

4. Assessment of the yellowfin sole stock in the Bering Sea and Aleutian Islands

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Executive Summary

Summary of Changes in Assessment Inputs

Changes to the input data

- 1) 2015 fishery age composition.
- 2) 2015 survey age composition.
- 3) 2016 trawl survey biomass point estimate and standard error.
- 4) Estimate of the discarded and retained portions of the 2015 catch.
- 5) Estimate of total catch made through the end of 2016. Catch of 150,000 t assumed for 2017 and 2018 projection.

Changes to the assessment methodology

Changes were made to the fishery weight-at-age where the average of the fishery aged samples from 2008-2014 were used for 2008-2016, replacing previous values that were time-invariant. The assessment updates last year's with results and management quantities that are higher than the 2015 assessment primarily due to a 48% increase in the 2016 survey biomass and the fishery weight-at-age changes. Yellowfin sole continue to be well-above B_{MSY} and the annual harvest remains below the ABC level. The female spawning stock is in a slow downward trend.

Summary of Results

Quantity	As estimated or <i>specified last year for:</i>		As estimated or <i>recommended this year for:</i>	
	2016	2017	2017	2018
M (natural mortality rate)	0.12	0.12	0.12	0.12
Tier	1a	1a	1a	1a
Projected total (age 6+) biomass (t)	2,170,000	2,086,200	2,290,100	2,202,300
Female spawning biomass (t)				
Projected	702,200	696,200	778,600	770,900
B_0	1,107,000		1,202,700	
B_{MSY}	435,000		424,000	
F_{OFL}	0.105	0.105	0.125	0.125
$maxF_{ABC}$	0.098	0.098	0.114	0.114
F_{ABC}	0.098	0.098	0.114	0.114
OFL (t)	228,100	219,200	287,000	276,000
maxABC (t)	211,700	203,500	260,800	250,800
ABC (t)	211,700	203,500	260,800	250,800
Status	As determined <i>last year for:</i>		As determined <i>this year for:</i>	
	2014	2015	2015	2016
Overfishing	No	n/a	No	n/a

Overfished	n/a	No	n/a	No
Approaching overfished	n/a	No	n/a	No

Projections are based on estimated catches of 150,000 t used in place of maximum ABC for 2017 and 2018.

Responses to SSC and Plan Team Comments on Assessments in General

General comments for all assessments:

The PT discussion relative to its choice of a preferred model for the BSAI northern rock sole assessment raised an important issue about using localized gear performance studies (e.g. Somerton and Munro 2001) to inform or fix estimates of catchability (Q). The PT pointed out that gear herding experiments can inform the estimation of Q , but support a very limited scope of inference given the broad spatial and temporal distributions of the factors influencing Q . Currently assessment authors are applying a variety of approaches to calculating Q including fixing the value, fitting it in the assessment model, fitting it with priors based on field studies, and estimating it as a temperature-dependent parameter. The SSC notes that Q relates survey abundance to stock size and fishing mortality to fishing effort for the stock area and survey or fishery time series, and as such is a direct scalar on the survey abundance estimates. Both the fish herding characteristics of the survey trawl and the timing of fish migrations (especially flatfish) impact Q , and these factors are known to be influenced by water temperature. The SSC recommends that assessment authors work with AFSC's survey program scientist to develop some objective criteria to inform the best approaches for calculating Q with respect to information provided by previous survey trawl performance studies (e.g. Somerton and Munro 2001), and fish-temperature relationships which may impact Q .

The AFSC survey scientists (RACE Division) met last summer (including the assessment scientist) to discuss implementing a field experiment regarding temperature and catchability. Dr. Somerton responded to a RFP from the Sustainable Fisheries Division (NOAA) and was awarded \$125,000 to be used in a 2016 field experiment implemented as an extension to the standard Bering Sea trawl survey. The RACE division scientists set out on a project to demonstrate that temperature influences the sampling efficiency of yellowfin sole by the eastern Bering Sea trawl by trawling in two areas with very different temperatures, that would not be too far apart, and near the cold pool. However, the results of last summer's experiment showed, instead, that sea-state had a much larger effect on yellowfin sole herding than bottom temperature. Dan Nichol then looked at the time-series of eastern Bering Sea yellowfin sole biomass estimates and found that they also were strongly correlated to wave height. To confuse things more, wave height was correlated to temperature. The experimental results then became a paper on wave height effect on yellowfin sole catchability (*The effects of wave-induced vessel motion on the geometry of a bottom survey trawl and the herding of yellowfin sole by Somerton, Weinberg, Munro and Rugolo, In Prep.*) and not the effects of temperature on catchability. However, this still did not answer the question of why you find a strong temperature effect on the survey catchability function in the stock assessment model. One possibility is that trawl sampling efficiency is more influenced by waves but yellowfin sole availability to the survey is more influenced by temperature.

Dr. Somerton retired at the end of October but proposed the following team effort to explore the issue further. Wayne Palsson will be the lead. 1. Peter Munro / Lou Rugolo will re-analyze the experimental data to show the relative effects of temperature and wave effects, because a strong contrast in both effects was present and an experimental design that was sufficiently balanced so that they are not confounded.

2. Dan Nichol has completed a CPUE versus waves analysis back to 2005. Although RACE surveys have recorded sea state data for many years it was not entered into the database earlier than 2005, so Dan will do some hand entry, then repeat the analysis with more years and considering wave height and temperature together in a model.

3. Dan Nichol also has some ideas about how to show that yellowfin sole availability changes with temperature using sex ratio, based on the notion that males arrive first and leave the spawning areas later than females.

In terms of the stock assessment, we will explore an additional waves parameter to fit the survey biomass. Of course, none of this is satisfactory for those who think that localized experiments are not useful due to their limited scope of inference being applied to problems that are large, both temporally and spatially.

Responses to SSC and Plan Team Comments Specific to this Assessment

No assessment-specific comments were received.

Introduction

The yellowfin sole (*Limanda aspera*) is one of the most abundant flatfish species in the eastern Bering Sea (EBS) and currently is the target of the largest flatfish fishery in the world. They inhabit the EBS shelf and are considered one stock. Abundance in the Aleutian Islands region is negligible.

Yellowfin sole are distributed in North American waters from off British Columbia, Canada, (approx. lat. 49° N) to the Chukchi Sea (about lat. 70° N) and south along the Asian coast to about lat. 35° N off the South Korean coast in the Sea of Japan. Adults exhibit a benthic lifestyle and occupy separate winter, spawning and summertime feeding distributions on the eastern Bering Sea shelf. From over-winter grounds near the shelf margins, adults begin a migration onto the inner shelf in April or early May each year for spawning and feeding. The directed fishery has typically occurred from winter through autumn (Wilderbuer et al. 1992). Yellowfin sole are managed as a single stock in the BSAI management area as there is presently no evidence of stock structure.

Fishery

Yellowfin sole have annually been caught with bottom trawls on the Bering Sea shelf since the fishery began in 1954 and were overexploited by foreign fisheries in 1959-62 when catches averaged 404,000 t annually (Fig. 4.1, top panel). As a result of reduced stock abundance, catches declined to an annual average of 117,800 t from 1963-71 and further declined to an annual average of 50,700 t from 1972-77. The lower yield in this latter period was partially due to the discontinuation of the U.S.S.R. fishery. In the early 1980s, after the stock condition had improved, catches again increased reaching a peak of over 227,000 t in 1985.

During the 1980s, there was also a major transition in the characteristics of the fishery. Yellowfin sole were traditionally taken exclusively by foreign fisheries and these fisheries continued to dominate through 1984. However, U.S. fisheries developed rapidly during the 1980s in the form of joint ventures, and during the last half of the decade began to dominate and then take all of the catch as the foreign fisheries were phased out of the EBS. Since 1990, only domestic harvesting and processing has occurred.

The management of the yellowfin sole fishery changed significantly in 2008 with the implementation of Amendment 80 to the BSAI Fisheries Management Plan. The Amendment directly allocated fishery resources among BSAI trawl harvesters in consideration of their historic harvest patterns and future

harvest needs in order to improve retention and utilization of fishery resources by the non-AFA trawl catcher/processor fleet. This was accomplished by extending the groundfish retention standards to all H&G vessels and also by providing the ability to form cooperatives within the newly formed Amendment 80 sector. In addition, Amendment 80 also mandated additional monitoring requirements which included observer coverage on all hauls, motion-compensating scales for weighing samples, flow scales to obtain accurate catch weight estimates for the entire catch, no mixing of hauls and no on-deck sorting. The partitioning of TAC and PSC (prohibited species catch) among cooperatives has significantly changed the way the annual catch has accumulated (Fig 4.1, bottom panel) and the rate of target catch per bycatch ton. There is now a more even and slow attainment of the annual catch relative to the pre-Amendment 80 fishing behavior.

Yellowfin sole are usually headed and gutted, frozen at sea, and then shipped to Asian countries for further processing (AFSC 2016). The first wholesale value of Alaska yellowfin sole totaled \$97.8 million in 2014. In 2010, following a comprehensive assessment process, the yellowfin sole fishery was certified under the Marine Stewardship Council environmental standard for sustainable and well-managed fisheries. The certification also applies to all the major flatfish fisheries in the BSAI and GOA. The total annual catch (t) since implementation of the MFCMA in 1977 is shown in Table 4.1.

Also in 2010, federally permitted vessels using non-pelagic trawl gear whose harvest results in flatfish retained catch that is greater than any other retained fishery category were required to use modified trawl gear. The modifications required the use of elevating devices to raise the section of the trawl warps between the doors and the trawl wing tips by 2.5 inches off the seafloor. The purpose of the management action was to reduce damage of non-target animals, particularly those that form habitat structure or support other fisheries while not substantially reducing flatfish catch rates or causing gear handling problems (Rose et al. 2010).

The 1997 catch of 181,389 t (retained and discards) was the largest since the fishery became completely domestic but was at lower levels from 1998 – 2010, averaging 94,004 t (Table 4.2). From 2011-2014 the catch increased, averaging 155,000 t. The 2013 catch totaled 165,000 t (73% of the ABC), the highest annual catch in the past 19 years. For 2016, the catch distribution has been spread out from January through May with the majority coming from 4 BSAI management areas (509, 513, 514, 516). As of mid-October 2016, the fishing season is ongoing. In order to estimate the total 2016 catch for the stock assessment model, the average proportion of the 2010-2015 cumulative catch attained by the 37th week of the year (mid-September) was applied to the 2016 catch amount at the same time period and results in a 2016 catch estimate of 130,000 t (62% of the ABC). The size composition of the 2016 catch for both males and females, from observer sampling, are shown in Figure 4.2, the catch proportions by month and area are shown in Figure 4.3, and maps of the locations where yellowfin sole were caught in 2016, by month (through mid-September), are shown in Figure 4.4. The average age of yellowfin sole in the 2015 catch is estimated at 12.6 and 12.56 years for females and males, respectively.

The time-series of catch in Table 4.1 also includes yellowfin sole that were discarded in domestic fisheries during the period 1987 to the present. Annual discard estimates were calculated from at-sea sampling (Table 4.2). The rate of discard has ranged from a low of 2% of the total catch in 2012 (and 2015) to 30% in 1992. The trend has been toward fuller retention of the catch in recent years, and with the advent of the Amendment 80 harvest practices, discarding is at its lowest level since these estimates have become available. Historically, discarding primarily occurred in the yellowfin sole directed fishery, with lesser amounts in the Pacific cod, Pollock, rock sole, flathead sole, and “other flatfish” fisheries (Table 4.3).

Data

The data used in this assessment include estimates of total catch, bottom trawl survey biomass estimates and their attendant 95% confidence intervals, catch-at-age from the fishery, and population age

composition estimates from the bottom trawl survey. Weight-at-age and proportion mature-at-age are also available from studies conducted during the bottom trawl surveys.

Data source	years
Fishery catch	1954-2016
Fishery age composition	1964-2015
Survey biomass and standard error, bottom temperature	1982-2016
Survey age composition	1979-2015
Annual length-at-age and weight-at-age	1979-2015
Maturity at age	Combined 1992 and 2012 samples

Fishery Catch and Catch-at-Age

This assessment uses fishery catch data from 1955- 2016 (shown for 1964-2016 in Table 4.1), including an estimate of the 2016 catch, and fishery catch-at-age (proportions) from 1964-2015 (Table 4.4, 1975-2015). The 2015 fishery age composition is primarily composed of fish older than 9 years with a large amount of 20+ fish.

Survey Biomass Estimates and Population Age Composition Estimates

Indices of relative abundance available from AFSC surveys have shown a major increase in the abundance of yellowfin sole during the late 1970s, increasing from 21 kg/ha in 1975 to 51 kg/ha in 1981 (Fig. 4.2 in Bakkala and Wilderbuer 1990). These increases have also been documented through Japanese commercial pair trawl data and catch-at-age modeling in past assessments (Bakkala and Wilderbuer 1990).

Since 1981, the survey CPUEs have fluctuated widely (Fig. 4.5). Biomass estimates for yellowfin sole from the annual bottom trawl survey on the eastern Bering Sea shelf are shown in Table 4.5 and Figure 4.6. The data show a doubling of survey biomass between 1975 and 1979 with a further increase to over 3.3 million t in 1981. Total survey abundance estimates fluctuated erratically from 1983 to 1990 with biomass ranging from as high as 3.5 million t in 1983 to as low as 1.9 million t in 1986. Biomass estimates since 1990 indicate an even trend at high levels of abundance for yellowfin sole, with the exception of the results from the 1999 and 2000 summer surveys, which were at lower levels. Surveys from 2001-2005 estimated an increase each year but the estimates since 2006 indicate a stable level with some annual variability. However, the 2012 estimate is a 19% decrease from 2011 and the 2013 and 2014 surveys have estimated a 17% increase over 2012. Similarly, there was a 24% decrease from 2014 to 2015 followed by a 48% increase from 2015 to 2016, the highest biomass estimate since 1984. Fluctuations of the magnitude shown between 1980 and 1990, 1998 and 1999, 2008 and 2009, 2011 and 2012, 2014 and 2015 and 2015 and 2016 are unreasonable considering the elements of slow growth and long life span of yellowfin sole combined with low to moderate exploitation rate, characteristics which should produce more gradual changes in abundance.

Variability of yellowfin sole survey biomass estimates (Fig. 4.6) is in part due to the availability of yellowfin sole to the survey area (Nichol, 1998). Yellowfin sole are known to undergo annual migrations from wintering areas off the shelf-slope break to near shore waters where they spawn throughout the spring and summer months (Nichol, 1995; Wakabayashi, 1989; Wilderbuer et al., 1992). Exploratory survey sampling in coastal waters of the eastern Bering Sea during early summer indicate that yellowfin sole concentrations can be greater in these shallower areas not covered by the standard AFSC survey than in the survey proper. Commercial bottom trawlers have commonly found high concentrations of yellowfin sole in areas such as near Togiak Bay (Low and Narita, 1990) and in more recent years from

Kuskokwim Bay to just south of Nunivak Island. The coastal areas are sufficiently large enough to offer a substantial refuge for yellowfin sole from the current survey.

Over the past 18 years, survey biomass estimates for yellowfin sole have shown a positive correlation with shelf bottom temperatures (Nichol, 1998); estimates have generally been lower during cold years. The 1999 survey, which was conducted in exceptionally cold waters, indicated a decline in biomass that was unrealistic. The bottom temperatures during the 2000 survey were much warmer than in 1999, and the biomass increased, but still did not approach estimates from earlier years. Average bottom temperature and biomass both increased again during the period 2001 – 2003, with the 2003 value the highest temperature and biomass observed over the 22 year time series up to that time. Given that both the 1999 and 2000 surveys were conducted two weeks earlier than previous surveys, it is possible that the time difference may also have affected the availability of yellowfin sole to the survey. If, for example, the timing of peak yellowfin sole spawning in nearshore waters corresponded to the time of the survey, a greater proportion of the population would be unavailable to the standard survey area. This pattern was observed again in 2009 and 2012 when the temperatures and the bottom trawl survey point estimates were lower. Summer shelf bottom temperatures in 2012 were the 2nd coldest recorded by the survey and the time-series and resulted in a 19% decline from 2011. In 2016 the Bering Sea had the highest recorded bottom temperature since measurements began in 1982 and the 2016 estimate of biomass was the highest in 32 years and 48% higher than the 2015 estimate.

We propose two possible reasons why survey biomass estimates are lower during years when bottom temperatures are low. First, catchability may be lower because yellowfin sole may be less active when cold. Less active fish may be less susceptible to herding, and escapement under the footrope of survey gear may increase if fish are less active. Secondly, bottom temperatures may influence the timing of the inshore spawning migrations of yellowfin sole and therefore affect their availability to the survey area. Because yellowfin sole spawning grounds include nearshore areas outside the survey area, availability of fish within the survey area can vary with the timing of this migration and the timing of the survey. In the case of 2016, a very warm year in the Bering Sea, it appears that a higher portion of the adult biomass was distributed on the shelf (outside of the spawning areas) relative to the average of all previous survey years, indicating earlier spawning migration (Fig 4.7).

Yellowfin sole population numbers-at-age estimated from the annual bottom trawl surveys are shown in Table 4.6 and their occurrence in trawl survey hauls and associated collections of lengths and age structures since 1982 are shown in Table 4.7. Their total tonnage caught in the resource assessment surveys since 1982 are listed in Table 4.8 and also in an appendix table with IPHC survey catches.

Age Determination

Yellowfin sole ages have been determined at the AFSC by using the break and burn method on collected otoliths since 1979 in surveys and from fisheries. In 2016 the age determination methods for yellowfin sole were validated using the bomb-produced uptake measurement of ¹⁴C method (Kastelle et al. 2016).

Length and Weight-at-Age

Past assessments of yellowfin sole have used sex-specific, time-invariant growth based on the average length-at-age and weight-at-length relationships from the time-series of survey observations summed over all years since 1982. These weight-at-age estimates were estimated from the following relationships:

Parameters of the von Bertalanffy growth curve have been estimated for yellowfin sole, by sex, from the trawl survey database as follows:

	L_{inf}	K	t_0	n
Males	33.7	0.161	-0.111	656
Females	37.8	0.137	0.112	709

A sex-specific length-weight relationship was also calculated from the survey database using the usual power function, $\text{weight (g)} = a \text{ Length(cm)}^b$, where a and b are parameters estimated to provide the best fit to the data (Fig. 4.8).

	a	b	n
males	0.00854	3.081	2,701
females	0.0054	3.227	3,662

These estimates of weight at length were applied to the annual trawl survey estimates of population length at age averaged over all years, by sex, to calculate the weight at each age (Fig. 4.8). Since the resulting estimates of weight-at-age were highly variable for fish older than 11 years, ages 11-20 were smoothed using a five year average smoothing method for 1982-2015.

Recent applications of dendrochronology (tree-ring techniques) have been used to develop biochronologies from the otolith growth increments of northern rock sole (*Lepidopsetta polyxystra*), yellowfin sole and Alaska plaice (*Pleuronectes quadrituberculatus*) in the eastern Bering Sea. These techniques ensure that all growth increments are assigned the correct calendar year, allowing for estimation of somatic growth by age and year for chronologies that span approximately 25 years (Matta et al. 2010). The analysis indicated that yellowfin sole somatic growth exhibits annual variability and is positively correlated with May bottom water temperature in the Bering Sea (Fig. 4.9).

The relationship between temperature and growth was further explored by reanalyzing yellowfin sole growth by age and year. Length-weight data collected when obtaining otolith (age) samples in RACE surveys ($n=7,000$ from 1987, 1994 and 1999-2009) also indicate that weight at age exhibits annual variability and is highly correlated with summer bottom water temperature observations with a lag of 2-3 years for the temperature effect to be seen (shown for age 5 fish in figure 4.10). These observations were then extended back to 1979 using survey population length-at-age estimates (since weight-at-age is a power function of the length-at-age, Clark et al. 1999, Walters and Wilderbuer 2000).

In order to incorporate time-varying (year effect on growth) and temperature-dependent growth functions into the age-structured stock assessment model we used the annual observed population mean weight-at-age (time-varying) from the trawl survey. These empirical data indicate good somatic growth correspondence with annual bottom temperature anomalies from 1982-2014 (Fig. 4.11).

Maturity-at-age

Maturity information collected from yellowfin sole females during the 1992 and 1993 eastern Bering Sea trawl surveys have been used in this assessment for the past 20 years (Table 4.10). Nichol (1995) estimated the age of 50% maturity at 10.5 years based on the histological examination of 639 ovaries. Maturity has recently been re-evaluated from a histological analysis of ovaries collected in 2012 (Table 4.10). Results were very similar to the earlier study with only a 2% difference in estimates of yellowfin sole female spawning biomass (TenBrink and Wilderbuer 2015). In addition, the SSC requested that the assessment use a maturity schedule that uses estimates derived from both the 1992 and the 2012 collections (Table 4.10). For yellowfin sole sexual maturity occurs well after the age of entry into the fishery. Yellowfin sole females are 82% selected to the fishery by age 10 whereas they have been found to be only 40% mature at this age.

Analytic Approach

Model Structure

The abundance, mortality, recruitment and selectivity of yellowfin sole were assessed with a stock assessment model using the AD Model Builder language (Fournier et al. 2012; Ianelli and Fournier 1998).

The conceptual model is a separable catch-age analysis that uses survey estimates of biomass and age composition as auxiliary information (Fournier and Archibald 1982). The assessment model simulates the dynamics of the population and compares the expected values of the population characteristics to the characteristics observed from surveys and fishery sampling programs. This is accomplished by the simultaneous estimation of the parameters in the model using the maximum likelihood estimation procedure. The fit of the simulated values to the observable characteristics is optimized by maximizing a log(likelihood) function given some distributional assumptions about the observed data.

The model starts at age one and fish older than twenty are allowed to accumulate into a plus group. Since the sex-specific weight-at-age for yellowfin sole diverges after age of maturity (about age 10 for 40% of the stock) with females growing larger than males, the current assessment model is coded to accommodate the sex-specific aspects of the population dynamics of yellowfin sole. The model allows for the input of sex-specific estimates of fishery and survey age composition and weight-at-age and provides sex-specific estimates of population numbers, fishing mortality, selectivity, fishery and survey age composition and allows for the estimation of sex-specific natural mortality and catchability. The model retains the utility to fit combined sex data inputs.

The suite of parameters estimated by the model are classified by three likelihood components:

Data component	Distributional assumption
Trawl fishery catch-at-age	Multinomial
Trawl survey population age composition	Multinomial
Trawl survey biomass estimates and S.E.	Log normal

The total log likelihood is the sum of the likelihoods for each data component (Table 4.11). The likelihood components may be weighted by an emphasis factor, however, equal emphasis was placed on fitting each likelihood component in the yellowfin sole assessment except for the catch. The AD Model Builder software fits the data components using automatic differentiation (Griewank and Corliss 1991) software developed as a set of libraries (AUTODIFF C++ library). Table 4.11 also presents the key equations used to model the yellowfin sole population dynamics in the Bering Sea and Table 4.12 provides a description of the variables used in Table 4.11.

Sharp increases in trawl survey abundance estimates for most species of Bering Sea flatfish between 1981 and 1982 indicate that the 83-112 trawl was more efficient for capturing these species than the 400-mesh eastern trawl used in 1975, and 1979-81. Allowing the model to tune to these early survey estimates would most likely underestimate the true pre-1982 biomass, thus exaggerating the degree to which biomass increased during that period. Although this underestimate would have little effect on the estimate of current yellowfin sole biomass, it would affect the spawner and recruitment estimates for the time-series. Hence, the pre-1982 survey biomass estimates were omitted from the analysis.

The model of yellowfin sole population dynamics was evaluated with respect to the observations of the time-series of survey and fishery age compositions and the survey biomass trend since 1982.

Parameters Estimated Outside the Assessment Model

Natural mortality (M) was initially estimated by a least squares analysis where catch-at-age data were fitted to Japanese pair trawl effort data while varying the catchability coefficient (q) and M simultaneously. The best fit to the data (the point where the residual variance was minimized) occurred at a M value of 0.12 (Bakkala and Weststad 1984). This was also the value which provided the best fit to the observable population characteristics when M was profiled over a range of values in the stock assessment model using data up to 1992 (Wilderbuer 1992). Since then, natural mortality has been estimated as a free parameter in some of the stock assessment model runs which have been evaluated for the past five years. A natural mortality value of 0.12 is used for both sexes in the base model presented in this assessment.

Yellowfin sole maturity schedules were estimated from in-situ observations from two studies as discussed in a previous section (Table 4.10).

Parameters Estimated Inside the Assessment Model

The parameters estimated by the model are presented below:

Fishing mortality	Selectivity	Survey catchability	Year class strength	Spawner-recruit	Total
64	260	2	103	2	431

The increase in the number of parameters estimated in this assessment compared to last year (6) can be accounted for by the input of another year of fishery data and the entry of another year class into the observed population and four more sex-specific fishery selectivity parameters.

Year class strengths

The population simulation specifies the numbers-at-age in the beginning year of the simulation, the number of recruits in each subsequent year, and the survival rate for each cohort as it moves through the population over time using the population dynamics equations given in Table 4.11.

Selectivity

Fishery and survey selectivity was modeled separately for males and females using the two parameter formulation of the logistic function (Table 4.11). The model was run with an asymptotic selectivity curve for the older fish in the fishery and survey, but still was allowed to estimate the shape of the logistic curve for young fish. The oldest year classes in the surveys and fisheries were truncated at 20 and allowed to accumulate into the age category 20+ years. A single selectivity curve, for both males and females, was fit for all years of survey data.

Given that there have been annual changes in management, vessel participation and most likely gear selectivity, time-varying fishing selectivity curves were estimated. A logistic equation was used to model fishery selectivity and is a function of time-varying parameters specifying the age and slope at 50% selection, ϕ_t and η_t , respectively. The fishing selectivity (S^f) for age a and year t is modeled as,

$$S_{a,t}^f = \left[1 + e^{\eta_t(a-\phi_t)} \right]^{-1}$$

where η_t and ϕ_t are time-varying and partitioned (for estimation) into parameters representing the mean and a vector of deviations (log-scale) conditioned to sum to zero. The deviations are constrained by a lognormal prior with a variance that was iteratively estimated. The process of iterating was to first set the variance to a high value (diffuse prior) of 0.5² and estimate the deviations. The next step was to compare the variability of model estimates. The variance of the model estimates were then rounded up slightly and fixed for subsequent runs. The 2016 values were fixed as the average of the 3 most recent years.

Fishing Mortality

The fishing mortality rates (F) for each age and year are calculated to approximate the catch weight by solving for F while still allowing for observation error in catch measurement. A large emphasis (300) was placed on the catch likelihood component to force the model to closely match the observed catch.

Survey Catchability

A past assessment (Wilderbuer and Nichol 2001) first examined the relationship between estimates of survey biomass and bottom water temperature. To better understand how water temperature may affect

the catchability of yellowfin sole to the survey trawl, catchability was estimated for each year in the stock assessment model as:

$$q = e^{-\alpha + \beta T}$$

where q is catchability, T is the average annual bottom water temperature anomaly at survey stations less than 100 m, and α and β are parameters estimated by the model. The catchability equation has two parts. The $e^{-\alpha}$ term is a constant or time-independent estimate of q . The second term, $e^{\beta T}$ is a time-varying (annual) q which responds to metabolic aspects of herding or distribution (availability) which can vary annually with bottom water temperature. The result of incorporating bottom temperature to estimate annual q is shown in Figure 4.12 (for the base model).

Spawner-Recruit Estimation

Annual recruitment estimates from 1978-2010 were constrained to fit a Ricker (1958) form of the stock recruitment relationship as follows:

$$R = \alpha S e^{-\beta S}$$

where R is age 1 recruitment, S is female spawning biomass (t) the previous year, and α and β are parameters estimated by the model. The spawner-recruit fitting is estimated in a later phase after initial estimates of survival, numbers-at-age and selectivity are obtained.

Results

Model Evaluation

The model evaluation for this stock assessment involved a two-step process. The first step was to evaluate the productivity of the yellowfin sole stock by an examination of which sets of years to include for spawner-recruit fitting. The second step evaluated various hypothesized states of nature by fitting natural mortality and catchability estimates in various combinations.

The SSC determined in December 2006 that yellowfin sole would be managed under the Tier 1 harvest guidelines, and therefore future harvest recommendations would be based on MSY and F_{MSY} values calculated from a spawner-recruit relationship. MSY is an equilibrium concept and its value is dependent on both the spawner-recruit estimates which are assumed to represent the equilibrium stock size-recruitment relationship and the model used to fit the estimates. In the yellowfin sole stock assessment model, a Ricker form of the stock-recruit relationship was fit to various combinations of these data and estimates of F_{MSY} and B_{MSY} were calculated, assuming that the fit to the stock-recruitment data represents the long-term productivity of the stock.

For this assessment, 2 different stock-recruitment time-series were investigated: the full time-series 1955-2008 (Model A) and the post-regime shift era, 1978-2008 (Model B) (Fig. 4.13) (see Joint Plan Team recommendations for September 2012). Very different estimates of the long-term sustainability of the stock (F_{MSY} and B_{MSY}) are obtained, depending on which years of stock-recruitment data are included in the fitting procedure (Table 4.13). When the entire time-series from 1955-2008 was fit, the large recruitments that occurred at low spawning stock sizes in the 1960s and early 1970s determined that the yellowfin sole stock was most productive at a smaller stock size with the result that F_{MSY} (0.163) is higher than $F_{35\%}$ ($F_{35\%} = 0.135$) and B_{MSY} is 336,000 (Model A). If we limit the analysis to consider only recruitments which occurred after the well-documented regime shift in 1977, a lower value of F_{MSY} is obtained (0.125) and B_{MSY} is 424,000 t. Table 4.13 indicates that the ABC values from the Model A harvest scenario for 2017 would be 105,300 t higher than Model B. Posterior distributions of F_{MSY} for these models indicate that this parameter is estimated with less uncertainty for Model A resulting in the

reduced buffer between ABC and OFL relative to Model B (9% for Model B versus < 1% for Model A, Table 4.13 and Fig 4.14).

It is important for the Tier 1 calculations to identify which subset of the stock recruitment data is used. Using the full time series to fit the spawner recruit curve estimates that the stock is most productive at a small stock size. Thus MSY and F_{MSY} are relatively high values and B_{MSY} is a lower value. If the stock was productive in the past at a small stock size because of non-density dependent factors (environment), then reducing the stock size to low levels could be detrimental to the long-term sustainability of the stock if the environment, and thus productivity, have changed from the earlier period. Since observations of yellowfin sole recruitment at low stock sizes are not available from multiple time periods, it is uncertain if future recruitment events at low stock conditions would be as productive as during the late 1960s-early 1970s.

Given the uncertainty of the productivity of yellowfin sole at low spawning stock sizes, and because the AFSC policy for reference point time-series selection is to use the post 1977 regime shift values unless there is a compelling reason to do otherwise, the productivity of yellowfin sole in this assessment is estimated by fitting the 1977-2008 spawner-recruit data in the model (Model B).

The second step in the model evaluation for this assessment entails the use of a single structural model to consider the uncertainty in the key parameters M and catchability. This is the Model which has been the model of choice is the past 7 assessments (Model 1) and operates by fixing M at 0.12 for both sexes and then estimates q using the relationship between survey catchability and the annual average water temperature at the sea floor (from survey stations at less than 100 m). The other models used in the evaluation represented various combinations of estimating M or q as free parameters with different amounts of uncertainty in the parameter estimates (Wilderbuer et al. 2010). The results are detailed in those assessments and are not repeated here except for the following observations.

Modeling survey catchability as a nonlinear function of bottom water temperature returns q estimates > 1.0 for years when the bottom temperature is anomalously warm (greater than the mean temperature) and less than 1.0 when below the temperature mean. These values are consistent with our hypothesis that more fish are available to the survey in warm years relative to cooler years due to the timing of the annual spawning migration to nearshore areas that occurs sooner in warm years.

Experiments examining the bridle efficiency of the Bering Sea survey trawl indicate that yellowfin sole are herded into the trawl path from an area between the wing tips of the net and the point where the bridles contact the seafloor (Somerton and Munro 2001). The herding experiments suggest that the survey trawl catchability is greater than 1.0. The likelihood profile of q from the model indicated a small variance with a narrow range of likely values with a low probability of q being equal to the value of 1.0 in a past assessment (Wilderbuer and Nichol 2003).

A model that allows M to be estimated as a free parameter for males with females fixed at 0.12 provided a better fit to the sex ratio estimated from the annual trawl survey age compositions than did the base model (both sexes fixed at $M = 0.12$). However, since the population sex ratio annually observed at the time of the survey is a function of the timing of the annual spawning in adjacent inshore areas, it is questionable that providing the best fit to these observations is really fitting the population sex ratio better. Thus, the model configuration which utilizes the relationship between annual seafloor temperature and survey catchability with M fixed at 0.12 for both sexes (Model 1) is the preferred model used to base the assessment of the condition of the Bering Sea yellowfin sole resource for the 2016 fishing season.

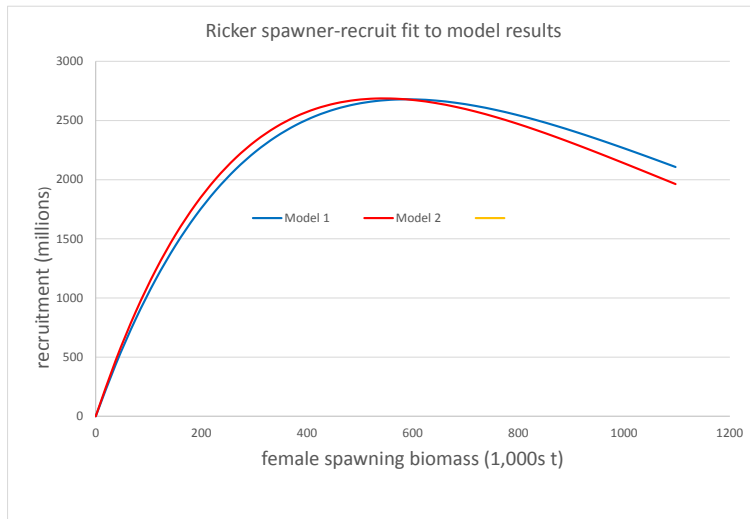
Time Series Results

Before presenting the preferred model results, a brief consideration of the inputs and changes to the assessment methodology relative to last year (Model 0) is given. Primary updates for Model 1 were the catch, the fishery and survey age compositions from 2015 and the 2016 survey biomass estimate. In their totality, these changes produced a Model 1 ABC estimate (using the same assessment model as last year)

that was 7% higher than Model 0, F_{ABC} that was <1% lower and FSB that was 13% larger. These increases were primarily due to the 48% increase in survey biomass from 2015 to 2016. In order to understand the effect of the new data components on the 2016 results (fishery weight at age data from 2008-2016), a piece-wise example of the model results is presented. Model 1 updated the previous assessment model as described above, Model 2 added the revised fishery weights at age from over 4,000 aged fish available from the fishery sampling.

	Model 0	Model 1	Model 2
F_{abc}	0.10	0.10	0.11
F_{OFL}	0.11	0.11	0.13
6+ biomass	2,169,70	2,333,450	2,290,140
ABC	211,694	226,278	260,826
OFL	228,053	248,423	287,051
FSB	702,179	791,479	778,569

Applying the average of the 2008-2014 fishery weight at age to the 2008-2016 fishery weight at age values in Model 2 caused the ABC and OFL to increase by 15% relative to Model 1 because of higher F_{ABC} and F_{OFL} values. This increase was due to refitting the spawner-recruit curve with the averaged fishery weight-at-age data which had the effect of moving the curve and B_{MSY} to the left (lower B_{msy} , higher F_{msy} for Tier 1 stocks).



The 2016 trawl survey point estimate increased 48% from 2015. This resulted in higher model estimates of population numbers at age and biomass for the time-series back to the mid-1960s relative to last year's assessment. In addition, the large 2003 year class (13 years old in 2016) is present in the population, but now past their cohort maximum. The model results indicate the stock has been in a slowly declining condition since the mid-1980s. The estimates of total biomass and ABC are higher than those used to manage the stock in 2016. Seven of the past 11 years have had negative bottom temperature anomalies in the Bering Sea but the last three have been above the mean. 2016 was the warmest year relative to temperatures collected since 1982 with an estimated value 2.5 deg. C above the long term mean. The temperature-dependent q adjustment for 2016 was 1.2.

Fishing Mortality and Selectivity

The assessment model estimates of the annual fishing mortality in terms of age-specific annual F and on fully selected ages are given in Tables 4.14 and 4.15, respectively. The full-selection F has averaged 0.08 over the period of 1978-2015 with a maximum of 0.12 in 1978 and a minimum of 0.04 in 2001.

Selectivities estimated by the model (Table 4.16, Fig. 4.15) indicate that both sexes of yellowfin sole are 50% selected by the fishery at about age 9 and nearly fully selected by age 13, with annual variability.

Abundance Trend

The model estimates q at an average value of 0.95 for the period 1982-2016 which results in the model estimate of the 2016 age 2+ total biomass at 2,457,000 t (Table 4.17). Model results indicate that yellowfin sole total biomass (age 2+) was at low levels during most of the 1960s and early 1970s (700,000-1,000,000 t) after a period of high exploitation (Table 4.17, Fig. 4.16, center left panel). Sustained above average recruitment from 1967-76 combined with light exploitation resulted in a biomass increase to a peak of 3.4 million t by 1985. The population biomass has since been in a slow decline as the strong 1981 and 1983 year-classes have passed through the population, with only the 1991, 1995 and 2003 year-classes at levels observed during the 1970s. The present biomass is estimated at 72% of the peak 1985 level.

The female spawning biomass has also declined since the peak in 1993, with a 2016 estimate of 775,100 t (28% decline). The spawning biomass has been in a gradual decline for the past 22 years and is 20% above the $B_{40\%}$ level and 1.8 times the B_{MSY} level (Fig. 4.16). The model estimate of yellowfin sole population numbers at age for all years is shown in Table 4.18 and the resulting fit to the observed fishery and survey age compositions input into the model are shown in the Figure 4.17. The fit to the trawl survey biomass estimates are shown in Figure 4.16. Allowing q to be correlated with annual bottom temperature provides a better fit to the bottom trawl survey estimates than using a q fixed at the average value (Fig. 4.18). Table 4.19 lists the numbers of female spawners estimated by the model for all ages and years. The estimated average age of yellowfin sole in the population is 6.7 years for males and females.

Both the trawl survey and the stock assessment model indicate that the yellowfin sole resource increased during the 1970s and early 1980s to a peak level during the mid-1980s. The yellowfin sole population biomass slowly decreased over the 22 years since the mid-1990s as the majority of year-classes during those years were below average strength. Above-average recruitment from the strong 2003 year-class is expected to maintain the abundance of yellowfin sole at a level above B_{MSY} in the near future and there also have been average to above average recruitment from 2006 to 2009. The stock assessment projection model indicates a mildly decreasing trend in female spawning biomass through 2023 if the fishing mortality rate continues at the same level as the average of the past 5 years (Fig. 4.22).

Recruitment Trends

The primary reason for the sustained increase in abundance of yellowfin sole during the 1970s and early 1980s was the recruitment of a series of stronger than average year classes spawned in 1967-76 (Figure 4.19 and Table 4.20). The 1981 year class was the strongest observed (and estimated) during the 47 year period analyzed and the 1983 year class was also very strong. Survey age composition estimates and the assessment model also estimate that the 1987 and 1988 year classes were average and the 1991 and 1995 year classes were above average. With the exception of these 4 year classes, recruitment from 15 of the following 19 years estimated from 1984-2005 (since the strong 1983 year-class) were below the 48 year average, which caused the population to gradually decline. The 2003 year-class has now been observed multiple times in the age compositions and is clearly a strong year class, similar to some of the strong recruitment mentioned above and are contributing to the reservoir of spawning fish in the current population. In addition, recruitment from 2006-2009 may also be average to above average but, at present, there is uncertainty due to a lack of repeated observations.

Historical Exploitation Rates

Based on results from the stock assessment model, annual average exploitation rates of yellowfin sole since 1977 ranged from 3 to 8% of the total biomass, and have averaged 5% (Table 4.15). Posterior distributions of selected parameters from the preferred stock assessment model used in the assessment are shown in Figure 4.20. The values and standard deviations of some selected model parameters are listed in Table 4.21.

Retrospective analysis

A within-model retrospective analysis is also included for the recommended assessment model where retrospective female spawning biomass is calculated by working backwards in time dropping data one year at a time and then comparing the “peeled” estimate to the reference stock assessment model used in the assessment (Fig. 4.21). The resulting pattern from the current assessment model was less than desirable.

Peculiar to the yellowfin sole assessment, in comparison to the northern rock sole and Alaska plaice assessments (that have nice patterns), is the large amount of variability in the annual survey biomass assessments for this stock due to the temperature-influenced availability to the survey. This large variability in the annual estimates can contribute to undesirable patterns since the earlier years are not fitting the same highly variable information as the current year. Exploratory model runs were made to examine the influence of fitting the survey biomass on the retrospective patterns. This was accomplished by making model runs that increased the survey biomass standard error by 10%, 20% and 30%, and also runs that attempted to decrease the influence of fitting the survey age composition by decreasing the effective n , and also a run with a fixed q (no bottom temp modeling). The models are listed below and are evaluated using Mohn’s test statistic.

Model Description	Mohn’s test statistic
Current stock assessment model (Model 1)	-0.193
Model 1 with survey standard error increased by 10%	-0.239
Model 1 with survey standard error increased by 20%	-0.211
Model 1 with survey standard error increased by 30%	-0.219
Down-weighted survey age comps, base st. dev values	-0.207
Up-weighted survey age comps (500) and down-weighted survey SE (increased st. dev value by 30%)	-0.238
survey q fixed at 1.05 for all runs	-0.186

A small gain was made by fixing q whereby all the runs became more similar, other runs increased the test statistic over the base value (lower is better). This is a work in progress to examine these patterns.

Harvest Recommendations

Since the peak value in 1984, estimates from the stock assessment model indicate the total biomass has slowly declined. The estimate of age 6+ total biomass for 2016 is 2,170,000 t.

The SSC has determined that yellowfin sole qualify as a Tier 1 stock and therefore the 2016 ABC is calculated using Tier 1 methodology. In 2006 the SSC selected the 1978-2001 data set for the Tier 1 harvest recommendation. Using this approach again for the 2017 harvest (now the 1978-2010 time-series) recommendation (Model B in Table 4.13), the $F_{ABC} = F_{\text{harmonic mean}} = 0.11$.

The Tier 1 harvest level is calculated as the product of the harmonic mean of F_{MSY} and the geometric mean of the 2016 biomass estimate, as follows:

$B_{gm} = e^{\ln \hat{B} - \frac{cv^2}{2}}$, where B_{gm} is the geometric mean of the 2017 biomass estimate, \hat{B} is the point estimate of the 2017 biomass from the stock assessment model and cv^2 is the coefficient of variation of the point estimate (a proxy for sigma);

and

$\bar{F}_{har} = e^{\ln \hat{F}_{msy} - \frac{\ln sd^2}{2}}$, where \bar{F}_{har} is the harmonic mean, \hat{F}_{msy} is the peak mode of the F_{MSY} distribution and sd^2 is the square of the standard deviation of the F_{MSY} distribution. This calculation gives a Tier 1 ABC harvest recommendation of **260,800 t** and an OFL of 287,000 t for 2017. This results in a 9% (26,230 t) buffer between ABC and OFL. The ABC value is 23% higher than last year, primarily due to an increasing survey estimate and changes to the spawner-recruit curve from the fishery weight-at-age modeling. The stock assessment analysis must also consider harvest limits, usually described as overfishing fishing mortality levels with corresponding yield amounts. Amendment 56 to the BSAI FMP sets the Tier 1 harvest limit at the F_{MSY} fishing mortality value. The overfishing fishing mortality values, ABC fishing mortality values and their corresponding yields are given as follows:

<u>Harvest level</u>	<u>F value</u>	<u>2017 Yield</u>
Tier 1 $F_{OFL} = F_{MSY}$	0.125	287,000 t
Tier 1 $F_{ABC} = F_{\text{harmonic mean}}$	0.11	260,800 t

Status Determination

A standard set of projections is required for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with the vector of 2016 numbers at age estimated in the assessment. This vector is then projected forward to the beginning of 2017 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end) catch for 2016. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest

alternatives that are likely to bracket the final TAC for 2017, are as follows (“ $max F_{ABC}$ ” refers to the maximum permissible value of F_{ABC} under Amendment 56):

Scenario 1: In all future years, F is set equal to $max F_{ABC}$. (Rationale: Historically, TAC has been constrained by ABC, so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, F is set equal to a constant fraction of $max F_{ABC}$, where this fraction is equal to the ratio of the F_{ABC} value for 2017 recommended in the assessment to the $max F_{ABC}$ for 2017. (Rationale: When F_{ABC} is set at a value below $max F_{ABC}$, it is often set at the value recommended in the stock assessment.)

Scenario 3: In all future years, F is set equal to 50% of $max F_{ABC}$. (Rationale: This scenario provides a likely lower bound on F_{ABC} that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 4: In all future years, F is set equal to the 2012-2016 average F . (Rationale: For some stocks, TAC can be well below ABC, and recent average F may provide a better indicator of F_{TAC} than F_{ABC} .)

Scenario 5: In all future years, F is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

Two other scenarios are needed to satisfy the MSFCMA’s requirement to determine whether a stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follow (for Tier 3 stocks, the MSY level is defined as $B_{35\%}$):

Scenario 6: In all future years, F is set equal to F_{OFL} . (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be above its MSY level in 2015 and above its MSY level in 2028 under this scenario, then the stock is not overfished.)

Scenario 7: In 2017 and 2018, F is set equal to $max F_{ABC}$, and in all subsequent years, F is set equal to F_{OFL} . (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2029 under this scenario, then the stock is not approaching an overfished condition.)

Simulation results shown in Table 4.22 indicate that yellowfin sole are not currently overfished and are not approaching an overfished condition. The projection of yellowfin sole female spawning biomass through 2029 is shown in Figure 4.22 and a phase plane figure of the estimated time-series of yellowfin sole female spawning biomass relative to the harvest control rule is shown in Figure 4.23.

Scenario Projections and Two-Year Ahead Overfishing Level

In addition to the seven standard harvest scenarios, Amendments 48/48 to the BSAI and GOA Groundfish Fishery Management Plans require projections of the likely OFL two years into the future. The 2016 numbers at age from the stock assessment model are projected to 2017 given the 2016 catch and then a 2017 catch of 150,000 t is applied to the projected 2017 population biomass to obtain the 2018 OFL.

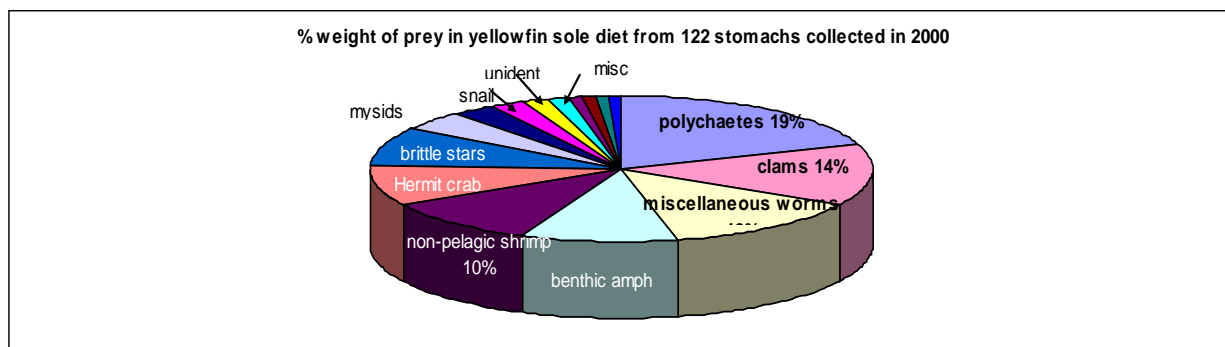
Tier 1 Projection		SSB	Geometric mean 6+ total biomass	ABC	OFL
Year	Catch				
2017	150,000	778,600	2,290,000	260,800	287,000
2018	150,000	770,900	2,202,300	250,800	276,000

Ecosystem Considerations

Ecosystem Effects on the stock

1) Prey availability/abundance trends

Yellowfin sole diet by life stage varies as follows: Larvae consume plankton and algae, early juveniles consume zooplankton, late juvenile stage and adults prey includes bivalves, polychaetes, amphipods, mollusks, euphausiids, shrimps, brittle stars, sculpins and miscellaneous crustaceans. Information is not available to assess the abundance trends of the benthic infauna of the Bering Sea shelf. The original description of infaunal distribution and abundance by Haflinger (1981) resulted from sampling conducted in 1975 and 1976 and has not been re-sampled since. The large populations of flatfish which have occupied the middle shelf of the Bering Sea over the past twenty-five years for summertime feeding do not appear food-limited. These populations have fluctuated due to the variability in recruitment success which suggests that the primary infaunal food source has been at an adequate level to sustain the yellowfin sole resource.



2) Predator population trends

As juveniles, it is well-documented from studies in other parts of the world that flatfish are prey for shrimp species in near shore areas. This has not been reported for Bering Sea yellowfin sole due to a lack of juvenile sampling and collections in near shore areas, but is thought to occur. As late juveniles they have been found in stomachs of Pacific cod and Pacific halibut; mostly small yellowfin sole ranging from 7 to 25 cm standard length..

Past, present and projected future population trends of these predator species can be found in their respective SAFE chapters in this volume and also from Annual reports compiled by the International Pacific Halibut Commission. Encounters between yellowfin sole and their predators may be limited since their distributions do not completely overlap in space and time.

3) Changes in habitat quality

Changes in the physical environment which may affect yellowfin sole distribution patterns, recruitment success and migration timing patterns are catalogued in the Ecosystem Considerations Report of this SAFE report. Habitat quality may be enhanced during years of favorable cross-shelf advection (juvenile survival) and warmer bottom water temperatures with reduced ice cover (higher metabolism with more active feeding).

Fishery Effects on the ecosystem

- 1) The yellowfin sole target fishery contribution to the total bycatch of other target species is shown for 1992-2015 in Table 4.23. The catch of non-target species from 2003-2015 is shown in Table 4.24. The yellowfin sole target fishery contribution to the total bycatch of prohibited species is

shown for 2013 and 2014 in Table 13 of the Economic SAFE (Appendix C) and is summarized for 2014 as follows:

Prohibited species	Yellowfin sole fishery % of total bycatch
Halibut mortality	38.6
Herring	13.2
Red King crab	3
<u>C. bairdi</u>	31.5
Other Tanner crab	54.8
Salmon	<1

- 2) Relative to the predator needs in space and time, the yellowfin sole target fishery has a low selectivity for fish 7-25 cm and therefore has minimal overlap with removals from predation.
- 3) The target fishery is not perceived to have an effect on the amount of large size target fish in the population due to its history of light to moderate exploitation (6%) over the past 30 years. Population age composition data indicate a large 20+ age group.
- 4) Yellowfin sole fishery discards are presented in the Catch History section.
- 5) It is unknown what effect the fishery has had on yellowfin sole maturity-at-age and fecundity.
- 6) Analysis of the benthic disturbance from the yellowfin sole fishery is available in the Preliminary draft of the Essential Fish Habitat Environmental Impact Statement.

Ecosystem effects on yellowfin sole			
Indicator	Observation	Interpretation	Evaluation
<i>Prey availability or abundance trends</i>			
Benthic infauna	Stomach contents	Stable, data limited	Unknown
<i>Predator population trends</i>			
Fish (Pacific cod, halibut, skates)	Stable	Possible increases to yellowfin sole mortality	
<i>Changes in habitat quality</i>			
Temperature regime	Cold years yellowfin sole catchability and herding may decrease, timing of migration may be prolonged	Likely to affect surveyed stock	No concern (dealt with in model)
Winter-spring environmental conditions	Affects pre-recruit survival	Probably a number of factors	Causes natural variability
Yellowfin sole effects on ecosystem			
Indicator	Observation	Interpretation	Evaluation
<i>Fishery contribution to bycatch</i>			
Prohibited species	Stable, heavily monitored	Minor contribution to mortality	No concern
Forage (including herring, Atka mackerel, cod, and pollock)	Stable, heavily monitored	Bycatch levels small relative to forage biomass	No concern
HAPC biota	Low bycatch levels of (spp)	Bycatch levels small relative to HAPC biota	No concern
Marine mammals and birds	Very minor direct-take	Safe	No concern
Sensitive non-target species	Likely minor impact	Data limited, likely to be safe	No concern
<i>Fishery concentration in space and time</i>	Low exploitation rate	Little detrimental effect	No concern
<i>Fishery effects on amount of large size target fish</i>	Low exploitation rate	Natural fluctuation	No concern
<i>Fishery contribution to discards and offal production</i>	Stable trend	Improving, but data limited	Possible concern
<i>Fishery effects on age-at-maturity and fecundity</i>	Unknown	NA	Possible concern

Data Gaps and Research Priorities

Isolation by distance genetic study to define stock structure in the planning stage. NPRB proposal to collect maturity in the northern Bering Sea for comparison with recent SE Bering Sea shelf samples.

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Tables

Table 4.1--Catch (t) of yellowfin sole 1964-2016. Catch for 2016 is an estimate through the end of 2016.

Year	Foreign	Domestic		Total
		JVP	DAP	
1964	111,777			111,777
1965	53,810			53,810
1966	102,353			102,353
1967	162,228			162,228
1968	84,189			84,189
1969	167,134			167,134
1970	133,079			133,079
1971	160,399			160,399
1972	47,856			47,856
1973	78,240			78,240
1974	42,235			42,235
1975	64,690			64,690
1976	56,221			56,221
1977	58,373			58,373
1978	138,433			138,433
1979	99,019			99,019
1980	77,768	9,623		87,391
1981	81,255	16,046		97,301
1982	78,331	17,381		95,712
1983	85,874	22,511		108,385
1984	126,762	32,764		159,526
1985	100,706	126,401		227,107
1986	57,197	151,400		208,597
1987	1,811	179,613	4	181,428
1988		213,323	9,833	223,156
1989		151,501	1,664	153,165
1990		69,677	14,293	83,970
1991			115,842	115,842
1992			149,569	149,569
1993			106,101	106,101
1994			144,544	144,544
1995			124,740	124,740
1996			129,659	129,659
1997			181,389	181,389
1998			101,201	101,201
1999			67,320	67,320
2000			83,850	83,850
2001			63,395	63,395
2002			73,000	73,000
2003			74,418	74,418
2004			69,046	69,046
2005			94,383	94,383
2006			99,068	99,068
2007			121,029	121,029
2008			148,894	148,894
2009			107,528	107,528
2010			118,624	118,624
2011			151,164	151,164
2012			147,183	147,183
2013			164,944	164,944
2014			156,778	156,778
2015			126,933	126,933
2016			130,500	130,500

Table 4.2 Estimates of retained and discarded (t) yellowfin sole caught in Bering Sea fisheries.

Year	Retained	Discarded
1987	3	1
1988	7,559	2,274
1989	1,279	385
1990	10,093	4,200
1991	89,054	26,788
1992	103,989	45,580
1993	76,798	26,838
1994	107,629	36,948
1995	96,718	28,022
1996	101,324	28,334
1997	149,570	31,818
1998	80,365	20,836
1999	55,202	12,118
2000	69,788	14,062
2001	54,759	8,635
2002	62,050	10,950
2003	63,732	10,686
2004	57,378	11,668
2005	85,321	9,062
2006	90,570	8,498
2007	109,084	11,945
2008	141,253	7,659
2009	92,488	5,733
2010	113,244	5,380
2011	146,419	4,745
2012	143,737	3,446
2013	158,781	6,163
2014	152,164	4,614
2015	123,065	3,871

Table 4.3. Discarded and retained catch of non-CDQ yellowfin sole, by fishery, in 2015.
Source: AKFIN.

Trip Target Name	Discarded	Retained
Atka Mackerel	<1	
Pollock - bottom	5	360
Pacific Cod	2,134	568
Alaska Plaice - BSAI	5	563
Other Flatfish - BSAI		
Halibut		
Rockfish	<1	<1
Flathead Sole	34	1,951
Kamchatka Flounder - BSAI	<1	<1
Pollock - midwater	97	401
Rock Sole - BSAI	207	12,654
Sablefish		
Greenland Turbot - BSAI	0	<1
Arrowtooth Flounder	<1	<1
Yellowfin Sole - BSAI	1,386	106,568

Table 4.4. Yellowfin sole fishery catch-at-age (proportions), 1975-2015.

	females										
	7	8	9	10	11	12	13	14	15	16	17+
1975	0.05	0.14	0.09	0.05	0.02	0.02	0.02	0.00	0.01	0.00	0.00
1976	0.04	0.07	0.17	0.10	0.07	0.01	0.01	0.00	0.01	0.00	0.00
1977	0.07	0.16	0.11	0.02	0.01	0.01	0.00	0.00	0.00	0.00	0.00
1978	0.05	0.13	0.12	0.09	0.09	0.03	0.01	0.01	0.00	0.00	0.00
1979	0.03	0.07	0.12	0.12	0.08	0.06	0.03	0.01	0.01	0.00	0.00
1980	0.06	0.04	0.06	0.11	0.10	0.07	0.07	0.04	0.02	0.01	0.03
1981	0.06	0.06	0.03	0.05	0.09	0.11	0.07	0.05	0.02	0.01	0.01
1982	0.03	0.07	0.06	0.05	0.09	0.09	0.05	0.03	0.02	0.01	0.00
1983	0.07	0.05	0.08	0.04	0.05	0.07	0.06	0.05	0.03	0.02	0.01
1984	0.03	0.04	0.05	0.09	0.04	0.06	0.05	0.06	0.03	0.01	0.02
1985	0.02	0.02	0.06	0.05	0.07	0.06	0.05	0.06	0.05	0.01	0.01
1986	0.03	0.03	0.04	0.09	0.07	0.06	0.04	0.03	0.04	0.03	0.06
1987	0.01	0.03	0.02	0.04	0.05	0.05	0.06	0.05	0.03	0.03	0.09
1988	0.02	0.03	0.07	0.02	0.04	0.05	0.04	0.06	0.05	0.02	0.12
1989	0.00	0.04	0.05	0.05	0.03	0.03	0.05	0.02	0.05	0.05	0.15
1990	0.02	0.01	0.13	0.03	0.06	0.03	0.02	0.01	0.01	0.08	0.09
1991	0.01	0.07	0.01	0.11	0.04	0.04	0.01	0.03	0.04	0.02	0.09
1992	0.01	0.02	0.09	0.02	0.14	0.04	0.04	0.01	0.03	0.02	0.12
1993	0.02	0.01	0.02	0.09	0.01	0.12	0.03	0.03	0.02	0.03	0.18
1994	0.02	0.03	0.03	0.03	0.16	0.00	0.10	0.01	0.04	0.02	0.13
1995	0.04	0.06	0.02	0.01	0.02	0.10	0.00	0.16	0.01	0.03	0.12
1996	0.01	0.04	0.06	0.02	0.03	0.03	0.07	0.01	0.11	0.01	0.11
1997	0.02	0.02	0.06	0.03	0.02	0.03	0.03	0.10	0.01	0.06	0.10
1998	0.02	0.03	0.08	0.04	0.02	0.04	0.04	0.12	0.01	0.07	0.13
1999	0.01	0.02	0.03	0.02	0.08	0.05	0.03	0.04	0.04	0.07	0.23
2000	0.00	0.01	0.05	0.03	0.03	0.07	0.08	0.05	0.02	0.04	0.22
2001	0.01	0.02	0.05	0.08	0.05	0.04	0.07	0.05	0.04	0.02	0.16
2002	0.01	0.02	0.03	0.04	0.06	0.04	0.03	0.04	0.05	0.02	0.21
2003	0.00	0.05	0.04	0.03	0.04	0.08	0.03	0.02	0.02	0.03	0.19
2004	0.01	0.01	0.10	0.05	0.03	0.02	0.05	0.02	0.01	0.05	0.19
2005	0.02	0.03	0.03	0.08	0.03	0.03	0.04	0.06	0.03	0.01	0.19
2006	0.07	0.05	0.03	0.04	0.14	0.02	0.00	0.01	0.03	0.01	0.10
2007	0.01	0.04	0.03	0.02	0.04	0.08	0.03	0.03	0.03	0.03	0.17
2008	0.02	0.04	0.04	0.05	0.03	0.03	0.07	0.04	0.02	0.03	0.17
2009	0.02	0.03	0.06	0.06	0.03	0.05	0.04	0.05	0.03	0.03	0.17
2010	0.04	0.03	0.04	0.04	0.04	0.03	0.04	0.04	0.06	0.03	0.17
2011	0.02	0.05	0.04	0.05	0.04	0.06	0.03	0.03	0.03	0.04	0.16
2012	0.02	0.03	0.05	0.04	0.05	0.02	0.05	0.02	0.01	0.02	0.16
2013	0.01	0.02	0.04	0.07	0.06	0.07	0.05	0.04	0.02	0.03	0.14
2014	0.01	0.02	0.04	0.05	0.05	0.06	0.04	0.03	0.04	0.03	0.18
2015	0.01	0.02	0.02	0.04	0.03	0.08	0.04	0.04	0.03	0.04	0.18

Table 4.4- continued.

	males										
	7	8	9	10	11	12	13	14	15	16	17+
1975	0.09	0.24	0.12	0.02	0.01	0.01	0.02	0.00	0.00	0.00	0.00
1976	0.05	0.04	0.14	0.11	0.04	0.02	0.02	0.00	0.01	0.00	0.00
1977	0.03	0.08	0.07	0.12	0.09	0.04	0.01	0.00	0.00	0.00	0.00
1978	0.05	0.09	0.07	0.07	0.07	0.03	0.02	0.01	0.00	0.00	0.00
1979	0.03	0.06	0.11	0.07	0.06	0.04	0.02	0.00	0.00	0.00	0.00
1980	0.04	0.02	0.03	0.05	0.05	0.05	0.03	0.01	0.01	0.01	0.01
1981	0.04	0.04	0.03	0.05	0.07	0.06	0.04	0.02	0.01	0.00	0.00
1982	0.04	0.06	0.05	0.05	0.07	0.07	0.04	0.02	0.01	0.00	0.00
1983	0.06	0.03	0.06	0.05	0.04	0.06	0.05	0.03	0.02	0.01	0.01
1984	0.01	0.06	0.04	0.08	0.04	0.05	0.05	0.10	0.04	0.02	0.01
1985	0.02	0.03	0.06	0.06	0.07	0.06	0.07	0.06	0.03	0.02	0.02
1986	0.03	0.02	0.05	0.06	0.04	0.04	0.04	0.02	0.04	0.04	0.05
1987	0.02	0.05	0.04	0.05	0.04	0.05	0.06	0.04	0.02	0.03	0.13
1988	0.03	0.04	0.09	0.02	0.04	0.04	0.02	0.05	0.04	0.01	0.08
1989	0.00	0.04	0.04	0.06	0.02	0.02	0.04	0.03	0.03	0.03	0.16
1990	0.05	0.01	0.18	0.02	0.05	0.02	0.03	0.04	0.00	0.04	0.07
1991	0.01	0.09	0.01	0.19	0.03	0.06	0.01	0.01	0.02	0.02	0.07
1992	0.01	0.03	0.10	0.03	0.14	0.02	0.02	0.02	0.01	0.01	0.04
1993	0.02	0.01	0.01	0.08	0.01	0.10	0.02	0.02	0.01	0.02	0.09
1994	0.02	0.04	0.03	0.02	0.11	0.01	0.09	0.01	0.03	0.01	0.05
1995	0.03	0.06	0.03	0.01	0.02	0.10	0.00	0.10	0.01	0.01	0.05
1996	0.02	0.06	0.04	0.02	0.01	0.03	0.10	0.02	0.07	0.01	0.06
1997	0.02	0.02	0.05	0.04	0.03	0.04	0.02	0.10	0.02	0.05	0.09
1998	0.02	0.02	0.02	0.05	0.04	0.03	0.01	0.02	0.03	0.01	0.12
1999	0.00	0.02	0.01	0.02	0.04	0.05	0.03	0.02	0.03	0.04	0.11
2000	0.00	0.02	0.05	0.01	0.02	0.05	0.03	0.03	0.02	0.05	0.13
2001	0.01	0.02	0.01	0.03	0.03	0.01	0.02	0.03	0.03	0.02	0.16
2002	0.00	0.02	0.03	0.04	0.09	0.03	0.02	0.05	0.01	0.01	0.14
2003	0.01	0.08	0.04	0.03	0.02	0.04	0.02	0.02	0.03	0.02	0.15
2004	0.01	0.02	0.09	0.02	0.02	0.02	0.03	0.01	0.01	0.02	0.19
2005	0.01	0.04	0.02	0.08	0.02	0.02	0.03	0.04	0.02	0.01	0.14
2006	0.06	0.05	0.03	0.05	0.06	0.02	0.01	0.02	0.03	0.01	0.09
2007	0.02	0.06	0.03	0.03	0.05	0.06	0.02	0.02	0.03	0.02	0.12
2008	0.02	0.03	0.06	0.02	0.03	0.02	0.07	0.01	0.02	0.02	0.11
2009	0.01	0.04	0.03	0.05	0.03	0.01	0.02	0.04	0.02	0.02	0.13
2010	0.06	0.03	0.06	0.02	0.05	0.04	0.02	0.01	0.02	0.01	0.10
2011	0.02	0.07	0.03	0.03	0.03	0.04	0.01	0.01	0.01	0.03	0.14
2012	0.02	0.04	0.08	0.04	0.03	0.03	0.04	0.04	0.01	0.02	0.16
2013	0.02	0.00	0.03	0.08	0.05	0.06	0.04	0.04	0.03	0.01	0.09
2014	0.02	0.04	0.03	0.04	0.06	0.02	0.04	0.04	0.04	0.01	0.13
2015	0.01	0.03	0.04	0.02	0.02	0.05	0.02	0.05	0.02	0.04	0.16

Table 4.5—Yellowfin sole biomass estimates (t) from the annual Bering Sea shelf bottom trawl survey and upper and lower 95% confidence intervals.

Year	Total	Lower CI	Upper CI
1982	3,377,800	2,571,000	4,184,600
1983	3,535,300	2,958,100	4,112,400
1984	3,141,200	2,636,800	3,645,600
1985	2,443,700	1,563,400	3,324,000
1986	1,909,900	1,480,700	2,339,000
1987	2,613,100	2,051,800	3,174,400
1988	2,402,400	1,808,400	2,996,300
1989	2,316,300	1,836,700	2,795,800
1990	2,183,800	1,886,200	2,479,400
1991	2,393,300	2,116,000	2,670,700
1992	2,172,900	1,898,900	2,690,600
1993	2,465,400	2,151,500	2,779,300
1994	2,610,500	2,266,800	2,954,100
1995	2,009,700	1,724,800	2,294,600
1996	2,298,600	1,749,900	2,847,300
1997	2,163,400	1,907,900	2,418,900
1998	2,329,600	2,033,130	2,626,070
1999	1,306,470	1,118,800	1,494,150
2000	1,581,900	1,382,000	1,781,800
2001	1,863,700	1,605,000	2,122,300
2002	2,016,700	1,740,700	2,292,700
2003	2,239,600	1,822,700	2,656,600
2004	2,530,600	2,147,900	2,913,300
2005	2,823,500	2,035,800	3,499,800
2006	2,133,070	1,818,253	2,447,932
2007	2,152,738	1,775,191	2,530,285
2008	2,099,521	1,599,100	2,600,000
2009	1,739,238	1,435,188	2,043,288
2010	2,367,830	1,807,430	2,928,230
2011	2,403,021	1,926,371	2,879,671
2012	1,951,400	1,675,982	2,226,819
2013	2,279,004	1,934,134	2,623,874
2014	2,512,250	2,058,018	2,966,482
2015	1,932,347	1,644,043	2,220,651
2016	2,859,811	2,532,202	3,187,421

Table 4.6. Yellowfin sole population numbers-at-age (millions) estimated from the annual bottom trawl surveys, 1982-2015.

Females

females

year/age	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17+
1979	21	113	150	442	616	386	555	801	626	528	219	274	59	35	29	15
1980	1	92	342	518	800	1055	413	661	880	651	765	285	113	33	23	23
1981	0	20	195	839	692	1321	1155	261	477	744	527	311	168	55	23	45
1982	38	183	349	1211	1485	1424	1619	843	829	832	704	409	246	159	51	84
1983	0	5	59	154	751	1413	843	1065	936	753	1155	866	295	160	60	54
1984	0	53	278	264	427	745	841	1111	1080	941	541	583	480	239	174	133
1985	0	3	105	442	587	406	632	915	441	518	545	384	298	321	205	127
1986	0	8	24	219	349	666	279	574	519	377	284	318	196	250	136	259
1987	0	0	70	120	803	458	843	259	376	599	356	449	243	270	247	688
1988	0	0	7	370	71	1495	560	557	184	239	351	208	360	273	219	886
1989	0	0	14	98	718	234	1337	593	446	74	179	308	234	238	183	565
1990	0	0	70	102	325	1066	192	1257	408	482	101	72	107	78	231	605
1991	0	10	127	248	123	405	896	151	1263	213	525	63	128	87	123	807
1992	0	19	247	485	520	213	286	938	94	825	75	309	129	137	170	715
1993	0	24	100	357	634	434	269	224	1314	78	866	157	165	69	68	674
1994	0	54	95	223	518	905	555	482	284	1170	516	44	274	142	42	588
1995	0	19	153	288	181	889	627	274	135	25	634	21	561	104	80	512
1996	0	16	154	809	288	279	434	517	206	146	151	602	116	637	47	619
1997	0	18	324	502	725	256	239	506	228	114	176	184	500	44	314	533
1998	0	10	83	479	420	900	260	203	370	413	369	170	176	265	67	1167
1999	0	3	65	198	175	185	727	104	107	245	190	186	72	102	175	425
2000	0	11	54	248	208	304	444	537	189	198	237	219	65	117	145	572
2001	0	1	71	239	522	248	403	415	654	374	83	191	154	127	189	617
2002	0	16	123	170	255	778	346	290	229	457	221	91	307	116	152	805
2003	0	15	115	241	251	287	1143	225	279	286	251	103	115	170	168	943
2004	10	33	192	430	560	441	217	966	221	212	218	219	106	20	167	1020
2005	0	53	167	194	602	433	213	487	834	196	144	191	324	170	53	1332

2006	0	67	302	376	276	634	470	176	325	738	133	133	71	156	175	514
2007	0	37	515	348	376	277	504	308	124	227	504	119	137	127	105	724
2008	0	24	115	736	621	546	359	355	198	117	259	350	153	79	85	732
2009	5	38	204	204	1187	609	488	259	210	218	129	138	196	88	43	444
2010	0	33	328	386	438	895	554	517	329	335	155	166	135	173	99	684
2011	0	14	243	539	707	463	769	410	457	204	226	149	142	145	186	619
2012	10	50	229	394	503	293	243	752	256	334	106	156	37	150	128	547
2013	0	4	88	269	420	531	256	221	409	406	358	119	135	133	133	770
2014	0	0	37	421	384	248	420	231	228	523	341	160	144	228	34	819
2015	0	23	3	167	467	350	308	287	249	149	282	258	135	99	80	592

Table 4.6.(continued)

males																
year/age	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17+
1979	21	115	143	390	381	303	583	847	604	406	349	247	54	76	29	36
1980	20	78	306	632	853	1221	457	558	616	568	444	370	147	18	8	8
1981	0	50	200	1047	640	1280	858	394	372	546	534	266	66	83	55	12
1982	89	193	428	1780	1781	1059	1673	644	774	463	471	482	302	8	24	8
1983	0	1	65	183	724	1729	808	1049	676	699	722	566	425	550	77	51
1984	0	68	246	323	497	734	830	612	788	718	358	379	201	316	122	106
1985	0	41	172	419	559	263	652	527	401	451	360	224	260	157	112	65
1986	0	13	47	108	373	652	262	327	284	335	211	205	115	210	82	252
1987	0	5	41	106	838	467	673	445	328	277	210	147	106	142	185	600
1988	0	2	10	435	49	1163	553	443	85	187	28	177	336	189	28	599
1989	0	2	23	181	788	177	1306	513	357	135	50	103	54	204	35	478
1990	0	11	47	121	316	888	195	1144	318	263	40	65	67	24	55	389
1991	0	0	103	354	139	275	1046	68	1137	328	244	74	64	60	53	420
1992	0	0	146	445	566	262	226	812	114	907	193	213	12	12	61	607
1993	0	20	52	233	646	393	279	247	1096	69	842	53	53	50	0	341
1994	4	22	71	166	427	953	656	308	191	822	26	622	46	132	11	303
1995	0	0	169	120	270	667	565	94	179	75	478	13	603	49	24	418
1996	0	76	95	837	244	227	425	344	331	141	139	399	61	449	125	495

1997	0	10	214	425	798	181	184	446	245	194	214	108	514	79	264	416
1998	0	48	70	351	569	832	159	226	204	272	346	140	157	191	113	814
1999	0	5	100	142	225	243	575	146	94	309	269	75	53	28	119	425
2000	0	0	36	219	259	143	509	583	78	215	133	77	92	78	66	547
2001	0	0	87	141	652	341	375	357	562	208	87	158	65	73	140	432
2002	0	58	72	158	309	758	318	333	262	442	194	120	220	161	133	507
2003	0	24	95	178	258	251	1074	238	363	53	284	173	10	71	57	682
2004	4	63	114	469	447	199	395	993	263	81	195	223	103	47	249	456
2005	0	49	166	187	474	476	204	288	972	123	142	121	133	69	93	726
2006	0	101	173	348	332	505	393	288	298	384	116	155	89	39	11	590
2007	0	58	481	352	405	284	545	209	166	252	338	101	133	72	59	620
2008	0	10	99	662	462	483	344	453	225	144	185	329	63	66	35	581
2009	0	65	144	289	946	462	555	248	249	217	78	31	195	30	29	363
2010	0	78	199	418	371	1032	462	510	171	189	159	53	117	151	78	678
2011	1	7	150	385	482	358	792	398	224	176	77	81	136	103	157	440
2012	0	69	274	352	344	273	238	425	297	179	98	67	91	34	100	2
2013	0	7	92	366	384	481	211	268	445	200	200	33	89	100	118	612
2014	0	0	0	9	366	396	286	338	310	251	400	206	193	20	192	841
2015	1	29	36	131	426	332	301	312	318	48	180	131	80	1	80	492

Table 4.7-Occurance of yellowfin sole in the Bering Sea trawl survey and collections of length and age structures and the number of otoliths aged from each survey.

Year	Total Hauls	Hauls w/Len	Number lengths	Hauls w/otoliths	Hauls w/ages	Number otoliths	Number ages
1982	334	246	37023	35	35	744	744
1983	353	256	33924	37	37	709	709
1984	355	271	33894	56	56	821	796
1985	357	261	33824	44	43	810	802
1986	354	249	30470	34	34	739	739
1987	357	224	31241	16	16	798	798
1988	373	254	27138	14	14	543	543
1989	374	236	29672	24	24	740	740
1990	371	251	30257	28	28	792	792
1991	372	248	27986	26	26	742	742
1992	356	229	23628	16	16	606	606
1993	375	242	26651	20	20	549	549
1994	375	269	24448	14	14	526	522
1995	376	254	22116	20	20	654	647
1996	375	247	27505	16	16	729	721
1997	376	262	26034	11	11	470	466
1998	375	310	34509	15	15	575	570
1999	373	276	28431	31	31	777	770
2000	372	255	24880	20	20	517	511
2001	375	251	26558	25	25	604	593
2002	375	246	26309	32	32	738	723
2003	376	241	27135	37	37	699	695
2004	375	251	26103	26	26	725	712
2005	373	251	24658	34	34	644	635
2006	376	246	28470	39	39	428	426
2007	376	247	24790	66	66	779	772
2008	375	238	25848	65	65	858	830
2009	376	235	22018	70	70	784	752
2010	376	228	20619	77	77	841	827
2011	376	228	21665	65	64	784	753
2012	376	242	23519	72	72	993	973
2013	376	232	23261	70	70	821	803
2014	376	219	20229	52	52	799	790
2015	376	223	20830	73	73	878	875
2016	376	238	92360	69		884	

Table 4.8—Total tonnage of yellowfin sole caught in resource assessment surveys in the eastern Bering Sea from 1977-2016.

Year	Research catch (t)
1977	60
1978	71
1979	147
1980	92
1981	74
1982	158
1983	254
1984	218
1985	105
1986	68
1987	92
1988	138
1989	148
1990	129
1991	118
1992	60
1993	95
1994	91
1995	95
1996	72
1997	76
1998	79
1999	61
2000	72
2001	75
2002	76
2003	78
2004	114
2005	94
2006	74
2007	74
2008	69
2009	60
2010	79
2011	77
2012	64
2013	75
2014	81
2015	64
2016	98

Table 4.10. Female yellowfin sole proportion mature at age from Nichol (1995) and TenBrink and Wilderbuer (2015).

Age	1992, 1993 samples	2012 samples	Combined
1	0.00	0	0
2	0.00	0	0
3	.001	0	0
4	.004	0	0
5	.008	0	0
6	.020	.01	0.01
7	.046	.03	0.04
8	.104	.09	0.10
9	.217	.21	0.21
10	.397	.43	0.41
11	.612	.68	0.65
12	.790	.86	0.83
13	.899	.94	0.92
14	.955	.98	0.97
15	.981	.99	0.99
16	.992	1.0	1.0
17	.997	1.0	1.0
18	1.0	1.0	1.0
19	1.0	1.0	1.0
20	1.0	1.0	1.0

Table 4.11. Key equations used in the population dynamics model.

$N_{t,1} = R_t = R_0 e^{\tau_t}$, $\tau_t \sim N(0, \delta_R^2)$	Recruitment 1956-75
$N_{t,1} = R_t = R_y e^{\tau_t}$, $\tau_t \sim N(0, \delta_R^2)$	Recruitment 1976-96
$C_{t,a} = \frac{F_{t,a}}{Z_{t,a}} (1 - e^{-z_{t,a}}) N_{t,a}$	Catch in year t for age a fish
$N_{t+1,a+1} = N_{t,a} e^{-z_{t,a}}$	Numbers of fish in year $t+1$ at age a
$N_{t+1,A} = N_{t,A-1} e^{-z_{t,A-1}} + N_{t,A} e^{-z_{t,A}}$	Numbers of fish in the “plus group”
$S_t = \sum N_{t,a} W_{t,a} \phi_a$	Spawning biomass
$Z_{t,a} = F_{t,a} + M$	Total mortality in year t at age a
$F_{t,a} = s_a \mu^F \exp^{\varepsilon_t^F}$, $\varepsilon_t^F \sim N(0, \sigma^{2F})$	Fishing mortality
$s_a = \frac{1}{1 + (e^{-\alpha + \beta a})}$	Age-specific fishing selectivity
$C_t = \sum C_{t,a}$	Total catch in numbers
$P_{t,a} = C_{t,a} / C_t$	Proportion at age in catch
$SurB_t = q \sum N_{t,a} W_{t,a} v_a$	Survey biomass

Table 4.11—continued.

$$qprior = \lambda \frac{0.5(\ln q_{est,t} - \ln q_{prior})^2}{\sigma_q^2} \quad \text{survey catchability prior (when estimated)}$$

$$mprior = \lambda \frac{0.5(\ln m_{est} - \ln m_{prior})^2}{\sigma_m^2} \quad \text{natural mortality prior (when estimated)}$$

$$reclike = \lambda \left(\sum_{i=1965}^{\text{endyear}} (R - R_i)^2 + \sum_{a=1}^{20} (R_{init} - R_{init,a})^2 + \frac{1}{2 \left(\left(\sum_{i=1965}^{\text{endyear}} R - R_i \right) \frac{1}{n+1} \right)} \right) \quad \text{recruitment likelihood}$$

$$catchlike = \lambda \sum_{i=\text{startyear}}^{\text{endyear}} (\ln C_{obs,i} - \ln C_{est,i})^2 \quad \text{catch likelihood}$$

$$surveylike = \lambda \frac{(\ln B - \ln \hat{B})^2}{2\sigma^2} \quad \text{survey likelihood}$$

$$SurvAgelike = \sum_{i,t} m_t P_{t,a} \ln \frac{\hat{P}_{t,a}}{P_{t,a}} \quad \text{survey age composition likelihood}$$

$$FishAgelike = \sum_{i,t} m_t P_{t,a} \ln \frac{\hat{P}_{t,a}}{P_{t,a}} \quad \text{fishery age composition likelihood}$$

Table 4.12. Variables used in the population dynamics model.

Variables

R_t	Age 1 recruitment in year t
R_0	Geometric mean value of age 1 recruitment, 1956-75
R_γ	Geometric mean value of age 1 recruitment, 1976-2014
τ_t	Recruitment deviation in year t
$N_{t,a}$	Number of fish in year t at age a
$C_{t,a}$	Catch numbers of fish in year t at age a
$P_{t,a}$	Proportion of the numbers of fish age a in year t
C_t	Total catch numbers in year t
$W_{t,a}$	Mean body weight (kg) of fish age a in year t
ϕ_a	Proportion of mature females at age a
$F_{t,a}$	Instantaneous annual fishing mortality of age a fish in year t
M	Instantaneous natural mortality, assumed constant over all ages and years
$Z_{t,a}$	Instantaneous total mortality for age a fish in year t
s_a	Age-specific fishing gear selectivity
μ^F	Median year-effect of fishing mortality
ε_t^F	The residual year-effect of fishing mortality
v_a	Age-specific survey selectivity
α	Slope parameter in the logistic selectivity equation
β	Age at 50% selectivity parameter in the logistic selectivity equation
σ_t	Standard error of the survey biomass in year t

Table 4.13. Models evaluated for stock productivity in the 2014 stock assessment of yellowfin sole

	Model A	Model B
Years included	1955-2010	1978-2010
Fmsy	0.163	0.125
Bmsy (t)	336,000	424,000
ABC (t)	366,100	260,800
OFL (t)	368,800	287,000
Buffer between ABC and OFL	<1%	9%

Table 4.15. Model estimates of yellowfin sole full selection fishing mortality and exploitation rate (catch/total biomass).

Year	Full selection F	Exploitation Rate
1964	0.28	0.13
1965	0.26	0.06
1966	0.45	0.12
1967	0.58	0.19
1968	0.47	0.11
1969	0.66	0.21
1970	0.67	0.17
1971	1.04	0.19
1972	0.31	0.05
1973	0.53	0.07
1974	0.16	0.03
1975	0.14	0.04
1976	0.13	0.03
1977	0.06	0.02
1978	0.12	0.05
1979	0.07	0.04
1980	0.08	0.03
1981	0.06	0.03
1982	0.05	0.03
1983	0.05	0.03
1984	0.07	0.05
1985	0.11	0.07
1986	0.10	0.07
1987	0.10	0.06
1988	0.13	0.08
1989	0.10	0.05
1990	0.04	0.03
1991	0.05	0.03
1992	0.08	0.05
1993	0.06	0.03
1994	0.07	0.05
1995	0.06	0.04
1996	0.07	0.04
1997	0.10	0.06
1998	0.06	0.04
1999	0.04	0.03
2000	0.05	0.03
2001	0.04	0.03
2002	0.04	0.03
2003	0.04	0.03
2004	0.04	0.02
2005	0.05	0.03
2006	0.05	0.03

2007	0.07	0.04
2008	0.08	0.05
2009	0.06	0.04
2010	0.07	0.04
2011	0.09	0.06
2012	0.09	0.05
2013	0.11	0.06
2014	0.11	0.06
2015	0.11	0.05
2016	0.08	0.05

Table 4.17. Model estimates of yellowfin sole age 2+ total biomass (t) and begin-year female spawning biomass (t) from the 2015 and 2016 stock assessments.

Year	2016 Assessment			Total biomass	2015 Assessment			
	Female spawning biomass	lower 95% C.I.	upper 95% C.I.		Female spawning biomass	lower 95% C.I.	upper 95% C.I.	Total biomass
1964	17,615	0	118,681	839,742	0	38,573	25,103	845,645
1965	36,707	0	132,549	835,631	801,316	869,946	49,136	839,968
1966	61,107	0	176,538	879,970	845,017	914,923	86,778	887,261
1967	75,435	24,122	227,197	863,836	828,184	899,488	112,346	874,716
1968	74,113	36,145	305,498	782,492	747,066	817,918	115,001	790,128
1969	69,945	38,949	412,165	810,677	772,126	849,228	119,277	823,558
1970	65,929	45,598	515,323	786,026	745,077	826,975	86,186	793,085
1971	59,749	44,716	635,426	846,563	799,225	893,901	58,853	840,912
1972	41,419	27,598	749,440	909,321	853,546	965,096	47,919	909,173
1973	48,795	35,178	813,345	1,155,700	1,087,642	1,223,758	52,638	1,152,600
1974	64,968	50,999	909,890	1,397,720	1,316,327	1,479,113	66,966	1,390,750
1975	114,443	96,313	989,376	1,739,260	1,641,040	1,837,480	116,928	1,729,150
1976	175,884	154,058	1,035,446	2,035,950	1,922,097	2,149,803	177,978	2,022,420
1977	267,530	239,644	1,024,757	2,336,020	2,207,505	2,464,535	268,408	2,319,060
1978	381,170	346,426	1,020,864	2,620,880	2,478,393	2,763,367	380,350	2,600,470
1979	494,992	453,142	969,031	2,771,670	2,617,756	2,925,584	492,288	2,748,190
1980	620,393	571,189	945,558	2,944,900	2,781,001	3,108,799	615,902	2,918,630
1981	735,618	679,316	959,799	3,103,120	2,931,069	3,275,171	729,260	3,074,160
1982	799,728	739,950	1,037,043	3,208,630	3,034,186	3,383,074	792,193	3,177,480
1983	895,921	830,945	1,116,056	3,183,300	3,008,550	3,358,050	887,476	3,151,850
1984	971,246	902,170	1,148,466	3,393,380	3,208,313	3,578,447	961,354	3,358,420
1985	1,013,620	940,507	1,151,663	3,394,790	3,204,282	3,585,298	1,002,360	3,358,010
1986	996,871	922,160	1,147,621	3,108,320	2,921,755	3,294,885	984,813	3,072,090
1987	986,120	909,314	1,084,286	3,058,180	2,868,449	3,247,911	973,220	3,020,440
1988	927,181	851,995	1,043,384	2,955,620	2,765,580	3,145,660	914,112	2,917,250
1989	896,354	820,891	978,903	2,994,610	2,796,796	3,192,424	882,781	2,953,320
1990	903,497	827,073	968,280	2,858,500	2,662,744	3,054,256	889,446	2,817,460
1991	977,265	897,625	955,946	2,964,280	2,761,988	3,166,572	962,178	2,921,120
1992	1,051,080	967,695	952,293	3,150,890	2,937,437	3,364,343	1,034,790	3,104,160
1993	1,079,390	993,276	950,364	3,167,600	2,948,384	3,386,816	1,062,210	3,118,100
1994	1,078,550	991,859	955,937	3,194,550	2,972,156	3,416,944	1,060,960	3,142,370
1995	1,072,910	984,971	981,901	2,972,780	2,757,673	3,187,887	1,054,750	2,922,000
1996	1,007,480	922,326	990,766	2,881,690	2,668,577	3,094,803	989,680	2,829,580
1997	968,198	883,618	1,005,017	2,881,350	2,664,581	3,098,119	950,181	2,825,950
1998	903,440	820,982	1,003,624	2,612,680	2,406,208	2,819,152	885,551	2,558,750
1999	891,856	809,586	976,873	2,434,810	2,236,710	2,632,910	873,513	2,382,480
2000	876,306	794,815	940,254	2,472,620	2,272,633	2,672,607	857,630	2,417,300
2001	868,908	788,103	915,992	2,397,250	2,201,205	2,593,295	849,655	2,341,700
2002	864,250	784,076	895,672	2,430,820	2,233,003	2,628,637	844,375	2,372,570
2003	869,246	788,965	882,511	2,618,170	2,406,359	2,829,981	848,353	2,553,240

2004	893,962	811,856	881,374	2,805,970	2,578,635	3,033,305	871,486	2,735,560
2005	905,612	822,422	848,169	2,906,560	2,671,057	3,142,063	881,745	2,835,630
2006	920,437	834,870	853,370	2,887,410	2,650,047	3,124,773	894,910	2,816,810
2007	921,166	833,788	862,526	2,895,700	2,654,321	3,137,079	894,163	2,827,450
2008	894,603	806,047	88,556	2,771,050	2,532,450	3,009,650	864,230	2,706,420
2009	858,763	770,266	88,497	2,606,720	2,373,090	2,840,350	821,948	2,532,050
2010	835,187	746,461	88,726	2,670,900	2,431,224	2,910,576	793,112	2,593,210
2011	815,498	726,321	89,177	2,701,250	2,454,740	2,947,760	767,772	2,614,460
2012	802,230	711,911	90,319	2,693,730	2,441,145	2,946,315	749,486	2,602,540
2013	799,268	706,687	92,581	2,647,970	2,392,068	2,903,872	743,198	2,549,110
2014	764,979	671,939	93,040	2,449,450	2,199,567	2,699,333	706,208	2,339,690
2015	767,803	671,170	96,633	2,451,240	2,188,420	2,714,060	697,207	2,313,020
2016	775,148	675,773	99,375	2,457,260	2,177,392	2,737,128		

Table 4.18—Model estimates of yellowfin sole population numbers at age (billions) for 1954-2016.

Females																				
	1	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17	18	19	20
1954	1.00	0.41	0.28	0.26	0.25	0.25	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24
1955	0.81	0.89	0.36	0.25	0.23	0.23	0.22	0.22	0.22	0.21	0.21	0.21	0.21	0.21	0.21	0.21	0.21	0.21	0.21	0.42
1956	0.60	0.72	0.79	0.32	0.22	0.20	0.20	0.19	0.19	0.19	0.19	0.19	0.19	0.19	0.19	0.19	0.19	0.19	0.19	0.56
1957	1.78	0.54	0.64	0.70	0.29	0.20	0.18	0.18	0.17	0.17	0.17	0.17	0.16	0.16	0.16	0.16	0.16	0.16	0.16	0.65
1958	1.22	1.58	0.47	0.57	0.62	0.25	0.18	0.16	0.16	0.15	0.15	0.15	0.14	0.14	0.14	0.14	0.14	0.14	0.14	0.71
1959	0.91	1.08	1.40	0.42	0.50	0.55	0.23	0.16	0.14	0.14	0.13	0.13	0.13	0.12	0.12	0.12	0.12	0.12	0.12	0.73
1960	0.85	0.81	0.96	1.24	0.37	0.45	0.48	0.20	0.13	0.11	0.11	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.67
1961	0.50	0.76	0.72	0.85	1.10	0.33	0.39	0.40	0.15	0.09	0.07	0.06	0.06	0.06	0.06	0.06	0.06	0.06	0.06	0.44
1962	0.93	0.44	0.67	0.64	0.75	0.96	0.28	0.29	0.24	0.07	0.03	0.02	0.02	0.02	0.02	0.02	0.02	0.02	0.02	0.16
1963	0.47	0.82	0.39	0.59	0.55	0.60	0.59	0.07	0.02	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
1964	0.43	0.42	0.73	0.34	0.52	0.48	0.52	0.49	0.06	0.01	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
1965	0.56	0.38	0.37	0.65	0.31	0.46	0.43	0.46	0.42	0.05	0.01	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
1966	0.60	0.50	0.34	0.33	0.57	0.27	0.41	0.37	0.38	0.31	0.03	0.01	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
1967	1.24	0.53	0.44	0.30	0.29	0.51	0.24	0.35	0.30	0.26	0.19	0.02	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
1968	1.88	1.10	0.47	0.39	0.26	0.26	0.44	0.19	0.23	0.16	0.13	0.09	0.01	0.00	0.00	0.00	0.00	0.00	0.00	0.00
1969	1.92	1.67	0.97	0.42	0.35	0.23	0.22	0.37	0.14	0.14	0.09	0.07	0.05	0.01	0.00	0.00	0.00	0.00	0.00	0.00
1970	2.53	1.71	1.48	0.86	0.37	0.31	0.21	0.19	0.29	0.09	0.07	0.04	0.03	0.02	0.00	0.00	0.00	0.00	0.00	0.00
1971	2.80	2.24	1.51	1.31	0.77	0.33	0.27	0.18	0.17	0.25	0.07	0.05	0.02	0.02	0.01	0.00	0.00	0.00	0.00	0.00
1972	2.20	2.48	1.99	1.34	1.17	0.68	0.26	0.11	0.06	0.05	0.08	0.02	0.02	0.01	0.01	0.00	0.00	0.00	0.00	0.00
1973	1.52	1.95	2.20	1.76	1.19	1.03	0.60	0.23	0.10	0.05	0.04	0.05	0.02	0.01	0.00	0.00	0.00	0.00	0.00	0.00
1974	2.04	1.35	1.73	1.95	1.56	1.06	0.92	0.53	0.20	0.08	0.04	0.03	0.03	0.01	0.01	0.00	0.00	0.00	0.00	0.00
1975	2.40	1.81	1.19	1.53	1.73	1.39	0.94	0.81	0.47	0.18	0.07	0.03	0.02	0.02	0.01	0.00	0.00	0.00	0.00	0.00
1976	1.58	2.13	1.61	1.06	1.36	1.53	1.22	0.81	0.68	0.37	0.14	0.05	0.02	0.02	0.02	0.00	0.00	0.00	0.00	0.00
1977	1.98	1.40	1.88	1.42	0.94	1.20	1.36	1.07	0.70	0.57	0.30	0.11	0.04	0.02	0.01	0.01	0.00	0.00	0.00	0.00
1978	1.30	1.76	1.24	1.67	1.26	0.83	1.05	1.17	0.92	0.59	0.48	0.25	0.09	0.03	0.02	0.01	0.01	0.00	0.00	0.00
1979	0.83	1.15	1.56	1.10	1.48	1.11	0.72	0.90	0.97	0.74	0.47	0.38	0.20	0.07	0.03	0.01	0.01	0.01	0.00	0.00
1980	1.58	0.73	1.02	1.38	0.97	1.31	0.98	0.63	0.77	0.82	0.62	0.39	0.31	0.16	0.06	0.02	0.01	0.01	0.01	0.00

Table 4.18—Model estimates of yellowfin sole population numbers at age (billions) for 1954-2016 (continued).

Females																				
	1	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17	18	19	20
1981	1.17	1.41	0.65	0.90	1.22	0.86	1.15	0.85	0.54	0.66	0.69	0.51	0.32	0.26	0.13	0.05	0.02	0.01	0.01	0.01
1982	3.38	1.04	1.25	0.58	0.80	1.08	0.76	1.01	0.74	0.46	0.55	0.58	0.43	0.27	0.21	0.11	0.04	0.02	0.01	0.01
1983	0.63	3.00	0.92	1.10	0.51	0.71	0.95	0.66	0.87	0.63	0.39	0.47	0.49	0.36	0.23	0.18	0.10	0.03	0.01	0.02
1984	2.79	0.56	2.66	0.82	0.98	0.45	0.62	0.82	0.57	0.74	0.53	0.33	0.40	0.41	0.31	0.19	0.15	0.08	0.03	0.03
1985	0.96	2.47	0.49	2.36	0.73	0.87	0.40	0.54	0.71	0.48	0.61	0.44	0.28	0.33	0.34	0.25	0.16	0.13	0.07	0.04
1986	0.73	0.85	2.19	0.44	2.09	0.64	0.76	0.34	0.46	0.58	0.38	0.49	0.35	0.22	0.26	0.27	0.20	0.13	0.10	0.09
1987	1.00	0.65	0.75	1.94	0.39	1.85	0.56	0.66	0.29	0.37	0.47	0.31	0.39	0.28	0.18	0.21	0.22	0.16	0.10	0.15
1988	1.36	0.89	0.58	0.67	1.72	0.34	1.64	0.50	0.57	0.24	0.31	0.38	0.25	0.32	0.23	0.14	0.17	0.17	0.13	0.20
1989	1.35	1.21	0.79	0.51	0.59	1.53	0.30	1.43	0.43	0.47	0.19	0.24	0.30	0.19	0.25	0.18	0.11	0.13	0.14	0.26
1990	0.67	1.20	1.07	0.70	0.45	0.53	1.35	0.27	1.25	0.36	0.39	0.16	0.20	0.24	0.16	0.20	0.14	0.09	0.11	0.32
1991	0.75	0.60	1.06	0.95	0.62	0.40	0.47	1.19	0.23	1.08	0.31	0.33	0.13	0.17	0.20	0.13	0.17	0.12	0.08	0.36
1992	1.65	0.66	0.53	0.94	0.84	0.55	0.36	0.41	1.04	0.20	0.93	0.26	0.28	0.11	0.14	0.17	0.11	0.14	0.10	0.37
1993	0.99	1.47	0.59	0.47	0.84	0.75	0.48	0.31	0.36	0.89	0.17	0.77	0.22	0.23	0.09	0.12	0.14	0.09	0.12	0.39
1994	0.83	0.87	1.30	0.52	0.42	0.74	0.66	0.43	0.27	0.31	0.76	0.14	0.65	0.18	0.19	0.08	0.10	0.12	0.08	0.42
1995	0.84	0.74	0.78	1.15	0.46	0.37	0.65	0.58	0.37	0.23	0.26	0.63	0.12	0.53	0.15	0.16	0.06	0.08	0.10	0.41
1996	2.07	0.74	0.65	0.69	1.02	0.41	0.32	0.57	0.49	0.31	0.19	0.21	0.53	0.10	0.45	0.13	0.13	0.05	0.07	0.43
1997	0.89	1.84	0.66	0.58	0.61	0.90	0.36	0.28	0.49	0.42	0.26	0.16	0.18	0.44	0.08	0.37	0.10	0.11	0.05	0.41
1998	0.75	0.79	1.63	0.58	0.51	0.54	0.79	0.31	0.24	0.41	0.34	0.21	0.13	0.14	0.35	0.07	0.30	0.08	0.09	0.36
1999	0.91	0.66	0.70	1.44	0.52	0.45	0.47	0.69	0.27	0.20	0.35	0.29	0.18	0.11	0.12	0.29	0.06	0.25	0.07	0.38
2000	1.30	0.81	0.59	0.62	1.28	0.46	0.40	0.42	0.61	0.23	0.18	0.29	0.25	0.15	0.09	0.10	0.25	0.05	0.21	0.38
2001	0.84	1.15	0.72	0.52	0.55	1.13	0.41	0.36	0.37	0.53	0.20	0.15	0.25	0.21	0.13	0.08	0.09	0.21	0.04	0.50
2002	1.18	0.75	1.02	0.64	0.46	0.49	1.00	0.36	0.31	0.32	0.45	0.17	0.13	0.21	0.18	0.11	0.07	0.07	0.18	0.46
2003	1.14	1.05	0.66	0.90	0.57	0.41	0.43	0.89	0.32	0.27	0.27	0.39	0.15	0.11	0.18	0.15	0.09	0.06	0.06	0.54
2004	1.92	1.01	0.93	0.59	0.80	0.50	0.36	0.38	0.78	0.27	0.23	0.23	0.33	0.12	0.09	0.15	0.13	0.08	0.05	0.51
2005	0.91	1.70	0.90	0.82	0.52	0.71	0.44	0.32	0.33	0.67	0.23	0.20	0.20	0.28	0.11	0.08	0.13	0.11	0.07	0.48
2006	1.11	0.81	1.51	0.80	0.73	0.46	0.63	0.39	0.28	0.29	0.57	0.20	0.17	0.17	0.24	0.09	0.07	0.11	0.09	0.46
2007	1.40	0.99	0.71	1.34	0.70	0.64	0.40	0.54	0.33	0.23	0.24	0.48	0.17	0.14	0.14	0.20	0.07	0.06	0.09	0.46

Females

	1	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17	18	19	20
2008	1.20	1.24	0.87	0.63	1.19	0.62	0.57	0.35	0.46	0.28	0.20	0.20	0.40	0.14	0.12	0.12	0.16	0.06	0.05	0.46
2009	1.22	1.06	1.10	0.77	0.56	1.05	0.55	0.49	0.30	0.39	0.23	0.16	0.17	0.33	0.11	0.10	0.10	0.13	0.05	0.41
2010	1.43	1.08	0.94	0.98	0.69	0.50	0.93	0.48	0.43	0.25	0.32	0.19	0.13	0.14	0.27	0.09	0.08	0.08	0.11	0.39
2011	0.65	1.27	0.96	0.83	0.87	0.61	0.44	0.81	0.42	0.36	0.21	0.27	0.16	0.11	0.11	0.23	0.08	0.07	0.07	0.41
2012	0.53	0.57	1.12	0.85	0.74	0.77	0.54	0.39	0.70	0.35	0.30	0.17	0.22	0.13	0.09	0.09	0.18	0.06	0.05	0.39
2013	1.11	0.47	0.51	1.00	0.76	0.65	0.68	0.47	0.33	0.59	0.29	0.25	0.14	0.18	0.11	0.07	0.08	0.15	0.05	0.36
2014	1.04	0.98	0.41	0.45	0.88	0.67	0.58	0.60	0.41	0.28	0.49	0.24	0.20	0.11	0.14	0.08	0.06	0.06	0.12	0.33
2015	1.11	0.92	0.87	0.37	0.40	0.78	0.59	0.51	0.52	0.35	0.23	0.40	0.19	0.16	0.09	0.11	0.07	0.05	0.05	0.36
2016	1.13	0.99	0.82	0.77	0.33	0.36	0.69	0.52	0.45	0.45	0.30	0.20	0.32	0.15	0.12	0.07	0.09	0.05	0.04	0.32

Table 4.18—Model estimates of yellowfin sole population numbers at age (billions) for 1954-2016 (continued).

Males

	1	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17	18	19	20
1954	1.00	0.69	0.33	0.27	0.26	0.25	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24	0.24
1955	0.81	0.89	0.61	0.29	0.24	0.23	0.22	0.22	0.22	0.21	0.21	0.21	0.21	0.21	0.21	0.21	0.21	0.21	0.21	0.42
1956	0.60	0.72	0.79	0.54	0.26	0.21	0.21	0.19	0.19	0.19	0.19	0.19	0.19	0.19	0.19	0.19	0.19	0.19	0.19	0.56
1957	1.78	0.54	0.64	0.70	0.48	0.23	0.19	0.18	0.17	0.17	0.17	0.16	0.16	0.16	0.16	0.16	0.16	0.16	0.16	0.65
1958	1.22	1.58	0.47	0.57	0.62	0.43	0.20	0.17	0.16	0.15	0.15	0.15	0.14	0.14	0.14	0.14	0.14	0.14	0.14	0.71
1959	0.91	1.08	1.40	0.42	0.50	0.55	0.38	0.18	0.15	0.14	0.13	0.13	0.13	0.12	0.12	0.12	0.12	0.12	0.12	0.73
1960	0.85	0.81	0.96	1.24	0.37	0.45	0.49	0.33	0.15	0.12	0.11	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.67
1961	0.50	0.76	0.72	0.85	1.10	0.33	0.39	0.43	0.28	0.12	0.08	0.07	0.06	0.06	0.06	0.06	0.06	0.06	0.06	0.44
1962	0.93	0.44	0.67	0.64	0.75	0.98	0.29	0.35	0.37	0.24	0.09	0.04	0.03	0.02	0.02	0.02	0.02	0.02	0.02	0.16
1963	0.47	0.82	0.39	0.60	0.57	0.67	0.87	0.26	0.31	0.33	0.21	0.07	0.02	0.00	0.00	0.00	0.00	0.00	0.00	0.00
1964	0.43	0.42	0.73	0.35	0.53	0.49	0.57	0.69	0.19	0.21	0.21	0.13	0.04	0.01	0.00	0.00	0.00	0.00	0.00	0.00
1965	0.56	0.38	0.37	0.64	0.25	0.36	0.33	0.38	0.46	0.13	0.14	0.14	0.09	0.03	0.01	0.00	0.00	0.00	0.00	0.00
1966	0.60	0.50	0.34	0.33	0.57	0.22	0.32	0.29	0.34	0.39	0.10	0.10	0.10	0.06	0.02	0.01	0.00	0.00	0.00	0.00
1967	1.24	0.53	0.44	0.30	0.29	0.51	0.20	0.28	0.26	0.28	0.30	0.06	0.06	0.06	0.03	0.01	0.00	0.00	0.00	0.00
1968	1.88	1.10	0.47	0.39	0.26	0.26	0.45	0.17	0.24	0.22	0.20	0.17	0.03	0.03	0.03	0.02	0.01	0.00	0.00	0.00
1969	1.92	1.67	0.97	0.42	0.35	0.23	0.23	0.40	0.15	0.21	0.19	0.16	0.12	0.02	0.02	0.02	0.01	0.00	0.00	0.00
1970	2.53	1.71	1.48	0.86	0.37	0.31	0.21	0.20	0.32	0.09	0.11	0.09	0.08	0.06	0.01	0.01	0.01	0.00	0.00	0.00
1971	2.80	2.24	1.51	1.31	0.77	0.33	0.25	0.12	0.09	0.14	0.04	0.05	0.04	0.03	0.03	0.00	0.00	0.00	0.00	0.00

1972 2.20 2.48 1.99 1.34 1.17 0.68 0.29 0.23 0.11 0.08 0.12 0.03 0.02 0.01 0.01 0.01 0.00 0.00 0.00 0.00

Table 4.18. Model estimates of yellowfin sole population numbers at age (billions) for 1954-2016 (continued).

Males																				
	1	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17	18	19	20
1973	1.52	1.95	2.20	1.76	1.19	1.03	0.59	0.23	0.16	0.07	0.05	0.08	0.02	0.01	0.01	0.01	0.01	0.00	0.00	0.00
1974	2.04	1.35	1.73	1.95	1.56	1.05	0.90	0.46	0.13	0.08	0.04	0.03	0.04	0.01	0.01	0.00	0.00	0.00	0.00	0.00
1975	2.40	1.81	1.19	1.53	1.73	1.37	0.90	0.73	0.35	0.10	0.06	0.03	0.02	0.03	0.01	0.01	0.00	0.00	0.00	0.00
1976	1.58	2.13	1.61	1.06	1.36	1.53	1.20	0.76	0.58	0.28	0.08	0.05	0.02	0.02	0.02	0.01	0.00	0.00	0.00	0.00
1977	1.98	1.40	1.88	1.42	0.94	1.20	1.35	1.06	0.66	0.48	0.22	0.06	0.04	0.02	0.01	0.02	0.00	0.00	0.00	0.00
1978	1.30	1.76	1.24	1.67	1.26	0.83	1.06	1.19	0.91	0.56	0.41	0.19	0.05	0.03	0.01	0.01	0.02	0.00	0.00	0.00
1979	0.83	1.15	1.56	1.10	1.48	1.12	0.73	0.92	1.00	0.75	0.44	0.32	0.15	0.04	0.02	0.01	0.01	0.01	0.00	0.01
1980	1.58	0.73	1.02	1.38	0.97	1.31	0.98	0.64	0.79	0.84	0.62	0.37	0.27	0.12	0.03	0.02	0.01	0.01	0.01	0.01
1981	1.17	1.41	0.65	0.90	1.22	0.86	1.16	0.86	0.56	0.68	0.73	0.53	0.31	0.22	0.10	0.03	0.02	0.01	0.01	0.01
1982	3.38	1.04	1.25	0.57	0.80	1.08	0.76	1.01	0.75	0.48	0.58	0.61	0.45	0.26	0.19	0.08	0.02	0.01	0.01	0.02
1983	0.63	3.00	0.92	1.10	0.51	0.71	0.95	0.66	0.87	0.64	0.41	0.49	0.52	0.38	0.22	0.16	0.07	0.02	0.01	0.02
1984	2.79	0.56	2.66	0.82	0.98	0.45	0.62	0.82	0.56	0.74	0.54	0.34	0.42	0.44	0.32	0.19	0.13	0.06	0.02	0.03
1985	0.96	2.47	0.49	2.36	0.72	0.87	0.40	0.54	0.70	0.47	0.61	0.45	0.28	0.34	0.36	0.26	0.15	0.11	0.05	0.04
1986	0.73	0.85	2.19	0.44	2.09	0.64	0.76	0.34	0.44	0.56	0.37	0.49	0.36	0.23	0.27	0.29	0.21	0.12	0.09	0.07
1987	1.00	0.65	0.76	1.94	0.39	1.85	0.56	0.65	0.28	0.36	0.45	0.30	0.39	0.28	0.18	0.22	0.23	0.17	0.10	0.12
1988	1.36	0.89	0.58	0.67	1.72	0.34	1.64	0.49	0.55	0.23	0.29	0.36	0.24	0.31	0.23	0.15	0.18	0.18	0.13	0.18
1989	1.35	1.21	0.79	0.51	0.59	1.53	0.30	1.43	0.41	0.44	0.18	0.22	0.28	0.19	0.24	0.18	0.11	0.14	0.14	0.24
1990	0.67	1.20	1.07	0.70	0.45	0.53	1.35	0.27	1.25	0.35	0.36	0.14	0.18	0.23	0.15	0.20	0.14	0.09	0.11	0.31
1991	0.75	0.60	1.06	0.95	0.62	0.40	0.47	1.19	0.23	1.07	0.30	0.31	0.12	0.15	0.19	0.13	0.17	0.12	0.08	0.36
1992	1.65	0.66	0.53	0.94	0.84	0.55	0.36	0.41	1.03	0.20	0.91	0.25	0.26	0.10	0.13	0.16	0.11	0.14	0.10	0.37
1993	0.99	1.47	0.59	0.47	0.84	0.75	0.48	0.31	0.35	0.86	0.16	0.74	0.21	0.21	0.08	0.11	0.13	0.09	0.12	0.39
1994	0.83	0.87	1.30	0.52	0.42	0.74	0.66	0.42	0.27	0.30	0.74	0.14	0.63	0.17	0.18	0.07	0.09	0.11	0.07	0.42
1995	0.84	0.74	0.78	1.15	0.46	0.37	0.65	0.57	0.36	0.23	0.25	0.61	0.11	0.52	0.14	0.15	0.06	0.07	0.09	0.41
1996	2.07	0.74	0.65	0.69	1.02	0.41	0.32	0.57	0.49	0.30	0.19	0.21	0.51	0.10	0.43	0.12	0.12	0.05	0.06	0.42
1997	0.89	1.84	0.66	0.58	0.61	0.90	0.36	0.28	0.48	0.41	0.25	0.16	0.17	0.42	0.08	0.36	0.10	0.10	0.04	0.40

Table 4.18. Model estimates of yellowfin sole population numbers at age (billions) for 1954-2016 (continued).

Males																				
	1	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17	18	19	20
1998	0.75	0.79	1.63	0.59	0.51	0.54	0.79	0.31	0.24	0.40	0.33	0.20	0.13	0.14	0.34	0.06	0.29	0.08	0.08	0.35
1999	0.91	0.66	0.70	1.44	0.52	0.46	0.48	0.69	0.27	0.20	0.34	0.28	0.17	0.11	0.12	0.28	0.05	0.24	0.07	0.36
2000	1.30	0.81	0.59	0.62	1.28	0.46	0.40	0.42	0.61	0.24	0.17	0.29	0.24	0.15	0.09	0.10	0.24	0.05	0.21	0.37
2001	0.84	1.15	0.72	0.52	0.55	1.14	0.41	0.36	0.37	0.53	0.20	0.15	0.24	0.20	0.12	0.08	0.08	0.20	0.04	0.48
2002	1.18	0.75	1.02	0.64	0.46	0.49	1.01	0.36	0.32	0.33	0.47	0.17	0.13	0.21	0.17	0.10	0.06	0.07	0.17	0.44
2003	1.14	1.05	0.66	0.90	0.57	0.41	0.43	0.89	0.32	0.27	0.28	0.40	0.15	0.11	0.18	0.15	0.09	0.05	0.06	0.52
2004	1.92	1.01	0.93	0.59	0.80	0.50	0.36	0.38	0.78	0.27	0.23	0.24	0.34	0.13	0.09	0.15	0.12	0.08	0.05	0.49
2005	0.91	1.70	0.90	0.82	0.52	0.71	0.44	0.32	0.34	0.67	0.23	0.20	0.20	0.29	0.11	0.08	0.13	0.11	0.06	0.46
2006	1.11	0.81	1.51	0.80	0.73	0.46	0.63	0.39	0.28	0.29	0.57	0.20	0.17	0.17	0.24	0.09	0.07	0.11	0.09	0.44
2007	1.40	0.99	0.71	1.34	0.71	0.65	0.40	0.54	0.33	0.23	0.24	0.48	0.17	0.14	0.14	0.20	0.08	0.06	0.09	0.45
2008	1.20	1.24	0.87	0.63	1.19	0.62	0.57	0.35	0.46	0.28	0.19	0.20	0.40	0.14	0.12	0.12	0.17	0.06	0.05	0.44
2009	1.22	1.06	1.10	0.77	0.56	1.05	0.55	0.49	0.30	0.38	0.23	0.16	0.16	0.33	0.11	0.10	0.10	0.14	0.05	0.40
2010	1.43	1.08	0.94	0.98	0.69	0.50	0.93	0.48	0.43	0.26	0.32	0.19	0.13	0.14	0.27	0.09	0.08	0.08	0.11	0.38
2011	0.65	1.27	0.96	0.83	0.87	0.61	0.44	0.81	0.41	0.36	0.21	0.27	0.16	0.11	0.11	0.23	0.08	0.07	0.07	0.41
2012	0.53	0.57	1.12	0.85	0.74	0.77	0.54	0.38	0.69	0.35	0.30	0.17	0.22	0.13	0.09	0.09	0.18	0.06	0.05	0.39
2013	1.11	0.47	0.51	1.00	0.76	0.65	0.68	0.47	0.33	0.58	0.28	0.24	0.14	0.18	0.10	0.07	0.08	0.15	0.05	0.36
2014	1.04	0.98	0.41	0.45	0.88	0.67	0.58	0.60	0.41	0.28	0.47	0.23	0.19	0.11	0.14	0.08	0.06	0.06	0.12	0.33
2015	1.11	0.92	0.87	0.37	0.40	0.78	0.59	0.51	0.52	0.35	0.23	0.39	0.18	0.15	0.09	0.11	0.07	0.05	0.05	0.35
2016	1.13	0.99	0.82	0.77	0.33	0.35	0.69	0.52	0.45	0.45	0.30	0.20	0.32	0.15	0.12	0.07	0.09	0.05	0.04	0.32

Table 4.19—Model estimates of the number of female spawners (millions) 1964-2016.

year/age	6	7	8	9	10	11	12	13	14	15	16	17	18	19	20
1964	5.3	20.0	47.9	12.4	4.8	1.5	0.3	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.7
1965	5.1	16.4	44.6	89.8	19.0	5.4	1.3	0.2	0.1	0.1	0.1	0.0	0.0	0.0	0.5
1966	3.0	15.7	36.2	80.2	127.3	20.6	4.8	1.0	0.1	0.1	0.0	0.0	0.0	0.0	0.4
1967	5.6	9.1	34.3	63.8	107.3	120.6	15.1	3.0	0.6	0.1	0.0	0.0	0.0	0.0	0.2
1968	2.8	16.7	18.0	48.1	66.5	84.7	76.6	8.3	1.6	0.3	0.0	0.0	0.0	0.0	0.1

1969	2.6	8.6	35.6	29.7	59.0	59.9	60.6	47.5	4.9	0.9	0.2	0.0	0.0	0.0	0.1
1970	3.4	7.8	18.6	61.0	36.9	48.4	36.4	31.2	22.9	2.3	0.4	0.1	0.0	0.0	0.0
1971	3.6	10.4	17.6	36.2	103.1	47.5	42.7	21.9	15.6	10.7	1.0	0.2	0.0	0.0	0.0
1972	7.5	10.0	10.4	12.3	22.0	50.5	19.0	14.9	7.2	5.0	3.4	0.3	0.1	0.0	0.0
1973	11.4	22.9	22.5	20.3	20.9	27.7	44.0	13.9	10.3	4.8	3.3	2.2	0.2	0.0	0.0
1974	11.6	35.0	51.5	43.4	33.4	25.7	24.0	28.7	7.9	5.5	2.5	1.7	1.2	0.1	0.0
1975	15.2	35.7	78.5	99.5	72.5	43.4	26.0	20.5	22.9	6.1	4.2	1.9	1.3	0.9	0.1
1976	16.9	46.7	78.9	144.7	154.1	88.4	43.0	22.4	16.7	18.0	4.8	3.3	1.5	1.0	0.8
1977	13.2	51.8	104.0	149.6	234.1	194.3	89.1	37.4	18.3	13.2	14.2	3.7	2.5	1.2	1.4
1978	9.1	40.1	113.6	195.5	245.2	307.5	207.9	83.0	32.9	15.6	11.2	11.8	3.1	2.1	2.1
1979	12.2	27.6	87.4	207.7	305.6	304.3	310.1	182.5	68.8	26.3	12.4	8.8	9.3	2.4	3.3
1980	14.4	37.3	61.0	164.2	337.9	397.4	322.3	286.2	158.9	58.0	22.0	10.3	7.3	7.7	4.8
1981	9.5	43.9	82.8	116.0	271.2	444.0	422.2	296.7	247.8	133.0	48.1	18.1	8.5	6.0	10.2
1982	11.9	28.9	97.5	157.5	191.9	358.1	476.0	393.5	260.6	210.5	112.0	40.2	15.1	7.0	13.5
1983	7.8	36.4	64.1	185.2	260.8	255.0	388.0	449.6	350.8	224.9	180.2	95.1	34.1	12.8	17.4
1984	5.0	23.7	80.0	120.7	304.7	345.2	275.6	365.9	400.2	302.3	192.3	152.8	80.6	28.8	25.6
1985	9.5	15.1	52.5	150.8	196.5	395.4	364.3	253.3	317.3	335.9	251.7	158.8	126.1	66.4	44.8
1986	7.1	29.1	33.4	98.2	241.5	248.5	404.3	323.5	212.0	256.9	269.8	200.5	126.3	100.2	88.3
1987	20.4	21.6	63.8	61.4	154.4	303.4	254.5	361.1	272.7	173.0	208.0	216.6	160.8	101.2	151.0
1988	3.8	62.5	48.1	121.8	100.9	199.8	315.3	228.8	305.4	223.0	140.3	167.2	174.0	129.0	202.3

year/age	6	7	8	9	10	11	12	13	14	15	16	17	18	19	20
1989	16.8	11.6	139.1	90.8	194.9	125.9	200.6	274.8	188.0	242.7	175.8	109.7	130.7	135.8	258.5
1990	5.8	51.7	25.9	267.7	150.3	252.6	130.9	180.6	233.0	154.2	197.5	141.9	88.5	105.2	317.5
1991	4.4	17.8	115.9	50.0	448.4	201.0	274.9	124.1	161.6	201.9	132.5	168.4	120.9	75.2	359.5
1992	6.0	13.6	39.9	223.0	83.7	599.0	218.1	259.6	110.5	139.3	172.6	112.4	142.7	102.3	367.7
1993	8.2	18.5	30.3	75.8	366.1	109.1	632.1	200.0	224.4	92.4	115.6	142.0	92.4	117.1	385.8
1994	8.1	25.1	41.3	58.3	127.2	490.6	118.4	594.5	176.8	191.6	78.2	96.9	119.0	77.3	420.6
1995	4.0	25.0	55.8	77.9	95.1	165.5	519.5	109.2	517.6	149.0	160.2	64.9	80.3	98.4	411.9
1996	4.5	12.4	55.3	105.2	127.4	124.5	176.6	483.4	95.9	440.1	125.7	134.1	54.2	67.0	426.0
1997	9.9	13.8	27.4	104.6	173.0	167.1	132.8	163.9	422.8	81.2	369.3	104.7	111.5	45.0	409.4
1998	5.9	30.3	30.3	51.5	169.5	222.2	173.5	119.5	138.8	346.2	65.9	297.4	84.2	89.5	365.0
1999	5.0	18.1	67.0	57.3	84.7	223.1	238.3	162.2	105.4	118.5	293.2	55.4	249.5	70.5	380.8

2000	5.1	15.4	40.6	129.7	96.8	114.0	243.4	225.9	144.9	91.1	101.6	249.4	47.1	211.8	383.0
2001	12.5	15.5	34.5	78.4	218.3	129.5	123.3	228.6	199.9	124.1	77.4	85.6	209.9	39.5	499.8
2002	5.4	38.3	34.8	66.3	131.5	292.6	141.3	117.2	204.9	173.5	106.8	66.1	73.0	178.8	459.4
2003	4.5	16.6	86.1	67.4	111.8	176.3	318.4	133.7	104.6	177.0	148.7	90.8	56.1	61.9	541.2
2004	5.5	13.9	37.1	165.9	112.9	149.6	191.9	301.8	119.6	90.5	152.0	126.6	77.3	47.7	512.4
2005	7.8	16.9	31.1	71.4	278.1	151.4	163.3	182.4	270.6	103.8	77.9	129.8	108.1	65.8	477.2
2006	5.1	23.9	37.7	59.4	118.8	369.6	163.5	153.4	161.5	231.8	88.2	65.7	109.3	90.8	456.5
2007	7.1	15.3	52.0	70.3	97.1	156.5	398.1	153.6	136.1	138.7	197.6	74.5	55.5	92.2	461.5
2008	6.9	21.7	34.0	98.6	115.3	126.9	166.2	367.7	133.9	114.7	116.0	163.9	61.8	45.9	458.2
2009	11.5	21.0	47.9	63.9	159.7	148.7	133.0	151.7	316.8	111.6	94.9	95.2	134.4	50.6	412.6
2010	5.5	35.5	46.8	91.0	104.8	209.0	158.5	123.6	133.2	269.1	94.1	79.3	79.5	112.1	386.3
2011	6.7	16.8	79.0	89.1	149.8	137.4	222.5	146.9	108.1	112.6	225.9	78.3	66.0	66.0	413.8
2012	8.4	20.5	37.4	149.7	145.2	193.4	143.5	202.0	125.8	89.5	92.5	184.0	63.7	53.6	390.0
2013	7.2	25.8	45.6	71.0	245.5	188.6	202.9	130.7	173.3	104.3	73.6	75.5	150.0	51.9	361.1
2014	7.4	22.1	57.9	87.5	116.6	315.0	194.1	180.8	109.7	140.7	84.0	58.8	60.3	119.6	329.2
2015	8.6	22.6	49.5	111.0	144.8	151.3	326.3	173.1	151.5	88.8	112.9	66.9	46.8	47.8	356.1
2016	3.9	26.5	50.7	95.7	187.0	193.3	161.4	297.1	146.6	123.1	71.3	89.8	53.1	37.1	320.2

Table 4.20. Model estimates of yellowfin sole age 5 recruitment (millions) from the 2015 and 2016 stock assessments.

Year class	2015 assessment	2016 assessment
1964	733	694
1965	734	742
1966	1,520	1,533
1967	2,316	2,331
1968	2,361	2,380
1969	3,100	3,126
1970	3,432	3,461
1971	2,691	2,716
1972	1,862	1,878
1973	2,496	2,519
1974	2,932	2,958
1975	1,929	1,947
1976	2,424	2,447
1977	1,584	1,599
1978	1,008	1,018
1979	1,937	1,956
1980	1,435	1,450
1981	4,138	4,182
1982	765	774
1983	3,407	3,448
1984	1,173	1,187
1985	897	909
1986	1,219	1,236
1987	1,659	1,684
1988	1,646	1,673
1989	817	831
1990	908	925
1991	2,004	2,045
1992	1,191	1,217
1993	1,002	1,027
1994	1,010	1,037
1995	2,489	2,560
1996	1,076	1,104
1997	910	926
1998	1,103	1,131
1999	1,553	1,603
2000	1,013	1,040
2001	1,419	1,457
2002	1,401	1,410
2003	2,431	2,373
2004	1,201	1,124
2005	1,352	1,374

2006	1,726	1,732
2007	1,414	1,478
2008	1,529	1,510
2009	1,662	1,767
2010		802

Table 4.21—Selected parameter estimates and their standard deviation from the preferred stock assessment model.

	parameter	value	std dev		parameter	value	std dev
	alpha (q-temp model)	0.05	0.04	1976	total biomass	2,035,300	52,815
	beta (q-temp model)	0.10	0.01	1977	total biomass	2,334,700	59,338
	mean_log_rec	0.80	0.09	1978	total biomass	2,619,100	65,422
	mean sel_slope_fsh (females)	1.15	0.08	1979	total biomass	2,769,500	70,420
	mean sel50_fsh (females)	8.79	0.26	1980	total biomass	2,942,300	74,915
	mean sel_slope_fsh_males	1.36	0.09	1981	total biomass	3,100,200	78,720
	mean sel50_fsh_males	8.29	0.24	1982	total biomass	3,205,400	80,444
	sel_slope_srv (females)	1.64	0.09	1983	total biomass	3,180,000	81,006
	sel50_srv (females)	5.02	0.07	1984	total biomass	3,389,700	86,035
	sel_slope_srv_males	-0.06	0.08	1985	total biomass	3,390,900	88,628
	sel50_srv_males	0.02	0.02	1986	total biomass	3,104,600	85,722
	Ricker SR logalpha	-4.31	0.52	1987	total biomass	3,054,400	87,620
	Ricker SR logbeta	-6.30	0.32	1988	total biomass	2,951,700	87,362
	Fmsy	0.12	0.04	1989	total biomass	2,990,500	91,313
	log (Fmsy)	-2.12	0.31	1990	total biomass	2,854,500	89,646
	ABC_biomass 2016	2,284,800	130,620	1991	total biomass	2,960,100	92,394
	ABC_biomass 2017	2,197,000	141,510	1992	total biomass	3,146,400	97,202
	msy	412,590	149,240	1993	total biomass	3,162,900	99,584
	Bmsy	423,740	79,683	1994	total biomass	3,189,700	100,950
1954	total biomass	2,493,500	153,250	1995	total biomass	2,968,100	97,374
1955	total biomass	2,459,800	138,050	1996	total biomass	2,877,100	96,036
1956	total biomass	2,416,100	121,140	1997	total biomass	2,876,600	97,574
1957	total biomass	2,362,300	103,180	1998	total biomass	2,608,200	92,775
1958	total biomass	2,319,800	84,973	1999	total biomass	2,430,600	88,900
1959	total biomass	2,272,000	67,827	2000	total biomass	2,468,300	89,658
1960	total biomass	2,092,700	52,949	2001	total biomass	2,393,000	88,147
1961	total biomass	1,651,000	38,419	2002	total biomass	2,426,500	89,025
1962	total biomass	1,142,900	24,658	2003	total biomass	2,613,500	95,104
1963	total biomass	804,730	16,254	2004	total biomass	2,801,000	101,570
1964	total biomass	846,580	16,374	2005	total biomass	2,901,300	105,270
1965	total biomass	840,790	16,174	2006	total biomass	2,882,200	106,240
1966	total biomass	887,570	16,956	2007	total biomass	2,890,300	108,350
1967	total biomass	874,190	17,448	2008	total biomass	2,765,700	106,770
1968	total biomass	797,970	17,304	2009	total biomass	2,601,400	104,520
1969	total biomass	828,470	18,982	2010	total biomass	2,665,200	107,490
1970	total biomass	794,050	20,425	2011	total biomass	2,695,000	110,870
1971	total biomass	850,500	23,369	2012	total biomass	2,687,200	114,340

1972	total biomass	910,620	27,049	2013	total biomass	2,641,100	117,110
1973	total biomass	1,156,000	32,580	2014	total biomass	2,442,600	114,940
1974	total biomass	1,397,000	38,503	2015	total biomass	2,443,400	122,080
1975	total biomass	1,739,200	45,993	2016	total biomass	2,448,200	130,120

Table 4.22. Projections of yellowfin sole female spawning biomass (1,000s t), catch (1,000s t) and full selection fishing mortality rate for seven future harvest scenarios.

Scenarios 1 and 2				Scenario 4			
Maximum Tier 3 ABC harvest permissible				1/2 Maximum Tier 3 ABC harvest permissible			
Female				Female			
Year	spawning biomass	catch	F	Year	spawning biomass	catch	F
2016	775.899	130.497	0.08	2016	775.899	130.497	0.08
2017	776.688	193.242	0.11	2017	791.193	96.6257	0.06
2018	755.529	185.717	0.11	2018	811.759	97.8799	0.06
2019	731.662	175.419	0.11	2019	826.833	96.8635	0.06
2020	695.451	163.333	0.11	2020	824.305	94.0948	0.06
2021	653.668	155.175	0.11	2021	809.326	92.5486	0.06
2022	621.874	145.477	0.11	2022	797.305	92.827	0.06
2023	606.268	141.295	0.11	2023	795.681	93.9998	0.06
2024	600.516	141.39	0.10	2024	799.836	95.8282	0.06
2025	602.543	144.195	0.10	2025	810.348	97.8358	0.06
2026	609.438	148.72	0.11	2026	825.078	100.183	0.06
2027	618.376	152.687	0.11	2027	841.682	102.396	0.06
2028	628.871	156.159	0.11	2028	860.7	104.324	0.06
2029	638.927	159.08	0.11	2029	879.673	106.111	0.06
Scenario 3				Scenario 5			
Harvest at average F over the past 5 years				No fishing			
Female				Female			
Year	spawning biomass	catch	F	Year	spawning biomass	catch	F
2016	775.899	130.497	0.08	2016	775.899	130.497	0.08
2017	784.461	141.866	0.08	2017	805.239	0	0
2018	786.857	129.812	0.08	2018	869.301	0	0
2019	787.535	126.411	0.08	2019	929.637	0	0
2020	772.242	121.004	0.08	2020	970.871	0	0
2021	746.783	117.583	0.08	2021	995.493	0	0
2022	726.358	116.872	0.08	2022	1018.59	0	0
2023	717.545	117.492	0.08	2023	1049.43	0	0
2024	715.68	119.119	0.08	2024	1082.95	0	0
2025	720.716	121.11	0.08	2025	1121.4	0	0
2026	730.527	123.623	0.08	2026	1162.29	0	0
2027	742.714	126.011	0.08	2027	1203.18	0	0
2028	757.329	128.087	0.08	2028	1246.29	0	0
2029	772.046	130.002	0.08	2029	1288.69	0	0

Table 4.22—continued.

Scenario 6				Scenario 7			
Determination of whether yellowfin sole are currently overfished				Determination of whether the stock is approaching an overfished condition			
			B35=548.000				B35=548
		Female				Female	
Year	spawning biomass	catch	F	Year	spawning biomass	catch	F
2016	775.899	130.497	0.08	2016	775.899	130.497	0.08
2017	771.194	229.012	0.14	2017	776.687	193.254	0.11
2018	735.013	215.954	0.14	2018	755.524	185.716	0.11
2019	698.174	200.396	0.14	2019	726.436	207.835	0.14
2020	651.712	183.659	0.14	2020	676.279	189.808	0.14
2021	604.532	160.74	0.13	2021	624.42	170.833	0.13
2022	573.035	148.831	0.12	2022	587.405	155.702	0.12
2023	558.741	144.877	0.12	2023	568.953	149.607	0.12
2024	554.514	145.674	0.11	2024	561.545	148.897	0.12
2025	557.745	149.359	0.12	2025	562.449	151.514	0.12
2026	565.436	155.153	0.12	2026	568.473	156.548	0.12
2027	574.558	160.741	0.12	2027	576.439	161.585	0.12
2028	584.34	165.551	0.12	2028	585.45	166.032	0.12
2029	593.012	169.449	0.12	2029	593.627	169.702	0.12

Table 4.23 (continued).

	2010	2011	2012	2013	2014	2015
Pollock	3,749	8,685	11,226	20,246	24,712	21,282
Arrowtooth Flounder	868	2,338	995	2,012	2,216	1,686
Pacific Cod	8,649	16,300	19,230	24,382	15,217	12,169
Groundfish, General	3,048					
Rock Sole	9,030	9,762	8,959	7,737	7,031	9,773
Flathead Sole	1,895	3,236	2,109	4,191	3,999	3,337
Sablefish		<1			<1	<1
Atka Mackerel		<1	<1	<1	<1	<1
Pacific ocean Perch		<1		17	<1	<1
Rex Sole						
Other flatfish			1,201	388	2,887	1,041
Squid		<1				
Dover Sole						
Thornyhead						
Shortraker/Rougheye						
Butter Sole						
Starry Flounder						
Northern Rockfish			<1			
Dusky Rockfish						
Yellowfin Sole	90,008	136,905	133,719	147,777	139,480	107,955
English Sole						
Unsp.demersal rockfish						
Greenland Turbot				335		42
Alaska Plaice		6	6		56	
Sculpin, General						1,083
Skate, General				16,006		1,073
Sharpchin Rockfish	10,749	18,340	13,613		14,347	
Bocaccio		1,808	1,924	1,922	1,261	
Rockfish, General		1,969	2,270	2,686	1,969	
Octopus						
Smelt, general						
Chilipepper				<1		
Eels			1.3			
Lingcod						
Jellyfish (unspecified)						
Snails						
Sea cucumber						
Korean horsehair crab						
Kamchatka flounder			110	147		427
Sharks						1

Table 4-24. Estimated non-target species catch (t) in the yellowfin sole fishery, 2003-2015 (PSC not included).

Row Labels	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012	2013	2014	2015
Benthic urochordata	1671.6	1701.5	674.5	520.1	114.5	347.6	204.7	156.0	133.0	140.8	197.4	116.1	230.1
Birds													
Bivalves	1.5	1.1	1.3	0.3	0.5	1.5	1.3	1.8	1.7	0.7	1.2	0.9	1.4
Brittle star unidentified	34.3	32.3	28.7	20.0	7.6	19.0	5.2	4.2	14.0	13.1	5.9	11.6	2.9
Capelin	0.0	4.5	0.0	0.1	0.3	0.2	0.3	0.7	3.8	2.3	0.2	1.3	1.8
Corals Bryozoans	0.2	0.0	1.2	9.4	0.2	8.3	0.3	0.5	0.9	0.7	3.0	0.8	0.1
Eelpouts	19.1	12.3	7.7	4.5	2.3	5.6	5.2	5.1	29.3	14.3	51.6	69.8	21.1
Eulachon	0.0	0.3	0.0	0.1	5.1	0.0	0.1	0.1	0.5	0.1	0.0	0.7	0.2
Greenlings	0.6	0.7	0.3	0.7	0.5	0.2	0.0	0.1	0.0	0.1		0.0	0.2
Grenadier					0.3		0.4						
Gunnels					0.0						0.0		0.0
Hermit crab unidentified	87.9	52.0	83.6	26.9	35.8	36.6	15.4	17.0	15.9	9.9	6.3	8.6	4.1
Invertebrate unidentified	556.5	625.8	421.2	177.2	40.0	70.4	30.6	25.9	65.4	121.3	25.2	44.4	6.2
Misc crabs	14.4	21.6	11.9	10.6	28.0	14.1	11.0	11.7	20.2	18.2	39.7	19.8	18.8
Misc crustaceans	0.0	0.2	0.2	2.3	1.4	0.7	1.3	0.9	0.5	0.4	0.6	0.2	0.6
Misc fish	95.8	91.2	66.2	42.5	71.2	66.3	48.8	29.2	40.0	86.2	48.2	69.3	34.8
Misc inverts (worms etc)	0.0	0.1	0.0	0.0	0.0	0.2	0.2	0.1	0.2	0.1	0.2	0.1	0.0
Other osmerids	4.2	4.3	0.5	0.6	35.8	9.8	0.8	2.8	2.1	4.7	1.0	9.2	4.8
Pacific Sand lance	0.0	0.2	0.1	0.0	0.0	0.0	0.0	0.0	0.4	0.2	0.0	0.0	0.1
Pacific Sandfish								0.0	0.0	0.0	0.0	0.1	0.1
Pandalid shrimp	0.2	0.9	0.1	0.8	0.1	0.3	0.5	0.7	2.3	0.6	2.1	1.0	0.2
Polychaete unidentified	0.0	0.1	0.0	0.4	0.1	0.2	0.1	0.1	0.2	0.1	2.0	0.1	0.1
Scypho jellies	111.9	298.7	115.6	46.8	42.4	145.8	223.2	152.4	307.2	179.3	463.2	805.0	352.0
Sea anemone unidentified	6.3	6.2	2.6	4.9	8.8	24.8	25.5	20.5	14.7	6.2	23.4	5.7	4.2
Sea pens whips	0.0	0.0	0.2	0.0	0.0	0.3	0.2	0.6	0.0	0.1	0.1	0.0	0.0
Sea star	1941.3	1868.0	1611.8	1308.6	1462.0	1829.0	683.7	795.6	1674.0	1732.7	1372.4	2106.5	1816.7
Snails	118.3	191.1	69.7	141.5	95.3	139.6	57.7	57.7	74.7	33.7	46.4	33.7	30.0
Sponge unidentified	11.3	6.8	12.2	3.1	0.4	6.8	69.4	16.5	15.1	14.1	16.6	1.5	2.2
Stichaeidae	0.1	0.0		0.0	0.8	0.2	0.0	0.2	0.4	0.1	0.1	0.4	0.5
Surf smelt						0.0							
urchins dollars cucumbers	2.3	0.3	2.5	0.8	3.4	4.9	7.5	1.3	1.0	0.7	0.8	0.5	0.4
Grand Total	4678	4920	3112	2322	1957	2732	1393	1302	2417	2381	2308	3307	2534

Table 4.25--Yellowfin sole TAC and ABC levels, 1980- 2016.

Year	TAC	ABC	Total catch
1980	117,000	169,000	87,391
1981	117,000	214,500	97,301
1982	117,000	214,500	95,712
1983	117,000	214,500	108,385
1984	230,000	310,000	159,526
1985	229,900	310,000	227,107
1986	209,500	230,000	208,597
1987	187,000	187,000	181,428
1988	254,000	254,000	223,156
1989	182,675	241,000	153,170
1990	207,650	278,900	80,584
1991	135,000	250,600	95,000
1992	235,000	372,000	159,038
1993	220,000	238,000	106,101
1994	150,325	230,000	144,544
1995	190,000	277,000	124,740
1996	200,000	278,000	129,659
1997	230,000	233,000	181,389
1998	220,000	220,000	101,201
1999	207,980	212,000	67,320
2000	123,262	191,000	83,850
2001	113,000	176,000	63,395
2002	86,000	115,000	72,999
2003	83,750	114,000	74,418
2004	86,075	114,000	69,046
2005	90,686	124,000	94,683
2006	95,701	121,000	99,068
2007	136,000	225,000	121,029
2008	225,000	248,000	148,894
2009	210,000	210,000	107,528
2010	219,000	219,000	118,624
2011	196,000	239,000	151,164
2012	202,000	203,000	147,183
2013	198,000	206,000	164,944
2014	184,000	239,800	156,778
2015	149,000	248,800	126,933
2016	144,000	211,700	130,500

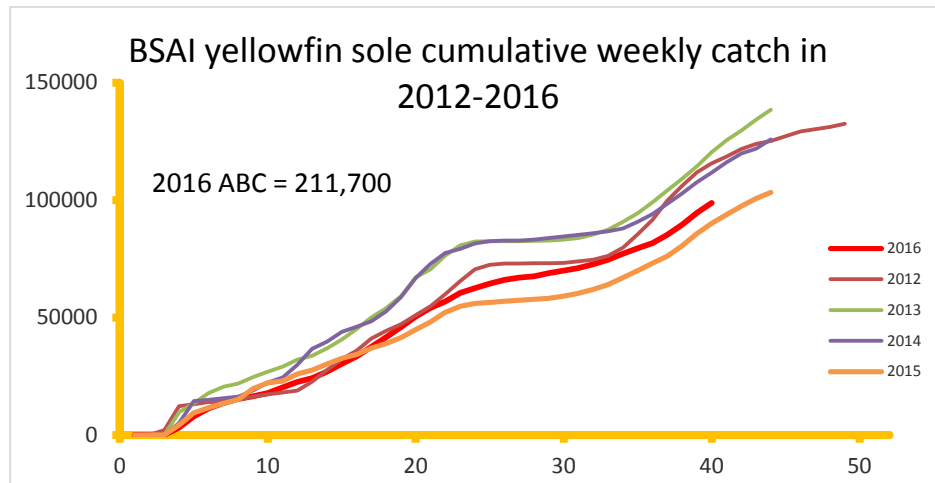
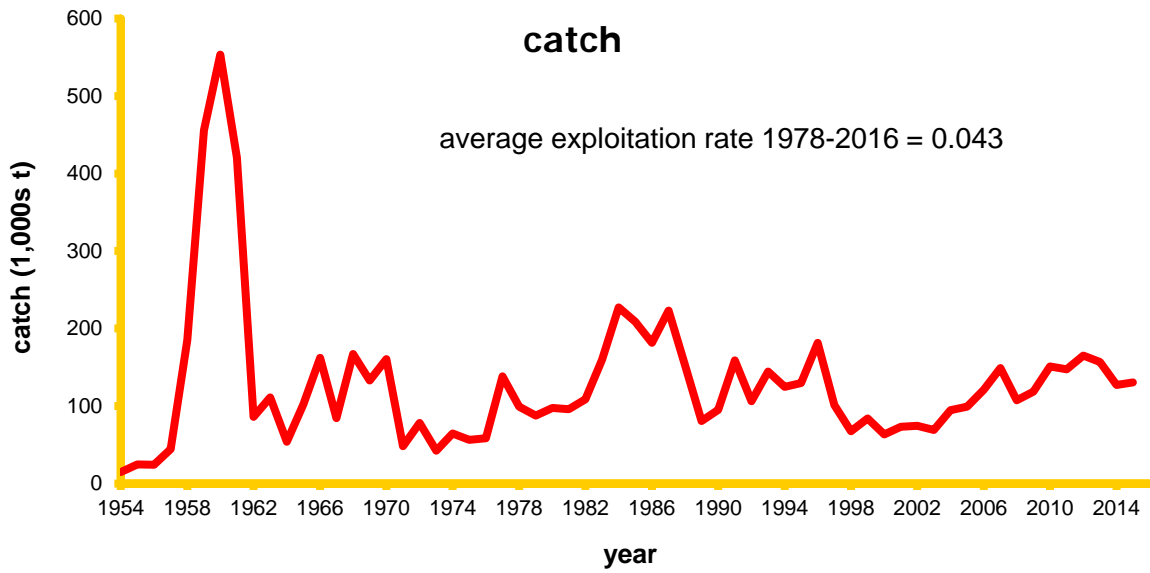


Figure 4.1—Yellowfin sole annual catch (1,000s t) in the Eastern Bering Sea from 1954-2016 (top panel) and catch by week (non CDQ) from 2010 – September 2016 (bottom panel).

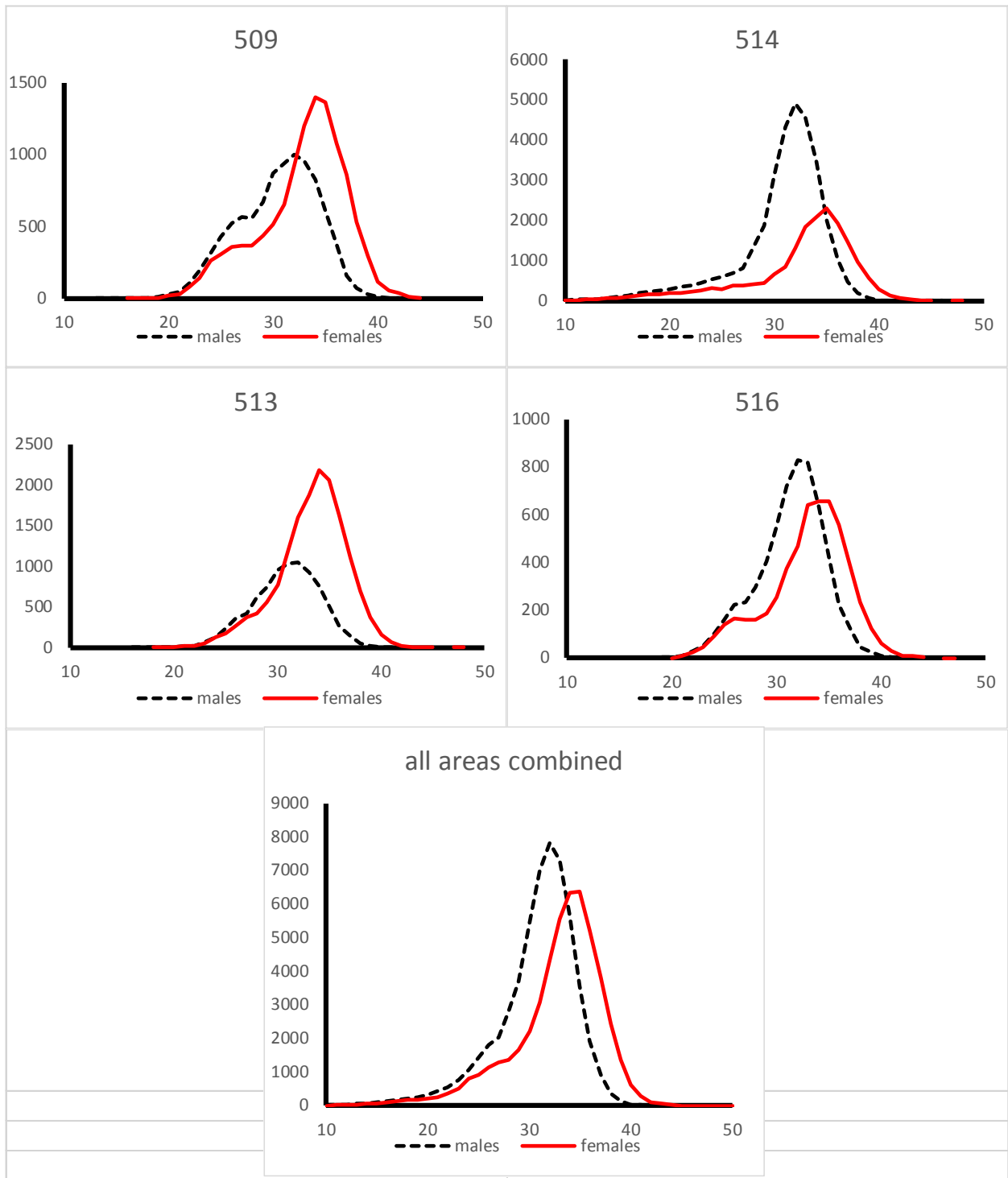
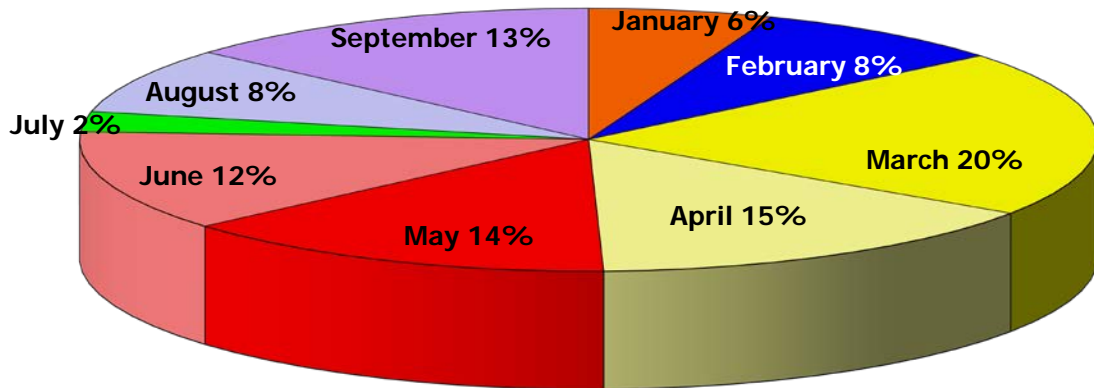


Figure 4.2--Size composition of the yellowfin sole catch in 2016 (through mid-September), by subarea and total.

yellowfin sole catch by month in 2016 through September 9



yellowfin sole catch by area in 2016 (through September 9)

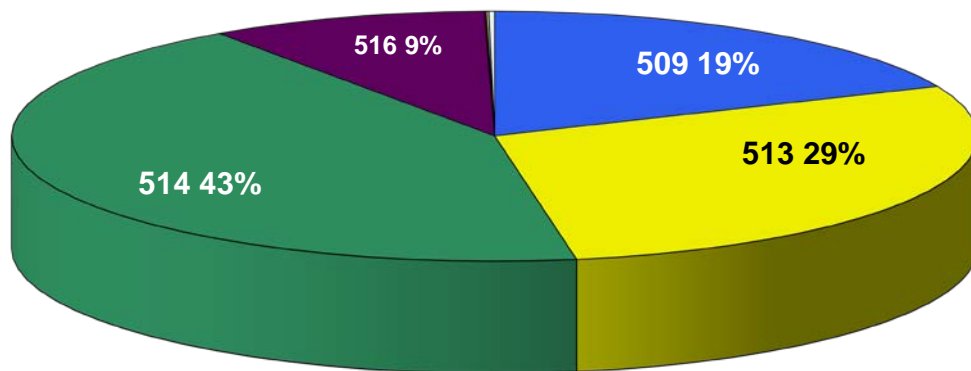
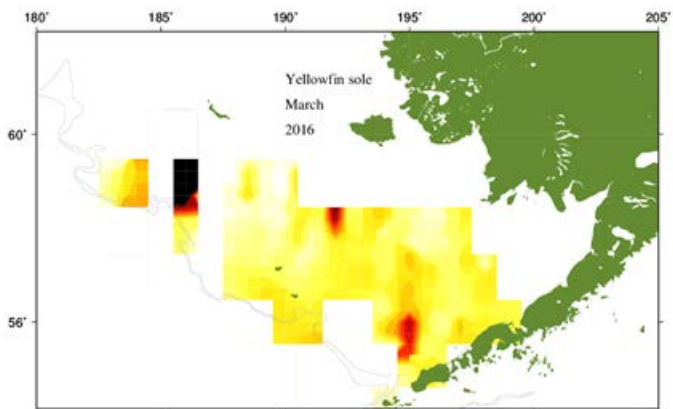
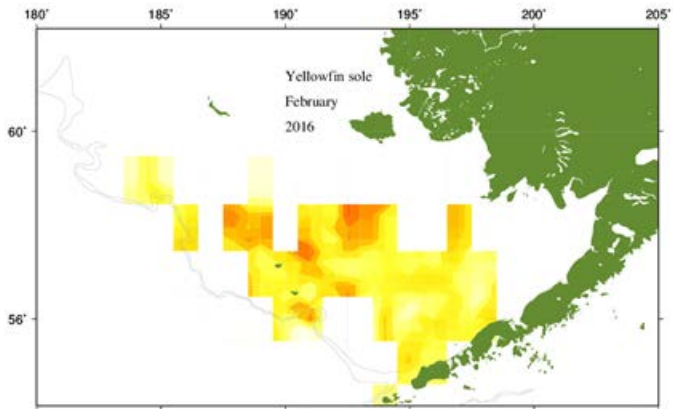
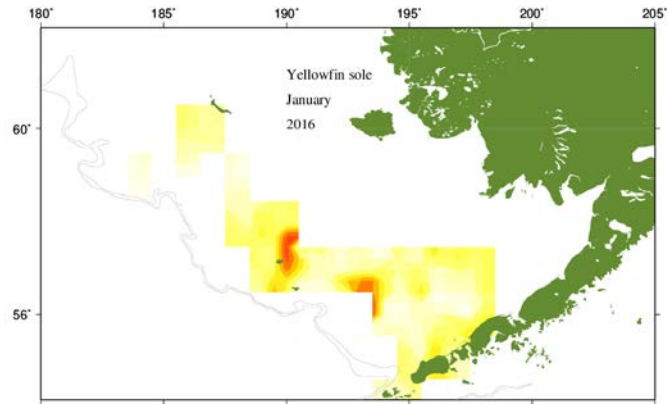
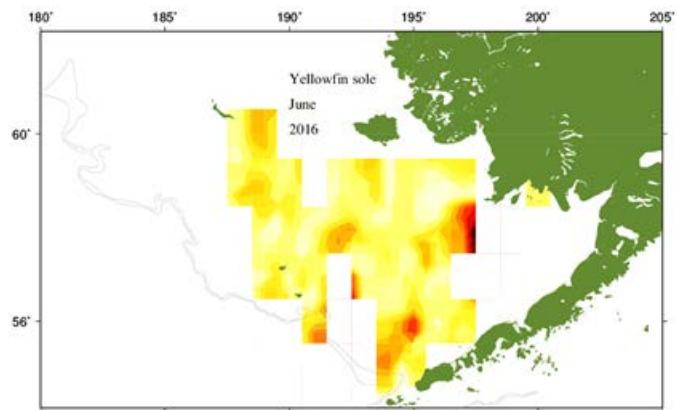
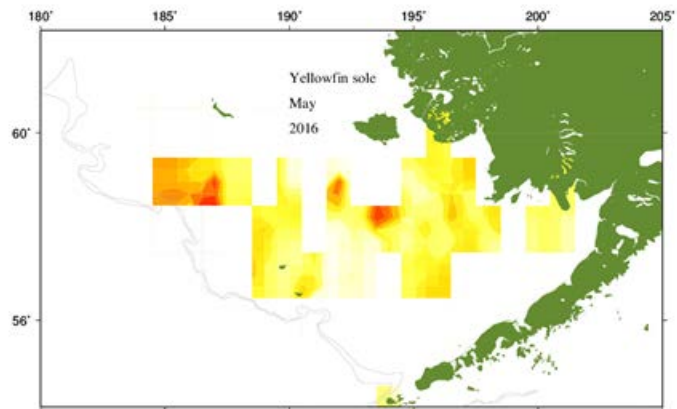
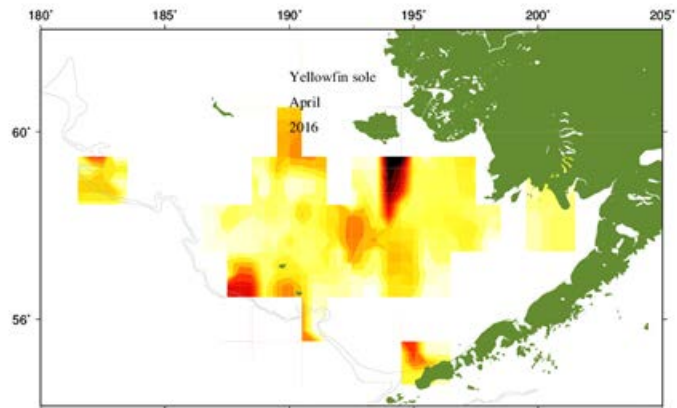


Figure 4.3 Yellowfin sole catch by month and area in the Eastern Bering Sea in 2015.





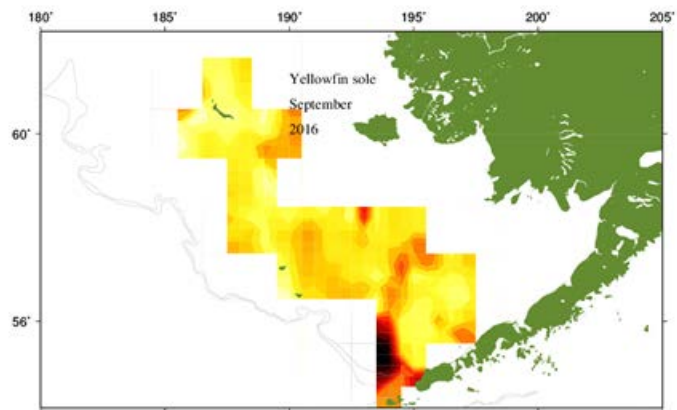
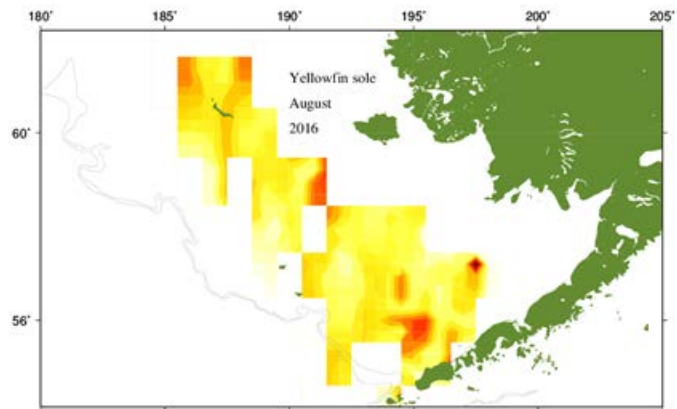
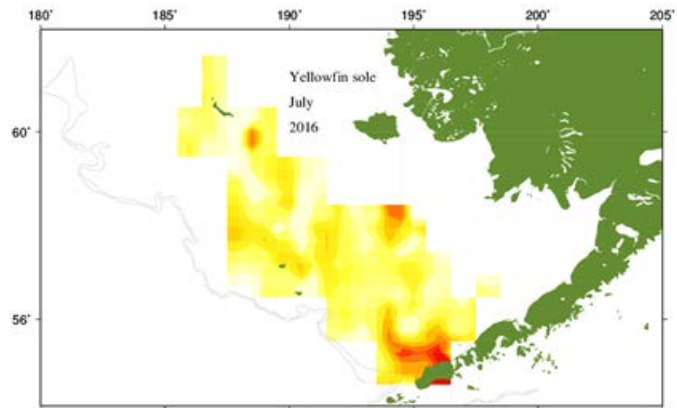


Figure 4.4— (Fishery locations by month).

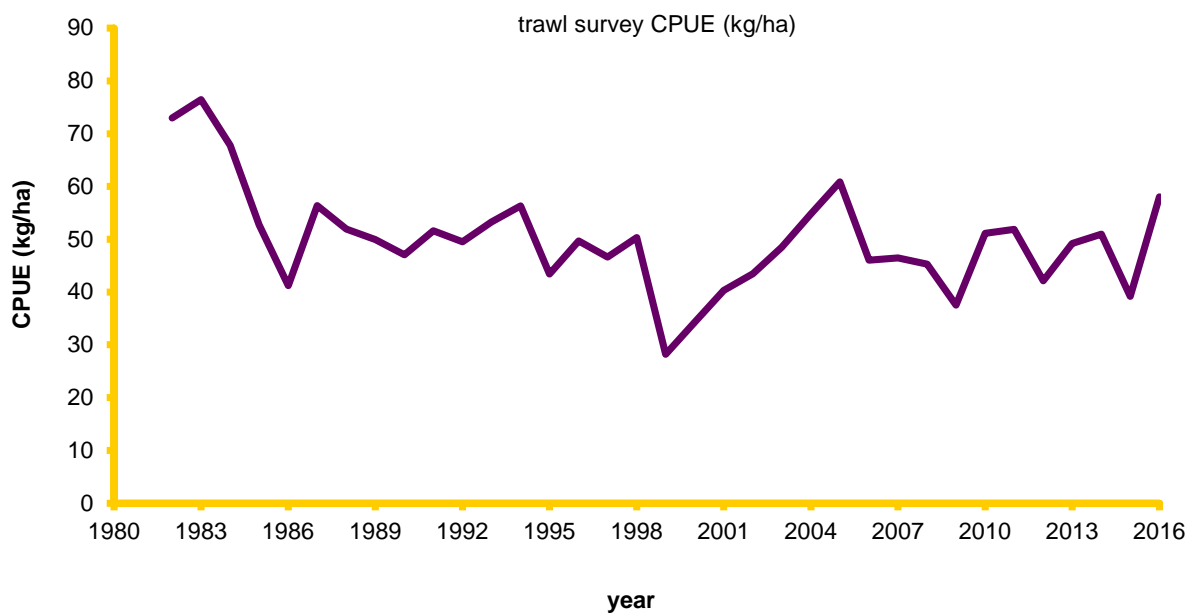


Figure 4.5.--Yellowfin sole CPUE (catch per unit effort in kg/ha) from the annual Bering Sea shelf trawl surveys, 1982-2015.

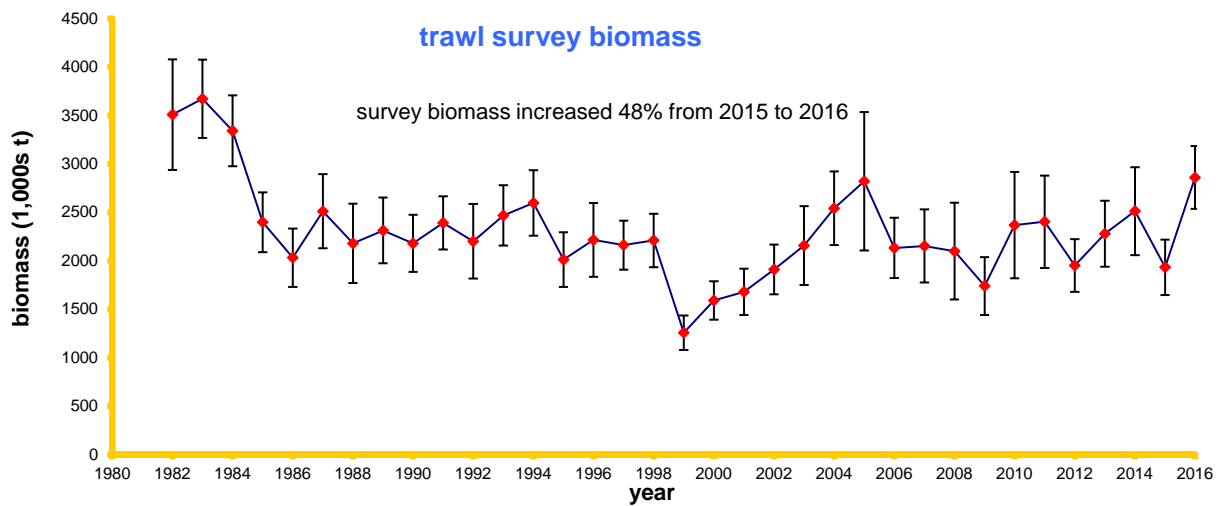


Figure 4.6.--Annual bottom trawl survey biomass point-estimates and 95% confidence intervals for yellowfin sole, 1982-2016.

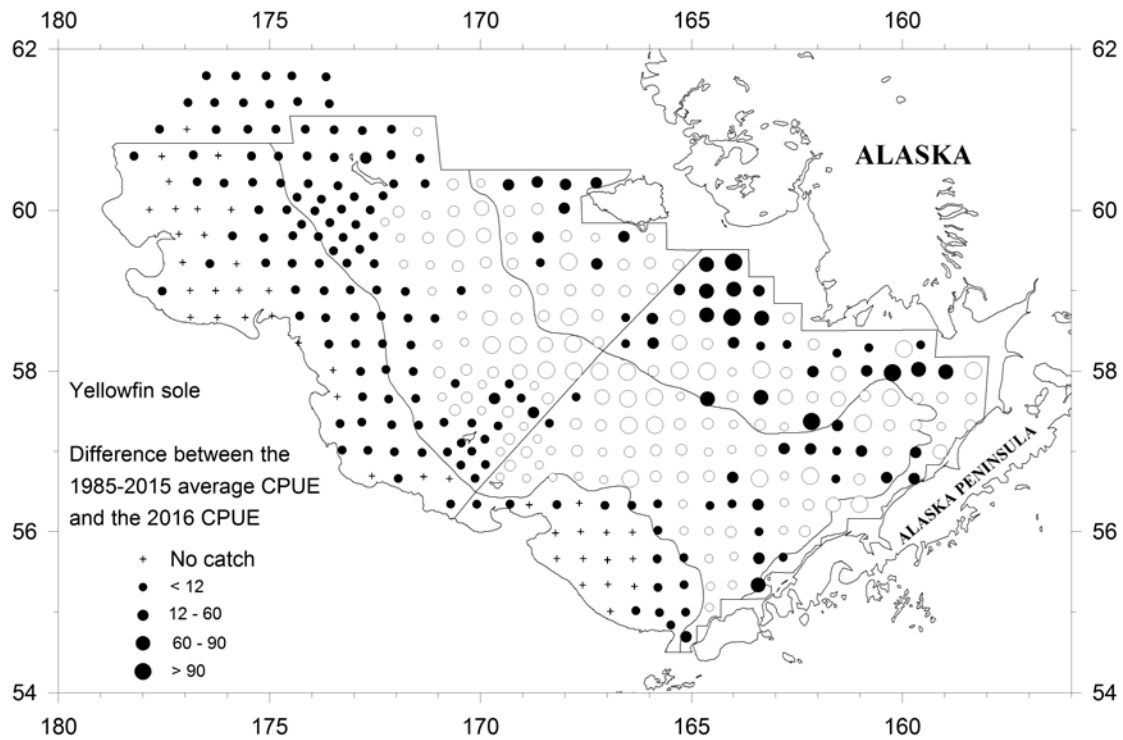


Figure 4.7.--Difference between the 1985-2015 average trawl survey CPUE for yellowfin sole and the 2016 survey CPUE. Open circles indicate that the magnitude of the catch was greater in 2016 than the long-term average, closed circles indicate the catch was greater in the long-term average than in 2016.

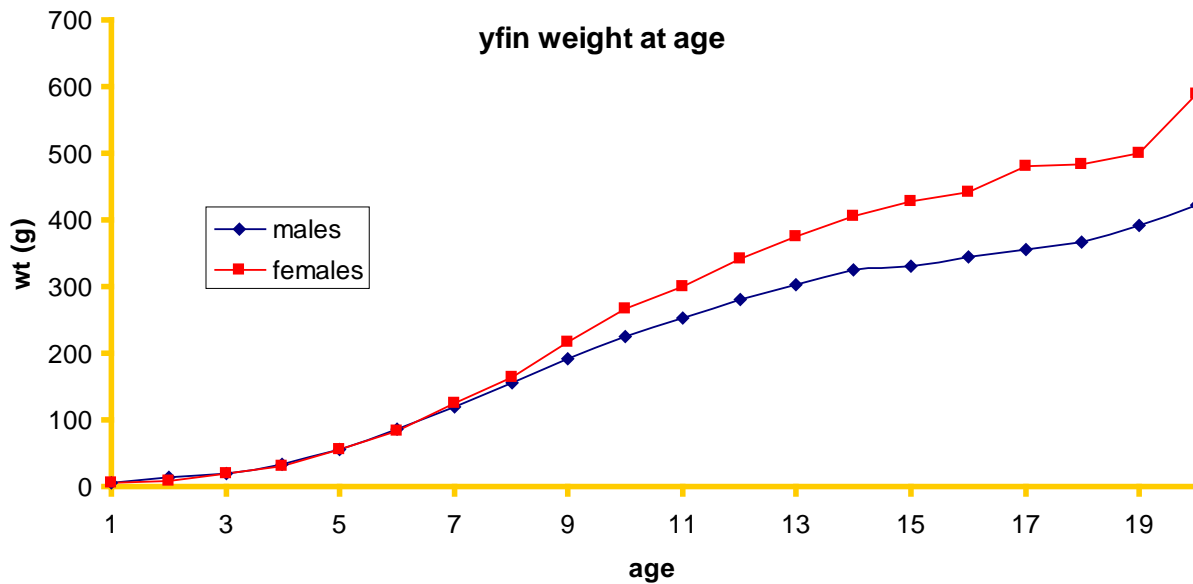


Figure 4.8--Estimates of average yellowfin sole weight-at-age (g) from trawl survey observations.

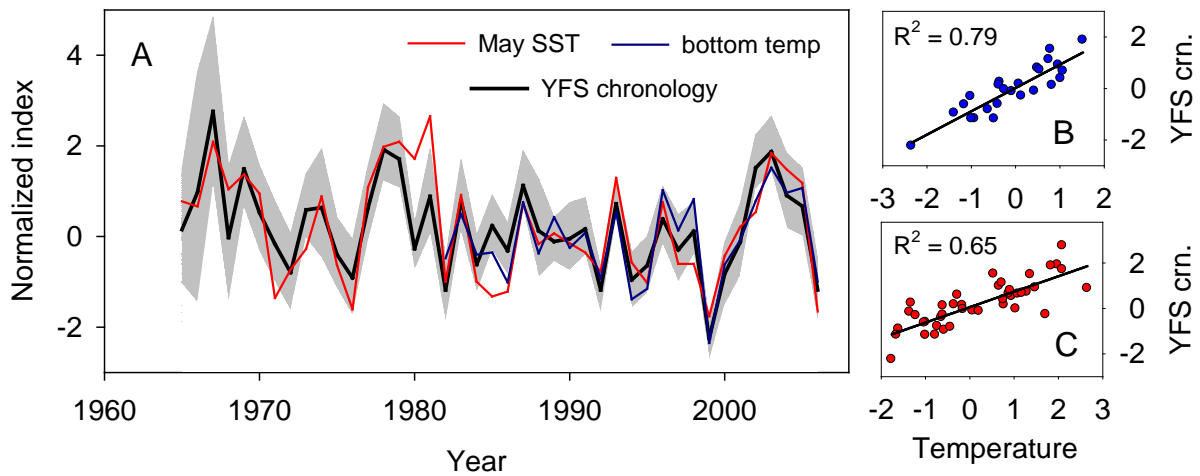


Figure 4.9--Master chronology for yellowfin sole and time series of mean summer bottom temperature and May sea surface temperature for the southeastern Bering Sea (Panel A). All data re normalized to a mean of 0 and standard deviation of 1. Correlations of chronologies with bottom temperature and sea surface temperature are shown in panels B and C, respectively. From Matta et al. 2010.

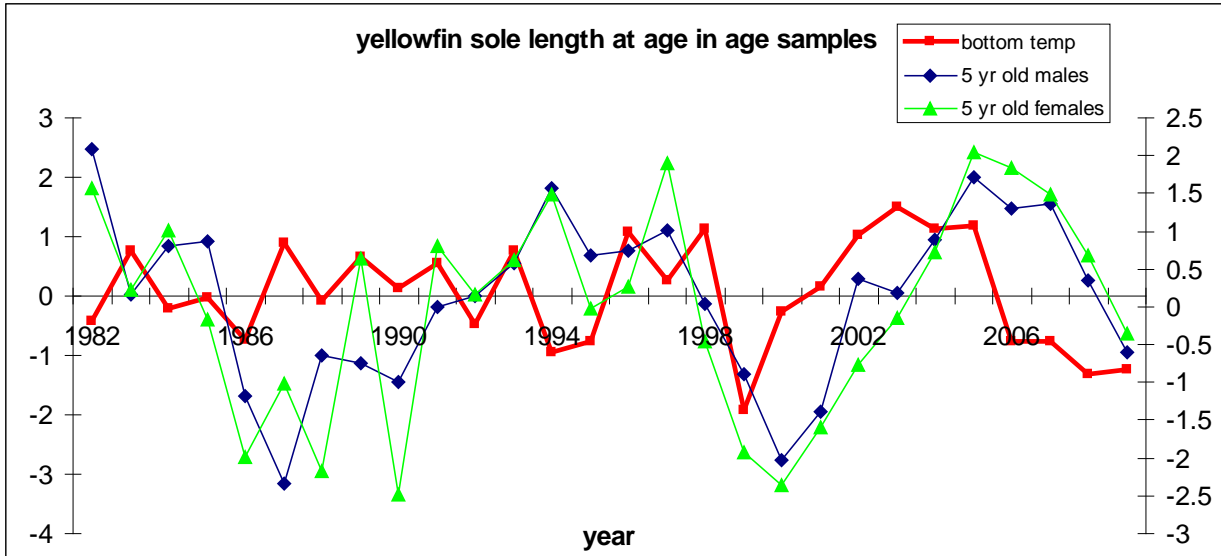


Figure 4.10—Yellowfin sole length-at-age anomalies, for males and females, and bottom temperature anomalies. Correspondence in these residuals is apparent with a 2-3 year lag effect from the mid-1990s to 2009. Late 1980s and early 1990s pattern may be a density-dependent response in growth from the large 1981 and 1983 year-classes.

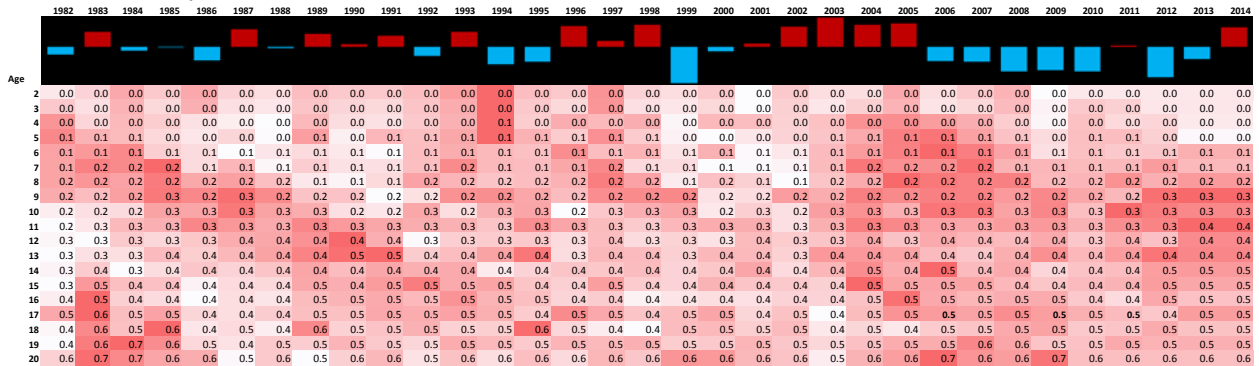


Figure 4.11—Results show the temperature anomalies (second row at top as bars) and observed values by age and year. Shadings within the matrix reflects relative weight-at-age (within a row) with darker red being heavier than average.

temperature-catchability model result

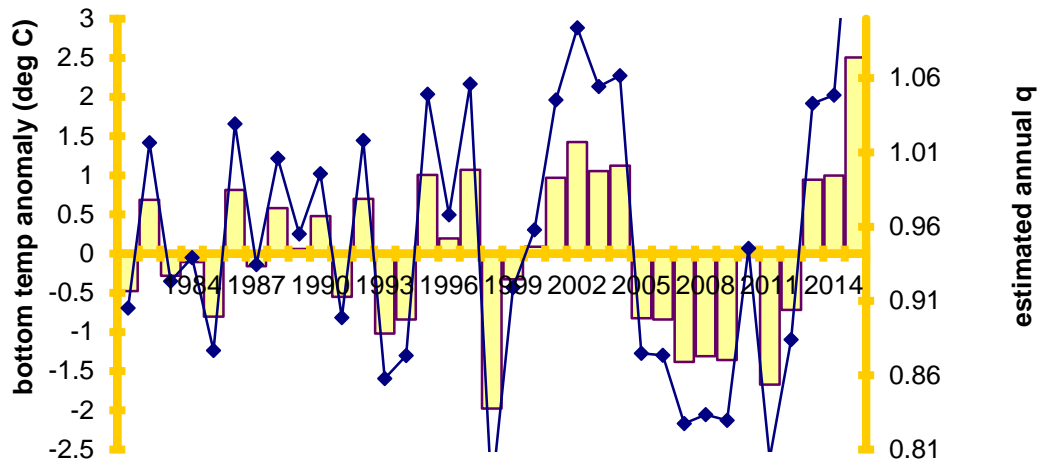


Figure 4.12.--Average bottom water temperature from stations less than or equal to 100 m in the Bering Sea trawl survey (bars) and the stock assessment model estimate of q for each year 1982-2016.

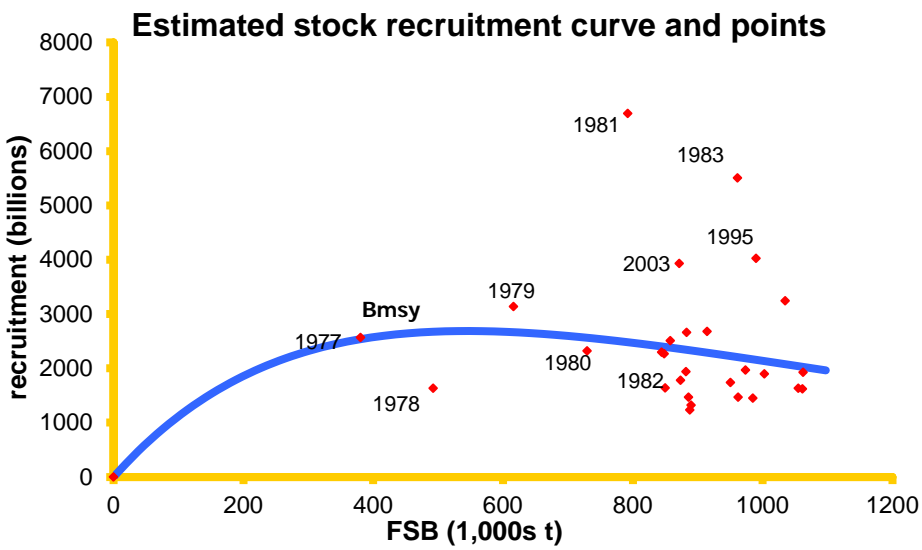
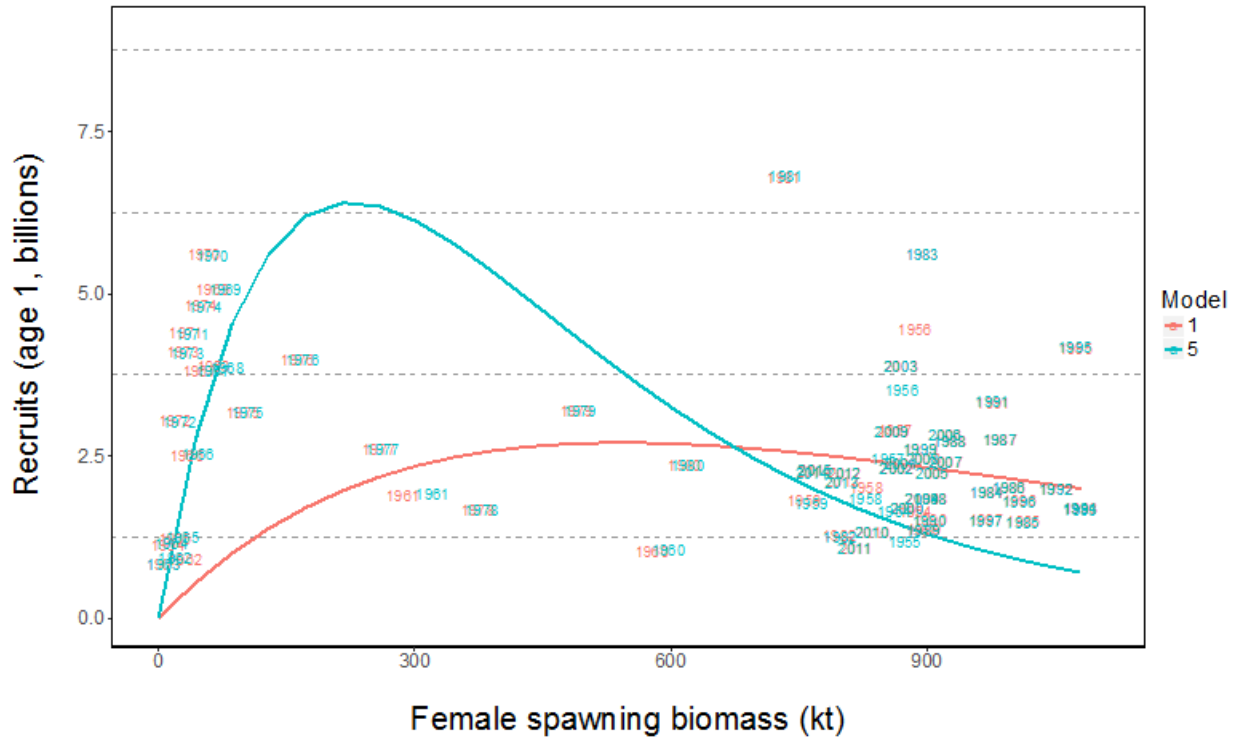


Figure 4.13--Fit of the Ricker (1958) stock recruitment model to two distinct stock recruitment time-series data sets (top panel), and the fit to the assessment preferred model (model B, lower panel).

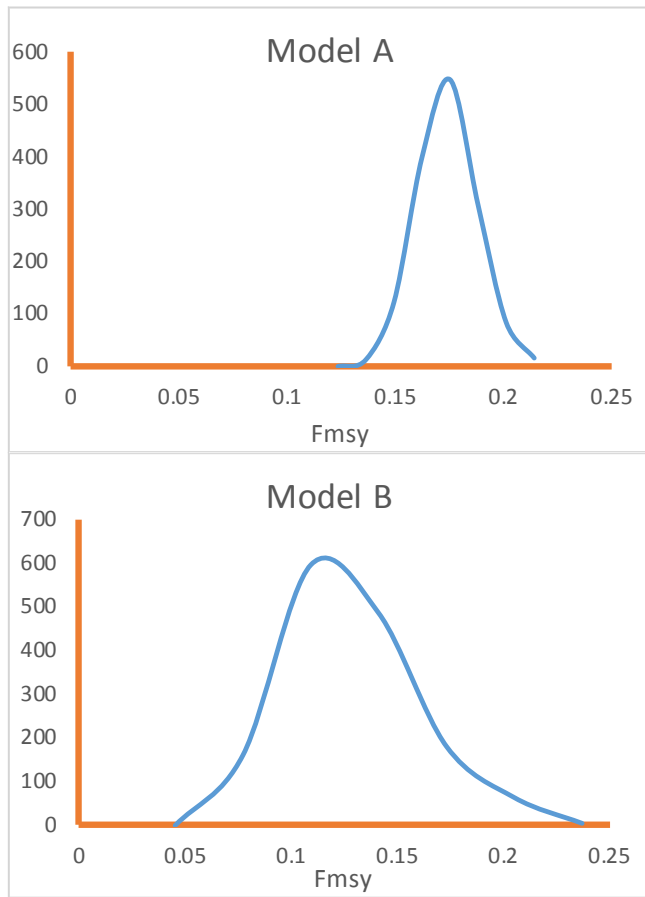


Figure 4.14--Posterior distributions of F_{msy} for the two models considered in the stock productivity analysis.

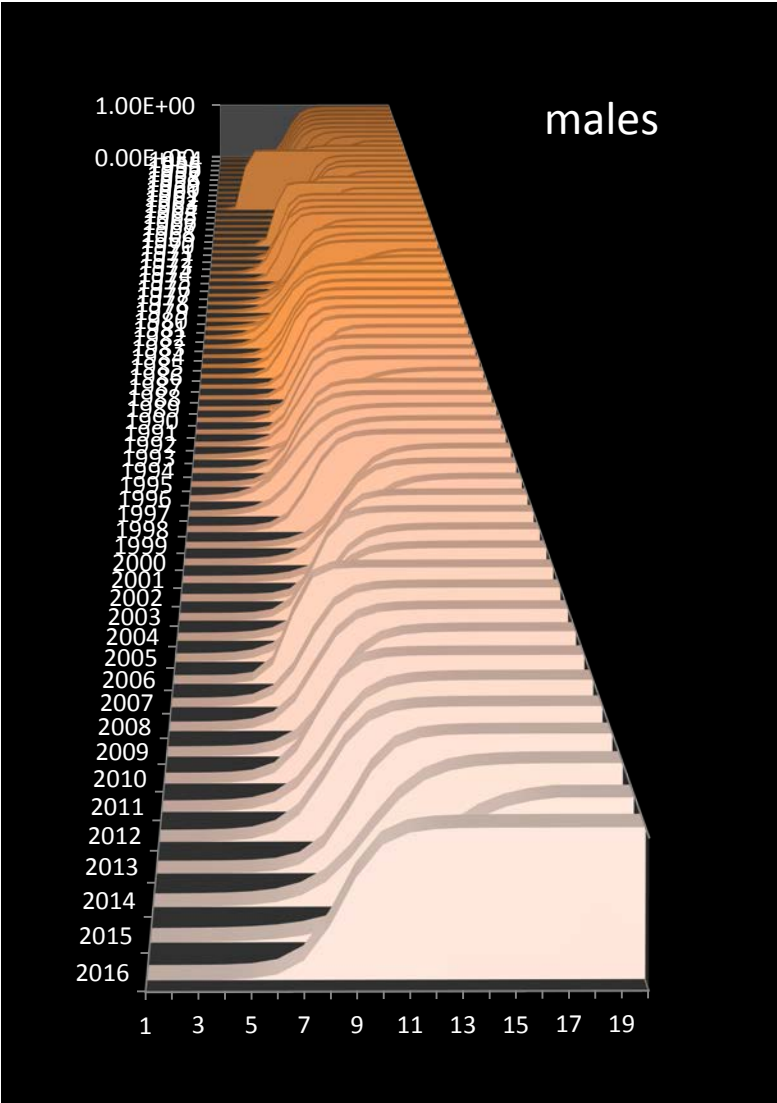


Figure 4.15a--Estimated male fishery selectivity by age and year.

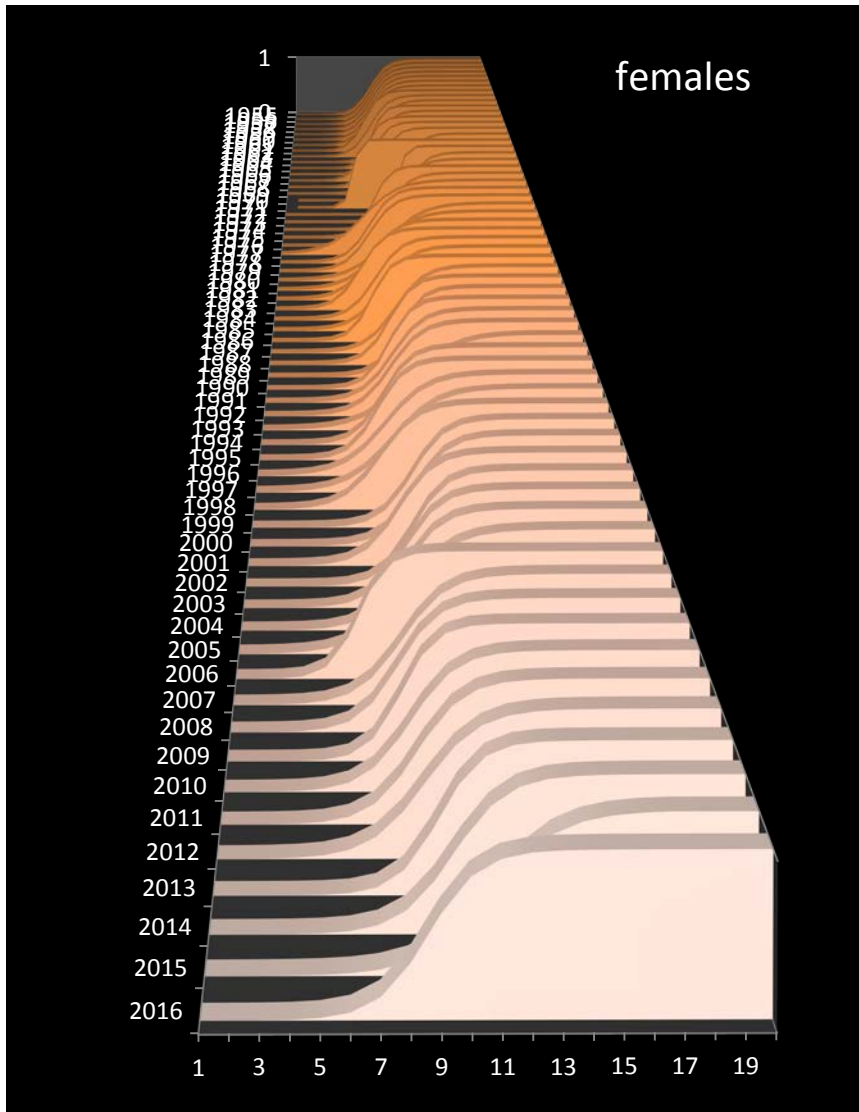


Figure 4.15b.--Estimated female fishery selectivity by age and year.

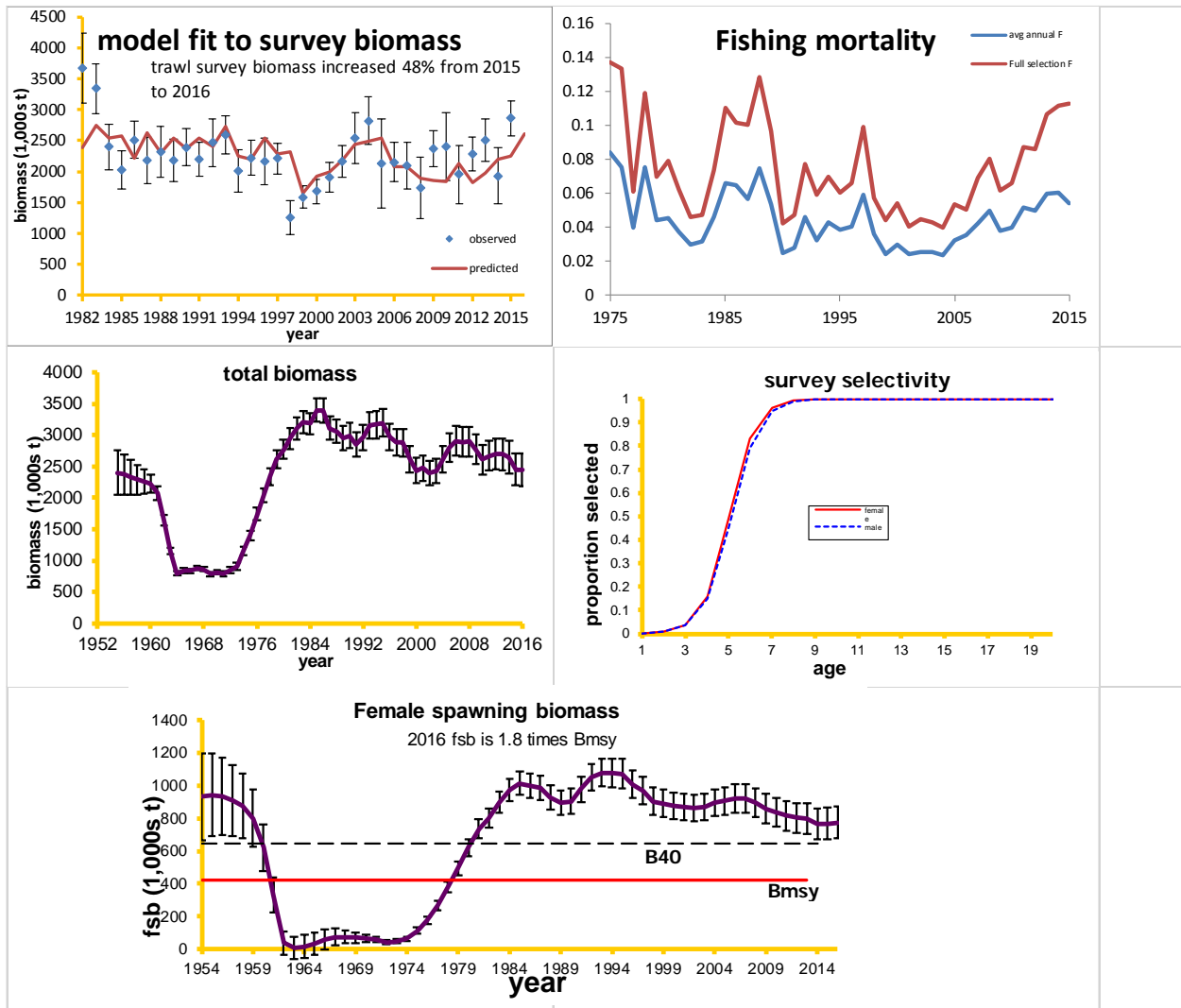


Figure 4.16. Model fit to the survey biomass estimates (top left panel), model estimate of the full selection fishing mortality rate throughout the time-series (top right panel), model estimate of total biomass (middle left panel), the model estimate of survey selectivity (middle right panel) and the estimate of female spawning biomass (bottom left panel).

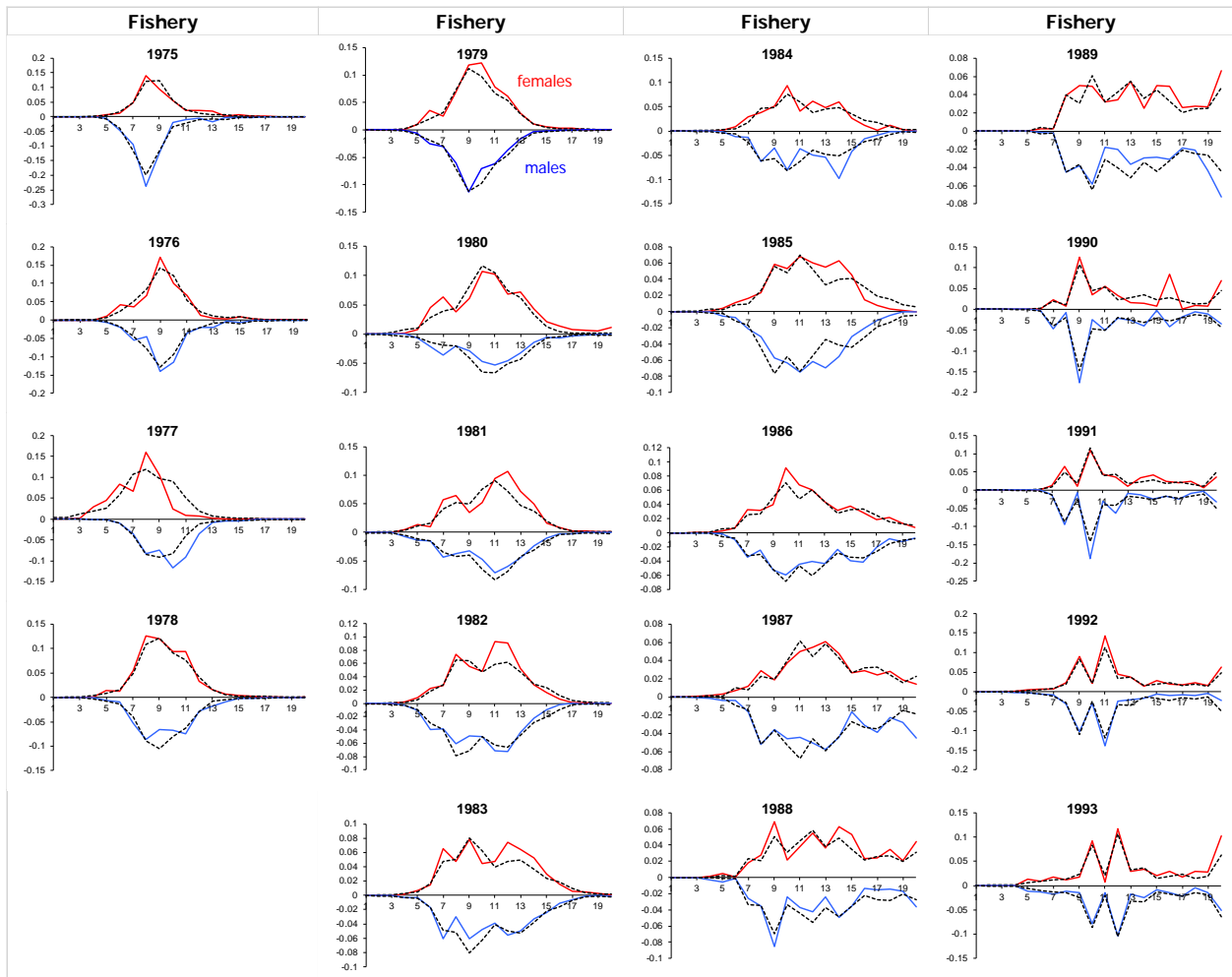


Figure 4.17. Stock assessment model fit to the time-series of fishery and survey age composition, by sex.

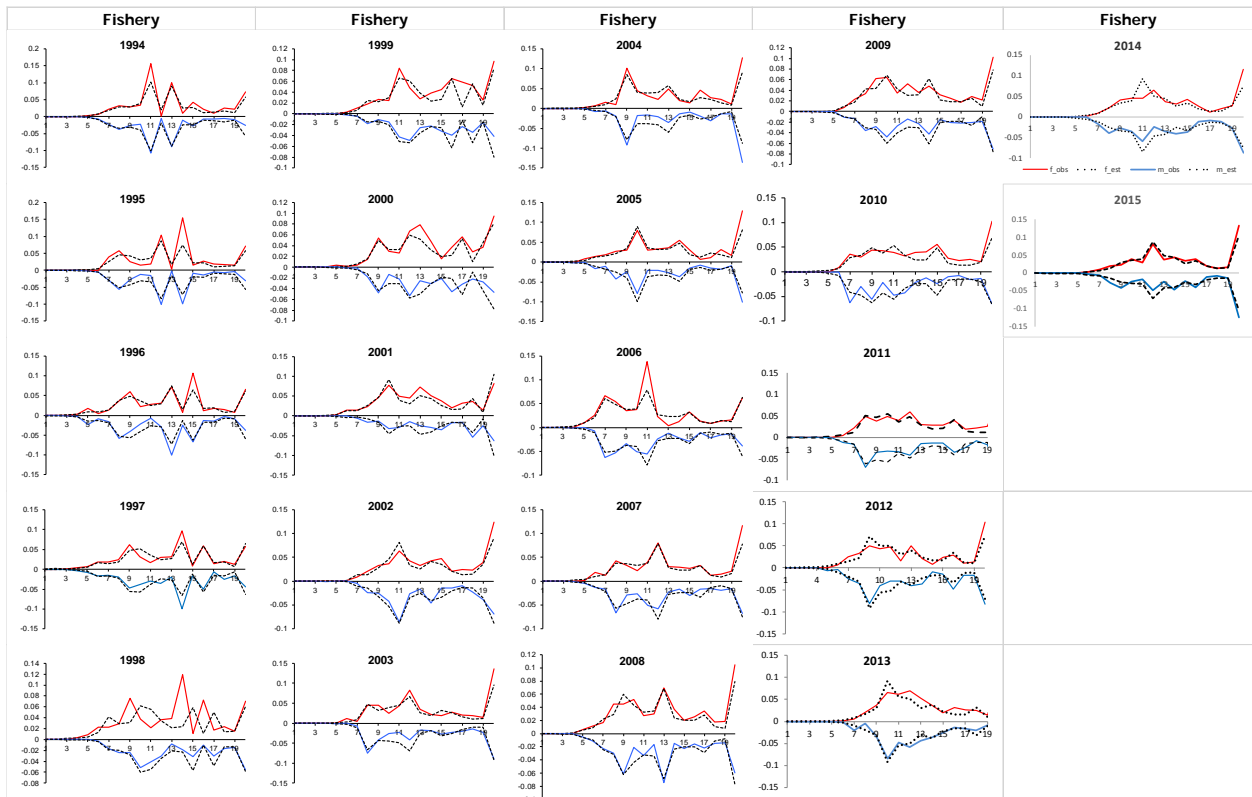


Figure 4.17 (continued).

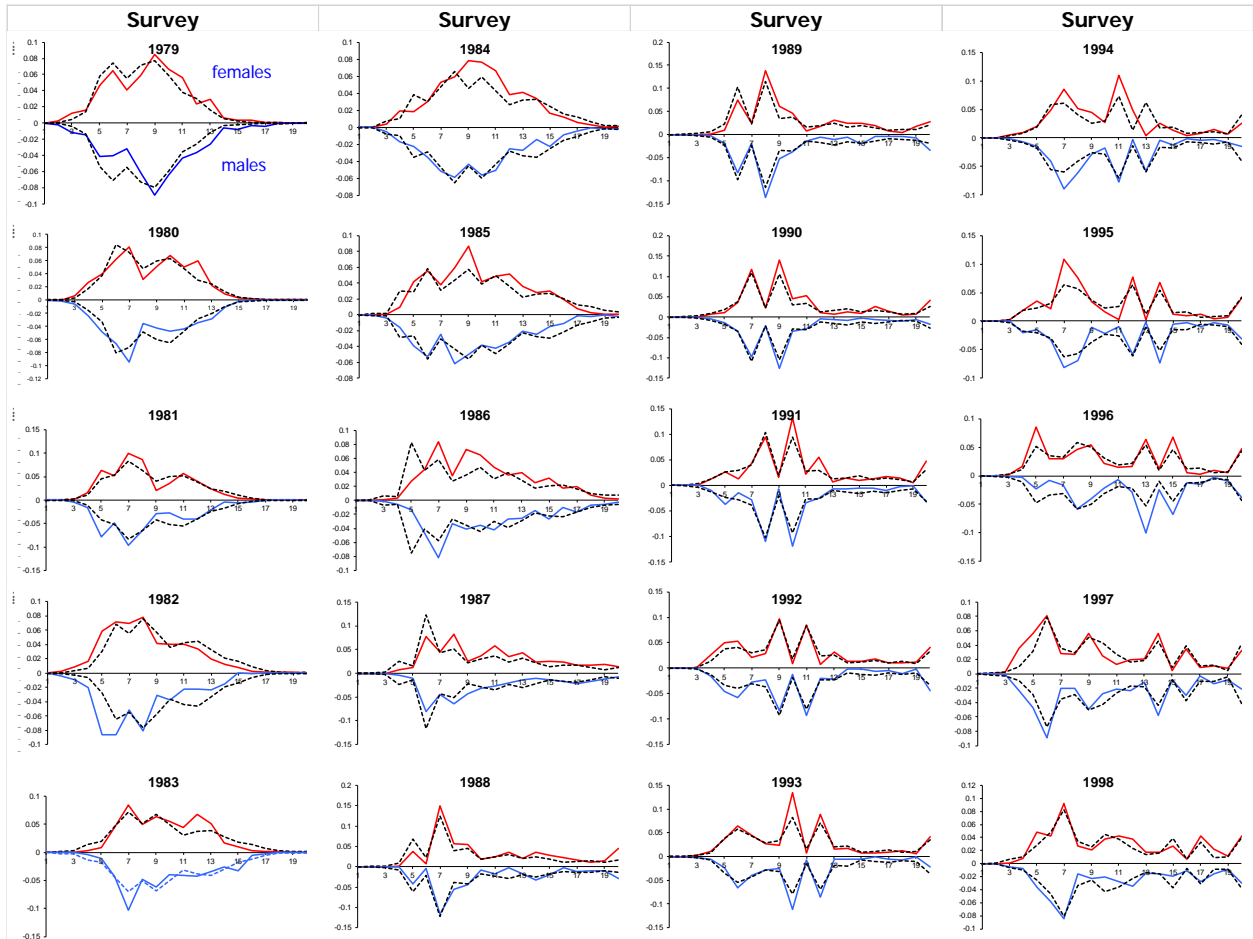


Figure 4.17 (continued).

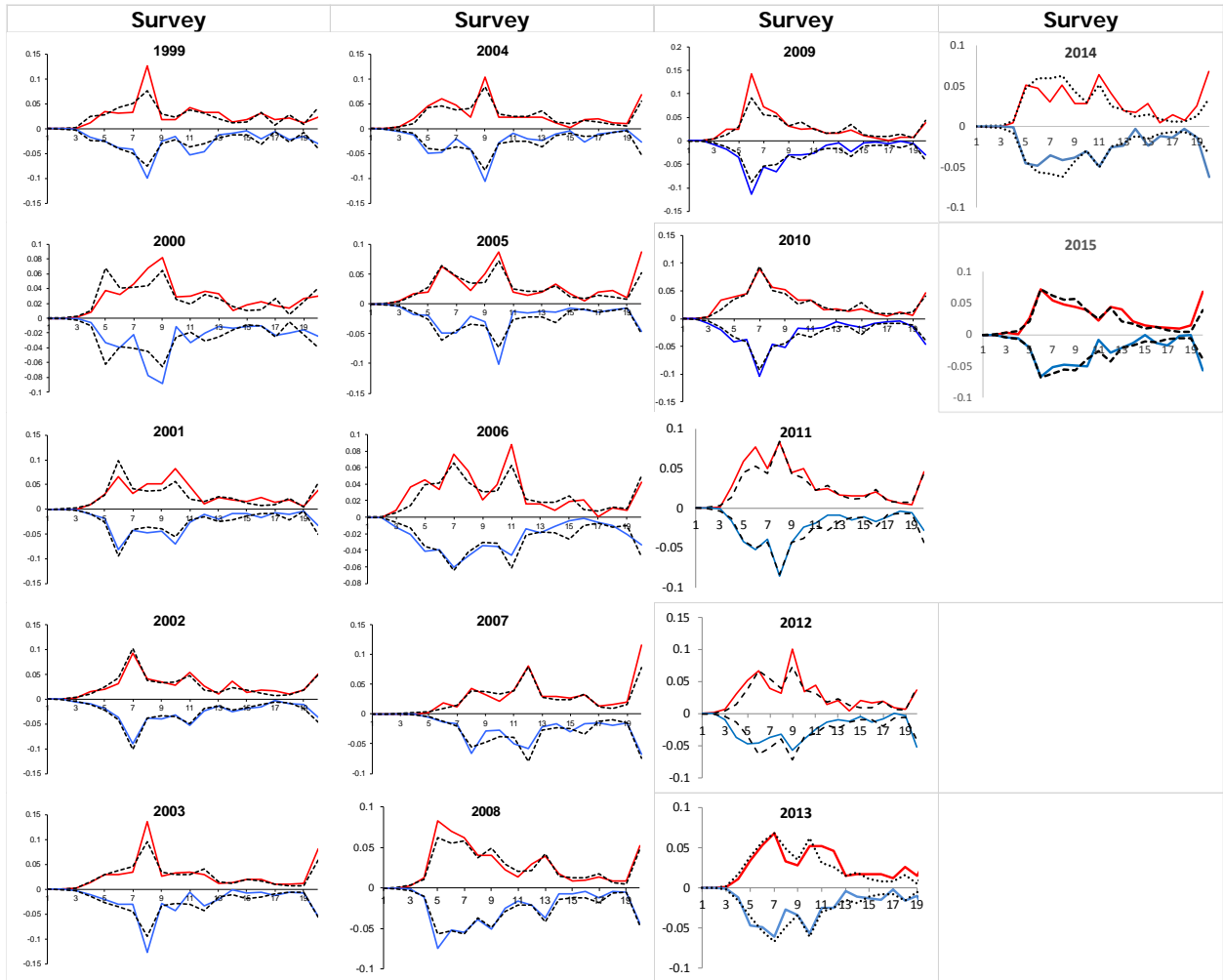


Figure 4.17 (continued).

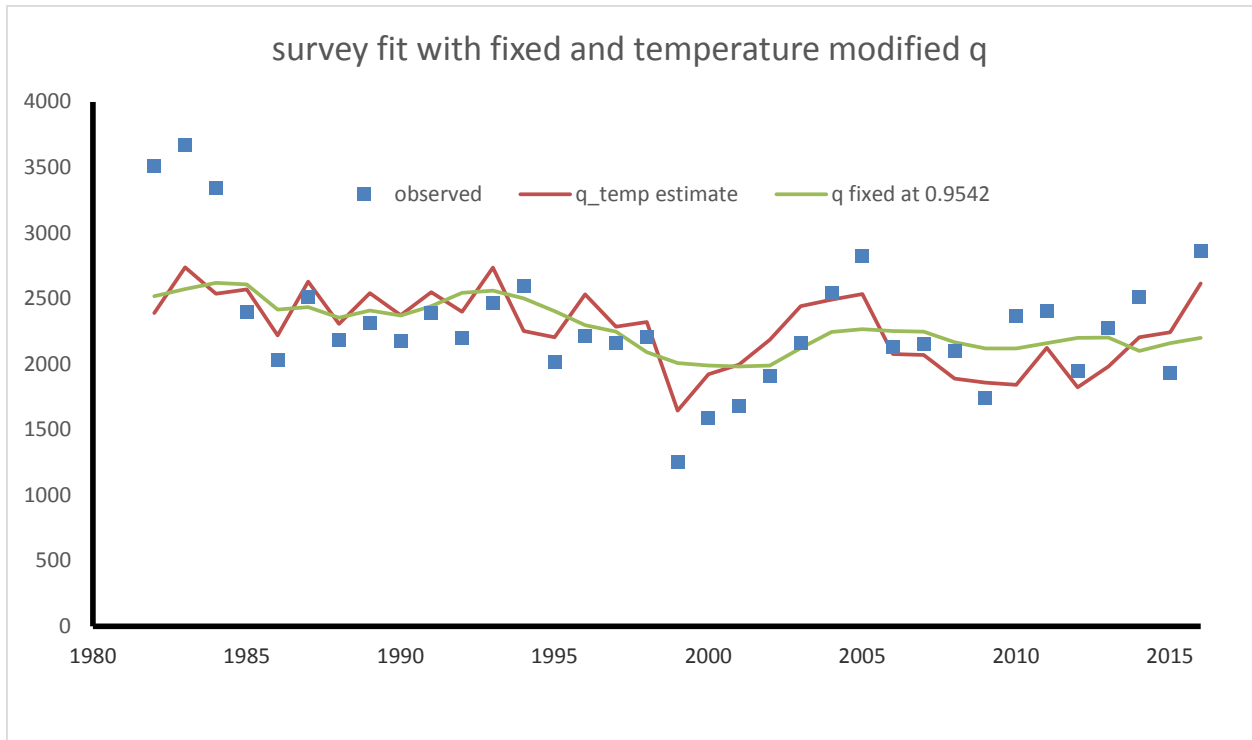


Figure 4.18.--Comparison of the fit to the survey biomass using a fixed q and the q -bottom temperature relationship.

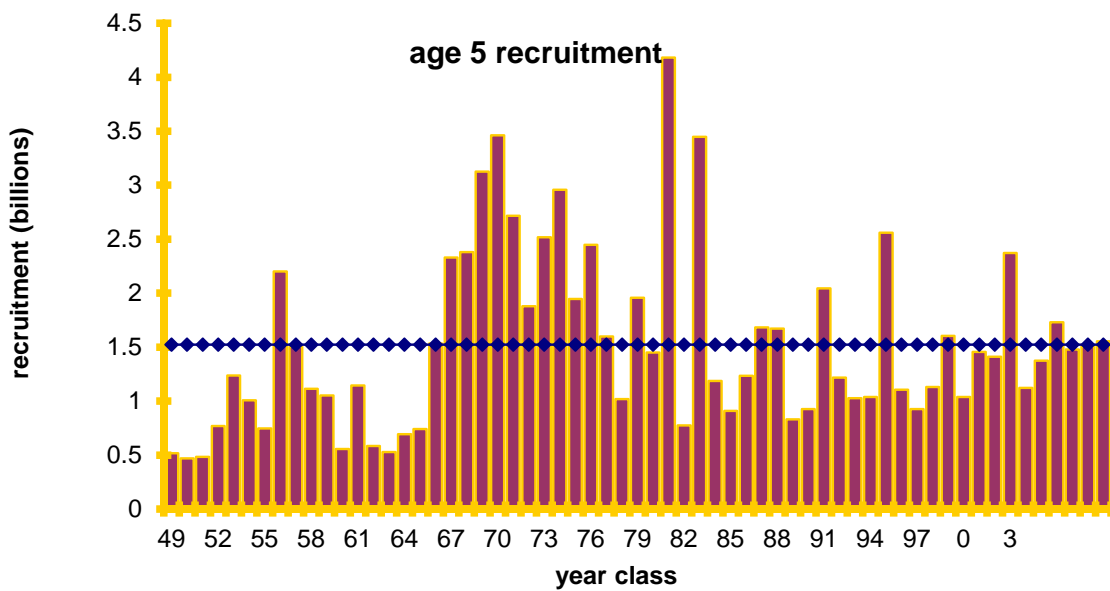


Figure 4.19.--Year class strength of age 5 yellowfin sole estimated by the stock assessment model. The dotted line is the average of the estimates from 62 years of recruitment.

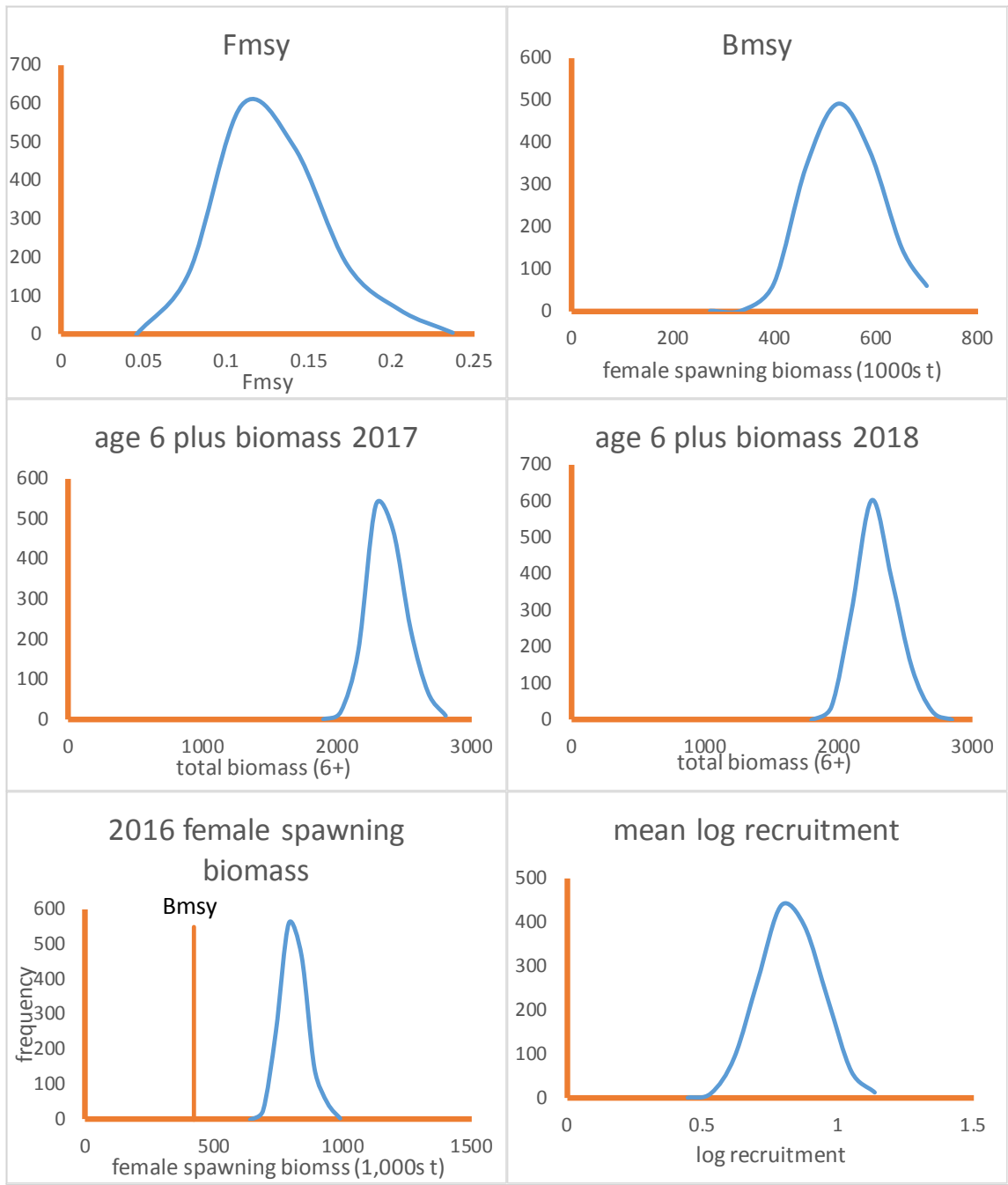


Figure 4.20.--Posterior distributions of some important parameters estimated by the preferred stock assessment model (from mcmc integration).

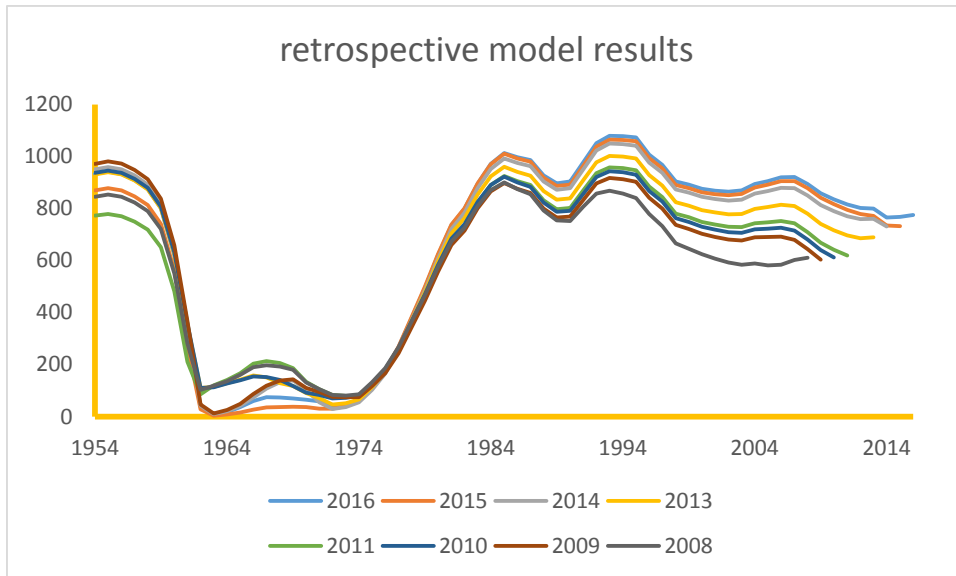


Figure 4.21—Retrospective plot of yellowfin sole female spawning biomass estimates (1,000s t), 2008-2016, from the recommended assessment model.

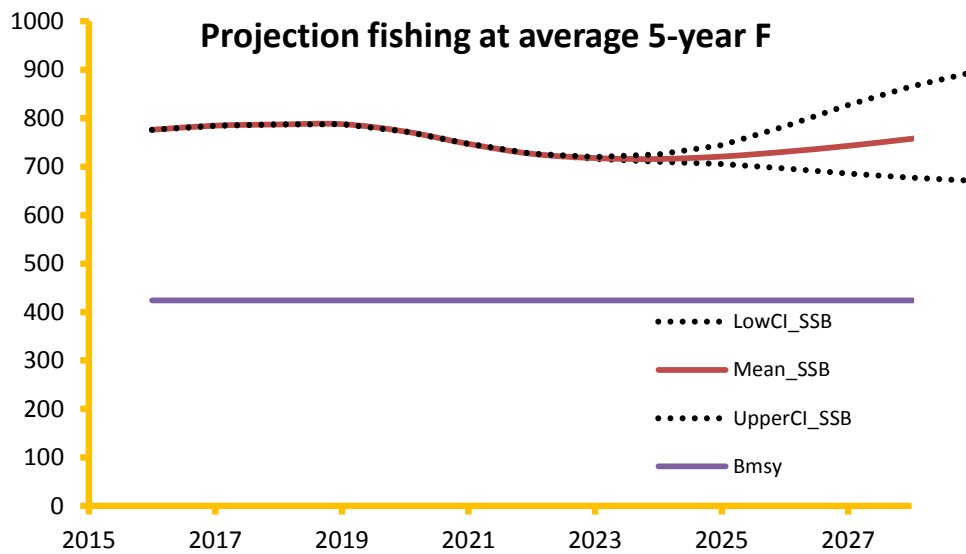


Figure 4.22.--Projection of yellowfin sole female spawning biomass (1,000s t) at the average full-selection F from the past 5 years (0.104) through 2029 with $B_{40\%}$ and B_{msy} levels indicated.

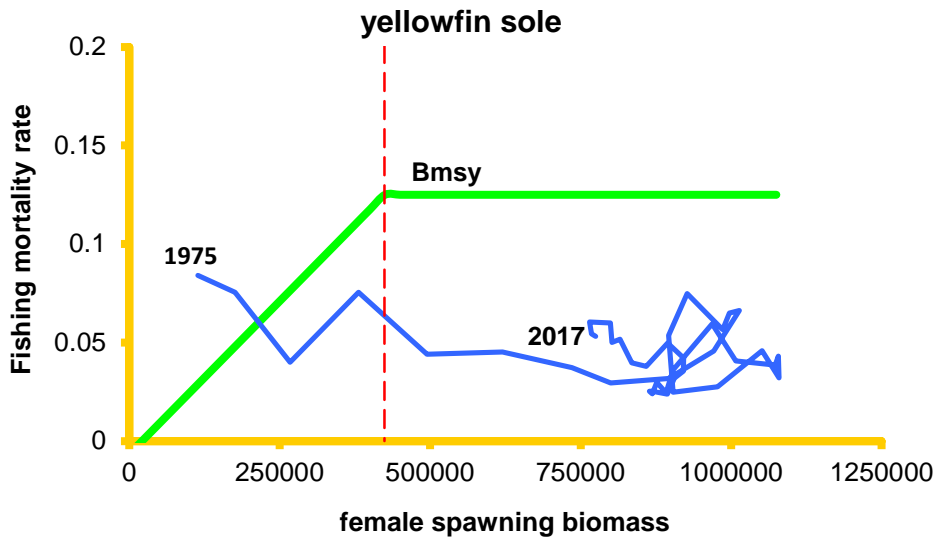


Figure 4.23.--Phase plane figure of the time-series of yellowfin sole female spawning biomass relative to the harvest control rule with 1975 and 2017 indicated.

Appendix

	IPHC research catch of yellowfin sole	
	number	weight (kg)
2007	707	502
2008	0	0
2009	0	0
2010	898	741