathoms to 140 fathoms.

Impacts of the Pacific cod fisheries on essential fish habitat were further analyzed in an environmental impact statement by NMFS (2005), followed by a 5 -year review in 2010 (NMFS 2010). A second 5 -year review is currently in progress.

## DATA GAPS AND RESEARCH PRIORITIES

Significant improvements in the quality of this assessment could be made if future research were directed toward closing certain data gaps. At this point, the most critical needs pertain to trawl survey catchability and selectivity, specifically: 1) to understand the factors determining these characteristics, 2) to understand whether/how these characteristics change over time, and 3) to obtain accurate estimates of these characteristics. Additional surveys of the NBS may prove helpful in this regard. Ageing also continues to be an issue, as the assessment models consistently estimate a positive ageing bias. Longer-
term research needs include improved understanding of: 1) the ecology of Pacific cod in the EBS, including spatial dynamics, trophic and other interspecific relationships, and the relationship between climate and recruitment; 2) ecology of species taken as bycatch in the Pacific cod fisheries, including estimation of biomass, carrying capacity, and resilience; and 3) ecology of species that interact with Pacific cod, including estimation of interaction strengths, biomass, carrying capacity, and resilience.

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Ongoing contributions: Rick Methot developed the SS software used to conduct the Pacific cod assessments over the last many years. NMFS Alaska Region provided the official catch time series. Numerous AFSC personnel and countless fishery observers collected nearly all of the raw data that were used in this assessment.

Reviewers: Carey McGilliard and the BSAI Groundfish Plan Team provided reviews of this assessment.

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TABLES

Table 2.1a-Summary of 1964-1980 catches (t) of Pacific cod in the EBS by fleet sector. "For." = foreign, "JV" = joint venture processing, "Dom." = domestic annual processing. Catches by gear are not available for these years. Catches may not always include discards.

| Year | For. | JV | Dom. | Total |
| :---: | :---: | ---: | ---: | ---: |
| 1964 | 13,408 | 0 | 0 | 13,408 |
| 1965 | 14,719 | 0 | 0 | 14,719 |
| 1966 | 18,200 | 0 | 0 | 18,200 |
| 1967 | 32,064 | 0 | 0 | 32,064 |
| 1968 | 57,902 | 0 | 0 | 57,902 |
| 1969 | 50,351 | 0 | 0 | 50,351 |
| 1970 | 70,094 | 0 | 0 | 70,094 |
| 1971 | 43,054 | 0 | 0 | 43,054 |
| 1972 | 42,905 | 0 | 0 | 42,905 |
| 1973 | 53,386 | 0 | 0 | 53,386 |
| 1974 | 62,462 | 0 | 0 | 62,462 |
| 1975 | 51,551 | 0 | 0 | 51,551 |
| 1976 | 50,481 | 0 | 0 | 50,481 |
| 1977 | 33,335 | 0 | 0 | 33,335 |
| 1978 | 42,512 | 0 | 31 | 42,543 |
| 1979 | 32,981 | 0 | 780 | 33,761 |
| 1980 | 35,058 | 8,370 | 2,433 | 45,861 |

Table 2.1b—Summary of 1981-1990 catches ( t ) of Pacific cod in the EBS by fleet sector, and gear type. All catches include discards. "LLine" = longline, "Subt." = sector subtotal. Breakdown of domestic annual processing by gear is not available prior to 1988.

|  | Foreign |  |  | Joint Venture |  |  | Domestic Annual Processing |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Trawl | LLine | Subt. | Trawl | Subt. | Trawl | LLine | Pot | Subt. | Total |  |
| 1981 | 30,347 | 5,851 | 36,198 | 7,410 | 7,410 | n/a | n/a | n/a | 12,899 | 56,507 |  |
| 1982 | 23,037 | 3,142 | 26,179 | 9,312 | 9,312 | n/a | n/a | n/a | 25,613 | 61,104 |  |
| 1983 | 32,790 | 6,445 | 39,235 | 9,662 | 9,662 | n/a | n/a | n/a | 45,904 | 94,801 |  |
| 1984 | 30,592 | 26,642 | 57,234 | 24,382 | 24,382 | n/a | n/a | n/a | 43,487 | 125,103 |  |
| 1985 | 19,596 | 36,742 | 56,338 | 35,634 | 35,634 | n/a | n/a | n/a | 51,475 | 143,447 |  |
| 1986 | 13,292 | 26,563 | 39,855 | 57,827 | 57,827 | n/a | n/a | n/a | 37,923 | 135,605 |  |
| 1987 | 7,718 | 47,028 | 54,746 | 47,722 | 47,722 | n/a | n/a | n/a | 47,435 | 149,903 |  |
| 1988 | 0 | 0 | 0 | 106,592 | 106,592 | 93,706 | 2,474 | 299 | 96,479 | 203,071 |  |
| 1989 | 0 | 0 | 0 | 44,612 | 44,612 | 119,631 | 13,935 | 145 | 133,711 | 178,323 |  |
| 1990 | 0 | 0 | 0 | 8,078 | 8,078 | 115,493 | 47,114 | 1,382 | 163,989 | 172,067 |  |

Table 2.1c—Summary of 1991-2016 catches ( t ) of Pacific cod in the EBS by gear type. The small catches taken by "other" gear types have been merged proportionally with the catches of the gear types shown. Pot catches for 2014-2016 include the State-managed fishery. Catches for 2016 are through September 25.

| Year | Trawl | Longline | Pot | Total |
| ---: | ---: | ---: | ---: | ---: |
| 1991 | 129,393 | 77,505 | 3,343 | 210,241 |
| 1992 | 77,276 | 79,420 | 7,514 | 164,210 |
| 1993 | 81,792 | 49,296 | 2,098 | 133,186 |
| 1994 | 85,294 | 78,898 | 8,071 | 172,263 |
| 1995 | 111,250 | 97,923 | 19,326 | 228,498 |
| 1996 | 92,029 | 88,996 | 28,042 | 209,067 |
| 1997 | 93,995 | 117,097 | 21,509 | 232,601 |
| 1998 | 60,855 | 84,426 | 13,249 | 158,529 |
| 1999 | 51,939 | 81,520 | 12,408 | 145,867 |
| 2000 | 53,841 | 81,678 | 15,856 | 151,376 |
| 2001 | 35,670 | 90,394 | 16,478 | 142,542 |
| 2002 | 51,118 | 100,371 | 15,067 | 166,555 |
| 2003 | 46,717 | 108,764 | 19,957 | 175,438 |
| 2004 | 57,866 | 108,618 | 17,264 | 183,748 |
| 2005 | 52,638 | 113,190 | 17,112 | 182,940 |
| 2006 | 53,236 | 96,613 | 18,969 | 168,818 |
| 2007 | 45,700 | 77,181 | 17,248 | 140,129 |
| 2008 | 33,497 | 88,936 | 17,368 | 139,802 |
| 2009 | 36,959 | 96,606 | 13,609 | 147,174 |
| 2010 | 41,298 | 81,848 | 19,723 | 142,869 |
| 2011 | 64,086 | 117,066 | 28,063 | 209,215 |
| 2012 | 75,534 | 128,513 | 28,737 | 232,784 |
| 2013 | 81,615 | 124,814 | 30,261 | 236,691 |
| 2014 | 72,260 | 127,311 | 39,193 | 238,763 |
| 2015 | 66,677 | 128,218 | 37,938 | 232,832 |
| 2016 | 69,786 | 98,691 | 39,314 | 207,791 |

Table 2.2—Discards ( t ) and discard rates (\%) of Pacific cod in the Pacific cod fishery, by area, gear, and year for the period 1991-2016 (2016 data are current through September 25). The small amounts of discards taken by other gear types have been merged proportionally into the gear types shown. Note that Amendment 49, which mandated increased retention and utilization, was implemented in 1998.

|  | Discard amount $(\mathrm{t})$ |  |  |  | Discard rate (\%) |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Trawl | Longline | Pot | Total | Trawl | Longline | Pot | All |
| 1991 | 1,278 | 1,493 | 4 | 2,774 | 4.11 | 2.62 | 0.26 | 3.10 |
| 1992 | 3,314 | 1,768 | 59 | 5,141 | 8.68 | 2.23 | 0.78 | 4.12 |
| 1993 | 5,449 | 2,234 | 25 | 7,708 | 12.89 | 4.54 | 1.21 | 8.24 |
| 1994 | 4,599 | 2,917 | 161 | 7,677 | 9.98 | 3.71 | 2.01 | 5.79 |
| 1995 | 7,987 | 3,669 | 222 | 11,877 | 12.24 | 3.77 | 1.15 | 6.54 |
| 1996 | 2,971 | 2,833 | 391 | 6,194 | 5.12 | 3.19 | 1.39 | 3.54 |
| 1997 | 3,327 | 3,183 | 79 | 6,590 | 5.42 | 2.72 | 0.37 | 3.30 |
| 1998 | 102 | 2,456 | 52 | 2,610 | 0.27 | 2.92 | 0.39 | 1.94 |
| 1999 | 353 | 1,285 | 52 | 1,691 | 0.95 | 1.58 | 0.42 | 1.29 |
| 2000 | 207 | 2,267 | 71 | 2,546 | 0.56 | 2.78 | 0.45 | 1.90 |
| 2001 | 142 | 1,531 | 52 | 1,726 | 0.76 | 1.70 | 0.32 | 1.38 |
| 2002 | 557 | 2,066 | 91 | 2,715 | 1.73 | 2.06 | 0.61 | 1.84 |
| 2003 | 240 | 1,771 | 159 | 2,170 | 0.79 | 1.63 | 0.80 | 1.36 |
| 2004 | 158 | 1,814 | 48 | 2,019 | 0.41 | 1.67 | 0.28 | 1.23 |
| 2005 | 86 | 2,599 | 61 | 2,747 | 0.26 | 2.30 | 0.36 | 1.68 |
| 2006 | 193 | 1,528 | 63 | 1,784 | 0.54 | 1.58 | 0.33 | 1.18 |
| 2007 | 238 | 1,373 | 45 | 1,656 | 0.74 | 1.78 | 0.26 | 1.31 |
| 2008 | 13 | 1,280 | 156 | 1,449 | 0.09 | 1.44 | 0.90 | 1.20 |
| 2009 | 126 | 1,503 | 16 | 1,645 | 1.02 | 1.56 | 0.12 | 1.34 |
| 2010 | 154 | 1,402 | 19 | 1,575 | 1.08 | 1.72 | 0.10 | 1.36 |
| 2011 | 121 | 1,860 | 32 | 2,013 | 0.42 | 1.59 | 0.11 | 1.16 |
| 2012 | 136 | 1,759 | 40 | 1,934 | 0.38 | 1.37 | 0.14 | 1.01 |
| 2013 | 220 | 3,066 | 90 | 3,376 | 0.58 | 2.46 | 0.30 | 1.75 |
| 2014 | 192 | 2,893 | 155 | 3,240 | 0.50 | 2.28 | 0.40 | 1.58 |
| 2015 | 141 | 2,374 | 104 | 2,618 | 0.43 | 1.85 | 0.27 | 1.32 |
| 2016 | 117 | 2,029 | 60 | 2,206 | 0.29 | 2.06 | 0.15 | 1.24 |

Table 2.3-History of BSAI (1977-2013) and EBS (2014-2016) Pacific cod catch, TAC, ABC, and OFL (t). Catch for 2016 is through September 25. Note that specifications through 2013 were for the combined BSAI region, so BSAI catch is shown rather than the EBS catches from Table 2.1 for the period 1977-2013. Source for historical specifications: NPFMC staff.

| Year | Catch | TAC | ABC | OFL |
| ---: | ---: | ---: | ---: | ---: |
| 1977 | 36,597 | 58,000 | - | - |
| 1978 | 4,838 | 70,500 | - | - |
| 1979 | 39,354 | 70,500 | - | - |
| 1980 | 51,649 | 70,700 | 148,000 | - |
| 1981 | 63,941 | 78,700 | 160,000 | - |
| 1982 | 69,501 | 78,700 | 168,000 | - |
| 1983 | 103,231 | 120,000 | 298,200 | - |
| 1984 | 133,084 | 210,000 | 291,300 | - |
| 1985 | 150,384 | 220,000 | 347,400 | - |
| 1986 | 142,511 | 229,000 | 249,300 | - |
| 1987 | 163,110 | 280,000 | 400,000 | - |
| 1988 | 208,236 | 200,000 | 385,300 | - |
| 1989 | 182,865 | 230,681 | 370,600 | - |
| 1990 | 179,608 | 227,000 | 417,000 | - |
| 1991 | 210,241 | 229,000 | 229,000 | - |
| 1992 | 164,210 | 182,000 | 182,000 | 188,000 |
| 1993 | 133,186 | 164,500 | 164,500 | 192,000 |
| 1994 | 172,263 | 191,000 | 191,000 | 228,000 |
| 1995 | 228,498 | 250,000 | 328,000 | 390,000 |
| 1996 | 209,067 | 270,000 | 305,000 | 420,000 |
| 1997 | 232,601 | 270,000 | 306,000 | 418,000 |
| 1998 | 158,529 | 210,000 | 210,000 | 336,000 |
| 1999 | 145,867 | 177,000 | 177,000 | 264,000 |
| 2000 | 151,376 | 193,000 | 193,000 | 240,000 |
| 2001 | 142,542 | 188,000 | 188,000 | 248,000 |
| 2002 | 166,555 | 200,000 | 223,000 | 294,000 |
| 2003 | 175,438 | 207,500 | 223,000 | 324,000 |
| 2004 | 183,748 | 215,500 | 223,000 | 350,000 |
| 2005 | 182,940 | 206,000 | 206,000 | 265,000 |
| 2006 | 168,818 | 194,000 | 194,000 | 230,000 |
| 2007 | 140,129 | 170,720 | 176,000 | 207,000 |
| 2008 | 139,802 | 170,720 | 176,000 | 207,000 |
| 2009 | 147,174 | 176,540 | 182,000 | 212,000 |
| 2010 | 142,869 | 168,780 | 174,000 | 205,000 |
| 2011 | 209,215 | 227,950 | 235,000 | 272,000 |
| 2012 | 232,784 | 261,000 | 314,000 | 369,000 |
| 2013 | 236,691 | 260,000 | 307,000 | 359,000 |
| 2014 | 238,763 | 246,897 | 255,000 | 299,000 |
| 2015 | 232,832 | 240,000 | 255,000 | 346,000 |
| 2016 | 207,791 | 238,680 | 255,000 | 390,000 |
|  |  |  |  |  |

Table 2.4. Amendments to the BSAI Fishery Management Plan (FMP) that reference Pacific cod explicitly (excerpted from Appendix A of the FMP).

Amendment 2, implemented January 12, 1982:
For Pacific cod, decreased maximum sustainable yield to $55,000 t$ from $58,700 t$, increased equilibrium yield to $160,000 t$ from 58,700 t, increased acceptable biological catch to $160,000 \mathrm{t}$ from $58,700 \mathrm{t}$, increased optimum yield to $78,700 \mathrm{t}$ from $58,700 \mathrm{t}$, increased reserves to $3,935 \mathrm{t}$ from $2,935 \mathrm{t}$, increased domestic annual processing (DAP) to $26,000 \mathrm{t}$ from 7,000 t , and increased DAH to $43,265 \mathrm{t}$ from 24,265 t .
Amendment 4, implemented May 9, 1983, supersedes Amendment 2:
For Pacific Cod, increased equilibrium yield and acceptable biological catch to $168,000 \mathrm{t}$ from $160,000 \mathrm{t}$, increased optimum yield to $120,000 \mathrm{t}$ from $78,700 \mathrm{t}$, increased reserves to $6,000 \mathrm{t}$ from $3,935 \mathrm{t}$, and increased TALFF to $70,735 \mathrm{t}$ from 31,500 t .
Amendment 10, implemented March 16, 1987: Established Bycatch Limitation Zones for domestic and foreign fisheries for yellowfin sole and other flatfish (including rock sole); an area closed to all trawling within Zone 1; red king crab, C. bairdi Tanner crab, and Pacific halibut PSC limits for DAH yellowfin sole and other flatfish fisheries; a C. bairdi PSC limit for foreign fisheries; and a red king crab PSC limit and scientific data collection requirement for U.S. vessels fishing for Pacific cod in Zone 1 waters shallower than 25 fathoms.
Amendment 24, implemented February 28, 1994, and effective through December 31, 1996:

1. Established the following gear allocations of BSAI Pacific cod TAC as follows: 2 percent to vessels using jig gear; 44.1 percent to vessels using hook-and-line or pot gear, and 53.9 percent to vessels using trawl gear.
2. Authorized the seasonal apportionment of the amount of Pacific cod allocated to gear groups. Criteria for seasonal apportionments and the seasons authorized to receive separate apportionments will be set forth in regulations.
Amendment 46, implemented January 1, 1997, superseded Amendment 24:
Replaced the three year Pacific cod allocation established with Amendment 24, with the following gear allocations in BSAI Pacific cod: 2 percent to vessels using jig gear; 51 percent to vessels using hook-and-line or pot gear; and 47 percent to vessels using trawl gear. The trawl apportionment will be divided 50 percent to catcher vessels and 50 percent to catcher processors. These allocations as well as the seasonal apportionment authority established in Amendment 24 will remain in effect until amended.
Amendment 49, implemented January 3, 1998:
Implemented an Increased Retention/Increased Utilization Program for pollock and Pacific cod beginning January 1, 1998 and rock sole and yellowfin sole beginning January 1, 2003.
Amendment 64, implemented September 1, 2000, revised Amendment 46: Allocated the Pacific cod Total Allowable Catch to the jig gear (2 percent), fixed gear (51 percent), and trawl gear (47 percent) sectors.
Amendment 67, implemented May 15, 2002, revised Amendment 39: Established participation and harvest requirements to qualify for a BSAI Pacific cod fishery endorsement for fixed gear vessels.
Amendment 77, implemented January 1, 2004, revised Amendment 64: Implemented a Pacific cod fixed gear allocation between hook and line catcher processors ( 80 percent), hook and line catcher vessels ( 0.3 percent), pot catcher processors ( 3.3 percent), pot catcher vessels ( 15 percent), and catcher vessels (pot or hook and line) less than 60 feet ( 1.4 percent).
Amendment 80, implemented on July 26, 2007, superseded Amendments 49 and 75:
3. Allocates non-pollock groundfish in the BSAI among trawl sectors
4. Creates a limited access privilege program to facilitate the formation of harvesting cooperative in the non-American Fisheries Act trawl catcher/processor sector.
Amendment 85, partially implemented on March 5, 2007, superseded Amendments 46 and 77:
Implemented a gear allocation among all non-CDQ fishery sectors participating in the directed fishery for Pacific cod. After deduction of the CDQ allocation, the Pacific cod TAC is apportioned to vessels using jig gear (1.4 percent); catcher processors using trawl gear listed in Section 208(e)(1)-(20) of the AFA (2.3 percent); catcher processors using trawl gear as defined in Section 219(a)(7) of the Consolidated Appropriations Act, 2005 (Public Law 108-447) (13.4 percent); catcher vessels using trawl gear ( 22.1 percent); catcher processors using hook-and-line gear ( 48.7 percent); catcher vessels $\geq 60$ ' LOA using hook-and-line gear ( 0.2 percent); catcher processors using pot gear ( 1.5 percent); catcher vessels $\geq 60$ ' LOA using pot gear ( 8.4 percent); and catcher vessels $<60^{\prime}$ LOA that use either hook-and-line gear or pot gear ( 2.0 percent).
Amendment 99, implemented on January 6, 2014 (effective February 6, 2014):
Allows holders of license limitation program (LLP) licenses endorsed to catch and process Pacific cod in the Bering Sea/Aleutian Islands hook-and-line fisheries to use their LLP license on larger newly built or existing vessels by:
5. Increasing the maximum vessel length limits of the LLP license, and
6. Waiving vessel length, weight, and horsepower limits of the American Fisheries Act.

Amendment 103, implemented November 14, 2014:
Revise the Pribilof Islands Habitat Conservation Zone to close to fishing for Pacific cod with pot gear (in addition to the closure to all trawling).

Table 2.5 (p. 1 of 4)— EBS catch (t) of Pacific cod by year, gear, and season for the years 1977-2016 as configured in Model 11.5. Because direct estimates of gear- and period-specific catches are not available for the years 1977-1980, the figures shown here are estimates derived by distributing each year's total catch according to the average proportion observed for each gear/period combination during the years 1981-1988. The small amounts of catch from "other" gear types have been merged into the gear types listed below proportionally.

| Year | Season | Trawl fishery |  |  | Longline fishery |  |  | Pot fishery |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Jan-Apr | May-Jul | Aug-Dec | Jan-Apr | May-Jul | Aug-Dec | Jan-Apr | May-Jul | Aug-Dec |
| 1977 | Jan-Feb | 5974 | 0 | 0 | 740 | 0 | 0 | 0 | 0 | 0 |
| 1977 | Mar-Apr | 5974 | 0 | 0 | 740 | 0 | 0 | 0 | 0 | 0 |
| 1977 | May-Jul | 0 | 7080 | 0 | 0 | 544 | 0 | 0 | 0 | 0 |
| 1977 | Aug-Oct | 0 | 0 | 5475 | 0 | 0 | 1733 | 0 | 0 | 0 |
| 1977 | Nov-Dec | 0 | 0 | 3429 | 0 | 0 | 1646 | 0 | 0 | 0 |
| 1978 | Jan-Feb | 7884 | 0 |  | 977 | 0 | 0 | 0 | 0 | 0 |
| 1978 | Mar-Apr | 7884 | 0 | 0 | 977 | 0 | 0 | 0 | 0 | 0 |
| 1978 | May-Jul | 0 | 9343 | 0 | 0 | 717 | 0 | 0 | 0 | 0 |
| 1978 | Aug-Oct | 0 | 0 | 7226 | 0 | 0 | 2286 | 0 | 0 | 0 |
| 1978 | Nov-Dec | 0 | 0 | 4526 | 0 | 0 | 2172 | 0 | 0 | 0 |
| 1979 | Jan-Feb | 6452 | 0 | 0 | 800 | 0 | 0 | 0 | 0 | 0 |
| 1979 | Mar-Apr | 6452 | 0 | 0 | 800 | 0 | 0 | 0 | 0 | 0 |
| 1979 | May-Jul | 0 | 7646 | 0 | 0 | 587 | 0 | 0 | 0 | 0 |
| 1979 | Aug-Oct | 0 | 0 | 5914 | 0 | 0 | 1871 | 0 | 0 | 0 |
| 1979 | Nov-Dec | 0 | 0 | 3704 | 0 | 0 | 1778 | 0 | 0 | 0 |
| 1980 | Jan-Feb | 7355 | 0 | 0 | 912 | 0 | 0 | 0 | 0 | 0 |
| 1980 | Mar-Apr | 7355 | 0 | 0 | 912 | 0 | 0 | 0 | 0 | 0 |
| 1980 | May-Jul | 0 | 8716 | 0 | 0 | 669 | 0 | 0 | 0 | 0 |
| 1980 | Aug-Oct | 0 | 0 | 6741 | 0 | 0 | 2133 | 0 | 0 | 0 |
| 1980 | Nov-Dec | 0 | 0 | 4222 | 0 | 0 | 2027 | 0 | 0 | 0 |
| 1981 | Jan-Feb | 6027 | 0 | 0 | 514 | 0 | 0 | 0 | 0 | 0 |
| 1981 | Mar-Apr | 6027 | 0 | 0 | 514 | 0 | 0 | 0 | 0 | 0 |
| 1981 | May-Jul | 0 | 12405 | 0 | 0 | 673 | 0 | 0 | 0 | 0 |
| 1981 | Aug-Oct | 0 | 0 | 15439 | 0 |  | 2179 | 0 | 0 | 0 |
| 1981 | Nov-Dec | 0 | 0 | 10743 | 0 | 0 | 1971 | 0 | 0 | 0 |
| 1982 | Jan-Feb | 8697 | 0 | 0 | 145 | 0 | 0 | 0 | 0 | 0 |
| 1982 | Mar-Apr | 8697 | 0 | 0 | 145 | 0 | 0 | 0 | 0 | 0 |
| 1982 | May-Jul | 0 | 16449 | 0 | 0 | 389 | 0 | 0 | 0 | 0 |
| 1982 | Aug-Oct | 0 | 0 | 14224 | 0 | 0 | 1312 | 0 | 0 | 0 |
| 1982 | Nov-Dec | 0 | 0 | 8174 | 0 |  | 1154 | 0 | 0 | 0 |
| 1983 | Jan-Feb | 16303 | 0 | 0 | 1176 | 0 | 0 | 0 | 0 | 0 |
| 1983 | Mar-Apr | 16303 | 0 | 0 | 1176 | 0 | 0 | 0 | 0 | 0 |
| 1983 | May-Jul | 0 | 24351 | 0 | 0 | 1087 | 0 | 0 | 0 | 0 |
| 1983 | Aug-Oct | 0 | 0 | 19453 | 0 | 0 | 1627 | 0 | 0 | 0 |
| 1983 | Nov-Dec | 0 | 0 | 11353 | 0 | 0 | 1378 | 0 | 0 | 0 |
| 1984 | Jan-Feb | 19295 | 0 | 0 | 2005 | 0 | 0 | 0 | 0 | 0 |
| 1984 | Mar-Apr | 19295 | 0 | 0 | 2005 | 0 | 0 | 0 | 0 | 0 |
| 1984 | May-Jul | 0 | 26290 | 0 | 0 | 2421 | 0 | 0 | 0 | 0 |
| 1984 | Aug-Oct | 0 | 0 | 20844 | 0 | 0 | 10463 | 0 | 0 | 0 |
| 1984 | Nov-Dec | 0 | 0 | 12523 | 0 | 0 | 9754 | 0 | 0 | 0 |
| 1985 | Jan-Feb | 22269 | 0 | 0 | 5481 | 0 | 0 | 0 | 0 | 0 |
| 1985 | Mar-Apr | 22269 | 0 | 0 | 5481 | 0 | 0 | 0 | 0 | 0 |
| 1985 | May-Jul | 0 | 30250 | 0 | 0 | 3881 | 0 | 0 | 0 | 0 |
| 1985 | Aug-Oct | 0 | 0 | 20713 | 0 | 0 | 11260 | 0 | 0 | 0 |
| 1985 | Nov-Dec | 0 | 0 | 11155 | 0 | 0 | 10690 | 0 | 0 | 0 |

Table 2.5 (p. 2 of 4)— EBS catch (t) of Pacific cod by year, gear, and season for the years 1977-2015 as configured in Model 11.5.

| Year | Season | Trawl fishery |  |  | Longline fishery |  |  | Pot fishery |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Jan-Apr | May-Jul | Aug-Dec | Jan-Apr | May-Jul | Aug-Dec | Jan-Apr | May-Jul | Aug-Dec |
| 1986 | Jan-Feb | 23914 | 0 | 0 | 3558 | 0 | 0 | 0 | 0 | 0 |
| 1986 | Mar-Apr | 23914 | 0 | 0 | 3558 | 0 | 0 | 0 | 0 | 0 |
| 1986 | May-Jul | 0 | 29689 | 0 | 0 | 2071 | 0 | 0 | 0 | 0 |
| 1986 | Aug-Oct | 0 | 0 | 20057 | 0 | 0 | 8785 | 0 | 0 | 0 |
| 1986 | Nov-Dec | 0 | 0 | 11191 | 0 | 0 | 8639 | 0 | 0 | 0 |
| 1987 | Jan-Feb | 25765 | 0 |  | 8379 | 0 | 0 | 0 | 0 | 0 |
| 1987 | Mar-Apr | 25765 | 0 | 0 | 8379 | 0 | 0 | 0 | 0 | 0 |
| 1987 | May-Jul | 0 | 23285 | 0 | 0 | 4671 | 0 | 0 | 0 | 0 |
| 1987 | Aug-Oct | 0 | 0 | 15932 | 0 | 0 | 13617 | 0 | 0 | 0 |
| 1987 | Nov-Dec | 0 | 0 | 10731 | 0 | 0 | 13376 | 0 | 0 | 0 |
| 1988 | Jan-Feb | 50988 | 0 | 0 | 214 | 0 |  | 0 | 0 | 0 |
| 1988 | Mar-Apr | 50988 | 0 | 0 | 214 | 0 | 0 | 0 | 0 | 0 |
| 1988 | May-Jul | 0 | 42602 | 0 | 0 | 571 | 0 | 0 | 0 | 0 |
| 1988 | Aug-Oct | 0 | 0 | 32137 | 0 | 0 | 1005 | 0 | 0 | 0 |
| 1988 | Nov-Dec | 0 | 0 | 23583 | 0 | 0 | 773 | 0 | 0 | 0 |
| 1989 | Jan-Feb | 50984 | 0 |  | 1524 | 0 | 0 | 13 | 0 | 0 |
| 1989 | Mar-Apr | 50984 | 0 | 0 | 1524 | 0 | 0 | 13 | 0 | 0 |
| 1989 | May-Jul | 0 | 36816 | 0 | 0 | 4074 | 0 | 0 | 49 | 0 |
| 1989 | Aug-Oct | 0 | 0 | 15561 | 0 | 0 | 4235 | 0 | 0 | 46 |
| 1989 | Nov-Dec | 0 | 0 | 9899 | 0 | 0 | 2579 | 0 | 0 | 25 |
| 1990 | Jan-Feb | 40658 | 0 | 0 | 5268 | 0 | 0 | 0 | 0 | 0 |
| 1990 | Mar-Apr | 40658 | 0 | 0 | 5268 | 0 | 0 | 0 | 0 | 0 |
| 1990 | May-Jul | 0 | 27930 | 0 | 0 | 13730 | 0 | 0 | 657 | 0 |
| 1990 | Aug-Oct | 0 | 0 | 9063 | 0 | 0 | 14197 | 0 | 0 | 526 |
| 1990 | Nov-Dec | 0 | 0 | 5262 | 0 | 0 | 8650 | 0 | 0 | 198 |
| 1991 | Jan-Feb | 34996 | 0 | 0 | 8229 | 0 | 0 | 20 | 0 | 0 |
| 1991 | Mar-Apr | 65276 | 0 | 0 | 12317 | 0 | 0 | 522 | 0 | 0 |
| 1991 | May-Jul | 0 | 16403 | 0 | 0 | 20115 | 0 | 0 | 410 | 0 |
| 1991 | Aug-Oct | 0 | 0 | 12271 | 0 | 0 | 21276 | 0 | 0 | 2306 |
| 1991 | Nov-Dec | 0 | 0 | 6420 | 0 | 0 | 9312 | 0 | 0 | 369 |
| 1992 | Jan-Feb | 23310 | 0 | 0 | 13660 | 0 | 0 | 13 | 0 | 0 |
| 1992 | Mar-Apr | 31836 | 0 | 0 | 22121 | 0 | 0 | 833 | 0 | 0 |
| 1992 | May-Jul | 0 | 11784 | 0 | 0 | 27051 | 0 | 0 | 5321 | 0 |
| 1992 | Aug-Oct | 0 | 0 | 8182 | 0 | 0 | 16319 | 0 | 0 | 1992 |
| 1992 | Nov-Dec | 0 | 0 | 1788 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | Jan-Feb | 27998 | 0 | 0 | 22396 | 0 | 0 | 24 | 0 | 0 |
| 1993 | Mar-Apr | 35294 | 0 | 0 | 21434 | 0 | 0 | 1597 | 0 | 0 |
| 1993 | May-Jul | 0 | 5552 | 0 | 0 | 4744 | 0 | 0 | 2093 | 0 |
| 1993 | Aug-Oct | 0 | 0 | 6944 | 0 | 0 | 3002 | 0 | 0 | 0 |
| 1993 | Nov-Dec | 0 | 0 | 1544 | 0 | 0 | 564 | 0 | 0 | 0 |
| 1994 | Jan-Feb | 13856 | 0 | 0 | 22458 | 0 | 0 | 0 | 0 | 0 |
| 1994 | Mar-Apr | 43634 | 0 | 0 | 29089 | 0 | 0 | 4159 | 0 | 0 |
| 1994 | May-Jul | 0 | 4453 | 0 | 0 | 6210 | 0 | 0 | 1792 | 0 |
| 1994 | Aug-Oct | 0 | 0 | 20070 | 0 | 0 | 20718 | 0 | 0 | 3133 |
| 1994 | Nov-Dec | 0 | 0 | 2691 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | Jan-Feb | 31939 | 0 | 0 | 29936 | 0 | 0 | 23 | 0 | 0 |
| 1995 | Mar-Apr | 58159 | 0 | 0 | 34516 | 0 | 0 | 7715 | 0 | 0 |
| 1995 | May-Jul | 0 | 1145 | 0 | 0 | 4161 | 0 | 0 | 7342 | 0 |
| 1995 | Aug-Oct | 0 | 0 | 19770 | 0 | 0 | 21305 | 0 | 0 | 2927 |
| 1995 | Nov-Dec | 0 | 0 | 119 | 0 | 0 | 8802 | 0 | 0 | 640 |

Table 2.5 (p. 3 of 4)— EBS catch (t) of Pacific cod by year, gear, and season for the years 1977-2015 as configured in Model 11.5.

| Year | Season | Trawl fishery |  |  | Longline fishery |  |  | Pot fishery |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Jan-Apr | May-Jul | Aug-Dec | Jan-Apr | May-Jul | Aug-Dec | Jan-Apr | May-Jul | Aug-Dec |
| 1996 | Jan-Feb | 21151 | 0 | 0 | 28835 | 0 | 0 | 25 | 0 | 0 |
| 1996 | Mar-Apr | 50436 | 0 | 0 | 29471 | 0 | 0 | 12571 | 0 | 0 |
| 1996 | May-Jul | 0 | 6797 | 0 | 0 | 4179 | 0 | 0 | 11600 | 0 |
| 1996 | Aug-Oct | 0 | 0 | 10543 | 0 | 0 | 23629 | 0 | 0 | 4347 |
| 1996 | Nov-Dec | 0 | 0 | 1475 | 0 | 0 | 3278 | 0 | 0 | 728 |
| 1997 | Jan-Feb | 25713 | 0 | 0 | 31971 | 0 | 0 | 30 | 0 | 0 |
| 1997 | Mar-Apr | 52321 | 0 | 0 | 30578 | 0 | 0 | 9639 | 0 | 0 |
| 1997 | May-Jul | 0 | 5174 | 0 | 0 | 8145 | 0 | 0 | 7352 | 0 |
| 1997 | Aug-Oct | 0 | 0 | 9321 | 0 | 0 | 21323 | 0 | 0 | 3780 |
| 1997 | Nov-Dec | 0 | 0 | 2366 | 0 | 0 | 24250 | 0 | 0 | 637 |
| 1998 | Jan-Feb | 15535 | 0 | 0 | 29256 | 0 | 0 | 1719 | 0 | 0 |
| 1998 | Mar-Apr | 27765 | 0 | 0 | 19060 | 0 | 0 | 5613 | 0 | 0 |
| 1998 | May-Jul | 0 | 4940 | 0 | 0 | 3709 | 0 | 0 | 5321 | 0 |
| 1998 | Aug-Oct | 0 | 0 | 12586 | 0 | 0 | 16155 | 0 | 0 | 1890 |
| 1998 | Nov-Dec | 0 | 0 | 1330 | 0 | 0 | 13196 | 0 | 0 | 454 |
| 1999 | Jan-Feb | 17660 | 0 | 0 | 30548 | 0 | 0 | 1900 | 0 | 0 |
| 1999 | Mar-Apr | 24661 | 0 | 0 | 20876 | 0 | 0 | 4937 | 0 | 0 |
| 1999 | May-Jul | 0 | 3028 | 0 | 0 | 3283 | 0 | 0 | 5420 | 0 |
| 1999 | Aug-Oct | 0 | 0 | 5658 | 0 | 0 | 20571 | 0 | 0 | 2054 |
| 1999 | Nov-Dec | 0 | 0 | 229 | 0 | 0 | 4986 | 0 | 0 | 56 |
| 2000 | Jan-Feb | 18935 | 0 | 0 | 30652 | 0 | 0 | 11647 | 0 | 0 |
| 2000 | Mar-Apr | 23194 | 0 | 0 | 8195 | 0 | 0 | 4105 | 0 | 0 |
| 2000 | May-Jul | 0 | 3800 | 0 | 0 | 1394 | 0 | 0 | 1077 | 0 |
| 2000 | Aug-Oct | 0 | 0 | 6199 | 0 | 0 | 22107 | 0 | 0 | 1667 |
| 2000 | Nov-Dec | 0 | 0 | 590 | 0 | 0 | 17816 | 0 | 0 | 0 |
| 2001 | Jan-Feb | 7962 | 0 | 0 | 18208 | 0 | 0 | 2206 | 0 | 0 |
| 2001 | Mar-Apr | 13895 | 0 | 0 | 16568 | 0 | 0 | 11279 | 0 | 0 |
| 2001 | May-Jul | 0 | 3500 | 0 | 0 | 3882 | 0 | 0 | 1005 | 0 |
| 2001 | Aug-Oct | 0 | 0 | 8904 | 0 | 0 | 30967 | 0 | 0 | 2970 |
| 2001 | Nov-Dec | 0 | 0 | 803 | 0 | 0 | 19752 | 0 | 0 | 641 |
| 2002 | Jan-Feb | 13410 | 0 | 0 | 35198 | 0 | 0 | 1845 | 0 | 0 |
| 2002 | Mar-Apr | 21130 | 0 | 0 | 14486 | 0 | 0 | 8407 | 0 | 0 |
| 2002 | May-Jul | 0 | 8163 | 0 | 0 | 1903 | 0 | 0 | 531 | 0 |
| 2002 | Aug-Oct | 0 | 0 | 8594 | 0 | 0 | 34463 | 0 | 0 | 2997 |
| 2002 | Nov-Dec | 0 | 0 | 291 | 0 | 0 | 14335 | 0 | 0 | 803 |
| 2003 | Jan-Feb | 15389 | 0 | 0 | 35435 | 0 | 0 | 11705 | 0 | 0 |
| 2003 | Mar-Apr | 16452 | 0 | 0 | 17100 | 0 | 0 | 1651 | 0 | 0 |
| 2003 | May-Jul | 0 | 6752 | 0 | 0 | 2748 | 0 | 0 | 454 | 0 |
| 2003 | Aug-Oct | 0 | 0 | 7793 | 0 | 0 | 35120 | 0 | 0 | 5141 |
| 2003 | Nov-Dec | 0 | 0 | 264 | 0 | 0 | 18004 | 0 | 0 | 1429 |
| 2004 | Jan-Feb | 21886 | 0 | 0 | 37436 | 0 | 0 | 9023 | 0 | 0 |
| 2004 | Mar-Apr | 17432 | 0 | 0 | 16627 | 0 | 0 | 2854 | 0 | 0 |
| 2004 | May-Jul | 0 | 9773 | 0 | 0 | 2919 | 0 | 0 | 946 | 0 |
| 2004 | Aug-Oct | 0 | 0 | 8766 | 0 | 0 | 31394 | 0 | 0 | 3841 |
| 2004 | Nov-Dec | 0 | 0 | 75 | 0 | 0 | 20181 | 0 | 0 | 596 |
| 2005 | Jan-Feb | 27361 | 0 | 0 | 46935 | 0 | 0 | 9033 | 0 | 0 |
| 2005 | Mar-Apr | 15119 | 0 | 0 | 6612 | 0 | 0 | 3114 | 0 | 0 |
| 2005 | May-Jul | 0 | 7410 | 0 | 0 | 3290 | 0 | 0 | 0 | 0 |
| 2005 | Aug-Oct | 0 | 0 | 2892 | 0 | 0 | 35350 | 0 | 0 | 4550 |
| 2005 | Nov-Dec | 0 | 0 | 113 | 0 | 0 | 20756 | 0 | 0 | 407 |

Table 2.5 (p. 4 of 4)— EBS catch (t) of Pacific cod by year, gear, and season for the years 1977-2015 as configured in Model 11.5. Aug-Oct and Nov-Dec catches for 2016 are extrapolated.

| Year | Season | Trawl fishery |  |  | Longline fishery |  |  | Pot fishery |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Jan-Apr | May-Jul | Aug-Dec | Jan-Apr | May-Jul | Aug-Dec | Jan-Apr | May-Jul | Aug-Dec |
| 2006 | Jan-Feb | 28611 | 0 | 0 | 45149 | 0 | 0 | 10608 | 0 | 0 |
| 2006 | Mar-Apr | 13901 | 0 | 0 | 6017 | 0 | 0 | 3297 | 0 | 0 |
| 2006 | May-Jul | 0 | 6347 | 0 | 0 | 1905 | 0 | 0 | 364 | 0 |
| 2006 | Aug-Oct | 0 | 0 | 4357 | 0 | 0 | 42493 | 0 | 0 | 3887 |
| 2006 | Nov-Dec | 0 | 0 | 70 | 0 | 0 | 1013 | 0 | 0 | 799 |
| 2007 | Jan-Feb | 15947 | 0 | 0 | 42943 | 0 | 0 | 10702 | 0 | 0 |
| 2007 | Mar-Apr | 16302 | 0 | 0 | 1917 | 0 | 0 | 1139 | 0 | 0 |
| 2007 | May-Jul | 0 | 10225 | 0 | 0 | 1213 | 0 | 0 | 479 | 0 |
| 2007 | Aug-Oct | 0 | 0 | 3190 | 0 | 0 | 30304 | 0 | 0 | 4922 |
| 2007 | Nov-Dec | 0 | 0 | 67 | 0 | 0 | 777 | 0 | 0 | 0 |
| 2008 | Jan-Feb | 15579 | 0 | 0 | 41627 | 0 | 0 | 8850 | 0 | 0 |
| 2008 | Mar-Apr | 7093 | 0 | 0 | 3657 | 0 | 0 | 1951 | 0 | 0 |
| 2008 | May-Jul | 0 | 3868 | 0 | 0 | 2665 | 0 | 0 | 225 | 0 |
| 2008 | Aug-Oct | 0 | 0 | 6306 | 0 | 0 | 33019 | 0 | 0 | 6218 |
| 2008 | Nov-Dec | 0 | 0 | 655 | 0 | 0 | 7966 | 0 | 0 | 124 |
| 2009 | Jan-Feb | 12194 | 0 | 0 | 44713 | 0 | 0 | 9395 | 0 | 0 |
| 2009 | Mar-Apr | 9602 | 0 | 0 | 3726 | 0 | 0 | 1722 | 0 | 0 |
| 2009 | May-Jul | 0 | 4174 | 0 | 0 | 2239 | 0 | 0 | 257 | 0 |
| 2009 | Aug-Oct | 0 | 0 | 10491 | 0 | 0 | 35381 | 0 | 0 | 1301 |
| 2009 | Nov-Dec | 0 | 0 | 403 | 0 | 0 | 10494 | 0 | 0 | 1081 |
| 2010 | Jan-Feb | 16351 | 0 | 0 | 40595 | 0 | 0 | 10695 | 0 | 0 |
| 2010 | Mar-Apr | 8148 | 0 | 0 | 2050 | 0 | 0 | 1726 | 0 | 0 |
| 2010 | May-Jul | 0 | 3982 | 0 | 0 | 2904 | 0 | 0 | 268 | 0 |
| 2010 | Aug-Oct | 0 | 0 | 9602 | 0 | 0 | 25115 | 0 | 0 | 5432 |
| 2010 | Nov-Dec | 0 | 0 | 1601 | 0 | 0 | 12616 | 0 | 0 | 1786 |
| 2011 | Jan-Feb | 21215 | 0 | 0 | 29312 | 0 | 0 | 15345 | 0 | 0 |
| 2011 | Mar-Apr | 20797 | 0 | 0 | 26006 | 0 | 0 | 2297 | 0 | 0 |
| 2011 | May-Jul | 0 | 7275 | 0 | 0 | 14044 | 0 | 0 | 594 | 0 |
| 2011 | Aug-Oct | 0 | 0 | 13355 | 0 | 0 | 31048 | 0 | 0 | 8954 |
| 2011 | Nov-Dec | 0 | 0 | 1728 | 0 | 0 | 17245 | 0 | 0 | 0 |
| 2012 | Jan-Feb | 39141 | 0 | 0 | 33808 | 0 | 0 | 19238 | 0 | 0 |
| 2012 | Mar-Apr | 14802 | 0 | 0 | 24489 | 0 | 0 | 2295 | 0 | 0 |
| 2012 | May-Jul | 0 | 8667 | 0 | 0 | 21241 | 0 | 0 | 794 | 0 |
| 2012 | Aug-Oct | 0 | 0 | 11670 | 0 | 0 | 27629 | 0 | 0 | 6232 |
| 2012 | Nov-Dec | 0 | 0 | 1058 | 0 | 0 | 20888 | 0 | 0 | 832 |
| 2013 | Jan-Feb | 35433 | 0 | 0 | 38744 | 0 | 0 | 19229 | 0 | 0 |
| 2013 | Mar-Apr | 16948 | 0 | 0 | 21974 | 0 | 0 | 3277 | 0 | 0 |
| 2013 | May-Jul | 0 | 5981 | 0 | 0 | 13880 | 0 | 0 | 0 | 0 |
| 2013 | Aug-Oct | 0 | 0 | 20904 | 0 | 0 | 26573 | 0 | 0 | 5892 |
| 2013 | Nov-Dec | 0 | 0 | 1608 | 0 | 0 | 22682 | 0 | 0 | 3567 |
| 2014 | Jan-Feb | 31400 | 0 | 0 | 32550 | 0 | 0 | 21523 | 0 | 0 |
| 2014 | Mar-Apr | 22055 | 0 | 0 | 26084 | 0 | 0 | 7124 | 0 | 0 |
| 2014 | May-Jul | 0 | 7069 | 0 | 0 | 21204 | 0 | 0 | 154 | 0 |
| 2014 | Aug-Oct | 0 | 0 | 11014 | 0 | 0 | 25890 | 0 | 0 | 6093 |
| 2014 | Nov-Dec | 0 | 0 | 990 | 0 | 0 | 22085 | 0 | 0 | 3528 |
| 2015 | Jan-Feb | 22015.6 | 0 | 0 | 27883.1 | 0 | 0 | 20031.6 | 0 | 0 |
| 2015 | Mar-Apr | 25510 | 0 | 0 | 27187.3 | 0 | 0 | 7904 | 0 | 0 |
| 2015 | May-Jul | 0 | 83.76 | 0 | 0 | 20844 | 0 | 0 | 160.726 | 0 |
| 2015 | Aug-Oct | 0 | 0 | 10615 | 0 | 0 | 30960 | 0 | 0 | 6771 |
| 2015 | Nov-Dec | 0 | 0 | 626 | 0 | 0 | 21221 | 0 | 0 | 3218 |

Table 2.6 (page 1 of 3)— Fishery CPUE as configured in Model 11.5. Units are $\mathrm{kg} /$ minute for trawl gear, $\mathrm{kg} / \mathrm{hook}$ for longline gear, and $\mathrm{kg} /$ pot for pot gear.

| Jan-Apr trawl fishery |  |  |  | May-Jul trawl fishery |  |  |  | Aug-Dec trawl fishery |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Season | CPUE | Sigma | Year | Season | CPUE | Sigma | Year | Season | CPUE | Sigma |
| 1991 | Jan-Feb | 55.864 | 0.091 | 1991 | May-Jul | 36.761 | 0.203 | 1991 | Aug-Oct | 71.702 | 0.603 |
| 1992 | Jan-Feb | 60.427 | 0.161 | 1992 | May-Jul | 38.568 | 0.291 | 1992 | Aug-Oct | 57.517 | 0.773 |
| 1993 | Jan-Feb | 62.047 | 0.157 | 1993 | May-Jul | 39.902 | 0.469 | 1993 | Aug-Oct | 113.970 | 0.503 |
| 1994 | Jan-Feb | 51.965 | 0.222 | 1994 | May-Jul | 26.767 | 0.248 | 1994 | Aug-Oct | 56.308 | 0.390 |
| 1995 | Jan-Feb | 88.482 | 0.122 | 1995 | May-Jul | 59.393 | 1.669 | 1995 | Aug-Oct | 60.164 | 0.323 |
| 1996 | Jan-Feb | 48.331 | 0.132 | 1996 | May-Jul | 29.174 | 0.314 | 1996 | Aug-Oct | 34.896 | 0.291 |
| 1997 | Jan-Feb | 75.605 | 0.122 | 1997 | May-Jul | 24.880 | 0.259 | 1997 | Aug-Oct | 62.619 | 0.567 |
| 1998 | Jan-Feb | 59.920 | 0.159 | 1998 | May-Jul | 26.245 | 0.303 | 1998 | Aug-Oct | 38.995 | 0.305 |
| 1999 | Jan-Feb | 42.399 | 0.119 | 1999 | May-Jul | 15.672 | 0.426 | 1999 | Aug-Oct | 20.611 | 0.367 |
| 2000 | Jan-Feb | 34.522 | 0.123 | 2000 | May-Jul | 32.694 | 0.293 | 2000 | Aug-Oct | 15.070 | 0.528 |
| 2001 | Jan-Feb | 25.452 | 0.166 | 2001 | May-Jul | 60.120 | 0.298 | 2001 | Aug-Oct | 16.662 | 0.249 |
| 2002 | Jan-Feb | 35.892 | 0.141 | 2002 | May-Jul | 39.985 | 0.209 | 2002 | Aug-Oct | 15.141 | 0.196 |
| 2003 | Jan-Feb | 24.642 | 0.169 | 2003 | May-Jul | 49.493 | 0.210 | 2003 | Aug-Oct | 19.171 | 0.156 |
| 2004 | Jan-Feb | 62.609 | 0.138 | 2004 | May-Jul | 34.588 | 0.163 | 2004 | Aug-Oct | 21.519 | 0.154 |
| 2005 | Jan-Feb | 43.993 | 0.116 | 2005 | May-Jul | 24.100 | 0.172 | 2005 | Aug-Oct | 15.932 | 0.834 |
| 2006 | Jan-Feb | 36.397 | 0.107 | 2006 | May-Jul | 30.653 | 0.186 | 2006 | Aug-Oct | 26.772 | 0.376 |
| 2007 | Jan-Feb | 30.849 | 0.095 | 2007 | May-Jul | 39.485 | 0.114 | 2007 | Aug-Oct | 18.147 | 0.681 |
| 2008 | Jan-Feb | 24.385 | 0.152 | 2008 | May-Jul | 40.650 | 0.251 | 2008 | Aug-Oct | 60.047 | 0.336 |
| 2009 | Jan-Feb | 37.853 | 0.171 | 2009 | May-Jul | 33.932 | 0.292 | 2009 | Aug-Oct | 54.154 | 0.226 |
| 2010 | Jan-Feb | 41.949 | 0.137 | 2010 | May-Jul | 32.031 | 0.336 | 2010 | Aug-Oct | 73.484 | 0.198 |
| 2011 | Jan-Feb | 50.737 | 0.110 | 2011 | May-Jul | 49.228 | 0.259 | 2011 | Aug-Oct | 56.918 | 0.202 |
| 2012 | Jan-Feb | 97.338 | 0.099 | 2012 | May-Jul | 117.192 | 0.247 | 2012 | Aug-Oct | 52.247 | 0.255 |
| 2013 | Jan-Feb | 67.061 | 0.081 | 2013 | May-Jul | 39.218 | 0.291 | 2013 | Aug-Oct | 82.463 | 0.140 |
| 2014 | Jan-Feb | 57.039 | 0.087 | 2014 | May-Jul | 53.157 | 0.224 | 2014 | Aug-Oct | 56.967 | 0.283 |
| 2015 | Jan-Feb | 49.877 | 0.112 | 2015 | May-Jul | 52.999 | 0.215 | 2015 | Aug-Oct | 86.387 | 0.266 |
| 2016 | Jan-Feb | 66.897 | 0.092 | 2016 | May-Jul | 31.841 | 0.306 | 2016 | Aug-Oct | 26.146 | 0.546 |
| 1991 | Mar-Apr | 61.454 | 0.059 |  |  |  |  | 1993 | Nov-Dec | 32.678 | 0.914 |
| 1992 | Mar-Apr | 48.269 | 0.069 |  |  |  |  | 1996 | Nov-Dec | 29.543 | 0.482 |
| 1993 | Mar-Apr | 48.840 | 0.073 |  |  |  |  | 1997 | Nov-Dec | 31.309 | 1.093 |
| 1994 | Mar-Apr | 52.428 | 0.053 |  |  |  |  | 1998 | Nov-Dec | 16.891 | 0.646 |
| 1995 | Mar-Apr | 55.463 | 0.061 |  |  |  |  | 1999 | Nov-Dec | 12.994 | 0.964 |
| 1996 | Mar-Apr | 33.954 | 0.051 |  |  |  |  | 2009 | Nov-Dec | 28.369 | 1.180 |
| 1997 | Mar-Apr | 45.985 | 0.062 |  |  |  |  | 2010 | Nov-Dec | 40.079 | 0.681 |
| 1998 | Mar-Apr | 31.809 | 0.072 |  |  |  |  | 2011 | Nov-Dec | 20.796 | 1.180 |
| 1999 | Mar-Apr | 35.675 | 0.086 |  |  |  |  | 2012 | Nov-Dec | 52.570 | 1.293 |
| 2000 | Mar-Apr | 31.397 | 0.085 |  |  |  |  | 2013 | Nov-Dec | 17.174 | 1.669 |
| 2001 | Mar-Apr | 21.213 | 0.106 |  |  |  |  | 2014 | Nov-Dec | 24.191 | 1.180 |
| 2002 | Mar-Apr | 26.640 | 0.103 |  |  |  |  |  |  |  |  |
| 2003 | Mar-Apr | 28.131 | 0.095 |  |  |  |  |  |  |  |  |
| 2004 | Mar-Apr | 42.816 | 0.116 |  |  |  |  |  |  |  |  |
| 2005 | Mar-Apr | 48.932 | 0.114 |  |  |  |  |  |  |  |  |
| 2006 | Mar-Apr | 56.188 | 0.141 |  |  |  |  |  |  |  |  |
| 2007 | Mar-Apr | 45.097 | 0.092 |  |  |  |  |  |  |  |  |
| 2008 | Mar-Apr | 40.343 | 0.196 |  |  |  |  |  |  |  |  |
| 2009 | Mar-Apr | 55.557 | 0.183 |  |  |  |  |  |  |  |  |
| 2010 | Mar-Apr | 55.766 | 0.266 |  |  |  |  |  |  |  |  |
| 2011 | Mar-Apr | 76.788 | 0.148 |  |  |  |  |  |  |  |  |
| 2012 | Mar-Apr | 76.796 | 0.154 |  |  |  |  |  |  |  |  |
| 2013 | Mar-Apr | 64.027 | 0.138 |  |  |  |  |  |  |  |  |
| 2014 | Mar-Apr | 61.816 | 0.101 |  |  |  |  |  |  |  |  |
| 2015 | Mar-Apr | 72.289 | 0.100 |  |  |  |  |  |  |  |  |
| 2016 | Mar-Apr | 85.530 | 0.129 |  |  |  |  |  |  |  |  |

Table 2.6 (page 2 of 3)— Fishery CPUE as configured in Model 11.5. Units are kg/minute for trawl gear, $\mathrm{kg} / \mathrm{hook}$ for longline gear, and $\mathrm{kg} /$ pot for pot gear.

| Jan-Apr longline fishery |  |  |  | May-Jul longline fishery |  |  |  | Aug-Dec longline fishery |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Season | CPUE | Sigma | Year | Season | CPUE | Sigma | Year | Season | CPUE | Sigma |
| 1991 | Jan-Feb | 1.124 | 0.156 | 1991 | May-Jul | 0.771 | 0.075 | 1991 | Aug-Oct | 0.595 | 0.062 |
| 1992 | Jan-Feb | 0.873 | 0.088 | 1992 | May-Jul | 0.530 | 0.052 | 1992 | Aug-Oct | 0.512 | 0.069 |
| 1993 | Jan-Feb | 0.654 | 0.066 | 1993 | May-Jul | 0.551 | 0.176 | 1994 | Aug-Oct | 0.576 | 0.068 |
| 1994 | Jan-Feb | 0.728 | 0.068 | 1994 | May-Jul | 0.713 | 0.133 | 1995 | Aug-Oct | 0.587 | 0.070 |
| 1995 | Jan-Feb | 0.895 | 0.069 | 1995 | May-Jul | 0.760 | 0.179 | 1996 | Aug-Oct | 0.542 | 0.060 |
| 1996 | Jan-Feb | 0.878 | 0.069 | 1996 | May-Jul | 0.669 | 0.178 | 1997 | Aug-Oct | 0.580 | 0.064 |
| 1997 | Jan-Feb | 0.989 | 0.072 | 1997 | May-Jul | 0.657 | 0.120 | 1998 | Aug-Oct | 0.398 | 0.064 |
| 1998 | Jan-Feb | 0.888 | 0.074 | 1998 | May-Jul | 0.496 | 0.184 | 1999 | Aug-Oct | 0.481 | 0.061 |
| 1999 | Jan-Feb | 0.743 | 0.067 | 1999 | May-Jul | 0.637 | 0.143 | 2000 | Aug-Oct | 0.404 | 0.053 |
| 2000 | Jan-Feb | 0.730 | 0.069 | 2000 | May-Jul | 0.610 | 0.169 | 2001 | Aug-Oct | 0.398 | 0.052 |
| 2001 | Jan-Feb | 0.586 | 0.079 | 2001 | May-Jul | 0.514 | 0.107 | 2002 | Aug-Oct | 0.372 | 0.046 |
| 2002 | Jan-Feb | 0.680 | 0.062 | 2002 | May-Jul | 0.405 | 0.137 | 2003 | Aug-Oct | 0.342 | 0.044 |
| 2003 | Jan-Feb | 0.517 | 0.053 | 2003 | May-Jul | 0.376 | 0.109 | 2004 | Aug-Oct | 0.312 | 0.048 |
| 2004 | Jan-Feb | 0.562 | 0.060 | 2004 | May-Jul | 0.367 | 0.115 | 2005 | Aug-Oct | 0.330 | 0.045 |
| 2005 | Jan-Feb | 0.626 | 0.055 | 2005 | May-Jul | 0.385 | 0.106 | 2006 | Aug-Oct | 0.391 | 0.047 |
| 2006 | Jan-Feb | 0.747 | 0.062 | 2006 | May-Jul | 0.366 | 0.162 | 2007 | Aug-Oct | 0.402 | 0.038 |
| 2007 | Jan-Feb | 0.734 | 0.045 | 2007 | May-Jul | 0.406 | 0.142 | 2008 | Aug-Oct | 0.307 | 0.048 |
| 2008 | Jan-Feb | 0.794 | 0.068 | 2008 | May-Jul | 0.366 | 0.140 | 2009 | Aug-Oct | 0.348 | 0.049 |
| 2009 | Jan-Feb | 0.893 | 0.069 | 2009 | May-Jul | 0.384 | 0.151 | 2010 | Aug-Oct | 0.352 | 0.060 |
| 2010 | Jan-Feb | 0.781 | 0.067 | 2010 | May-Jul | 0.419 | 0.156 | 2011 | Aug-Oct | 0.369 | 0.059 |
| 2011 | Jan-Feb | 0.716 | 0.083 | 2011 | May-Jul | 0.374 | 0.088 | 2012 | Aug-Oct | 0.321 | 0.060 |
| 2012 | Jan-Feb | 0.774 | 0.081 | 2012 | May-Jul | 0.429 | 0.080 | 2013 | Aug-Oct | 0.355 | 0.057 |
| 2013 | Jan-Feb | 0.736 | 0.062 | 2013 | May-Jul | 0.424 | 0.091 | 2014 | Aug-Oct | 0.360 | 0.058 |
| 2014 | Jan-Feb | 0.599 | 0.068 | 2014 | May-Jul | 0.356 | 0.064 | 2015 | Aug-Oct | 0.372 | 0.055 |
| 2015 | Jan-Feb | 0.576 | 0.072 | 2015 | May-Jul | 0.459 | 0.079 | 2016 | Aug-Oct | 0.419 | 0.123 |
| 2016 | Jan-Feb | 0.665 | 0.073 | 2016 | May-Jul | 0.440 | 0.080 | 1991 | Nov-Dec | 0.551 | 0.093 |
| 1991 | Mar-Apr | 0.993 | 0.110 |  |  |  |  | 1995 | Nov-Dec | 0.648 | 0.110 |
| 1992 | Mar-Apr | 0.858 | 0.071 |  |  |  |  | 1996 | Nov-Dec | 0.590 | 0.277 |
| 1993 | Mar-Apr | 0.669 | 0.061 |  |  |  |  | 1997 | Nov-Dec | 0.577 | 0.073 |
| 1994 | Mar-Apr | 0.735 | 0.060 |  |  |  |  | 1998 | Nov-Dec | 0.501 | 0.072 |
| 1995 | Mar-Apr | 0.841 | 0.062 |  |  |  |  | 1999 | Nov-Dec | 0.541 | 0.120 |
| 1996 | Mar-Apr | 0.756 | 0.067 |  |  |  |  | 2000 | Nov-Dec | 0.416 | 0.067 |
| 1997 | Mar-Apr | 0.829 | 0.078 |  |  |  |  | 2001 | Nov-Dec | 0.432 | 0.065 |
| 1998 | Mar-Apr | 0.619 | 0.075 |  |  |  |  | 2002 | Nov-Dec | 0.394 | 0.072 |
| 1999 | Mar-Apr | 0.617 | 0.067 |  |  |  |  | 2003 | Nov-Dec | 0.365 | 0.060 |
| 2000 | Mar-Apr | 0.617 | 0.097 |  |  |  |  | 2004 | Nov-Dec | 0.441 | 0.065 |
| 2001 | Mar-Apr | 0.539 | 0.073 |  |  |  |  | 2005 | Nov-Dec | 0.385 | 0.064 |
| 2002 | Mar-Apr | 0.676 | 0.082 |  |  |  |  | 2006 | Nov-Dec | 0.433 | 0.214 |
| 2003 | Mar-Apr | 0.530 | 0.068 |  |  |  |  | 2007 | Nov-Dec | 0.449 | 0.332 |
| 2004 | Mar-Apr | 0.579 | 0.076 |  |  |  |  | 2008 | Nov-Dec | 0.449 | 0.087 |
| 2005 | Mar-Apr | 0.678 | 0.113 |  |  |  |  | 2009 | Nov-Dec | 0.428 | 0.090 |
| 2006 | Mar-Apr | 0.796 | 0.112 |  |  |  |  | 2010 | Nov-Dec | 0.447 | 0.087 |
| 2007 | Mar-Apr | 0.693 | 0.155 |  |  |  |  | 2011 | Nov-Dec | 0.447 | 0.086 |
| 2008 | Mar-Apr | 0.774 | 0.145 |  |  |  |  | 2012 | Nov-Dec | 0.476 | 0.077 |
| 2009 | Mar-Apr | 1.159 | 0.172 |  |  |  |  | 2013 | Nov-Dec | 0.479 | 0.071 |
| 2010 | Mar-Apr | 0.829 | 0.195 |  |  |  |  | 2014 | Nov-Dec | 0.439 | 0.067 |
| 2011 | Mar-Apr | 0.703 | 0.072 |  |  |  |  | 2015 | Nov-Dec | 0.463 | 0.076 |
| 2012 | Mar-Apr | 0.597 | 0.082 |  |  |  |  |  |  |  |  |
| 2013 | Mar-Apr | 0.659 | 0.083 |  |  |  |  |  |  |  |  |
| 2014 | Mar-Apr | 0.523 | 0.071 |  |  |  |  |  |  |  |  |
| 2015 | Mar-Apr | 0.571 | 0.073 |  |  |  |  |  |  |  |  |
| 2016 | Mar-Apr | 0.562 | 0.085 |  |  |  |  |  |  |  |  |

Table 2.6 (page 3 of 3)— Fishery CPUE as configured in Model 11.5. Units are kg/minute for trawl gear, $\mathrm{kg} / \mathrm{hook}$ for longline gear, and $\mathrm{kg} /$ pot for pot gear.

| Jan-Apr pot fishery |  |  |  | May-Jul pot fishery |  |  |  | Aug-Dec pot fishery |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Season | CPUE | Sigma | Year | Season | CPUE | Sigma | Year | Season | CPUE | Sigma |
| 2000 | Jan-Feb | 56.553 | 0.152 | 1991 | May-Jul | 64.037 | 0.251 | 1991 | Aug-Oct | 88.556 | 0.132 |
| 2001 | Jan-Feb | 72.207 | 0.503 | 1992 | May-Jul | 66.730 | 0.077 | 1992 | Aug-Oct | 30.252 | 0.113 |
| 2002 | Jan-Feb | 81.893 | 0.264 | 1993 | May-Jul | 90.669 | 0.228 | 1994 | Aug-Oct | 97.172 | 0.151 |
| 2003 | Jan-Feb | 73.858 | 0.139 | 1994 | May-Jul | 75.421 | 0.173 | 1995 | Aug-Oct | 57.783 | 0.153 |
| 2004 | Jan-Feb | 78.980 | 0.170 | 1995 | May-Jul | 72.065 | 0.098 | 1996 | Aug-Oct | 49.758 | 0.136 |
| 2005 | Jan-Feb | 85.328 | 0.168 | 1996 | May-Jul | 55.819 | 0.089 | 1997 | Aug-Oct | 47.938 | 0.167 |
| 2006 | Jan-Feb | 83.292 | 0.154 | 1997 | May-Jul | 46.843 | 0.114 | 1998 | Aug-Oct | 32.057 | 0.281 |
| 2007 | Jan-Feb | 64.671 | 0.109 | 1998 | May-Jul | 49.999 | 0.129 | 1999 | Aug-Oct | 37.675 | 0.213 |
| 2008 | Jan-Feb | 81.642 | 0.208 | 1999 | May-Jul | 47.466 | 0.124 | 2001 | Aug-Oct | 46.493 | 0.169 |
| 2009 | Jan-Feb | 92.345 | 0.189 |  |  |  |  | 2002 | Aug-Oct | 42.331 | 0.189 |
| 2010 | Jan-Feb | 88.535 | 0.167 |  |  |  |  | 2003 | Aug-Oct | 57.632 | 0.174 |
| 2011 | Jan-Feb | 130.718 | 0.153 |  |  |  |  | 2004 | Aug-Oct | 48.802 | 0.210 |
| 2012 | Jan-Feb | 138.710 | 0.148 |  |  |  |  | 2005 | Aug-Oct | 45.872 | 0.192 |
| 2013 | Jan-Feb | 128.974 | 0.143 |  |  |  |  | 2006 | Aug-Oct | 55.342 | 0.185 |
| 2014 | Jan-Feb | 105.380 | 0.145 |  |  |  |  | 2007 | Aug-Oct | 65.356 | 0.151 |
| 2015 | Jan-Feb | 105.052 | 0.129 |  |  |  |  | 2008 | Aug-Oct | 57.252 | 0.164 |
| 2016 | Jan-Feb | 97.702 | 0.122 |  |  |  |  | 2009 | Aug-Oct | 72.836 | 0.266 |
| 1992 | Mar-Apr | 86.412 | 0.422 |  |  |  |  | 2010 | Aug-Oct | 82.936 | 0.210 |
| 1993 | Mar-Apr | 84.191 | 0.136 |  |  |  |  | 2011 | Aug-Oct | 81.445 | 0.148 |
| 1994 | Mar-Apr | 89.313 | 0.107 |  |  |  |  | 2012 | Aug-Oct | 64.934 | 0.130 |
| 1995 | Mar-Apr | 91.679 | 0.094 |  |  |  |  | 2013 | Aug-Oct | 87.471 | 0.128 |
| 1996 | Mar-Apr | 73.485 | 0.077 |  |  |  |  | 2014 | Aug-Oct | 77.822 | 0.162 |
| 1997 | Mar-Apr | 93.226 | 0.120 |  |  |  |  | 2015 | Aug-Oct | 82.978 | 0.113 |
| 1998 | Mar-Apr | 77.558 | 0.184 |  |  |  |  | 1991 | Nov-Dec | 91.633 | 0.262 |
| 1999 | Mar-Apr | 67.604 | 0.195 |  |  |  |  | 1995 | Nov-Dec | 53.251 | 0.188 |
| 2000 | Mar-Apr | 45.310 | 0.163 |  |  |  |  | 1996 | Nov-Dec | 46.456 | 0.422 |
| 2001 | Mar-Apr | 69.247 | 0.137 |  |  |  |  | 1997 | Nov-Dec | 41.829 | 0.413 |
| 2002 | Mar-Apr | 61.628 | 0.176 |  |  |  |  | 1998 | Nov-Dec | 41.138 | 0.802 |
| 2004 | Mar-Apr | 65.936 | 0.390 |  |  |  |  | 2001 | Nov-Dec | 40.740 | 0.631 |
| 2006 | Mar-Apr | 116.202 | 0.422 |  |  |  |  | 2002 | Nov-Dec | 55.955 | 0.417 |
| 2014 | Mar-Apr | 183.575 | 0.353 |  |  |  |  | 2003 | Nov-Dec | 60.093 | 0.334 |
| 2015 | Mar-Apr | 133.103 | 0.173 |  |  |  |  | 2004 | Nov-Dec | 66.375 | 0.451 |
| 2016 | Mar-Apr | 118.028 | 0.244 |  |  |  |  | 2006 | Nov-Dec | 37.187 | 0.422 |
|  |  |  |  |  |  |  |  | 2010 | Nov-Dec | 104.985 | 0.373 |
|  |  |  |  |  |  |  |  | 2013 | Nov-Dec | 90.404 | 0.213 |
|  |  |  |  |  |  |  |  | 2014 | Nov-Dec | 69.205 | 0.210 |
|  |  |  |  |  |  |  |  | 2015 | Nov-Dec | 71.605 | 0.220 |

Table 2.7— Total biomass and abundance, with standard deviations, as estimated by EBS shelf bottom trawl surveys, 1982-2016. For biomass, lower and upper $95 \%$ confidence intervals are also shown.

|  | Biomass $(\mathrm{t})$ |  |  |  |  | Abundance (1000s of fish) |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
| Year | Estimate | Std. error | L95\% CI | U95\% CI | Estimate | Std. error |  |
| 1982 | $1,013,061$ | 73,621 | 867,292 | $1,158,831$ | 583,781 | 38,064 |  |
| 1983 | $1,187,096$ | 120,958 | 942,640 | $1,431,553$ | 752,456 | 80,566 |  |
| 1984 | $1,013,558$ | 62,513 | 889,782 | $1,137,334$ | 651,058 | 47,126 |  |
| 1985 | $1,001,112$ | 55,845 | 890,540 | $1,111,684$ | 841,108 | 113,438 |  |
| 1986 | $1,118,006$ | 69,626 | 980,146 | $1,255,866$ | 838,217 | 83,855 |  |
| 1987 | $1,027,518$ | 63,670 | 901,452 | $1,153,584$ | 677,054 | 44,120 |  |
| 1988 | 960,962 | 76,961 | 808,579 | $1,113,344$ | 507,560 | 35,581 |  |
| 1989 | 833,473 | 62,713 | 709,300 | 957,645 | 292,247 | 19,986 |  |
| 1990 | 691,256 | 51,455 | 589,376 | 793,136 | 423,835 | 36,466 |  |
| 1991 | 514,407 | 38,039 | 439,090 | 589,725 | 488,892 | 51,108 |  |
| 1992 | 529,049 | 44,616 | 440,708 | 617,390 | 577,560 | 68,603 |  |
| 1993 | 663,308 | 53,143 | 558,085 | 768,531 | 810,608 | 99,259 |  |
| 1994 | $1,360,790$ | 247,737 | 865,316 | $1,856,263$ | $1,232,175$ | 152,212 |  |
| 1995 | $1,002,961$ | 91,622 | 821,550 | $1,184,372$ | 757,910 | 75,473 |  |
| 1996 | 889,366 | 87,521 | 716,076 | $1,062,657$ | 607,198 | 88,384 |  |
| 1997 | 604,439 | 68,120 | 468,199 | 740,678 | 485,643 | 70,802 |  |
| 1998 | 534,150 | 42,937 | 449,135 | 619,165 | 514,339 | 46,852 |  |
| 1999 | 569,765 | 49,471 | 471,811 | 667,718 | 488,337 | 45,289 |  |
| 2000 | 531,171 | 43,160 | 445,714 | 616,627 | 483,808 | 44,188 |  |
| 2001 | 811,816 | 73,211 | 665,394 | 958,239 | 960,917 | 91,898 |  |
| 2002 | 584,565 | 63,820 | 456,926 | 712,205 | 536,342 | 53,802 |  |
| 2003 | 590,973 | 62,121 | 466,732 | 715,214 | 498,873 | 62,220 |  |
| 2004 | 562,309 | 33,739 | 495,505 | 629,113 | 397,948 | 34,332 |  |
| 2005 | 606,050 | 43,056 | 520,799 | 691,301 | 450,705 | 63,363 |  |
| 2006 | 517,698 | 28,341 | 461,583 | 573,813 | 394,024 | 23,785 |  |
| 2007 | 423,704 | 34,811 | 354,081 | 493,326 | 733,402 | 195,956 |  |
| 2008 | 403,125 | 26,822 | 350,018 | 456,232 | 476,697 | 49,413 |  |
| 2009 | 421,291 | 34,969 | 352,053 | 490,530 | 716,637 | 62,705 |  |
| 2010 | 860,210 | 102,307 | 657,642 | $1,062,778$ | 887,836 | 117,022 |  |
| 2011 | 896,039 | 66,843 | 763,690 | $1,028,388$ | 836,822 | 79,207 |  |
| 2012 | 890,665 | 100,473 | 689,718 | $1,091,612$ | 987,973 | 91,589 |  |
| 2013 | 791,958 | 73,952 | 644,054 | 939,862 | 750,889 | 124,917 |  |
| 2014 | $1,079,712$ | 153,299 | 769,895 | $1,389,528$ | $1,122,144$ | 143,618 |  |
| 2015 | $1,102,261$ | 150,981 | 800,299 | $1,404,223$ | 982,470 | 113,501 |  |
| 2016 | 944,621 | 76,948 | 790,725 | $1,098,516$ | 640,359 | 61,639 |  |
|  |  |  |  |  |  |  |  |

Table 2.8 (page 1 of 4)—Trawl survey size composition, by year and cm (sample size in column 2).

| Year | Nact |  | 45 | 56 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 882 | 10548 |  | 00 | 0 | 0 | 0 | 1 | 8 | 9 | 19 | 26 | 52 | 59 | 109 | 66 | 51 | 52 | 46 | 19 | 8 | 9 | 2 | 8 | 18 | 25 | 40 | 67 | 87 | 123 | 193 | 22 |
| 1983 | 13149 |  | 00 | 00 | 0 | 0 | 7 | 96 | 290 | 455 | 458 | 484 | 461 | 433 | 394 | 252 | 250 | 120 | 74 | 44 | 29 | 9 | 5 | 18 | 34 | 46 | 56 | 100 | 125 | 146 | 173 |
| 1984 | 12145 |  | 00 | 00 | 0 | 0 | 6 | 25 | 36 | 55 | 43 | 27 | 25 | 26 | 30 | 46 | 32 | 64 | 72 | 90 | 125 | 231 | 312 | 379 | 460 | 574 | 600 | 647 | 569 | 477 | 394 |
| 1985 | 16880 |  | 00 | 0 | 0 | 0 | 4 | 56 | 102 | 179 | 145 | 216 | 287 | 304 | 372 | 503 | 507 | 526 | 647 | 559 | 555 | 32 | 212 | 130 | 91 | 100 | 06 | 59 | 220 | 216 | 272 |
| 1986 | 15378 |  | 00 | 0 | 0 | 1 | 23 | 38 | 93 | 133 | 130 | 202 | 175 | 177 | 150 | 93 | 34 | 27 | 20 | 22 | 72 | 114 | 218 | 360 | 449 | 697 | 629 | 616 | 638 | 653 | 580 |
| 198 | 10599 |  | 00 | 0 | 0 | 0 | 0 | 14 | 4 | 7 | 24 | 38 | 60 | 81 | 108 | 121 | 121 | 153 | 124 | 80 | 61 | 47 | 63 | 76 | 117 | 124 | 200 | 274 | 302 | 325 | 291 |
| 1988 | 9991 |  | 00 | 0 | 0 | 0 | 0 |  | 8 | 7 | 28 | 13 | 27 | 26 | 23 | 42 | 27 | 18 | 26 | 35 | 48 | 68 | 77 | 88 | 86 | 109 | 83 | 124 | 122 | 137 | 179 |
| 1989 | 10001 |  | 00 | 00 | 0 | 0 | 3 | 3 | 19 | 47 | 37 | 70 | 86 | 108 | 105 | 101 | 66 | 39 | 19 | 21 | 30 | 4 | 15 | 16 | 35 | 13 | 34 | 30 | 24 | 33 | 37 |
| 1990 | 5630 |  | 00 | 00 | 0 | 0 | 26 | 71 | 104 | 154 | 150 | 185 | 236 | 259 | 205 | 149 | 117 | 89 | 57 | 35 | 41 | 42 | 33 | 47 | 76 | 77 | 95 | 103 | 97 | 92 | 118 |
| 1991 | 7220 |  | 00 | 0 | 0 | 0 | 6 | 31 | 94 | 112 | 140 | 137 | 163 | 133 | 136 | 128 | 107 | 135 | 86 | 72 | 72 | 78 | 100 | 97 | 166 | 192 | 265 | 285 | 325 | 289 | 372 |
| 199 | 9599 |  | 0 | 0 | 0 | 0 | 0 | 1 | 17 | 81 | 183 | 91 | 75 | 150 | 198 | 221 | 233 | 247 | 215 | 227 | 111 | 119 | 135 | 182 | 264 | 288 | 302 | 349 | 373 | 348 | 31 |
| 199 | 10402 |  | 00 | 0 | 0 | 1 | 2 | 29 | 81 | 191 | 423 | 293 | 403 | 354 | 32 | 318 | 343 | 31 | 321 | 215 | 134 | 97 | 61 | 55 | 66 | 85 | 94 | 173 | 206 | 230 | 290 |
| 199 | 13924 |  | 00 | 0 | 0 | 0 | 3 | 10 | 5 | 27 | 42 | 76 | 91 | 100 | 100 | 116 | 136 | 11 | 103 | 91 | 131 | 120 | 171 | 154 | 205 | 321 | 430 | 552 | 639 | 732 | 767 |
| 199 | 9208 |  | 00 | 0 | 0 | 0 | 3 | 12 | 15 | 13 | 19 | 41 | 37 | 42 | 56 | 59 | 81 | 68 | 34 | 24 | 19 | 37 | 47 | 89 | 108 | 158 | 194 | 228 | 218 | 245 | 225 |
| 199 | 9349 |  | 00 | 00 | 0 | 0 | 1 | 2 | 11 | 9 | 23 | 33 | 48 | 64 | 53 | 66 | 69 | 64 | 54 | 36 | 20 | 22 | 23 | 58 | 65 | 129 | 163 | 194 | 229 | 275 | 237 |
| 199 | 9177 |  | 00 | 0 | 0 | 0 | 8 | 17 | 65 | 114 | 167 | 193 | 192 | 196 | 212 | 284 | 226 | 218 | 226 | 178 | 105 | 58 | 41 | 41 | 34 | 70 | 109 | 103 | 154 | 223 | 231 |
| 1998 | 9572 |  | 00 | 0 | 0 | 0 | 1 | 4 | 23 | 53 | 84 | 117 | 104 | 136 | 91 | 45 | 22 | 6 | 4 | 17 | 25 | 57 | 72 | 182 | 276 | 381 | 494 | 599 | 628 | 614 | 13 |
| 1999 | 11695 |  | 00 | 0 | 0 | 0 | 1 | 15 | 53 | 100 | 109 | 122 | 94 | 113 | 78 | 42 | 30 | 41 | 49 | 39 | 53 | 109 | 110 | 196 | 228 | 222 | 310 | 268 | 295 | 308 | 240 |
| 2000 | 12547 |  | 00 | 00 | 410 | 10 | 23 | 51 | 99 | 137 | 298 | 478 | 582 | 442 | 278 | 274 | 141 | 87 | 33 | 9 | 12 | 25 | 39 | 77 | 119 | 170 | 197 | 220 | 258 | 305 | 22 |
| 200 | 19748 |  | 00 | 00 | 0 | 5 | 6 | 27 | 63 | 127 | 204 | 312 | 449 | 658 | 710 | 766 | 678 | 662 | 440 | 349 | 219 | 136 | 112 | 160 | 226 | 314 | 365 | 507 | 657 | 832 | 22 |
| 200 | 12237 |  | 00 | 0 | 0 | 1 | 3 | 6 | 21 | 43 | 63 | 80 | 101 | 159 | 112 | 166 | 111 | 71 | 51 | 35 | 17 | 42 | 63 | 106 | 160 | 240 | 268 | 434 | 474 | 555 | 553 |
| 200 | 12353 |  | 00 | 0 | 0 | 1 | 3 | 5 | 11 | 56 | 92 | 138 | 205 | 232 | 206 | 249 | 254 | 282 | 252 | 237 | 199 | 218 | 154 | 120 | 66 | 57 | 59 | 79 | 57 | 115 | 144 |
| 200 | 10811 |  | 02 | 2 | 0 | 0 | 1 | 4 | 20 | 45 | 86 | 152 | 106 | 194 | 187 | 215 | 211 | 135 | 144 | 111 | 65 | 56 | 72 | 92 | 104 | 188 | 196 | 219 | 238 | 273 | 30 |
| 2005 | 11288 |  | 00 | 0 | 0 | 0 | 0 | 1 | 4 | 22 | 43 | 87 | 138 | 201 | 248 | 304 | 284 | 301 | 290 | 362 | 362 | 387 | 376 | 289 | 210 | 137 | 135 | 141 | 115 | 158 | 178 |
| 2006 | 12131 |  | 01 | 1 | 4 | 7 | 40 | 101 | 336 | 405 | 427 | 453 | 401 | 343 | 330 | 359 | 280 | 243 | 146 | 105 | 65 | 54 | 56 | 55 | 64 | 86 | 115 | 168 | 189 | 246 | 243 |
| 2007 | 12809 |  | 00 | 00 | 0 | 7 | 7 | 129 | 481 | 1163 | 1425 | 1398 | 1141 | 730 | 715 | 511 | 326 | 400 | 230 | 121 | 122 | 42 | 44 | 65 | 86 | 124 | 117 | 154 | 122 | 140 | 14 |
| 2008 | 12985 |  | 00 | 0 | 0 | 0 | 6 | 54 | 169 | 350 | 380 | 390 | 350 | 312 | 227 | 151 | 75 | 40 | 21 | 40 | 70 | 162 | 307 | 479 | 550 | 707 | 745 | 719 | 682 | 559 | 461 |
| 20 | 16679 |  | 10 | 00 | 73 | 36 | 106 | 401 | 971 | 1058 | 1087 | 878 | 744 | 650 | 485 | 460 | 318 | 219 | 114 | 35 | 28 | 33 | 82 | 93 | 173 | 253 | 336 | 396 | 467 | 436 | 33 |
| 2010 | 7564 |  | 00 | 00 | 0 | 0 | 1 | 5 | 18 | 24 | 29 | 50 | 50 | 56 | 46 | 31 | 15 | 17 | 9 | 13 | 31 | 60 | 126 | 193 | 241 | 355 | 431 | 417 | 394 | 394 | 32 |
| 201 | 20739 |  | 00 | 00 | 0 | 0 | 8 | 20 | 76 | 142 | 257 | 306 | 385 | 413 | 597 | 627 | 905 | 886 | 851 | 536 | 286 | 109 | 34 | 37 | 55 | 48 | 56 | 72 | 121 | 136 | 188 |
| 2012 | 13076 |  | 00 | 0 | 0 | 0 | 74 | 379 | 686 | 732 | 563 | 424 | 417 | 310 | 410 | 396 | 208 | 129 | 48 | 31 | 10 | 28 | 37 | 59 | 84 | 178 | 259 | 269 | 358 | 352 | 390 |
| 2013 | 18691 |  | 00 | 0 | 0 | 1 | 9 | 50 | 116 | 146 | 207 | 222 | 283 | 239 | 177 | 127 | 35 | 22 | 63 | 86 | 268 | 398 | 653 | 785 | 982 | 1078 | 840 | 908 | 652 | 658 | 415 |
| 2014 | 17944 |  | 00 | 00 | 1 | 0 | 1 | 9 | 90 | 117 | 239 | 340 | 466 | 519 | 657 | 498 | 608 | 490 | 520 | 308 | 218 | 111 | 103 | 91 | 72 | 96 | 221 | 247 | 419 | 331 | 484 |
| 2015 | 19316 |  | 00 | 00 | 0 | 0 | 0 | 11 | 42 | 42 | 85 | 77 | 52 | 47 | 57 | 57 | 60 | 74 | 85 | 69 | 77 | 76 | 78 | 81 | 122 | 177 | 277 | 385 | 524 | 722 | 906 |
| 2016 | 17170 |  |  | 00 | 0 | 1 | 0 | 5 | 17 | 27 | 54 | 61 | 40 | 22 | 52 | 66 | 99 | 152 | 228 | 277 | 261 | 229 | 154 | 109 | 58 | 35 | 40 | 68 | 87 | 105 | 13 |

Table 2.8 (page 2 of 4)—Trawl survey size composition, by year and cm .

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 305 | 317 | 23 | 197 | 144 | 146 | 126 | 137 | 180 | 203 | 282 | 302 |  | 328 |  | 280 | 284 | 270 | 254 |  | 278 |  |  | 225 | 260 | 264 |  |  |  |
|  | 165 | 213 | 145 | 12 | 107 | 61 | 62 | 86 | 94 | 143 | 157 |  | 269 |  |  | 298 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 345 | 295 | 220 | 155 | 107 | 102 | 88 | 59 |  | 75 | 91 |  |  |  |  |  |  | 95 |  | 142 | 129 | 156 | 167 |  | 198 |  | 215 |  |  |  |
|  | 300 | 309 | 312 | 28 | 343 | 351 | 389 | 413 | 514 | 500 | 514 | 482 | 470 | 359 | 323 |  |  | 168 | 128 | 96 | 93 | 103 | 101 |  | 85 | 87 | 90 | 85 |  | 110 |
| 1986 | 557 | 448 | 402 | 34 | 332 | 220 | 194 | 138 | 126 | 136 | 163 |  | 216 | 205 | 246 | 218 |  | 269 | 258 | 275 | 288 | 299 | 226 | 25 | 251 | 175 | 171 | 120 |  |  |
|  | 280 | 207 | 235 | 201 | 172 | 186 | 22 | 210 | 293 | 327 | 330 | 330 | 322 |  | 252 |  |  |  |  | 133 |  | 146 | 140 |  | 123 | 92 | 139 | 136 |  |  |
|  | 190 | 269 | 21 | 195 | 211 | 141 | 184 | 165 | 239 | 222 | 197 | 319 | 27 | 29 | 27 | 24 | 308 | 266 | 229 | 250 | 25 |  | 220 | 214 | 22 | 194 | 199 | 166 | 207 | 165 |
|  | 70 | 33 | 107 | 109 | 13 | 115 | 125 | 101 | 115 | 115 | 139 | 176 | 165 | 176 | 183 | 176 | 200 | 253 | 23 | 260 | 247 | 234 | 326 | 29 | 219 | 222 | 197 | 290 | 186 |  |
|  |  | 80 | 113 |  | 67 |  | 67 |  |  |  |  |  |  |  | 39 |  |  |  |  |  |  | 66 |  |  | 72 | 75 | 85 | 89 |  | 78 |
|  |  |  | 261 | 195 | 173 | 143 | 118 | - | 68 | - | 61 |  |  | S3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  | 215 | 176 | 149 | 125 | 180 | 146 | 21 | 188 | 22 |  | 18 | 186 |  |  |  |  |  | 8 | 78 |  |  |  |  |  |  |  |  |  |
|  |  |  | 246 | 22 | 196 |  |  |  |  |  |  |  | 27 |  |  |  |  |  |  | 113 | 118 | 108 |  |  |  |  |  |  |  | 52 |
| 199 | 673 |  | 472 | 36 | 288 | 19 |  | 133 |  |  | 188 |  | 23 | 256 |  | 299 |  | 189 | 230 | 188 |  | 175 | 219 | 251 | 2 | 162 | , | 153 |  |  |
|  | 198 | 155 | 217 | 24 | 23 | 314 | 378 | 371 | 41 | 421 | 39 | 342 | 33 | 293 | 199 | 189 | 15 | 142 |  | 98 | 108 | 95 | 88 | 93 | 86 | 72 | 93 | 9 | 104 |  |
| 19 | 251 | 191 | 200 | 168 | 157 | 16 | 15 | 176 | 21 | 238 | 28 | 261 | 29 | 320 | 301 | 29 | 32 | 272 | 282 | 282 | 24 | 25 | 20 | 16 | 152 | 13 | 14 | 99 |  | 86 |
| 1997 | 222 | 174 | 159 | 15 | 138 | 145 | 136 | 125 | 12 | 135 | 135 | 17 | 19 | 22 |  | 17 |  | 150 | 180 | 187 | 160 | 167 | 12 | 213 | 164 | 173 | 123 | 130 | 107 |  |
|  | 537 | 346 | 260 | 22 | 166 | 147 | 134 | 101 | 119 | 117 | 13 | 12 | 169 | 119 | 115 | 133 |  | 94 | 89 | 82 | 82 | 72 | 61 | 79 | 89 | 75 | 6 | 77 | 87 | 85 |
| 19 | 22 | 197 | 191 | 24 | 290 | 308 | 382 | 486 | 509 |  | 55 |  | 39 |  |  | 23 |  | 165 | 14 | 14 | 117 | 117 | 93 | 104 | 92 | 85 | 71 | 117 | 86 |  |
| 20 | 197 | 18 | 188 | 174 | 199 | 22 | 256 | 267 | 303 | 306 | 34 |  | 355 | 32 | 391 | 34 |  | 262 |  | 239 | 25 | 194 | 20 | 183 | 159 | 15 | 149 |  |  | 90 |
| 200 | 921 | 806 | 700 | 51 | 40 | 30 | 218 | 18 | 176 | 152 | 157 | 186 | 22 | 280 | 230 | 26 | 250 | 230 | 262 | 273 | 257 | 235 | 219 | 225 | 189 | 208 | 184 | 149 | 197 |  |
|  | 520 | 381 | 400 | 31 | 295 | 25 | 28 | 25 | 40 | 35 | 453 | 393 | 38 |  | 330 | 188 | 227 | 183 | 166 | 137 | 162 | 129 |  | 89 | 109 | 121 | 125 |  | 111 |  |
|  | 316 | 216 | 319 | 240 |  | 29 | 318 | 361 | 342 | 389 | 45 |  | 461 |  | 39 | 27 |  |  | 246 | 260 | 198 | 185 | 166 | 148 | 124 |  | 138 | 116 | 96 | 70 |
|  | 317 | 310 | 335 | 313 | 32 | 25 | 242 | 21 | 208 | 188 | 181 |  | 148 |  |  | 170 | 205 | 198 |  | 182 |  | 186 | 167 | 89 | 143 | 56 | 167 | 148 | 143 |  |
|  | 197 | 19 | 207 | 23 | 288 | 252 |  |  |  | 207 | 21 |  | 20 | 16 | 19 |  |  | 12 |  | 129 |  |  |  | 101 | 98 | 100 | 117 | 84 | 18 |  |
|  | 26 | 245 | 303 | 26 | 298 | 25 | 244 | 209 | 200 | 161 | 171 | 145 | 15 | 12 | 157 | 147 | 191 | 169 | 17 | 145 | 174 | 137 | 182 | 105 | 128 | 0 | 97 | 105 | 95 |  |
|  | 124 | 114 | 92 | 93 | 76 | 60 | 73 | 77 | 74 | 68 | 82 | 76 | 85 | 79 | 8 | 60 | 75 | 74 | 82 | 68 | 72 | 59 | 54 |  | 52 | 47 | 61 | 50 | 60 | 49 |
|  | 341 | 2 | 200 | 161 | 151 | 133 | 130 | 117 | 143 | 129 | 138 | 138 | 139 | 113 | 135 | 12 | 124 | 127 | 134 | 11 | 108 | 101 | 11 | 1 | 113 | 103 | 113 | 91 |  |  |
|  | 306 | 221 | 21 | 215 | 225 | 302 | 304 | 362 | 380 | 379 | 347 | 33 | 280 | 289 | 247 | 18 | 14 | 144 | 117 | 10 | 93 | 82 | 75 | 78 | 85 | 88 | 72 | 85 | 77 |  |
| 201 | 269 | 183 | 165 | 106 | 95 | 64 | 75 | 78 | 124 | 132 | 232 | 15 | 165 | 160 | 157 | 12 | 1 | 106 | 147 | 114 | 156 | 151 | 140 | 95 | 140 | 112 | 101 | 71 | 90 | 58 |
| 2011 | 16 | 23 | 229 | 272 | 287 | 403 | 457 | 673 | 80 | 859 | 925 | 872 | 790 | 63 | 511 | 347 | 349 | 278 | 265 | 18 | 230 | 225 | 265 | 18 | 276 | 241 | 301 | 228 | 294 | 184 |
| 2012 | 279 | 309 | 190 | 158 | 98 | 81 | 61 | 46 | 63 | 59 | 85 | 81 | 130 |  | 96 | 188 | 239 | 28 | 379 | 32 | 408 | 309 | 316 | 218 | 198 | 168 | 164 | 97 | 12 | 86 |
| 2013 | 310 | 240 | 180 | 174 | 145 | 126 | 184 | 153 | 230 | 292 | 36 | 43 | 519 | 40 | 386 | 34 | 32 | 25 | 25 | 195 | 210 | 136 | 192 | 142 | 214 | 193 | 234 | 192 | 212 | 203 |
|  | 460 | 498 | 349 | 311 | 184 | 190 | 145 | 203 | 282 | 444 | 458 | 655 | 675 | 608 | 559 | 492 | 425 | 285 | 21 | 203 | 20 | 182 | 165 | 192 | 249 | 247 | 198 | 191 | 20 |  |
| 20 | 1055 | 111 | 987 | 939 | 766 | 575 | 498 | 286 | 267 | 200 | 377 | 373 | 500 | 47 | 469 | 426 | 4 | 320 | 352 | 34 | 318 | 337 | 389 | 337 | 433 | 331 | 300 | 219 | 245 | 158 |
| 2016 | 180 | 164 | 230 | 251 | 299 | 283 | 33 | 388 | 471 | 5 | 611 | 812 | 892 | 863 | 883 | 761 | 685 | 538 | 409 | 422 | 295 | 293 | 277 | 267 | 248 | 264 | 247 | 226 | 232 | 228 |

Table 2.8 (page 3 of 4)—Trawl survey size composition, by year and cm .


Table 2.8 (page 4 of 4)—Trawl survey size composition, by year and cm.

| Year | 99 | 100 | 101 | 102 | 103 | 104 | 105 | 106 | 107 | 108 | 109 | 110 | 111 | 112 | 113 | 114 | 115 | 116 | 117 | 118 | 119 | 120+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 1 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 4 | 2 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 2 | 4 | 3 | 0 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 1 | 1 | 3 | 3 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 8 | 3 | 1 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 11 | 10 | 22 | 1 | 22 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 1 | 5 | 0 | 6 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 3 | 0 | 0 | 0 | 6 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 1 | 4 | 3 | 3 | 3 | 3 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 1993 | 4 | 1 | 2 | 2 | 1 | 8 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 1 | 2 | 9 | 6 | 3 | 1 | 7 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 3 | 5 | 1 | 3 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 2 | 4 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 1 | 2 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 1 | 1 | 1 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 3 | 2 | 1 | 0 | 2 | 1 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 1 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 5 | 1 | 1 | 0 | 2 | 1 | 3 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 1 | 1 | 0 | 0 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 2 | 0 | 1 | 0 | 5 | 0 | 1 | 0 | 0 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 4 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 7 | 5 | 3 | 2 | 3 | 2 | 0 | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 2 | 3 | 2 | 8 | 1 | 2 | 1 | 2 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2008 | 2 | 4 | 3 | 7 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 1 | 2 | 1 | 1 | 0 | 1 | 2 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 2 | 2 | 1 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 2 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $0$ | 0 |
| 2014 | 0 | 2 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 2 | 1 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 1 | 3 | 2 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 2.9—Age compositions observed by the EBS shelf bottom trawl survey, 1994-2015. "Nact" = actual sample size (these get rescaled so that the average across all age compositions equals 300).

| Year | Nact | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 715 | 0.000005 | 0.088672 | 0.382767 | 0.171412 | 0.122294 | 0.118182 | 0.080798 | 0.020837 | 0.007176 | 0.004734 | 0.001400 | 0.000877 | 0.000847 |
| 1995 | 571 | 0.000011 | 0.052455 | 0.264340 | 0.420604 | 0.099263 | 0.078477 | 0.049525 | 0.016434 | 0.009218 | 0.006001 | 0.001592 | 0.000861 | 0.001218 |
| 1996 | 711 | 0.000009 | 0.056039 | 0.207937 | 0.202719 | 0.292993 | 0.135159 | 0.057671 | 0.028893 | 0.010285 | 0.004356 | 0.001891 | 0.001134 | 0.000913 |
| 1997 | 719 | 0.000000 | 0.255019 | 0.168807 | 0.183404 | 0.156924 | 0.119989 | 0.077764 | 0.022284 | 0.010205 | 0.003092 | 0.001292 | 0.000837 | 0.000384 |
| 1998 | 635 | 0.000004 | 0.076678 | 0.440954 | 0.203769 | 0.112415 | 0.056560 | 0.059534 | 0.028349 | 0.015990 | 0.004092 | 0.000810 | 0.000620 | 0.000225 |
| 1999 | 860 | 0.000003 | 0.079333 | 0.199608 | 0.302578 | 0.231813 | 0.080458 | 0.057700 | 0.027426 | 0.012165 | 0.005398 | 0.001337 | 0.001540 | 0.000641 |
| 2000 | 860 | 0.000016 | 0.234045 | 0.126952 | 0.150320 | 0.241908 | 0.147327 | 0.061410 | 0.013983 | 0.013878 | 0.005623 | 0.002827 | 0.001265 | 0.000446 |
| 2001 | 920 | 0.000010 | 0.289356 | 0.235498 | 0.193613 | 0.090842 | 0.083317 | 0.068087 | 0.026483 | 0.007913 | 0.002233 | 0.001513 | 0.000830 | 0.000305 |
| 2002 | 870 | 0.000057 | 0.079969 | 0.187981 | 0.317798 | 0.233323 | 0.071814 | 0.058708 | 0.033935 | 0.010407 | 0.003866 | 0.001138 | 0.000496 | 0.000508 |
| 2003 | 1263 | 0.000010 | 0.175004 | 0.156251 | 0.250572 | 0.209411 | 0.118871 | 0.041031 | 0.030074 | 0.013595 | 0.003643 | 0.000534 | 0.000536 | 0.000467 |
| 2004 | 995 | 0.000016 | 0.143715 | 0.165800 | 0.270793 | 0.128216 | 0.127925 | 0.090590 | 0.039844 | 0.019027 | 0.008659 | 0.002185 | 0.002584 | 0.000645 |
| 2005 | 1279 | 0.000000 | 0.183283 | 0.244428 | 0.209260 | 0.121129 | 0.065286 | 0.079441 | 0.054992 | 0.023756 | 0.010480 | 0.003625 | 0.003628 | 0.000691 |
| 2006 | 1300 | 0.000000 | 0.324413 | 0.142773 | 0.164963 | 0.121408 | 0.092865 | 0.063362 | 0.046415 | 0.028492 | 0.009924 | 0.003059 | 0.001347 | 0.000979 |
| 2007 | 1441 | 0.000000 | 0.700419 | 0.095563 | 0.067128 | 0.041366 | 0.045974 | 0.017598 | 0.014302 | 0.008393 | 0.005034 | 0.001740 | 0.001509 | 0.000972 |
| 2008 | 1213 | 0.000144 | 0.213306 | 0.445262 | 0.144892 | 0.082666 | 0.048588 | 0.032949 | 0.010242 | 0.010253 | 0.005786 | 0.002760 | 0.001363 | 0.001791 |
| 2009 | 1412 | 0.000675 | 0.454268 | 0.189424 | 0.230908 | 0.064146 | 0.028780 | 0.014629 | 0.009463 | 0.003920 | 0.002059 | 0.000825 | 0.000575 | 0.000328 |
| 2010 | 1292 | 0.000000 | 0.046504 | 0.479394 | 0.179317 | 0.203241 | 0.064417 | 0.014561 | 0.007700 | 0.002561 | 0.001271 | 0.000380 | 0.000517 | 0.000138 |
| 2011 | 1253 | 0.000030 | 0.290446 | 0.073000 | 0.388141 | 0.111090 | 0.095573 | 0.027843 | 0.006911 | 0.003347 | 0.001640 | 0.000971 | 0.000559 | 0.000 |
| 2012 | 1301 | 0.000045 | 0.365988 | 0.234280 | 0.058292 | 0.237219 | 0.061719 | 0.030655 | 0.007422 | 0.002046 | 0.001548 | 0.000467 | 0.000156 | 0.000162 |
| 2013 | 1418 | 0.000000 | 0.107227 | 0.426997 | 0.178038 | 0.108369 | 0.112914 | 0.050391 | 0.010939 | 0.003598 | 0.000810 | 0.000197 | 0.000291 | 0.000230 |
| 2014 | 1223 | 0.000048 | 0.278522 | 0.185139 | 0.236592 | 0.201442 | 0.048010 | 0.035751 | 0.010180 | 0.002252 | 0.000916 | 0.000709 | 0.000144 | 0.000295 |
| 2015 | 856 | 0.000000 | 0.068339 | 0.400726 | 0.219958 | 0.185469 | 0.088264 | 0.020851 | 0.013564 | 0.001652 | 0.000615 | 0.000264 | 0.000134 | 0.000164 |

Table 2.10—Mean size (cm) at age from age-length key applied to respective size compositions, and sample sizes. Mean lengths for samples of size zero result from application of area-specific long-term average age-length keys. These data are used in Model 11.5 only.

## Average length (cm) at age:

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1994 | 11.00 | 19.02 | 31.77 | 39.93 | 49.49 | 58.15 | 64.18 | 70.91 | 81.58 | 86.82 | 94.76 | 90.32 | 95.75 |
| 1995 | 11.00 | 17.36 | 32.35 | 43.22 | 53.13 | 62.11 | 69.90 | 74.72 | 81.81 | 85.01 | 92.00 | 91.34 | 95.51 |
| 1996 | 11.00 | 17.67 | 31.64 | 41.44 | 50.30 | 57.76 | 67.26 | 75.94 | 82.19 | 88.56 | 90.26 | 90.32 | 95.78 |
| 1997 | 0.00 | 17.22 | 31.87 | 42.02 | 51.78 | 59.84 | 64.98 | 72.43 | 79.32 | 86.65 | 91.96 | 92.31 | 93.81 |
| 1998 | 11.00 | 15.52 | 30.77 | 37.87 | 49.37 | 59.07 | 66.40 | 70.45 | 77.63 | 89.24 | 89.18 | 92.04 | 91.13 |
| 1999 | 11.00 | 15.83 | 29.66 | 40.34 | 46.26 | 56.80 | 65.53 | 71.51 | 79.82 | 82.59 | 92.01 | 90.47 | 96.05 |
| 2000 | 11.00 | 15.26 | 30.33 | 38.99 | 47.70 | 53.76 | 59.87 | 73.39 | 74.57 | 79.79 | 82.41 | 81.72 | 94.11 |
| 2001 | 11.00 | 17.89 | 31.36 | 36.70 | 48.31 | 55.35 | 62.03 | 66.01 | 76.95 | 82.27 | 78.58 | 89.10 | 92.23 |
| 2002 | 11.00 | 16.54 | 30.08 | 36.95 | 46.92 | 55.84 | 62.72 | 68.82 | 72.00 | 79.77 | 92.23 | 89.97 | 94.65 |
| 2003 | 11.00 | 18.00 | 29.81 | 40.87 | 48.29 | 56.52 | 65.36 | 70.44 | 75.30 | 81.52 | 84.94 | 84.17 | 79.15 |
| 2004 | 11.00 | 17.24 | 30.21 | 37.98 | 49.00 | 57.04 | 64.11 | 71.15 | 75.61 | 83.31 | 88.13 | 86.20 | 94.39 |
| 2005 | 0.00 | 18.59 | 26.70 | 39.16 | 48.56 | 57.03 | 64.12 | 72.34 | 78.60 | 81.76 | 88.33 | 87.14 | 93.66 |
| 2006 | 0.00 | 15.33 | 30.89 | 38.55 | 47.57 | 55.93 | 65.02 | 73.82 | 82.42 | 85.67 | 88.74 | 94.13 | 96.82 |
| 2007 | 0.00 | 15.06 | 31.03 | 41.18 | 50.61 | 59.35 | 66.64 | 74.75 | 81.58 | 84.28 | 94.22 | 87.83 | 91.35 |
| 2008 | 11.00 | 15.37 | 29.77 | 41.31 | 53.38 | 60.88 | 66.05 | 72.84 | 79.09 | 84.33 | 89.80 | 95.07 | 91.31 |
| 2009 | 11.00 | 14.14 | 31.10 | 42.51 | 51.63 | 59.79 | 65.93 | 71.91 | 75.60 | 83.69 | 90.02 | 88.82 | 90.16 |
| 2010 | 0.00 | 15.55 | 30.51 | 43.47 | 53.84 | 59.23 | 66.33 | 70.81 | 81.50 | 81.92 | 89.97 | 84.17 | 99.36 |
| 2011 | 11.00 | 18.19 | 33.07 | 43.94 | 53.86 | 62.52 | 67.87 | 73.11 | 77.22 | 85.25 | 85.84 | 83.83 | 95.47 |
| 2012 | 11.00 | 14.02 | 32.06 | 44.55 | 53.52 | 61.30 | 68.51 | 73.17 | 81.31 | 83.57 | 91.96 | 93.52 | 96.88 |
| 2013 | 0.00 | 16.27 | 28.90 | 43.88 | 50.64 | 61.98 | 66.60 | 74.93 | 77.31 | 83.71 | 88.35 | 87.26 | 95.00 |
| 2014 | 11.00 | 17.65 | 31.84 | 44.08 | 51.56 | 61.06 | 67.36 | 71.05 | 79.98 | 86.87 | 86.10 | 93.74 | 89.19 |
| 2015 | 0.00 | 20.66 | 34.62 | 43.60 | 54.73 | 61.61 | 72.31 | 74.48 | 81.88 | 80.65 | 94.37 | 97.02 | 95.99 |

Number of samples at age ( 0 indicates mean length inferred from long-term average age-length key):

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1994 | 0 | 40 | 213 | 143 | 109 | 89 | 73 | 26 | 12 | 7 | 1 | 2 | 0 |
| 1995 | 0 | 23 | 138 | 194 | 89 | 55 | 38 | 14 | 9 | 6 | 1 | 1 | 3 |
| 1996 | 0 | 34 | 143 | 138 | 183 | 101 | 65 | 37 | 5 | 2 | 0 | 1 | 2 |
| 1997 | 0 | 94 | 92 | 109 | 125 | 120 | 110 | 38 | 21 | 5 | 3 | 2 | 0 |
| 1998 | 0 | 56 | 145 | 97 | 94 | 73 | 88 | 47 | 28 | 6 | 0 | 1 | 0 |
| 1999 | 0 | 81 | 162 | 188 | 155 | 100 | 70 | 43 | 16 | 8 | 0 | 1 | 0 |
| 2000 | 0 | 111 | 100 | 130 | 202 | 175 | 82 | 21 | 19 | 6 | 6 | 1 | 0 |
| 2001 | 0 | 163 | 155 | 153 | 132 | 123 | 117 | 42 | 15 | 6 | 4 | 5 | 2 |
| 2002 | 1 | 72 | 153 | 201 | 186 | 80 | 87 | 61 | 14 | 6 | 2 | 0 | 1 |
| 2003 | 0 | 163 | 197 | 191 | 189 | 191 | 129 | 110 | 66 | 17 | 1 | 4 | 2 |
| 2004 | 0 | 141 | 133 | 197 | 128 | 150 | 129 | 59 | 32 | 17 | 4 | 4 | 0 |
| 2005 | 0 | 141 | 218 | 238 | 171 | 112 | 146 | 120 | 73 | 29 | 18 | 10 | 2 |
| 2006 | 0 | 205 | 176 | 179 | 168 | 155 | 140 | 133 | 93 | 36 | 10 | 4 | 1 |
| 2007 | 0 | 252 | 203 | 189 | 154 | 207 | 104 | 119 | 75 | 62 | 21 | 12 | 13 |
| 2008 | 0 | 141 | 262 | 244 | 188 | 134 | 97 | 45 | 45 | 28 | 13 | 8 | 8 |
| 2009 | 0 | 222 | 259 | 325 | 186 | 133 | 100 | 81 | 47 | 23 | 13 | 12 | 9 |
| 2010 | 0 | 105 | 344 | 229 | 296 | 144 | 70 | 48 | 30 | 13 | 5 | 7 | 0 |
| 2011 | 0 | 185 | 148 | 315 | 178 | 218 | 107 | 40 | 20 | 12 | 11 | 8 | 10 |
| 2012 | 0 | 162 | 289 | 129 | 284 | 161 | 150 | 55 | 30 | 20 | 11 | 3 | 4 |
| 2013 | 0 | 133 | 289 | 264 | 171 | 271 | 163 | 81 | 25 | 10 | 3 | 4 | 3 |
| 2014 | 0 | 156 | 152 | 234 | 283 | 134 | 165 | 57 | 23 | 9 | 8 | 0 | 2 |
| 2015 | 0 | 98 | 147 | 145 | 156 | 136 | 79 | 69 | 15 | 5 | 3 | 0 | 2 |

Table 2.11—Mean weight at age (kg) as estimated by the trawl survey (no weight data prior to 1998).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1998 | 0.000 | 0.031 | 0.286 | 0.563 | 1.339 | 2.404 | 3.522 | 4.273 | 5.865 | 9.246 | 9.227 | 10.228 | 9.924 |
| 1999 | 0.000 | 0.027 | 0.193 | 0.510 | 0.787 | 1.505 | 2.365 | 3.118 | 4.412 | 4.914 | 6.915 | 6.555 | 8.030 |
| 2000 | 0.000 | 0.026 | 0.249 | 0.570 | 1.107 | 1.641 | 2.339 | 4.575 | 4.821 | 6.023 | 6.700 | 6.516 | 10.401 |
| 2001 | 0.000 | 0.047 | 0.290 | 0.482 | 1.174 | 1.823 | 2.634 | 3.222 | 5.291 | 6.566 | 5.660 | 8.499 | 9.524 |
| 2002 | 0.000 | 0.039 | 0.256 | 0.488 | 1.030 | 1.776 | 2.555 | 3.418 | 3.937 | 5.425 | 8.546 | 7.907 | 9.317 |
| 2003 | 0.000 | 0.050 | 0.257 | 0.716 | 1.232 | 2.055 | 3.296 | 4.206 | 5.225 | 6.765 | 7.729 | 7.504 | 6.330 |
| 2004 | 0.000 | 0.043 | 0.266 | 0.557 | 1.270 | 2.074 | 3.026 | 4.237 | 5.157 | 7.056 | 8.461 | 7.877 | 10.590 |
| 2005 | 0.000 | 0.054 | 0.175 | 0.608 | 1.224 | 2.066 | 3.027 | 4.483 | 5.872 | 6.678 | 8.587 | 8.218 | 10.462 |
| 2006 | 0.000 | 0.028 | 0.277 | 0.570 | 1.129 | 1.912 | 3.122 | 4.720 | 6.757 | 7.663 | 8.592 | 10.412 | 11.595 |
| 2007 | 0.000 | 0.022 | 0.247 | 0.638 | 1.271 | 2.166 | 3.193 | 4.688 | 6.281 | 7.006 | 10.171 | 8.042 |  |
| 2008 | 0.000 | 0.029 | 0.247 | 0.715 | 1.640 | 2.512 | 3.272 | 4.493 | 5.868 | 7.224 | 8.857 | 10.654 | 9.452 |
| 2009 | 0.000 | 0.022 | 0.276 | 0.760 | 1.426 | 2.294 | 3.148 | 4.171 | 4.905 | 6.817 | 8.632 | 8.266 | 8.762 |
| 2010 | 0.000 | 0.030 | 0.267 | 0.838 | 1.677 | 2.284 | 3.294 | 4.071 | 6.418 | 6.526 | 8.841 | 7.123 | 12.192 |
| 2011 | 0.000 | 0.049 | 0.348 | 0.879 | 1.709 | 2.780 | 3.635 | 4.634 | 5.540 | 7.652 | 7.826 | 7.243 | 11.133 |
| 2012 | 0.000 | 0.021 | 0.304 | 0.878 | 1.590 | 2.465 | 3.531 | 4.368 | 6.143 | 6.712 | 9.143 | 9.655 | 10.827 |
| 2013 | 0.000 | 0.035 | 0.222 | 0.859 | 1.367 | 2.629 | 3.320 | 4.862 | 5.381 | 6.961 | 8.288 | 7.962 | 10.518 |
| 2014 | 0.000 | 0.046 | 0.308 | 0.881 | 1.461 | 2.522 | 3.466 | 4.116 | 6.035 | 7.880 | 7.659 | 10.079 | 8.658 |
| 2015 | 0.000 | 0.075 | 0.392 | 0.819 | 1.695 | 2.475 | 4.134 | 4.543 | 6.153 | 5.862 | 9.689 | 10.590 | 10.376 |
| Ave: | 0.000 | 0.037 | 0.271 | 0.680 | 1.335 | 2.189 | 3.171 | 4.279 | 5.676 | 6.982 | 8.523 | 8.612 | 9.973 |

Table 2.12- Relative population numbers, with standard deviations and $95 \%$ confidence intervals, as estimated by NMFS longline surveys, 19972015.

|  | Relative population number |  |  |  |
| ---: | ---: | ---: | ---: | ---: |
| Year | Estimate | Std. error | L95\% CI | U95\% CI |
| 1997 | 204,250 | 20,290 | 163,671 | 244,830 |
| 1999 | 139,390 | 14,690 | 110,009 | 168,770 |
| 2001 | 168,872 | 22,719 | 123,435 | 214,310 |
| 2003 | 203,096 | 25,236 | 152,624 | 253,568 |
| 2005 | 109,534 | 23,052 | 63,430 | 155,638 |
| 2007 | 119,105 | 16,525 | 86,055 | 152,155 |
| 2009 | 95,553 | 21,171 | 53,211 | 137,895 |
| 2011 | 143,786 | 26,141 | 91,504 | 196,069 |
| 2013 | 171,225 | 41,944 | 87,337 | 255,113 |
| 2015 | 157,996 | 30,499 | 96,998 | 218,994 |

Table 2.13—NMFS longline survey size composition, by year and cm (sample size in column 2 of top panel). No fish smaller than 37 cm have been observed in this survey.

| Year | Nact | 37 | 38 | 39 | 40 | 41 | 42 | 43 | 44 | 45 | 46 | 47 | 48 | 49 | 50 | 51 | 52 | 53 | 54 | 55 | 56 | 57 | 58 | 59 | 60 | 61 | 62 | 63 | 64 | 65 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1997 | 9671 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 4 | 1 | 8 | 12 | 18 | 38 | 47 | 79 | 109 | 136 | 149 | 201 | 198 | 187 | 177 | 234 | 221 | 275 | 311 | 393 | 411 | 526 |
| 1999 | 9860 | 4 | 0 | 1 | 5 | 5 | 4 | 35 | 38 | 92 | 121 | 194 | 221 | 289 | 269 | 188 | 259 | 224 | 199 | 270 | 253 | 280 | 270 | 308 | 302 | 472 | 405 | 449 | 452 | 530 |
| 2001 | 10313 | 0 | 0 | 3 | 1 | 2 | 3 | 7 | 10 | 14 | 24 | 34 | 36 | 80 | 89 | 94 | 153 | 169 | 212 | 247 | 313 | 294 | 305 | 371 | 386 | 398 | 445 | 426 | 525 | 542 |
| 2003 | 9984 | 1 | 0 | 1 | 0 | 5 | 11 | 16 | 30 | 51 | 82 | 120 | 175 | 268 | 280 | 390 | 384 | 460 | 481 | 460 | 542 | 449 | 378 | 423 | 467 | 501 | 454 | 539 | 461 | 466 |
| 2005 | 6506 | 0 | 0 | 0 | 3 | 3 | 0 | 10 | 10 | 10 | 15 | 12 | 46 | 38 | 68 | 107 | 125 | 177 | 159 | 191 | 171 | 193 | 205 | 214 | 237 | 247 | 300 | 265 | 331 | 307 |
| 2007 | 6964 | 0 | 0 | 0 | 1 | 2 | 4 | 9 | 19 | 36 | 56 | 68 | 99 | 102 | 104 | 158 | 132 | 136 | 127 | 129 | 169 | 164 | 236 | 227 | 273 | 257 | 277 | 352 | 266 | 284 |
| 2009 | 7135 | 1 | 2 | 1 | 6 | 11 | 23 | 40 | 73 | 108 | 193 | 255 | 240 | 291 | 315 | 273 | 251 | 182 | 223 | 185 | 191 | 222 | 198 | 211 | 257 | 295 | 274 | 256 | 267 | 335 |
| 2011 | 8137 | 3 | 0 | 1 | 4 | 5 | 10 | 46 | 42 | 78 | 101 | 104 | 148 | 127 | 116 | 111 | 143 | 155 | 155 | 202 | 212 | 292 | 356 | 382 | 477 | 527 | 565 | 491 | 487 | 465 |
| 2013 | 8003 | 0 | 0 | 0 | 0 | 2 | 1 | 1 | 3 | 4 | 14 | 23 | 75 | 79 | 111 | 142 | 166 | 180 | 196 | 169 | 163 | 169 | 187 | 220 | 229 | 287 | 359 | 463 | 469 | 536 |
| 2015 | 9453 | 0 | 0 | 0 | 0 | 1 | 2 | 1 | 5 | 12 | 26 | 30 | 90 | 93 | 149 | 205 | 238 | 324 | 412 | 482 | 539 | 596 | 626 | 590 | 526 | 428 | 425 | 386 | 365 | 350 |


| Year | 66 | 67 | 68 | 69 | 70 | 71 | 72 | 73 | 74 | 75 | 76 | 77 | 78 | 79 | 80 | 81 | 82 | 83 | 84 | 85 | 86 | 87 | 88 | 89 | 90 | 91 | 92 | 93 | 94 | 95 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 495 | 608 | 621 | 696 | 534 | 565 | 468 | 462 | 310 | 281 | 204 | 161 | 117 | 97 | 66 | 50 | 36 | 19 | 18 | 17 | 15 | 13 | 7 | 8 | 16 | 8 | 4 | 4 | 5 | 6 |
| 1999 | 441 | 422 | 367 | 361 | 337 | 305 | 258 | 224 | 161 | 203 | 121 | 99 | 82 | 53 | 47 | 33 | 31 | 18 | 35 | 11 | 8 | 3 | 28 | 9 | 29 | 2 | 3 | 2 | 5 |  |
| 2001 | 497 | 568 | 593 | 541 | 490 | 409 | 436 | 292 | 291 | 185 | 146 | 154 | 120 | 100 | 47 | 62 | 48 | 38 | 24 | 28 | 8 | 14 | 4 | 10 | 5 | 8 | 1 | 2 | 1 | 0 |
| 2003 | 345 | 297 | 231 | 167 | 175 | 148 | 160 | 148 | 74 | 39 | 116 | 19 | 29 | 25 | 27 | 9 | 19 | 12 | 8 | 3 | 7 | 8 | 5 | 4 | 2 | 4 | 2 | 1 | 2 |  |
| 2005 | 314 | 329 | 260 | 247 | 260 | 247 | 224 | 168 | 191 | 130 | 109 | 124 | 57 | 72 | 66 | 58 | 40 | 28 | 28 | 16 | 15 | 20 | 6 | 14 | 11 | 7 | 4 | 6 | 3 | 3 |
| 2007 | 300 | 279 | 220 | 248 | 227 | 139 | 204 | 161 | 172 | 140 | 94 | 121 | 107 | 98 | 90 | 72 | 68 | 74 | 46 | 49 | 54 | 35 | 39 | 41 | 22 | 20 | 16 | 19 | 17 | 20 |
| 2009 | 241 | 209 | 137 | 156 | 109 | 144 | 131 | 142 | 123 | 91 | 67 | 57 | 73 | 31 | 44 | 21 | 10 | 26 | 15 | 11 | 10 | 5 | 10 | 7 | 10 | 2 | 17 | 11 | 9 | 1 |
| 2011 | 462 | 406 | 307 | 230 | 187 | 169 | 135 | 74 | 62 | 63 | 40 | 30 | 16 | 13 | 12 | 8 | 12 | 12 | 11 | 4 | 3 | 5 | 13 | 10 | 3 | 10 | 6 | 2 | 4 | 9 |
| 2013 | 508 | 548 | 540 | 482 | 383 | 342 | 259 | 184 | 154 | 87 | 85 | 61 | 26 | 32 | 20 | 10 | 3 | 8 | 1 | 5 | 3 | 2 | 2 | 1 | 1 | 0 | 0 | 0 | 1 | 1 |
| 2015 | 306 | 244 | 256 | 235 | 248 | 194 | 156 | 142 | 133 | 117 | 85 | 55 | 68 | 54 | 46 | 50 | 29 | 21 | 20 | 16 | 13 | 14 | 7 | 9 | 6 | 5 | 6 | 2 | 4 | 1 |


| Year | 96 | 97 | 98 | 99 | 100 | 101 | 102 | 103 | 104 | 105 | 106 | 107 | 108 | 109 | 110 | 111 | 112 | 113 | 114 | 115 | 116 | 117 | 118 |  | 120+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 6 | 2 | 2 | 3 | 1 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 2 | 1 | 2 | 0 | 3 | 2 | 1 | 2 | 6 | 0 | 2 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 3 | 2 | 0 | 0 | 2 | 1 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 1 | 0 | 1 | 1 | 0 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 15 | 14 | 11 | 17 | 9 | 4 | 4 | 3 | 2 | 3 | 1 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 2 | 2 | 9 | 6 | 4 | 5 | 2 | 0 | 3 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 6 | 1 | 1 | 1 | 1 | 0 | 2 | 1 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 1 | 1 | 1 | 3 | 1 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 2.14a—Key features shared by Models 11.5 and 16.1.
Features common to both models
Time-invariant natural mortality, survey catchability, and mean length at age
Parameters governing width of length-at-age distribution (for a given mean) estimated internally
Ageing bias parameters estimated internally
Survey size composition data used in all years, including years with age composition data

Table 2.14b—Key features that differ between Models 11.5 and 16.1. "Internal" means that the parameter was estimated inside the assessment model; "external" means that the parameter was estimated outside the assessment model.

| Features that differ between models | Model 11.5 | Model 16.1 |
| :--- | :--- | :--- |
| Seasons per year | 5 (for catch), 3 (for fishery selectivity) | 1 |
| Number of initial age groups estimated | 3 | 20 |
| Natural mortality rate estimation | External (Jensen 1996) | Internal |
| Trawl survey catchability estimation | External (based on Nichol et al. 2007, 2009 assessment) | Internal |
| Mean length at age functional form | Von Bertalanffy (3 parameters, internal) | Richards (4 parameters, internal) |
| Mean length at age data | Included, but not used for estimation | Not included |
| Fishery CPUE data | Included, but not used for estimation | Not included |
| Weight at age | Internal length at age, external weight at length (seasonal) | External |
| SD of log age 0 recruitment (oR) | External (based on 2009 assessment) | Internal |
| "Fballpark" (like a weak prior on F) | Used | Not used |
| Selectivity functional form | Double normal (fishery and trawl survey) | Logistic (fishery and trawl survey) |
| Selectivity basis | Length (fishery), age (trawl survey) | Age (fishery and trawl survey) |
| Selectivity structure | Gear (3) and season (3) | None |
| Time-varying fishery selectivity | Estimated independently for 2 to 7 "blocks" of years | None |
| Time-varying survey selectivity | Annual dev sor the ascending_width parameter | None |

Table 2.15—Annual offsets to the base values of the $\alpha$ and $\beta$ weight-at-length parameters.

| Year: | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\alpha$ offset: | $2.07 \mathrm{E}-06$ | $-2.45 \mathrm{E}-06$ | $1.34 \mathrm{E}-06$ | $-2.64 \mathrm{E}-07$ | $6.80 \mathrm{E}-07$ | $2.67 \mathrm{E}-06$ | $3.22 \mathrm{E}-07$ | $1.14 \mathrm{E}-05$ |
| $\beta$ offset: | $-7.16 \mathrm{E}-02$ | $1.40 \mathrm{E}-01$ | $-4.80 \mathrm{E}-02$ | $7.43 \mathrm{E}-03$ | $-3.27 \mathrm{E}-02$ | $-8.56 \mathrm{E}-02$ | $-3.78 \mathrm{E}-03$ | $-2.77 \mathrm{E}-01$ |


| Year: | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\alpha$ offset: | $-1.02 \mathrm{E}-06$ | $-2.25 \mathrm{E}-06$ | $-2.20 \mathrm{E}-07$ | $-2.17 \mathrm{E}-06$ | $-1.26 \mathrm{E}-06$ | $1.17 \mathrm{E}-06$ | $1.64 \mathrm{E}-06$ | $2.12 \mathrm{E}-07$ |
| $\beta$ offset: | $5.76 \mathrm{E}-02$ | $1.30 \mathrm{E}-01$ | $1.61 \mathrm{E}-02$ | $1.33 \mathrm{E}-01$ | $8.09 \mathrm{E}-02$ | $-3.03 \mathrm{E}-02$ | $-6.19 \mathrm{E}-02$ | $-1.79 \mathrm{E}-02$ |


| Year: | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\alpha$ offset: | $2.62 \mathrm{E}-06$ | $4.70 \mathrm{E}-07$ | $-9.75 \mathrm{E}-07$ | $7.50 \mathrm{E}-06$ | $1.10 \mathrm{E}-06$ | $1.58 \mathrm{E}-06$ | $1.83 \mathrm{E}-06$ | $2.07 \mathrm{E}-06$ |
| $\beta$ offset: | $-7.33 \mathrm{E}-02$ | $-1.78 \mathrm{E}-02$ | $4.83 \mathrm{E}-02$ | $-2.00 \mathrm{E}-01$ | $-5.35 \mathrm{E}-02$ | $-7.12 \mathrm{E}-02$ | $-6.93 \mathrm{E}-02$ | $-6.46 \mathrm{E}-02$ |


| Year: | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\alpha$ offset: | $3.93 \mathrm{E}-06$ | $1.36 \mathrm{E}-06$ | $-3.40 \mathrm{E}-07$ | $2.02 \mathrm{E}-06$ | $-9.51 \mathrm{E}-09$ | $9.20 \mathrm{E}-07$ | $4.56 \mathrm{E}-07$ | $4.16 \mathrm{E}-06$ |
| $\beta$ offset: | $-1.23 \mathrm{E}-01$ | $-4.93 \mathrm{E}-02$ | $1.51 \mathrm{E}-02$ | $-7.27 \mathrm{E}-02$ | $4.11 \mathrm{E}-03$ | $-3.40 \mathrm{E}-02$ | $-1.27 \mathrm{E}-02$ | $-1.30 \mathrm{E}-01$ |


| Year: | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\alpha$ offset: | $-7.18 \mathrm{E}-07$ | $1.10 \mathrm{E}-06$ | $5.78 \mathrm{E}-07$ | $2.85 \mathrm{E}-06$ | $-7.54 \mathrm{E}-07$ | $-1.92 \mathrm{E}-06$ | $-2.15 \mathrm{E}-06$ |
| $\beta$ offset: | $3.71 \mathrm{E}-02$ | $-4.36 \mathrm{E}-02$ | $-2.88 \mathrm{E}-02$ | $-1.06 \mathrm{E}-01$ | $2.96 \mathrm{E}-02$ | $9.14 \mathrm{E}-02$ | $1.01 \mathrm{E}-01$ |

Table 2.16—Begin-year mean weights at age, interpolated from the survey mean weights at age shown in Table 2.11.

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1999 | 0 | 0.013 | 0.112 | 0.398 | 0.675 | 1.422 | 2.384 | 3.320 | 4.342 | 5.390 | 8.080 | 7.891 | 9.129 |
| 2000 | 0 | 0.013 | 0.138 | 0.381 | 0.808 | 1.214 | 1.922 | 3.470 | 3.970 | 5.217 | 5.807 | 6.715 | 8.478 |
| 2001 | 0 | 0.024 | 0.158 | 0.366 | 0.872 | 1.465 | 2.138 | 2.780 | 4.933 | 5.694 | 5.842 | 7.600 |  |
| 2002 | 0 | 0.020 | 0.152 | 0.389 | 0.756 | 1.475 | 2.189 | 3.026 | 3.579 | 5.358 | 7.556 | 6.784 | 8.908 |
| 2003 | 0 | 0.025 | 0.148 | 0.486 | 0.860 | 1.543 | 2.536 | 3.380 | 4.321 | 5.351 | 6.577 | 8.025 | 7.118 |
| 2004 | 0 | 0.022 | 0.158 | 0.407 | 0.993 | 1.653 | 2.541 | 3.767 | 4.681 | 6.140 | 7.613 | 7.803 | 9.047 |
| 2005 | 0 | 0.027 | 0.109 | 0.437 | 0.891 | 1.668 | 2.550 | 3.754 | 5.055 | 5.918 | 7.821 | 8.340 |  |
| 2006 | 0 | 0.014 | 0.165 | 0.372 | 0.869 | 1.568 | 2.594 | 3.874 | 5.620 | 6.768 | 7.635 | 9.500 | 9.907 |
| 2007 | 0 | 0.011 | 0.138 | 0.458 | 0.921 | 1.648 | 2.553 | 3.905 | 5.501 | 6.881 | 8.917 | 8.317 | 9.805 |
| 2008 | 0 | 0.014 | 0.135 | 0.481 | 1.139 | 1.892 | 2.719 | 3.843 | 5.278 | 6.753 | 7.932 | 10.413 | 8.747 |
| 2009 | 0 | 0.011 | 0.153 | 0.503 | 1.071 | 1.967 | 2.830 | 3.721 | 4.699 | 6.343 | 7.928 | 8.562 | 9.708 |
| 2010 | 0 | 0.015 | 0.144 | 0.557 | 1.218 | 1.855 | 2.794 | 3.610 | 5.294 | 5.715 | 7.829 | 7.878 | 10.229 |
| 2011 | 0 | 0.025 | 0.189 | 0.573 | 1.273 | 2.228 | 2.959 | 3.964 | 4.806 | 7.035 | 7.176 | 8.042 | 9.128 |
| 2012 | 0 | 0.010 | 0.176 | 0.613 | 1.234 | 2.087 | 3.155 | 4.002 | 5.389 | 6.126 | 8.398 | 8.741 | 9.035 |
| 2013 | 0 | 0.017 | 0.122 | 0.581 | 1.123 | 2.110 | 2.892 | 4.197 | 4.874 | 6.552 | 7.500 | 8.553 | 10.087 |
| 2014 | 0 | 0.023 | 0.171 | 0.551 | 1.160 | 1.945 | 3.047 | 3.718 | 5.449 | 6.630 | 7.310 | 9.183 | 8.310 |
| 2015 | 0 | 0.038 | 0.219 | 0.563 | 1.288 | 1.968 | 3.328 | 4.004 | 5.134 | 5.949 | 8.784 | 9.124 | 10.227 |
| Ave: | 0 | 0.019 | 0.152 | 0.477 | 1.009 | 1.747 | 2.655 | 3.667 | 4.878 | 6.107 | 7.571 | 8.322 | 9.121 |

Table 2.17-Input multinomial sample sizes for length composition data as specified in the stock assessment models (S1...S5 = seasons 1-5, Srv. = shelf trawl survey). Sample sizes for the NMFS longline survey length composition data in Model 16.7 are shown in the main text.

| Year | Model 11.5 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | M16.x |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Trawl fishery |  |  |  |  | Longline fishery |  |  |  |  | Pot fishery |  |  |  |  | Srv. | Fish. Srv. |  |
|  | S1 | S2 | S3 | S4 | S5 | S1 | S2 | S3 | S4 | S5 | S1 | S2 | S3 | S4 | S5 |  |  |  |
| 1977 |  |  | 10 | 13 |  |  |  |  |  |  |  |  |  |  |  |  | 2 |  |
| 1978 |  |  |  | 34 |  | 8 | 23 |  | 41 | 17 |  |  |  |  |  |  | 12 |  |
| 1979 |  |  | 16 |  | 6 | 73 | 24 | 31 | 12 | 19 |  |  |  |  |  |  | 17 |  |
| 1980 | 23 | 63 |  |  |  | 8 | 6 | 29 | 13 | 18 |  |  |  |  |  |  | 15 |  |
| 1981 |  |  | 50 |  | 15 | 7 | 5 | 26 |  | 11 |  |  |  |  |  |  | 11 |  |
| 1982 |  | 25 | 20 | 5 | 13 |  | 12 | 16 | 34 | 19 |  |  |  |  |  | 238 | 13 | 250 |
| 1983 | 19 | 70 | 27 | 11 | 149 | 82 | 86 | 47 | 53 | 58 |  |  |  |  |  | 297 | 56 | 312 |
| 1984 | 77 | 97 | 90 | 22 | 34 | 67 | 90 | 81 | 189 | 726 |  |  |  |  |  | 274 | 138 | 288 |
| 1985 | 73 | 243 | 10 | 16 | 6 | 311 | 67 | 7 | 372 | 1070 |  |  |  |  |  | 381 | 204 | 400 |
| 1986 | 84 | 199 | 78 | 45 |  | 227 | 28 | 97 | 200 | 940 |  |  | 11 | 13 |  | 347 | 178 | 365 |
| 1987 | 253 | 176 | 102 | 152 | 80 | 687 | 199 | 100 | 613 | 1257 |  |  | 5 | 15 |  | 239 | 339 | 251 |
| 1988 | 720 | 317 | 34 | 6 | 34 | 12 |  |  |  |  |  |  |  |  |  | 226 | 105 | 237 |
| 1989 | 619 |  | 67 |  | 12 |  |  |  | 37 |  |  |  |  | 9 |  | 226 | 70 | 237 |
| 1990 | 220 | 562 | 273 | 5 |  | 14 | 81 | 616 | 620 | 304 |  |  | 7 | 71 |  | 127 | 260 | 134 |
| 1991 | 426 | 1018 | 53 |  |  | 165 | 245 | 555 | 913 | 285 |  |  | 17 | 118 | 13 | 163 | 357 | 171 |
| 1992 | 106 | 729 | 55 |  |  | 392 | 723 | 1029 | 535 |  | 6 | 10 | 244 | 115 |  | 217 | 369 | 228 |
| 1993 | 165 | 902 |  |  |  | 487 | 718 | 83 |  |  |  | 91 | 36 |  |  | 235 | 232 | 247 |
| 1994 | 109 | 1343 | 82 |  |  | 591 | 853 | 180 | 438 |  |  | 203 | 105 | 68 |  | 314 | 372 | 330 |
| 1995 | 88 | 890 |  | 8 |  | 600 | 770 | 100 | 492 | 216 | 7 | 268 | 338 | 95 | 61 | 208 | 368 | 218 |
| 1996 | 66 | 1287 | 95 | 40 | 14 | 738 | 737 | 103 | 742 | 37 |  | 433 | 456 | 177 | 20 | 211 | 463 | 222 |
| 1997 | 126 | 1098 | 29 |  |  | 751 | 796 | 265 | 829 | 708 |  | 269 | 343 | 126 | 23 | 207 | 502 | 218 |
| 1998 | 75 | 939 | 31 | 38 | 5 | 644 | 574 | 111 | 987 | 858 |  | 211 | 240 | 50 |  | 216 | 446 | 227 |
| 1999 | 238 | 565 | 12 | 15 |  | 740 | 789 | 239 | 976 | 246 |  | 118 | 292 | 83 |  | 264 | 404 | 277 |
| 2000 | 198 | 527 | 36 |  |  | 684 | 395 | 130 | 1265 | 830 | 304 | 168 |  |  |  | 283 | 425 | 298 |
| 2001 | 74 | 305 | 41 | 52 |  | 557 | 670 | 327 | 1420 | 855 | 27 | 291 | 19 | 138 | 9 | 446 | 448 | 469 |
| 2002 | 161 | 316 | 90 | 121 |  | 980 | 549 | 210 | 1714 | 699 | 80 | 162 | 16 | 125 | 16 | 276 | 491 | 290 |
| 2003 | 121 | 414 | 100 | 149 |  | 1277 | 802 | 322 | 1896 | 1006 | 264 | 13 |  | 135 | 39 | 279 | 612 | 293 |
| 2004 | 146 | 255 | 134 | 85 |  | 1043 | 668 | 278 | 1662 | 832 | 158 | 34 | 14 | 116 | 18 | 244 | 497 | 257 |
| 2005 | 205 | 272 | 112 |  |  | 1216 | 299 | 315 | 1659 | 818 | 143 | 22 |  | 136 |  | 255 | 487 | 268 |
| 2006 | 278 | 157 | 82 | 13 |  | 960 | 294 | 151 | 1659 | 81 | 200 | 49 | 11 | 138 | 29 | 274 | 384 | 288 |
| 2007 | 188 | 211 | 145 |  |  | 882 | 75 |  | 1218 | 56 | 210 | 23 |  | 100 |  | 289 | 299 | 304 |
| 2008 | 165 | 92 | 32 | 21 |  | 805 | 190 | 207 | 1551 | 462 | 120 | 26 |  | 123 |  | 293 | 355 | 308 |
| 2009 | 85 | 57 | 27 | 66 |  | 721 | 116 | 163 | 1483 | 431 | 122 | 21 |  | 52 | 15 | 377 | 315 | 396 |
| 2010 | 163 | 37 | 17 | 58 |  | 775 | 75 | 148 | 960 | 434 | 142 |  |  | 114 | 37 | 171 | 277 | 179 |
| 2011 | 242 | 138 | 36 | 84 |  | 492 | 667 | 419 | 1019 | 441 | 164 |  |  | 168 |  | 468 | 363 | 492 |
| 2012 | 328 | 125 | 45 | 27 |  | 573 | 542 | 556 | 1021 | 580 | 203 | 28 |  | 238 |  | 295 | 400 | 310 |
| 2013 | 465 | 168 | 30 | 123 |  | 890 | 507 | 421 | 1092 | 692 | 127 | 9 |  | 195 | 78 | 422 | 503 | 443 |
| 2014 | 435 | 348 | 58 | 51 |  | 760 | 689 | 853 | 996 | 768 | 146 | 21 |  | 114 | 61 | 405 | 497 | 426 |
| 2015 | 241 | 327 | 49 | 40 |  | 673 | 654 | 550 | 1171 | 623 | 193 | 66 | 16 | 221 | 47 | 436 | 456 | 458 |
| 2016 | 369 | 193 | 30 | 10 |  | 658 | 495 | 550 | 228 |  | 169 | 36 |  |  |  | 388 | 257 | 407 |

Table 2.18—Objective function components and parameter counts.

| Component | M11.5 | M16.1 | M16.6 | M16.7 | M16.8 | M16.9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Catch | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Equilibrium catch | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Survey abundance index | -3.95 | -23.52 | -25.21 | -34.29 | -41.36 | -18.72 |
| Size composition | 5242.98 | 1378.92 | 1372.94 | 1636.85 | 1218.48 | 1187.99 |
| Age composition | 153.94 | 243.81 | 241.40 | 252.32 | 127.95 | 238.82 |
| Recruitment | 21.18 | 3.38 | 4.25 | 4.78 | 0.72 | 0.89 |
| "Softbounds" | 0.02 | 0.01 | 0.01 | 0.01 | 0.00 | 0.00 |
| Deviations | 20.71 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 |
| "F ballpark" | 0.00 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| TOTAL | 5434.88 | 1602.60 | 1593.39 | 1859.67 | 1305.79 | 1408.97 |


| Size composition | M11.5 | M16.1 | M16.6 | M16.7 | M16.8 | M16.9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Jan-Apr trawl fishery | 1134.15 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| May-Jul trawl fishery | 213.12 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| Aug-Dec trawl fishery | 257.56 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| Jan-Apr longl. fishery | 795.94 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| May-Jul longl. fishery | 264.57 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| Aug-Dec longl. fishery | 1129.12 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| Jan-Apr pot fishery | 149.33 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| May-Jul pot fishery | 70.47 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| Aug-Dec pot fishery | 262.89 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| Fishery | $\mathrm{n} / \mathrm{a}$ | 366.40 | 364.60 | 365.10 | 338.72 | 192.69 |
| Trawl survey | 965.82 | 1012.52 | 1008.34 | 1020.80 | 879.76 | 995.30 |
| NMFS LL survey | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | 250.94 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |


| Survey abundance index | M11.5 | M16.1 | M16.6 | M16.7 | M16.8 | M16.9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Trawl survey | -3.95 | -23.52 | -25.21 | -25.15 | -41.36 | -18.72 |
| NMFS LL survey | n/a | n/a | n/a | -9.15 | n/a | n/a |


| Parameter type | M11.5 | M16.1 | M16.6 | M16.7 | M16.8 | M16.9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| True parameters: | 115 | 18 | 18 | 21 | 17 | 18 |
| Constrained dev s: | 75 | 59 | 59 | 59 | 92 | 139 |
| Total: | 190 | 77 | 77 | 80 | 109 | 157 |

Table 2.19—Log-scale standard errors of the data (oave), root mean squared errors (RMSE), mean normalized residuals (MNR), standard deviations of normalized residuals (SDNR), and observed:expected correlations (Corr.) for fishery CPUE and survey relative abundance time series. Fishery CPUE data are not used in fitting Model 11.5 and are not included at all in Models 16.x; fishery CPUE results are shown for comparison only.

| Model | Fleet | oave | RMSE | MNR | SDNR | Corr. |
| :---: | :--- | ---: | ---: | ---: | ---: | ---: |
| 11.5 | Jan-Apr trawl fishery | 0.08 | 0.48 | 0.57 | 4.02 | 0.17 |
| 11.5 | May-Jul trawl fishery | 0.25 | 0.42 | -0.16 | 1.70 | 0.19 |
| 11.5 | Aug-Dec trawl fishery | 0.57 | 0.69 | 0.17 | 2.31 | 0.12 |
| 11.5 | Jan-Apr longline fishery | 0.08 | 0.39 | 0.23 | 4.68 | -0.18 |
| 11.5 | May-Jul longline fishery | 0.20 | 0.29 | 0.35 | 2.61 | 0.46 |
| 11.5 | Aug-Dec longline fishery | 0.12 | 0.27 | 0.12 | 4.12 | 0.30 |
| 11.5 | Jan-Apr pot fishery | 0.12 | 0.35 | 0.18 | 2.05 | 0.23 |
| 11.5 | May-Jul pot fishery | 0.14 | 0.21 | 0.04 | 1.47 | 0.23 |
| 11.5 | Aug-Dec pot fishery | 0.32 | 0.39 | 0.01 | 2.06 | 0.14 |
| 11.5 | Shelf trawl survey | 0.11 | 0.23 | 1.04 | 1.82 | 0.78 |
| 16.1 | Shelf trawl survey | 0.11 | 0.19 | 0.07 | 1.79 | 0.79 |
| 16.6 | Shelf trawl survey | 0.11 | 0.19 | 0.10 | 1.76 | 0.79 |
| 16.7 | Shelf trawl survey | 0.11 | 0.18 | 0.11 | 1.76 | 0.80 |
| 16.8 | Shelf trawl survey | 0.11 | 0.16 | 0.11 | 1.47 | 0.85 |
| 16.9 | Shelf trawl survey | 0.11 | 0.20 | 0.08 | 1.86 | 0.78 |
| 16.7 | NMFS longline survey | 0.16 | 0.25 | -0.27 | 1.42 | 0.60 |

Table 2.20—Ratios of effective sample size to input sample size for each fishery and survey size composition time series. Nrec = number of records, Ninp = input sample size, Neff = effective sample size, $\mathrm{A}(\cdot)=$ arithmetic mean, $\mathrm{H}(\cdot)=$ harmonic mean. Note that the arithmetic mean input sample size for the trawl survey size composition data in Model 11.5 (285) differs from that of the other models (300).

| Model | Fleet | Nrec | A(Ninp) | Ratios |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | A(Neff/Ninp) | A(Neff)/A(Ninp) | H(Neff)/A(Ninp) |
| 11.5 | Jan-Apr trawl fish. | 70 | 312 | 4.97 | 3.02 | 1.71 |
| 11.5 | May-Jul trawl fish. | 36 | 61 | 9.24 | 7.49 | 3.34 |
| 11.5 | Aug-Dec trawl fish. | 39 | 43 | 12.67 | 6.07 | 3.36 |
| 11.5 | Jan-Apr longl. fish. | 74 | 474 | 8.36 | 3.99 | 1.20 |
| 11.5 | May-Jul longl. fish. | 36 | 261 | 9.35 | 5.00 | 2.91 |
| 11.5 | Aug-Dec longl. fish. | 69 | 674 | 6.34 | 3.11 | 0.91 |
| 11.5 | Jan-Apr pot fish. | 42 | 128 | 13.89 | 10.10 | 3.87 |
| 11.5 | May-Jul pot fish. | 17 | 128 | 17.97 | 7.81 | 1.86 |
| 11.5 | Aug-Dec pot fish. | 41 | 86 | 10.12 | 7.41 | 2.89 |
| 16.1 | Fishery | 40 | 300 | 8.68 | 5.83 | 1.88 |
| 16.6 | Fishery | 40 | 300 | 8.73 | 5.87 | 1.90 |
| 16.7 | Fishery | 40 | 300 | 10.25 | 8.47 | 1.89 |
| 16.8 | Fishery | 40 | 300 | 10.12 | 8.24 | 1.91 |
| 16.9 | Fishery | 40 | 300 | 16.43 | 8.82 | 3.48 |
| 11.5 | Trawl survey | 35 | 285 | 1.95 | 1.65 | 1.02 |
| 16.1 | Trawl survey | 35 | 300 | 1.82 | 1.56 | 1.00 |
| 16.6 | Trawl survey | 35 | 300 | 1.83 | 1.56 | 1.01 |
| 16.7 | Trawl survey | 35 | 300 | 1.84 | 1.57 | 1.00 |
| 16.8 | Trawl survey | 35 | 300 | 2.26 | 1.90 | 1.15 |
| 16.9 | Trawl survey | 35 | 300 | 1.87 | 1.59 | 1.03 |
| 16.7 | NMFS LL survey | 10 | 300 | 1.79 | 1.59 | 1.01 |

Table 2.21-Input sample size, effective sample size, and ratio thereof for each year of age composition data from the bottom trawl survey. Last two rows show arithmetic and harmonic means. Color scale extends from red (low) to green (high) in each row.

| Year | Input N | Effective N |  |  |  |  |  | Ratio |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | M11.5 | M16.1 | M16.6 | M16.7 | M16.8 | M16.9 | M11.5 | M16.1 | M16.6 | M16.7 | M16.8 | M16.9 |
| 1994 | 204 | 428 | 186 | 209 | 233 | 237 | 163 | 2.10 | 0.91 | 0.49 | 1.26 | 1.13 | 0.70 |
| 1995 | 163 | 37 | 29 | 29 | 24 | 54 | 31 | 0.23 | 0.18 | 0.79 | 0.82 | 1.85 | 1.29 |
| 1996 | 203 | 365 | 68 | 79 | 60 | 598 | 83 | 1.80 | 0.34 | 0.22 | 0.87 | 7.55 | 1.39 |
| 1997 | 205 | 154 | 51 | 54 | 62 | 194 | 45 | 0.75 | 0.25 | 0.35 | 1.23 | 3.61 | 0.72 |
| 1998 | 181 | 1245 | 93 | 83 | 103 | 1229 | 97 | 6.88 | 0.51 | 0.07 | 1.11 | 14.77 | 0.94 |
| 1999 | 246 | 124 | 61 | 55 | 50 | 94 | 68 | 0.50 | 0.25 | 0.45 | 0.83 | 1.70 | 1.35 |
| 2000 | 246 | 114 | 62 | 53 | 42 | 60 | 82 | 0.46 | 0.25 | 0.46 | 0.67 | 1.15 | 1.96 |
| 2001 | 263 | 103 | 37 | 39 | 38 | 74 | 37 | 0.39 | 0.14 | 0.37 | 1.03 | 1.91 | 0.97 |
| 2002 | 248 | 88 | 40 | 38 | 39 | 96 | 40 | 0.35 | 0.16 | 0.43 | 0.98 | 2.53 | 1.04 |
| 2003 | 361 | 280 | 824 | 986 | 935 | 224 | 707 | 0.78 | 2.28 | 3.52 | 1.13 | 0.23 | 0.76 |
| 2004 | 284 | 31 | 34 | 34 | 34 | 50 | 35 | 0.11 | 0.12 | 1.11 | 0.97 | 1.46 | 1.04 |
| 2005 | 365 | 365 | 183 | 182 | 170 | 321 | 169 | 1.00 | 0.50 | 0.50 | 0.93 | 1.76 | 0.99 |
| 2006 | 371 | 141 | 51 | 52 | 57 | 404 | 55 | 0.38 | 0.14 | 0.37 | 1.11 | 7.82 | 0.97 |
| 2007 | 412 | 58 | 11 | 11 | 10 | 74 | 12 | 0.14 | 0.03 | 0.19 | 0.93 | 6.72 | 1.17 |
| 2008 | 346 | 261 | 135 | 136 | 153 | 838 | 127 | 0.75 | 0.39 | 0.52 | 1.13 | 6.18 | 0.83 |
| 2009 | 403 | 96 | 162 | 139 | 130 | 395 | 165 | 0.24 | 0.40 | 1.46 | 0.81 | 2.84 | 1.27 |
| 2010 | 369 | 101 | 210 | 260 | 241 | 171 | 285 | 0.27 | 0.57 | 2.57 | 1.15 | 0.66 | 1.18 |
| 2011 | 358 | 144 | 121 | 117 | 110 | 106 | 110 | 0.40 | 0.34 | 0.81 | 0.90 | 0.90 | 1.00 |
| 2012 | 372 | 92 | 76 | 78 | 69 | 97 | 91 | 0.25 | 0.20 | 0.85 | 0.91 | 1.24 | 1.32 |
| 2013 | 405 | 113 | 127 | 125 | 112 | 137 | 135 | 0.28 | 0.31 | 1.10 | 0.88 | 1.10 | 1.21 |
| 2014 | 349 | 416 | 290 | 311 | 370 | 323 | 259 | 1.19 | 0.83 | 0.75 | 1.27 | 1.04 | 0.70 |
| 2015 | 244 | 312 | 201 | 206 | 222 | 415 | 202 | 1.28 | 0.82 | 0.66 | 1.11 | 2.01 | 0.91 |
| Mean | 300 | 230 | 139 | 149 | 148 | 282 | 136 | 0.93 | 0.45 | 0.82 | 1.00 | 3.19 | 1.08 |
| Harm. | 277 | 112 | 59 | 59 | 56 | 132 | 62 | 0.38 | 0.19 | 0.40 | 0.98 | 1.33 | 1.01 |

Table 2.22a-Biological parameters, ageing bias, recruitment parameters (except annual devs), initial fishing mortality, and log catchability estimated by at least one of the stock assessment models ("-" in the SD column means that the parameter value was assumed rather than estimated (specifically, natural mortality, $\sigma$ (recruitment), and $\ln ($ trawl survey catchability), all in Model 11.5), and " $\mathrm{n} / \mathrm{a}$ " means that the parameter was not used in the respective model).

| Parameter | Model 11.5 |  | Model 16.1 |  | Model 16.6 |  | Model 16.7 |  | Model 16.8 |  | Model 16.9 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Est. | SD | Est. | SD | Est. | SD | Est. | SD | Est. | SD | Est. | SD |
| Natural mortality | 0.340 |  | 0.378 | 0.012 | 0.363 | 0.013 | 0.344 | 0.012 | 0.375 | 0.012 | 0.376 | 0.012 |
| Length at age 1 (cm) | 14.352 | 0.106 | 16.399 | 0.088 | 16.401 | 0.088 | 16.449 | 0.088 | 16.360 | 0.088 | 16.381 | 0.088 |
| Asymptotic length (cm) | 92.747 | 0.494 | 98.412 | 1.826 | 99.387 | 1.901 | 101.132 | 1.814 | 100.396 | 1.984 | 97.914 | 1.778 |
| Brody growth coefficient | 0.239 | 0.002 | 0.200 | 0.012 | 0.197 | 0.012 | 0.200 | 0.011 | 0.195 | 0.012 | 0.195 | 0.012 |
| Richards growth coefficient | n/a | n/a | 1.054 | 0.048 | 1.050 | 0.048 | 1.014 | 0.043 | 1.050 | 0.048 | 1.077 | 0.050 |
| SD of length at age 1 (cm) | 3.605 | 0.067 | 3.424 | 0.058 | 3.425 | 0.058 | 3.479 | 0.057 | 3.422 | 0.058 | 3.403 | 0.058 |
| SD of length at age 20 (cm) | 9.616 | 0.154 | 9.663 | 0.275 | 9.717 | 0.282 | 8.851 | 0.219 | 9.551 | 0.296 | 9.984 | 0.289 |
| Ageing bias at age 1 (years) | 0.336 | 0.013 | 0.325 | 0.012 | 0.321 | 0.013 | 0.308 | 0.014 | 0.323 | 0.013 | 0.328 | 0.012 |
| Ageing bias at age 20 (years) | 0.322 | 0.145 | 0.323 | 0.153 | 0.351 | 0.154 | 0.527 | 0.154 | 0.351 | 0.160 | 0.313 | 0.150 |
| $\ln$ (mean post-1976 recruitment) | 13.171 | 0.019 | 13.620 | 0.104 | 13.220 | 0.104 | 13.011 | 0.094 | 13.555 | 0.094 | 13.593 | 0.103 |
| $\sigma$ (recruitment) | 0.570 |  | 0.631 | 0.066 | 0.638 | 0.066 | 0.638 | 0.066 | 0.602 | 0.065 | 0.610 | 0.061 |
| $\ln$ (pre-1977 recruitment offset) | -1.137 | 0.130 | -1.047 | 0.226 | -1.099 | 0.216 | -1.172 | 0.198 | -1.098 | 0.220 | -0.748 | 0.203 |
| Initial F (Jan-Apr trawl fishery) | 0.664 | 0.141 | n/a | n/a | n/a | n/a | n/a | n/a | n/a | n/a | n/a | n/a |
| Initial F (fishery) | n/a | n/a | 0.127 | 0.045 | 0.155 | 0.056 | 0.188 | 0.071 | 0.149 | 0.056 | 0.073 | 0.021 |
| $\ln$ (trawl survey catchability) | -0.261 |  | -0.487 | 0.062 | -0.133 | 0.065 | 0.033 | 0.056 | -0.408 | 0.056 | -0.496 | 0.061 |
| $\ln ($ NMFS LL survey catchability) | n/a | n/a | n/a | n/a | n/a | n/a | 0.410 | 0.071 | n/a | n/a | n/a | n/a |

Table 2.22b—Initial age composition devs estimated by the stock assessment models.

| Parameter | Model 11.5 |  | Model 16.1 |  | Model 16.6 |  | Model 16.7 |  | Model 16.8 |  | Model 16.9 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Est. | SD | Est. | SD | Est. | SD | Est. | SD | Est. | SD | Est. | SD |
| itial age $20 \ln$ (abundance) dev | n/a | n/a | -0.006 | 0.630 | -0.005 | 0.636 | -0.004 | 0.636 | -0.004 | 0.600 | -0.008 | 0.608 |
| Initial age $19 \ln$ (abundance) dev | n/a | n/a | -0.004 | 0.630 | -0.003 | 0.637 | -0.003 | 0.637 | -0.003 | 0.601 | -0.005 | 0.609 |
| Initial age $18 \ln$ (abundance) dev | n/a | n/a | -0.006 | 0.629 | -0.006 | 0.636 | -0.005 | 0.636 | -0.005 | 0.600 | -0.007 | 0.608 |
| Initial age $17 \ln$ (abundance) dev | a | n/a | -0.010 | 0.628 | -0.009 | 0.635 | -0.009 | 0.635 | -0.008 | 0.599 | -0.012 | 0.607 |
| Initial age $16 \ln$ (abundance) dev | /a | n/a | -0.016 | 0.626 | -0.016 | 0.633 | -0.014 | 0.633 | -0.013 | 0.598 | -0.018 | 0.605 |
| Initial age $15 \ln$ (abundance) dev | n/a | n | -0.027 | 0.623 | -0.025 | 0.630 | -0.024 | 0.630 | -0.022 | 0.595 | -0.027 | 0.602 |
| Initial age $14 \ln$ (abundance) dev | n/a | n | -0.043 | 0.619 | -0.041 | 0.626 | -0.039 | 0.626 | -0.036 | 0.591 | -0.042 | 0.598 |
| Initial age $13 \ln$ (abundance) dev | a | n/a | -0.069 | 0.612 | -0.066 | 0.619 | -0.063 | 0.620 | -0.060 | 0.586 | -0.063 | 0.592 |
| Initial age $12 \ln$ (abundance) dev | n/a | $\mathrm{n} /$ | -0.109 | 0.602 | -0.103 | 0.610 | -0.101 | 0.611 | -0.096 | 0.577 | -0.093 | 0.585 |
| Initial age $11 \ln$ (abundance) dev | n/a | $\mathrm{n} /$ | -0.163 | 0.589 | -0.157 | 0.597 | -0.155 | 0.600 | -0.147 | 0.566 | -0.134 | 0.575 |
| Initial age $10 \ln$ (abundance) dev | n/a | n/a | -0.239 | 0.574 | -0.232 | 0.582 | -0.232 | 0.585 | -0.220 | 0.552 | -0.188 | 0.563 |
| Initial age $9 \ln$ (abundance) dev | a | n/a | -0.332 | 0.556 | -0.328 | 0.563 | -0.332 | 0.567 | -0.312 | 0.535 | -0.253 | 0.549 |
| Initial age $8 \ln$ (abundance) dev | n/a | n/a | -0.437 | 0.537 | -0.442 | 0.543 | -0.451 | 0.547 | -0.418 | 0.517 | -0.323 | 0.536 |
| Initial age $7 \ln$ (abundance) dev | n/a | n/a | -0.542 | 0.519 | -0.560 | 0.523 | -0.578 | 0.525 | -0.526 | 0.499 | -0.386 | 0.524 |
| Initial age $6 \ln$ (abundance) dev | a | /a | -0.616 | 0.504 | -0.650 | 0.505 | -0.680 | 0.503 | -0.609 | 0.483 | -0.412 | 0.517 |
| Initial age $5 \ln$ (abundance) dev | n/a | n/a | -0.580 | 0.497 | -0.628 | 0.495 | -0.678 | 0.488 | -0.588 | 0.473 | -0.345 | 0.518 |
| Initial age $4 \ln$ (abundance) dev | n/a | n/a | -0.194 | 0.481 | -0.246 | 0.478 | -0.309 | 0.471 | -0.227 | 0.459 | -0.064 | 0.523 |
| Initial age $3 \ln$ (abundance) dev | 1.305 | 0.187 | -0.068 | 0.471 | -0.092 | 0.463 | -0.021 | 0.435 | -0.075 | 0.447 | 0.052 | 0.481 |
| Initial age $2 \ln$ (abundance) dev | -0.699 | 0.418 | -0.118 | 0.521 | -0.153 | 0.516 | -0.220 | 0.498 | -0.137 | 0.493 | -0.253 | 0.538 |
| Initial age $1 \ln$ (abundance) dev | 1.399 | 0.210 | 0.771 | 0.524 | 0.744 | 0.513 | 0.790 | 0.451 | 0.725 | 0.498 | 0.133 | 0.543 |

Table 2.22c—Annual log-scale recruitment devs estimated by the stock assessment models. Color scale extends from red (low) to green (high) in each column.

|  | Model 11.5 |  | Model 16.1 |  | Model 16.6 |  | Model 16.7 |  | Model 16.8 |  | Model 16.9 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | st. | SD | Est. | SD | Est. | SD | Est. | SD | Est. | SD | Est. | SD |
| 19 | 1.363 | 0.108 | 02 | 0.21 | 0.935 | 0.212 | 0.723 | 0.200 | 0.9 | 0.20 | 1.081 | 0.214 |
| 1978 | 0.483 | 0.207 | 0.505 | . 26 | 0.483 | 0.25 | 0.446 | 0.2 | 0.468 | 0.2 | 0.587 | 0.286 |
| 1979 | 0.7 | 0.10 | 0.496 | 0.147 | . 81 | 0.1 | 0.442 | 0.1 | 0.469 | 0.1 | 0.600 | 0.156 |
| 19 | -0.329 | 0.130 | -0. | 0.1 | -0.28 | 0.1 | -0.24 | 0.1 | -0.36 | 0.1 | 39 | 0.162 |
| 1981 | -0. | 0.141 | -0.883 | 0.142 | -0.883 | 0.142 | -0.85 | 0.1 | -0.725 | 0.1 | -0.913 | 0.148 |
| 82 | . 981 | 0.041 | 0.784 | 0.051 | . 82 | 0.051 | 0.740 | 0.050 | 0.740 | 0.05 | . 845 | 0.054 |
| 1983 | -0.509 | 0.110 | -0.584 | 0.12 | -0.580 | 0.125 | -0.529 | 0.11 | -0.338 | 0.12 | -0.582 | 0.133 |
| 1984 | 0.777 | 0.045 | 0.758 | 0.051 | 0.766 | 0.050 | 0.74 | 0.049 | 0.690 | 0.06 | 0.785 | 0.054 |
| 1985 | -0.053 | . 07 | -0.222 | 0.090 | -0.202 | 0.090 | -0.145 | 0.083 | -0.118 | 0.09 | -0.138 | 0.093 |
| 1986 | -0.749 | 09 | -0.64 | 0.10 | -0.61 | 0.102 | -0.55 | 0.095 | -0.702 | 0.1 | -0.575 |  |
| 1987 | -1.094 | 0.105 | -1.528 | . 18 | -1. | 0.17 | -1.371 | 0.16 | -1.56 | 0.2 | -1.6 | 0.201 |
| 1988 | -0.202 | . 05 | -0.50 | 0.0 | -0. | 0.09 | -0.4 | 0.0 | -0.1 | 0.0 | -0.5 | 0.109 |
| 1989 | 0.534 | 0.040 | 0.534 | 0.05 | 0.530 | 0. | 0.496 | 0.0 | 0.442 | 0.0 | 0.544 | 0.063 |
| 1990 | 0.340 | 0.044 | 0.342 | 0.0 | 0.33 | 0.065 | 0.3 | 0.0 | 0.4 | 0.0 | 23 | 0.071 |
| 1991 | -0.282 | 0.06 | -0.053 | 0.07 | -0.078 | 0.078 | -0.025 | 0.072 | -0.176 | 0.09 | -0.081 | 0.08 |
| 1992 | 0.623 | 0.032 | 0.7 | 0.04 | 0. | 0.04 | 0.73 | 0.03 | 0.6 | 0.04 | 0.824 | . 04 |
| 1993 | -0.409 | 0.05 | -0.149 | 0.06 | -0. | 0.06 | -0.288 | 0.06 | -0.08 | 0.07 | -0. | . 072 |
| 1994 | -0.33 | 0.05 | -0.258 | 0.07 | -0. | 0.06 | -0.253 | 0.06 | -0.268 | 0.07 | -0.24 | 0.075 |
| 1995 | -0.256 | 0.05 | -0.340 | 0.07 | -0.439 | 0.07 | -0.36 | 0.06 | -0.20 | 0.08 | -0.370 | 0.083 |
| 1996 | 0.64 | 03 | 0.665 | . 04 | 0.57 | 0.04 | 0.5 | 0.03 | 0.6 | 0.0 | 0.72 | 0.043 |
| 1997 | -0.243 | . 05 | -0. | 0.0 | -0.180 | 0.06 | -0.2 | 0.0 | -0.03 | 0.0 | -0. | . 0.0 |
| 1998 | -0.273 | . 04 | -0.222 | 0.06 | -0.25 | 0.06 | -0.16 | 0.0 | -0.10 | 0.0 | -0.28 | 0.073 |
| 1999 | 376 | 03 | 0.483 | 0.0 | 0.482 | 0.0 | 0.519 | 0.0 |  | 0.0 |  |  |
| 2000 | -0.104 | 0.03 | 0.1 |  | 0.213 | 0.0 | 0.2 | 0.04 | 0.0 | 0.06 | 0.159 | 0.0 |
| 2001 | -0.877 | 0.05 | -0. | 0.068 | -0.601 | 0.06 | -0.56 | 0.06 | -0.438 | 0.071 | -0.695 | 0.07 |
| 2002 | -0.345 | 0.038 | -0.35 | 0.05 | -0. | 0.052 | -0.266 | 0.048 | -0.393 | 0.06 | -0.392 | . 05 |
| 2003 | -0.5 | 04 | -0. | 0.05 | -0. | 0.05 | -0.47 | 0.05 | -0.388 | 0.06 | -0. | 0.058 |
| 2004 | -0. | 0.04 | -0. | 0.06 | -0 | 0.06 | -0.6 | 0.05 | -0.7 | 0.0 | -0.788 | 0.063 |
| 2005 | -0.60 | 0.046 | -0.410 | 0.05 | -0.3 | 0.0 | -0.30 | 0.0 | -0.5 | 0.0 | -0.44 | 0.05 |
| 2006 | 0.650 | 0.02 | 0.75 | 0.03 | 0.8 | 0.03 | 0.78 | 0.0 | 0.60 | 0.0 | 0.76 | 0.039 |
| 2007 | -0.533 | 0.05 | -0.0 | 0.0 | -0.00 | 0.0 | -0.02 | 0.0 | -0.109 | 0.0 | -0.06 | 0.060 |
| 2008 | 1.101 | 0.028 | 1.106 | 0.0 | 1.150 | 0. | 1.116 | 0.030 | 1.0 | 0.0 |  | 0.040 |
| 2009 | -0.649 | 0.09 | -0.9 | 0.1 | -0.8 | 0.1 | -0.86 | 0.1 | -0.8 | 0.1 | -1. | 0.1 |
| 2010 | 0.60 | 0.04 | 0.631 | 0.05 | 0.644 | 0.0 | 0.614 | 0.0 | 0.4 | 0.0 | 0.630 | 0 |
| 2011 | 0.99 | 0.04 | . 6 | 0.05 | 1.038 | 0.04 | 0.957 | 0.0 | 0.8 | 0.05 | 1.106 | . 0 |
| 2012 | 0.223 | 0.07 | 0.17 | 0.07 | 0.16 | 0.07 | 0.123 | 0.07 | 0.21 | 0.08 | 0.21 | . 07 |
| 2013 | 1.117 | 0.06 | 0.998 | 0.06 | 0.982 | 0.06 | 0.927 | 0.059 | 0.85 | 0.07 | 1.02 | 0.061 |
| 2014 | -0.884 | 0.15 | -0.966 | 0.14 | -0.983 | 0.14 | -1.02 | 0.14 | -1.001 | 0.1 | -0.94 | . 1 |
| 2015 | -0.622 | 0.20 | -0.804 | 0.19 | -0.820 | 0.198 | -0.836 | 0.19 | -0.786 | 0.28 | -0.789 | 0.19 |

Table 2.22d—Fishery selectivity parameters estimated by Model 11.5.

| Parameter | Est. | SD |
| :--- | ---: | ---: |
| P3_May-Jul_Trawl | 5.635 | 0.103 |
| P2_Jan-Apr_Longline | -6.912 | 14.211 |
| P4_Jan-Apr_Longline | 5.032 | 0.137 |
| P3_May-Jul_Longline | 5.082 | 0.043 |
| P2_Aug-Dec_Longline | -2.095 | 0.267 |
| P4_Aug-Dec_Longline | 4.952 | 0.347 |
| P2_Jan-Apr_Pot | -9.392 | 15.244 |
| P3_Jan-Apr_Pot | 5.033 | 0.046 |
| P4_Jan-Apr_Pot | 4.351 | 0.283 |
| P3_May-Jul_Pot | 4.920 | 0.084 |
| P1_Jan-Apr_Trawl_1977 | 68.881 | 3.174 |
| P1_Jan-Apr_Trawl_1985 | 75.859 | 1.714 |
| P1_Jan-Apr_Trawl_1990 | 68.955 | 1.108 |
| P1_Jan-Apr_Trawl_1995 | 73.966 | 0.960 |
| P1_Jan-Apr_Trawl_2000 | 78.219 | 1.231 |
| P1_Jan-Apr_Trawl_2005 | 78.243 | 0.708 |
| P3_Jan-Apr_Trawl_1977 | 6.173 | 0.179 |
| P3_Jan-Apr_Trawl_1985 | 6.604 | 0.079 |
| P3_Jan-Apr_Trawl_1990 | 6.094 | 0.059 |
| P3_Jan-Apr_Trawl_1995 | 6.299 | 0.047 |
| P3_Jan-Apr_Trawl_2000 | 6.304 | 0.062 |
| P3_Jan-Apr_Trawl_2005 | 6.022 | 0.037 |
| P1_May-Jul_Trawl_1977 | 50.183 | 1.717 |
| P1_May-Jul_Trawl_1985 | 51.339 | 1.748 |
| P1_May-Jul_Trawl_1990 | 62.003 | 1.546 |
| P1_May-Jul_Trawl_2000 | 53.046 | 1.523 |
| P1_May-Jul_Trawl_2005 | 58.497 | 1.421 |
| P1_Aug-Dec_Trawl_1977 | 62.702 | 4.089 |
| P1_Aug-Dec_Trawl_1980 | 81.845 | 5.841 |
| P1_Aug-Dec_Trawl_1985 | 85.878 | 5.340 |
| P1_Aug-Dec_Trawl_1990 | 75.459 | 33.666 |
| P1_Aug-Dec_Trawl_1995 | 102.473 |  |
| P1_Aug-Dec_Trawl_2000 | 56.927 | 1.470 |
| P3_Aug-Dec_Trawl_1977 | 5.556 | 0.332 |
| P3_Aug-Dec_Trawl_1980 | 6.664 | 0.237 |
| P3_Aug-Dec_Trawl_1985 | 6.593 | 0.234 |
| P3_Aug-Dec_Trawl_1990 | 6.318 | 1.927 |
| P3_Aug-Dec_Traw__1995 | 7.021 | 0.090 |
| P3_Aug-Dec_Trawl_2000 | 5.257 | 0.149 |
| P1_Jan-Apr_Longline_1977 | 59.291 | 2.087 |
| P1_Jan-Apr_Longline_1980 | 72.266 | 2.556 |
| P1_Jan-Apr_Longline_1985 | 74.994 | 0.927 |
| P1_Jan-Apr_Longline_1990 | 66.279 | 0.478 |
| P1_Jan-Apr_Longline_1995 | 65.878 | 0.427 |
| P1_Jan-Apr_Longline_2000 | 63.750 | 0.442 |
| P1_Jan-Apr_Longline_2005 | 67.679 | 0.334 |
| P3_Jan-Apr_Longline_1977 | 5.168 | 0.211 |
| P3_Jan-Apr_Longline_1980 | 5.907 | 0.184 |
| P3_Jan-Apr_Longline_1985 | 5.852 | 0.069 |
| P3_Jan-Apr_Longline_1990 | 5.242 | 0.046 |
| P3_Jan-Apr_Longline_1995 | 5.318 | 0.040 |
|  |  |  |


| Parameter | Est. | SD |
| :--- | ---: | ---: |
| P3_Jan-Apr_Longline_2000 | 5.383 | 0.041 |
| P3_Jan-Apr_Longline_2005 | 5.324 | 0.027 |
| P6_Jan-Apr_Longline_1977 | -1.259 | 0.799 |
| P6_Jan-Apr_Longline_1980 | 0.518 | 1.096 |
| P6_Jan-Apr_Longline_1985 | -1.072 | 0.422 |
| P6_Jan-Apr_Longline_1990 | -0.464 | 0.137 |
| P6_Jan-Apr_Longline_1995 | -0.636 | 0.139 |
| P6_Jan-Apr_Longline_2000 | -1.144 | 0.144 |
| P6_Jan-Apr_Longline_2005 | -0.810 | 0.139 |
| P1_May-Jul_Longline_1977 | 64.073 | 2.218 |
| P1_May-Jul_Longline_1980 | 62.984 | 1.372 |
| P1_May-Jul_Longline_1985 | 63.802 | 1.139 |
| P1_May-Jul_Longline_1990 | 64.180 | 0.478 |
| P1_May-Jul_Longline_2000 | 60.315 | 0.550 |
| P1_May-Jul_Longline_2005 | 65.578 | 0.479 |
| P1_Aug-Dec_Longline_1977 | 60.968 | 2.260 |
| P1_Aug-Dec_Longline_1980 | 69.217 | 1.678 |
| P1_Aug-Dec_Longline_1985 | 64.151 | 0.785 |
| P1_Aug-Dec_Longline_1990 | 67.114 | 0.726 |
| P1_Aug-Dec_Longline_1995 | 69.475 | 0.714 |
| P1_Aug-Dec_Longline_2000 | 63.592 | 0.434 |
| P1_Aug-Dec_Longline_2005 | 63.965 | 0.327 |
| P3_Aug-Dec_Longline_1977 | 4.570 | 0.325 |
| P3_Aug-Dec_Longline_1980 | 5.380 | 0.143 |
| P3_Aug-Dec_Longline_1985 | 4.845 | 0.092 |
| P3_Aug-Dec_Longline_1990 | 5.035 | 0.077 |
| P3_Aug-Dec_Longline_1995 | 5.507 | 0.054 |
| P3_Aug-Dec_Longline_2000 | 5.182 | 0.042 |
| P3_Aug-Dec_Longline_2005 | 4.990 | 0.032 |
| P6_Aug-Dec_Longline_1977 | -2.381 | 1.919 |
| P6_Aug-Dec_Longline_1980 | 0.823 | 0.853 |
| P6_Aug-Dec_Longline_1985 | 0.294 | 0.243 |
| P6_Aug-Dec_Longline_1990 | 2.687 | 1.114 |
| P6_Aug-Dec_Longline_1995 | 9.566 | 11.507 |
| P6_Aug-Dec_Longline_2000 | -0.275 | 0.181 |
| P6_Aug-Dec_Longline_2005 | 4.631 | 5.905 |
| P1_Jan-Apr_Pot_1977 | 69.059 | 0.918 |
| P1_Jan-Apr_Pot_1995 | 68.684 | 0.535 |
| P1_Jan-Apr_Pot_2000 | 68.328 | 0.506 |
| P1_Jan-Apr_Pot_2005 | 69.474 | 0.486 |
| P6_Jan-Apr_Pot_1977 | 0.200 | 0.551 |
| P6_Jan-Apr_Pot_1995 | -0.202 | 0.252 |
| P6_Jan-Apr_Pot_2000 | -0.543 | 0.233 |
| P6_Jan-Apr_Pot_2005 | 0.190 | 0.223 |
| P1_May-Jul_Pot_1977 | 0.874 |  |
| P1_May-Jul_Pot_1995 | 0.879 | 0.735 |
| P1_Aug-Dec_Pot_1977 | 0.091 |  |
| P1_Aug-Dec_Pot_2000 | 1.203 |  |
| P3_Aug-Dec_Pot_1977 | 0.651 |  |
| P3_Aug-Dec_Pot_2000 | 0.121 |  |

Table 2.22e—Survey selectivity parameters as estimated by Model 11.5. Color scale extends from red (low) to green (high).

| Parameter | Estimate | St. dev. |
| :--- | ---: | ---: |
| P1 | 1.265 | 0.052 |
| P2 | -2.682 | 0.346 |
| P3 | -2.384 | 0.427 |
| P4 | 2.522 | 0.384 |
| P5 | -9.992 | - |
| P6 | -1.002 | 0.312 |
| P3_dev_1982 | -0.040 | 0.033 |
| P3_dev_1983 | -0.032 | 0.018 |
| P3_dev_1984 | -0.075 | 0.028 |
| P3_dev_1985 | 0.013 | 0.022 |
| P3_dev_1986 | -0.039 | 0.024 |
| P3_dev_1987 | 0.026 | 0.038 |
| P3_dev_1988 | -0.070 | 0.032 |
| P3_dev_1989 | -0.112 | 0.019 |
| P3_dev_1990 | -0.023 | 0.022 |
| P3_dev_1991 | -0.033 | 0.023 |
| P3_de__1992 | 0.092 | 0.040 |
| P3_dev_1993 | 0.054 | 0.029 |
| P3_dev_1994 | -0.032 | 0.022 |
| P3_dev_1995 | -0.084 | 0.021 |
| P3_dev_1996 | -0.106 | 0.019 |
| P3_dev_1997 | -0.055 | 0.016 |
| P3_dev_1998 | -0.069 | 0.020 |
| P3_dev_1999 | -0.072 | 0.018 |
| P3_dev_2000 | -0.027 | 0.017 |
| P3_dev_2001 | 0.168 | 0.037 |
| P3_dev_2002 | -0.012 | 0.024 |
| P3_dev_2003 | 0.009 | 0.020 |
| P3_dev_2004 | -0.009 | 0.021 |
| P3_de__2005 | 0.057 | 0.028 |
| P3_dev_2006 | 0.175 | 0.037 |
| P3_dev_2007 | 0.211 | 0.036 |
| P3_dev_2008 | 0.124 | 0.037 |
| P3_dev_2009 | 0.009 | 0.017 |
| P3_dev_2010 | -0.046 | 0.024 |
| P3_dev_2011 | 0.031 | 0.020 |
| P3_dev_2012 | 0.041 | 0.021 |
| P3_dev_2013 | -0.039 | 0.018 |
| P3_dev_2014 | -0.017 | 0.017 |
|  |  |  |

Table 2.22 f —Main selectivity parameters estimated by Models 16.x.

| Parameter | Model 16.1 |  | Model 16.6 |  | Model 16.7 |  | Model 16.8 |  | Model 16.9 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Est. | SD | Est. | SD | Est. | SD | Est. | SD | Est. | SD |
| Fishery A50\% | 4.319 | 0.046 | 4.324 | 0.046 | 4.303 | 0.040 | 4.327 | 0.047 | 4.496 | 64.179 |
| Fishery A95\%-A50\% | 1.156 | 0.032 | 1.158 | 0.032 | 1.193 | 0.029 | 1.164 | 0.032 | 1.264 | 12.234 |
| Trawl survey A50\% | 1.010 | 0.006 | 1.006 | 0.006 | 1.001 | 0.006 | 1.001 | 0.002 | 1.008 | 0.006 |
| Trawl survey A95\%-A50\% | 0.289 | 0.050 | 0.289 | 0.050 | 0.289 | 0.050 | 0.010 |  | 0.289 | 0.051 |
| NMFS LL survey A50\% | n/a | n/a | n/a | n/a | 4.009 | 0.078 | n/a | n/a | n/a | n/a |
| NMFS LL survey A95\%-A50\% | n/a | n/a | n/a | n/a | 0.190 | 1.656 | n/a | n/a | n/a | n/a |

Table 2.22g-Selectivity devs as estimated by Models 16.8 and 16.9. Color scale extends from red (low) to green (high). Standard deviations in Model 16.9 reflect logit-transformed parameters and wide bounds.

| Year | Model 16.8Survey A50\% dev |  | Model 16.9 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Fishery A50\% dev |  | Fishery A95\%-A50\% dev |  |
|  | Est. | SD | Est. | SD | Est. | SD |
| 1977 | n/a | n/a | -0.296 | 13.003 | -1.211 | 24.658 |
| 1978 | n/a | n/a | -0.199 | 12.968 | -0.635 | 13.206 |
| 1979 | n/a | n/a | -0.328 | 12.969 | -1.407 | 20.606 |
| 1980 | n/a | n/a | -0.290 | 13.011 | -1.292 | 22.732 |
| 1981 | n/a | n/a | -0.472 | 12.981 | -0.811 | 13.624 |
| 1982 | 0.0015 | 0.0012 | 0.262 | 12.970 | 1.256 | 14.080 |
| 1983 | 0.0001 | 0.0009 | 0.225 | 12.968 | 3.732 | 36.191 |
| 1984 | 0.0027 | 0.0011 | 0.106 | 12.968 | 4.926 | 32.618 |
| 1985 | -0.0004 | 0.0010 | -0.176 | 12.968 | -0.691 | 13.157 |
| 1986 | 0.0008 | 0.0010 | -0.012 | 12.968 | 0.331 | 13.151 |
| 1987 | -0.0004 | 0.0012 | 0.031 | 12.968 | 0.265 | 13.150 |
| 1988 | 0.0014 | 0.0015 | -0.110 | 12.968 | 1.694 | 14.847 |
| 1989 | 0.0033 | 0.0010 | 0.068 | 12.968 | 4.489 | 33.790 |
| 1990 | 0.0003 | 0.0010 | 0.232 | 12.968 | 4.856 | 32.791 |
| 1991 | 0.0010 | 0.0010 | -0.031 | 12.968 | 0.157 | 13.152 |
| 1992 | -0.0023 | 0.0014 | -0.130 | 12.968 | -0.437 | 13.148 |
| 1993 | -0.0013 | 0.0010 | -0.062 | 12.968 | 0.054 | 13.149 |
| 1994 | 0.0014 | 0.0010 | 0.016 | 12.968 | 0.211 | 13.148 |
| 1995 | 0.0022 | 0.0010 | -0.069 | 12.968 | -0.138 | 13.149 |
| 1996 | 0.0028 | 0.0010 | 0.072 | 12.968 | 0.207 | 13.149 |
| 1997 | 0.0010 | 0.0009 | 0.118 | 12.968 | 0.608 | 13.150 |
| 1998 | 0.0021 | 0.0010 | 0.025 | 12.968 | 0.176 | 13.149 |
| 1999 | 0.0020 | 0.0009 | -0.098 | 12.968 | -0.223 | 13.148 |
| 2000 | 0.0009 | 0.0009 | -0.045 | 12.968 | -0.639 | 13.159 |
| 2001 | -0.0066 | 0.0048 | -0.038 | 12.968 | -0.280 | 13.149 |
| 2002 | 0.0010 | 0.0010 | -0.133 | 12.968 | -0.417 | 13.148 |
| 2003 | -0.0013 | 0.0010 | -0.089 | 12.968 | -0.570 | 13.149 |
| 2004 | 0.0003 | 0.0009 | -0.092 | 12.972 | -1.299 | 22.665 |
| 2005 | -0.0022 | 0.0012 | -0.048 | 12.968 | -0.231 | 13.149 |
| 2006 | -0.0073 | 0.0076 | -0.051 | 12.968 | -0.457 | 13.150 |
| 2007 | -0.0358 | 2.2942 | 0.037 | 12.968 | -0.144 | 13.149 |
| 2008 | -0.0010 | 0.0010 | -0.041 | 12.968 | -0.359 | 13.148 |
| 2009 | -0.0010 | 0.0009 | -0.103 | 12.968 | -0.963 | 13.159 |
| 2010 | 0.0005 | 0.0011 | -0.070 | 12.996 | -1.430 | 20.052 |
| 2011 | -0.0014 | 0.0009 | 0.082 | 12.968 | -0.009 | 13.149 |
| 2012 | -0.0017 | 0.0010 | -0.076 | 12.989 | -1.402 | 20.778 |
| 2013 | 0.0011 | 0.0009 | -0.122 | 12.968 | -0.441 | 13.149 |
| 2014 | -0.0004 | 0.0009 | -0.070 | 12.968 | -0.595 | 13.148 |
| 2015 | n/a | n/a | -0.079 | 12.983 | -1.418 | 20.406 |
| 2016 | n/a | n/a | 0.004 | 12.968 | -0.192 | 13.150 |

Table 2.23a—Annual fishing mortality rates as estimated by the models. Color scale extends from red (low) to green (high) in each column.

|  | Model 11.5 |  | Model 16.1 |  | Model 16.6 |  | Model 16.7 |  | Model 16.8 |  | Model 16.9 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Yea | Est. | SD | Est. | SD | Est. | SD | Est. | SD | Est. | SD | Est. | SD |
| 1977 | 0.081 | 0.014 | 0.189 | 0.069 | 0.2 | 0.090 | 0.307 | 0.112 | 0.219 | 0.08 | . 88 | 0.026 |
| 1978 | 0.091 | 0.01 | 0.247 | 0.09 | 0.313 | 0.1 | 03 | 0.1 | 0.291 | 0.1 | 0.115 | 0.034 |
| 79 | 0.064 | 0.009 | 0.190 | 0.069 | 0.245 | 0.091 | 0.315 | 0.11 | 0.226 | 0.086 | 0.083 | 22 |
| 0 | 0.055 | 0.008 | 0.2 | 0.06 | 0.2 | 0.087 | 0.3 | 0.104 | 0.256 | 0.08 | 0.104 | 0.028 |
| 1 | 0.051 | 0.00 | 0.151 | 0.03 | 0.1 | 0.0 | 0. | 0.0 | 0. | 0.0 | . 056 | 0.013 |
| 982 | 0.040 | 0.004 | 0.088 | 0.011 | 0.096 | 0.012 | 0.12 | 0.01 | 0.103 | 0.01 | 0.171 | 0.1 |
| 1983 | 0.056 | 0.005 | 0.104 | 0.011 | 0.111 | 0.011 | 0.139 | 0.01 | 0.119 | 0.01 | 0.159 | 0.05 |
| 1984 | 0.075 | 0.005 | 0.124 | 0.01 | 0.1 | 0.01 | 0.18 | 0.01 | 0.141 | 0.012 | 0.138 | 0.019 |
| 1985 | 0.096 | 0.005 | 0.151 | 0.013 | 0.168 | 0.01 | 0.202 | 0.015 | 0.171 | 0.014 | 0.131 | 0.0 |
| 1986 | 0.093 | 0.005 | 0.151 | 0.012 | 0.170 | 0.013 | 0.200 | 0.014 | 0.171 | 0.013 | 0.14 | 0.012 |
| 1987 | 0.108 | 0.005 | 0.159 | 0.011 | 0.181 | 0.01 | 0.212 | 0.01 | 0.178 | 0.012 | 0.160 | 0.015 |
| 1988 | 0.144 | 0.007 | 0.229 | 0.01 | 0.242 | 0.016 | 0.280 | 0.016 | 0.251 | 0.016 | 0.195 | . 0 |
| 1989 | 0.134 | 0.006 | 0.201 | 0.013 | 0.205 | 0.012 | 0.235 | 0.013 | 0.220 | 0.013 | 0.215 | 0.02 |
| 1990 | 0.144 | 0.006 | 0.218 | 0.013 | 0.229 | 0.013 | 0.260 | 0.013 | 0.237 | 0.014 | 0.272 | . 026 |
| 1991 | 0.223 | 0.009 | 0.347 | 0.02 | 0.404 | 0.02 | 0.458 | 0.02 | 0.381 | 0.023 | 0.315 | . 025 |
| 1992 | 0.223 | 0.009 | 0.378 | 0.02 | 0.487 | 0.03 | 0.560 | 0.03 | 0.414 | 0.030 | 0.313 | 0.023 |
| 19 | 0.187 | 0.008 | 0.334 | 0.025 | 0.373 | 0.028 | 0.432 | 0.028 | 0.349 | 0.026 | 0.285 | 0.029 |
| 1994 | 0.221 | 0.016 | 0.357 | 0.024 | 0.402 | 0.026 | 0.467 | 0.026 | 0.377 | 0.024 | 0.379 | 0.042 |
| 1995 | 0.325 | 0.012 | 0.461 | 0.030 | 0.509 | 0.032 | 0.591 | 0.032 | 0.484 | 0.029 | 0.423 | 0.036 |
| 1996 | 0.296 | 0.01 | 0.41 | 0.028 | 0.470 | 0.03 | 0.543 | 0.030 | 0.458 | 0.029 | 0.516 | 0.054 |
| 1997 | 0.338 | 0.013 | 0.4 | 0.027 | 0.518 | 0.034 | 0.616 | 0.033 | 0.480 | 0.030 | 0.586 | 0.084 |
| 1998 | 0.265 | 0.011 | 0.285 | 0.019 | 0.416 | 0.029 | 0.5 | 0.031 | 0.323 | 0.02 | 0.290 | 0.029 |
| 1999 | 0.253 | 0.011 | 0.44 | 0.03 | 0.425 | 0.031 | 0.522 | 0.034 | 0.511 | 0.03 | 0.389 | 0.03 |
| 2000 | 0.241 | 0.009 | 0.422 | 0.030 | 0.408 | 0.03 | 0.491 | 0.032 | 0.471 | 0.033 | 0.408 | 0.0 |
| 2001 | 0.207 | 0.007 | 0.340 | 0.022 | 0.326 | 0.022 | 0.395 | 0.023 | 0.372 | 0.024 | 0.311 | 0.031 |
| 2002 | 0.250 | 0.008 | 0. | . 02 | 0.39 | 0.02 | 0. | 0.027 | 0.472 | 0.03 | 0.363 | 0.025 |
| 2003 | 0.268 | 0.008 | 0.4 | 0.02 | 0.4 | 0.02 | 0.5 | 0.028 | 0.4 | 0.02 | 0.359 | 0.02 |
| 2004 | 0.293 | 0.009 | 0.401 | 0.02 | 0.401 | 0.023 | 0.465 | 0.023 | 0.398 | 0.022 | 0.365 | 0.021 |
| 2005 | 0.325 | 0.010 | 0.416 | 0.024 | 0.410 | 0.022 | 0.475 | 0.022 | 0.422 | 0.023 | 0.409 | 0.029 |
| 2006 | 0.364 | 0.013 | 0.449 | 0.029 | 0.469 | 0.027 | 0.5 | 0.026 | 0.450 | 0.027 | 0.4 | 0.03 |
| 20 | 0.354 | 0.01 | 0.42 | 0.03 | 0.455 | 0.028 | 0.540 | 0.027 | 0.429 | 0.028 | 0 | 0.047 |
| 2008 | 0.399 | 0.015 | 0.483 | 0.037 | 0.5 | 0.038 | 0.685 | 0.038 | 0.471 | 0.032 | 0.510 | . 0 |
| 2009 | 0.447 | 0.019 | 0.652 | . 05 | 0.68 | . 05 | 0.85 | 0.05 | 0.660 | 0.05 | 0.646 | . 066 |
| 010 | 0.369 | 0.01 | 0.46 | 0.04 | 0.5 | 0.04 | 0.6 | 0.04 | 0.5 | 0.0 | 0.810 | . 1 |
| 2011 | 0.43 | 0.020 | 0.4 | 0.035 | 0.5 | 0.04 | 0.6 | 0.04 | 0.5 | 0.0 | 0.6 | 0.232 |
| 2012 | 0.398 | 0.019 | 0.403 | 0.036 | 0.496 | 0.040 | 0.620 | 0.04 | 0.506 | 0.03 | 0.48 | 0.05 |
| 2013 | 0.343 | 0.017 | 0.318 | 0.028 | 0.40 | 0.033 | 0.518 | 0.03 | 0.405 | 0.031 | 0.282 | 0.02 |
| 2014 | 0.325 | 0.018 | 0.351 | 0.034 | 0.453 | 0.042 | 0.597 | 0.047 | 0.467 | 0.041 | 0.345 | 0.03 |
| 2015 | 0.269 | 0.016 | 0.281 | 0.027 | 0.391 | 0.038 | 0.517 | 0.044 | 0.388 | 0.036 | 0.303 | 0.03 |
| 2016 | 0.260 | 0.017 | 0.289 | 0.029 | 0.343 | 0.034 | 0.462 | 0.043 | 0.416 | 0.042 | 0.300 | 0.052 |

Table 2.23b-Model 11.5 estimates of seasonal full-selection fishing mortality rates, on an annual time scale. Sea1=Jan-Feb, Sea2=Mar-Apr, Sea3=May-Jul, Sea4=Aug-Oct, Sea5=Nov-Dec.

|  | Trawl fishery |  |  |  |  | Longline fishery |  |  |  |  | Pot fishery |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Sea1 | Sea2 | Sea3 | Sea4 | Sea5 | Sea1 | Sea2 | Sea3 | Sea4 | Sea5 | Sea1 | Sea2 | Sea3 | Sea4 | Sea5 |
| 1977 | 0.086 | 0.089 | 0.055 | 0.049 | 0.042 | 0.017 | 0.017 | 0.006 | 0.024 | 0.032 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0.098 | 0.101 | 0.066 | 0.056 | 0.050 | 0.017 | 0.017 | 0.006 | 0.025 | 0.034 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0.070 | 0.072 | 0.043 | 0.039 | 0.033 | 0.013 | 0.013 | 0.005 | 0.019 | 0.025 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 0.062 | 0.062 | 0.031 | 0.041 | 0.034 | 0.010 | 0.010 | 0.004 | 0.013 | 0.017 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0.034 | 0.033 | 0.033 | 0.064 | 0.060 | 0.004 | 0.004 | 0.002 | 0.009 | 0.010 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0.035 | 0.035 | 0.036 | 0.045 | 0.036 | 0.001 | 0.001 | 0.001 | 0.004 | 0.005 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 0.055 | 0.057 | 0.051 | 0.053 | 0.044 | 0.005 | 0.005 | 0.003 | 0.004 | 0.005 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 0.062 | 0.066 | 0.057 | 0.056 | 0.049 | 0.007 | 0.008 | 0.006 | 0.027 | 0.038 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 0.079 | 0.084 | 0.067 | 0.065 | 0.051 | 0.024 | 0.026 | 0.010 | 0.034 | 0.048 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 0.088 | 0.094 | 0.067 | 0.065 | 0.053 | 0.017 | 0.019 | 0.005 | 0.027 | 0.039 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 0.097 | 0.103 | 0.053 | 0.053 | 0.052 | 0.043 | 0.046 | 0.013 | 0.043 | 0.061 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 0.196 | 0.211 | 0.103 | 0.113 | 0.120 | 0.001 | 0.001 | 0.002 | 0.003 | 0.004 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 0.209 | 0.227 | 0.100 | 0.059 | 0.054 | 0.008 | 0.009 | 0.012 | 0.015 | 0.013 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 0.178 | 0.195 | 0.093 | 0.033 | 0.029 | 0.032 | 0.035 | 0.048 | 0.052 | 0.047 | 0.000 | 0.000 | 0.002 | 0.002 | 0.001 |
| 1991 | 0.182 | 0.382 | 0.068 | 0.056 | 0.044 | 0.063 | 0.106 | 0.088 | 0.100 | 0.066 | 0.000 | 0.004 | 0.002 | 0.011 | 0.003 |
| 1992 | 0.150 | 0.225 | 0.056 | 0.044 | 0.014 | 0.135 | 0.241 | 0.143 | 0.093 | 0.000 | 0.000 | 0.009 | 0.030 | 0.011 | 0.000 |
| 1993 | 0.191 | 0.260 | 0.026 | 0.036 | 0.011 | 0.228 | 0.232 | 0.025 | 0.016 | 0.004 | 0.000 | 0.018 | 0.013 | 0.000 | 0.000 |
| 1994 | 0.088 | 0.297 | 0.020 | 0.100 | 0.019 | 0.194 | 0.267 | 0.030 | 0.105 | 0.000 | 0.000 | 0.041 | 0.010 | 0.016 | 0.000 |
| 1995 | 0.216 | 0.435 | 0.005 | 0.199 | 0.002 | 0.248 | 0.317 | 0.021 | 0.110 | 0.065 | 0.000 | 0.078 | 0.040 | 0.015 | 0.005 |
| 1996 | 0.146 | 0.381 | 0.031 | 0.109 | 0.022 | 0.243 | 0.270 | 0.021 | 0.122 | 0.024 | 0.000 | 0.130 | 0.063 | 0.023 | 0.005 |
| 1997 | 0.183 | 0.413 | 0.026 | 0.101 | 0.037 | 0.273 | 0.291 | 0.044 | 0.118 | 0.196 | 0.000 | 0.101 | 0.042 | 0.021 | 0.005 |
| 1998 | 0.124 | 0.242 | 0.027 | 0.143 | 0.022 | 0.290 | 0.209 | 0.022 | 0.098 | 0.116 | 0.018 | 0.066 | 0.033 | 0.012 | 0.004 |
| 1999 | 0.149 | 0.227 | 0.017 | 0.067 | 0.004 | 0.334 | 0.249 | 0.021 | 0.129 | 0.044 | 0.023 | 0.065 | 0.036 | 0.013 | 0.001 |
| 2000 | 0.175 | 0.230 | 0.017 | 0.027 | 0.004 | 0.314 | 0.088 | 0.007 | 0.128 | 0.147 | 0.142 | 0.053 | 0.007 | 0.009 | 0.000 |
| 2001 | 0.068 | 0.125 | 0.015 | 0.039 | 0.005 | 0.167 | 0.162 | 0.019 | 0.174 | 0.162 | 0.023 | 0.124 | 0.006 | 0.015 | 0.005 |
| 2002 | 0.112 | 0.191 | 0.035 | 0.038 | 0.002 | 0.338 | 0.152 | 0.010 | 0.203 | 0.122 | 0.019 | 0.096 | 0.003 | 0.016 | 0.006 |
| 2003 | 0.132 | 0.152 | 0.030 | 0.035 | 0.002 | 0.351 | 0.182 | 0.014 | 0.207 | 0.154 | 0.130 | 0.020 | 0.003 | 0.027 | 0.011 |
| 2004 | 0.193 | 0.168 | 0.048 | 0.044 | 0.001 | 0.378 | 0.184 | 0.016 | 0.201 | 0.192 | 0.101 | 0.035 | 0.006 | 0.022 | 0.005 |
| 2005 | 0.283 | 0.174 | 0.044 | 0.017 | 0.001 | 0.536 | 0.085 | 0.022 | 0.232 | 0.203 | 0.104 | 0.040 | 0.000 | 0.030 | 0.004 |
| 2006 | 0.350 | 0.193 | 0.044 | 0.030 | 0.001 | 0.632 | 0.095 | 0.016 | 0.337 | 0.012 | 0.149 | 0.053 | 0.003 | 0.031 | 0.009 |
| 2007 | 0.231 | 0.267 | 0.083 | 0.026 | 0.001 | 0.709 | 0.035 | 0.012 | 0.278 | 0.010 | 0.179 | 0.021 | 0.005 | 0.046 | 0.000 |
| 2008 | 0.262 | 0.134 | 0.035 | 0.055 | 0.008 | 0.783 | 0.077 | 0.029 | 0.343 | 0.122 | 0.170 | 0.042 | 0.003 | 0.066 | 0.002 |
| 2009 | 0.237 | 0.208 | 0.035 | 0.080 | 0.004 | 0.944 | 0.086 | 0.026 | 0.366 | 0.150 | 0.209 | 0.043 | 0.003 | 0.014 | 0.016 |
| 2010 | 0.309 | 0.162 | 0.027 | 0.060 | 0.014 | 0.724 | 0.037 | 0.028 | 0.204 | 0.142 | 0.221 | 0.037 | 0.003 | 0.045 | 0.021 |
| 2011 | 0.320 | 0.333 | 0.040 | 0.067 | 0.012 | 0.391 | 0.368 | 0.113 | 0.213 | 0.160 | 0.234 | 0.037 | 0.005 | 0.064 | 0.000 |
| 2012 | 0.490 | 0.195 | 0.040 | 0.051 | 0.007 | 0.365 | 0.274 | 0.133 | 0.149 | 0.158 | 0.245 | 0.030 | 0.005 | 0.034 | 0.006 |
| 2013 | 0.363 | 0.184 | 0.026 | 0.086 | 0.009 | 0.345 | 0.210 | 0.077 | 0.133 | 0.162 | 0.192 | 0.035 | 0.000 | 0.030 | 0.026 |
| 2014 | 0.287 | 0.216 | 0.027 | 0.039 | 0.005 | 0.279 | 0.240 | 0.110 | 0.117 | 0.138 | 0.203 | 0.073 | 0.001 | 0.028 | 0.023 |
| 2015 | 0.176 | 0.214 | 0.026 | 0.033 | 0.003 | 0.207 | 0.211 | 0.090 | 0.117 | 0.112 | 0.168 | 0.069 | 0.001 | 0.026 | 0.017 |
| 2016 | 0.238 | 0.160 | 0.024 | 0.029 | 0.002 | 0.221 | 0.158 | 0.085 | 0.105 | 0.101 | 0.132 | 0.132 | 0.000 | 0.023 | 0.016 |

Table 2.24—Summary of key management reference points from the standard projection algorithm (last seven rows are from SS). All biomass figures are in t . Color scale: red = row minimum, green = row maximum.

| Quantity | M11.5 | M16.1 | M16.6 | M16.7 | M16.8 | M16.9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| B100\% | 788,000 | 668,000 | 620,000 | 609,000 | 631,000 | 681,000 |
| B40\% | 315,000 | 267,000 | 248,000 | 243,000 | 252,000 | 272,000 |
| B35\% | 276,000 | 234,000 | 217,000 | 213,000 | 221,000 | 238,000 |
| B(2017) | 440,000 | 380,000 | 327,000 | 242,000 | 268,000 | 393,000 |
| B(2018) | 462,000 | 393,000 | 340,000 | 267,000 | 289,000 | 408,000 |
| B(2017)/B100\% | 0.56 | 0.57 | 0.53 | 0.40 | 0.42 | 0.58 |
| B(2018)/B100\% | 0.59 | 0.59 | 0.55 | 0.44 | 0.46 | 0.60 |
| F40\% | 0.28 | 0.29 | 0.31 | 0.29 | 0.29 | 0.32 |
| F35\% | 0.34 | 0.36 | 0.38 | 0.35 | 0.35 | 0.38 |
| maxFABC(2017) | 0.28 | 0.29 | 0.31 | 0.29 | 0.29 | 0.32 |
| maxFABC(2018) | 0.28 | 0.29 | 0.31 | 0.29 | 0.29 | 0.32 |
| maxABC(2017) | 338,000 | 265,000 | 239,000 | 170,000 | 191,000 | 276,000 |
| maxABC(2018) | 325,000 | 280,000 | 255,000 | 192,000 | 211,000 | 302,000 |
| FOFL(2017) | 0.34 | 0.36 | 0.38 | 0.35 | 0.35 | 0.38 |
| FOFL(2018) | 0.34 | 0.36 | 0.38 | 0.35 | 0.35 | 0.38 |
| OFL(2017) | 396,000 | 314,000 | 284,000 | 200,000 | 226,000 | 327,000 |
| OFL(2018) | 381,000 | 331,000 | 302,000 | 228,000 | 249,000 | 357,000 |
| $\operatorname{Pr}(m a x A B C(2017)>\operatorname{truOFL(2017))}$ | 0.01 | 0.08 | 0.08 | 0.08 | 0.08 | 0.09 |
| $\operatorname{Pr(maxABC(2018)>\text {truOFL(2018))}} 2$ | 0.03 | 0.10 | 0.10 | 0.08 | 0.10 | 0.10 |
| $\operatorname{Pr(B(2017)<B20\% )}$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| $\operatorname{Pr}(B(2018)<B 20 \%)$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| $\operatorname{Pr}(B(2019)<B 20 \%)$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| $\operatorname{Pr}(B(2020)<B 20 \%)$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| $\operatorname{Pr(B(2021)<B20\% )}$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |

## Legend:

B100\% = equilibrium unfished spawning biomass
B40\% = 40\% of B100\% (the inflection point of the harvest control rules in Tier 3)
$B 35 \%=35 \%$ of B100\% (the BMSY proxy for Tier 3)
$\mathrm{B}($ year $)=$ projected spawning biomass for year
B (year)/B100\% = ratio of spawning biomass to B100\%
F40\% = fishing mortality that reduces equilibrium spawning per recruit to $40 \%$ of unfished F35\% = fishing mortality that reduces equilibrium spawning per recruit to $35 \%$ of unfished maxFABC (year) = maximum permissible ABC fishing mortality rate under Tier 3 maxABC(year) = maximum permissible ABC under Tier 3
FOFL(year) = OFL fishing mortality rate under Tier 3
OFL(year) = OFL under Tier 3
$\operatorname{Pr}(\operatorname{maxABC}($ year $)>t r u O F L($ year $))=$ probability that maxABC is greater than the "true" OFL
$\operatorname{Pr}(\mathrm{B}($ year $)<\mathrm{B} 20 \%)=$ probability that spawning biomass is less than $20 \%$ of unfished

Table 2.25-Schedules of length (cm) at age and selectivity at age as defined by parameter estimates from Model 16.6.

|  | Begin-year length |  | Mid-year length |  | Selectivity |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | Mean | SD | Mean | SD | Fishery | Survey |
| 0 | 0.001 | 3.425 | 5.468 | 3.425 | 0.000 | 0.000 |
| 1 | 10.934 | 3.425 | 16.401 | 3.425 | 0.000 | 0.486 |
| 2 | 24.577 | 4.045 | 31.866 | 4.598 | 0.003 | 1.000 |
| 3 | 38.398 | 5.093 | 44.268 | 5.538 | 0.033 | 1.000 |
| 4 | 49.552 | 5.939 | 54.316 | 6.300 | 0.305 | 1.000 |
| 5 | 58.614 | 6.626 | 62.495 | 6.920 | 0.848 | 1.000 |
| 6 | 66.001 | 7.186 | 69.169 | 7.426 | 0.986 | 1.000 |
| 7 | 72.033 | 7.643 | 74.623 | 7.839 | 0.999 | 1.000 |
| 8 | 76.966 | 8.017 | 79.086 | 8.178 | 1.000 | 1.000 |
| 9 | 81.004 | 8.323 | 82.739 | 8.455 | 1.000 | 1.000 |
| 10 | 84.311 | 8.574 | 85.733 | 8.682 | 1.000 | 1.000 |
| 11 | 87.020 | 8.779 | 88.186 | 8.868 | 1.000 | 1.000 |
| 12 | 89.241 | 8.948 | 90.197 | 9.020 | 1.000 | 1.000 |
| 13 | 91.063 | 9.086 | 91.847 | 9.145 | 1.000 | 1.000 |
| 14 | 92.556 | 9.199 | 93.199 | 9.248 | 1.000 | 1.000 |
| 15 | 93.782 | 9.292 | 94.309 | 9.332 | 1.000 | 1.000 |
| 16 | 94.787 | 9.368 | 95.220 | 9.401 | 1.000 | 1.000 |
| 17 | 95.612 | 9.431 | 95.967 | 9.458 | 1.000 | 1.000 |
| 18 | 96.288 | 9.482 | 96.580 | 9.504 | 1.000 | 1.000 |
| 19 | 96.844 | 9.524 | 97.083 | 9.542 | 1.000 | 1.000 |
| 20 | 97.738 | 9.717 | 97.893 | 9.717 | 1.000 | 1.000 |

Table 2.26-Begin-year weight at age as defined by input weight-at-length parameters and length-at-age parameters estimated by Model 16.6.

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 0.00 | 0.02 | 0.18 | 0.69 | 1.52 | 2.55 | 3.68 | 4.82 | 5.91 | 6.93 | 7.84 | 8.65 | 9.35 | 9.95 | 10.46 | 10.90 | 11.26 | 11.56 | 11.82 | 12.03 | 12.38 |
| 1978 | 0.00 | 0.01 | 0.15 | 0.63 | 1.45 | 2.51 | 3.72 | 4.96 | 6.17 | 7.31 | 8.34 | 9.26 | 10.06 | 10.76 | 11.35 | 11.85 | 12.27 | 12.63 | 12.92 | 13.17 | 13.58 |
| 1979 | 0.00 | 0.02 | 0.17 | 0.68 | 1.51 | 2.54 | 3.67 | 4.82 | 5.93 | 6.95 | 7.88 | 8.69 | 9.40 | 10.01 | 10.53 | 10.97 | 11.34 | 11.65 | 11.91 | 12.12 | 12.47 |
| 1980 | 0.00 | 0.01 | 0.16 | 0.65 | 1.45 | 2.46 | 3.58 | 4.72 | 5.82 | 6.85 | 7.77 | 8.59 | 9.31 | 9.92 | 10.45 | 10.89 | 11.27 | 11.58 | 11.84 | 12.05 | 12.41 |
| 1981 | 0.00 | 0.02 | 0.17 | 0.66 | 1.45 | 2.45 | 3.55 | 4.67 | 5.74 | 6.74 | 7.64 | 8.44 | 9.13 | 9.73 | 10.24 | 10.66 | 11.03 | 11.33 | 11.58 | 11.79 | 12.13 |
| 1982 | 0.00 | 0.02 | 0.18 | 0.71 | 1.55 | 2.59 | 3.74 | 4.89 | 5.99 | 7.02 | 7.94 | 8.75 | 9.46 | 10.06 | 10.58 | 11.01 | 11.38 | 11.69 | 11.94 | 12.15 | 12.50 |
| 1983 | 0.00 | 0.02 | 0.17 | 0.69 | 1.53 | 2.60 | 3.78 | 4.99 | 6.15 | 7.23 | 8.20 | 9.06 | 9.81 | 10.46 | 11.01 | 11.48 | 11.87 | 12.20 | 12.47 | 12.70 | 13.07 |
| 1984 | 0.00 | 0.02 | 0.20 | 0.72 | 1.49 | 2.42 | 3.41 | 4.39 | 5.32 | 6.17 | 6.92 | 7.59 | 8.16 | 8.65 | 9.07 | 9.42 | 9.71 | 9.96 | 10.16 | 10.33 | 10.61 |
| 1985 | 0.00 | 0.01 | 0.16 | 0.67 | 1.51 | 2.59 | 3.80 | 5.03 | 6.23 | 7.34 | 8.35 | 9.25 | 10.03 | 10.70 | 11.28 | 11.77 | 12.18 | 12.52 | 12.80 | 13.04 | 13.43 |
| 1986 | 0.00 | 0.01 | 0.15 | 0.64 | 1.48 | 2.57 | 3.80 | 5.06 | 6.30 | 7.45 | 8.50 | 9.43 | 10.25 | 10.96 | 11.56 | 12.07 | 12.50 | 12.86 | 13.15 | 13.40 | 13.82 |
| 1987 | 0.00 | 0.02 | 0.17 | 0.67 | 1.51 | 2.57 | 3.74 | 4.94 | 6.10 | 7.17 | 8.15 | 9.01 | 9.76 | 10.41 | 10.96 | 11.43 | 11.82 | 12.15 | 12.42 | 12.65 | 13.03 |
| 1988 | 0.00 | 0.01 | 0.16 | 0.66 | 1.53 | 2.66 | 3.92 | 5.23 | 6.51 | 7.70 | 8.79 | 9.75 | 10.60 | 11.33 | 11.95 | 12.48 | 12.92 | 13.29 | 13.60 | 13.86 | 14.29 |
| 1989 | 0.00 | 0.01 | 0.17 | 0.69 | 1.57 | 2.70 | 3.97 | 5.27 | 6.53 | 7.71 | 8.78 | 9.73 | 10.56 | 11.27 | 11.88 | 12.40 | 12.83 | 13.20 | 13.50 | 13.75 | 14.17 |
| 1990 | 0.00 | 0.02 | 0.18 | 0.71 | 1.58 | 2.66 | 3.86 | 5.08 | 6.25 | 7.33 | 8.31 | 9.18 | 9.94 | 10.59 | 11.14 | 11.61 | 12.00 | 12.33 | 12.60 | 12.83 | 13.21 |
| 1991 | 0.00 | 0.02 | 0.17 | 0.68 | 1.49 | 2.51 | 3.62 | 4.74 | 5.83 | 6.83 | 7.73 | 8.53 | 9.22 | 9.82 | 10.33 | 10.76 | 11.12 | 11.42 | 11.67 | 11.88 | 12.22 |
| 1992 | 0.00 | 0.02 | 0.16 | 0.64 | 1.42 | 2.41 | 3.50 | 4.61 | 5.67 | 6.66 | 7.56 | 8.35 | 9.04 | 9.63 | 10.13 | 10.56 | 10.92 | 11.22 | 11.47 | 11.68 | 12.02 |
| 1993 | 0.00 | 0.02 | 0.19 | 0.74 | 1.62 | 2.71 | 3.91 | 5.12 | 6.28 | 7.36 | 8.33 | 9.19 | 9.93 | 10.57 | 11.12 | 11.58 | 11.96 | 12.29 | 12.55 | 12.78 | 13.15 |
| 1994 | 0.00 | 0.02 | 0.17 | 0.67 | 1.49 | 2.52 | 3.66 | 4.81 | 5.93 | 6.96 | 7.90 | 8.72 | 9.44 | 10.06 | 10.59 | 11.04 | 11.41 | 11.72 | 11.99 | 12.20 | 12.56 |
| 1995 | 0.00 | 0.01 | 0.16 | 0.65 | 1.47 | 2.52 | 3.69 | 4.88 | 6.04 | 7.12 | 8.09 | 8.96 | 9.72 | 10.37 | 10.92 | 11.39 | 11.79 | 12.12 | 12.39 | 12.62 | 13.00 |
| 1996 | 0.00 | 0.02 | 0.20 | 0.74 | 1.56 | 2.56 | 3.64 | 4.72 | 5.75 | 6.69 | 7.53 | 8.27 | 8.92 | 9.47 | 9.94 | 10.33 | 10.66 | 10.94 | 11.17 | 11.36 | 11.67 |
| 1997 | 0.00 | 0.02 | 0.17 | 0.65 | 1.43 | 2.40 | 3.47 | 4.55 | 5.59 | 6.56 | 7.43 | 8.20 | 8.87 | 9.44 | 9.93 | 10.35 | 10.69 | 10.98 | 11.22 | 11.43 | 11.76 |
| 1998 | 0.00 | 0.02 | 0.17 | 0.65 | 1.42 | 2.39 | 3.45 | 4.52 | 5.54 | 6.49 | 7.35 | 8.10 | 8.76 | 9.33 | 9.81 | 10.21 | 10.55 | 10.84 | 11.08 | 11.27 | 11.60 |
| 1999 | 0.00 | 0.02 | 0.17 | 0.68 | 1.48 | 2.49 | 3.60 | 4.71 | 5.79 | 6.78 | 7.67 | 8.46 | 9.15 | 9.74 | 10.24 | 10.67 | 11.02 | 11.32 | 11.57 | 11.77 | 12.11 |
| 2000 | 0.00 | 0.02 | 0.18 | 0.71 | 1.56 | 2.62 | 3.79 | 4.96 | 6.09 | 7.14 | 8.08 | 8.92 | 9.64 | 10.27 | 10.80 | 11.24 | 11.62 | 11.93 | 12.20 | 12.41 | 12.77 |
| 2001 | 0.00 | 0.02 | 0.19 | 0.71 | 1.54 | 2.56 | 3.67 | 4.79 | 5.86 | 6.85 | 7.74 | 8.52 | 9.20 | 9.78 | 10.28 | 10.70 | 11.05 | 11.34 | 11.58 | 11.79 | 12.12 |
| 2002 | 0.00 | 0.02 | 0.17 | 0.68 | 1.50 | 2.53 | 3.66 | 4.81 | 5.91 | 6.93 | 7.85 | 8.67 | 9.38 | 9.98 | 10.50 | 10.94 | 11.31 | 11.62 | 11.87 | 12.08 | 12.44 |
| 2003 | 0.00 | 0.02 | 0.16 | 0.66 | 1.47 | 2.50 | 3.64 | 4.81 | 5.93 | 6.98 | 7.93 | 8.77 | 9.50 | 10.13 | 10.67 | 11.12 | 11.50 | 11.82 | 12.09 | 12.31 | 12.68 |
| 2004 | 0.00 | 0.02 | 0.18 | 0.69 | 1.50 | 2.52 | 3.64 | 4.76 | 5.85 | 6.85 | 7.75 | 8.55 | 9.24 | 9.84 | 10.34 | 10.77 | 11.13 | 11.43 | 11.68 | 11.89 | 12.23 |
| 2005 | 0.00 | 0.02 | 0.17 | 0.67 | 1.49 | 2.54 | 3.69 | 4.87 | 6.01 | 7.07 | 8.02 | 8.87 | 9.60 | 10.24 | 10.78 | 11.24 | 11.62 | 11.94 | 12.21 | 12.43 | 12.80 |
| 2006 | 0.00 | 0.02 | 0.17 | 0.68 | 1.50 | 2.53 | 3.67 | 4.82 | 5.93 | 6.96 | 7.89 | 8.71 | 9.42 | 10.04 | 10.56 | 11.00 | 11.38 | 11.69 | 11.95 | 12.16 | 12.52 |
| 2007 | 0.00 | 0.02 | 0.17 | 0.68 | 1.51 | 2.57 | 3.73 | 4.91 | 6.04 | 7.10 | 8.06 | 8.90 | 9.64 | 10.27 | 10.81 | 11.27 | 11.65 | 11.97 | 12.24 | 12.46 | 12.83 |
| 2008 | 0.00 | 0.02 | 0.19 | 0.71 | 1.53 | 2.55 | 3.65 | 4.76 | 5.82 | 6.79 | 7.67 | 8.44 | 9.11 | 9.69 | 10.18 | 10.60 | 10.94 | 11.23 | 11.47 | 11.67 | 12.00 |
| 2009 | 0.00 | 0.01 | 0.16 | 0.66 | 1.49 | 2.54 | 3.71 | 4.91 | 6.07 | 7.15 | 8.12 | 8.99 | 9.74 | 10.40 | 10.95 | 11.42 | 11.81 | 12.14 | 12.42 | 12.65 | 13.03 |
| 2010 | 0.00 | 0.02 | 0.17 | 0.67 | 1.48 | 2.50 | 3.62 | 4.75 | 5.84 | 6.85 | 7.76 | 8.57 | 9.27 | 9.87 | 10.38 | 10.82 | 11.18 | 11.49 | 11.74 | 11.95 | 12.30 |
| 2011 | 0.00 | 0.02 | 0.17 | 0.66 | 1.45 | 2.45 | 3.55 | 4.67 | 5.75 | 6.75 | 7.65 | 8.45 | 9.14 | 9.74 | 10.25 | 10.68 | 11.04 | 11.34 | 11.59 | 11.80 | 12.15 |
| 2012 | 0.00 | 0.02 | 0.18 | 0.67 | 1.46 | 2.44 | 3.50 | 4.58 | 5.61 | 6.55 | 7.41 | 8.16 | 8.82 | 9.38 | 9.86 | 10.26 | 10.60 | 10.88 | 11.12 | 11.31 | 11.64 |
| 2013 | 0.00 | 0.01 | 0.16 | 0.64 | 1.43 | 2.45 | 3.57 | 4.72 | 5.83 | 6.86 | 7.80 | 8.63 | 9.35 | 9.97 | 10.51 | 10.96 | 11.33 | 11.65 | 11.91 | 12.13 | 12.49 |
| 2014 | 0.00 | 0.01 | 0.15 | 0.61 | 1.39 | 2.40 | 3.53 | 4.69 | 5.81 | 6.86 | 7.82 | 8.67 | 9.41 | 10.05 | 10.59 | 11.06 | 11.44 | 11.77 | 12.04 | 12.26 | 12.64 |
| 2015 | 0.00 | 0.01 | 0.14 | 0.59 | 1.36 | 2.34 | 3.45 | 4.59 | 5.69 | 6.73 | 7.67 | 8.50 | 9.23 | 9.86 | 10.40 | 10.85 | 11.23 | 11.55 | 11.82 | 12.04 | 12.41 |
| 2016 | 0.00 | 0.02 | 0.16 | 0.66 | 1.47 | 2.50 | 3.64 | 4.79 | 5.91 | 6.95 | 7.89 | 8.72 | 9.44 | 10.07 | 10.60 | 11.05 | 11.42 | 11.74 | 12.00 | 12.22 | 12.58 |

Table 2.27-Time series of EBS Pacific cod age 0+ biomass, age 3+ biomass, female spawning biomass ( t ), and standard deviation of spawning biomass ("SB SD") as estimated by the final models in last year's and this year's assessments. Spawning biomasses listed for 2016 under last year's assessment and for 2017 under this year's assessment represent output from the standard projection model.

| Year | Last year's assessment |  |  |  | This year's assessment |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 0+ | Age 3+ | Spawn. | SB SD | Age 0+ | Age 3+ | Spawn. | SB SD |
| 1977 | 580,639 | 572,253 | 163,291 | 31,819 | 238,614 | 223,326 | 70,865 | 24,355 |
| 1978 | 662,973 | 613,704 | 180,863 | 31,815 | 256,462 | 223,849 | 69,420 | 24,426 |
| 1979 | 837,745 | 721,585 | 209,322 | 32,738 | 370,312 | 265,037 | 70,557 | 23,663 |
| 1980 | 1,208,950 | 1,156,370 | 265,060 | 34,971 | 570,378 | 505,965 | 92,073 | 24,019 |
| 1981 | 1,629,440 | 1,569,570 | 369,954 | 38,337 | 827,796 | 765,577 | 145,981 | 25,966 |
| 1982 | 1,972,970 | 1,951,710 | 518,100 | 42,278 | 1,106,220 | 1,073,460 | 247,693 | 31,990 |
| 1983 | 2,151,420 | 2,131,750 | 658,400 | 44,272 | 1,237,100 | 1,210,540 | 350,393 | 36,843 |
| 1984 | 2,163,410 | 2,085,600 | 730,720 | 42,467 | 1,190,670 | 1,090,430 | 371,180 | 33,482 |
| 1985 | 2,139,140 | 2,115,600 | 726,420 | 37,989 | 1,276,680 | 1,246,900 | 420,649 | 35,009 |
| 1986 | 2,086,050 | 2,021,450 | 689,270 | 32,820 | 1,269,260 | 1,194,720 | 406,577 | 31,529 |
| 1987 | 2,063,010 | 2,035,030 | 663,430 | 28,260 | 1,291,840 | 1,259,500 | 395,026 | 27,811 |
| 1988 | 1,990,860 | 1,976,770 | 640,555 | 24,608 | 1,315,560 | 1,296,100 | 417,063 | 26,771 |
| 1989 | 1,792,190 | 1,780,400 | 601,150 | 21,610 | 1,200,380 | 1,189,160 | 402,900 | 24,351 |
| 1990 | 1,564,990 | 1,536,690 | 552,990 | 18,953 | 1,016,270 | 982,785 | 365,345 | 20,398 |
| 1991 | 1,367,790 | 1,314,930 | 477,333 | 16,285 | 833,926 | 762,154 | 292,052 | 15,778 |
| 1992 | 1,232,480 | 1,190,500 | 384,888 | 13,820 | 719,402 | 665,888 | 207,754 | 12,834 |
| 1993 | 1,222,170 | 1,195,220 | 337,244 | 12,072 | 830,487 | 780,122 | 192,784 | 12,988 |
| 1994 | 1,269,250 | 1,214,270 | 350,966 | 11,363 | 859,959 | 780,070 | 200,617 | 12,101 |
| 1995 | 1,295,990 | 1,275,180 | 353,929 | 11,328 | 913,253 | 881,497 | 222,817 | 12,672 |
| 1996 | 1,236,950 | 1,214,240 | 347,596 | 11,460 | 912,478 | 876,930 | 223,066 | 13,106 |
| 1997 | 1,156,000 | 1,128,240 | 336,292 | 11,391 | 785,469 | 753,258 | 214,083 | 12,595 |
| 1998 | 1,051,930 | 995,227 | 307,581 | 11,089 | 685,519 | 616,558 | 188,334 | 12,288 |
| 1999 | 1,072,690 | 1,048,190 | 292,121 | 10,714 | 717,289 | 681,281 | 182,377 | 12,549 |
| 2000 | 1,110,350 | 1,084,210 | 291,681 | 10,439 | 777,932 | 737,965 | 189,306 | 13,136 |
| 2001 | 1,126,320 | 1,081,630 | 318,496 | 10,281 | 792,518 | 719,162 | 196,919 | 12,876 |
| 2002 | 1,148,480 | 1,121,510 | 324,413 | 9,812 | 826,840 | 777,016 | 210,966 | 12,865 |
| 2003 | 1,124,710 | 1,110,720 | 316,221 | 9,043 | 825,474 | 802,516 | 211,567 | 12,515 |
| 2004 | 1,046,300 | 1,024,300 | 304,265 | 8,256 | 801,338 | 769,480 | 215,587 | 11,985 |
| 2005 | 929,101 | 911,637 | 272,196 | 7,602 | 727,157 | 701,947 | 213,810 | 11,370 |
| 2006 | 806,650 | 790,942 | 230,281 | 7,023 | 629,130 | 605,741 | 189,101 | 10,217 |
| 2007 | 703,317 | 681,507 | 196,128 | 6,482 | 556,424 | 518,554 | 160,504 | 9,330 |
| 2008 | 667,841 | 609,888 | 171,337 | 6,100 | 566,147 | 466,989 | 137,495 | 8,453 |
| 2009 | 723,102 | 696,689 | 156,561 | 6,034 | 635,332 | 585,479 | 127,669 | 8,726 |
| 2010 | 841,372 | 750,699 | 168,038 | 6,599 | 785,604 | 665,880 | 139,011 | 9,933 |
| 2011 | 1,057,930 | 1,037,380 | 213,789 | 8,274 | 949,486 | 925,047 | 185,652 | 12,560 |
| 2012 | 1,197,200 | 1,135,900 | 254,174 | 11,074 | 1,033,000 | 944,495 | 224,467 | 15,964 |
| 2013 | 1,324,910 | 1,241,520 | 304,236 | 14,942 | 1,080,380 | 978,122 | 258,446 | 19,443 |
| 2014 | 1,494,390 | 1,445,130 | 348,402 | 19,718 | 1,136,650 | 1,088,360 | 273,303 | 22,390 |
| 2015 | 1,666,970 | 1,585,980 | 401,573 | 25,678 | 1,185,890 | 1,101,870 | 284,191 | 25,368 |
| 2016 | 1,831,620 | 1,817,980 | 466,000 | 30,739 | 1,324,040 | 1,308,360 | 337,455 | 31,215 |
| 2017 |  |  |  |  | 1,255,550 | 1,233,720 | 326,592 | 35,425 |

Table 2.28-Time series of age 0 recruitment ( 1000 s of fish), with standard deviations, as estimated by the final models in last year's and this year's assessments.

|  | Last year's values |  | This year's values |  |
| ---: | ---: | ---: | ---: | ---: |
| Year | Recruits | Std. dev. | Recruits | Std. dev. |
| 1977 | $1,756,770$ | 193,401 | $1,144,750$ | 291,512 |
| 1978 | 735,150 | 154,302 | 728,494 | 206,031 |
| 1979 | 911,872 | 95,777 | 727,199 | 136,045 |
| 1980 | 316,193 | 42,086 | 338,565 | 59,776 |
| 1981 | 186,119 | 27,168 | 185,901 | 34,070 |
| 1982 | $1,199,810$ | 47,679 | 982,656 | 119,243 |
| 1983 | 270,491 | 30,591 | 251,708 | 42,442 |
| 1984 | 975,814 | 42,279 | 966,891 | 112,849 |
| 1985 | 420,228 | 29,856 | 367,518 | 50,174 |
| 1986 | 206,951 | 19,138 | 243,353 | 34,115 |
| 1987 | 148,111 | 15,608 | 101,666 | 20,766 |
| 1988 | 367,125 | 20,676 | 277,455 | 37,886 |
| 1989 | 769,057 | 30,920 | 763,595 | 86,361 |
| 1990 | 627,254 | 27,833 | 625,925 | 71,589 |
| 1991 | 338,528 | 20,762 | 415,703 | 51,114 |
| 1992 | 836,942 | 26,123 | 928,333 | 95,226 |
| 1993 | 295,068 | 16,994 | 368,560 | 43,036 |
| 1994 | 322,162 | 16,048 | 319,622 | 37,706 |
| 1995 | 349,179 | 18,802 | 289,947 | 35,339 |
| 1996 | 858,353 | 26,888 | 798,434 | 85,113 |
| 1997 | 350,987 | 17,271 | 375,721 | 43,067 |
| 1998 | 343,807 | 16,119 | 348,699 | 38,932 |
| 1999 | 662,988 | 19,273 | 727,777 | 72,993 |
| 2000 | 406,171 | 13,814 | 556,135 | 56,753 |
| 2001 | 187,567 | 10,495 | 246,470 | 27,034 |
| 2002 | 320,855 | 11,944 | 332,437 | 34,816 |
| 2003 | 252,583 | 11,704 | 279,904 | 29,449 |
| 2004 | 221,685 | 11,169 | 234,562 | 25,592 |
| 2005 | 252,378 | 12,323 | 317,928 | 34,223 |
| 2006 | 886,192 | 27,543 | $1,023,420$ | 106,295 |
| 2007 | 271,215 | 17,279 | 447,919 | 51,147 |
| 2008 | $1,405,850$ | 53,534 | $1,420,070$ | 153,244 |
| 2009 | 237,580 | 24,864 | 183,956 | 27,608 |
| 2010 | 837,777 | 48,260 | 856,424 | 96,804 |
| 2011 | $1,248,580$ | 78,359 | $1,269,710$ | 147,292 |
| 2012 | 646,490 | 59,830 | 528,928 | 66,385 |
| 2013 | $1,261,180$ | 112,033 | $1,200,650$ | 141,608 |
| 2014 | 159,532 | 40,506 | 168,227 | 29,891 |
| 2015 |  |  | 197,947 | 44,945 |
| Average | 574,858 |  | 552,389 |  |
|  |  |  |  |  |
|  |  |  |  |  |

Table 2.29—Numbers (1000s) at age as estimated by Model 16.6.

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 1144750 | 219490 | 62271 | 46037 | 27320 | 12376 | 7388 | 4826 | 3235 | 2159 | 1417 | 909 | 572 | 354 | 216 | 131 | 79 | 47 | 28 | 17 | 25 |
| 1978 | 728494 | 796658 | 152741 | 43308 | 31778 | 17648 | 7002 | 4041 | 2631 | 1763 | 1177 | 772 | 496 | 312 | 193 | 118 | 71 | 43 | 26 | 15 | 23 |
| 1979 | 727199 | 506976 | 554378 | 106206 | 29825 | 20099 | 9415 | 3577 | 2056 | 1338 | 897 | 599 | 393 | 252 | 159 | 98 | 60 | 36 | 22 | 13 | 19 |
| 1980 | 338565 | 506075 | 352800 | 385551 | 73311 | 19264 | 11367 | 5147 | 1950 | 1120 | 729 | 489 | 326 | 214 | 137 | 86 | 53 | 33 | 20 | 12 | 18 |
| 1981 | 185901 | 235616 | 352171 | 245341 | 265873 | 46930 | 10628 | 6038 | 2725 | 1032 | 593 | 386 | 259 | 173 | 113 | 73 | 46 | 28 | 17 | 10 | 16 |
| 1982 | 982656 | 129373 | 163965 | 244967 | 169727 | 175244 | 28081 | 6205 | 3517 | 1587 | 601 | 345 | 225 | 151 | 101 | 66 | 42 | 27 | 16 | 10 | 15 |
| 1983 | 251708 | 683856 | 90032 | 114078 | 169935 | 114715 | 112439 | 17780 | 3924 | 2224 | 1003 | 380 | 218 | 142 | 95 | 64 | 42 | 27 | 17 | 10 | 16 |
| 1984 | 966891 | 175170 | 475902 | 62637 | 79097 | 114337 | 72682 | 70159 | 11079 | 2445 | 1386 | 625 | 237 | 136 | 89 | 59 | 40 | 26 | 17 | 10 | 16 |
| 1985 | 367518 | 672884 | 121902 | 331058 | 43372 | 52569 | 70011 | 43586 | 41992 | 6630 | 1463 | 829 | 374 | 142 | 81 | 53 | 36 | 24 | 16 | 10 | 16 |
| 1986 | 243353 | 255765 | 468261 | 84796 | 229106 | 28679 | 31735 | 41296 | 25654 | 24712 | 3901 | 861 | 488 | 220 | 83 | 48 | 31 | 21 | 14 | 9 | 15 |
| 1987 | 101666 | 169355 | 177987 | 325726 | 58679 | 151402 | 17284 | 18684 | 24260 | 15068 | 14514 | 2292 | 506 | 287 | 129 | 49 | 28 | 18 | 12 | 8 | 14 |
| 1988 | 277455 | 70752 | 117854 | 123805 | 225313 | 38638 | 90344 | 10059 | 10848 | 14082 | 8747 | 8425 | 1330 | 294 | 166 | 75 | 28 | 16 | 11 | 7 | 13 |
| 1989 | 763595 | 193088 | 49236 | 81964 | 85466 | 145639 | 21899 | 49519 | 5496 | 5926 | 7693 | 4778 | 4602 | 727 | 160 | 91 | 41 | 16 | 9 | 6 | 1 |
| 1990 | 625925 | 531405 | 134369 | 34245 | 56653 | 55880 | 85212 | 12456 | 28092 | 3117 | 3361 | 4363 | 2710 | 2610 | 412 | 91 | 52 | 23 | 9 | 5 | 10 |
| 1991 | 415703 | 435596 | 369801 | 93453 | 23651 | 36763 | 32017 | 47300 | 6894 | 15544 | 1725 | 1860 | 2414 | 1499 | 1444 | 228 | 50 | 29 | 13 | 5 | 8 |
| 1992 | 928333 | 289297 | 303117 | 257073 | 64166 | 14553 | 18170 | 14966 | 21996 | 3204 | 7225 | 802 | 864 | 1122 | 697 | 671 | 106 | 23 | 13 | 6 | 6 |
| 1993 | 368560 | 646047 | 201309 | 210669 | 176017 | 38487 | 6699 | 7820 | 6401 | 9403 | 1370 | 3088 | 343 | 369 | 480 | 298 | 287 | 45 | 10 | 6 | 5 |
| 1994 | 319622 | 256489 | 449566 | 139955 | 144795 | 109318 | 19519 | 3227 | 3749 | 3067 | 4506 | 656 | 1480 | 164 | 177 | 230 | 143 | 138 | 22 | 5 | 5 |
| 1995 | 289947 | 222432 | 178483 | 312524 | 96099 | 89137 | 54097 | 9137 | 1503 | 1745 | 1428 | 2097 | 306 | 689 | 76 | 82 | 107 | 66 | 64 | 10 | 5 |
| 1996 | 798434 | 201780 | 154780 | 124040 | 213832 | 57267 | 40298 | 22797 | 3825 | 629 | 730 | 597 | 878 | 128 | 288 | 32 | 34 | 45 | 28 | 27 | 6 |
| 1997 | 375721 | 555647 | 140410 | 107578 | 84979 | 128936 | 26753 | 17642 | 9920 | 1664 | 274 | 318 | 260 | 382 | 56 | 125 | 14 | 15 | 19 | 12 | 14 |
| 1998 | 348699 | 261472 | 386647 | 97578 | 73582 | 50492 | 57819 | 11168 | 7316 | 4112 | 690 | 113 | 132 | 108 | 158 | 23 | 52 | 6 | 6 | 8 | 11 |
| 1999 | 727777 | 242668 | 181949 | 268775 | 66971 | 45107 | 24695 | 26699 | 5130 | 3359 | 1888 | 317 | 52 | 60 | 49 | 73 | 11 | 24 | 3 | 3 | 9 |
| 2000 | 556135 | 506476 | 168864 | 126478 | 184416 | 40947 | 21902 | 11308 | 12159 | 2335 | 1529 | 859 | 144 | 24 | 28 | 23 | 33 | 5 | 11 | 1 | 5 |
| 2001 | 246470 | 387027 | 352439 | 117387 | 86828 | 113316 | 20158 | 10191 | 5234 | 5626 | 1080 | 707 | 398 | 67 | 11 | 13 | 10 | 15 | 2 | 5 | 3 |
| 2002 | 332437 | 171524 | 269323 | 245055 | 80807 | 54700 | 59791 | 10167 | 5119 | 2628 | 2825 | 542 | 355 | 200 | 33 | 6 | 6 | 5 | 8 | 1 |  |
| 2003 | 279904 | 231350 | 119358 | 187231 | 168325 | 49904 | 27309 | 28278 | 4785 | 2408 | 1236 | 1329 | 255 | 167 | 94 | 16 | 3 | 3 | 2 | 4 | 2 |
| 2004 | 234562 | 194791 | 160988 | 82970 | 128474 | 102979 | 24272 | 12529 | 12904 | 2182 | 1098 | 564 | 606 | 116 | 76 | 43 | 7 | 1 | 1 | 1 | 3 |
| 2005 | 317928 | 163237 | 135549 | 111914 | 56974 | 79122 | 51019 | 11377 | 5843 | 6015 | 1017 | 512 | 263 | 282 | 54 | 36 | 20 | 3 | 1 | 1 | 2 |
| 2006 | 1023420 | 221253 | 113591 | 94228 | 76826 | 34990 | 38897 | 23700 | 5258 | 2699 | 2778 | 470 | 236 | 121 | 130 | 25 | 16 | 9 | 2 | 0 | 1 |
| 2007 | 447919 | 712218 | 153961 | 78951 | 64558 | 46345 | 16366 | 17053 | 10328 | 2290 | 1176 | 1210 | 205 | 103 | 53 | 57 | 11 | 7 | 4 | 1 | 1 |
| 2008 | 1420070 | 311717 | 495604 | 107013 | 54116 | 39109 | 21934 | 7274 | 7535 | 4562 | 1011 | 519 | 534 | 90 | 45 | 23 | 25 | 5 | 3 | 2 | 1 |
| 2009 | 183956 | 988255 | 216906 | 344380 | 73092 | 31740 | 16917 | 8780 | 2891 | 2993 | 1812 | 402 | 206 | 212 | 36 | 18 | 9 | 10 | 2 | 1 | 1 |
| 2010 | 856424 | 128019 | 687652 | 150670 | 234223 | 41241 | 12327 | 5975 | 3074 | 1011 | 1047 | 634 | 141 | 72 | 74 | 13 | 6 | 3 | 3 | 1 | 1 |
| 2011 | 1269710 | 596003 | 89082 | 477875 | 103032 | 138870 | 18384 | 5110 | 2460 | 1265 | 416 | 431 | 261 | 58 | 30 | 31 | 5 | 3 | 1 | 1 | 1 |
| 2012 | 528928 | 883616 | 414728 | 61905 | 326699 | 60941 | 61490 | 7562 | 2088 | 1005 | 516 | 170 | 176 | 106 | 24 | 12 | 12 | 2 | 1 | 1 | 1 |
| 2013 | 1200650 | 368092 | 614867 | 288232 | 42373 | 195419 | 27841 | 26230 | 3205 | 884 | 426 | 219 | 72 | 75 | 45 | 10 | 5 | 5 | 1 | 0 | 1 |
| 2014 | 168227 | 835560 | 256143 | 427434 | 197899 | 26067 | 96518 | 13004 | 12188 | 1489 | 411 | 198 | 102 | 33 | 35 | 21 | 5 | 2 | 2 | 0 | 0 |
| 2015 | 197947 | 117073 | 581433 | 178038 | 292994 | 119937 | 12351 | 42954 | 5754 | 5390 | 658 | 182 | 87 | 45 | 15 | 15 | 9 | 2 | 1 | 1 | 0 |
| 2016 | 551005 | 137756 | 81467 | 404206 | 122293 | 180956 | 59890 | 5843 | 20218 | 2707 | 2536 | 310 | 85 | 41 | 21 | 7 | 7 | 4 | 1 | 0 | 1 |

Table 2.30—Model 16.6 estimates of "effective" fishing mortality $\left(=-\ln \left(N_{a+1, t+1} / N_{a, t}\right)-M\right)$ at age and year.

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 0.000 | 0.001 | 0.008 | 0.075 | 0.207 | 0.241 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 |
| 1978 | 0.000 | 0.001 | 0.010 | 0.096 | 0.266 | 0.309 | 0.313 | 0.313 | 0.313 | 0.313 | 0.313 | 0.313 | 0.313 | 0.313 | 0.313 | 0.313 | 0.313 | 0.313 | 0.313 |
| 1979 | 0.000 | 0.001 | 0.008 | 0.075 | 0.207 | 0.241 | 0.244 | 0.245 | 0.245 | 0.245 | 0.245 | 0.245 | 0.245 | 0.245 | 0.245 | 0.245 | 0.245 | 0.245 | 0.245 |
| 1980 | 0.000 | 0.001 | 0.009 | 0.084 | 0.232 | 0.270 | 0.274 | 0.274 | 0.274 | 0.274 | 0.274 | 0.274 | 0.274 | 0.274 | 0.274 | 0.274 | 0.274 | 0.274 | 0.274 |
| 1981 | 0.000 | 0.000 | 0.006 | 0.054 | 0.151 | 0.176 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 | 0.178 |
| 1982 | 0.000 | 0.000 | 0.003 | 0.029 | 0.081 | 0.095 | 0.096 | 0.096 | 0.096 | 0.096 | 0.096 | 0.096 | 0.096 | 0.096 | 0.096 | 0.096 | 0.096 | 0.096 | 0.096 |
| 1983 | 0.000 | 0.000 | 0.004 | 0.034 | 0.094 | 0.109 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 | 0.111 |
| 1984 | 0.000 | 0.000 | 0.005 | 0.046 | 0.128 | 0.149 | 0.151 | 0.151 | 0.151 | 0.151 | 0.151 | 0.151 | 0.151 | 0.151 | 0.151 | 0.151 | 0.151 | 0.151 | 0.151 |
| 1985 | 0.000 | 0.000 | 0.006 | 0.051 | 0.142 | 0.165 | 0.168 | 0.168 | 0.168 | 0.168 | 0.168 | 0.168 | 0.168 | 0.168 | 0.168 | 0.168 | 0.168 | 0.168 | 0.168 |
| 1986 | 0.000 | 0.000 | 0.006 | 0.052 | 0.144 | 0.167 | 0.169 | 0.170 | 0.170 | 0.170 | 0.170 | 0.170 | 0.170 | 0.170 | 0.170 | 0.170 | 0.170 | 0.170 | 0.170 |
| 1987 | 0.000 | 0.000 | 0.006 | 0.055 | 0.154 | 0.179 | 0.181 | 0.181 | 0.181 | 0.181 | 0.181 | 0.181 | 0.181 | 0.181 | 0.181 | 0.181 | 0.181 | 0.181 | 0.181 |
| 1988 | 0.000 | 0.001 | 0.008 | 0.074 | 0.205 | 0.239 | 0.242 | 0.242 | 0.242 | 0.242 | 0.242 | 0.242 | 0.242 | 0.242 | 0.242 | 0.242 | 0.242 | 0.242 | 0.242 |
| 1989 | 0.000 | 0.001 | 0.007 | 0.062 | 0.173 | 0.202 | 0.204 | 0.205 | 0.205 | 0.205 | 0.205 | 0.205 | 0.205 | 0.205 | 0.205 | 0.205 | 0.205 | 0.205 | 0.205 |
| 1990 | 0.000 | 0.001 | 0.008 | 0.070 | 0.194 | 0.226 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 |
| 1991 | 0.000 | 0.001 | 0.013 | 0.123 | 0.342 | 0.398 | 0.403 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 |
| 1992 | 0.000 | 0.001 | 0.016 | 0.149 | 0.413 | 0.481 | 0.487 | 0.487 | 0.487 | 0.487 | 0.487 | 0.487 | 0.487 | 0.487 | 0.487 | 0.487 | 0.487 | 0.487 | 0.487 |
| 1993 | 0.000 | 0.001 | 0.012 | 0.114 | 0.316 | 0.368 | 0.373 | 0.373 | 0.373 | 0.373 | 0.373 | 0.373 | 0.373 | 0.373 | 0.373 | 0.373 | 0.373 | 0.373 | 0.373 |
| 1994 | 0.000 | 0.001 | 0.013 | 0.123 | 0.341 | 0.397 | 0.402 | 0.402 | 0.402 | 0.402 | 0.402 | 0.402 | 0.402 | 0.402 | 0.402 | 0.402 | 0.402 | 0.402 | 0.402 |
| 1995 | 0.000 | 0.001 | 0.017 | 0.155 | 0.431 | 0.502 | 0.508 | 0.509 | 0.509 | 0.509 | 0.509 | 0.509 | 0.509 | 0.509 | 0.509 | 0.509 | 0.509 | 0.509 | 0.509 |
| 1996 | 0.000 | 0.001 | 0.016 | 0.143 | 0.399 | 0.464 | 0.470 | 0.470 | 0.470 | 0.470 | 0.470 | 0.470 | 0.470 | 0.470 | 0.470 | 0.470 | 0.470 | 0.470 | 0.470 |
| 1997 | 0.000 | 0.001 | 0.017 | 0.158 | 0.439 | 0.511 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 |
| 1998 | 0.000 | 0.001 | 0.014 | 0.127 | 0.353 | 0.410 | 0.416 | 0.416 | 0.416 | 0.416 | 0.416 | 0.416 | 0.416 | 0.416 | 0.416 | 0.416 | 0.416 | 0.416 | 0.416 |
| 1999 | 0.000 | 0.001 | 0.014 | 0.129 | 0.360 | 0.419 | 0.424 | 0.424 | 0.425 | 0.425 | 0.425 | 0.425 | 0.425 | 0.425 | 0.425 | 0.425 | 0.425 | 0.425 | 0.425 |
| 2000 | 0.000 | 0.001 | 0.014 | 0.125 | 0.346 | 0.403 | 0.408 | 0.408 | 0.408 | 0.408 | 0.408 | 0.408 | 0.408 | 0.408 | 0.408 | 0.408 | 0.408 | 0.408 | 0.408 |
| 2001 | 0.000 | 0.001 | 0.011 | 0.100 | 0.277 | 0.322 | 0.326 | 0.326 | 0.326 | 0.326 | 0.326 | 0.326 | 0.326 | 0.326 | 0.326 | 0.326 | 0.326 | 0.326 | 0.326 |
| 2002 | 0.000 | 0.001 | 0.013 | 0.119 | 0.332 | 0.386 | 0.391 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 | 0.392 |
| 2003 | 0.000 | 0.001 | 0.014 | 0.129 | 0.358 | 0.417 | 0.422 | 0.422 | 0.423 | 0.423 | 0.423 | 0.423 | 0.423 | 0.423 | 0.423 | 0.423 | 0.423 | 0.423 | 0.423 |
| 2004 | 0.000 | 0.001 | 0.013 | 0.122 | 0.340 | 0.395 | 0.400 | 0.401 | 0.401 | 0.401 | 0.401 | 0.401 | 0.401 | 0.401 | 0.401 | 0.401 | 0.401 | 0.401 | 0.401 |
| 2005 | 0.000 | 0.001 | 0.014 | 0.125 | 0.348 | 0.404 | 0.409 | 0.410 | 0.410 | 0.410 | 0.410 | 0.410 | 0.410 | 0.410 | 0.410 | 0.410 | 0.410 | 0.410 | 0.410 |
| 2006 | 0.000 | 0.001 | 0.016 | 0.143 | 0.397 | 0.462 | 0.468 | 0.469 | 0.469 | 0.469 | 0.469 | 0.469 | 0.469 | 0.469 | 0.469 | 0.469 | 0.469 | 0.469 | 0.469 |
| 2007 | 0.000 | 0.001 | 0.015 | 0.139 | 0.386 | 0.448 | 0.454 | 0.455 | 0.455 | 0.455 | 0.455 | 0.455 | 0.455 | 0.455 | 0.455 | 0.455 | 0.455 | 0.455 | 0.455 |
| 2008 | 0.000 | 0.002 | 0.019 | 0.171 | 0.476 | 0.553 | 0.560 | 0.561 | 0.561 | 0.561 | 0.561 | 0.561 | 0.561 | 0.561 | 0.561 | 0.561 | 0.561 | 0.561 | 0.561 |
| 2009 | 0.000 | 0.002 | 0.023 | 0.210 | 0.583 | 0.678 | 0.687 | 0.688 | 0.688 | 0.688 | 0.688 | 0.688 | 0.688 | 0.688 | 0.688 | 0.688 | 0.688 | 0.688 | 0.688 |
| 2010 | 0.000 | 0.001 | 0.018 | 0.160 | 0.445 | 0.518 | 0.525 | 0.525 | 0.525 | 0.525 | 0.525 | 0.525 | 0.525 | 0.525 | 0.525 | 0.525 | 0.525 | 0.525 | 0.525 |
| 2011 | 0.000 | 0.001 | 0.018 | 0.163 | 0.452 | 0.526 | 0.533 | 0.533 | 0.533 | 0.533 | 0.533 | 0.533 | 0.533 | 0.533 | 0.533 | 0.533 | 0.533 | 0.533 | 0.533 |
| 2012 | 0.000 | 0.001 | 0.017 | 0.151 | 0.421 | 0.489 | 0.496 | 0.496 | 0.496 | 0.496 | 0.496 | 0.496 | 0.496 | 0.496 | 0.496 | 0.496 | 0.496 | 0.496 | 0.496 |
| 2013 | 0.000 | 0.001 | 0.014 | 0.123 | 0.343 | 0.399 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 | 0.404 |
| 2014 | 0.000 | 0.001 | 0.015 | 0.138 | 0.384 | 0.447 | 0.453 | 0.453 | 0.453 | 0.453 | 0.453 | 0.453 | 0.453 | 0.453 | 0.453 | 0.453 | 0.453 | 0.453 | 0.453 |
| 2015 | 0.000 | 0.001 | 0.013 | 0.119 | 0.332 | 0.386 | 0.391 | 0.391 | 0.391 | 0.391 | 0.391 | 0.391 | 0.391 | 0.391 | 0.391 | 0.391 | 0.391 | 0.391 | 0.391 |
| 2016 | 0.000 | 0.001 | 0.011 | 0.105 | 0.291 | 0.338 | 0.343 | 0.343 | 0.343 | 0.343 | 0.343 | 0.343 | 0.343 | 0.343 | 0.343 | 0.343 | 0.343 | 0.343 | 0.343 |

Table 2.31—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=\max F_{A B C}$ in 2017-2029 (Scenario 1), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 239,000 | 239,000 | 239,000 | 239,000 | 0 |
| 2018 | 246,000 | 246,000 | 246,000 | 246,000 | 0 |
| 2019 | 195,000 | 195,000 | 195,000 | 195,000 | 1 |
| 2020 | 157,000 | 158,000 | 158,000 | 158,000 | 166 |
| 2021 | 141,000 | 144,000 | 145,000 | 153,000 | 4,118 |
| 2022 | 131,000 | 148,000 | 153,000 | 187,000 | 18,899 |
| 2023 | 108,000 | 151,000 | 160,000 | 236,000 | 42,541 |
| 2024 | 94,000 | 157,000 | 163,000 | 262,000 | 52,774 |
| 2025 | 87,000 | 161,000 | 166,000 | 258,000 | 56,866 |
| 2026 | 85,100 | 165,000 | 170,000 | 272,000 | 59,397 |
| 2027 | 85,300 | 166,000 | 170,000 | 272,000 | 59,395 |
| 2028 | 84,600 | 165,000 | 170,000 | 270,000 | 58,033 |
| 2029 | 87,600 | 165,000 | 169,000 | 268,000 | 56,955 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 324,000 | 324,000 | 324,000 | 324,000 | 0 |
| 2018 | 325,000 | 325,000 | 325,000 | 325,000 | 0 |
| 2019 | 289,000 | 289,000 | 289,000 | 289,000 | 45 |
| 2020 | 247,000 | 248,000 | 248,000 | 250,000 | 847 |
| 2021 | 224,000 | 228,000 | 230,000 | 240,000 | 5,514 |
| 2022 | 213,000 | 228,000 | 232,000 | 266,000 | 17,850 |
| 2023 | 199,000 | 231,000 | 239,000 | 306,000 | 35,836 |
| 2024 | 188,000 | 236,000 | 245,000 | 345,000 | 49,193 |
| 2025 | 180,000 | 239,000 | 250,000 | 353,000 | 56,361 |
| 2026 | 178,000 | 240,000 | 254,000 | 361,000 | 60,190 |
| 2027 | 179,000 | 242,000 | 255,000 | 369,000 | 61,330 |
| 2028 | 178,000 | 242,000 | 255,000 | 373,000 | 60,156 |
| 2029 | 181,000 | 243,000 | 254,000 | 371,000 | 58,630 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2018 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2019 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2020 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2021 | 0.28 | 0.28 | 0.29 | 0.30 | 0.01 |
| 2022 | 0.26 | 0.28 | 0.29 | 0.31 | 0.02 |
| 2023 | 0.25 | 0.29 | 0.29 | 0.31 | 0.02 |
| 2024 | 0.23 | 0.29 | 0.28 | 0.31 | 0.03 |
| 2025 | 0.22 | 0.30 | 0.28 | 0.31 | 0.03 |
| 2026 | 0.22 | 0.30 | 0.28 | 0.31 | 0.03 |
| 2027 | 0.22 | 0.30 | 0.29 | 0.31 | 0.03 |
| 2028 | 0.22 | 0.30 | 0.29 | 0.31 | 0.03 |
| 2029 | 0.22 | 0.30 | 0.29 | 0.31 | 0.03 |

Table 2.32—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that catches in 2017-2018 are less than ABC by amounts predicted from past performance, but that $F=\max F_{A B C}$ in 2019-2029 (Scenario 2), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 203,000 | 203,000 | 203,000 | 203,000 | 0 |
| 2018 | 212,000 | 212,000 | 212,000 | 212,000 | 0 |
| 2019 | 210,000 | 210,000 | 210,000 | 210,000 | 1 |
| 2020 | 167,000 | 168,000 | 168,000 | 168,000 | 49 |
| 2021 | 152,000 | 155,000 | 156,000 | 164,000 | 3,633 |
| 2022 | 135,000 | 152,000 | 157,000 | 190,000 | 18,453 |
| 2023 | 109,000 | 152,000 | 161,000 | 237,000 | 42,525 |
| 2024 | 94,200 | 158,000 | 163,000 | 262,000 | 52,881 |
| 2025 | 86,900 | 161,000 | 166,000 | 259,000 | 56,949 |
| 2026 | 85,000 | 165,000 | 170,000 | 273,000 | 59,441 |
| 2027 | 85,200 | 166,000 | 170,000 | 272,000 | 59,414 |
| 2028 | 84,600 | 165,000 | 170,000 | 270,000 | 58,039 |
| 2029 | 87,600 | 165,000 | 169,000 | 268,000 | 56,957 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 327,000 | 327,000 | 327,000 | 327,000 | 0 |
| 2018 | 340,000 | 340,000 | 340,000 | 340,000 | 0 |
| 2019 | 313,000 | 313,000 | 313,000 | 313,000 | 45 |
| 2020 | 264,000 | 265,000 | 265,000 | 266,000 | 854 |
| 2021 | 234,000 | 238,000 | 240,000 | 250,000 | 5,585 |
| 2022 | 217,000 | 232,000 | 236,000 | 271,000 | 18,113 |
| 2023 | 201,000 | 232,000 | 241,000 | 309,000 | 36,201 |
| 2024 | 188,000 | 236,000 | 246,000 | 347,000 | 49,478 |
| 2025 | 180,000 | 238,000 | 250,000 | 354,000 | 56,520 |
| 2026 | 178,000 | 240,000 | 254,000 | 361,000 | 60,259 |
| 2027 | 179,000 | 242,000 | 255,000 | 369,000 | 61,354 |
| 2028 | 178,000 | 242,000 | 255,000 | 373,000 | 60,162 |
| 2029 | 181,000 | 243,000 | 254,000 | 371,000 | 58,630 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.26 | 0.26 | 0.26 | 0.26 | 0.00 |
| 2018 | 0.25 | 0.25 | 0.25 | 0.25 | 0.00 |
| 2019 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2020 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2021 | 0.29 | 0.30 | 0.30 | 0.31 | 0.01 |
| 2022 | 0.27 | 0.29 | 0.29 | 0.31 | 0.01 |
| 2023 | 0.25 | 0.29 | 0.29 | 0.31 | 0.02 |
| 2024 | 0.23 | 0.29 | 0.28 | 0.31 | 0.03 |
| 2025 | 0.22 | 0.30 | 0.28 | 0.31 | 0.03 |
| 2026 | 0.22 | 0.30 | 0.28 | 0.31 | 0.03 |
| 2027 | 0.22 | 0.30 | 0.29 | 0.31 | 0.03 |
| 2028 | 0.22 | 0.30 | 0.29 | 0.31 | 0.03 |
| 2029 | 0.22 | 0.30 | 0.29 | 0.31 | 0.03 |

Table 2.33—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that the upper bound on $F_{A B C}$ is set the most recent five-year average fishing mortality rate in 2017-2029 (Scenario 3), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 312,000 | 312,000 | 312,000 | 312,000 | 0 |
| 2018 | 298,000 | 298,000 | 298,000 | 298,000 | 0 |
| 2019 | 217,000 | 217,000 | 217,000 | 217,000 | 1 |
| 2020 | 169,000 | 169,000 | 169,000 | 169,000 | 67 |
| 2021 | 168,000 | 169,000 | 169,000 | 172,000 | 1,593 |
| 2022 | 159,000 | 173,000 | 178,000 | 211,000 | 18,085 |
| 2023 | 132,000 | 173,000 | 184,000 | 272,000 | 46,616 |
| 2024 | 119,000 | 176,000 | 187,000 | 298,000 | 56,860 |
| 2025 | 112,000 | 179,000 | 190,000 | 293,000 | 60,389 |
| 2026 | 110,000 | 180,000 | 192,000 | 308,000 | 63,334 |
| 2027 | 110,000 | 182,000 | 192,000 | 308,000 | 62,855 |
| 2028 | 110,000 | 182,000 | 191,000 | 307,000 | 60,791 |
| 2029 | 111,000 | 180,000 | 190,000 | 301,000 | 59,621 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 319,000 | 319,000 | 319,000 | 319,000 | 0 |
| 2018 | 295,000 | 295,000 | 295,000 | 295,000 | 0 |
| 2019 | 243,000 | 243,000 | 243,000 | 243,000 | 45 |
| 2020 | 197,000 | 198,000 | 198,000 | 200,000 | 854 |
| 2021 | 176,000 | 181,000 | 182,000 | 193,000 | 5,754 |
| 2022 | 163,000 | 179,000 | 184,000 | 220,000 | 19,144 |
| 2023 | 145,000 | 180,000 | 189,000 | 259,000 | 37,552 |
| 2024 | 132,000 | 185,000 | 193,000 | 290,000 | 49,457 |
| 2025 | 122,000 | 187,000 | 196,000 | 291,000 | 55,061 |
| 2026 | 120,000 | 189,000 | 198,000 | 303,000 | 57,808 |
| 2027 | 121,000 | 190,000 | 199,000 | 303,000 | 58,061 |
| 2028 | 122,000 | 189,000 | 198,000 | 307,000 | 56,577 |
| 2029 | 122,000 | 188,000 | 197,000 | 304,000 | 55,334 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |
| 2018 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |
| 2019 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |
| 2020 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |
| 2021 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |
| 2022 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |
| 2023 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |
| 2024 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |
| 2025 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |
| 2026 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |
| 2027 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |
| 2028 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |
| 2029 | 0.42 | 0.42 | 0.42 | 0.42 | 0.00 |

Table 2.34—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that the upper bound on $F_{A B C}$ is set at $F_{60 \%}$ in 2017-2029 (Scenario 4), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 120,000 | 120,000 | 120,000 | 120,000 | 0 |
| 2018 | 138,000 | 138,000 | 138,000 | 138,000 | 0 |
| 2019 | 124,000 | 124,000 | 124,000 | 124,000 | 0 |
| 2020 | 109,000 | 109,000 | 109,000 | 109,000 | 23 |
| 2021 | 108,000 | 108,000 | 108,000 | 109,000 | 555 |
| 2022 | 104,000 | 109,000 | 111,000 | 123,000 | 6,578 |
| 2023 | 92,500 | 109,000 | 114,000 | 151,000 | 19,329 |
| 2024 | 84,800 | 110,000 | 116,000 | 168,000 | 26,226 |
| 2025 | 79,600 | 113,000 | 118,000 | 171,000 | 29,776 |
| 2026 | 76,500 | 115,000 | 120,000 | 173,000 | 32,245 |
| 2027 | 76,300 | 116,000 | 121,000 | 180,000 | 33,101 |
| 2028 | 75,800 | 116,000 | 121,000 | 179,000 | 32,658 |
| 2029 | 76,700 | 116,000 | 121,000 | 179,000 | 32,006 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 332,000 | 332,000 | 332,000 | 332,000 | 0 |
| 2018 | 375,000 | 375,000 | 375,000 | 375,000 | 0 |
| 2019 | 377,000 | 377,000 | 377,000 | 377,000 | 45 |
| 2020 | 354,000 | 355,000 | 355,000 | 356,000 | 854 |
| 2021 | 333,000 | 337,000 | 339,000 | 349,000 | 5,767 |
| 2022 | 315,000 | 331,000 | 336,000 | 374,000 | 19,573 |
| 2023 | 294,000 | 331,000 | 341,000 | 421,000 | 41,776 |
| 2024 | 272,000 | 337,000 | 348,000 | 475,000 | 62,665 |
| 2025 | 255,000 | 342,000 | 355,000 | 499,000 | 77,093 |
| 2026 | 244,000 | 346,000 | 360,000 | 506,000 | 85,916 |
| 2027 | 241,000 | 351,000 | 364,000 | 526,000 | 90,325 |
| 2028 | 239,000 | 353,000 | 365,000 | 533,000 | 90,966 |
| 2029 | 237,000 | 353,000 | 365,000 | 532,000 | 89,731 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2018 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2019 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2020 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2021 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2022 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2023 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2024 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2025 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2026 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2027 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2028 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2029 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |

Table 2.35—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=0$ in 2017-2029 (Scenario 5), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 0 | 0 | 0 | 0 | 0 |
| 2023 | 0 | 0 | 0 | 0 | 0 |
| 2024 | 0 | 0 | 0 | 0 | 0 |
| 2025 | 0 | 0 | 0 | 0 | 0 |
| 2026 | 0 | 0 | 0 | 0 | 0 |
| 2027 | 0 | 0 | 0 | 0 | 0 |
| 2028 | 0 | 0 | 0 | 0 | 0 |
| 2029 | 0 | 0 | 0 | 0 | 0 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 339,000 | 339,000 | 339,000 | 339,000 | 0 |
| 2018 | 427,000 | 427,000 | 427,000 | 427,000 | 0 |
| 2019 | 481,000 | 481,000 | 481,000 | 481,000 | 45 |
| 2020 | 498,000 | 499,000 | 499,000 | 500,000 | 854 |
| 2021 | 499,000 | 504,000 | 505,000 | 516,000 | 5,774 |
| 2022 | 495,000 | 512,000 | 517,000 | 555,000 | 19,804 |
| 2023 | 483,000 | 522,000 | 533,000 | 618,000 | 44,282 |
| 2024 | 464,000 | 535,000 | 550,000 | 694,000 | 72,096 |
| 2025 | 444,000 | 550,000 | 565,000 | 750,000 | 95,588 |
| 2026 | 427,000 | 561,000 | 578,000 | 782,000 | 112,729 |
| 2027 | 420,000 | 569,000 | 588,000 | 805,000 | 123,962 |
| 2028 | 416,000 | 579,000 | 595,000 | 826,000 | 129,665 |
| 2029 | 414,000 | 586,000 | 600,000 | 841,000 | 131,269 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2018 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2019 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2020 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2021 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2022 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2023 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2024 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2025 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2026 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2027 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2028 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2029 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |

Table 2.36—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=F_{\text {OFL }}$ in 2017-2029 (Scenario 6), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 284,000 | 284,000 | 284,000 | 284,000 | 0 |
| 2018 | 279,000 | 279,000 | 279,000 | 279,000 | 0 |
| 2019 | 210,000 | 210,000 | 210,000 | 210,000 | 1 |
| 2020 | 147,000 | 147,000 | 147,000 | 148,000 | 561 |
| 2021 | 139,000 | 143,000 | 144,000 | 152,000 | 4,582 |
| 2022 | 134,000 | 153,000 | 159,000 | 210,000 | 23,811 |
| 2023 | 110,000 | 157,000 | 170,000 | 264,000 | 51,301 |
| 2024 | 96,600 | 165,000 | 175,000 | 288,000 | 61,456 |
| 2025 | 90,500 | 167,000 | 178,000 | 284,000 | 65,178 |
| 2026 | 88,300 | 169,000 | 180,000 | 298,000 | 67,581 |
| 2027 | 89,000 | 170,000 | 180,000 | 298,000 | 67,273 |
| 2028 | 88,200 | 170,000 | 179,000 | 295,000 | 65,668 |
| 2029 | 90,900 | 169,000 | 178,000 | 294,000 | 64,586 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 321,000 | 321,000 | 321,000 | 321,000 | 0 |
| 2018 | 306,000 | 306,000 | 306,000 | 306,000 | 0 |
| 2019 | 260,000 | 260,000 | 260,000 | 260,000 | 45 |
| 2020 | 217,000 | 218,000 | 218,000 | 219,000 | 813 |
| 2021 | 202,000 | 206,000 | 207,000 | 217,000 | 5,336 |
| 2022 | 194,000 | 209,000 | 213,000 | 246,000 | 17,334 |
| 2023 | 183,000 | 214,000 | 221,000 | 286,000 | 33,613 |
| 2024 | 173,000 | 219,000 | 227,000 | 315,000 | 44,371 |
| 2025 | 166,000 | 220,000 | 230,000 | 319,000 | 49,640 |
| 2026 | 165,000 | 222,000 | 232,000 | 328,000 | 52,446 |
| 2027 | 165,000 | 222,000 | 232,000 | 331,000 | 52,936 |
| 2028 | 164,000 | 222,000 | 231,000 | 331,000 | 51,467 |
| 2029 | 167,000 | 221,000 | 231,000 | 330,000 | 50,099 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.38 | 0.38 | 0.38 | 0.38 | 0.00 |
| 2018 | 0.38 | 0.38 | 0.38 | 0.38 | 0.00 |
| 2019 | 0.38 | 0.38 | 0.38 | 0.38 | 0.00 |
| 2020 | 0.33 | 0.33 | 0.33 | 0.33 | 0.00 |
| 2021 | 0.30 | 0.31 | 0.31 | 0.33 | 0.01 |
| 2022 | 0.29 | 0.32 | 0.32 | 0.38 | 0.02 |
| 2023 | 0.27 | 0.32 | 0.33 | 0.38 | 0.04 |
| 2024 | 0.26 | 0.33 | 0.33 | 0.38 | 0.04 |
| 2025 | 0.25 | 0.33 | 0.33 | 0.38 | 0.05 |
| 2026 | 0.24 | 0.34 | 0.33 | 0.38 | 0.05 |
| 2027 | 0.25 | 0.34 | 0.33 | 0.38 | 0.05 |
| 2028 | 0.24 | 0.34 | 0.33 | 0.38 | 0.05 |
| 2029 | 0.25 | 0.33 | 0.33 | 0.38 | 0.05 |

Table 2.37—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=\max F_{A B C}$ in each year 2017-2018 and $F=F_{\text {OFL }}$ thereafter (Scenario 7), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 239,000 | 239,000 | 239,000 | 239,000 | 0 |
| 2018 | 246,000 | 246,000 | 246,000 | 246,000 | 0 |
| 2019 | 231,000 | 231,000 | 231,000 | 231,000 | 1 |
| 2020 | 168,000 | 168,000 | 168,000 | 169,000 | 587 |
| 2021 | 147,000 | 151,000 | 152,000 | 160,000 | 4,651 |
| 2022 | 136,000 | 155,000 | 161,000 | 211,000 | 23,682 |
| 2023 | 111,000 | 158,000 | 171,000 | 264,000 | 51,246 |
| 2024 | 96,400 | 165,000 | 174,000 | 288,000 | 61,482 |
| 2025 | 90,300 | 167,000 | 178,000 | 284,000 | 65,207 |
| 2026 | 88,200 | 169,000 | 180,000 | 297,000 | 67,595 |
| 2027 | 89,000 | 170,000 | 180,000 | 298,000 | 67,276 |
| 2028 | 88,200 | 170,000 | 179,000 | 295,000 | 65,668 |
| 2029 | 90,900 | 169,000 | 178,000 | 294,000 | 64,585 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 324,000 | 324,000 | 324,000 | 324,000 | 0 |
| 2018 | 325,000 | 325,000 | 325,000 | 325,000 | 0 |
| 2019 | 286,000 | 286,000 | 286,000 | 286,000 | 45 |
| 2020 | 233,000 | 234,000 | 234,000 | 235,000 | 810 |
| 2021 | 208,000 | 212,000 | 213,000 | 224,000 | 5,319 |
| 2022 | 197,000 | 211,000 | 215,000 | 248,000 | 17,314 |
| 2023 | 183,000 | 214,000 | 222,000 | 286,000 | 33,641 |
| 2024 | 173,000 | 219,000 | 226,000 | 315,000 | 44,407 |
| 2025 | 166,000 | 220,000 | 230,000 | 319,000 | 49,657 |
| 2026 | 164,000 | 222,000 | 232,000 | 328,000 | 52,447 |
| 2027 | 165,000 | 222,000 | 232,000 | 330,000 | 52,932 |
| 2028 | 164,000 | 222,000 | 231,000 | 331,000 | 51,463 |
| 2029 | 167,000 | 221,000 | 231,000 | 330,000 | 50,096 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2018 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2019 | 0.38 | 0.38 | 0.38 | 0.38 | 0.00 |
| 2020 | 0.35 | 0.35 | 0.36 | 0.36 | 0.00 |
| 2021 | 0.31 | 0.32 | 0.32 | 0.34 | 0.01 |
| 2022 | 0.30 | 0.32 | 0.32 | 0.38 | 0.02 |
| 2023 | 0.27 | 0.32 | 0.33 | 0.38 | 0.04 |
| 2024 | 0.26 | 0.33 | 0.33 | 0.38 | 0.04 |
| 2025 | 0.25 | 0.33 | 0.33 | 0.38 | 0.05 |
| 2026 | 0.24 | 0.34 | 0.33 | 0.38 | 0.05 |
| 2027 | 0.25 | 0.34 | 0.33 | 0.38 | 0.05 |
| 2028 | 0.24 | 0.34 | 0.33 | 0.38 | 0.05 |
| 2029 | 0.25 | 0.33 | 0.33 | 0.38 | 0.05 |

Table 2.38a (page 1 of 2)—Incidental catch (t) of FMP species taken in the EBS trawl fishery for Pacific cod, expressed as a proportion of the incidental catch of that species taken in all FMP EBS fisheries, 1991-2016 (2016 data current through October 9). Color shading: red = row minimum, green = row maximum (minima and maxima computed across both pages of the table).

| Species/group | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alaska Plaice |  |  |  |  |  |  |  |  |  |  |  | 0.03 | 0.03 |
| Arrowtooth Flounder | 0.06 | 0.24 | 0.21 | 0.16 | 0.20 | 0.26 | 0.24 | 0.11 | 0.12 | 0.11 | 0.19 | 0.38 | 0.43 |
| Atka Mackerel | 0.12 | 0.20 | 0.01 | 0.02 | 0.01 | 0.11 | 0.22 | 0.82 | 0.11 | 0.19 | 0.27 | 0.35 | 0.75 |
| Flathead Sole |  |  |  |  | 0.32 | 0.30 | 0.29 | 0.16 | 0.21 | 0.19 | 0.11 | 0.22 | 0.23 |
| Flounder | 0.02 | 0.07 | 0.13 | 0.14 |  |  |  |  |  |  |  |  |  |
| Greenland Turbot | 0.02 | 0.03 | 0.04 | 0.02 | 0.06 | 0.06 | 0.02 | 0.09 | 0.01 | 0.05 | 0.03 | 0.04 | 0.11 |
| Kamchatka Flounder |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Northern Rockfish |  |  |  |  |  |  |  |  |  |  |  | 0.40 | 0.24 |
| Octopus |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Other Flatfish |  |  |  |  | 0.06 | 0.07 | 0.06 | 0.04 | 0.03 | 0.05 | 0.05 | 0.14 | 0.33 |
| Other Rockfish | 0.04 | 0.28 | 0.23 | 0.05 | 0.14 | 0.11 | 0.07 | 0.35 | 0.14 | 0.19 | 0.02 | 0.11 | 0.28 |
| Other Species |  |  |  |  |  |  |  |  |  |  |  |  | 0.12 |
| Pacific Cod | 0.08 | 0.14 | 0.17 | 0.16 | 0.20 | 0.12 | 0.17 | 0.03 | 0.12 | 0.06 | 0.06 | 0.16 | 0.09 |
| Pacific Ocean Perch | 0.21 | 0.24 | 0.27 | 0.23 | 0.29 | 0.19 | 0.39 | 0.27 | 0.14 | 0.53 | 0.04 | 0.02 | 0.04 |
| Pollock | 0.05 | 0.12 | 0.24 | 0.20 | 0.21 | 0.25 | 0.30 | 0.23 | 0.45 | 0.30 | 0.18 | 0.27 | 0.38 |
| Rock Sole | 0.04 | 0.08 | 0.12 | 0.18 | 0.34 | 0.31 | 0.30 | 0.22 | 0.31 | 0.19 | 0.27 | 0.22 | 0.27 |
| Rougheye Rockfish |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sablefish | 0.01 | 0.01 | conf | 0.01 | 0.00 | 0.04 | 0.00 | 0.03 | 0.06 | 0.08 | 0.04 | 0.10 | 0.19 |
| Sculpin |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shark |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sharpchin/Northern Rockfish |  | 0.29 | 0.30 |  |  |  |  |  |  |  | 0.12 |  |  |
| Short/Rough/Sharp/North Rockfish | 0.26 | 0.58 | 0.18 | 0.12 | 0.17 | 0.13 | 0.45 | 0.27 | 0.16 | 0.28 | 0.46 |  |  |
| Shortraker Rockfish |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shortraker/Rougheye Rockfish |  | 0.02 | conf |  |  |  |  |  |  |  | conf | 0.05 | 0.05 |
| Skate |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Squid | 0.01 | 0.03 | 0.00 | 0.28 | 0.00 | conf | 0.00 | 0.01 | 0.00 | 0.00 | 0.01 | 0.00 | 0.02 |
| Yellowfin Sole | 0.00 | 0.01 | 0.03 | 0.04 | 0.01 | 0.05 | 0.02 | 0.02 | 0.03 | 0.07 | 0.05 | 0.09 | 0.06 |

Table 2.38a (page 2 of 2)—Incidental catch ( $t$ ) of FMP species taken in the EBS trawl fishery for Pacific cod, expressed as a proportion of the incidental catch of that species taken in all FMP EBS fisheries, 1991-2016 (2016 data current through October 9). Color shading: red = row minimum, green = row maximum (minima and maxima computed across both pages of the table).

| Species/group | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alaska Plaice | 0.05 | 0.04 | 0.02 | 0.03 | 0.00 | 0.01 | 0.00 | 0.01 | 0.01 | 0.05 | 0.02 | 0.00 | 0.00 |
| Arrowtooth Flounder | 0.52 | 0.41 | 0.45 | 0.23 | 0.08 | 0.08 | 0.07 | 0.06 | 0.07 | 0.07 | 0.08 | 0.08 | 0.09 |
| Atka Mackerel | 0.76 | 0.36 | 0.24 | 0.17 | 0.10 | 0.40 | 0.39 | 0.35 | 0.27 | 0.06 | 0.09 | 0.03 | 0.90 |
| Flathead Sole | 0.33 | 0.23 | 0.41 | 0.39 | 0.11 | 0.07 | 0.06 | 0.09 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 |
| Flounder |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Greenland Turbot | 0.17 | 0.05 | 0.11 | 0.21 | 0.01 | 0.00 | 0.03 | 0.00 | 0.02 | 0.01 | 0.01 | 0.00 | 0.00 |
| Kamchatka Flounder |  |  |  |  |  |  |  | 0.01 | 0.01 | 0.02 | 0.04 | 0.02 | 0.03 |
| Northern Rockfish | 0.59 | 0.31 | 0.28 | 0.08 | 0.05 | 0.03 | 0.20 | 0.06 | 0.11 | 0.01 | 0.01 | 0.00 | conf |
| Octopus |  |  |  |  |  |  |  | 0.03 | 0.01 | 0.02 | 0.02 | 0.01 | 0.00 |
| Other Flatfish | 0.49 | 0.35 | 0.20 | 0.07 | 0.03 | 0.03 | 0.04 | 0.03 | 0.03 | 0.03 | 0.01 | 0.07 | 0.11 |
| Other Rockfish | 0.33 | 0.32 | 0.24 | 0.06 | 0.06 | 0.02 | 0.04 | 0.03 | 0.32 | 0.03 | 0.03 | 0.07 | 0.05 |
| Other Species | 0.12 | 0.07 | 0.11 | 0.17 | 0.04 | 0.03 | 0.03 |  |  |  |  |  |  |
| Pacific Cod | 0.07 | 0.03 | 0.08 | 0.12 | 0.01 | 0.06 | 0.06 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.04 |
| Pacific Ocean Perch | 0.26 | 0.22 | 0.08 | 0.04 | 0.01 | 0.01 | 0.00 | 0.01 | 0.07 | 0.00 | conf | 0.00 | 0.00 |
| Pollock | 0.39 | 0.36 | 0.53 | 0.75 | 0.43 | 0.37 | 0.32 | 0.42 | 0.29 | 0.22 | 0.17 | 0.05 | 0.10 |
| Rock Sole | 0.30 | 0.37 | 0.35 | 0.27 | 0.14 | 0.07 | 0.12 | 0.18 | 0.16 | 0.07 | 0.21 | 0.30 | 0.25 |
| Rougheye Rockfish | 0.12 | 0.05 |  |  | conf |  | conf |  |  |  |  |  | conf |
| Sablefish | 0.32 | 0.09 | 0.00 | 0.00 | 0.00 | conf | conf | conf | conf | 0.02 |  | 0.01 | 0.00 |
| Sculpin |  |  |  |  |  |  |  | 0.07 | 0.07 | 0.07 | 0.07 | 0.07 | 0.13 |
| Shark |  |  |  |  |  |  |  |  |  | 0.00 | 0.00 | 0.01 | conf |
| Sharpchin/Northern Rockfish Short/Rough/Sharp/North Rockfish |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shortraker Rockfish | conf |  | conf | conf |  | conf |  |  |  | conf |  |  | conf |
| Shortraker/Rougheye Rockfish |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Skate |  |  |  |  |  |  |  | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 |
| Squid | 0.01 | 0.00 | conf | 0.00 | conf |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | conf |
| Yellowfin Sole | 0.11 | 0.11 | 0.08 | 0.03 | 0.01 | 0.00 | 0.01 | 0.01 | 0.01 | 0.06 | 0.02 | 0.00 | 0.00 |

Table 2.38b (page 1 of 2)—Incidental catch (t) of FMP species taken in the EBS longline fishery for Pacific cod, expressed as a proportion of the incidental catch of that species taken in all FMP EBS fisheries, 1991-2016 (2016 data current through October 9). Color shading: red = row minimum, green = row maximum (minima and maxima computed across both pages of the table).

| Species/group | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alaska Plaice |  |  |  |  |  |  |  |  |  |  |  | 0.00 | 0.00 |
| Arrowtooth Flounder | 0.11 | 0.15 | 0.08 | 0.10 | 0.20 | 0.15 | 0.23 | 0.11 | 0.08 | 0.14 | 0.14 | 0.11 | 0.11 |
| Atka Mackerel | 0.00 | 0.01 | 0.01 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.01 | 0.00 |
| Flathead Sole |  | conf |  |  | 0.03 | 0.03 | 0.03 | 0.05 | 0.06 | 0.06 | 0.08 | 0.09 | 0.09 |
| Flounder | 0.01 | 0.01 | 0.01 | 0.01 |  |  |  |  |  |  |  |  |  |
| Greenland Turbot | 0.10 | 0.20 | 0.05 | 0.09 | 0.14 | 0.19 | 0.19 | 0.07 | 0.04 | 0.12 | 0.05 | 0.08 | 0.17 |
| Kamchatka Flounder |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Northern Rockfish |  |  |  |  |  |  |  |  |  |  |  | 0.08 | 0.09 |
| Octopus |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Other Flatfish |  |  |  |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.01 | 0.06 | 0.04 |
| Other Rockfish | 0.03 | 0.12 | 0.14 | 0.08 | 0.06 | 0.20 | 0.11 | 0.08 | 0.20 | 0.12 | 0.35 | 0.25 | 0.11 |
| Other Species |  |  |  |  |  |  |  |  |  |  |  |  | 0.56 |
| Pacific Cod | 0.09 | 0.08 | 0.07 | 0.10 | 0.09 | 0.11 | 0.16 | 0.70 | 0.42 | 0.69 | 0.62 | 0.58 | 0.65 |
| Pacific Ocean Perch | 0.00 | 0.01 | 0.00 | 0.00 | 0.05 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Pollock | 0.01 | 0.02 | 0.02 | 0.02 | 0.03 | 0.03 | 0.04 | 0.04 | 0.02 | 0.05 | 0.06 | 0.04 | 0.05 |
| Rock Sole | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Rougheye Rockfish |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sablefish | 0.05 | 0.73 | 0.07 | 0.08 | 0.12 | 0.13 | 0.12 | 0.08 | 0.08 | 0.39 | 0.32 | 0.20 | 0.30 |
| Sculpin |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shark |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sharpchin/Northern Rockfish |  | 0.01 | 0.01 |  |  |  |  |  |  |  | 0.05 |  |  |
| Short/Rough/Sharp/North Rockfish | 0.03 | 0.05 | 0.04 | 0.11 | 0.18 | 0.18 | 0.05 | 0.14 | 0.03 | 0.19 | 0.05 |  |  |
| Shortraker Rockfish |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shortraker/Rougheye Rockfish |  | 0.10 | 0.19 |  |  |  |  |  |  |  | 0.74 | 0.19 | 0.20 |
| Skate |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Squid |  |  |  |  |  |  |  |  |  | 0.00 |  | conf | conf |
| Yellowfin Sole | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.07 | 0.05 | 0.05 |

Table 2.38b (page 2 of 2)—Incidental catch (t) of FMP species taken in the EBS longline fishery for Pacific cod, expressed as a proportion of the incidental catch of that species taken in all FMP EBS fisheries, 1991-2016 (2016 data current through October 9). Color shading: red = row minimum, green = row maximum (minima and maxima computed across both pages of the table).

| Species/group | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alaska Plaice | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Arrowtooth Flounder | 0.09 | 0.14 | 0.12 | 0.16 | 0.18 | 0.20 | 0.16 | 0.22 | 0.28 | 0.14 | 0.21 | 0.33 | 0.19 |
| Atka Mackerel | 0.01 | 0.00 | 0.00 | 0.01 | 0.01 | 0.01 | 0.02 | 0.01 | 0.01 | 0.13 | 0.17 | 0.02 | 0.05 |
| Flathead Sole | 0.11 | 0.14 | 0.11 | 0.06 | 0.14 | 0.12 | 0.14 | 0.18 | 0.17 | 0.24 | 0.39 | 0.39 | 0.33 |
| Flounder |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Greenland Turbot | 0.11 | 0.12 | 0.13 | 0.15 | 0.11 | 0.15 | 0.16 | 0.16 | 0.11 | 0.01 | 0.02 | 0.13 | 0.28 |
| Kamchatka Flounder |  |  |  |  |  |  |  | 0.14 | 0.11 | 0.07 | 0.09 | 0.13 | 0.15 |
| Northern Rockfish | 0.05 | 0.08 | 0.04 | 0.11 | 0.30 | 0.56 | 0.68 | 0.39 | 0.16 | 0.38 | 0.43 | 0.48 | 0.31 |
| Octopus |  |  |  |  |  |  |  | 0.05 | 0.12 | 0.11 | 0.06 | 0.08 | 0.15 |
| Other Flatfish | 0.06 | 0.09 | 0.07 | 0.01 | 0.01 | 0.06 | 0.07 | 0.02 | 0.03 | 0.01 | 0.01 | 0.04 | 0.02 |
| Other Rockfish | 0.23 | 0.26 | 0.24 | 0.24 | 0.19 | 0.16 | 0.61 | 0.38 | 0.21 | 0.37 | 0.50 | 0.51 | 0.21 |
| Other Species | 0.65 | 0.68 | 0.56 | 0.44 | 0.51 | 0.51 | 0.53 |  |  |  |  |  |  |
| Pacific Cod | 0.75 | 0.82 | 0.63 | 0.68 | 0.63 | 0.73 | 0.51 | 0.78 | 0.66 | 0.63 | 0.76 | 0.78 | 0.81 |
| Pacific Ocean Perch | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
| Pollock | 0.03 | 0.03 | 0.03 | 0.03 | 0.11 | 0.10 | 0.22 | 0.17 | 0.09 | 0.11 | 0.04 | 0.06 | 0.06 |
| Rock Sole | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.01 | 0.01 | 0.03 | 0.05 | 0.02 |
| Rougheye Rockfish | 0.14 | 0.33 | 0.68 | 0.48 | 0.39 | 0.31 | 0.18 | 0.27 | 0.19 | 0.17 | 0.24 | 0.45 | 0.37 |
| Sablefish | 0.16 | 0.22 | 0.21 | 0.16 | 0.03 | 0.04 | 0.04 | 0.19 | 0.02 | 0.10 | 0.07 | 0.26 | 0.70 |
| Sculpin |  |  |  |  |  |  |  | 0.25 | 0.25 | 0.17 | 0.31 | 0.38 | 0.33 |
| Shark |  |  |  |  |  |  |  | 0.31 | 0.33 | 0.40 | 0.41 | 0.44 | 0.33 |
| Sharpchin/Northern Rockfish Short/Rough/Sharp/North Rockfish |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shortraker Rockfish | 0.12 | 0.31 | 0.20 | 0.63 | 0.12 | 0.64 | 0.33 | 0.29 | 0.22 | 0.09 | 0.08 | 0.24 | 0.13 |
| Shortraker/Rougheye Rockfish |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Skate |  |  |  |  |  |  |  | 0.76 | 0.76 | 0.78 | 0.82 | 0.89 | 0.91 |
| Squid | conf | conf |  |  | conf | conf | conf | 0.00 |  | 0.00 | 0.00 | 0.00 |  |
| Yellowfin Sole | 0.04 | 0.07 | 0.05 | 0.02 | 0.06 | 0.09 | 0.03 | 0.14 | 0.20 | 0.23 | 0.39 | 0.47 | 0.35 |

Table 2.38c (page 1 of 2)—Incidental catch (t) of FMP species taken in the EBS pot fishery for Pacific cod, expressed as a proportion of the incidental catch of that species taken in all FMP EBS fisheries, 1991-2016 (2016 data current through October 9). Color shading: red = row minimum, green = row maximum (minima and maxima computed across both pages of the table).

| Species/group | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alaska Plaice |  |  |  |  |  |  |  |  |  |  |  | conf | conf |
| Arrowtooth Flounder | 0.00 | 0.00 | conf | conf | 0.00 | 0.00 | 0.00 | 0.00 | conf | conf | conf | 0.02 | 0.00 |
| Atka Mackerel | 0.00 | 0.03 | conf | 0.05 | 0.23 | 0.11 | 0.29 | 0.03 | conf | conf | conf | conf | 0.06 |
| Flathead Sole |  |  |  |  | conf | 0.00 | conf | conf | 0.00 | conf | conf | conf | 0.00 |
| Flounder | conf | 0.00 | conf | conf |  |  |  |  |  |  |  |  |  |
| Greenland Turbot | conf | conf |  | conf | 0.00 | 0.00 | conf | conf | conf |  | conf | conf | 0.00 |
| Kamchatka Flounder |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Northern Rockfish |  |  |  |  |  |  |  |  |  |  |  | conf | 0.02 |
| Octopus |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Other Flatfish |  |  |  |  | conf | 0.00 | 0.00 | conf | conf | conf | conf | conf | 0.00 |
| Other Rockfish | 0.00 | 0.00 | conf | 0.01 | 0.02 | 0.04 | 0.06 | 0.03 | conf | conf | conf | conf | 0.07 |
| Other Species |  |  |  |  |  |  |  |  |  |  |  |  | 0.02 |
| Pacific Cod | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.02 | 0.00 | 0.01 | 0.02 | 0.02 | 0.02 | 0.03 | 0.06 |
| Pacific Ocean Perch | conf | conf | conf | conf | 0.00 | 0.00 | conf | conf | conf | conf | conf | conf | 0.00 |
| Pollock | 0.00 | 0.00 | conf | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | conf | 0.00 | 0.00 |
| Rock Sole | 0.00 | 0.00 | conf | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | conf | conf | 0.00 |
| Rougheye Rockfish |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sablefish | conf | conf |  | conf | conf | conf | conf | conf | conf |  | conf | conf | 0.00 |
| Sculpin |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shark |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sharpchin/Northern Rockfish |  | conf | conf |  |  |  |  |  |  |  | conf |  |  |
| Short/Rough/Sharp/North Rockfish | 0.00 |  |  | conf | 0.01 | 0.00 | 0.00 | conf | conf | conf | conf |  |  |
| Shortraker Rockfish |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shortraker/Rougheye Rockfish |  | conf | conf |  |  |  |  |  |  |  | conf | conf | 0.00 |
| Skate |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Squid |  |  |  |  | conf | conf |  |  | conf |  | conf | conf |  |
| Yellowfin Sole | 0.00 | 0.00 | conf | conf | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | conf | 0.00 | 0.01 |

Table 2.38c (page 2 of 2)—Incidental catch (t) of FMP species taken in the EBS pot fishery for Pacific cod, expressed as a proportion of the incidental catch of that species taken in all FMP EBS fisheries, 1991-2016 (2016 data current through October 9). Color shading: red = row minimum, green = row maximum (minima and maxima computed across both pages of the table).

| Species/group | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alaska Plaice | conf | conf | conf | conf |  |  |  |  | 0.00 | conf | conf | conf | conf |
| Arrowtooth Flounder | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Atka Mackerel | 0.03 | 0.17 | 0.29 | 0.11 | 0.68 | 0.03 | 0.56 | 0.11 | 0.05 | 0.19 | 0.36 | 0.06 | 0.02 |
| Flathead Sole | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Flounder |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Greenland Turbot |  | conf | 0.00 | conf | conf | conf |  | 0.00 |  |  | conf | conf |  |
| Kamchatka Flounder |  |  |  |  |  |  |  | 0.00 | 0.00 | 0.00 | conf | 0.00 |  |
| Northern Rockfish | 0.01 | 0.02 | 0.01 | 0.02 | 0.11 | 0.06 | 0.02 | 0.02 | 0.01 | 0.00 | 0.01 | 0.01 | 0.00 |
| Octopus |  |  |  |  |  |  |  | 0.88 | 0.81 | 0.85 | 0.90 | 0.85 | 0.74 |
| Other Flatfish | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Other Rockfish | 0.04 | 0.10 | 0.12 | 0.01 | 0.02 | 0.00 | 0.02 | 0.04 | 0.02 | 0.11 | 0.05 | 0.05 | 0.03 |
| Other Species | 0.01 | 0.01 | 0.02 | 0.01 | 0.02 | 0.01 | 0.01 |  |  |  |  |  |  |
| Pacific Cod | 0.02 | 0.02 | 0.03 | 0.02 | 0.08 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.04 | 0.03 | 0.02 |
| Pacific Ocean Perch | 0.00 | 0.00 | 0.00 | conf | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  | 0.00 |  |
| Pollock | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Rock Sole | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Rougheye Rockfish | 0.00 | 0.01 |  |  |  |  |  |  |  |  |  |  |  |
| Sablefish | 0.01 | 0.00 | 0.08 |  |  |  |  | conf |  | 0.00 |  |  |  |
| Sculpin |  |  |  |  |  |  |  | 0.03 | 0.02 | 0.06 | 0.07 | 0.06 | 0.05 |
| Shark |  |  |  |  |  |  |  | conf | 0.00 |  |  |  |  |
| Sharpchin/Northern Rockfish Short/Rough/Sharp/North Rockfish |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shortraker Rockfish |  |  |  |  |  |  | 0.00 |  |  |  | conf |  |  |
| Shortraker/Rougheye Rockfish |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Skate |  |  |  |  |  |  |  | 0.00 | conf |  | 0.00 | conf | conf |
| Squid |  | conf |  | conf |  |  | conf |  |  |  | conf |  | conf |
| Yellowfin Sole | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.01 | 0.00 | 0.01 | 0.01 | 0.05 | 0.08 | 0.08 | 0.03 |

Table 2.39—Incidental catch (t) of selected members of the former "Other Species" complex taken in the EBS fisheries for Pacific cod (all gears), expressed as a proportion of the incidental catch of that species taken in all FMP EBS fisheries, 1991-2016 (2016 data current through September 25). Color shading: red = row minimum, green = row maximum (minima and maxima computed across both panels of the table).

| Species Common Name | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| octopus, North Pacific |  |  |  |  |  |  |  |  | conf | conf | conf | 0.73 | 0.81 |
| shark, other |  |  |  |  |  |  |  |  |  |  |  | conf | 0.66 |
| shark, Pacific sleeper |  |  |  |  |  |  |  |  | conf | conf | 0.05 | 0.09 | 0.50 |
| shark, salmon |  |  |  |  |  |  |  |  | conf |  |  |  | conf |
| shark, spiny dogfish |  |  |  |  |  |  |  |  |  | 0.91 | 0.42 | 0.92 | 0.99 |
| skate, Alaska <br> skate, big |  |  |  |  |  |  |  |  |  |  |  |  |  |
| skate, longnose |  |  |  |  |  |  |  |  |  |  |  | conf | 0.71 |
| skate, other |  |  |  |  |  |  |  |  | 0.16 | 0.04 | conf | 0.43 | 0.81 |
| squid, majestic | 0.01 | 0.03 | 0.00 | 0.28 | 0.00 | conf | 0.00 | 0.01 | 0.00 | 0.00 | 0.01 | 0.00 | 0.02 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Species Common Name | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| octopus, North Pacific | 0.82 | 0.88 | 0.94 | 0.89 | 0.90 | 0.66 | 0.84 | 0.97 | 0.94 | 0.98 | 0.97 | 0.93 | 0.89 |
| shark, other | 0.35 | 0.52 | 0.01 | 0.16 | 0.42 | 0.72 | 0.48 | 0.75 | 0.41 | 0.79 | 0.79 | 0.61 | conf |
| shark, Pacific sleeper | 0.56 | 0.60 | 0.41 | 0.20 | 0.17 | 0.30 | 0.59 | 0.50 | 0.22 | 0.39 | 0.61 | 0.61 | 0.51 |
| shark, salmon | conf | 0.07 | 0.02 |  |  |  |  |  |  | 0.01 |  | conf |  |
| shark, spiny dogfish | 0.97 | 0.98 | 0.97 | 0.83 | 0.63 | 0.95 | 0.93 | 0.93 | 0.98 | 0.80 | 0.83 | 0.89 | 0.87 |
| skate, Alaska |  |  |  |  |  |  | 0.26 |  |  |  |  |  |  |
| skate, big | 0.84 | 0.72 | 0.92 | 0.73 | 0.71 | 0.51 | 0.73 |  |  |  |  |  |  |
| skate, longnose | 0.55 | 0.97 | 0.67 | 0.37 | 1.00 | 0.67 | 0.49 |  |  |  |  |  |  |
| skate, other | 0.85 | 0.85 | 0.78 | 0.75 | 0.70 | 0.67 | 0.93 | 0.78 | 0.77 | 0.79 | 0.83 | 0.90 | 0.93 |
| squid, majestic | 0.01 | 0.00 | conf | 0.00 | conf | conf | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | conf |

Table 2.40-Incidental catch (herring and halibut in t , salmon and crab in number of individuals) of prohibited species taken in the EBS fisheries for Pacific cod (all gears), expressed as a proportion of the incidental catch of that species taken in all FMP EBS fisheries, 1991-2016 (2016 data current through October 2). Color shading: red = row minimum, green = row maximum (minima and maxima computed across both panels of the table). Note that all entries for 2003 are marked " $\mathrm{n} / \mathrm{a}$ ", due to an error in the database that was discovered too late to be corrected in time for this assessment.

| Species Group Name | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Bairdi Tanner Crab <br> Blue King Crab <br> Chinook Salmon <br> Golden (Brown) King Crab <br> Halibut <br> Herring <br> Non-Chinook Salmon <br> Opilio Tanner (Snow) Crab <br> Other King Crab <br> Red King Crab | 0.20 | 0.10 | 0.07 | 0.13 | 0.14 | 0.22 | 0.15 | 0.10 | 0.17 | 0.16 | 0.14 | 0.29 | n/a |
|  |  |  |  |  |  |  |  |  |  |  |  |  | n/a |
|  | 0.09 | 0.12 | 0.13 | 0.18 | 0.34 | 0.09 | 0.09 | 0.04 | 0.14 | 0.20 | 0.08 | 0.04 | n/a |
|  |  |  |  |  |  |  |  |  |  |  |  |  | n/a |
|  | 0.52 | 0.64 | 0.49 | 0.67 | 0.71 | 0.74 | 0.71 | 0.68 | 0.66 | 0.69 | 0.63 | 0.69 | n/a |
|  | conf | 0.01 | 0.03 | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 | 0.02 | n /a |
|  | 0.00 | 0.00 | 0.00 | 0.01 | 0.04 | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 | 0.03 | 0.01 | n/a |
|  | 0.02 | 0.02 | 0.02 | 0.04 | 0.05 | 0.10 | 0.17 | 0.17 | 0.33 | 0.12 | 0.12 | 0.34 | n/a |
|  | 0.02 | 0.12 | 0.01 | 0.08 | 0.16 | 0.66 | 0.54 | 0.73 | 0.35 | 0.33 | 0.58 | 0.69 | n/a |
|  | 0.31 | 0.07 | 0.01 | 0.01 | 0.17 | 0.78 | 0.36 | 0.23 | 0.18 | 0.38 | 0.26 | 0.32 | n/a |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Species Group Name <br> Bairdi Tanner Crab <br> Blue King Crab <br> Chinook Salmon <br> Golden (Brown) King Crab <br> Halibut <br> Herring <br> Non-Chinook Salmon <br> Opilio Tanner (Snow) Crab <br> Other King Crab <br> Red King Crab | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
|  | 0.30 | 0.19 | 0.40 | 0.57 | 0.66 | 0.53 | 0.45 | 0.27 | 0.22 | 0.26 | 0.50 | 0.61 | 0.69 |
|  | 0.94 | 0.95 | 0.70 | 1.00 | 0.87 | 0.89 | 1.00 | 0.56 | 0.77 | 0.32 | 0.57 | 0.81 | 0.94 |
|  | 0.08 | 0.04 | 0.03 | 0.04 | 0.03 | 0.01 | 0.04 | 0.00 | 0.05 | 0.04 | 0.05 | 0.05 | 0.09 |
|  | 0.00 | 0.21 | 0.01 | 0.00 | 0.00 | 0.01 | 0.02 | 0.00 | 0.02 | 0.03 | 0.03 | 0.02 | 0.01 |
|  | 0.73 | 0.72 | 0.68 | 0.64 | 0.66 | 0.63 | 0.65 | 0.68 | 0.69 | 0.67 | 0.63 | 0.66 | 0.60 |
|  | 0.01 | 0.03 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
|  | 0.01 | 0.00 | 0.02 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  | 0.14 | 0.08 | 0.34 | 0.53 | 0.47 | 0.52 | 0.31 | 0.20 | 0.08 | 0.06 | 0.19 | 0.23 | 0.23 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0.14 | 0.16 | 0.14 | 0.32 | 0.29 | 0.10 | 0.06 | 0.35 | 0.26 | 0.76 | 0.82 | 0.90 | 0.44 |

Table 2.41a (page 1 of 2)—Incidental catch (t) of non-target species groups-other than birds-taken in the EBS fisheries for Pacific cod (all gears), expressed as a proportion of the incidental catch of that species group taken in all FMP EBS fisheries, 2004-2016 (2016 data are current through September 25). Color shading: red = row minimum, green = row maximum.

| Species Group Name | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Benthic urochordata | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 | 0.01 | 0.04 | 0.20 | 0.28 | 0.20 | 0.29 | 0.27 | 0.06 |
| Bivalves | 0.84 | 0.72 | 0.77 | 0.61 | 0.85 | 0.85 | 0.47 | 0.82 | 0.92 | 0.88 | 0.79 | 0.81 | 0.67 |
| Brittle star unidentified | 0.02 | 0.01 | 0.02 | 0.04 | 0.01 | 0.05 | 0.01 | 0.07 | 0.06 | 0.09 | 0.04 | 0.08 | 0.05 |
| Capelin | 0.02 |  |  | 0.00 | 0.00 |  | 0.00 | 0.00 | 0.00 | 0.10 | 0.02 | 0.00 |  |
| Corals Bryozoans - Corals Bryozoans Unidentified | 0.46 | 0.40 | 0.09 | 0.86 | 0.15 | 0.85 | 0.24 | 0.50 | 0.92 | 0.25 | 0.46 | 0.60 | 0.36 |
| Corals Bryozoans - Red Tree Coral | 0.62 | 0.44 | 0.01 | 0.96 | 0.12 | 0.33 |  |  |  |  |  |  |  |
| Dark Rockfish |  |  |  |  | 0.98 | 0.96 | 0.96 | 0.98 | 0.79 | 0.98 | 0.25 | 0.77 | 0.23 |
| Eelpouts | 0.36 | 0.46 | 0.29 | 0.11 | 0.09 | 0.04 | 0.05 | 0.05 | 0.14 | 0.16 | 0.28 | 0.58 | 0.47 |
| Eulachon | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |  | conf | 0.00 | 0.00 |  |  |  |  |
| Giant Grenadier | 0.07 | 0.12 | 0.05 | 0.07 | 0.10 | 0.07 | 0.12 | 0.28 | 0.21 | 0.12 | 0.19 | 0.04 | 0.10 |
| Greenlings | 0.64 | 0.56 | 0.62 | 0.20 | 0.73 | 0.74 | 0.78 | 0.64 | 0.77 | 1.00 | 0.84 | 0.69 | 0.66 |
| Grenadier - Pacific Grenadier | 0.70 | 0.00 |  |  |  |  | 0.05 |  |  |  |  |  |  |
| Grenadier - Ratail Grenadier Unidentified | 0.11 | 0.15 | 0.09 | 0.23 | 0.40 | 0.08 | 0.19 | 0.04 | 0.04 | 0.76 | 0.06 | 0.61 | 0.07 |
| Gunnels | 1.00 | 1.00 |  | 0.03 |  |  |  |  |  |  |  |  |  |
| Hermit crab unidentified | 0.05 | 0.01 | 0.04 | 0.03 | 0.02 | 0.06 | 0.02 | 0.03 | 0.06 | 0.08 | 0.06 | 0.15 | 0.13 |
| Invertebrate unidentified | 0.01 | 0.01 | 0.08 | 0.22 | 0.02 | 0.24 | 0.41 | 0.38 | 0.21 | 0.25 | 0.30 | 0.27 | 0.47 |
| Lanternfishes (myctophidae) | conf |  |  |  |  |  |  |  |  |  |  |  |  |
| Large Sculpins - Bigmouth Sculpin |  |  |  |  | 0.35 | 0.46 | 0.46 | 0.56 | 0.58 | 0.54 | 0.66 | 0.68 | 0.83 |
| Large Sculpins - Brown Irish Lord |  |  |  |  | 1.00 | 1.00 |  |  |  |  |  |  |  |
| Large Sculpins - Great Sculpin |  |  |  |  | 0.18 | 0.16 | 0.16 | 0.28 | 0.23 | 0.23 | 0.35 | 0.39 | 0.37 |
| Large Sculpins - Hemilepidotus Unidentified |  |  |  |  | 0.93 | 0.99 | 1.00 | 0.99 | 0.99 | 0.98 | 0.98 | 0.98 | 0.97 |
| Large Sculpins - Myoxocephalus Unidentified |  |  |  |  | 0.23 | 0.60 | 0.93 | 0.73 | 0.96 | 0.79 | 0.81 | 0.89 | 0.90 |
| Large Sculpins - Plain Sculpin |  |  |  |  | 0.01 | 0.01 | 0.02 | 0.02 | 0.03 | 0.06 | 0.07 | 0.07 | 0.06 |
| Large Sculpins - Red Irish Lord |  |  |  |  | 0.11 | 0.63 | 0.90 | 0.86 | 1.00 | 0.07 | 0.26 | 0.25 | 0.17 |
| Large Sculpins - Warty Sculpin |  |  |  |  | 0.22 | 0.15 | 0.32 | 0.21 | 0.09 | 0.12 | 0.12 | 0.06 | 0.09 |
| Large Sculpins - Yellow Irish Lord |  |  |  |  | 0.51 | 0.33 | 0.52 | 0.42 | 0.64 | 0.48 | 0.43 | 0.56 | 0.55 |
| Large Sculpins | 0.60 | 0.48 | 0.44 | 0.36 |  |  |  |  |  |  |  |  |  |
| Misc crabs | 0.11 | 0.15 | 0.42 | 0.38 | 0.17 | 0.12 | 0.21 | 0.12 | 0.16 | 0.21 | 0.23 | 0.20 | 0.18 |
| Misc crustaceans | 0.20 | 0.53 | 0.14 | 0.18 | 0.03 | 0.02 | 0.05 | 0.04 | 0.03 | 0.10 | 0.08 | 0.05 | 0.13 |
| Misc fish | 0.42 | 0.37 | 0.28 | 0.17 | 0.14 | 0.18 | 0.20 | 0.24 | 0.33 | 0.24 | 0.52 | 0.44 | 0.53 |
| Misc inverts (worms etc) | 0.02 | 0.01 | 0.06 | 0.32 | 0.01 | 0.02 | 0.00 | 0.01 | 0.00 | conf | 0.02 | 0.03 |  |
| Other Sculpins | 0.64 | 0.55 | 0.60 | 0.40 | 0.70 | 0.58 | 0.69 | 0.57 | 0.75 | 0.30 | 0.50 | 0.42 | 0.46 |
| Other osmerids | 0.06 | 0.00 | 0.01 | 0.00 |  | 0.00 | 0.00 | 0.00 | conf | 0.00 |  | 0.00 |  |

Table 2.41a (page 2 of 2)—Incidental catch (t) of non-target species groups-other than birds-taken in the EBS fisheries for Pacific cod (all gears), expressed as a proportion of the incidental catch of that species group taken in all FMP EBS fisheries, 2004-2016 (2016 data are current through September 25). Color shading: red = row minimum, green = row maximum.

| Species Group Name | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Pacific Sand lance | 0.31 | 0.56 | 0.03 | 0.12 | 0.21 |  | 0.01 | 0.01 | conf |  | conf | 0.09 | 0.09 |
| Pacific Sandfish |  |  |  |  |  |  | 0.28 |  |  |  | 0.07 | 0.19 | 0.32 |
| Pandalid shrimp | 0.18 | 0.01 | 0.02 | 0.10 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Polychaete unidentified | 0.01 | 0.71 | 0.00 | 0.01 | 0.00 | 0.10 | 0.04 | 0.08 | 0.37 | 0.04 | 0.28 | 0.43 | 0.39 |
| Scypho jellies | 0.08 | 0.05 | 0.02 | 0.03 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.05 | 0.01 | 0.02 | 0.01 |
| Sea anemone unidentified | 0.62 | 0.86 | 0.79 | 0.36 | 0.51 | 0.74 | 0.62 | 0.76 | 0.86 | 0.76 | 0.81 | 0.86 | 0.85 |
| Sea pens whips | 0.90 | 0.93 | 0.86 | 0.60 | 0.86 | 0.93 | 0.87 | 0.90 | 0.91 | 0.96 | 0.96 | 0.96 | 0.96 |
| Sea star | 0.14 | 0.16 | 0.12 | 0.09 | 0.07 | 0.15 | 0.11 | 0.09 | 0.20 | 0.15 | 0.22 | 0.20 | 0.22 |
| Snails | 0.07 | 0.11 | 0.07 | 0.10 | 0.09 | 0.25 | 0.17 | 0.17 | 0.25 | 0.31 | 0.43 | 0.46 | 0.69 |
| Sponge unidentified | 0.09 | 0.06 | 0.15 | 0.08 | 0.08 | 0.07 | 0.03 | 0.06 | 0.15 | 0.09 | 0.09 | 0.17 | 0.19 |
| Stichaeidae | 0.06 | 0.11 | 0.06 | 0.01 | 0.04 | 0.00 | conf |  |  |  |  | 0.03 |  |
| Urchins dollars cucumbers | 0.48 | 0.48 | 0.40 | 0.24 | 0.18 | 0.08 | 0.15 | 0.35 | 0.41 | 0.26 | 0.37 | 0.48 | 0.52 |

Table 2.41b—Incidental catch ( t ) of bird species groups taken in the EBS fisheries for Pacific cod (all gears), expressed as a proportion of the incidental catch of that species group taken in all FMP EBS fisheries, 2004-2016 (2016 data are current through September 25).

| Species Group Name | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Auklets |  |  |  |  |  |  |  |  | 1.00 |  | 0.33 | 1.00 | 1.00 |
| Black-footed Albatross | 1.00 | 1.00 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 |  |  |  |  |  |  |
| Cormorant |  |  | 1.00 |  |  |  |  |  |  |  |  |  |  |
| Gull | 1.00 | 1.00 | 0.97 | 0.60 | 0.99 | 0.90 | 0.89 | 1.00 | 1.00 | 0.96 | 1.00 | 1.00 | 0.96 |
| Kittiwake | 0.31 | 0.44 |  |  |  | 0.34 |  | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| Laysan Albatross | 0.76 | 0.37 | 0.01 | 0.00 | 0.68 | 0.00 | 0.31 | 0.94 | 0.71 | 0.20 | 0.75 | 0.17 | 0.67 |
| Murre | 0.41 | 0.01 | 0.65 | 0.73 | 1.00 | 1.00 |  |  | 1.00 |  |  |  | 0.22 |
| Northern Fulmar | 0.91 | 0.86 | 0.57 | 0.74 | 0.75 | 0.86 | 0.85 | 0.86 | 0.79 | 0.87 | 0.82 | 0.84 | 0.93 |
| Other Alcid | 1.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| Other |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Puffin |  |  |  |  |  |  | 1.00 |  |  |  |  |  |  |
| Shearwaters | 0.99 | 0.37 | 0.95 | 0.94 | 0.99 | 0.87 | 0.85 | 0.72 | 0.90 | 0.69 | 0.36 | 0.66 | 0.93 |
| Short-tailed Albatross |  |  |  |  |  |  | 1.00 | 1.00 |  |  | 0.32 |  |  |
| Storm Petrels |  |  | 0.33 |  |  |  |  |  |  |  |  |  |  |
| Unidentified Albatross |  |  | 1.00 |  |  |  |  |  |  |  | 0.92 |  |  |
| Unidentified | 0.96 | 0.98 | 0.97 | 0.96 | 1.00 | 0.92 | 0.94 | 1.00 | 0.95 | 0.94 | 1.00 | 0.94 | 0.99 |

## FIGURES



Figure 2.1a. EBS maps showing each 400 square km cell with trawl hauls containing Pacific cod from at least 3 distinct vessels by season in 2015-2016, overlaid against NMFS 3-digit statistical areas.


Figure 2.1b. EBS maps showing each 400 square km cell with longline sets containing Pacific cod from at least 3 distinct vessels by season in 2015-2016, overlaid against NMFS 3-digit statistical areas.


Figure 2.1c. EBS maps showing each 400 square km cell with pot sets containing Pacific cod from at least 3 distinct vessels by season in 2015-2016, overlaid against NMFS 3-digit statistical areas.


Figure 2.2. Maps showing each 400 square km cell with pot sets containing Pacific cod from at least 3 distinct vessels by season in 2015-2016, overlaid against NMFS 3-digit statistical areas.


Figure 2.3-Time series of fishery catch per unit effort, by gear and season.


Figure 2.4a—Model fits to the trawl survey abundance time series, with 95\% CI for the observations.


Figure 2.4b—Model 16.7 fits to the NMFS longline survey abundance time series, with 95\% CI for the observations.


Figure 2.5a (page 1 of 2)—Fit to trawl survey age composition data obtained by Model 11.5 (grey = observed, green = estimated).


Age (yr)
Figure 2.5a (page 2 of 2)—Fit to trawl survey age composition data obtained by Model 11.5 (grey = observed, green = estimated).


Figure 2.5b (page 1 of 2)—Fit to trawl survey age composition data obtained by Model 16.1 (grey = observed, green = estimated).


Age (yr)

Figure 2.5b (page 2 of 2)—Fit to trawl survey age composition data obtained by Model 16.1 (grey = observed, green = estimated).


Figure 2.5c (page 1 of 2)—Fit to trawl survey age composition data obtained by Model 16.6 (grey = observed, green = estimated).


Age (yr)
Figure 2.5c (page 2 of 2)—Fit to trawl survey age composition data obtained by Model 16.6 (grey = observed, green = estimated).


Figure 2.5d (page 1 of 2)—Fit to trawl survey age composition data obtained by Model 16.7 (grey = observed, green = estimated).


Age (yr)
Figure 2.5d (page 2 of 2)—Fit to trawl survey age composition data obtained by Model 16.7 (grey = observed, green = estimated).


Figure 2.5 e (page 1 of 2)—Fit to trawl survey age composition data obtained by Model 16.8 (grey = observed, green = estimated).


Age (yr)
Figure 2.5e (page 2 of 2)—Fit to trawl survey age composition data obtained by Model 16.8 (grey = observed, green = estimated).


Figure 2.5 f (page 1 of 2)—Fit to trawl survey age composition data obtained by Model 16.9 (grey = observed, green = estimated).


Age (yr)
Figure 2.5 f (page 2 of 2)—Fit to trawl survey age composition data obtained by Model 16.9 (grey = observed, green = estimated).

## Model 11.5



Model 16.6


Model 16.8


Model 16.1


Model 16.7


Model 16.9


Figure 2.6-Time-aggregated fits to the trawl survey age composition data.

Model 11.5


Figure 2.7a—Time-aggregated fits to the size composition data for Model 11.5.

Model 16.1


Model 16.6


Figure 2.7b-Time-aggregated fits to the size composition data for Models 16.1 and 16.6.


Figure 2.7c-Time-aggregated fits to the size composition data for Model 16.7.

Model 16.8


Model 16.9


Figure 2.7d—Time-aggregated fits to the size composition data for Models 16.8 and 16.9.


Figure 2.8-Estimates of mean size at ages 1-3 from Models 1 and 16.x, compared to long-term average survey size $(0-50 \mathrm{~cm})$ composition. The mean sizes at age $1-3$ from the five models in the $16 . x$ series are so similar that only the minimum and maximum (at each age) from these five models is shown.


Figure 2.9 (page 1 of 2)—Fit to mean-size-at-age data from Model 11.5 (not used in Models 16.x). Black = observed, green = estimated.


Age (yr)

Figure 2.9 (page 2 of 2)—Fit to mean-size-at-age data from Model 11.5 (not used in Models 16.x). Black = observed, green = estimated.


Figure 2.10-Time series of estimated log recruitment deviations as estimated by the models, with $95 \%$ confidence intervals.


Figure 2.11a-Time series of spawning biomass relative to $B_{100 \%}$ as estimated by the models.


Figure 2.11b—Time series of total biomass ( t ) as estimated by the models, with survey biomass shown for comparison.


Model 16.6


Model 16.8


Model 16.1


Model 16.7


Model 16.9


Figure 2.12a-Survey (trawl, unless indicated otherwise) selectivity at age as estimated by the models (base values shown for Models 11.5 and 16.8, which have annually varying selectivity).

## Model 11.5



Model 16.8


Figure 2.12b—Annually varying trawl survey selectivity for Models 11.5 and 16.8.


Figure 2.13a-Fishery selectivity at length (cm) as estimated by Model 11.5. Rows represent gear types (trawl, longline, and pot, respectively), and columns represent seasons (Jan-Apr, May-Jul, and Aug-Dec, respectively).


## Model 16.6



Model 16.8


Model 16.7


Model 16.9


Figure 2.13b-Base fishery selectivity at age as estimated by models in the 16.x series (base values shown for Model 16.9, which has annually varying selectivity).


Figure 2.13c—Annually varying trawl survey selectivity for Model 16.9.


Figure 2.14a-Retrospective analysis of spawning biomass estimates from Model 11.5. Top panel: spawning biomass time series with $95 \%$ confidence intervals from the current version of Model 11.5 (2016) and 10 retrospective runs (2006-2015) obtained by dropping one year of data at a time. Bottom panel: change in spawning biomass relative to the current version of Model 11.5 for each of 10 retrospective runs. Mohn's $\rho=0.475$.


Figure 2.14b—Retrospective analysis of spawning biomass estimates from Model 16.1. Top panel: spawning biomass time series with $95 \%$ confidence intervals from the current version of Model 16.1 (2016) and 10 retrospective runs (2006-2015) obtained by dropping one year of data at a time. Bottom panel: change in spawning biomass relative to the current version of Model 16.1 for each of 10 retrospective runs. Mohn's $\rho=0.194$.


Figure 2.14c—Retrospective analysis of spawning biomass estimates from Model 16.6. Top panel: spawning biomass time series with $95 \%$ confidence intervals from the current version of Model 16.6 (2016) and 10 retrospective runs (2006-2015) obtained by dropping one year of data at a time. Bottom panel: change in spawning biomass relative to the current version of Model 16.6 for each of 10 retrospective runs. Mohn's $\rho=0.147$.


Figure 2.14d—Retrospective analysis of spawning biomass estimates from Model 16.7. Top panel: spawning biomass time series with $95 \%$ confidence intervals from the current version of Model 16.7 (2016) and 10 retrospective runs (2006-2015) obtained by dropping one year of data at a time. Bottom panel: change in spawning biomass relative to the current version of Model 16.7 for each of 10 retrospective runs. Mohn's $\rho=0.144$.


Figure 2.14e—Retrospective analysis of spawning biomass estimates from Model 16.8. Top panel: spawning biomass time series with $95 \%$ confidence intervals from the current version of Model 16.8 (2016) and 10 retrospective runs (2006-2015) obtained by dropping one year of data at a time. Bottom panel: change in spawning biomass relative to the current version of Model 16.8 for each of 10 retrospective runs. Mohn's $\rho=0.094$.


Figure 2.14f—Retrospective analysis of spawning biomass estimates from Model 16.9. Top panel: spawning biomass time series with $95 \%$ confidence intervals from the current version of Model 16.9 (2016) and 10 retrospective runs (2006-2015) obtained by dropping one year of data at a time. Bottom panel: change in spawning biomass relative to the current version of Model 16.9 for each of 10 retrospective runs. Mohn's $\rho=0.250$.


Figure 2.15-Between-assessment retrospective analysis of spawning biomass estimates. Top panel: spawning biomass time series with $95 \%$ confidence intervals from each of the final models in the assessments from 2006-2016 (assuming that Model 16.6 is adopted as the final model this year). Bottom panel: change in spawning biomass relative to this year's final model for the final model from each of the 10 previous assessments. Mohn's $\rho=0.885$.


Figure 2.16-Time series of age $0+$ and female spawning biomass as estimated by Model 16.6. Survey biomass is shown for comparison.


Figure 2.17—Time series of recruitment at age 0 as estimated Model 16.6.


Figure 2.18-Trajectory of Pacific cod fishing mortality and female spawning biomass as estimated by Model 16.6, 1977-2018 (yellow square = current year, magenta squares = first two projection years).


Figure 2.19—Environmental effects on recruitment. Upper panel: Estimated log recruitment devs (age 0) versus same-year October-December average of the NPI, with regression line and $95 \%$ confidence interval. Middle panel: Distribution of the regression slope, as generated by a cross-validation analysis. Lower panel: Correlation between individual data points and regression slope. See text for details.

# APPENDIX 2.1: PRELIMINARY ASSESSMENT OF THE PACIFIC COD STOCK IN THE EASTERN BERING SEA 

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## Introduction

This document represents an effort to respond to comments made by the Joint Team Subcommittee on Pacific cod models (JTS), and the SSC on last year's assessment of the Pacific cod (Gadus macrocephalus) stock in the eastern Bering Sea (EBS, Thompson 2015). Many of those comments were informed by the results of a CIE review of the EBS Pacific cod assessment conducted during February 16-19, 2016. The website located at http://tinyurl.com/Pcod-cie-2016 contains every file vetted during the review process as well as the final reports from the three reviewers.

Responses to SSC and Plan Team comments on assessments in general
SSC1 (10/15 minutes): "The Team Procedures document clarifies that the proposed development and testing of a naming convention should focus on tracking the modeling configurations used for a particular stock assessment. The rationale for this request is two-fold. First, it will help us understand how long it has been since a benchmark change in model configuration has occurred; second, it will help the reviewers and public to track model changes. Of the options presented in the Joint Plan Teams minutes, the SSC agrees that Option 4 has several advantages and recommends that this Option be advanced next year." As in last year's final assessment, Option 4a was used to number models in this preliminary assessment.

SSC2 (12/15 minutes): "The SSC reminds the authors and PTs to follow the model numbering scheme adopted at the December 2014 meeting." Given that comment SSC1 superseded the model numbering scheme adopted at the December 2014 meeting, it seems reasonable to assume that inclusion of this comment in the $12 / 15$ minutes was an error.

SSC3 (12/15 minutes): "Many assessments are currently exploring ways to improve model performance by re-weighting historic survey data. The SSC encourages the authors and PTs to refer to the forthcoming CAPAM data-weighting workshop report." Results described by Punt (in press) were used to choose a data-weighting method for Model 16.5.

SSC4 (12/15 minutes): "The SSC recommends that assessment authors work with AFSC’s survey program scientist to develop some objective criteria to inform the best approaches for calculating $Q$ with respect to information provided by previous survey trawl performance studies (e.g. Somerton and Munro 2001), and fish-temperature relationships which may impact $Q$." The recent paper by Weinberg et al. (2016) is an example of the suggested collaboration.

Responses to SSC and Plan Team comments specific to Eastern Bering Sea Pacific cod
Note: Following the procedure initiated in 2014, the task of developing recommendations for models to be included in this year's preliminary Pacific cod assessments (subject to review and potential revision by the SSC) was delegated to the JTS rather than the full Joint Plan Teams.

SSC5 (12/15 minutes): "The SSC was encouraged by the author’s explanation that dome-shaped selectivity may, in part, be explained by the possibility that some of older fish may be residing in the northern Bering Sea (NBS) at the time of the survey. This is supported by the size composition of the fish in the 2010 NBS trawl survey, which suggested that up to 40\% of the fish in some larger size classes reside in this area, although the overall proportion in the NBS was small. The SSC encourages the author to further examine Pacific cod catches from trawl surveys conducted triennially by the National Marine Fisheries Service (NMFS) (1976-1991) and by the Alaska Department of Fish \& Game (1996 to the present) to monitor the distribution and abundance of red king crab and demersal fish (see: Hamazaki, T., Fair, L., Watson, L., Brennan, E., 2005. Analyses of Bering Sea bottom-trawl surveys in Norton Sound: absence of regime shift effect on epifauna and demersal fish. ICES Journal of Marine Science 62, 1597-1602). While the 2010 bottom trawl survey in the NBS found relatively few Pacific cod (3\% of total biomass), it is possible that the proportion of Pacific cod that are outside the standard survey area was higher in other years. A second possibility is that older Pacific cod migrate to nearshore areas to feed in the summer, making them unavailable to the survey." The JTS recommended postponing this examination until 2017, when another survey of the northern Bering Sea is scheduled.

SSC6 (12/15 minutes):"The SSC noted that the iteratively tuned, time-varying parameters in the model have not been updated since 2009. The author confirmed that the currently assumed standard deviations of two dev vectors (log of age-0 recruitment and a parameter corresponding to the ascending part of the selectivity curve) may no longer match the standard deviations of these vectors, which could contribute to retrospective bias. The SSC looks forward to a new paper on this issue that the author is preparing." The paper is in revision following initial journal review.

SSC7 (12/15 minutes): "While the model selection criteria proposed by the author are reasonable, we note that these criteria do not take into account the model fit itself. Model fit and retrospective performance should be more strongly considered in the selection of a final model for specifications." Although selection of a final model is not addressed in this preliminary assessment, retrospective analyses are presented for all models.

SSC8 (12/15 minutes): "Although the SSC has repeatedly stressed the need to incrementally evaluate model changes, the SSC did not intend this to imply an automatic preference for the status quo model (as implied by the authors criterion \#1) if alternatives with better performance are available." This comment will be addressed in the final assessment.

JTS1 (5/16 minutes): "For the BS, the subcommittee recommended that the following models be developed for this year's preliminary assessment:

- Model 1: BS Model 11.5, the final model from 2015 (same as the final models from 2011-2014)
- Model 2: Like BS Model 15.6, but simplified as follows:
o Weight abundance indices more heavily than sizecomps.
o Use the simplest selectivity form that gives a reasonable fit.
o Do not allow survey selectivity to vary with time.
o Do not allow survey catchability to vary with time.
o Force trawl survey selectivity to be asymptotic.
o Do not allow strange selectivity patterns.
o Use empirical weight at age.
- Model 3: Like BS Model 15.6, but including the IPHC longline survey data and other features, specifically:
o Do not allow strange selectivity patterns.
o Estimate catchability of new surveys internally with non-restrictive priors.
o Include additional data sets to increase confidence in model results.
o Include IPHC longline survey, with 'extra SD.'
- Model 4: Like Model 3 above, but including the NMFS longline survey instead of the IPHC longline survey.
- Model 5: Like Models 3 and 4 above, but including both the IPHC and NMFS longline survey data and two features not included in either Model 3 or 4, specifically:
o Start including fishery agecomp data.
o Use empirical weight at age.
- Model 6: Like Model 5 above, but including two features not included in Model 5, specifically:
o Use either Francis or harmonic mean weighting.
o Explore age-specific M (e.g., using Lorenzen function)."
All of the requested models are included in this preliminary assessment (see also comment SSC9). Note that some points in the above lists of features may be somewhat duplicative, but were included by the JTS in order to address specific comments made by CIE reviewers. For Model 6, harmonic mean weighting (Punt in press) and the age-specific natural mortality function proposed by Lorenzen $(1996,2011)$ were used. As noted in the JTS meeting minutes, the model numbers used above were intended just as placeholders, until final model numbers could be assigned, following the adopted model numbering convention (see comment SSC1). Application of the numbering convention resulted in the following model numbers:
$\begin{array}{lcccccc}\text { JTS "placeholder" model number: } & 1 & 2 & 3 & 4 & 5 & 6 \\ \text { Final model number: } & 11.5 & 16.1 & 16.2 & 16.3 & 16.4 & 16.5\end{array}$
JTS2 (5/16 minutes): "For the EBS, the JTS recommended that the following non-model analysis be conducted for this year's preliminary assessment:
- Non-model analysis 1: Verify that the trawl survey data sometimes include age 0 fish."

Although very rare ( 5 records in 1984 and 1 record in 2002), the trawl survey data do sometimes include age 0 fish, as confirmed this summer by AFSC RACE and Age and Growth personnel (pers. commun., Dan Nichol (RACE) and Delsa Anderl (Age and Growth)).

SSC9 (6/16 minutes): "The SSC accepts the JTS recommendations for models to bring forward in the 2016 assessment...." See comment JTS1.

SSC10 (6/16 minutes): "The SSC agrees with CIE recommendations to use all reasonable data sources that are available, although the use of the longline survey data in the model has been attempted in the past with little success. As the author noted, survey indices were generally negatively correlated with model-estimated biomass in past assessments. The use of 'extra SD' in the proposed models for both regions is a reasonable approach to deal with this issue." Internally estimated increments to the logscale standard errors for the IPHC and NMFS longline survey indices are reported in Table 2.1.8.

SSC11 (6/16 minutes): "The SSC encourages the use of empirical weight-at-age data in some of the model variants, but notes that this requires precise aging data." Empirical weight-at-age data are used in

Models 16.1, 16.4, and 16.5. Some issues involved in generating these data are discussed in the "Data" section.

SSC12 (6/16 minutes): "The SSC encourages the author to conduct a retrospective analysis across historically used models in addition to the standard retrospective analysis using the current model." The requested analysis is not included in this preliminary assessment. It may be noted that there have been no changes in the accepted model since 2011. Barring any changes in this request, the analysis will be included in the final assessment.

SSC13 (6/16 minutes):"The SSC encourages further work (outside the model) to examine potential causes for the apparent dome-shaped selectivity in most models. Research on these older 'missing' fish could include analysis of existing northern Bering Sea survey data, as noted in last December's minutes, and an analysis of slope survey data to examine if older fish descend to deeper waters as suggested in public testimony." See comment SSC5.

## Data

The data used in this preliminary assessment are identical to those used in last year's final assessment (Thompson 2015), except for:

- the addition of "empirical" weight-at-age data in Models 16.1, 16.4, and 16.5;
- the addition of IPHC survey data (abundance index and size composition) in Models 16.2, 16.4, and 16.5; and
- the addition of NMFS longline survey data (abundance index and size composition) in Models $16.3,16.4$, and 16.5 .

The following table summarizes the sources, types, and years of data included in the data file for one or more of the stock assessment models (italics denote data not included in last year's assessment):

| Source | Type | Years |
| :--- | :--- | :--- |
| Fishery | Catch biomass | $1977-2015$ |
| Fishery | Catch size composition | $1977-2015$ |
| Fishery | Catch per unit effort | $1991-2015$ |
| Fishery | Empirical weight at age | $2008-2011$ |
| EBS shelf bottom trawl survey | Relative abundance | $1982-2015$ |
| EBS shelf bottom trawl survey | Size composition | $1982-2015$ |
| EBS shelf bottom trawl survey | Age composition | $1994-2014$ |
| EBS shelf bottom trawl survey | Mean size at age | $1994-2014$ |
| EBS shelf bottom trawl survey | Empirical weight at age | $1998-2014$ |
| IPHC longline survey | Relative abundance | $1997-2014$ |
| IPHC longline survey | Size composition | $2008-2009,2011-2015$ |
| NMFS longline survey | Relative abundance | $1997-2015$ (odd years only) |
| NMFS longline survey | Size composition | $1997-2015$ (odd years only) |

Empirical weight-at-age estimates were computed using a two-stage bootstrap procedure (J. Ianelli, AFSC, pers. commun.) from the available age data, resulting in the values shown in Table 2.1.1. Four possible concerns might be noted with respect to these data:
6. No smoothing was applied to the estimates, even though they exhibit a fair amount of variability. For example, in the set of mid-year survey estimates, $18 \%$ of the cells differ from their respective age-specific time series average by $20 \%$ or more (not counting age 0 ); and in the set of fishery estimates, $34 \%$ of the cells differ from their respective age-specific time series average by $20 \%$ or more (not counting ages 0 or 1 ).
7. Age data exist for only 17 of the 34 years in the survey time series and only 4 of the 39 years in the fishery time series. Long-term averages were used for all years with no age data.
8. The fishery age data come primarily from the longline fishery, and may not be representative of the overall fishery.
9. Because the trawl survey takes place in summer, beginning-of-year population weights at age were calculated by averaging mid-year weight(age,year) and mid-year weight(age-1,year-1), implying that weight at age changes linearly within each one-year interval.

Relative abundance data from the IPHC and NMFS longline surveys are shown in Table 2.1.2, and size composition data from those two surveys are shown in Table 2.1.3.

Because the models presented in this preliminary assessment include various methods for tuning the input sample sizes for size and age composition data (see next section), a review of the current methods for specifying these input sample sizes is presented here: For the 2007 assessment, the harmonic means from a bootstrap analysis of the available fishery length data from 1990-2006 were computed. The harmonic means were smaller than the actual sample sizes, but still ranged well into the thousands. Analysis of the harmonic means revealed that, except when the actual sample size was very small (less than about 400), they tended to be very nearly proportional to the actual sample sizes, with the coefficient of proportionality dependent on whether the data were collected prior to 1999. For the years prior to 1999 the ratio was consistently very close to 0.16 , and for the years after 1998 the ratio was consistently very close to 0.34 . Thus, ever since the 2007 assessment (with some minor modifications through the years), input sample sizes have been set according to the following three-step process. First, records with actual sample sizes less than 400 are omitted. Second, sample sizes for fishery length compositions from years prior to 1999 are tentatively set at $16 \%$ of the actual sample sizes, and sample sizes for fishery length compositions since 1999 and sample sizes for all survey length compositions are tentatively set at $34 \%$ of the actual sample sizes. Third, all sample sizes are adjusted proportionally so that the average is 300 . Age composition input sample sizes are obtained by scaling the number of otoliths read so that the average is 300 .

## Model structures

All of the models presented in this preliminary assessment were developed using Stock Synthesis (SS, Methot and Wetzel 2013). The version used to run all models was SS V3.24u, as compiled on 8/29/2014. Stock Synthesis is programmed using the ADMB software package (Fournier et al. 2012). The user manual for SS V3.24s, along with a "change log" documenting revisions between V3.24s and V3.24u, is available at:
https://www.st.nmfs.noaa.gov/Assets/science_program/SS_User_Manual_3.24s.pdf.

## Developing the models requested by the Joint Team Subcommittee

Six models are presented in this preliminary assessment. Model 11.5 has been the accepted model since 2011. The other five models (Models 16.1-16.5) are all variants of Model 15.6, which was introduced in last year's preliminary assessment (where it was labeled "Model 6"). Details of Models 11.5 and 15.6 are described in their respective subsections below. The distinguishing features of Models 16.1-16.5 were listed above (see comment JPT1 under "Responses to SSC and Plan Team comments specific to Eastern Bering Sea Pacific cod," above).

In the minutes of its May 2016 meeting, the JTS recognized that some of the terms used in the descriptions of its requested models were somewhat subjective and that, in making those requests, the assessment author would need to determine:

1. How to measure the "weight" assigned to abundance indices and size composition data in the same units (Model 16.1).
2. What constitutes a "reasonable fit" to the size/age composition data (Model 16.1).
3. What constitutes a "strange" selectivity pattern (Models 16.1-16.5).

These issues were addressed as follows:

1. The relative "weight" assigned to abundance indices and size composition data was determined by comparing the average spawning biomasses from three models:
A. a model with a specified set of likelihood "emphasis" $(\lambda)$ values, with each $\lambda \geq 1.0$;
B. a model in which $\lambda$ for the abundance data was set equal to 0.01 while each $\lambda$ for the size composition data (fishery and survey) was left at the value specified in model A; and
C. a model in which each $\lambda$ for the size composition data (fishery and survey) was set equal to 0.01 while each $\lambda$ for the abundance data was left at the value specified in model B. Model B was taken to represent model A with the abundance data "turned off," while model C was taken to represent model A with the size composition data "turned off" (a $\lambda$ value of 0.01 rather than 0 was used for to represent "turning off" a data component because some parameters might prove inestimable if that data component were removed entirely). The abundance data in model A were determined to receive greater weight than the size composition data in that model if the absolute value of the proportional change in spawning biomass between models $B$ and $A$ exceeded the analogous value between models C and A . The JTS requested that this criterion (giving greater weight to abundance data than size composition data) be included in Model 16.1 only. As it turned out, the default $\lambda$ value of 1.0 for all data components was sufficient to satisfy this criterion, so no adjustments to any of the $\lambda$ values were necessary.
2. To focus on the ability of a particular functional form to fit the data, independent of the absolute values of the sample sizes specified for the associated multinomial distribution or $\lambda$ values, weighted coefficients of determination $\left(R^{2}\right)$, computed on both the raw and logit scales, were used to measure goodness of fit (the equations below are written in terms of age composition; the equations for size compositions are analogous):

$$
R^{2}=\sum_{y=y \min }^{y \max }\left(w_{y} \cdot\left(\frac{\sum_{a=0}^{a \max }\left(\text { Pobs }_{a, y}-\text { Pest }_{a, y}\right)^{2}}{1-\frac{\sum_{a=0}^{a x a x}}{}\left(\text { Pobs }_{a, y}-\text { Pobs }_{a v e, y}\right)^{2}}\right)\right),
$$

and

$$
R^{2}=\sum_{y=y \min }^{y \max }\left(w_{y} \cdot\left(1-\frac{\sum_{a=0}^{a \max }\left(\operatorname{logit}\left(\text { Pobs }_{a, y}\right)-\operatorname{logit}\left(\text { Pest }_{a, y}\right)\right)^{2}}{\sum_{a=0}^{\operatorname{amax}}\left(\operatorname{logit}\left(\text { Pobs }_{a, y}\right)-\operatorname{logit}\left(\text { Pobs }_{a v e, y}\right)\right)^{2}}\right)\right),
$$

where

$$
w_{y}=\frac{n_{y}}{\sum_{i=y \min }^{y \max } n_{i}}
$$

Pobs $_{a, y}$ represents the observed proportion at age $a$ in year $y$, Pobs $_{a v e, y}$ represents the average (across ages) observed proportion in year $y$, Pest $_{a, y}$ represents the estimated proportion at age $a$ in year $y$, and $n_{y}$ represents the specified multinomial sample size in year $y$. To guard against the possibility of achieving misleadingly high $R^{2}$ values by extending the size or age range beyond the sizes or ages actually observed, the data were filtered by removing all records with Pobs ${ }_{a, y}<$ 0.001 prior to computing the $R^{2}$ values. A fit was determined to be "reasonable" if it yielded both an $R^{2}$ value of at least 0.99 on the raw scale and an $R^{2}$ value of at least 0.70 on the logit scale. As with \#1 above, the JTS requested that this criterion (simplest selectivity function that gives a reasonable fit) be included in Model 16.1 only. Because the "random walk with respect to age" selectivity function gave a reasonable fit, the function was simplified in successive steps first by removing all time-variability, then by switching to a double-normal function, and finally by switching to a logistic function. The logistic function (for both the fishery and the survey) gave a reasonable fit to the fishery size composition data, the survey size composition data, and the survey age composition data, so it was retained as the final functional form.
3. In general, a "strange" selectivity pattern was defined here as one which was non-monotonic (i.e., where the signs of adjacent first differences changed), particularly if the first differences associated with sign changes were large (in absolute value), and particularly if sign changes in first differences occurred at relatively early ages. Specifically, an index of "strangeness" was defined as follows:
A. Age-specific weighting factors $P_{a}$ were calculated as the equilibrium unfished numbers at age expressed as a proportion of equilibrium unfished numbers.
B. For each year, age-specific first differences in selectivity $\Delta_{a, y}$ were calculated.
C. "Strangeness" was then calculated as:

$$
\left(\frac{1}{y \max -y \min +1}\right) \cdot \sum_{y=y \min }^{y \max } \sqrt{\sum_{a=2}^{\operatorname{amax}}\left(P_{a} \cdot\left(\left(\operatorname{sign}\left(\Delta_{a, y}\right) \neq \operatorname{sign}\left(\Delta_{a-1, y}\right)\right) \cdot\left(\Delta_{a}\right)^{2}\right)\right)}
$$

where the expression $\operatorname{sign}\left(\Delta_{a, y}\right) \neq \operatorname{sign}\left(\Delta_{a-1, y}\right)$ returned a value of 1 if the sign of $\Delta_{a, y}$ differed from the sign of $\Delta_{a-1, y}$ and a value of 0 otherwise. This index attains a minimum of 0 when selectivity is constant across age (or varies monotonically) and a maximum of 1 if selectivity alternates between values of 0 and 1 at all pairs of adjacent ages.
A time series of selectivity at age (for a given fleet) was determined to be "strange" if the index described above exceeded a value of 0.05. If a model produced a "strange" selectivity pattern, the standard deviations of the prior distributions for the selectivity parameters and the standard deviations of any selectivity dev vectors were decreased proportionally relative to the values estimated for Model 15.6 in last year's assessment until the threshold value of 0.05 was satisfied.

As in previous assessments, development of the final versions of all models included calculation of the Hessian matrix and a requirement that all models pass a "jitter" test of 50 runs. In the event that a jitter run produced a better value for the objective function than the base run, then:
4. The model was re-run starting from the final parameter file from the best jitter run.
5. The resulting new control file, with the parameter estimates from the best jitter run incorporated as starting values, became the new base run.
6. The entire process (starting with a new set of jitter runs) was repeated until no jitter run produced a better value for the objective function than the most recent base run.

One difference from previous assessments is that, for this preliminary assessment, an attempt was made to standardize the bounds within which individual parameters were "jittered." Specifically, once a model was ready to be subjected to the jitter test, the bounds for each parameter in the model were adjusted to match the $99.9 \%$ confidence interval (based on the normal approximation obtained by inverting the Hessian matrix). A jitter rate (equal to half the standard deviation of the logit-scale distribution from which "jittered" parameter values are drawn) was set at 1.0 for all models. Standardizing the jittering process in this manner may not explore parameter space as thoroughly as in previous assessments; however, it should make the jitter rate more interpretable, and show the extent to which the identified minimum (local or otherwise) is well behaved.

Except for selectivity parameters and annual catchability deviations (trawl survey only) in Models 16.216.5 and dev vectors in all models, all parameters were estimated with uniform prior distributions.

All selectivity devs were assumed to be additive (SS automatically assumes log recruitment devs to be additive).

Parameters estimated outside the assessment model (e.g., weight-at-length parameters, maturity-at-age parameters, ageing error matrix, trawl survey catchability in Model 11.5) were likewise described in last year's final assessment (Thompson 2015), and were not re-estimated for this preliminary assessment.

## Model 11.5: main features

Some of the main features characterizing Model 11.5 are as follow:

1. Age- and time-invariant natural mortality, estimated outside the model
2. Parameters governing time-invariant mean length at age estimated internally
3. Parameters governing width of length-at-age distribution (for a given mean) estimated internally
4. Ageing bias parameters estimated internally
5. Gear-and-season-specific catch and selectivity for the fisheries
6. Double normal selectivity for the fisheries and survey, with parameterization as follows:
7. beginning_of_peak_region (where the curve first reaches a value of 1.0)
8. width_of_peak_region (where the curve first departs from a value of 1.0)
9. ascending_width (equal to twice the variance of the underlying normal distribution)
10. descending_width (equal to twice the variance of the underlying normal distribution)
11. initial_selectivity (at minimum length/age)
12. final_selectivity (at maximum length/age)

All parameters except beginning_of_peak_region are transformed: The ascending_width and descending_width are log-transformed and the other three parameters are logit-transformed.
7. Length-based selectivity for the fisheries
8. Age-based selectivity for the survey
9. Fishery selectivity estimated for "blocks" of years
10. Survey selectivity constant over time, except with annual devs for the ascending_width parameter
11. Survey size composition data used in all years, including those years with age composition data (at the request of Plan Team members, inclusion of survey size composition data in all years was instituted in the 2011 assessment and has been retained ever since, based on the view that the costs of double-counting are outweighed by the benefits of including this information for estimation of growth parameters)
12. Fishery CPUE data included but not used for estimation
13. Mean size at age included but not used for estimation

## Model 11.5: iterative tuning

Iterative tuning of time-varying parameters
The standard deviations of the two dev vectors in Model 11.5 (the log of age 0 recruitment and the survey ascending_width parameter, both additive) were estimated iteratively during the 2009 assessment by tuning the specified $\sigma$ term for each vector to the standard deviation of the elements in that vector. Although this method is more justifiable than simply guessing at the value of $\sigma$, it is known to be biased low, and in the worst case may return a value of zero even when the true value is substantially greater than zero (Maunder and Deriso 2003, Thompson in prep.).

Per request of the BSAI Plan Team, the values of these $\sigma$ terms ( 0.57 and 0.07 , respectively) have been held constant in Model 11.5 and its predecessors ever since the 2009 assessment.

Iterative tuning of survey catchability
Survey catchability was estimated iteratively during the 2009 assessment by tuning $Q$ so that the average of the product of $Q$ and survey selectivity across the $60-81 \mathrm{~cm}$ size range matched the point estimate of 0.47 given by Nichol et al. (2007).

Per request of the BSAI Plan Team, this value of $Q(0.77)$ has been held constant in Model 11.5 and its predecessors ever since the 2009 assessment.

## Model 15.6: main features

Note that Model 15.6 was not among the models requested by the JTS and SSC for this preliminary assessment. However, it provides the starting point for Models 16.1-16.5, so it is appropriate to review its features.

Except for procedures related to iterative tuning (see next section), the main differences between Model 15.6 and Model 11.5 were as follow:

1. Each year consisted of a single season instead of five.
2. A single fishery was defined instead of nine season-and-gear-specific fisheries.
3. Composition data were given a weight of unity if the harmonic mean of the effective sample size was greater than the mean input sample size of 300 ; otherwise, composition data were weighted by tuning the mean input sample size to the harmonic mean of the effective sample size.
4. The survey was assumed to sample age 1 fish at true age 1.5 instead of 1.41667 .
5. Initial abundances were estimated for the first 20 age groups instead of the first three.
6. The natural mortality rate was estimated internally.
7. The SS feature known as "Fballpark" was turned off (this feature, which functions something like a very weak prior distribution on the fishing mortality rate in some specified year, did not appear to be providing any benefit in terms of model performance, and what little impact it had on resulting estimates was not easily justified).
8. The base value of survey catchability was estimated internally.
9. Survey catchability was allowed to vary annually.
10. Selectivity for both the fishery and the survey were allowed to vary annually.
11. Selectivity for both the fishery and survey was modeled using a random walk with respect to age (SS selectivity-at-age pattern \#17) instead of the usual double normal.
12. Selectivity at ages $9+$ was constrained to equal selectivity at age 8 for both the fishery and the survey.

## Model 15.6: iterative tuning

Note that the iterative tuning described in this section pertains to the development of Model 15.6 in last year's preliminary assessment. The values resulting from last year's tuning were, with a very few exceptions, retained for Models 16.1-16.5.

All iterative tuning procedures described below were undertaken simultaneously.

## Iterative tuning of prior distributions for selectivity parameters

Initially, the model was run with recruitment as the only time-varying quantity, with the standard deviation of log-scale recruitment estimated internally (i.e., as a free parameter), and with large standard deviations in the prior distributions for all selectivity parameters.

Once the initial model converged, a pair of transformed logistic curves was fit to the point estimates of the fishery and survey selectivity schedules (a transformed logistic curve was used because the selectivity parameters in pattern \#17 consist of the backward first differences of selectivity on the log scale, rather than selectivity itself; Thompson and Palsson 2013). The respective transformed logistic curve (fishery or survey) was then used to specify a new set of means for the selectivity prior distributions (one for each age). A constant (across age) prior standard deviation was then computed such that no age had a prior CV (on the selectivity scale, not the transformed scale) less than $50 \%$, and at least one age had a prior CV of exactly $50 \%$.

The model was then run with the new set of prior means and constant prior standard deviations (one for the fishery, one for the survey), then a new pair of transformed logistic curves was fit to the results, and the process was repeated until convergence was achieved.

## Iterative tuning of time-varying catchability

Although conceptually similar to a $d e v$ vector, SS treats each annual deviation in $\ln (\mathrm{Q})$ as a true parameter, with its own prior distribution. Because SS works in terms of $\ln (Q)$ rather than $Q$, normal prior distributions were assumed for all annual deviations. To be parsimonious, a single $\sigma$ was assumed for all such prior distributions.

Unlike the size composition or age composition data sets, the time series of survey abundance data includes not only a series of expected values, but a corresponding series of standard errors as well. This fact formed the basis for the iterative tuning of the $\sigma$ term for time-varying $Q$ in Model 15.6. The procedure involved iteratively adjusting $\sigma$ until the root-mean-squared-standardized-residual for survey abundance equaled unity.

Iterative tuning of time-varying parameters other than catchability
The following algorithm was used in Model 15.6 (Thompson in prep.; note that this is a multivariate generalization of one of the methods mentioned by Methot and Taylor (2011, viz., the third method listed on p. 1749)):

1. Set initial guesses for the os.
2. Run SS.
3. Compute the covariance matrix (V1) of the set of $d e v$ vectors (e.g., element $\{i, j\}$ is equal to the covariance between the subsets of the ith dev vector and the $j$ th dev vector consisting of years that those two vectors have in common).
4. Compute the covariance matrix of the parameters (the negative inverse of the Hessian matrix).
5. Extract the part of the covariance matrix of the parameters corresponding to the dev vectors, using only those years common to all dev vectors.
6. Average the values in the matrix obtained in step 5 across years to obtain an "average" covariance matrix (V2).
7. Compute the vector of os corresponding to $\mathbf{V} 1+\mathbf{V} 2$.
8. Return to step 2 and repeat until the os converge.

To speed the above algorithm, the os obtained in step 7 were sometimes substituted with values obtained by extrapolation or interpolation based on previous runs.

Unfortunately, given the way that selectivity pattern \#17 is implemented in SS, large gradients can result, particularly if sufficiently large devs occur at or adjacent to the age of peak selectivity. In the event that a large gradient appeared to be unavoidable during the tuning process, selectivity $d e v$ vectors were eliminated, one at a time (usually starting at the oldest ages and working downward), until the large gradients disappeared.

## Results

## Overview

The following table summarizes the status of the stock as estimated by the six models ("Value" is the point estimate, "SD" is the standard deviation of the point estimate, "CV" is the ratio of SD to the point estimate, "FSB 2016" is female spawning biomass in 2016 (t), and "Bratio 2016" is the ratio of FSB 2016 to $B_{100 \%}$; color shading for FSB 2016 and Bratio 2016 extends from red (low) to green (high) for each quantity):

|  | Model 11.5 |  |  | Model 16.1 |  |  | Model 16.2 |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Quantity | Value | SD | CV | Value | SD | CV | Value | SD | CV |
| FSB 2016 | 457,341 | 30,739 | 0.07 | 414,941 | 40,176 | 0.10 | 399,149 | 67,976 | 0.17 |
| Bratio 2016 | 0.61 | 0.03 | 0.06 | 0.57 | 0.06 | 0.10 | 0.46 | 0.07 | 0.15 |


|  | Model 16.3 |  | Model 16.4 |  |  | Model 16.5 |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Quantity | Value | SD | CV | Value | SD | CV | Value | SD |
| CV |  |  |  |  |  |  |  |  |
| FSB 2016 | 196,753 | 25,016 | 0.13 | 154,877 | 15,482 | 0.10 | 133,142 | 12,167 |
| Bratio 2016 | 0.21 | 0.03 | 0.14 | 0.14 | 0.02 | 0.12 | 0.09 | 0.01 |

The six models span wide ranges for these quantities. Estimates of FSB 2016 range from 133,000 t (Model 16.5) to 457,000 t (Model 11.5), and estimates of Bratio 2016 range from 0.09 (Model 16.5) to 0.61 (Model Model 11.5). The quantities FSB 2016 and Bratio 2016 tend to covary directly in these models.

## Goodness of fit

Objective function values and parameter counts are shown for each model in Table 2.1.4a, and multipliers used to adjust multinomial sample sizes are shown in Table 2.1.4b. Objective function values are not
directly comparable across models, because different data files are used for some models, different constraints are imposed, and the number and types of parameters vary considerably.

Figure 2.1.1a shows the fits of all six models to the trawl survey abundance data; Figure 2.1.1b shows the fits of Models 16.2, 16.4, and 16.5 to the IPHC longline survey abundance data; and Figure 2.1.1c shows the fits of Models 16.3, 16.4, and 16.5 to the NMFS longline survey abundance data.

Table 2.1.5 shows goodness of fit for the survey abundance data. Four measures are shown: root mean squared error (for comparison, the average log-scale standard error "oave" is also shown), mean normalized residual, standard deviation of normalized residuals, and correlation (observed:estimated). For the trawl survey data, Models 16.2-16.5 all give root mean squared errors close to oave. Models 16.1-16.5 all give mean normalized residuals close to zero, standard deviation of normalized residuals close to unity, and correlations greater close to 0.90 or better. The three models that use the IPHC longline survey data all give mean normalized residuals close to zero and standard deviation of normalized residuals close to unity (note that these models inflate the input $\sigma$ values by an internally estimated amount, and the resulting estimates of oave are fairly high, in the 0.42-0.46 range). However, as with previous attempts to use the IPHC longline survey data, all three of these models give negative correlations. The three models that use the NMFS longline survey data all fit those data fairly well, although the mean normalized residuals from all three of these models is substantially negative, ranging from -0.14 to -0.22 (note that, although these models were all given the opportunity to inflate the input $\sigma$ values by an internally estimated amount, Model 16.3 estimated this additional amount at a very small value ( 0.01 ), and the estimates from Models 16.4 and 16.5 tended to become pinned at the lower bound of zero, so estimation of this additional $\sigma$ was ultimately turned off in the latter two models).

Sample size ratios for the size composition data are shown in Table 2.1.6 (note that input sample sizes are the same for all models except Model 16.5). These results can be summarized as follows:

- Measured as the ratio of the arithmetic mean effective sample size to the arithmetic mean input, the models give values well in excess of unity for all components.
- Measured as the ratio of the harmonic mean effective sample size to the arithmetic mean input sample size, all models give noticeably smaller values, but still in excess of unity in most cases. Exceptions consist of the Aug-Dec longline fishery in Model 11.5, and all components in Model 16.5 , which was tuned explicitly so as to set these ratios equal to unity.

Sample size ratios for the survey age composition data are shown in Table 2.1.7a (all models) and for the fishery age composition data in Table 2.1.7b (Models 16.4 and 16.5 only). Note that input sample sizes for the survey data differ for several models: For Models 11.5 and 16.1, input sample sizes were scaled to the conventional mean of 300 ; for Models 16.2-16.4, input sample sizes were left at the values tuned in last year's assessment for Model 15.6 so that $\mathrm{H}(\mathrm{Neff}) / \mathrm{A}(\mathrm{Ninp})=$; and for Model 16.5, arithmetic mean input sample sizes were tuned in this year's assessment so that H (Neff)/A(Ninp)=1. The input sample sizes for the fishery data also differ between the two models that use those data: For Model 16.4, mean input sample sizes were assumed equal to mean input sample size for the survey agecomp data; while for Model 16.5, input sample sizes were tuned in this year's assessment so that $\mathrm{H}(\mathrm{Neff}) / \mathrm{A}(\mathrm{Ninp})=1$. The results can be summarized as follows:

- Measured as the ratio of the arithmetic means, Models 16.2-16.5 give values greater than unity for the survey age composition data (Models 11.5 and 16.1 do not), and Model 16.5 is the only one of the two models using fishery age composition data to achieve a value greater than unity.
- Measured as the ratio of the harmonic mean effective sample size to the arithmetic mean input sample size, Model 16.5 gives values essentially equal to unity for both the survey and fishery
age composition data (as this was the tuning criterion for that model), while the other models all give values much less than unity. Note that Punt (in press) concluded that the harmonic mean was a much more appropriate numerator than the arithmetic mean.

Figure 2.1.2 shows the fits to the survey age composition data (all models), and Figure 2.1 .3 shows the fits to the fishery age composition data (Models 16.4 and 16.5 only).

## Parameter estimates, time series, and retrospective analysis

Table 2.1.8 lists key parameters estimated internally in at least one of the models, along with their standard deviations.

In Model 16.5, the natural mortality rate $M$ varies as a function of age, following the approach described by Lorenzen $(1996,2011)$. The entry for this model in Table 2.1.8 corresponds to the value at the age at $50 \%$ maturity (rounded to the nearest integer, 5). The full schedule of $M$ values for Model 16.5 is shown below:

| Age: | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $M:$ | 1.022 | 0.548 | 0.337 | 0.259 | 0.218 | 0.194 | 0.178 | 0.167 | 0.159 | 0.153 | 0.149 |


| Age: | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |

$M:$

The estimates of log catchability for the trawl survey shown in Table 2.1.8 map into the following estimates of catchability on the natural scale, spanning the range 0.643 (Model 16.1) to 1.590 (Model 16.5):

| Model 11.5 |  | Model 16.1 |  | Model 16.2 |  | Model 16.3 |  | Model 16.4 |  | Model 16.5 |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
| Est. | SD | Est. | SD | Est. | SD | Est. | SD | Est. | SD | Est. |  |
| 0.770 | $\mathrm{n} / \mathrm{a}$ | 0.643 | 0.063 | 1.050 | 0.108 | 1.581 | 0.075 | 1.343 | 0.065 | 1.590 |  |

Selectivity schedules are plotted for the fishery in Figure 2.1.4, the trawl survey in Figure 2.1.5a, the IPHC longline survey in Figure 2.1.5b, and the NMFS longline survey in Figure 2.1.5c. All models estimate strongly domed trawl survey selectivity schedules, which is difficult to reconcile with the results of field experiments summarized by Weinberg et al. (2016).

Time series estimated by the models are shown for total biomass, female spawning biomass relative to $B_{100 \%}$, age 0 recruitment, and fishing mortality relative to $F_{40 \%}$ in Figures 2.1.6, 2.1.7, 2.1.8, and 2.1.9, respectively.

Figure 2.1 .10 shows 10 -year retrospectives of spawning biomass for each of the models. Mohn's $\rho$ (revised) values for the models are shown below:

$$
\begin{array}{cccccc}
\text { Model 11.5 } & \text { Model 16.1 } & \text { Model 16.2 } & \text { Model 16.3 } & \text { Model 16.4 } & \text { Model } 16.5 \\
0.475 & 0.108 & 0.122 & -0.069 & 0.047 & 0.130
\end{array}
$$

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## Tables

Table 2.1.1a—Empirical weight at age for the population (kg). Weights in years with no data were assumed equal to the time series average.
Mid-year population (assumed to be represented by the survey)

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 0.00998 | 0.03031 | 0.28786 | 0.57498 | 1.34596 | 2.41074 | 3.63180 | 4.21474 | 6.07145 | 9.48271 | 9.63297 | 10.35847 | 10.34591 |
| 1999 | 0.00899 | 0.02975 | 0.23180 | 0.64063 | 1.00586 | 1.94912 | 3.19931 | 4.24325 | 5.92678 | 6.62555 | 10.28628 | 9.30312 | 11.01461 |
| 2000 | 0.00923 | 0.02719 | 0.26119 | 0.55903 | 1.15590 | 1.75550 | 2.38551 | 4.65000 | 4.96850 | 7.55933 | 7.04082 | 6.69292 | 11.11449 |
| 2001 | 0.01002 | 0.04835 | 0.29901 | 0.50036 | 1.20808 | 1.89331 | 2.69627 | 3.39956 | 5.52989 | 7.36904 | 5.72057 | 8.71575 | 10.28275 |
| 2002 | 0.00980 | 0.03695 | 0.25876 | 0.49530 | 1.08671 | 1.88860 | 2.87333 | 3.85336 | 4.53517 | 6.51294 | 10.38147 | 10.12309 | 11.28232 |
| 2003 | 0.00999 | 0.05025 | 0.26101 | 0.74333 | 1.27478 | 2.11556 | 3.38217 | 4.36719 | 5.33931 | 7.32482 | 7.66614 | 7.54419 | 6.11988 |
| 2004 | 0.01015 | 0.04374 | 0.26757 | 0.56628 | 1.30774 | 2.12083 | 3.23492 | 4.16120 | 5.16134 | 7.67440 | 8.71412 | 8.39726 | 11.14933 |
| 2005 | 0.00973 | 0.05328 | 0.17234 | 0.60838 | 1.23215 | 2.05120 | 3.08502 | 4.52856 | 5.96756 | 6.86777 | 9.20336 | 8.45074 | 10.31994 |
| 2006 | 0.00968 | 0.02849 | 0.27966 | 0.58066 | 1.14618 | 1.91756 | 3.11939 | 4.68658 | 6.79608 | 8.00201 | 8.82361 | 10.45918 | 11.62473 |
| 2007 | 0.00973 | 0.02702 | 0.28484 | 0.72057 | 1.44073 | 2.41451 | 3.53216 | 5.01613 | 6.90555 | 7.39105 | 10.65904 | 9.62044 | 9.89080 |
| 2008 | 0.00985 | 0.02844 | 0.24745 | 0.71837 | 1.68031 | 2.59784 | 3.36087 | 4.60989 | 6.17281 | 6.84603 | 8.54395 | 10.83814 | 9.66511 |
| 2009 | 0.00949 | 0.02148 | 0.27761 | 0.76664 | 1.45560 | 2.34835 | 3.25543 | 4.21250 | 5.32347 | 6.70273 | 8.77372 | 8.44027 | 9.28363 |
| 2010 | 0.00972 | 0.02982 | 0.26814 | 0.84713 | 1.69584 | 2.33270 | 3.32758 | 4.10257 | 6.34880 | 6.54702 | 9.02960 | 8.11057 | 11.81749 |
| 2011 | 0.00979 | 0.05044 | 0.35786 | 0.88458 | 1.70856 | 2.79529 | 3.63364 | 4.59066 | 5.51827 | 7.80137 | 7.22967 | 7.33689 | 11.18761 |
| 2012 | 0.00984 | 0.02155 | 0.31056 | 0.90135 | 1.62013 | 2.50125 | 3.58963 | 4.38997 | 6.08762 | 6.56512 | 9.62029 | 9.96183 | 10.90289 |
| 2013 | 0.00968 | 0.02978 | 0.22017 | 0.87182 | 1.38144 | 2.67502 | 3.34309 | 4.96482 | 5.40016 | 6.77607 | 8.93127 | 7.92271 | 10.71269 |
| 2014 | 0.01000 | 0.04617 | 0.31459 | 0.90396 | 1.48265 | 2.56694 | 3.47574 | 4.15903 | 5.91011 | 7.44386 | 8.21912 | 10.23339 | 8.25589 |
| Ave: | 0.00974 | 0.03651 | 0.27661 | 0.69849 | 1.36889 | 2.26085 | 3.25998 | 4.42322 | 5.85840 | 7.44757 | 8.91945 | 9.10880 | 10.28900 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 0 | 0.01986 | 0.13105 | 0.46425 | 0.79042 | 1.64754 | 2.80502 | 3.93752 | 5.07076 | 6.34850 | 9.88449 | 9.46805 | 10.68654 |
| 2000 | 0 | 0.01809 | 0.14547 | 0.39542 | 0.89826 | 1.38068 | 2.16731 | 3.92465 | 4.60587 | 6.74305 | 6.83318 | 8.48960 | 10.20881 |
| 2001 | 0 | 0.02879 | 0.16310 | 0.38077 | 0.88356 | 1.52460 | 2.22588 | 2.89254 | 5.08994 | 6.16877 | 6.63995 | 7.87829 | 8.48784 |
| 2002 | 0 | 0.02348 | 0.15356 | 0.39715 | 0.79353 | 1.54834 | 2.38332 | 3.27481 | 3.96737 | 6.02142 | 8.87525 | 7.92183 | 9.99904 |
| 2003 | 0 | 0.03003 | 0.14898 | 0.50104 | 0.88504 | 1.60113 | 2.63539 | 3.62026 | 4.59633 | 5.93000 | 7.08954 | 8.96283 | 8.12148 |
| 2004 | 0 | 0.02686 | 0.15891 | 0.41364 | 1.02554 | 1.69780 | 2.67524 | 3.77169 | 4.76426 | 6.50685 | 8.01947 | 8.03170 | 9.34676 |
| 2005 | 0 | 0.03172 | 0.10804 | 0.43797 | 0.89921 | 1.67947 | 2.60293 | 3.88174 | 5.06438 | 6.01455 | 8.43888 | 8.58243 | 9.35860 |
| 2006 | 0 | 0.01911 | 0.16647 | 0.37650 | 0.87728 | 1.57486 | 2.58529 | 3.88580 | 5.66232 | 6.98479 | 7.84569 | 9.83127 | 10.03773 |
| 2007 | 0 | 0.01835 | 0.15667 | 0.50011 | 1.01070 | 1.78035 | 2.72486 | 4.06776 | 5.79606 | 7.09357 | 9.33052 | 9.22202 | 10.17499 |
| 2008 | 0 | 0.01908 | 0.13723 | 0.50161 | 1.20044 | 2.01929 | 2.88769 | 4.07103 | 5.59447 | 6.87579 | 7.96750 | 10.74859 | 9.64277 |
| 2009 | 0 | 0.01566 | 0.15302 | 0.50704 | 1.08699 | 2.01433 | 2.92663 | 3.78669 | 4.96668 | 6.43777 | 7.80988 | 8.49211 | 10.06088 |
| 2010 | 0 | 0.01966 | 0.14481 | 0.56237 | 1.23124 | 1.89415 | 2.83796 | 3.67900 | 5.28065 | 5.93525 | 7.86616 | 8.44215 | 10.12888 |
| 2011 | 0 | 0.03008 | 0.19384 | 0.57636 | 1.27785 | 2.24557 | 2.98317 | 3.95912 | 4.81042 | 7.07509 | 6.88835 | 8.18324 | 9.64909 |
| 2012 | 0 | 0.01567 | 0.18050 | 0.62961 | 1.25236 | 2.10491 | 3.19246 | 4.01181 | 5.33914 | 6.04170 | 8.71083 | 8.59575 | 9.11989 |
| 2013 | 0 | 0.01981 | 0.12086 | 0.59119 | 1.14140 | 2.14758 | 2.92217 | 4.27722 | 4.89507 | 6.43185 | 7.74820 | 8.77150 | 10.33726 |
| 2014 | 0 | 0.02793 | 0.17219 | 0.56206 | 1.17724 | 1.97419 | 3.07538 | 3.75106 | 5.43746 | 6.42201 | 7.49760 | 9.58233 | 8.08930 |
| Ave: | 0 | 0.02276 | 0.15217 | 0.48732 | 1.02694 | 1.80217 | 2.72692 | 3.79954 | 5.05883 | 6.43943 | 7.96534 | 8.82523 | 9.59062 |

Table 2.1.1b—Empirical weight at age for the fishery (kg). Weights at age in years with no data were assumed equal to the time series average.

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 0 | 0 | 0.00066 | 1.42044 | 2.00646 | 2.93810 | 3.78537 | 5.02224 | 6.66598 | 7.14621 | 8.50707 | 10.00366 | 5.22370 |
| 2009 | 0 | 0 | 0.52358 | 1.48214 | 2.13895 | 3.09177 | 3.98118 | 5.25889 | 5.53492 | 8.92676 | 8.71459 | 7.87592 | 7.99262 |
| 2010 | 0 | 0 | 0.78678 | 1.63473 | 2.33971 | 3.04616 | 3.96101 | 5.37651 | 5.92141 | 5.51816 | 11.94570 | 3.82506 | 4.14191 |
| 2011 | 0 | 0 | 0.00066 | 1.27767 | 2.21042 | 3.24410 | 4.25569 | 5.63710 | 7.52856 | 6.17703 | 3.01784 | 4.44490 | 3.53656 |
| Ave: | 0 | 0 | 0.65518 | 1.45374 | 2.17388 | 3.08003 | 3.99581 | 5.32368 | 6.41272 | 6.94204 | 8.04630 | 6.53738 | 5.22370 |

Table 2.1.2—Relative abundance data for the IPHC and NMFS longline surveys, with log-scale standard errors ( $\sigma$ ). Note that the $\sigma$ values shown here may be incremented by an amount estimated by any of the models that use these data (Models 16.2-16.5).

| IPHC longline survey |  |  |
| ---: | ---: | ---: |
| Year |  | RPN |
| 1997 | 61,309 | 0.062 |
| 1998 | 85,429 | 0.115 |
| 1999 | 12,907 | 0.294 |
| 2000 | 72,237 | 0.097 |
| 2001 | 85,096 | 0.093 |
| 2002 | 101,998 | 0.107 |
| 2003 | 111,880 | 0.079 |
| 2004 | 116,604 | 0.097 |
| 2005 | 67,446 | 0.092 |
| 2006 | 109,217 | 0.083 |
| 2007 | 107,141 | 0.083 |
| 2008 | 114,508 | 0.077 |
| 2009 | 104,931 | 0.092 |
| 2010 | 76,881 | 0.112 |
| 2011 | 75,284 | 0.094 |
| 2012 | 78,135 | 0.083 |
| 2013 | 84,194 | 0.078 |
| 2014 | 87,472 | 0.062 |


| NMFS longline survey |  |  |
| ---: | ---: | ---: |
| Year | RPN | $\sigma$ |
| 1997 | 174,388 | 0.108 |
| 1999 | 122,984 | 0.106 |
| 2001 | 142,531 | 0.132 |
| 2003 | 173,070 | 0.115 |
| 2005 | 89,561 | 0.216 |
| 2007 | 102,653 | 0.146 |
| 2009 | 82,798 | 0.231 |
| 2011 | 120,673 | 0.188 |
| 2013 | 154,310 | 0.244 |
| 2015 | 125,796 | 0.206 |

Table 2.1.3a-Size (cm) composition data from the IPHC longline survey. No fish were observed at lengths smaller than 21 cm .

| Len | 2008 | 2009 | 2011 | 2012 | 2013 | 2014 | 2015 | Len | 2008 | 2009 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 71 | 141 | 180 | 149 | 162 | 338 | 241 | 343 |
| 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 72 | 165 | 158 | 154 | 163 | 323 | 235 | 287 |
| 23 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 73 | 170 | 145 | 168 | 164 | 294 | 223 | 271 |
| 24 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 74 | 145 | 139 | 125 | 131 | 235 | 225 | 251 |
| 25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 75 | 125 | 135 | 123 | 141 | 207 | 238 | 203 |
| 26 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 76 | 103 | 109 | 93 | 125 | 156 | 177 | 177 |
| 27 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 77 | 114 | 142 | 82 | 118 | 173 | 187 | 149 |
| 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 78 | 107 | 114 | 59 | 105 | 130 | 185 | 144 |
| 29 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 79 | 101 | 103 | 45 | 86 | 100 | 138 | 127 |
| 30 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 80 | 99 | 92 | 51 | 69 | 97 | 135 | 120 |
| 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 81 | 75 | 75 | 50 | 69 | 76 | 100 | 112 |
| 32 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 82 | 94 | 97 | 48 | 59 | 86 | 106 | 98 |
| 33 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 83 | 106 | 77 | 47 | 50 | 63 | 77 | 93 |
| 34 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 84 | 93 | 83 | 42 | 46 | 51 | 56 | 75 |
| 35 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 85 | 75 | 84 | 35 | 52 | 57 | 60 | 76 |
| 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 86 | 91 | 69 | 39 | 34 | 50 | 51 | 73 |
| 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 87 | 101 | 76 | 39 | 34 | 37 | 40 | 62 |
| 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 88 | 96 | 78 | 33 | 31 | 39 | 34 | 51 |
| 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 89 | 75 | 71 | 17 | 46 | 25 | 20 | 55 |
| 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 90 | 97 | 61 | 29 | 45 | 28 | 30 | 48 |
| 41 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 91 | 93 | 66 | 29 | 28 | 26 | 21 | 34 |
| 42 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 92 | 91 | 57 | 28 | 22 | 28 | 17 | 28 |
| 43 | 0 | 4 | 1 | 0 | 2 | 1 | 0 | 93 | 87 | 68 | 17 | 33 | 31 | 20 | 25 |
| 44 | 1 | 4 | 3 | 2 | 3 | 0 | 1 | 94 | 81 | 58 | 14 | 29 | 13 | 20 | 12 |
| 45 | 1 | 4 | 4 | 2 | 1 | 2 | 3 | 95 | 74 | 73 | 16 | 27 | 16 | 19 | 18 |
| 46 | 3 | 17 | 2 | 2 | 0 | 2 | 2 | 96 | 55 | 54 | 18 | 15 | 12 | 11 | 12 |
| 47 | 4 | 18 | 8 | 4 | 4 | 4 | 7 | 97 | 74 | 68 | 21 | 13 | 14 | 9 | 12 |
| 48 | 4 | 28 | 4 | 6 | 5 | 14 | 9 | 98 | 64 | 39 | 24 | 14 | 11 | 13 | 10 |
| 49 | 7 | 23 | 11 | 8 | 13 | 7 | 23 | 99 | 51 | 60 | 14 | 17 | 12 | 7 | 11 |
| 50 | 6 | 40 | 17 | 9 | 10 | 19 | 25 | 100 | 44 | 40 | 20 | 15 | 5 | 2 | 14 |
| 51 | 12 | 47 | 15 | 21 | 16 | 20 | 42 | 101 | 39 | 45 | 8 | 8 | 9 | 6 | 7 |
| 52 | 15 | 48 | 25 | 44 | 36 | 30 | 34 | 102 | 23 | 43 | 9 | 16 | 4 | 4 | 9 |
| 53 | 16 | 63 | 20 | 61 | 33 | 27 | 60 | 103 | 15 | 38 | 8 | 15 | 7 | 3 |  |
| 54 | 22 | 49 | 17 | 85 | 35 | 43 | 97 | 104 | 18 | 18 | 6 | 6 | 3 | 2 | 3 |
| 55 | 42 | 58 | 37 | 101 | 55 | 65 | 91 | 105 | 17 | 23 | 11 | 5 | 5 | 2 | 2 |
| 56 | 31 | 69 | 47 | 101 | 61 | 64 | 125 | 106 | 7 | 10 | 6 | 1 | 4 | 0 | 2 |
| 57 | 67 | 90 | 47 | 109 | 105 | 94 | 179 | 107 | 7 | 16 | 4 | 6 | 1 | 1 | 2 |
| 58 | 69 | 104 | 76 | 139 | 128 | 116 | 210 | 108 | 3 | 11 | 3 | 2 | 2 | 0 | 0 |
| 59 | 75 | 137 | 85 | 127 | 154 | 143 | 246 | 109 | 2 | 5 | 7 | 1 | 0 | 0 | 0 |
| 60 | 101 | 126 | 111 | 125 | 204 | 189 | 260 | 110 | 0 | 1 | 3 | 0 | 0 | 0 | 0 |
| 61 | 113 | 176 | 146 | 164 | 238 | 222 | 293 | 111 | 2 | 3 | 1 | 0 | 1 | 1 | 0 |
| 62 | 156 | 173 | 154 | 120 | 277 | 275 | 307 | 112 | 3 | 2 | 1 | 0 | 0 | 1 |  |
| 63 | 161 | 195 | 164 | 174 | 345 | 250 | 289 | 113 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 64 | 142 | 186 | 167 | 166 | 343 | 260 | 278 | 114 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 65 | 160 | 204 | 184 | 204 | 389 | 288 | 270 | 115 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| 66 | 154 | 187 | 220 | 155 | 439 | 240 | 281 | 116 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 67 | 154 | 194 | 235 | 189 | 415 | 232 | 293 | 117 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 68 | 179 | 203 | 193 | 168 | 441 | 246 | 264 | 118 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 69 | 188 | 206 | 210 | 171 | 389 | 229 | 271 | 119 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 70 | 186 | 183 | 201 | 182 | 400 | 242 | 252 | 120 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 2.1.3b—Size (cm) composition data from the NMFS longline survey (page 1 of 2). No fish were observed at lengths smaller than 21 cm .

| Len | 1997 | 1999 | 2001 | 2003 | 2005 | 2007 | 2009 | 2011 | 2013 | 2015 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 23 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 26 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 29 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 30 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 32 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 33 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 34 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 35 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 37 | 0 | 1 | 0 | 1 | 0 | 0 | 1 | 2 | 0 | 0 |
| 38 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 2 | 1 | 0 |
| 39 | 0 | 1 | 4 | 1 | 0 | 2 | 3 | 1 | 1 | 0 |
| 40 | 0 | 3 | 2 | 0 | 3 | 2 | 9 | 6 | 0 | 0 |
| 41 | 0 | 7 | 4 | 13 | 5 | 5 | 14 | 17 | 2 | 1 |
| 42 | 6 | 6 | 5 | 15 | 2 | 9 | 26 | 32 | 2 | 2 |
| 43 | 1 | 40 | 12 | 24 | 9 | 29 | 44 | 66 | 1 | 1 |
| 44 | 6 | 39 | 12 | 40 | 15 | 49 | 88 | 130 | 8 | 4 |
| 45 | 4 | 80 | 21 | 74 | 15 | 70 | 112 | 184 | 6 | 15 |
| 46 | 10 | 126 | 30 | 93 | 22 | 95 | 184 | 199 | 20 | 25 |
| 47 | 21 | 191 | 46 | 137 | 16 | 118 | 217 | 225 | 30 | 45 |
| 48 | 28 | 196 | 57 | 179 | 48 | 143 | 215 | 189 | 71 | 75 |
| 49 | 48 | 238 | 90 | 258 | 37 | 178 | 259 | 207 | 89 | 107 |
| 50 | 70 | 260 | 83 | 273 | 79 | 150 | 282 | 213 | 102 | 153 |
| 51 | 89 | 250 | 104 | 367 | 101 | 202 | 270 | 196 | 141 | 183 |
| 52 | 113 | 275 | 157 | 388 | 117 | 191 | 240 | 178 | 161 | 228 |
| 53 | 164 | 268 | 199 | 413 | 158 | 197 | 215 | 177 | 163 | 297 |
| 54 | 160 | 251 | 210 | 460 | 152 | 154 | 244 | 183 | 168 | 355 |
| 55 | 227 | 316 | 263 | 447 | 175 | 161 | 212 | 217 | 151 | 431 |
| 56 | 216 | 356 | 315 | 470 | 163 | 192 | 204 | 242 | 143 | 522 |
| 57 | 232 | 346 | 335 | 437 | 201 | 176 | 215 | 288 | 151 | 538 |
| 58 | 244 | 303 | 354 | 398 | 215 | 226 | 219 | 330 | 178 | 604 |
| 59 | 270 | 322 | 384 | 434 | 229 | 216 | 246 | 348 | 195 | 530 |
| 60 | 274 | 362 | 412 | 464 | 247 | 243 | 254 | 406 | 238 | 520 |
| 61 | 338 | 417 | 440 | 473 | 248 | 254 | 278 | 445 | 305 | 404 |
| 62 | 385 | 401 | 480 | 501 | 273 | 244 | 296 | 442 | 388 | 428 |
| 63 | 410 | 457 | 482 | 484 | 274 | 301 | 277 | 412 | 475 | 386 |
| 64 | 423 | 428 | 488 | 479 | 317 | 265 | 270 | 386 | 477 | 384 |
| 65 | 546 | 498 | 517 | 427 | 297 | 262 | 260 | 384 | 535 | 345 |
| 66 | 479 | 439 | 496 | 350 | 316 | 236 | 225 | 358 | 513 | 321 |
| 67 | 561 | 404 | 577 | 325 | 306 | 243 | 187 | 317 | 529 | 283 |
| 68 | 602 | 367 | 558 | 276 | 263 | 188 | 167 | 269 | 533 | 258 |
| 69 | 581 | 338 | 489 | 209 | 273 | 204 | 174 | 223 | 483 | 250 |
| 70 | 481 | 296 | 447 | 187 | 272 | 194 | 127 | 167 | 385 | 271 |
|  |  |  |  |  |  |  |  |  |  |  |

Table 2.1.3b—Size (cm) composition data from the NMFS longline survey (page 2 of 2).

| Len | 1997 | 1999 | 2001 | 2003 | 2005 | 2007 | 2009 | 2011 | 2013 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 71 | 490 | 255 | 376 | 151 | 225 | 136 | 130 | 162 | 313 | 232 |
| 72 | 395 | 214 | 380 | 113 | 197 | 156 | 113 | 125 | 267 | 189 |
| 73 | 389 | 197 | 280 | 97 | 171 | 143 | 116 | 99 | 182 | 164 |
| 74 | 276 | 160 | 245 | 95 | 181 | 136 | 112 | 52 | 164 | 152 |
| 75 | 236 | 167 | 180 | 66 | 144 | 99 | 93 | 52 | 109 | 121 |
| 76 | 164 | 115 | 142 | 52 | 102 | 77 | 78 | 39 | 72 | 102 |
| 77 | 144 | 87 | 111 | 48 | 128 | 95 | 64 | 26 | 45 | 63 |
| 78 | 101 | 78 | 123 | 37 | 67 | 83 | 50 | 18 | 35 | 75 |
| 79 | 70 | 54 | 80 | 36 | 74 | 76 | 49 | 11 | 38 | 57 |
| 80 | 66 | 46 | 59 | 30 | 68 | 62 | 46 | 12 | 28 | 51 |
| 81 | 55 | 36 | 52 | 30 | 55 | 57 | 27 | 11 | 20 | 47 |
| 82 | 32 | 28 | 37 | 31 | 44 | 58 | 25 | 9 | 8 | 44 |
| 83 | 28 | 19 | 30 | 18 | 30 | 66 | 31 | 7 | 12 | 25 |
| 84 | 29 | 20 | 25 | 8 | 37 | 41 | 23 | 9 | 5 | 23 |
| 85 | 24 | 15 | 28 | 10 | 18 | 42 | 18 | 4 | 13 | 25 |
| 86 | 17 | 13 | 18 | 9 | 21 | 46 | 10 | 4 | 5 | 20 |
| 87 | 23 | 4 | 8 | 10 | 15 | 39 | 7 | 5 | 6 | 18 |
| 88 | 16 | 16 | 6 | 8 | 13 | 43 | 7 | 8 | 3 | 10 |
| 89 | 16 | 8 | 15 | 5 | 15 | 43 | 9 | 7 | 4 | 16 |
| 90 | 18 | 13 | 10 | 4 | 13 | 31 | 7 | 2 | 4 | 8 |
| 91 | 12 | 3 | 5 | 6 | 9 | 30 | 7 | 6 | 0 | 7 |
| 92 | 7 | 5 | 2 | 4 | 6 | 22 | 10 | 5 | 4 | 9 |
| 93 | 8 | 3 | 3 | 2 | 7 | 26 | 9 | 1 | 2 | 4 |
| 94 | 9 | 3 | 3 | 3 | 5 | 23 | 7 | 2 | 4 | 7 |
| 95 | 13 | 1 | 0 | 2 | 4 | 25 | 3 | 4 | 2 | 5 |
| 96 | 11 | 2 | 6 | 2 | 1 | 20 | 4 | 5 | 2 | 0 |
| 97 | 6 | 2 | 4 | 1 | 1 | 17 | 7 | 1 | 2 | 1 |
| 98 | 3 | 1 | 1 | 2 | 1 | 16 | 6 | 1 | 1 | 1 |
| 99 | 6 | 0 | 1 | 1 | 1 | 15 | 7 | 2 | 0 | 3 |
| 100 | 3 | 2 | 4 | 2 | 0 | 12 | 2 | 1 | 1 | 2 |
| 101 | 3 | 2 | 1 | 1 | 1 | 6 | 5 | 0 | 1 | 2 |
| 102 | 3 | 1 | 2 | 1 | 0 | 4 | 1 | 1 | 0 | 2 |
| 103 | 1 | 2 | 1 | 1 | 2 | 5 | 1 | 1 | 1 | 2 |
| 104 | 3 | 3 | 1 | 0 | 0 | 3 | 7 | 0 | 0 | 0 |
| 105 | 1 | 0 | 0 | 0 | 1 | 4 | 3 | 2 | 2 | 0 |
| 106 | 1 | 2 | 0 | 1 | 1 | 2 | 0 | 1 | 2 | 0 |
| 107 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 1 | 0 |
| 108 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| 109 | 1 | 1 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 |
| 110 | 1 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 |
| 111 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 112 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 113 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 114 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 115 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 116 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 117 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 118 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 119 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 120 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 2.1.4a-Objective function values and parameter counts. Note that fishery CPUE likelihoods are calculated, but not used, in Model 11.5.

|  | Aggregated data components |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Obj. function component | M11.5 | M16.1 | M16.2 | M16.3 | M16.4 | M16.5 |
| Catch | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Equilibrium catch | 0.01 | 0.00 | 0.00 | 0.00 | 0.01 | 0.03 |
| Survey abundance index | -6.87 | -20.68 | -65.07 | -68.95 | -72.68 | -63.49 |
| Size composition | 5235.34 | 1332.77 | 1203.53 | 1359.81 | 1595.14 | 2144.84 |
| Age composition | 145.88 | 230.60 | 87.74 | 67.26 | 111.19 | 72.49 |
| Recruitment | 22.19 | 4.55 | -4.05 | -0.40 | 5.28 | 44.64 |
| Priors | 0.00 | 0.00 | 158.73 | 304.00 | 480.69 | 784.12 |
| "Softbounds" | 0.03 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |
| Deviations | 20.31 | 0.00 | 96.61 | 55.82 | 59.85 | 118.88 |
| "F ballpark" | 0.00 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| Total | 5416.88 | 1547.24 | 1477.49 | 1717.55 | 2179.47 | 3101.51 |


|  | Abundance index, broken down by fleet |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Fleet | M11.5 | M16.1 | M16.2 | M16.3 | M16.4 | M16.5 |
| Fishery |  |  |  |  |  |  |
| Shelf trawl survey | -6.87 | -20.68 | -60.23 | -56.56 | -53.86 | -45.64 |
| IPHC longline survey |  |  | -4.84 |  | -13.44 | -13.85 |
| NMFS longline survey |  |  |  | -12.39 | -5.39 | -3.99 |
| Total | -6.87 | -20.68 | -65.07 | -68.95 | -72.68 | -63.49 |


|  | Size composition, broken down by fleet |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Fleet | M11.5 | M16.1 | M16.2 | M16.3 | M16.4 | M16.5 |
| Fishery | 4306.84 | 361.13 | 199.16 | 184.48 | 233.94 | 390.63 |
| Shelf trawl survey | 928.51 | 971.64 | 869.23 | 835.76 | 857.90 | 988.61 |
| IPHC longline survey |  |  | 135.14 |  | 364.40 | 493.74 |
| NMFS longline survey |  |  |  | 339.58 | 138.90 | 271.86 |
| Total | 5235.34 | 1332.77 | 1203.53 | 1359.81 | 1595.14 | 2144.84 |


| Fleet | Age composition, broken down by fleet |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | M11.5 | M16.1 | M16.2 | M16.3 | M16.4 | M16.5 |
| Fishery |  |  |  |  | 37.97 | 13.58 |
| Shelf trawl survey | 145.88 | 230.60 | 87.74 | 67.26 | 73.22 | 58.91 |
| IPHC longline survey NMFS longline survey |  |  |  |  |  |  |
| Total | 145.88 | 230.60 | 87.74 | 67.26 | 111.19 | 72.49 |
| Parameter counts | M11.5 | M16.1 | M16.2 | M16.3 | M16.4 | M16.5 |
| Unconstrained parameters | 115 | 18 | 15 | 15 | 16 | 16 |
| Parameters with priors | 0 | 0 | 55 | 55 | 62 | 62 |
| Constrained deviations | 73 | 58 | 286 | 286 | 286 | 286 |
| Total | 188 | 76 | 356 | 356 | 364 | 364 |

Table 2.1.4b—Multinomial sample size multipliers.

|  | Sizecomp multinomial sample size multipliers |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| Model | Fishery | Trawl survey | IPHC longline survey | NMFS longline survey |
| 11.5 | 1 | 1 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| 16.1 | 1 | 1 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| 16.2 | 1 | 1 | 1 | $\mathrm{n} / \mathrm{a}$ |
| 16.3 | 1 | 1 | $\mathrm{n} / \mathrm{a}$ | 1 |
| 16.4 | 1 | 1 | 1 | 1 |
| 16.5 | 2.01 | 1.07 | 1.52 | 3.65 |


|  | Agecomp multinomial sample size multipliers |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| Model | Fishery | Trawl survey | IPHC longline survey | NMFS longline survey |
| 11.5 | $\mathrm{n} / \mathrm{a}$ | 1 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| 16.1 | $\mathrm{n} / \mathrm{a}$ | 1 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| 16.2 | $\mathrm{n} / \mathrm{a}$ | 0.492 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| 16.3 | $\mathrm{n} / \mathrm{a}$ | 0.492 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| 16.4 | 0.492 | 0.492 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| 16.5 | 0.12 | 0.30 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |

Table 2.1.5—Various goodness-of-fit measures for survey abundance data. $\quad$ $\mathbf{~ a v e}=$ mean log-scale standard error, RMSE = root mean squared error, MNR = mean normalized residual, SDNR = standard deviation of normalized residuals, Corr. = correlation (observed:estimated).

| Model | Survey | oave | RMSE | MNR | SDNR | Corr. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 11.5 | Trawl | 0.11 | 0.22 | 0.95 | 1.80 | 0.78 |
| 16.1 | Trawl | 0.11 | 0.19 | 0.07 | 1.82 | 0.78 |
| 16.2 | Trawl | 0.11 | 0.11 | 0.09 | 1.00 | 0.93 |
| 16.3 | Trawl | 0.11 | 0.13 | 0.10 | 1.10 | 0.91 |
| 16.4 | Trawl | 0.11 | 0.14 | 0.10 | 1.17 | 0.90 |
| 16.5 | Trawl | 0.11 | 0.15 | 0.07 | 1.36 | 0.88 |
| 16.2 | IPHC LL | 0.43 | 0.56 | -0.05 | 1.07 | -0.12 |
| 16.4 | IPHC LL | 0.42 | 0.55 | -0.06 | 1.08 | -0.14 |
| 16.5 | IPHC LL | 0.46 | 0.58 | -0.05 | 1.07 | -0.14 |
| 16.3 | NMFS LL | 0.18 | 0.19 | -0.22 | 0.99 | 0.70 |
| 16.4 | NMFS LL | 0.17 | 0.16 | -0.19 | 0.96 | 0.77 |
| 16.5 | NMFS LL | 0.17 | 0.15 | -0.14 | 0.93 | 0.82 |

Table 2.1.6—Statistics related to effective sample sizes (Neff) for length composition data. Nrec = no. records, $\mathrm{A}(\cdot)=$ arithmetic mean, $\mathrm{H}(\cdot)=$ harmonic mean, $\mathrm{Ninp}=$ input sample size. Input sample sizes were adjusted for Model 16.5 (tuned so that $\mathrm{H}(\mathrm{Neff}) / \mathrm{A}(\mathrm{Ninp})=1.00$ ).

| Model | Fleet | Nrec | A(Ninp) | Ratios |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | A(Neff)/A(Ninp) | H(Neff)/A(Ninp) |
| 11.5 | Jan-Apr trawl fish. | 68 | 314 | 2.92 | 1.53 |
| 11.5 | May-Jul trawl fish. | 35 | 62 | 7.26 | 3.32 |
| 11.5 | Aug-Dec trawl fish. | 38 | 44 | 6.00 | 3.24 |
| 11.5 | Jan-Apr longline fish. | 72 | 476 | 3.99 | 1.18 |
| 11.5 | May-Jul longline fish. | 35 | 252 | 5.16 | 3.00 |
| 11.5 | Aug-Dec longline fish. | 67 | 673 | 3.09 | 0.89 |
| 11.5 | Jan-Apr pot fish. | 40 | 129 | 9.71 | 3.37 |
| 11.5 | May-Jul pot fish. | 17 | 129 | 7.72 | 1.72 |
| 11.5 | Aug-Dec pot fish. | 40 | 84 | 7.25 | 2.75 |
| 16.1 | Fishery | 39 | 300 | 5.61 | 1.86 |
| 16.2 | Fishery | 39 | 300 | 10.31 | 2.35 |
| 16.3 | Fishery | 39 | 300 | 14.34 | 2.17 |
| 16.4 | Fishery | 39 | 300 | 11.25 | 1.91 |
| 16.5 | Fishery | 39 | 603 | 5.87 | 1.00 |
| 11.5 | Trawl survey | 34 | 286 | 1.66 | 1.03 |
| 16.1 | Trawl survey | 34 | 300 | 1.57 | 1.01 |
| 16.2 | Trawl survey | 34 | 300 | 1.88 | 1.15 |
| 16.3 | Trawl survey | 34 | 300 | 2.01 | 1.17 |
| 16.4 | Trawl survey | 34 | 300 | 1.97 | 1.14 |
| 16.5 | Trawl survey | 34 | 321 | 1.75 | 1.00 |
| 16.2 | IPHC longline survey | 7 | 300 | 2.41 | 2.03 |
| 16.4 | IPHC longline survey | 7 | 300 | 2.58 | 2.16 |
| 16.5 | IPHC longline survey | 7 | 1094 | 1.13 | 1.00 |
| 16.3 | NMFS longline survey | 10 | 300 | 1.93 | 1.31 |
| 16.4 | NMFS longline survey | 10 | 300 | 1.80 | 1.28 |
| 16.5 | NMFS longline survey | 10 | 456 | 1.31 | 1.00 |

Table 2.1.7a—Statistics related to effective sample size (Eff. N) for survey age composition data. "In. $\mathrm{N} "=$ input sample size, Mean = arithmetic mean, Harm. = harmonic mean, Ratio1 = arithmetic mean effective sample size divided by arithmetic mean input sample size, Ratio2 = harmonic mean effective sample size divided by arithmetic mean input sample size. For Models 16.2-16.4, arithmetic mean input sample sizes were left at the values tuned in last year's assessment for Model 15.6 so that H(Neff)/A(Ninp)=1 (tan shading). For Model 16.5, arithmetic mean input sample sizes were tuned in this year's assessment so that $\mathrm{H}(\mathrm{Neff}) / \mathrm{A}(\mathrm{Ninp})=1$ (green shading).

Trawl survey age compositions

| Year | Model 11.5 |  | Model 16.1 |  | Model 16.2 |  | Model 16.3 |  | Model 16.4 |  | Model 16.5 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | In. N | Eff. N | In. N | Eff. N | In. N | Eff. N | In. N | Eff. N | In. N | Eff. N | In. N | Eff. N |
| 1994 | 201 | 437 | 201 | 209 | 99 | 211 | 99 | 210 | 99 | 155 | 60 | 186 |
| 1995 | 160 | 37 | 160 | 29 | 79 | 39 | 79 | 47 | 79 | 62 | 48 | 44 |
| 1996 | 200 | 342 | 200 | 69 | 98 | 156 | 98 | 240 | 98 | 198 | 60 | 103 |
| 1997 | 202 | 149 | 202 | 47 | 99 | 226 | 99 | 279 | 99 | 175 | 61 | 147 |
| 1998 | 178 | 1116 | 178 | 89 | 88 | 160 | 88 | 1913 | 88 | 1346 | 53 | 800 |
| 1999 | 241 | 125 | 241 | 59 | 119 | 79 | 119 | 111 | 119 | 76 | 72 | 83 |
| 2000 | 241 | 115 | 241 | 60 | 119 | 84 | 119 | 55 | 119 | 48 | 72 | 44 |
| 2001 | 258 | 99 | 258 | 37 | 127 | 73 | 127 | 85 | 127 | 79 | 77 | 89 |
| 2002 | 244 | 90 | 244 | 40 | 120 | 52 | 120 | 77 | 120 | 62 | 73 | 57 |
| 2003 | 354 | 266 | 354 | 797 | 174 | 1699 | 174 | 613 | 174 | 792 | 106 | 1212 |
| 2004 | 279 | 31 | 279 | 35 | 137 | 38 | 137 | 47 | 137 | 43 | 84 | 44 |
| 2005 | 359 | 395 | 359 | 184 | 177 | 388 | 177 | 379 | 177 | 360 | 108 | 319 |
| 2006 | 365 | 147 | 365 | 54 | 180 | 98 | 180 | 177 | 180 | 130 | 110 | 85 |
| 2007 | 404 | 61 | 404 | 11 | 199 | 34 | 199 | 477 | 199 | 270 | 121 | 107 |
| 2008 | 340 | 250 | 340 | 137 | 167 | 375 | 167 | 278 | 167 | 379 | 102 | 107 |
| 2009 | 396 | 94 | 396 | 168 | 195 | 214 | 195 | 303 | 195 | 500 | 119 | 210 |
| 2010 | 363 | 94 | 363 | 210 | 179 | 218 | 179 | 190 | 179 | 190 | 109 | 124 |
| 2011 | 352 | 151 | 352 | 121 | 173 | 99 | 173 | 92 | 173 | 120 | 106 | 46 |
| 2012 | 365 | 98 | 365 | 82 | 180 | 79 | 180 | 97 | 180 | 107 | 110 | 59 |
| 2013 | 398 | 122 | 398 | 141 | 196 | 107 | 196 | 116 | 196 | 95 | 119 | 85 |
| 2014 | 399 | 483 | 399 | 285 | 196 | 417 | 196 | 392 | 196 | 355 | 120 | 369 |
| Mean | 300 | 224 | 300 | 136 | 148 | 231 | 148 | 294 | 148 | 264 | 90 | 206 |
| Harm. |  | 109 |  | 58 |  | 95 |  | 128 |  | 119 |  | 90 |
| Ratio1 |  | 0.75 |  | 0.45 |  | 1.56 |  | 1.99 |  | 1.79 |  | 2.29 |
| Ratio2 |  | 0.36 |  | 0.19 |  | 0.64 |  | 0.87 |  | 0.81 |  | 1.00 |

Table 2.1.7b—Statistics related to effective sample size (Eff. N) for fishery age composition data. "In. N" = input sample size, Mean = arithmetic mean, Harm. = harmonic mean, Ratio1 = arithmetic mean effective sample size divided by arithmetic mean input sample size, Ratio2 = harmonic mean effective sample size divided by arithmetic mean input sample size. For Model 16.4, arithmetic mean input sample size for the fishery agecomp data was assumed equal to arithmetic mean input sample size for the survey agecomp data (purple shading). For Model 16.5, arithmetic mean input sample sizes were tuned in this year's assessment so that $\mathrm{H}(\mathrm{Neff}) / \mathrm{A}(\mathrm{Ninp})=1$ (green shading).

Fishery age compositions

| Year |  |  |  |  | Model 16.4 |  | Model 16.5 |  |
| :---: | :---: | :--- | :--- | :--- | :--- | :---: | :---: | :---: |
|  |  |  |  |  | In. N | Eff. N | In. N | Eff. N |
| 2008 |  |  |  |  | 130 | 75 | 32 | 59 |
| 2009 |  |  |  |  | 127 | 44 | 31 | 25 |
| 2010 |  |  |  | 111 | 71 | 27 | 31 |  |
| 2011 |  |  |  |  | 222 | 79 | 54 | 41 |
| Mean |  |  |  |  | 148 | 67 | 36 | 39 |
| Harm. |  |  |  |  | 64 |  | 35 |  |
| Ratio1 |  |  |  |  |  | 0.46 |  | 1.08 |
| Ratio2 |  |  |  |  |  |  |  |  |

Table 2.1.8—Estimates ("Est.") of key parameters and their standard deviations ("SD"). A blank indicates that the parameter (row) was not used in that model (column). A "_" symbol under SD. indicates that the parameter (row) was fixed (not estimated) in that model (column).

| Parameter | Model 11.5 |  | Model 16.1 |  | Model 16.2 |  | Model 16.3 |  | Model 16.4 |  | Model 16.5 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Est. | SD | Est. | SD | Est. | SD | Est. | SD | Est. | SD | Est. | SD |
| Natural mortality | 0.340 |  | 0.373 | 0.012 | 0.300 | 0.020 | 0.230 | 0.015 | 0.216 | 0.013 | 0.194 | 0.010 |
| Length at age 1 (cm) | 14.244 | 0.104 | 16.323 | 0.086 | 16.397 | 0.087 | 16.392 | 0.087 | 16.420 | 0.088 | 16.465 | 0.086 |
| Asymptotic length (cm) | 92.513 | 0.493 | 98.211 | 1.848 | 97.879 | 1.343 | 95.326 | 1.335 | 98.524 | 1.242 | 98.169 | 0.847 |
| Brody growth coefficient | 0.240 | 0.002 | 0.199 | 0.012 | 0.214 | 0.010 | 0.229 | 0.011 | 0.209 | 0.009 | 0.222 | 0.007 |
| Richards growth coefficient |  |  | 1.058 | 0.049 | 0.985 | 0.044 | 0.961 | 0.043 | 1.031 | 0.039 | 0.986 | 0.032 |
| SD of length at age 1 (cm) | 3.537 | 0.066 | 3.375 | 0.057 | 3.489 | 0.057 | 3.508 | 0.057 | 3.566 | 0.058 | 3.619 | 0.055 |
| SD of length at age 20 (cm) | 9.776 | 0.152 | 9.863 | 0.279 | 7.688 | 0.228 | 7.293 | 0.211 | 6.959 | 0.200 | 6.651 | 0.147 |
| Ageing bias at age 1 (years) | 0.333 | 0.013 | 0.320 | 0.013 | 0.287 | 0.025 | 0.285 | 0.027 | 0.295 | 0.026 | 0.277 | 0.032 |
| Ageing bias at age 20 (years) | 0.354 | 0.148 | 0.340 | 0.159 | 0.703 | 0.254 | 0.753 | 0.264 | 0.281 | 0.235 | 0.910 | 0.306 |
| ln(mean post-1976 recruitment) | 13.196 | 0.019 | 13.580 | 0.104 | 12.949 | 0.167 | 12.328 | 0.107 | 12.458 | 0.093 | 13.563 | 0.145 |
| Sigma_R | 0.570 |  | 0.644 | 0.068 | 0.603 |  | 0.603 |  | 0.603 |  | 0.603 |  |
| $\ln$ (pre-1977 recruitment offset) | -1.151 | 0.130 | -1.071 | 0.228 | -0.559 | 0.172 | -0.616 | 0.137 | -0.699 | 0.126 | -0.718 | 0.096 |
| Initial F (Jan-Apr trawl fishery) | 0.657 | 0.140 |  |  |  |  |  |  |  |  |  |  |
| Initial F (fishery) |  |  | 0.126 | 0.045 | 0.080 | 0.020 | 0.087 | 0.020 | 0.082 | 0.016 | 0.069 | 0.012 |
| "Extra SD" for NMFS LL survey |  |  |  |  | 0.335 | 0.079 |  |  | 0.000 |  | 0.000 |  |
| "Extra SD" for IPHC LL survey |  |  |  |  |  |  | 0.011 | 0.041 | 0.316 | 0.076 | 0.355 | 0.082 |
| Base ln(Q) for trawl survey | -0.261 | - | -0.441 | 0.063 | 0.049 | 0.108 | 0.458 | 0.074 | 0.295 | 0.065 | 0.464 | 0.046 |
| Base $\ln (\mathrm{Q})$ for NMFS LL survey |  |  |  |  | -0.002 | 0.170 |  |  | 0.068 | 0.066 | 0.354 | 0.057 |
| Base $\ln (\mathrm{Q})$ for IPHC LL survey |  |  |  |  |  |  | 0.324 | 0.081 | 0.324 | 0.158 | 0.562 | 0.141 |

Figures



Figure 2.1.1a—Model fits to the trawl survey abundance time series. Upper panel: Models 11.5, 16.1, and 16.2. Lower panel: Models 16.3-16.5. Survey time series shows $95 \%$ confidence interval.


Figure 2.1.1b—Model fits to the IPHC longline survey abundance time series (Models 16.2, 16.4, and 16.5 only). Survey time series shows $95 \%$ confidence interval, which differs between models.


Figure 2.1.1c—Model fits to the NMFS longline survey abundance time series (Models 16.3, 16.4, and 16.5 only). Survey time series shows $95 \%$ confidence interval, which differs between models.


Figure 2.1.2a—Model 11.5 fits to trawl survey age composition data (page 1 of 2 ).


Age (yr)

Figure 2.1.2a—Model 11.5 fits to trawl survey age composition data (page 2 of 2 ).


Figure 2.1.2b—Model 16.1 fits to trawl survey age composition data (page 1 of 2 ).


## Age (yr)

Figure 2.1.2b—Model 16.1 fits to trawl survey age composition data (page 2 of 2).


Figure 2.1.2c-Model 16.2 fits to trawl survey age composition data (page 1 of 2).


Age (yr)
Figure 2.1.2c—Model 16.2 fits to trawl survey age composition data (page 2 of 2 ).


Figure 2.1.2d—Model 16.3 fits to trawl survey age composition data (page 1 of 2 ).


Age (yr)
Figure 2.1.2d—Model 16.3 fits to trawl survey age composition data (page 2 of 2).


Figure 2.1.2e—Model 16.4 fits to trawl survey age composition data (page 1 of 2).


## Age (yr)

Figure 2.1.2e—Model 16.4 fits to trawl survey age composition data (page 2 of 2 ).


Figure 2.1.2f-Model 16.5 fits to trawl survey age composition data (page 1 of 2 ).


Age (yr)
Figure 2.1.2f—Model 16.5 fits to trawl survey age composition data (page 2 of 2 ).

Model 16.4


Model 16.5


Figure 2.1.3-Model fits to fishery age composition data (Models 16.4 and 16.5 only).


Figure 2.1.4a—Gear-and-season-specific fishery selectivity as estimated by Model 11.5.

Model 16.1


Figure 2.1.4b—Fishery selectivity as estimated by Model 16.1.

Model 16.2


Model 16.3


Figure 2.1.4c—Fishery selectivity as estimated by Models 16.2 and 16.3.

Model 16.4


Model 16.5


Figure 2.1.4d—Fishery selectivity as estimated by Models 16.4 and 16.5.

Model 11.5


Model 16.1


Figure 2.1.5a—Trawl survey selectivity (page 1 of 3).

Model 16.2


Model 16.3


Figure 2.1.5a—Trawl survey selectivity (page 2 of 3 ).

Model 16.4


Model 16.5


Figure 2.1.5a—Trawl survey selectivity (page 3 of 3).

## Model 16.2



Model 16.4


Model 16.5


Figure 2.1.5b—IPHC longline survey selectivity.

## Model 16.3



Model 16.4


Model 16.5


Figure 2.1.5c—NMFS longline survey selectivity.


Figure 2.1.6-Total biomass time series as estimated by each of the models. Survey biomass (with 95\% confidence interval) shown for comparison.


Figure 2.1.7—Time series of spawning biomass relative to $B_{100 \%}$ for each of the models, with $95 \%$ confidence intervals.


Figure 2.1.8—Age 0 recruitment (1000s of fish) for each model.


Figure 2.1.9—Time series of the ratio of full-selection fishing morality to $F_{40 \%}$.


Figure 2.1.10a—Ten-year spawning biomass retrospective analysis of Model 11.5.


Figure 2.1.10b—Ten-year spawning biomass retrospective analysis of Model 16.1.


Figure 2.1.10c—Ten-year spawning biomass retrospective analysis of Model 16.2.


Figure 2.1.10d—Ten-year spawning biomass retrospective analysis of Model 16.3.


Figure 2.1.10e—Ten-year spawning biomass retrospective analysis of Model 16.4.


Figure 2.1.10f—Ten-year spawning biomass retrospective analysis of Model 16.5.

# APPENDIX 2.2: BSAI PACIFIC COD ECONOMIC PERFORMANCE REPORT FOR 2015 

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Pacific cod is the second largest species in terms of catch in the Bering Sea \& Aleutian Island (BSAI) region. Pacific cod accounted for $13 \%$ of the BSAI's FMP groundfish harvest and $75 \%$ of the total Pacific cod harvest in Alaska. Retained catch of Pacific cod decreased $1 \%$ to 231 thousand t in 2015, and though down from its peak of 241 thousand t in 2012, is $35 \%$ higher than the 2006-2010 average (Table 2.2.1). The products made from BSAI Pacific cod had a first-wholesale value of $\$ 362$ million in 2015, which was up from $\$ 354$ million in 2014 and above the 2006-2010 average of $\$ 300$ million (Table 2.2.2). The higher revenue in recent years is largely the result of increased catch and production levels as the average first-wholesale price of Pacific cod products have declined in recent.

Cod is an iconic fishery with a long history of production across much of the globe. Global catch was consistently over 2 million through the 1980s, but began to taper off in the 1990s as cod stocks began to collapse in the northwest Atlantic Ocean. Over roughly the same period, the U.S. catch of Pacific cod (caught in Alaska) grew to approximately 250 thousand tons where it remained throughout the early to mid-2000s. European catch of Atlantic cod in the Barents Sea (conducted mostly by Russia, Norway, and Iceland) slowed and global catch hit a low in 2007 at 1.13 million t. U.S. Pacific cod's share of global catch was at a high at just over $20 \%$ in the early 2000s. Since 2007 global catch has grown to 1.85 million t in 2014 as catch in the Barents Sea has rebounded and U.S. catch has remained strong at over 300 thousand t since 2011. European Atlantic cod and U.S. Pacific cod remain the two major sources supplying the cod market over the past decade accounting for roughly $75 \%$ and $20 \%$, respectively. Atlantic cod and Pacific cod are substitutes in the global market. Because of cod's long history global demand is present in a number of geographical regions, but Europe and the U.S. are the primary consumer markets for many Pacific cod products. The market for cod is also indirectly affected by activity in the pollock fisheries which experienced a similar period of decline in 2008-2010 before rebounding. Cod and pollock are commonly used to produce breaded fish portions. Alaska caught Pacific cod in the BSAI became certified by the Marine Stewardship Council (MSC) in 2010, a NGO based third-party sustainability certification, which some buyers seek.

The Pacific cod total allowable catch (TAC) is allocated to multiple sectors (fleets). CDQ entities receive $10 \%$ of the total BSAI quota. The largest sectoral allocation goes to the Freezer longline CPs which receive roughly $44 \%$ of the total BSAI cod quota ( $48.7 \%$ non-CDQ quota). While not an official catch share program, the Freezer longline CPs have formed a voluntary cooperative that allows them to form private contracts among members to distribute the sectoral allocation. The remaining large sectors are the trawl CPs, trawl CVs, the pot gear CVs and some smaller sideboard limits to cover the catch of Pacific cod while targeting other species. The CVs (collectively referred to as the inshore sector) make deliveries to shore-based processors, and catcher/processors process catch at-sea before going directly to the wholesale markets. Among the at-sea CPs, catch is distributed approximately three-quarters to the hook-and-line and one quarter to trawl. The inshore sector accounts for $25 \%-30 \%$ of the total BSAI Pacific cod catch of which approximately two-thirds is caught by the trawl and one-third by the pot gear sectors. The retained catch of the inshore sector decreased $3 \%$ increase to 61 thousand $t$. The value of these deliveries (shoreside ex-vessel value) totaled $\$ 29.4$ million in 2015, which was down $21 \%$ from 2014, as ex-vessel prices also decreased $7 \%$ to an average of $\$ 0.249$ per pound. Changes in ex-vessel prices over time
generally reflect changes in the corresponding wholesale prices. Catch from the fixed gear vessels (which includes hook-and-line and pot gear) typically receive a slightly higher price from processors because they incur less damage when caught. The fixed gear price premium has varied over time but recently has been about $\$ 0.03$ per pound.

The first-wholesale value of Pacific cod products was down 2\% to $\$ 362.1$ million in 2015, though revenues in recent years remain high as result of increased catch levels. The average price of Pacific cod products in 2015 increased 5\% to $\$ 1.364$ driven by an increase in the H\&G price. Changes in global catch and production account for much the trends in the cod markets. In particular, the average first-wholesale prices peak at over $\$ 1.80$ per pound in 2007-2008 and subsequent declined precipitously in 2009 to $\$ 1.20$ per pound as markets priced in consecutive years of approximately 100 thousand $t$ increases in the Barents Sea cod catch in 2009-2011; coupled with reduced demand from the recession. Average firstwholesale prices since have fluctuated between approximately $\$ 1.20$ and $\$ 1.55$ per pound. Head and gut ( $\mathrm{H} \& \mathrm{G}$ ) production is the focus of the BSAI processors but a significant amount of fillets are produced as well. H\&G typically constitutes over $80 \%$ of value and fillets over $10 \%$ of value. Shoreside processors produce the majority of the fillets. Almost all of the at-sea sector's catch is processed into H\&G. Other product types are not produced in significant quantities. At-sea head and gut prices tend to be about 20\%$30 \%$ higher, in part because of the shorter period of time between catch and freezing, and in part because the at-sea sector is disproportionately caught by hook-and-line which yields a better price. Head \& gut prices bottomed out at $\$ 1.049$ in 2013, a year in which Barents Sea cod catch increased roughly 240 thousand t (an increase that is approximately the size of Alaska's cod total catch) but have since rebounded to \$1.365. Fillet Prices have steady declined from over \$3 in 2011 to \$2.465 in 2015.
U.S. exports of cod have risen almost proportionally with increasing U.S. cod production. More than $90 \%$ of the exports are H\&G, most of which goes to China for secondary processing and re-export. China's rise as a re-processor is fairly recent. Between 2001 and 2011 exports to China have increased nearly 10 fold. Japan and Europe (mostly Germany and the Netherlands) are also important export destinations. Approximately $30 \%$ of Alaska's cod production is estimated to remain in the U.S. In 2016 Norway and Russia maintained their Barents Sea TAC at 2015 levels despite recommendations by ICES to reduce the TAC by roughly $10 \%$. Reports indicate that marginal reduction in the Barents Sea catch is planned to take effect in 2017, but it is sufficiently small that it may not impact prices much.

Table 2.2.1. Bering Sea \& Aleutian Islands Pacific cod catch and ex-vessel data. Total and retained catch (thousand metric tons), number of vessel, catcher/processor (CP) hook-and-line H\&L share of catch, CP trawl share of catch, shoreside pot gear share of catch, shoreside trawl share of catch, shoreside ex-vessel value (million US\$), fixed gear and trawl price (US\$ per pound), and shoreside number of vessel; 20062010 average and 2011-2015.

|  | Avg 06-10 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total catch K mt | 177.2 | 220.1 | 250.9 | 250.3 | 249.3 | 242 |
| Retained Federal catch K mt | 170.9 | 216.5 | 241.2 | 238.8 | 232.1 | 230.9 |
| Vessels \# | 201.2 | 171 | 177 | 178 | 156 | 150 |
| CP H\&L share of BSAI catch | 53\% | 53\% | 54\% | 51\% | 53\% | 56\% |
| CP trawl share of BSAI catch | 18\% | 15\% | 15\% | 18\% | 15\% | 15\% |
| Shoreside fed total catch K mt | 46.8 | 65 | 70 | 67 | 67 | 61 |
| Shoreside catcher vessels \# | 61.2 | 54 | 55 | 50 | 47 | 49 |
| CV pot gear share of BSAI catch | 9\% | 11\% | 9\% | 9\% | 10\% | 9\% |
| CV trawl share of BSAI catch | 18\% | 18\% | 19\% | 17\% | 17\% | 16\% |
| Shoreside ex-vessel value M \$ | \$34.86 | \$34.04 | \$44.72 | \$34.04 | \$37.08 | \$29.40 |
| Shoreside ex-vessel price lb \$ | \$0.379 | \$0.275 | \$0.318 | \$0.244 | \$0.273 | \$0.249 |
| Shoreside fixed gear ex-vessel price premium | \$0.06 | \$0.06 | \$0.01 | \$0.01 | \$0.03 | \$0.03 |

Source: NMFS Alaska Region Blend and Catch-accounting System estimates; NMFS Alaska Region At-sea Production Reports; and ADF\&G Commercial Operators Annual Reports (COAR). Data compiled and provided by the Alaska Fisheries Information Network (AKFIN).

Table 2.2.2. Bering Sea \& Aleutian Islands Pacific cod first-wholesale market data. First-wholesale production (thousand metric tons), value (million US\$), price (US\$ per pound); fillet and head and gut volume (thousand metric tons), value share, and price (US\$ per pound); At-sea share of value and at-sea shoreside price difference (US\$ per pound); 2006-2010 average and 2011-2015.

|  | Avg 06-10 | 2011 | 2012 | 2013 |  | 2014 |  | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| All products volume K mt | 85.74 | 107.39 | 122.75 | 121.70 |  | 123.51 |  | 120.40 |
| All products Value M \$ | \$ 299.7 | \$ 366.0 | \$ 381.1 | \$ 303.7 | \$ | 353.8 | \$ | 362.1 |
| All products price lb \$ | \$ 1.586 | \$ 1.546 | \$ 1.408 | \$ 1.132 | \$ | 1.299 | \$ | 1.364 |
| Fillets volume K mt | 4.34 | 6.57 | 6.76 | 8.79 |  | 8.42 |  | 6.28 |
| Fillets value share | 10.1\% | 12.1\% | 12.1\% | 18.1\% |  | 14.1\% |  | 9.4\% |
| Fillets price lb \$ | \$ 3.182 | \$ 3.059 | \$ 3.100 | \$ 2.836 | \$ | 2.683 | \$ | 2.465 |
| Head \& Gut volume K mt | 70.41 | 88.78 | 104.24 | 97.76 |  | 100.56 |  | 100.76 |
| Head \& Gut value share | 82.0\% | 81.0\% | 82.4\% | 74.5\% |  | 78.8\% |  | 83.7\% |
| Head \& Gut price lb \$ | \$ 1.584 | \$ 1.514 | \$ 1.366 | \$ 1.049 | \$ | 1.257 | \$ | 1.365 |
| At-sea value share | 74.5\% | 74.2\% | 70.8\% | 68.7\% |  | 69.0\% |  | 77.1\% |
| At-sea price premium (\$/lb) | \$ 0.00 | \$ -0.04 | \$ -0.13 | \$ -0.28 | \$ | -0.01 | \$ | -0.12 |

Source: NMFS Alaska Region Blend and Catch-accounting System estimates; NMFS Alaska Region At-sea Production Reports; and ADF\&G Commercial Operators Annual Reports (COAR). Data compiled and provided by the Alaska Fisheries Information Network (AKFIN).

Table 2.2.3. Cod U.S. trade and global market data. Global production (thousand metric tons), U.S. share of global production, and Europe's share of global production; U.S. export volume (thousand metric tons), value (million US\$), and price (US\$ per pound); U.S. cod consumption (estimated), and share of domestic production remaining in the U.S. (estimated); and the share of U.S. export volume and value for head and gut (H\&G), fillets, China, Japan, and Germany and Netherlands; 2006-2010 average and 20112016.

|  | Avg 06-10 | 2011 | 2012 | 2013 | 2014 | 2015 | $\begin{array}{r} 2016 \\ \text { (thru June) } \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Global cod catch K mt | 1,209 | 1,505 | 1,600 | 1,828 | 1,850 | - | - |
| U.S. P. cod share of global catch Europe share of global catch | 19.0\% | 20.0\% | 20.4\% | 16.9\% | 17.6\% | - | - |
|  | 71.8\% | 73.1\% | 73.2\% | 76.7\% | 76.0\% | - | - |
| Pacific cod share of U.S. catch U.S. cod consumption K mt (est.) | 96.7\% | 97.4\% | 98.6\% | 99.3\% | 99.3\% | - | - |
|  | 80 | 88 | 98 | 105 | 115 | 108 | - |
| Share of U.S. cod not exported | 24\% | 24\% | 30\% | 31\% | 31\% | 26\% | - |
| Export volume K mt | 86.6 | 110.8 | 111.1 | 101.8 | 107.3 | 113.2 | 71.7 |
| Export value M US\$ | \$ 266.1 | \$ 371.3 | \$ 363.6 | \$ 308.0 | \$ 314.2 | \$ 334.9 | \$ 204.3 |
| Export price lb US\$ | \$ 1.393 | \$ 1.520 | \$ 1.485 | \$ 1.373 | \$ 1.328 | \$ 1.342 | \$ 1.293 |
| volume share value share | 71\% | 74\% | 80\% | 91\% | 92\% | 91\% | 94\% |
|  | 69\% | 75\% | 80\% | 89\% | 91\% | 90\% | 93\% |
| Fillets volume share | 13\% | 9\% | 9\% | 4\% | 2\% | 3\% | 3\% |
| Filue share | 16\% | 12\% | 11\% | 5\% | 4\% | 4\% | 4\% |
| China volume share | 23\% | 39\% | 46\% | 51\% | 54\% | 53\% | 64\% |
| value share | 21\% | 37\% | 43\% | 48\% | 51\% | 51\% | 61\% |
| Japan value shar | 18\% | 20\% | 16\% | 13\% | 16\% | 13\% | 9\% |
|  | 18\% | 20\% | 16\% | 13\% | 16\% | 14\% | 9\% |
| Netherlands volume share \& Germany value share | 11\% | 10\% | 8\% | 8\% | 9\% | 8\% | 5\% |
|  | 13\% | 11\% | 9\% | 9\% | 10\% | 8\% | 5\% |

Notes: Pacific cod in this table is for all U.S. Unless noted, `cod' in this table refers to Atlantic and Pacific cod. Russia, Norway, and Iceland account for the majority of Europe's cod catch which is largely focused in the Barents Sea.
Source: FAO Fisheries \& Aquaculture Dept. Statistics http://www.fao.org/fishery/statistics/en. NOAA Fisheries, Fisheries Statistics Division, Foreign Trade Division of the U.S. Census Bureau, http://www.st.nmfs.noaa.gov/commercial-fisheries/foreign-trade/index. U.S. Department of Agriculture http://www.ers.usda.gov/data-products/agricultural-exchange-rate-data-set.aspx.

## APPENDIX 2.3: HISTORY OF PREVIOUS EBS PACIFIC COD MODEL STRUCTURES DEVELOPED UNDER STOCK SYNTHESIS

For 2005 and beyond, the SSC’s accepted model from the final assessment is shown in bold red.

## Pre-2005

## Timeline

- Pre-1985: Simple projections of current survey numbers at age
- 1985: Projections based on 1979-1985 survey numbers at age
- 1986-1991: ad hoc separable age-structured FORTRAN model
- 1992: FORTRAN-based Stock Synthesis (SS), with age-based data
o Strong 1989 cohort "disappears;" production ageing ceased
- 1993-2003: Models continued to be developed using SS, with length-based data only
- 2001: CIE review of code for proposed "ALASKA" (Age-, Length-, and Area-Structured Kalman Assessment) model and methodology for decision-theoretic estimation of OFL and ABC
o Although review was favorable, use of ALASKA was postponed "temporarily"
- 2004: Models continued to be developed using SS, with length- and age-based data
o New age data, based on revised ageing protocol
o Agecomp data used in "marginal" form
Main features of the early Stock Synthesis EBS Pacific cod models
- $\quad$ Start year $=1977$
- Three seasons (Jan-May, Jun-Aug, Sep-Dec)
- Four fisheries (Jan-May trawl, Jun-Dec trawl, longline, pot)
- $M$ constant at 0.37
- $\quad Q$ constant at 1.00
- Efforts at internal estimation of $M, Q$ unsuccessful
- Double-logistic selectivity for all fleets (fisheries and survey)
- No fleets constrained to exhibit asymptotic selectivity
- Sizecomp input sample size = square root of true sample size
- Survey index standard deviations set to values reported by RACE Division


## 2005

This assessment marked the first application of ADMB-based Stock Synthesis to EBS Pacific cod
Three models were included:

- Model 1 was identical to the 2004 final model (configured under FORTRAN-based SS), except for use of new maturity schedule developed by Stark
- Model 2 was configured under ADMB-based SS, and was designed to be as close as possible to Model 1 given the limitations of the respective software packages, except:
o Nonuniform priors used throughout
o $M$ fixed at $0.37, Q$ fixed at 1.00
- Model 3 was identical to Model 2 except that $M$ and $Q$ were estimated internally

Weight-length and length-age data examined for evidence of sexual dimorphism; none found.

## 2006

Nine models were included, consisting of 2005 final model and a 3-way factorial design of alternative models (the factorial models all differed from the 2005 final model in that they estimated trawl survey $Q$ internally -in the 2005 final model, it was fixed at 1.0; and they estimated all selectivity parameters except for selectivity at the minimum size bin internally—in the 2005 final model, a few selectivity parameters were fixed externally):

- Model 0 was identical to 2005 final model
- Model A1 was identical to Model 0 except as noted above, with:
o NMFS longline survey data omitted
o Double logistic selectivity
o Prior emphasis = 1.0
- Model A2 was identical to Model 0 except as noted above, with:
o NMFS longline survey data omitted
o Double logistic selectivity
o Prior emphasis $=0.5$
- Model B1 was identical to Model 0 except as noted above, with:
o NMFS longline survey data omitted
o Double normal (four parameter) selectivity
o Prior emphasis $=1.0$
- Model B2 was identical to Model 0 except as noted above, with:
o NMFS longline survey data omitted
o Double normal (four parameter) selectivity
o Prior emphasis $=0.5$
- Model C1 was identical to Model 0 except as noted above, with:
o NMFS longline survey data included
o Double logistic selectivity
o Prior emphasis $=1.0$
- Model C2 was identical to Model 0 except as noted above, with:
o NMFS longline survey data included
o Double logistic selectivity
o Prior emphasis $=0.5$
- Model D1 was identical to Model 0 except as noted above, with:
o NMFS longline survey data included
o Double normal (four parameter) selectivity
o Prior emphasis $=1.0$
- Model D2 was identical to Model 0 except as noted above, with:
o NMFS longline survey data included
o Double normal (four parameter) selectivity
o Prior emphasis $=0.5$


## 2007

## Technical workshop

SS introduced a six-parameter form of the double normal selectivity curve (the previous version used only four parameters). This functional form is constructed from two underlying and linearly rescaled normal distributions, with a horizontal line segment joining the two peaks. As configured in SS, the equation uses the following six parameters:

1. beginning_of_peak_region (where the curve first reaches a value of 1.0)
2. width_of_peak_region (where the curve first departs from a value of 1.0)
3. ascending_width (equal to twice the variance of the underlying normal distribution)
4. descending_width (equal to twice the variance of the underlying normal distribution)
5. initial_selectivity (at minimum length/age)
6. final_selectivity (at maximum length/age)

All but beginning_of_peak_region are transformed: The ascending_width and descending_width are logtransformed and the other three parameters are logit-transformed.

Model 0 was prepared ahead of workshop:

- $M$ estimated internally
- Length-at-age parameters estimated internally
- Disequilibrium initial age structure
- Regime shift recruitment offset estimated internally
- Start year changed from 1964 to 1976
- New six-parameter double normal selectivity function used
- Prior distributions reflect $50 \%$ CV for most parameters

Twenty-one other models were prepared ahead of workshop, each of which was based on Model 0:

- Two models to examine inside/outside growth estimation:
o Model 1 was identical to Model 0 except length-at-age parameters estimated outside the model
o Model 2 was identical to Model 0 except standard deviation of length at age 12 estimated internally
- Two models to examine $M$ conditional on $Q$, vice-versa:
o Model 3 was identical to Model 0 except $M$ fixed at 0.37 and $Q$ free
o Model 4 was identical to Model 0 except $Q$ fixed at 0.75 and $M$ free
- Six models to examine effects of prior distributions:
o Model 5 was identical to Model 0 except $30 \%$ CV instead of $50 \%$
o Model 6 was identical to Model 0 except $40 \%$ CV instead of $50 \%$
o Model 7 was identical to Model 0 except emphasis $=0.2$ instead of 1.0
o Model 8 was identical to Model 0 except emphasis $=0.4$ instead of 1.0
o Model 9 was identical to Model 0 except emphasis $=0.6$ instead of 1.0
o Model 10 was identical to Model 0 except emphasis $=0.8$ instead of 1.0
- Four models to examine effects of asymptotic selectivity:
o Model 11 was identical to Model 0 except Jan-May trawl fishery selectivity forced asymptotic
o Model 12 was identical to Model 0 except longline fishery selectivity forced asymptotic
o Model 13 was identical to Model 0 except pot fishery selectivity forced asymptotic
o Model 14 was identical to Model 0 except shelf trawl survey selectivity forced asymptotic
- One model to examine estimation of stock-recruit relationship:
o Model 15 was identical to Model 0 except parameters of a Ricker stock-recruitment relationship estimated internally
- Six models to address EBS-specific comments from the public:
o Model 16 was identical to Model 0 except input $N$ determined by iterative re-weighting
o Model 17 was identical to Model 0 except input $N$ for mean-size-at-age data decreased by an order of magnitude
o Model 18 was identical to Model 0 except standard error from the shelf trawl survey doubled
o Model 19 was identical to Model 0 except all age data removed
o Model 20 was identical to Model 0 except slope survey data removed
o Model 21 was identical to Model 0 except start year changed to 1982
An immense factorial grid of fixed $M \times Q$ models also prepared ahead of workshop, for which only partial results were presented

Eight models were developed during the workshop itself:

- Model 22 was identical to Model 0 except "old" (pre-Stark) maturity schedule used
- Model 23 was identical to Model 0 except priors turned off and separate $M$ estimated for ages 1-2
- Model 24 was identical to Model 0 except priors turned off and longline fishery CPUE included as an index of abundance
- Model 25 was identical to Model 0 except priors turned off and Pcod bycatch from IPHC survey included as an index of abundance
- Model 26 was identical to Model 0 except priors turned off and either $Q(=0.75)$ or $M(=0.37)$ fixed
- Model 27 was identical to Model 0 except all priors turned off other than that for Jan-May trawl selectivity in largest size bin
- Model 28 was identical to Model 0 except survey selectivity forced asymptotic and $Q$ fixed at 0.5
- Model 29 was identical to Model 0 except separate $M$ estimated for ages $9+$


## Preliminary assessment

In general:

- Agecomp data presented as "age conditioned on length" (i.e., not marginals)
- Length-at-age SD a linear function of age
- Annual devs for length at age 1 , sigma $=0.11$
- Annual devs for recruitment, sigma=0.6, 1973-2005
- Annual devs for ascending selectivity, sigma=0.4
- All parameters estimated internally
- Except selectivity parameters pinned against bounds
- Uniform priors used exclusively
- Monotone selectivity for Jan-May trawl fishery
- All other selectivities new "double normal"

Four models were included, all of which were identical to the 2006 final model except as specified above and below:

- Model 1:
o Estimated effect of 1976 regime shift on median recruitment
o Added a large constant to fishery CPUE sigmas
- Model 2 was identical to Model 1 except age-dependent $M$ estimated for ages 8+
- Model 3 was identical to Model 1 except that it did not add the large constant to longline CPUE sigmas
- Model 4 was identical to Model 1 except:
o Effect of regime shift assumed to be zero
o Did not add large constant to longline CPUE sigmas
o Zero emphasis placed on initial catch and age composition
o Iteratively re-weighted input sigmas and input $N$

Also attempted but not included:

- Simplified model with only a single fishery and no seasons

Final assessment
Four models were included:

- Model 1 (comparisons to 2006 final model in parentheses):
o $M$ fixed at 0.34 ( $M$ fixed at 0.37 in 2006)
o Length-at-age parameters estimated internally (fixed at point estimates from data in 2006)
o Start year set at 1977 (start year set at 1964 in 2006)
o Three age groups in initial state vector estimated (initial state vector assumed to be in equilibrium in 2006)
o 6-parameter double normal selectivity (4-parameter version used in 2006)
o Uniform priors used exclusively (informative normal priors used for many parameters in 2006)
o Fishery selectivities constant across all years (approximately decadal "time blocks" used in 2006)
o Ascending limb of survey selectivity varies annually with $\sigma=0.2$ (survey selectivity assumed to be constant in 2006)
o Survey selectivity based on age (length-based selectivity used in 2006)
o Some fishery selectivities forced asymptotic (all selectivities free in 2006)
o Fishery CPUE data included for comparison (not included in 2006)
o Age-based maturity schedule (length-based schedule used in 2006)
o All fisheries seasonally structured (trawl partially seasonal, other gears non-seasonal in 2006)
o Trawl survey abundance measured in numbers (abundance measured in biomass in 2006)
o Multinomial $N$ based on rescaled bootstrap (sample size set equal to square root of actual $N$ in 2006)
- Model 2 was identical to Model 1 except $M$ fixed at 0.37
- Model 3 was identical to Model 1 except $M$ estimated internally
- Model 4 was identical to Model 1 except:
o $M$ estimated internally
o Survey selectivities forced to be asymptotic
o Age data ignored
o Start year set at 1982; 1977 regime shift ignored
o Length-based maturity used
o Length-based survey selectivity used
o Sigma=0.4 for annual deviations in selectivity parameters
o Initial catch ignored in estimating initial fishing mortality


## 2008

## Preliminary assessment

Five models were included:

- Model 1 was identical to the 2007 final model
- Model 2 was identical to Model 1 except growth parameter $L 2$ estimated externally
- Model 3 was identical to Model 1 except exponential-logistic selectivity used instead of double normal
- Model 4 was identical to 2007 Model 4
- Model 5 was identical to Model 1 except:
o Fishery selectivity blocks (5 yr, $10 \mathrm{yr}, 20 \mathrm{yr}$, or no blocks) chosen by AIC
o Lower bound of descending "width" $=5.0$
o Regime-specific recruitment "dev" vectors
o "SigmaR" set equal (iteratively) to stdev(dev) from current regime
o Seasonal weight-length, based on fishery data
o Number of free initial ages chosen by AIC
o Size-at-age data used if modes ambiguous


## Final assessment

Eight models were included:

- Model A1 was identical to Model 5 from September except lower bound on selectivity descending
"width" parameter relaxed so as not to be constraining
- Model A2 was identical to Model A1, except without age data
- Model B1 was identical to Model A1, except:
o "Asymptotic algorithm" used to determine which fisheries will be forced to exhibit asymptotic selectivity
o "Constant-parameters-across-blocks algorithm" used to determine which selectivity parameters can be held constant across blocks
- Model B2 was identical to Model B1, except without age data
- Model C1 was identical to Model B1, except with M estimated internally
- Model D2 was identical to Model B1, except:
o No age data
o Maturity modeled as function of length rather than age
o M estimated iteratively, based on mat. at len and len. at age
- Model E2 was identical to Model B1, except:
o No age data
o Post-1981 trawl survey selectivity forced to be asymptotic
o $M$ estimated internally
- Model F2 was identical to Model 4 from the final assessment for 2007, except start year = 1977


## 2009

## Preliminary assessment

Eight models were included, based on factorial design of the following:

- Selectivity functional form: double normal or exponential-logistic?
- Catchability: free or fixed at 1.0 ?
- Survey selectivity estimation: free or forced asymptotic?

Partial results were presented for a model with a prior distribution for $Q$ based on archival tags (the prior had virtually no impact, which was why only partial results were presented)

Other features explored but not included in the above models:

- Fixing trawl survey catchability at the mean of the above normal prior distribution
- Allowing trawl survey catchability to vary as a random walk
- Fixing trawl survey catchability at a value of 1.00 for the pre-1982 portion of the time series, but allowing it to be estimated freely for the post-1981 portion of the time series
- Reducing the number of survey selectivity parameters subject to annual deviations
- Use of additive, rather than multiplicative, deviations for certain survey selectivity parameters
- Decreasing the value of the $\sigma$ parameter used to constrain annual survey selectivity deviations
- Turning off annual deviations in survey selectivity parameters for the three most recent years
- Turning off all annual deviations in survey selectivity parameters
- Forcing trawl survey selectivity to peak at age 6.5 , the approximate mid-point of the size range of 6081 cm spanned by the results of Nichol et al. (2007)
- Imposing a beta prior distribution on the shape parameter of the exponential-logistic selectivity function in the trawl survey.


## Final assessment

Fourteen models were included (all new since the preliminary assessment except for Model A1):

- Models without mean-size-at-age data:
o Model A1 was identical to the 2008 final model, with the addition of new data, including the first available fishery agecomp data (from the 2008 Jan-May longline fishery)
o Model A2 was identical to Model A1, except all agecomp data omitted
o Model A3 was identical to Model A1, except 2008 Jan-May longline fishery agecomp data omitted
o Model F2 was identical to Model F2 from the final assessment for 2008
- Models with mean-size-at-age data and agecomp data:
o Model B1 was identical to Model A1 except:
- Survey selectivity held constant for most recent two years
- Cohort-specific growth included
- Input standard deviations of all "dev" vectors were set iteratively by matching the standard deviations of the set of estimated devs
- Standard deviation of length at age was estimated outside the model as a linear function of mean length at age
- Selectivity at maximum size or age was treated as a controllable parameter
- $Q$ for the post-1981 trawl survey was fixed at the value that sets the average (weighted by numbers at length) of the product of $Q$ and selectivity for the $60-81 \mathrm{~cm}$ size range equal to the point estimate of 0.47 obtained by Nichol et al. (2007)
- Potential ageing bias was accounted for in the ageing error matrix by examining alternative bias values in increments of 0.1 for ages 2 and above (age-specific bias values were also examined, but did not improve the fit significantly).
o Model C1 was identical to Model B1 except:
- Input standard deviations for all "dev" vectors and the amount of ageing bias fixed at the values obtained iteratively in Model B1
- Catchability itself (rather than the average product of catchability and selectivity for the $60-81 \mathrm{~cm}$ size range) set equal to 0.47
o Model D1 was identical to Model B1 except:
- Input standard deviations for all "dev" vectors and the amount of ageing bias fixed at the values obtained iteratively in Model B1
- Selectivity at maximum size or age was removed from the set of controllable parameters (instead, selectivity at maximum size or age becomes a function of other selectivity parameters)
o Model E1 was identical to Model B1 except:
- Input standard deviations for all "dev" vectors and the amount of ageing bias fixed at the values obtained iteratively in Model B1
- Selectivity at maximum size or age for all non-asymptotic fleets was set equal to a single value that was constant across fleets
o Model G1 was identical to Model B1 except:
- Input standard deviations for all "dev" vectors and the amount of ageing bias fixed at the values obtained iteratively in Model B1
- Survey selectivity was held constant across all years (i.e., no selectivity devs are estimated for any years)
- Models with mean-size-at-age data and without agecomp data:
o Models B2, C2, D2, E2, and G2 were identical to their B1, C1, D1, E1, and G1 counterparts except that agecomp data were ignored and the corresponding sizecomp data were active.


## 2010

## Preliminary assessment

Six models were included:

- Model 1 was identical to the 2009 final model
- Model 2 was identical to Model 1 except:
o Input standard deviations for all "dev" vectors fixed at the values obtained iteratively in Model 1
o IPHC survey data omitted
o Fishery age data omitted
o Traditional 3 -or- 5 cm size bins replaced with 1 cm size bins
o Traditional 3-season structure replaced with new, 5 -season structure
o Spawn time changed from beginning of season 1 to beginning of season 2
- Model 3 was identical to Model 2 except:
o Non-uniform prior distributions used for selectivity parameters and $Q$
- Model 4 was identical to Model 2 except:
o All age data omitted
o Maturity schedule was length-based rather than age-based
- Model 5 was identical to Model 4 except:
o Parameters governing spread of lengths at age around mean length at age estimated internally
- Model 6 was identical to Model 5 except:
o Cohort-specific growth replaced by annual variability in each of the three von Bertalanffy parameters


## Final assessment

Three models were included:

- Model A was identical to Model 1 from the preliminary assessment
- Model B was identical Model 2 from the preliminary assessment, except cohort-specific growth replaced by constant growth
- Model C: same as Model 4 from the preliminary assessment, except cohort-specific growth replaced by constant growth


## 2011

## CIE review

Exploratory model developed prior to review, which was the same as the 2010 final model, except:
o All sizecomp data turned on
o Nine season $\times$ gear fisheries consolidated into five seasonal fisheries
o Pre-1982 trawl survey data omitted
o Mean-size-at-age data omitted
o Fishery CPUE data omitted
o Average input $N$ set to 100 for all fisheries and the survey
o First reference age for length-at-age relationship set at 0.833333
o Richards growth implemented
o Ageing bias estimated internally
o Selectivities modeled as random walks with age (constant for ages 8+)
Twelve new models were developed during the review itself:

- Model 1 was identical to the 2010 final model except:
o Length at age 0 constrained to be positive
o Richards growth implemented
- Model 2 was identical to the 2010 final model except length at age 0 constrained to be positive
- Model 3 was identical to the 2010 final model except:
o All time blocks removed
o All selectivity parameters freed except fishery selectivity at initial age
o All selectivity parameters initialized at mid-point of bounds
- Model 4 was identical to the 2010 final model except:
o All time blocks removed
o Emphasis on fishery sizecomps set to 0.001
- Model 5 was identical to the 2010 final model except:
o Richards growth implemented
o Ageing bias estimated internally
- Model 6 was identical to Model 4 except time blocks included
- Model 7 was identical to the 2010 final model except $Q$ estimated internally
- Model 8 was identical to the 2010 final model except $M$ estimated internally with an informative prior
- Model 9 was identical to the 2010 final model except tail compression increased
- Model 10 was identical to the 2010 final model except mean-size-at-age data turned off
- Model 11 was the same the "exploratory" model except:
o Pre-1982 trawl survey data included
o All time blocks removed
o Fishery CPUE data included (but not used for estimation)
o Input $N$ set as in the 2010 final model
o First reference age for length-at-age relationship set at as in the 2010 final model
- Model 12 was identical to Model 11 except two iterations of survey variance and input $N$ reweighting added


## Preliminary assessment

Seven models were included:

- Model 1 was identical to the 2010 final model
- Model 2a was identical to Model 1 except for use of spline-based selectivity
- Model 2 b was identical to Model 1 except for omission of pre-1982 survey data
- Model 3 was identical to Model 2b except:
o Ageing bias estimated internally rather than by trial and error
o First reference age for length-at-age relationship (amin) set at 1.0
o Standard deviation of length at age amin tuned iteratively to match the value predicted externally by regression
- Model 4 was identical to Model 2b except:
o All agecomp data turned off
o All sizecomp data turned on
o First reference age for length-at-age relationship (amin) set at 1.0
o Parameters governing standard deviation of length at age estimated internally
- Model A was identical to Model 2 b except:
o First reference age in the mean length-at-age relationship was set at 1.41667 , to coincide with age 1 at the time of year when the survey takes place (in Models 1-2b, first reference age was set at 0 ; in Models $3-4$, it was set at 1 )
o Richards growth equation was used (in Models 1-4, von Bertalanffy was used)
o Ageing bias was estimated internally (as in Model 3; in Models 1-2 and 4, ageing bias was left at the values specified in the 2009 and 2010 assessments-although this was irrelevant for Model 4, which did not attempt to fit the age data)
o $\sigma_{\mathrm{R}}$ was estimated internally (in Models 1-4, this parameter was left at the value used in the 2009 and 2010 assessments)
o Fishery selectivity curves were defined for each of the five seasons, but were not stratified by gear type (in Models 1-4, seasons 1-2 and 4-5 were lumped into a pair of "super" seasons, and fisheries were also gear-specific)
o Selectivity curve for the fishery that came closest to being asymptotic on its own (in this case, the season 4 fishery) was forced to be asymptotic by fixing both width_of_peak_region and final_selectivity at a value of 10.0 and descending_width at a value of 0.0 (in Models 1-4, the Jan-Apr trawl fishery was forced to exhibit asymptotic selectivity)
o Survey selectivity was modeled as a function of length (in Models 1-4, survey selectivity was modeled as a function of age)
o Number of estimated year class strengths in the initial numbers-at-age vector was set at 10 (in Models 1-4, only 3 elements were estimated)
o The following parameters were tuned iteratively:
- Standard deviation of length at the first reference age was tuned iteratively to match the value from the regression of standard deviation against length at age presented in the final assessment for 2010 (as in Model 3; in Models 1-2, this parameter was set at 0.01 because the first reference age was 0 ; in Model 4 , it was estimated internally)
- Base value for $Q$ was tuned iteratively to set the average of the product of $Q$ and survey selectivity across the $60-81 \mathrm{~cm}$ range equal to 0.47 , corresponding to the Nichol et al. (2007) estimate (in Models 1-4, the base value was left at the value used in the 2009 and 2010 assessments)
- $Q$ was given annual (but not random walk) devs, with $\sigma \mathrm{dev}$ tuned iteratively to set the root-mean-squared-standardized-residual of the survey abundance estimates equal to 1.0 (in Models 1-4, $Q$ was constant)
- All estimated selectivity parameters were given annual random walk devs with $\sigma d e v$ tuned iteratively to match the standard deviation of the estimated devs, except that the devs for any selectivity parameter with a tuned $\sigma$ dev less than 0.005 were removed (in Models 1-4, certain fishery selectivity parameters were estimated independently in pre-specified blocks of years; the only time-varying selectivity parameter for the
survey was ascending_width, which had annual—but not random walk-devs with $\sigma d e v$ set at the value used in the 2009 and 2010 assessments)
- Age composition "variance adjustment" multiplier was tuned iteratively to set the mean effective sample size equal to the mean input sample size (in Models 1-4, this multiplier was fixed at 1.0)
- Model 5 was identical to Model A except that it used the time series of selectivity parameters estimated (using random walk devs) in Model A to identify appropriate breakpoints for defining block-specific selectivity parameters

Other model features explored but not included in any of the above:

- Annually varying Brody growth parameter
- Annually varying length at the first reference age
- Internal estimation of standard deviation of length at age
- Ordinary (not random walk) devs for annually varying selectivity parameters
- One selectivity parameter for each age (up to some age-plus group) and fleet, either with ordinary or random walk devs or constant
- Not forcing any fleet to exhibit asymptotic selectivity
- Internal estimation of survey catchability
- Iterative re-weighting of size composition likelihood components
- Internal estimation of the natural mortality rate
- Changing the SS parameter comp_tail_compression (the tails of each age or size composition record are compressed until the specified amount was reached; sometimes referred to as "dynamic binning")
- Changing the SS parameter add_to_comp (this amount was added to each element of each age or size composition vector-both observed and expected, which avoids taking the logarithm of zero and may also have robustness-related attributes)
- Internal estimation of ageing error variances


## Final assessment

Five models were included:

- Model 1 was identical to the 2010 final model (and Model 1 from the preliminary assessment)
- Model 2 b was identical to Model 2 b from the preliminary assessment
- Model 3 was identical to Model 3 from the preliminary assessment
- Model 4 was identical to Model 4 from the preliminary assessment
- Model 3b was identical to Model 3 from the preliminary assessment except:
o Parameters governing variability in length at age estimated internally
o All sizecomp data turned on
o Mean-size-at-age data turned off


## 2012

## Preliminary assessment

Five primary and nine secondary models were included (names of secondary models have decimal points; full results presented for primary models only):

- Model 1 was identical to the 2011 final model
o Model 1.1: Same as Model 1, except survey catchability estimated internally
o Model 1.2: Same as Model 1, except ageing bias parameters fixed at GOA values
o Model 1.3 Same as Model 1, except with revised weight-length representation
- Model 2 was identical to Model 1, except survey catchability re-tuned to match archival tag data
- Model 3 was identical to Model 1, except new fishery selectivity period beginning in 2008
- Model 4 was identical to Model 4 from the final assessment for 2011
o Model Pre5.1: Same as Model 1.3, except for three minor changes to the data file
o Model Pre5.2: Same as Model Pre5.1, except ages 1-10 in the initial vector estimated individually
o Model Pre5.3: Same as Model Pre5.2, except Richards growth curve used
o Model Pre5.4: Same as Model Pre5.3, except $\sigma$ for recruitment devs estimated internally as a free parameter
o Model Pre5.5: Same as Model Pre5.4, except survey selectivity modeled as a function of length
o Model Pre5.6: Same as Model Pre5.5, except fisheries defined by season only (not season-and-gear)
- Model 5: Same as Model Pre5.6, except four quantities estimated iteratively:
o Survey catchability tuned to match archival tag data
o Agecomp $N$ tuned to set the mean ratio of effective $N$ to input $N$ equal to 1
o Selectivity dev sigmas tuned according to the new method described in Annex 2.1.1 of the SAFE chapter


## Final assessment

Four models were included:

- Model 1 was identical to the 2011 final model
- Model 2 was identical to Model 1 except $Q$ was estimated freely
- Model 3 was identical to Model 1 except:
o Ageing bias was not estimated
o All agecomp data are ignored
- Model 4 was identical to Model 5 from the the preliminary assessment


## 2013

## Preliminary assessment

Four models were included:

- Model 1 was identical to the 2012 final model
- Model 2 was identical to Model 4 from the final 2012 assessment except $Q$ estimated internally using a non-constraining uniform prior distribution
- Model 3 was identical to Model 4 from the final 2012 assessment except:
o $Q$ estimated internally using a prior distribution based on archival tagging data
o Survey selectivity forced asymptotic
- Model 4 was identical to Model 4 from the final 2012 assessment


## Final assessment

Due to a protracted government shutdown during the peak of the final assessment season, only one model was presented:

- The unnumbered model was identical to the 2012 final model


## 2014

## Preliminary assessment

Six models were included:

- Model 1 was identical to the 2011-2013 final models
- Model 2 was the identical to Model 5 from the 2012 preliminary assessment (also identical to Model 4 in the 2012 final assessment and the 2013 preliminary assessment)
- Model 3 was identical to Model 2, except that survey catchability $Q$ was fixed at 1.0
- Model 4 was identical to Model 2, except that $Q$ was estimated with a uniform prior and with an internally estimated constant added to each year's log-scale survey abundance standard deviation
- Model 5 was identical to Model 2, except that Q was fixed at 1.0, survey selectivity was forced to be asymptotic, and the natural mortality rate $M$ was estimated freely
- Model 6 was a substantially new model, with the following differences from Model 1:
o Each year consisted of a single season instead of five
0 A single fishery was defined instead of nine season-and-gear-specific fisheries
0 The survey was assumed to sample age 1 fish at true age 1.5 instead of 1.41667
o Initial abundances were estimated for the first ten age groups instead of the first three
0 The natural mortality rate was estimated internally
o The base value of survey catchability was estimated internally
0 Length at age 1.5 was allowed to vary annually
O Survey catchability was allowed to vary annually
0 Selectivity for both the fishery and the survey were allowed to vary annually
O Selectivity for both the fishery and survey was modeled using a random walk with respect to age (SS selectivity-at-age pattern \#17) instead of the usual double normal
o Several quantities were tuned iteratively: prior distributions for selectivity parameters, catchability, and time-varying parameters other than catchability


## Final assessment

Two models were included:

- Model 1 was identical to the 2011-2013 final models
- Model 2 was identical to Model 2 from the preliminary assessment, except that the $L 1$ growth parameter was not allowed to vary with time


## 2015

## Preliminary assessment

Eight models were included.

## Group A:

- Model 0 was the same as Model 1 from the 2014 final assessment.
- Model 7 was the same as Model 0, but with composition data weighted by Equation TA1.8 of Francis (2011).
- Model 8 was the same as Model 0, but with Richards growth (Model 0 used von Bertalanffy growth, which is a special case of Richards growth).


## Subgroup B1:

- Model 2 was the same as Model 2 from the 2014 final assessment.
- Model 3 was the same as Model 2, but with composition data weighted by tuning the mean input sample size to the harmonic mean of the effective sample size, and with time-varying survey catchability $(Q)$ turned off.
- Model 4 was the same as Model 2, but with 20 age groups estimated in the initial numbers-at-age vector (Model 2 estimated 10 age groups in the initial numbers-at-age vector).

For all models in Subgroup B1, selectivity prior distributions and the parameters governing timevariability in recruitment, selectivity, and survey catchability were not re-tuned. That is, they were left at the values estimated for Model 2 during the 2014 assessment, except that time variability in survey catchability was turned off in Model 3. Note that the tuning for Model 2 was performed during the 2014 preliminary assessment (where it was labeled Model 6), and was not updated during the final 2014 assessment.

Subgroup B2:

- Model 5 was based on Model 2, but had a number of differences (described below), one of which was that SS runs were accepted even if the gradient was large, so long as the estimated covariance matrix of the parameters appeared reasonable.
- Model 6 was the same as Model 5, except that SS runs were accepted only if the gradient was small. In the event that a large gradient was obtained, age-specific selectivity dev vectors were removed, one at a time, until the large gradient disappeared.

Except for some procedures related to iterative tuning (see next set paragraph), the differences between Model 5 and Model 2 were as follow:

- Composition data were given a weight of unity if the harmonic mean of the effective sample size was greater than the mean input sample size of 300 ; otherwise, composition data were weighted by tuning the mean input sample size to the harmonic mean of the effective sample size.
- 20 age groups were estimated in the initial numbers-at-age vector.
- Selectivity at ages $9+$ was constrained to equal selectivity at age 8 for both the fishery and the survey.
- A superfluous selectivity parameter was fixed at the mean of the prior (in Model 2, the estimate of this parameter automatically went to the mean of the prior).
- The SS feature known as "Fballpark" was turned off (this feature, which functions something like a very weak prior distribution on the fishing mortality rate in some specified year, did not appear to be providing any benefit in terms of model performance, and what little impact it had on resulting estimates was not easily justified).
- SS runs were accepted even if the gradient was large, so long as the estimated covariance matrix of the parameters appeared reasonable (i.e., all values were numeric, no values were unbelievably large).

Iterative tuning of prior distributions for selectivity parameters and time-varying catchability in Model 5 proceeded as in Model 2, except that all iterative tuning procedures were undertaken simultaneously, rather than in the phased approach used for Model 2. For time-varying recruitment and selectivity, the approach used in Model 2, which was based on the method of Thompson and Lauth (2012), was not retained in Model 5. For a univariate model, if the method of Thompson and Lauth (2012) returns a nonzero estimate of $\sigma$, there is reason to believe that this estimate will be unbiased. However, the method
carries a fairly high probability of returning a "false negative;" that is, returning a zero estimate for $\sigma$ when the true value is non-zero (Thompson in prep.). To reduce this bias toward under-parameterization, the following algorithm was used in Model 5 (Thompson in prep.; note that this is a multivariate generalization of one of the methods mentioned by Methot and Taylor (2011, viz., the third method listed on $p .1749)$ ):

1. Set initial guesses for the os.
2. Run SS.
3. Compute the covariance matrix (V1) of the set of $d e v$ vectors (e.g., element $\{i, j\}$ is equal to the covariance between the subsets of the $i$ th $d e v$ vector and the $j$ th $d e v$ vector consisting of years that those two vectors have in common).
4. Compute the covariance matrix of the parameters (the negative inverse of the Hessian matrix).
5. Extract the part of the covariance matrix of the parameters corresponding to the dev vectors, using only those years common to all dev vectors.
6. Average the values in the matrix obtained in step 5 across years to obtain an "average" covariance matrix (V2).
7. Compute the vector of os corresponding to $\mathbf{V} 1+\mathbf{V} 2$.
8. Return to step 2 and repeat until the os converge.

To speed the above algorithm, the os obtained in step 7 were sometimes substituted with values obtained by extrapolation or interpolation based on previous runs.

As noted above, the procedure used in Model 5 for iterative tuning of time-varying $Q$ was the same as that used in Model 2. However, unlike Model 2, this procedure resulted in time-varying $Q$ being "tuned out" in Model 5. Model 6, which also used this procedure, ended up retaining time-varying $Q$.

## Final assessment

The final assessment included the same two models that were featured in the 2014 final assessment:

- Model 11.5 was identical to the 2011-2014 final models
- Model 14.2 was identical to Model 2 from the 2014 final assessment


## APPENDIX 2.4: SUPPLEMENTAL CATCH DATA

NMFS Alaska Region has made substantial progress in developing a database documenting many of the removals of FMP species that have resulted from activities outside of fisheries prosecuted under the BSAI Groundfish FMP, including removals resulting from scientific research, subsistence fishing, personal use, recreational fishing, exempted fishing permit activities, and commercial fisheries other than those managed under the BSAI groundfish FMP. Estimates for EBS Pacific cod from this dataset are shown in Table 2.4.1.

Although many sources of removal are documented in Table 2.4.1, the time series is highly incomplete for many of these. Cells shaded gray represent data contained in the NMFS database. Other entries represent extrapolations for years in which the respective activity was known or presumed to have taken place, where each extrapolated value consists of the time series average of the official data for the corresponding activity. In the case of surveys, years with missing values were identified from the literature or by contacting individuals knowledgeable about the survey (the NMFS database contains names of contact persons for most activities); in the case of fisheries, it was assumed that the activity occurred every year.

In the 2012 analysis (Attachment 2.4 of Thompson and Lauth 2012), the supplemental catch data were used to provide estimates of potential impacts of these data in the event that they were included in the catch time series used in the assessment model. The results of that analysis indicated that $F_{40 \%}$ increased by about 0.01 and that the one-year-ahead catch corresponding to harvesting at $F_{40 \%}$ decreased by about $4,000 \mathrm{t}$. Note that this is a separate issue from the effects of taking other removals "off the top" when specifying an ABC for the groundfish fishery; the former accounts for the impact on reference points, while the latter accounts for the fact that "other" removals will continue to occur.

The average of the total removals in Table 2.4.1 for the last three complete years (2013-2015) is 8,878 t .
It should be emphasized that these calculations are provided purely for purposes of comparison and discussion, as NMFS and the Council continue to refine policy pertaining to treatment of removals from sources other than the directed groundfish fishery.

## Reference

Thompson, G. G., and R. R. Lauth. 2012. Assessment of the Pacific cod stock in the Eastern Bering Sea and Aleutian Islands Area. In Plan Team for Groundfish Fisheries of the Bering Sea/Aleutian Islands (compiler), Stock assessment and fishery evaluation report for the groundfish resources of the Bering Sea/Aleutian Islands regions, p. 245-544. North Pacific Fishery Management Council, 605 W. 4th Avenue Suite 306, Anchorage, AK 99501.

Table 2.4.1—Total removals of Pacific cod ( t ) from activities not related to directed fishing. Cells shaded gray represent data contained in the NMFS database. Other entries represent extrapolations for years in which the respective activity was known or presumed to have taken place.

| Activity | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Aleutian Island Bottom Trawl Survey |  |  |  | 2 |  |  | 2 |  |  | 2 |  |  |  |  | 2 |  |  | 2 |  |
| Annual Longline Survey |  |  |  |  |  | 28 | 28 | 28 | 28 | 28 | 28 | 28 | 28 | 28 | 28 | 28 | 28 | 28 |  |
| Bait for Crab Fishery | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 |
| Bering Sea Acoustic Survey |  |  | 0 |  |  | 0 |  |  | 0 |  |  | 0 |  |  | 0 |  |  | 0 |  |
| Bering Sea Slope Survey |  |  | 1 |  | 1 | 1 |  |  | 1 |  |  | 1 |  |  | 1 |  |  |  |  |
| Eastern Bering Sea Bottom Trawl Survey | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 |
| Gulf of Alaska Bottom Trawl Survey |  |  |  |  |  |  |  | 0 |  |  | 0 |  |  | 0 |  |  | 0 |  |  |
| IPHC Annual Longline Survey |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Large-Mesh Trawl Survey |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 1 |  |  | 1 | 1 |
| Northern Bering Sea Bottom Trawl Survey |  |  | 1 |  | 1 | 1 |  |  | 1 |  |  | 1 |  |  | 1 |  |  |  |  |
| Pollock EFP 11-01 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Pribilof Islands Crab Survey |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| St. Mathews Crab Survey |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 9 |
| Subsistence Fishery | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 1 | 2 | 0 | 2 | 5 | 2 |
| Summer EBS Survey with Russia |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 |  |


| Activity | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Aleutian Island Bottom Trawl Survey |  | 2 |  |  | 2 |  | 2 |  | 2 |  | 2 |  |  |  | 2 |  | 1 |  | 2 |  |
| Annual Longline Survey |  | 38 |  | 30 |  | 28 |  | 30 |  | 23 |  | 25 |  | 20 |  | 24 |  | 27 |  | 32 |
| Bait for Crab Fishery | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 6547 | 1737 | 4544 | 6697 | 6618 | 9452 | 10233 |
| Bering Sea Acoustic Survey | 0 | 0 |  | 0 | 0 |  | 0 |  | 0 |  | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |  |
| Bering Sea Slope Survey |  |  |  |  | 1 |  | 1 |  | 1 |  |  |  | 1 |  | 2 |  | 1 | 1 | 1 |  |
| Eastern Bering Sea Bottom Trawl Survey | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 40 | 38 | 42 | 52 | 33 | 39 | 39 |
| Gulf of Alaska Bottom Trawl Survey | 0 |  |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |
| IPHC Annual Longline Survey |  |  | 35 | 35 | 35 | 35 | 35 | 35 | 35 | 35 | 35 | 35 | 35 | 35 | 32 | 20 | 17 | 29 | 52 | 59 |
| Large-Mesh Trawl Survey |  |  |  | 1 | 1 |  |  | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 2 | 1 | 1 | 1 |
| Northern Bering Sea Bottom Trawl Survey |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |  |  |  |  |  |
| Pollock EFP 11-01 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 11 | 307 |  |  |  |
| Pribilof Islands Crab Survey |  |  |  |  |  |  |  | 5 |  | 5 |  |  | 5 |  |  | 5 |  |  |  |  |
| St. Mathews Crab Survey |  |  | 9 |  |  | 9 |  |  | 9 |  |  | 9 |  |  | 9 |  |  | 9 |  |  |
| Subsistence Fishery | 2 | 2 | 1 | 0 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| Summer EBS Survey with Russia |  |  |  |  |  |  | 0 |  | 0 |  |  | 0 | 0 | 0 | 0 |  | 0 |  |  |  |

## APPENDIX 2.5: PARALLEL RESULTS FOR THE "HARVEST RECOMMENDATIONS" SECTION, BASED ON MODEL 11.5

The results presented in the "Harvest Recommendations" section of the main text are based on Model 16.6. Because the structure of this model differs substantively from Model 11.5 (the model accepted for the last five years by the SSC), a set of parallel results for the items in that section, based on Model 11.5, is provided here.

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For a stock exploited by multiple gear types, estimation of $F_{35 \%}$ and $F_{40 \%}$ requires an assumption regarding the apportionment of fishing mortality among those gear types. For this assessment, the apportionment was based on Model 11.5's estimates of fishing mortality by gear for the five most recent complete years of data (2011-2015). The average fishing mortality rates for those years implied that total fishing mortality was divided among the three main gear types according to the following percentages: trawl $32.6 \%$, longline $51.8 \%$, and pot $15.6 \%$. This apportionment results in estimates of $F_{35 \%}$ and $F_{40 \%}$ equal to 0.34 and 0.28 , respectively. Model 11.5's estimates of $B_{100 \%}, B_{40 \%}$, and $B_{35 \%}$ are $788,000 \mathrm{t}$, 315,000 t, and 276,000 t, respectively.

## Specification of OFL and Maximum Permissible ABC

Given the assumptions of Scenario 2 (below), female spawning biomass for 2017 and 2018 is estimated by Model 11.5 to be well above the $B_{40 \%}$ value of 315,000 t, thereby placing Pacific cod in sub-tier "a" of Tier 3 for both 2017 and 2018. Given this, Model 11.5 estimates OFL, maximum permissible ABC, and the associated fishing mortality rates for 2017 and 2018 as follows:

| Year | Overfishing Level | Maximum Permissible ABC |
| ---: | ---: | ---: |
| 2017 | OFL $=396,000 \mathrm{t}$ | $\operatorname{maxABC}=338,000 \mathrm{t}$ |
| 2018 | OFL $=381,000 \mathrm{t}$ | $\operatorname{maxABC}=325,000 \mathrm{t}$ |
| 2017 | $F O F L=0.34$ | $\operatorname{maxFABC}=0.28$ |
| 2018 | $F O F L=0.34$ | $\operatorname{maxFABC}=0.28$ |

The age $0+$ biomass projections for 2017 and 2018 from Model 11.5 (using SS rather than the standard projection model) are 1,760,000 t and 1,580,000 t. For comparison, the age 3+ biomass projections for 2017 and 2018 from Model 11.5 (again using SS) are 1,750,000 t and 1,540,000 t.

## Standard Harvest Scenarios, Projection Methodology, and Projection Results

The standard harvest scenarios and projection methodology were the same as described for Model 16.6 in the main text. Projections corresponding to the standard scenarios are shown for Model 11.5 in Tables 2.5.30-2.5.36 (table numbering is kept the same as in the main text, so as to facilitate comparisons).

## Status Determination

Methodology for status determination is as described in the main text. The status with respect to overfishing is independent of model choice for next year's specifications, as it depends entirely on the previous year's catch and OFL.

Based on the criteria described in the main text Tables 2.5.35 and 2.5.36, the stock is not overfished and is not approaching an overfished condition.

Table 2.5.30—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=\max F_{A B C}$ in 2017-2029 (Scenario 1), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 338,000 | 338,000 | 338,000 | 338,000 | 0 |
| 2018 | 312,000 | 312,000 | 312,000 | 312,000 | 0 |
| 2019 | 267,000 | 267,000 | 267,000 | 267,000 | 4 |
| 2020 | 232,000 | 233,000 | 233,000 | 236,000 | 1,264 |
| 2021 | 196,000 | 209,000 | 214,000 | 245,000 | 16,419 |
| 2022 | 163,000 | 197,000 | 207,000 | 279,000 | 38,109 |
| 2023 | 140,000 | 202,000 | 212,000 | 321,000 | 56,840 |
| 2024 | 125,000 | 213,000 | 219,000 | 338,000 | 67,951 |
| 2025 | 117,000 | 219,000 | 225,000 | 345,000 | 73,411 |
| 2026 | 118,000 | 225,000 | 228,000 | 356,000 | 75,089 |
| 2027 | 118,000 | 229,000 | 230,000 | 358,000 | 74,578 |
| 2028 | 121,000 | 227,000 | 230,000 | 357,000 | 73,106 |
| 2029 | 121,000 | 229,000 | 230,000 | 358,000 | 72,325 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 431,000 | 431,000 | 431,000 | 431,000 | 0 |
| 2018 | 432,000 | 432,000 | 432,000 | 432,000 | 0 |
| 2019 | 394,000 | 394,000 | 394,000 | 394,000 | 44 |
| 2020 | 340,000 | 341,000 | 341,000 | 343,000 | 983 |
| 2021 | 299,000 | 304,000 | 306,000 | 318,000 | 6,532 |
| 2022 | 273,000 | 290,000 | 295,000 | 335,000 | 20,644 |
| 2023 | 254,000 | 289,000 | 299,000 | 375,000 | 40,654 |
| 2024 | 239,000 | 295,000 | 307,000 | 423,000 | 57,415 |
| 2025 | 230,000 | 301,000 | 314,000 | 440,000 | 67,371 |
| 2026 | 228,000 | 303,000 | 320,000 | 447,000 | 72,770 |
| 2027 | 230,000 | 309,000 | 323,000 | 461,000 | 74,884 |
| 2028 | 231,000 | 308,000 | 324,000 | 465,000 | 74,125 |
| 2029 | 233,000 | 308,000 | 324,000 | 467,000 | 72,460 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.28 | 0.28 | 0.28 | 0.28 | 0.00 |
| 2018 | 0.28 | 0.28 | 0.28 | 0.28 | 0.00 |
| 2019 | 0.28 | 0.28 | 0.28 | 0.28 | 0.00 |
| 2020 | 0.28 | 0.28 | 0.28 | 0.28 | 0.00 |
| 2021 | 0.27 | 0.27 | 0.27 | 0.28 | 0.00 |
| 2022 | 0.24 | 0.26 | 0.26 | 0.28 | 0.01 |
| 2023 | 0.22 | 0.26 | 0.26 | 0.28 | 0.02 |
| 2024 | 0.21 | 0.26 | 0.26 | 0.28 | 0.03 |
| 2025 | 0.20 | 0.27 | 0.26 | 0.28 | 0.03 |
| 2026 | 0.20 | 0.27 | 0.26 | 0.28 | 0.03 |
| 2027 | 0.20 | 0.28 | 0.26 | 0.28 | 0.03 |
| 2028 | 0.20 | 0.27 | 0.26 | 0.28 | 0.03 |
| 2029 | 0.20 | 0.27 | 0.26 | 0.28 | 0.03 |

Table 2.5.31—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that catch in 2017 is less than ABC by an amount predicted from past performance, but that $F=\max F_{A B C}$ in 2018-2029 (Scenario 2), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 262,000 | 262,000 | 262,000 | 262,000 | 0 |
| 2018 | 254,000 | 254,000 | 254,000 | 254,000 | 0 |
| 2019 | 290,000 | 290,000 | 290,000 | 290,000 | 4 |
| 2020 | 247,000 | 248,000 | 249,000 | 251,000 | 1,264 |
| 2021 | 216,000 | 226,000 | 230,000 | 255,000 | 13,599 |
| 2022 | 172,000 | 208,000 | 217,000 | 284,000 | 37,344 |
| 2023 | 144,000 | 206,000 | 216,000 | 325,000 | 56,664 |
| 2024 | 126,000 | 215,000 | 220,000 | 339,000 | 68,040 |
| 2025 | 118,000 | 220,000 | 225,000 | 347,000 | 73,534 |
| 2026 | 118,000 | 225,000 | 228,000 | 357,000 | 75,186 |
| 2027 | 118,000 | 229,000 | 230,000 | 358,000 | 74,636 |
| 2028 | 121,000 | 227,000 | 230,000 | 357,000 | 73,137 |
| 2029 | 121,000 | 229,000 | 230,000 | 358,000 | 72,339 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 440,000 | 440,000 | 440,000 | 440,000 | 0 |
| 2018 | 462,000 | 462,000 | 462,000 | 462,000 | 0 |
| 2019 | 432,000 | 432,000 | 432,000 | 432,000 | 44 |
| 2020 | 368,000 | 368,000 | 369,000 | 370,000 | 983 |
| 2021 | 316,000 | 322,000 | 323,000 | 336,000 | 6,919 |
| 2022 | 282,000 | 299,000 | 305,000 | 346,000 | 21,591 |
| 2023 | 258,000 | 293,000 | 303,000 | 382,000 | 41,660 |
| 2024 | 241,000 | 297,000 | 309,000 | 427,000 | 58,213 |
| 2025 | 230,000 | 301,000 | 315,000 | 443,000 | 67,870 |
| 2026 | 228,000 | 303,000 | 320,000 | 447,000 | 73,029 |
| 2027 | 230,000 | 309,000 | 323,000 | 461,000 | 75,002 |
| 2028 | 231,000 | 308,000 | 324,000 | 465,000 | 74,173 |
| 2029 | 233,000 | 308,000 | 324,000 | 467,000 | 72,477 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.21 | 0.21 | 0.21 | 0.21 | 0.00 |
| 2018 | 0.22 | 0.22 | 0.22 | 0.22 | 0.00 |
| 2019 | 0.28 | 0.28 | 0.28 | 0.28 | 0.00 |
| 2020 | 0.28 | 0.28 | 0.28 | 0.28 | 0.00 |
| 2021 | 0.28 | 0.28 | 0.28 | 0.28 | 0.00 |
| 2022 | 0.25 | 0.27 | 0.27 | 0.28 | 0.01 |
| 2023 | 0.23 | 0.26 | 0.26 | 0.28 | 0.02 |
| 2024 | 0.21 | 0.26 | 0.26 | 0.28 | 0.03 |
| 2025 | 0.20 | 0.27 | 0.26 | 0.28 | 0.03 |
| 2026 | 0.20 | 0.27 | 0.26 | 0.28 | 0.03 |
| 2027 | 0.20 | 0.28 | 0.26 | 0.28 | 0.03 |
| 2028 | 0.20 | 0.27 | 0.26 | 0.28 | 0.03 |
| 2029 | 0.20 | 0.27 | 0.26 | 0.28 | 0.03 |

Table 2.5.32—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that the upper bound on $F_{A B C}$ is set the most recent five-year average fishing mortality rate in 2017-2029 (Scenario 3), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 365,000 | 365,000 | 365,000 | 365,000 | 0 |
| 2018 | 331,000 | 331,000 | 331,000 | 331,000 | 0 |
| 2019 | 280,000 | 280,000 | 280,000 | 280,000 | 4 |
| 2020 | 241,000 | 242,000 | 242,000 | 245,000 | 1,374 |
| 2021 | 213,000 | 224,000 | 228,000 | 255,000 | 14,739 |
| 2022 | 191,000 | 220,000 | 228,000 | 291,000 | 33,585 |
| 2023 | 172,000 | 223,000 | 233,000 | 336,000 | 50,865 |
| 2024 | 159,000 | 228,000 | 238,000 | 352,000 | 61,363 |
| 2025 | 151,000 | 231,000 | 242,000 | 359,000 | 66,900 |
| 2026 | 150,000 | 234,000 | 245,000 | 371,000 | 69,070 |
| 2027 | 150,000 | 236,000 | 245,000 | 370,000 | 68,639 |
| 2028 | 152,000 | 234,000 | 245,000 | 370,000 | 66,993 |
| 2029 | 151,000 | 235,000 | 244,000 | 369,000 | 66,159 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 428,000 | 428,000 | 428,000 | 428,000 | 0 |
| 2018 | 422,000 | 422,000 | 422,000 | 422,000 | 0 |
| 2019 | 379,000 | 379,000 | 379,000 | 379,000 | 44 |
| 2020 | 322,000 | 323,000 | 323,000 | 325,000 | 983 |
| 2021 | 280,000 | 285,000 | 287,000 | 300,000 | 6,898 |
| 2022 | 250,000 | 269,000 | 274,000 | 317,000 | 22,179 |
| 2023 | 224,000 | 264,000 | 274,000 | 356,000 | 43,704 |
| 2024 | 203,000 | 268,000 | 279,000 | 402,000 | 61,599 |
| 2025 | 189,000 | 273,000 | 284,000 | 416,000 | 72,182 |
| 2026 | 184,000 | 275,000 | 289,000 | 423,000 | 77,726 |
| 2027 | 183,000 | 280,000 | 291,000 | 432,000 | 79,663 |
| 2028 | 184,000 | 280,000 | 292,000 | 438,000 | 78,782 |
| 2029 | 183,000 | 280,000 | 291,000 | 438,000 | 77,148 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2018 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2019 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2020 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2021 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2022 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2023 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2024 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2025 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2026 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2027 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2028 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2029 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |

Table 2.5.33—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that the upper bound on $F_{A B C}$ is set at $F_{60 \%}$ in 2017-2029 (Scenario 4), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 176,000 | 176,000 | 176,000 | 176,000 | 0 |
| 2018 | 178,000 | 178,000 | 178,000 | 178,000 | 0 |
| 2019 | 166,000 | 166,000 | 166,000 | 166,000 | 2 |
| 2020 | 153,000 | 153,000 | 153,000 | 155,000 | 630 |
| 2021 | 141,000 | 146,000 | 148,000 | 161,000 | 6,890 |
| 2022 | 130,000 | 144,000 | 148,000 | 179,000 | 16,420 |
| 2023 | 120,000 | 146,000 | 151,000 | 205,000 | 26,408 |
| 2024 | 112,000 | 150,000 | 155,000 | 219,000 | 33,680 |
| 2025 | 106,000 | 152,000 | 158,000 | 224,000 | 38,191 |
| 2026 | 105,000 | 153,000 | 160,000 | 235,000 | 40,514 |
| 2027 | 103,000 | 155,000 | 161,000 | 235,000 | 41,165 |
| 2028 | 105,000 | 156,000 | 161,000 | 237,000 | 40,672 |
| 2029 | 104,000 | 157,000 | 162,000 | 236,000 | 40,161 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 450,000 | 450,000 | 450,000 | 450,000 | 0 |
| 2018 | 498,000 | 498,000 | 498,000 | 498,000 | 0 |
| 2019 | 500,000 | 500,000 | 500,000 | 500,000 | 44 |
| 2020 | 468,000 | 469,000 | 469,000 | 471,000 | 985 |
| 2021 | 432,000 | 437,000 | 439,000 | 452,000 | 7,036 |
| 2022 | 401,000 | 421,000 | 427,000 | 473,000 | 23,733 |
| 2023 | 371,000 | 416,000 | 428,000 | 524,000 | 50,269 |
| 2024 | 343,000 | 421,000 | 436,000 | 591,000 | 76,480 |
| 2025 | 322,000 | 428,000 | 444,000 | 624,000 | 95,296 |
| 2026 | 308,000 | 436,000 | 452,000 | 640,000 | 106,998 |
| 2027 | 306,000 | 442,000 | 458,000 | 665,000 | 113,140 |
| 2028 | 304,000 | 445,000 | 461,000 | 669,000 | 114,582 |
| 2029 | 301,000 | 448,000 | 463,000 | 672,000 | 113,442 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |
| 2018 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |
| 2019 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |
| 2020 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |
| 2021 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |
| 2022 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |
| 2023 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |
| 2024 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |
| 2025 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |
| 2026 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |
| 2027 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |
| 2028 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |
| 2029 | 0.14 | 0.14 | 0.14 | 0.14 | 0.00 |

Table 2.5.34—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=0$ in 2017-2029 (Scenario 5), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 0 | 0 | 0 | 0 | 0 |
| 2023 | 0 | 0 | 0 | 0 | 0 |
| 2024 | 0 | 0 | 0 | 0 | 0 |
| 2025 | 0 | 0 | 0 | 0 | 0 |
| 2026 | 0 | 0 | 0 | 0 | 0 |
| 2027 | 0 | 0 | 0 | 0 | 0 |
| 2028 | 0 | 0 | 0 | 0 | 0 |
| 2029 | 0 | 0 | 0 | 0 | 0 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 469,000 | 469,000 | 469,000 | 469,000 | 0 |
| 2018 | 574,000 | 574,000 | 574,000 | 574,000 | 0 |
| 2019 | 634,000 | 634,000 | 634,000 | 634,000 | 44 |
| 2020 | 648,000 | 649,000 | 649,000 | 651,000 | 987 |
| 2021 | 640,000 | 645,000 | 647,000 | 661,000 | 7,154 |
| 2022 | 625,000 | 646,000 | 652,000 | 700,000 | 25,135 |
| 2023 | 603,000 | 652,000 | 666,000 | 775,000 | 56,696 |
| 2024 | 576,000 | 668,000 | 686,000 | 873,000 | 92,585 |
| 2025 | 552,000 | 686,000 | 706,000 | 945,000 | 122,848 |
| 2026 | 530,000 | 702,000 | 724,000 | 988,000 | 144,846 |
| 2027 | 524,000 | 715,000 | 739,000 | $1,020,000$ | 159,213 |
| 2028 | 520,000 | 727,000 | 750,000 | $1,050,000$ | 166,537 |
| 2029 | 522,000 | 738,000 | 757,000 | $1,060,000$ | 168,671 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2018 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2019 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2020 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2021 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2022 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2023 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2024 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2025 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2026 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2027 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2028 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2029 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |

Table 2.5.35—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=F_{\text {OFL }}$ in 2017-2029 (Scenario 6), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 396,000 | 396,000 | 396,000 | 396,000 | 0 |
| 2018 | 353,000 | 353,000 | 353,000 | 353,000 | 0 |
| 2019 | 293,000 | 293,000 | 293,000 | 293,000 | 5 |
| 2020 | 241,000 | 243,000 | 243,000 | 247,000 | 2,131 |
| 2021 | 189,000 | 203,000 | 208,000 | 243,000 | 18,956 |
| 2022 | 162,000 | 198,000 | 211,000 | 301,000 | 45,354 |
| 2023 | 142,000 | 209,000 | 223,000 | 358,000 | 67,524 |
| 2024 | 128,000 | 222,000 | 234,000 | 373,000 | 79,141 |
| 2025 | 121,000 | 228,000 | 241,000 | 382,000 | 84,426 |
| 2026 | 122,000 | 232,000 | 244,000 | 393,000 | 85,613 |
| 2027 | 123,000 | 235,000 | 245,000 | 391,000 | 84,663 |
| 2028 | 124,000 | 234,000 | 244,000 | 396,000 | 83,011 |
| 2029 | 127,000 | 237,000 | 244,000 | 391,000 | 82,306 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 424,000 | 424,000 | 424,000 | 424,000 | 0 |
| 2018 | 410,000 | 410,000 | 410,000 | 410,000 | 0 |
| 2019 | 361,000 | 361,000 | 361,000 | 361,000 | 44 |
| 2020 | 304,000 | 304,000 | 304,000 | 306,000 | 887 |
| 2021 | 267,000 | 272,000 | 273,000 | 285,000 | 6,143 |
| 2022 | 248,000 | 264,000 | 269,000 | 306,000 | 19,393 |
| 2023 | 233,000 | 267,000 | 275,000 | 348,000 | 37,300 |
| 2024 | 221,000 | 275,000 | 283,000 | 385,000 | 51,225 |
| 2025 | 214,000 | 280,000 | 290,000 | 395,000 | 58,871 |
| 2026 | 213,000 | 282,000 | 294,000 | 408,000 | 62,832 |
| 2027 | 215,000 | 284,000 | 296,000 | 412,000 | 64,070 |
| 2028 | 215,000 | 284,000 | 296,000 | 419,000 | 62,861 |
| 2029 | 216,000 | 284,000 | 295,000 | 418,000 | 61,231 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.34 | 0.34 | 0.34 | 0.34 | 0.00 |
| 2018 | 0.34 | 0.34 | 0.34 | 0.34 | 0.00 |
| 2019 | 0.34 | 0.34 | 0.34 | 0.34 | 0.00 |
| 2020 | 0.32 | 0.32 | 0.32 | 0.33 | 0.00 |
| 2021 | 0.28 | 0.29 | 0.29 | 0.30 | 0.01 |
| 2022 | 0.26 | 0.28 | 0.28 | 0.33 | 0.02 |
| 2023 | 0.24 | 0.28 | 0.29 | 0.34 | 0.03 |
| 2024 | 0.23 | 0.29 | 0.29 | 0.34 | 0.04 |
| 2025 | 0.22 | 0.30 | 0.29 | 0.34 | 0.04 |
| 2026 | 0.22 | 0.30 | 0.29 | 0.34 | 0.04 |
| 2027 | 0.22 | 0.30 | 0.29 | 0.34 | 0.04 |
| 2028 | 0.22 | 0.30 | 0.30 | 0.34 | 0.04 |
| 2029 | 0.22 | 0.30 | 0.29 | 0.34 | 0.04 |

Table 2.5.36—Projections for EBS Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=\max F_{A B C}$ in each year 2017-2018 and $F=F_{\text {OFL }}$ thereafter (Scenario 7), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 338,000 | 338,000 | 338,000 | 338,000 | 0 |
| 2018 | 312,000 | 312,000 | 312,000 | 312,000 | 0 |
| 2019 | 313,000 | 313,000 | 313,000 | 313,000 | 5 |
| 2020 | 263,000 | 264,000 | 264,000 | 267,000 | 1,505 |
| 2021 | 201,000 | 215,000 | 220,000 | 257,000 | 19,688 |
| 2022 | 166,000 | 203,000 | 215,000 | 307,000 | 45,333 |
| 2023 | 143,000 | 210,000 | 224,000 | 358,000 | 67,298 |
| 2024 | 129,000 | 221,000 | 234,000 | 372,000 | 79,034 |
| 2025 | 121,000 | 228,000 | 240,000 | 382,000 | 84,413 |
| 2026 | 122,000 | 232,000 | 244,000 | 393,000 | 85,629 |
| 2027 | 123,000 | 235,000 | 245,000 | 391,000 | 84,676 |
| 2028 | 124,000 | 234,000 | 244,000 | 396,000 | 83,017 |
| 2029 | 127,000 | 237,000 | 244,000 | 391,000 | 82,308 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2017 | 431,000 | 431,000 | 431,000 | 431,000 | 0 |
| 2018 | 432,000 | 432,000 | 432,000 | 432,000 | 0 |
| 2019 | 388,000 | 388,000 | 388,000 | 388,000 | 44 |
| 2020 | 321,000 | 322,000 | 322,000 | 324,000 | 983 |
| 2021 | 276,000 | 281,000 | 283,000 | 295,000 | 6,288 |
| 2022 | 252,000 | 268,000 | 273,000 | 310,000 | 19,442 |
| 2023 | 234,000 | 268,000 | 277,000 | 349,000 | 37,311 |
| 2024 | 222,000 | 275,000 | 284,000 | 386,000 | 51,224 |
| 2025 | 214,000 | 280,000 | 290,000 | 396,000 | 58,865 |
| 2026 | 213,000 | 282,000 | 294,000 | 408,000 | 62,822 |
| 2027 | 214,000 | 284,000 | 296,000 | 412,000 | 64,058 |
| 2028 | 215,000 | 284,000 | 296,000 | 419,000 | 62,850 |
| 2029 | 216,000 | 284,000 | 295,000 | 418,000 | 61,222 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2017 | 0.28 | 0.28 | 0.28 | 0.28 | 0.00 |
| 2018 | 0.28 | 0.28 | 0.28 | 0.28 | 0.00 |
| 2019 | 0.34 | 0.34 | 0.34 | 0.34 | 0.00 |
| 2020 | 0.34 | 0.34 | 0.34 | 0.34 | 0.00 |
| 2021 | 0.29 | 0.30 | 0.30 | 0.31 | 0.01 |
| 2022 | 0.26 | 0.28 | 0.29 | 0.33 | 0.02 |
| 2023 | 0.24 | 0.28 | 0.29 | 0.34 | 0.03 |
| 2024 | 0.23 | 0.29 | 0.29 | 0.34 | 0.04 |
| 2025 | 0.22 | 0.30 | 0.29 | 0.34 | 0.04 |
| 2026 | 0.22 | 0.30 | 0.29 | 0.34 | 0.04 |
| 2027 | 0.22 | 0.30 | 0.29 | 0.34 | 0.04 |
| 2028 | 0.22 | 0.30 | 0.29 | 0.34 | 0.04 |
| 2029 | 0.22 | 0.30 | 0.29 | 0.34 | 0.04 |

