# 12. Assessment of the Pacific ocean perch stock in the Bering Sea/Aleutian Islands 

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## Executive Summary

The last full assessment for Pacific ocean perch (POP) was presented to the Plan Team in 2014. The following changes were made to POP assessment relative to the November 2014 SAFE:

## Summary of Changes in Assessment Inputs

Changes in the Input Data

1) Catch data was updated through 2015, and total catch for 2016 was projected.
2) The 2016 AI survey biomass estimate and length composition was included in the assessment.
3) The 2014 AI survey age composition was included in the assessment.
4) The eastern Bering Sea slope survey biomass estimates, age compositions (through 2012) and length composition (for 2016) were included in the assessment.
5) The 2014 and 2015 fishery length compositions were included in the assessment.
6) The 1968-1979 Fishery CPUE indices were not used in the assessment.
7) The fishery age and length composition data were recomputed to weight the length composition within subareas by the observed subarea catch.
8) The length-at-age, weights-at-age, and age-to-length conversion matrix were updated based on data from the NMFS AI trawl survey beginning in 1991.

Changes in the Assessment Methodology

1) In the 2014 assessment, the weights for the age/length composition data were obtained such that the standardized deviation of normalized residuals was a constant value (1) for all composition data types. Several methods for weighting the composition data were considered in this assessment, with the preferred model using the McAllister-Ianelli method.

## Summary of Results

A summary of the 2016 assessment recommended ABCs relative to the 2015 recommendations is shown below. BSAI Pacific ocean perch are not overfished or approaching an overfished condition. The recommended 2017 ABC and OFL are 43,723 t and $53,152 \mathrm{t}$, which are $38 \%$ increases from the maximum ABC and OFL specified last year for 2017 of $31,724 \mathrm{t}$ and $38,589 \mathrm{t}$. The 2016 AI survey biomass is large and consistent with the survey biomass estimates in 2010, 2012, and 2014, and the size composition data continue to show relatively strong recent cohorts. The mode is better able to fit the large AI survey biomass estimates since 2010, although the model total biomass is still lower the survey biomass estimates. A summary of the recommended ABCs and OFLs from this assessment relative the ABC and OFL specified last year is shown below:

| Quantity | As estimated or specified last year for: |  | As estimated or recommended this year for: |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | 2017 | 2017 | 2018 |
| $M$ (natural mortality rate) | 0.062 | 0.062 | 0.058 | 0.058 |
| Tier | 3a | 3a | 3 a | 3 a |
| Projected total (age 3+) biomass (t) | 557,886 | 542,162 | 767,767 | 753,302 |
| Female spawning biomass (t) |  |  |  |  |
| Projected | 222,369 | 211,339 | 314,489 | 307,808 |
| B100\% | 423,008 | 423,008 | 536,713 | 536,713 |
| $\mathrm{B}_{40 \%}$ | 169,203 | 169,203 | 214,685 | 214,685 |
| $B_{35 \%}$ | 148,053 | 148,053 | 187,849 | 187,849 |
| $F_{\text {OFL }}$ | 0.109 | 0.109 | 0.101 | 0.101 |
| $\operatorname{maxF}_{\text {ABC }}$ | 0.089 | 0.089 | 0.082 | 0.082 |
| $F_{\text {ABC }}$ | 0.089 | 0.089 | 0.082 | 0.082 |
| OFL (t) | 40,529 | 38,589 | 53,152 | 51,950 |
| maxABC (t) | 33,320 | 31,724 | 43,723 | 42,735 |
| ABC (t) | 33,320 | 31,724 | 43,723 | 42,735 |
|  | As determined | ear for: | As determined | ear for: |
| Status | 2014 | 2015 | 2015 | 2016 |
| Overfishing | No | n/a | No | n/a |
| Overfished | n/a |  | n/a | No |
| Approaching overfished | n/a |  | n/a | No |

*Projections are based on estimated catches of $30,835 \mathrm{t}$ and $30,139 \mathrm{t}$ used in place of maximum permissible ABC for 2017 and 2018.

## Area Apportionment

The ABC for BSAI Pacific ocean perch is currently apportioned among four areas: the western, central, and eastern Aleutian Islands, and eastern Bering Sea. A random effects model was used to smooth the time series of subarea survey biomass and obtain the proportions. Additionally, the smoothed biomass estimated for the EBS slope was adjusted to account for differences in estimated catchability and selectivity between the AI and EBS trawl surveys. The following table gives the projected OFLs and apportioned ABCs for 2017 and 2018 (using the adjusted distribution above), and the recent OFLs, ABCs, TACs, and catches.

| Area | Year | Age 3 Bio (t) | OFL | ABC | TAC | Catch ${ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BSAI | 2015 | 577,967 | 42,558 | 34,988 | 32,021 | 31,425 |
|  | 2016 | 557,886 | 40,529 | 33,320 | 31,900 | 24,796 |
|  | 2017 | 767,767 | 53,152 | 43,723 |  |  |
|  | 2018 | 753,302 | 51,950 | 42,735 |  |  |
| Eastern Bering Sea | 2015 |  |  | 8,771 | 8,021 | 7,918 |
|  | 2016 |  |  | 8,353 | 8000 | 3,743 |
|  | 2017 |  |  | 11,789 | n/a | n/a |
|  | 2018 |  |  | 11,523 | n/a | n/a |
| Eastern Aleutian Islands | 2015 |  |  | 8,312 | 8,000 | 7,865 |
|  | 2016 |  |  | 7,916 | 7900 | 5,780 |
|  | 2017 |  |  | 10,441 | n/a | n/a |
|  | 2018 |  |  | 10,205 | n/a | n/a |
| Central Aleutian Islands | 2015 |  |  | 7,723 | 7,000 | 6,834 |
|  | 2016 |  |  | 7,355 | 7000 | 6,608 |
|  | 2017 |  |  | 8,113 | n/a | n/a |
|  | 2018 |  |  | 7,930 | n/a | n/a |
| Western Aleutian Islands | 2015 |  |  | 10,182 | 9,000 | 8,808 |
|  | 2016 |  |  | 9,696 | 9000 | 8,663 |
|  | 2017 |  |  | 13,380 | n/a | n/a |
|  | 2018 |  |  | 13,077 | n/a | n/a |

${ }^{1}$ Catch through October 10, 2016

## Responses to SSC and Plan Team Comments on Assessments in General

(Joint Plan Team, November, 2014) For assessments involving age-structured models, this year's CIE review of BSAI and GOA rockfish assessments included three main recommendations for future research:

1. Selectivity/fit to plus group (e.g., explore dome-shaped selectivity, cubic splines)
2. Reevaluation of natural mortality
3. Alternative statistical models for survey data (e.g., GAM, GLM, hurdle models)

The Team agreed that development of alternative survey estimators is a high priority, but concluded that this priority is not specific to rockfish, and should be explored in a Center-wide initiative (see "Alternative statistical models for survey data" under Joint Team minutes). For the remaining two items, the Team recommended that selectivity and fit to the plus group should be given priority over reevaluation of the natural mortality rate.

Selectivity curves and natural mortality rates were evaluated in the 2014 assessment. The development of alternative survey estimators (i.e., model-based standardization of survey catch data) affects all NPFMC assessments that use survey data. Potential methodologies have been discussed in a limited number of meetings in 2014 among AFSC scientists, and between AFSC scientists and NWFSC scientists. Recently, scientists at the NWFSC has developed geostatistical models for survey standardization.

The minutes of the September, 2016 meeting of the Joint Groundfish Plan Team indicate that a workgroup is currently being formed to evaluate statistical models for survey standardization.
(GOA Plan Team, November 2015) The Team recommends an evaluation on how best to tailor the RE model to accommodate multiple indices.
Although this comment originated from the GOA Plan Team, it is also relevant the BSAI assessments. The random effects model is applied to the biomass estimates of the AI trawl survey and EBS slope survey to obtain ABC apportionments. In previous assessments, a simple summation of the smoothed estimates was done, implying that the catchability and selectivity of the two surveys were equivalent. The recommended model in this assessment estimates catchability and selectivity for both surveys, and this information was used to adjust the smoothed EBS slope survey index into units consistent with the AI survey.
(SSC, December 2015) Many assessments are currently exploring ways to improve model performance by re-weighting historic survey data. The SSC encourages the authors and PTs to refer to the forthcoming CAPAM data-weighting workshop report.
(SSC, October 2016) The SSC recommends that the Gulf of Alaska Groundfish Plan Team (GOA GPT), BSAI GPT, and CPT encourage the continued use of multiple approaches to data weighting (not just the Francis (2011) method, but also including the harmonic mean and others).
In this assessment, we evaluate several methods for weighting the age and length composition data. Weighting of the survey biomass indices has been deferred until an evaluation of model-based vs designbased survey estimators is conducted.
(SSC, October 2016) The SSC requests that stock assessment authors bookmark their assessment documents and commends those that have already adopted this practice.

Bookmarks for the major sections of the assessment were added to the 2016 document.

## Responses to SSC and Plan Team Comments Specific to this Assessment

(SSC, December, 2014) The SSC provides the following recommendations to the assessment author;

1) Evaluate whether fishery CPUE data (1968-1979) is necessary and consider removing it in future models.
2) Examine the evidence supporting the selectivity changes in the most recent years in the model. The shift from dome-shaped to asymptotic selectivity around 2010 appears to correspond with a divergence in modeled and survey estimated biomass.
3) Explore a better prior for catchability through empirical studies and determine how to use the EBS slope survey biomass estimates.
4) Explore estimates of biological parameters like maturity to see if there are trends in these estimates.
5) Continue to evaluate potential sources for the retrospective trend including the impacts of estimating survey catchability in the model.
6) Explore potential causes for survey biomass residual pattern

Initial responses to these recommendations were provided in the 2015 assessment. Where applicable, responses to this list are updated below (the numbers refer to the original list in the 2014 SSC minutes).

1) Evaluation of removal of the CPUE index is evaluated in this assessment.
2) Empirical studies on densities of rockfish in untrawlable and trawlable grounds are ongoing. Once completed, they should help inform a prior distribution of survey catchability.

Inclusion of the EBS slope survey estimates is evaluated in this assessment.
6) In this assessment, the model is better able to fit the recent high AI survey biomass. This is likely due to the addition of new data showing continued high biomass and recent year classes, and evaluation of data-weighting methods for the age and length composition data.

## Introduction

Pacific ocean perch (Sebastes alutus) inhabit the outer continental shelf and upper slope regions of the North Pacific Ocean and Bering Sea. Pacific ocean perch, and four other associated species of rockfish (northern rockfish, S. polyspinis; rougheye rockfish, S. aleutianus; shortraker rockfish, S. borealis; and sharpchin rockfish, S. zacentrus) were managed as a complex in the two distinct areas from 1979 to 1990. Known as the POP complex, these five species were managed as a single entity with a single TAC (total allowable catch). In 1991, the North Pacific Fishery Management Council separated POP from the other red rockfish in order to provide protection from possible overfishing. Of the five species in the former POP complex, S. alutus has historically been the most abundant rockfish in this region and has contributed most to the commercial rockfish catch.

## Information on Stock Structure

A variety of types of research can be used to infer stock structure of POP, including age and length compositions, growth patterns and other life-history information, and genetic studies. Spatial differences in age or length compositions can be used to infer differences in recruitment patterns that may correspond to population structure. In Queen Charlotte Sound, British Columbia, Gunderson (1972) found substantial differences in the mean lengths of POP in fishery hauls taken at similar depths which were related to differences in growth rates and concluded that POP likely form aggregations with distinct biological characteristics. In a subsequent study, Gunderson (1977) found differences in size and age composition between Moresby Gully and two other gullies in Queen Charlotte Sound. Westrheim (1970, 1973) recognized "British Columbia" and "Gulf of Alaska" POP stocks off the western coast of Canada based upon spatial differences in length frequencies, age frequencies, and growth patterns observed from a trawl survey. In a study that has influenced management off Alaska, Chikuni (1975) recognized distinct POP stocks in four areas - eastern Pacific (British Columbia), Gulf of Alaska, Aleutian Islands, and Bering Sea. However, Chikuni (1975) states that the eastern Bering Sea (EBS) stock likely receives larvae from both the Gulf of Alaska (GOA) and Aleutian Islands (AI) stock, and the AI stock likely receives larvae from the GOA stock.

An alternative approach to evaluating stock structure involves examination of rockfish life-history stages directly. Stock differentiation occurs from separation at key life-history stages. Because many rockfish species are not thought to exhibit large-scale movements as adults, movement to new areas and boundaries of discrete stocks may depend largely upon the pelagic larval and juvenile life-history stages. Simulation modeling of ocean currents in the Alaska region suggest that larval dispersal may occur over very broad areas, and may be dependent on month of parturition (Stockhausen and Herman 2007).

Analysis of field samples of rockfish larvae are hindered by difficulties in indentifying species. Analyses of archived Sebastes larvae was undertaken by Dr. Art Kendall revealed that species identification based on morphological characteristics is difficult because of overlapping characteristics among species, as few rockfish species in the north Pacific have published descriptions of the complete larval developmental series. However, all of the larvae examined could be assigned to four morphs identified by Kendall (1991), where each morph is associated with one or more species. Rockfish identification can be aided by studies that combine genetic and morphometric techniques and information has been developed to identify individual species based on allozymes (Seeb and Kendall 1991) and mitochondrial DNA (Gharrett et al. 2001, Rocha-Olivares 1998). The Ocean Carrying Capacity (OCC) field program, conducted by the Auke Bay laboratory, uses surface trawls to collect juvenile salmon and incidentally collects juvenile rockfish. These juvenile rockfish are large enough (approximately 25 mm and larger) to allow extraction of a tissue sample for genetic analysis without impeding morphometric studies. In 2002, species identifications were made for an initial sample of 55 juveniles with both morphometric and genetic techniques. The two techniques showed initial agreement on 39 of the 55 specimens, and the genetic results motivated re-evaluation of some of the morphological species identifications. Forty of the
specimens were identified as POP, and showed considerably more morphological variation for this species than previously documented.

Because stocks are, by definition, reproductively isolated population units, it is expected that different stocks would show differences in genetic material due to random drift or natural selection. Thus, analysis of genetic material from North Pacific rockfish is currently an active area of research.

Seeb and Gunderson (1988) used protein electrophoresis to infer genetic differences based upon differences in allozymes from POP collected from Washington to the Aleutian Islands. Discrete genetic stock groups were not observed, but instead gradual genetic variation occurred that was consistent with the isolation by distance model. The study included several samples in Queen Charlotte Sound where Gunderson (1972, 1977) found differences in size compositions and growth characteristics. Seeb and Gunderson (1988) concluded that the gene flow with Queen Charlotte Sound is sufficient to prevent genetic differentiation, but adult migrations were insufficient to prevent localized differences in length and age compositions. More recent studies of POP using microsatellite DNA revealed population structure at small spatial scales, consistent with the work of Gunderson (1972, 1977). These findings suggest that adult POP do not migrate far from their natal grounds and larvae are entrained by currents in localized retention areas (Withler et al. 2001).

Interpretations of stock structure are influenced by the technique used to assess genetic analysis differentiation, as illustrated by the differing conclusions produced from the POP allozyme work of Seeb and Gunderson (1988) and the microsatellite work of Withler et al. (2001). Note that these two techniques assess components of the genome that diverge on very different time scales and that, in this case, microsatellites are much more sensitive to genetic isolation. Protein electrophoresis examines DNA variation only indirectly via allozyme frequencies, and does not recognize situations where differences in DNA may result in identical allozymes (Park and Moran 1994). In addition, many microsatellite loci may be selectively neutral or near-neutral, whereas allozymes are central metabolic pathway enzymes and do not have quite the latitude to produce viable mutations. The mutation rate of microsatellite alleles can be orders of magnitude higher than allozyme locus mutation rates. Most current studies on rockfish genetic population structure involve direct examination of either mitochondrial DNA (mtDNA) or microsatellite DNA.

Dr. Anthony Gharrett of the Juneau Center of Fisheries and Ocean Sciences has examined the mtDNA and microsatellite variation for POP samples collected in the GOA and BSAI. The POP mtDNA analysis was performed on 124 fish collected from six regions ranging from southeast Alaska to the Bering Sea slope and central Aleutian Islands. No population structure was observed, as most fish (102) were characterized by a common haplotype. Preliminary results from an analysis of 10 microsatellite loci from the six regions resulted in 7 loci with significant heterogeneity in the distribution of allele frequencies. Additionally, the sample in each region was statistically distinct from those in adjacent regions, suggesting population structure on a relatively fine spatial scale consistent with the results on Gunderson (1972, 1977) and Wither et al. (2001). Ongoing genetic research with POP is focusing on increasing the sample sizes and collection sites for the microsatellite analysis in order to further refine our perception of stock structure.

## Fishery

POP were highly sought by Japanese and Soviet fisheries and supported a major trawl fishery throughout the 1960s. Catches in the eastern Bering Sea peaked at 47,000 (metric tons, t) in 1961; the peak catch in the Aleutian Islands region occurred in 1965 at 109,100 t. These stocks were not productive enough to support such large removals. Catches continued to decline throughout the 1960s and 1970s, reaching their lowest levels in the mid 1980s. With the gradual phase-out of the foreign fishery in the 200 -mile
U.S. Exclusive Economic Zone (EEZ), a small joint-venture fishery developed but was soon replaced by a domestic fishery by 1990. In 1990 the domestic fishery recorded the highest POP removals since 1977. The OFLs, ABCs, TACs, and catches by management complex from 1977 to 2000 (when POP were managed as separate stocks in the EBS and AI) are shown in Table 1. Note that in some years, POP were managed in the "POP complex" management group, which also included rougheye rockfish, shortraker rockfish, northern rockfish, and sharpchin rockfish. In 2002 POP were managed as a single stock across the BSAI (with the ABC subdivided between the EBS and AI subareas, and the BSAI OFLs, ABCs, TACs, and catches for this period is shown in Table 2. The ABCs, TACs, and catches from 1988 to 2016 are shown in Table 2. The catches of POP from 1977 by fishery type (i.e., foreign, joint venture, or domestic) is shown in Table 3.

Estimates of retained and discarded POP from the fishery have been available since 1990 (Table 4). From 1990-2009, the eastern Bering Sea region generally showed a higher discard rate than in the Aleutian Islands region, with the average rates $33 \%$ and $14 \%$, respectively. From 2010-2015, bycatch rates in the the eastern Bering Sea and the Aleitian Islands were low, averaging 9\% and 2\% respectively.

Initial age-structured assessments for BSAI POP modeled separate selectivity curves for the foreign and domestic fisheries (Ianelli and Ito 1992), although examination of the distribution of observer catch reveals interannual changes in the depth and areas in which POP are observed to be caught within the foreign and domestic periods. For example, POP are predominately taken in depths between 200 m and 300 m , although during the late 1970s-early-1980s a relatively large portion of POP were observed to be captured at depths greater than 300 m (Table 5, Figure 1). Additionally, from 1999 through the early 2000s the proportion caught between 100 m and 200 m increased from $\sim 20 \%$ in the early to mid 1990s to $\sim 30 \%$, and since the mid-2000s the proportion caught between 200 m and 300 m has increased. The area of capture has changed as well; during the late 1970s POP were predominately captured in the western Aleutians, whereas from the early 1980s to the mid-1990s POP were captured predominately in the eastern Aleutians. Establishment of area-specific TACs in the mid-1990s redistributed the POP catch such that about $50 \%$ of the current catch is now taken in the western Aleutians (Table 6, Figure 1). Note that the extent to which the patterns of observed catch can be used as a proxy for patterns in total catch is dependent upon the degree to which the observer sampling represents the true fishery. In particular, the proportions of total POP caught that were actually sampled by observers were very low in the foreign fishery, due to low sampling ratio prior to 1984 (Megrey and Wespestad 1990).

Non-commercial catches are shown in Appendix A.

## Data

## Fishery Data

Catch per unit effort (CPUE) data from Japanese trawl fisheries indicate that POP stock abundance declined to very low levels in the Aleutian Islands region (Ito 1986). By 1977, CPUE values had dropped by more than $90-95 \%$ from those of the early 1960s. Japanese CPUE data after 1977, however, is not a good index of stock abundance because most of the fishing effort has been directed to species other than POP. Standardizing and partitioning total groundfish effort into effort directed solely toward POP is difficult. Increased quota restrictions, effort shifts to different target species, and rapid improvements in fishing technology undoubtedly affect our estimates of effective fishing effort. The catch per unit effort data from a select group of trawl vessels targeting POP from 1968-1979 have historically been included in the assessment. Removal of this data (due to redundancy with other information, and unclear methodology for the standardization) was evaluated in this assessment, with the recommended model not including this data.

Length measurements and otoliths read from the EBS and AI management areas (Tables 7 and 8) were combined to create fishery age/size composition matrices, with the length composition within management subareas weighted by the estimated catch numbers from observed tows. Age and/or length composition were not included for several years due due to low samples sizes of fish measured (years 1973-1976, 1985-1986), and/or otoliths read (years 1984-86). In 1982, the method for aging otoliths at the Alaska Fisheries Science Center changed from surface reading to the break and burn method (Betty Goetz, Alaska Fisheries Science Center, pers. comm.), as the latter method is considered more accurate for older fish (Tagart 1984). The time at which the otoliths collected from 1977 to 1982 were read is not known for many vessels and cruises. However, the information available suggests that otoliths from 1977 to 1980 were read prior to 1981, whereas otoliths from 1981 and 1982 were read after 1982. Thus, fishery otoliths from 1977 to 1980 were not used because they were believed to be read by surface ageing and thought to be biased.

Beginning in 1998, samples of otoliths from the fishery catch have been read almost annually or biennially, and show relatively strong year classes from 1984-1988. Fishery age compositions from 20052013 indicate several strong recent year classes from 1995-2000 (Figure 2).

## Survey Data

Cooperative U.S. - Japan trawl surveys were conducted in the AI 1980, 1983, and 1986, and have been used in previous BSAI POP assessments. However, differences exist in gear design and vessels used between these surveys and the NMFS surveys beginning in 1991 (Skip Zenger, National Marine Fisheries Service, personal communication). For example, the Japanese nets used in the cooperative surveys varied between years and included large roller gear, in contrast to the poly-nor'eastern nets used in the current surveys (Ronholt et al 1994). Given the difficulty of documenting the methodologies for these surveys, and standardizing these surveys with the NMFS surveys, this assessment model is conducted with only the NMFS surveys.

The Aleutian Islands survey biomass estimates were used as an index of abundance for the BSAI POP stock. Since 2000 the survey has occurred biennially, although the 2008 survey was canceled due to a lack of funding. Note that there is wide variability among survey estimates from the portion of the southern Bering Sea portion of the survey (from $165^{\circ} \mathrm{W}$ to $170^{\circ} \mathrm{W}$ ), as the post-1991 coefficients of variation (CVs) range from 0.41 to 0.63 (Table 9). The biomass indices in this region increased from $1,501 \mathrm{t}$ in 1991 to $18,217 \mathrm{t}$ in 1994, and have since ranged between $12,099 \mathrm{t}$ (1997) and 87,952 t (2016). The biomass indices of Pacific ocean perch in the Aleutian Islands management area region ( $170^{\circ} \mathrm{W}$ to $170^{\circ} \mathrm{E}$ ) appears to be less variable, with CVs ranging from 0.11 in 2016 to 0.24 in 1994. The biomass estimates for the AI area have ranged between a low of $342,785 \mathrm{t}$ in 1991 and $894,551 \mathrm{t}$ in 2016. From 2010-2016, the total AI survey biomasses have exceeded $900,000 \mathrm{t}$ for each survey, whereas the survey estimates prior to 2010 have not exceeded $665,000 \mathrm{t}$.

The 2016 survey biomass index of $982,503 \mathrm{t}$ is a slight increase from the 2014 estimate of $970,968 \mathrm{t}$ (Table 9). A decline of $34 \%$ in the estimated biomasses between these two surveys was observed in central AI, whereas increases were observed in the other survey areas. Maps of survey CPUE are shown in Figure 3, and indicate relatively high abundance throughout much of the Aleutian Islands. The coefficient of variation (CV) for the 2016 survey of 0.11 was the lowest CV observed.

Age composition data exists for each Aleutian Islands survey, and the length measurements and otoliths read are shown in Table 10. The survey age compositions from 1991-2000 indicate relatively strong year classes in 1977, 1984, and 1988. Recent age composition data from 2004-2012 indicate relatively strong year classes from 1996 to 2000, and the 2014 age composition indicates relative strong 2004 and 2005
year classes (Figure 4).
The current EBS slope survey was initiated as a biennial survey in 2002. The most recent slope survey prior to 2002, excluding some preliminary tows in 2000 intended for evaluating survey gear, was in 1991. Previous slope survey results have not been used in the BSAI model due to high CVs, relatively small population sizes compared to the AI biomass estimates, and lack of recent surveys. However, the biomass indices in the EBS slope survey have been increasing, ranging from 76,665 tin 2002 to 357,369 t in the 2016 survey, with CVs ranging from 0.68 in 2016 to 0.53 in 2002 (Table 9). EBS survey CPUE from the 2016, 2012, and 2010 surveys are shown in Figure 5. The slope survey was not conducted in 2008 or 2014 due to lack of funding. This assessment evaluates the incorporating this time series (and associated age and length composition data), with the recommended model included these data. Age composition data for the EBS survey is available for all survey years except 2016 (Figure 6).

## Biological data

A large number of samples are collected from the surveys for age determination, length-weight relationships, sex ratio information, and for estimating the length distribution of the population. The age compositions for inclusion in the model were estimated outside the model by constructing age-length keys for each year and using them to estimate the survey age distribution from the estimated survey length distribution from the same year. Because the survey length distributions are used to create the survey age distributions, the survey length distributions are removed from the model in years in which we have survey ages. The survey age data were based on the break and burn method of ageing POP, so they were treated as unbiased but measured with error. Kimura and Lyons (1991) reported that the percent agreement between readers varies from $60 \%$ for age 3 fish to $13 \%$ for age 25 fish data. The information on percent agreement was used to derive the variability of observed age around the "true" age, assuming a normal distribution. The mean number of fish at age available to the survey or fishery is multiplied by the aging error matrix to produce the expected observed survey or fishery age compositions.

Aging methods have improved since the start of the time series. Historically, POP age determinations were done using scales and surface readings from otoliths. These gave estimates of natural mortality of about 0.15 and longevity of about 30 years (Gunderson 1977). Based on the now accepted break and burn method of age determination using otoliths, Chilton and Beamish (1982) determined the maximum age of S. alutus to be 90 years. Using similar information, Archibald et al. (1981) concluded that natural mortality for POP should be on the order of 0.05 .

Aleutian Islands survey data from 1991 through 2014 were used to estimate growth curves. The resulting von Bertalannfy growth parameters were $L_{\text {inf }}=41.62 \mathrm{~cm}, \mathrm{k}=0.14$, and $\mathrm{t}_{0}=-1.270$. Growth information from the Aleutian Islands was used to convert estimated numbers-at-age within the model to estimated numbers-at-length.

A conversion matrix was created to convert modeled number at ages to modeled number at length bin, and consists of the proportion of each age that is expected in each length bin. This matrix was created by fitting a polynomial relationship to the observed CV in length at each age (obtained for each survey from 1991-2014 by the multiplying the estimated survey length distribution by the age-length key), and the predicted relationship was used to produce variation around the predicted size at age from the von Bertalanffy relationship. The resulting CVs of length at age of the transition matrix decrease from 0.14 at age 3 to 0.07 at age 40 .

The estimated length(cm)-weight(g) relationship for Aleutian Islands POP was estimated with survey information from the same years, with the length-weight parameters estimated as $a=1.0 \times 10^{-5}$ and $b=$
3.07, where weight $=a^{*}$ (length $)^{b}$. The Aleutian Islands length-weight relationship was used to produce estimated weights at age.

The following table summarizes the data available for the recommended BSAI POP model:

| Component | BSAI |
| :--- | :--- |
| Fishery catch | $1960-2016$ |
| Fishery age composition | $1981-82,1990,1998,2000-2009,2011,2013$ |
| Fishery size composition | $1964-72,1983-1984,1987-1989,1991-1997,1999,2010,2012$, |
|  | $2014-2015$ |
| AI Survey age composition | $1991,1994,1997,2000,2002,2004,2006,2010,2012,2014$ |
| AI Survey length composition | 2016 |
| AI Survey biomass estimates | $1991,1994,1997,2000,2002,2004,2006,2010,2012,2014,2016$ |
| EBS Survey age composition | $2002,2004,2008,2010,2012$ |
| EBS Survey length composition | 2016 |
| EBS Survey biomass estimates | $2002,2004,2008,2010,2012,2016$ |

## Analytic Approach

## Model Structure

An age-structured population dynamics model, implemented in the software program AD Model Builder, was used to obtain estimates of recruitment, numbers at age, and catch at age. Population size in numbers at age $a$ in year $t$ was modeled as

$$
N_{t, a}=N_{t-1, a-1} e^{-Z_{t-1, a-1}} \quad 3<a<A, \quad 1960<t \leq T
$$

where $Z$ is the sum of the instantaneous fishing mortality rate ( $F_{t, a}$ ) and the natural mortality rate ( $M$ ), $A$ is the maximum number of age groups modeled in the population, and $T$ is the terminal year of the analysis (defined as 2016).

The numbers at age $A$ are a "pooled" group consisting of fish of age $A$ and older, and are estimated as

$$
N_{t, A}=N_{t-1, A-1} e^{-Z_{t-1, A-1}}+N_{t-1, A} e^{-Z_{t-1, A}}
$$

The plus group was set to 40+, following a sensitivity analysis conducted in the 2012 stock assessment (Spencer and Ianelli 2012)

The numbers at age in the first year of the model are estimated as

$$
N_{a}=R_{0} e^{-M(a-3)}
$$

where $R_{0}$ is the number of age 3 recruits for an unfished population, thus producing an age structure in equilibrium with an unfished stock. It is generally thought that little fishing for rockfish occurred prior to 1960, so an equilibrium unfished age-structure seems reasonable.

The total numbers of age 3 fish (recruitment) from 1960 to 2013 are estimated as parameters in the model, and are modeled with a lognormal distribution

$$
N_{t, 3}=e^{\mu_{R}+v_{t}}
$$

where $v_{t}$ is a time-variant deviation with a log-scale recruitment standard deviation of $\sigma_{r}$. Little information exists to determine the year-class strength for the three most recent cohorts (2014-2016), which were set to the estimated mean recruitment (based upon the log-scale mean, and the value of $\sigma_{\mathrm{r}}$ ).

The fishing mortality rate for a specific age and time $\left(F_{t, a}\right)$ is modeled as the product of a $s_{a, t}{ }_{a}$ and a yearspecific fully-selected fishing mortality rate $f$. The fully selected mortality rate is modeled as the product of a mean $\left(\mu_{f}\right)$ and a year-specific deviation $\left(\varepsilon_{t}\right)$, thus $F_{t, a}$ is

$$
F_{t, a}=s_{a, t}^{f} f_{t}=s_{a, t}^{f} e^{\left(\mu_{f}+\varepsilon_{t}\right)}
$$

The mean number-at-age for each year was computed as

$$
\bar{N}_{t, a}=N_{t, a}\left(1-e^{-z_{t, a}}\right) / Z_{t, a}
$$

Catch biomass-at-age was computed as the product of mean numbers at age, instantaneous fishing mortality, and weight at age.

In previous assessment, the Aleutian Islands trawl survey catchability incorporated the processes of availability (either areal or vertical) and vulnerability to the gear. The introduction of the EBS trawl survey catchability requires consideration of how much of the BSAI stock is "available" to the each survey. The availability ( $a_{\mathrm{Al,t}}$ ) in each year to the AI survey was obtained by using the random effects model to smooth the AI and EBs survey biomass and computing the proportion of the total smoothed biomass in the AI area. The predicted survey biomass for the AI trawl survey biomass $\hat{B}_{A l, t}^{t w l}$ was computed as

$$
\hat{B}_{A l, t}^{t w l}=a_{A l, t} q^{t w l} \sum_{a}\left(\bar{N}_{t, a} s_{a}^{t w l} W_{a}\right)
$$

where $W_{a}$ is the population weight-at-age, $s_{a}^{t w l}$ is the survey selectivity, and $q^{t w l}$ is the trawl survey catchability. The The predicted survey biomass for the EBS trawl survey biomass $\hat{B}_{E B S, t}^{t w l}$ is similar but model availability as ( $1-a_{\mathrm{AI}, \mathrm{t}}$ ):

$$
\hat{B}_{E B S, t}^{t w l}=\left(1-a_{A I, t}\right) q^{t w l} \sum_{a}\left(\bar{N}_{t, a} S_{a}^{t w l} W_{a}\right)
$$

Selectivities for the AI and EBS trawl surveys were modeled with logistic functions.
To facilitate parameter estimation, prior distributions were used for the survey catchability and the natural mortality rate $M$. A lognormal distribution was also used for the natural mortality rate $M$, with the mean set to 0.05 and the CV set to 0.05 . The standard deviation of $\log$ recruits, $\sigma_{\mathrm{r}}$, was fixed at 0.75 . Similar, the prior distribution for Aleutian Islands survey selectivity followed a lognormal distribution with a mean of 1.0 and a coefficient of variation (CV) of 0.45 . EBS survey selectivity was estimated freely.

Fishery selectivity was modeled with a bicubic spline with 4 year nodes and 5 age nodes, for a total of 20 selectivity parameters. Values at these nodes are the log-scale fishery selectivity and estimated as parameters, and fishery selectivity at ages and years between the nodes are interpolated with the bicubic spline. The smoothness of the surface is controlled by the number of nodes, and also by a series of
penalties estimated within the model. Four types of penalties were used: 1) smoothness across the ages (modeled with the sum of second differences); 2) the slope of the rate of decline when selectivity decreases with age (modeled with the sum of first differences); 3) the smoothness across years (modeled with the sum of second differences); and 4) the smoothness across years (modeled with the first difference; this addresses situations in which the selectivity across years was relatively smooth but also non-constant, as would occur with a trend).

## Sample sizes for age and length composition data

The "models" in this assessment differ in the types of data included and the weighting of the age and length composition data (rather than structural changes in the modelling equations):

Model 0) The 2014 model results. This is shown in some plots as a basis for comparing the new models.
Model 14) The 2014 model with data updated through 2016. The weighting of the age and length composition data was unchanged from 2014.

Model 16.1) Incorporation of the EBS slope survey biomass estimates and age and length composition data. The data weighting was unchanged from the 2014 model, with weights for the EBS age and length composition data set to 1 .

Model 16.2) Model 16.1, but with removal of the CPUE time series.

Models 4-6 involve different methods for reweighting the age and length composition data. In each of these methods, the multinomial sample size $N_{j, y}$ for data type $j$ and year $y$ is computed as

$$
N_{j, y}=w_{j} \tilde{N}_{j, y}
$$

where $\tilde{N}_{j, y}$ is the original "first stage" sample size (set to the square root of fish lengthed or aged), and $w_{j}$ is a weight for data type $j$. The weights are a function of the fit of to the age and length composition data, and iterated in successive model runs until they converge. Note that this method preserves the relative weighting between years within a given data type.

Model 16.3) Model 16.2, but computes the weights as the harmonic mean of the ratio of effective sample size to first stage sample size (method TA1.1 in Francis (2011); often referred to as the "McAllisterIanelli method").

Model 16.4) Model 16.2, but computes the weights as the inverse of the variance of the standardized residuals (method TA1.2 in Francis (2011); this method was used in the 2014 assessment).

Model 16.5) Model 16.2, bit computes the weights as the variance of a standardized residual between the means of observed and predicted ages (or lengths) (i.e., one residual is computed for each year within a data type. This is method TA1.8 in Francis (2011) and often referred to as the "Francis method".

Because the differences between the "models" above pertain to differences in the input data, standard model selection criteria such as AIC do not apply. The root mean squared error (RMSE) was used to evaluate the relative size of residuals within data types across the different models:

$$
R M S E=\sqrt{\frac{\sum_{n}(\ln (y)-\ln (\hat{y}))^{2}}{n}}
$$

where $y$ and $\hat{y}$ are the observed and estimated values, respectively, of a series length $n$.

## Parameters Estimated Outside the Assessment Model

The parameters estimated independently include the age error matrix, the age-length conversion matrix, individual weight at age, and the proportion of the stock available to the AI survey. The calculations for these quantities are described above.

## Parameters Estimated Inside the Assessment Model

Parameter estimation is facilitated by comparing the model output to several observed quantities, such as the age and length composition of the survey and fishery catch, the survey biomass, and the catch biomass. The general approach is to assume that deviations between model estimates and observed quantities are attributable to observation error and can be described with statistical distributions. Each data component provides a contribution to a total log-likelihood function, and parameter values that minimize the negative log-likelihood are selected.

The likelihood of the initial recruitments were modeled with a lognormal distribution, yielding the following negative log-likelihood (excluding some constant terms)

$$
\lambda_{1}\left[\sum_{t=1}^{n} \frac{\left(v_{t}+\sigma_{r}^{2} / 2\right)^{2}}{2 \sigma_{r}^{2}}+n \ln \left(\sigma_{r}\right)\right]
$$

where $n$ is the number of years where recruitment is estimated. The adjustment of adding $\sigma^{2} / 2$ to the deviation was made in order to produce deviations from the mean, rather than the median, recruitment. If $\sigma_{\mathrm{r}}$ is fixed, the term $n \ln \left(\sigma_{\mathrm{r}}\right)$ adds a constant value to the negative log-likelihood. .

The likelihoods of the fishery and survey age and length compositions were modeled with a multinomial distribution. The negative log of the multinomial function (excluding constant terms) for the fishery length composition data, with the addition of a term that scales the likelihood, is

$$
-n_{f, t, l} \sum_{s, t, l}\left(p_{f, t, l} \ln \left(\hat{p}_{f, t, l}\right)-p_{f, t, l} \ln \left(p_{f, t, l}\right)\right)
$$

where $n$ is the reweighted sample size, and $p_{f, t, l}$ and $\hat{p}_{f, t, l}$ are the observed and estimated proportion at length in the fishery by year and length. The likelihood for the age and length proportions in the survey, $p_{\text {surv }, \text { ta }}$ and $p_{\text {surv, } t, l}$, respectively, follow similar equations.

The negative log-likelihood of the survey biomass was modeled with a lognormal distribution:

$$
\lambda_{2} \sum_{t}\left(\ln \left(\text { obs_biom }_{t}\right)-\ln \left(\text { pred_biom }_{t}\right)\right)^{2} / 2 c v_{t}^{2}
$$

where obs_biom ${ }_{t}$ is the observed survey biomass at time $t, c v_{t}$ is the coefficient of variation of the survey biomass in year $t$, and $\lambda_{2}$ is a weighting factor. The negative log-likelihood of the catch biomass was modeled with a lognormal distribution:

$$
\lambda_{3} \sum_{t}\left(\ln \left(o b s_{-} c a t_{t}\right)-\ln \left(\text { pred }_{-} c a t_{t}\right)\right)^{2}
$$

where obs_cat $t_{t}$ and pred_cat ${ }_{t}$ are the observed and predicted catch. The "observed" catch for 2016 is obtained by estimating the Oct-Dec catch (based on the remaining ABC available after October, and the average proportion in recent years of the remaining ABC caught from Oct-Dec) and adding this to the observed catch through October. Because the catch biomass is generally thought to be observed with higher precision that other variables, $\lambda_{3}$ is given a very high weight so as to fit the catch biomass nearly exactly.

A maturity ogive was fit within the assessment model to samples collected in 2010 from fishery and survey vessels ( $n=280$; TenBrink and Spencer 2013) and in 2004 by fishery observers ( $n=165$ ). The samples were analyzed using histological methods. Parameters of the logistic equation were estimated by maximizing the bionomial likelihood within the assessment model. The number of fish sampled and number of mature fish by age for each collection were the input data, thus weighting the two collections by sample size. Due to the low number of young fish, high weights were applied to age 3 and 4 fish in order to preclude the logistic equation from predicting a high proportion of mature fish at age 0 . The estimated age at $50 \%$ maturity is 9.1 years.

The overall negative log-likelihood function, excluding the priors on $M$ and survey catchability, the penalties on time-varying fishery selectivity parameters, and the maturity ogive parameters, is

$$
\begin{aligned}
& \lambda_{1}\left[\sum_{t=1}^{n} \frac{\left(v_{t}+\sigma_{r}^{2} / 2\right)^{2}}{2 \sigma_{r}^{2}}+n \ln \left(\sigma_{r}\right)\right]+ \\
& \lambda_{2} \sum_{t}\left(\ln \left(o b s_{-} \text {biom }_{t}\right)-\ln \left(\text { pred }-b i o m_{t}\right)\right)^{2} / 2 c v_{t}^{2}+ \\
& -n_{f, t, l} \sum_{s, t, l}\left(p_{f, t, l} \ln \left(\hat{p}_{f, t, l}\right)-p_{f, t, l} \ln \left(p_{f, t, l}\right)\right)+ \\
& -n_{f, t, a} \sum_{s, t, l}\left(p_{f, t, a} \ln \left(\hat{p}_{f, t, a}\right)-p_{f, t, a} \ln \left(p_{f, t, a}\right)\right)+ \\
& -n_{\text {surv }, t, a} \sum_{s, t, a}\left(p_{\text {surv }, t, a} \ln \left(\hat{p}_{\text {surv }, t, a}\right)-p_{\text {surv }, t, a} \ln \left(p_{\text {surv }, t, a}\right)\right)+ \\
& -n_{\text {surv }, t, l} \sum_{s, t, a}\left(p_{\text {surv }, t, l} \ln \left(\hat{p}_{\text {surv }, t, l}\right)-p_{\text {surv }, t, l} \ln \left(p_{\text {surv }, t, l}\right)\right)+ \\
& \lambda_{3} \sum_{t}\left(\ln \left(o b s_{-} \text {cat } t\right)-\ln \left(\text { pred }_{-} c a t_{t}\right)\right)^{2}
\end{aligned}
$$

For the model run in this analysis, $\lambda_{1}, \lambda_{2}$, and $\lambda_{3}$ were assigned weights of 1,1 , and 500 , reflecting a strong emphasis on fitting the catch data. The negative log-likelihood function was minimized by varying
the following parameters (using the bicubic fishery selectivity, and inclusion the EBS slope survey and exclusion of the CPUE index):

| Parameter type | Number |
| :--- | :---: |
| 1) Fishing mortality mean | 1 |
| 2) Fishing mortality deviations | 57 |
| 3) Recruitment mean | 1 |
| 4) Recruitment deviations | 54 |
| 5) Unfished recruitment | 1 |
| 6) Biomass survey catchabilities | 2 |
| 7) Fishery selectivity parameters | 20 |
| 8) Survey selectivity parameters | 4 |
| 9) Natural mortality rate | 1 |
| 10) Maturity parameters | 2 |
| Total parameters | 143 |

Finally, a Monte Carlo Markov Chain (MCMC) algorithm was used to obtain estimates of parameter uncertainty (Gelman et al. 1995). One million MCMC simulations were conducted, with every 1,000th sample saved for the sample from the posterior distribution after excluding the first 50,000 simulations. Ninety-five percent confidence intervals were produced as the values corresponding to the $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of the MCMC evaluation. For this assessment, confidence intervals on total biomass, spawning biomass, and recruitment strength are presented.

## Results

## Model Evaluation

All models estimate increased biomass in recent years relative to the model 0 (Figure 7). Comparison of models 16.1 and 16.2 illustrate the effect of removing the CPUE index. Even though the CPUE index covered the years 1968 - 1979, the result of removing the CPUE index is slightly increased biomass in recent years. This occurs from a decrease in the AI survey catchability from 1.41 in Model 16.1 to 1.33 in Model 16.2, as not constraining the model to fit the decline in CPUE in the 1970s allows for flexibility in adjusting survey catchability to fit the AI survey biomass. The decline in biomass in the 1960s and 1970s from the catch data and is very similar between these two models, suggesting that the information in the CPUE index is somewhat redundant.

The three models that re-weight the age and length data are shown in blue in Figure 7. Models 16.2 and 16.4 are very similar to each other and to Model 16.2 (which used the weights from 2014 assessment). Model 16.5, which used the "Francis" weights, estimates reduced total biomass, which results from increasing the AI survey catchability to 1.93 .

All the models that use data updated through 2016 fit the recent high AI survey biomass estimates relative to the model 0 (Figure 8), whereas the fit to the EBS survey is nearly identical across models (Figure 9). Models 16.3 and 16.4 provide very similar fits to the AI survey, and use of the Francis data weights (model 16.5) provides the best fit to these recent estimates. However, this fit comes at a cost of degraded fit to the fishery length composition and the AI survey age composition data, and indicated by the RMSE values (Table 11). The data weights for the fishery length composition and the AI survey age composition in Model 16.5 are greatly reduced relative to the other models, with the fishery length composition data
being nearly removed from the model (Figure 10).
A potential concern with the Francis method is that unreliable estimates of the variance of the residuals may be obtained with data types with a small number of years (as a single residual is computed for each year). For the POP model, there are 5 years of age composition data for the EBS slope survey, and 1 available year available each for the EBS and AI survey. In this assessment, the weights used for the survey length composition data was paired to the age composition from the same survey. It is unclear how the choice of pairing would affect the model results or, more generally, how sensitive the results of the Francis method are to small numbers available years for some data types.

Models 16.3 and 16.4 provide very similar results. We recommend model 16.3 (the McAllister-Ianelli method), partly because its common usage in other assessment models eases communication of the methodology. Estimated values of model parameters and their standard deviations are shown in Table 12.

A retrospective analysis was conducted on model 16.3 to evaluate the effect of recent data on estimated spawning stock biomass. For the current assessment model, a series of model runs were conducted in which the end year of the model was varied from 2016 to 2006, and this was accomplished by sequentially dropping age and length composition data, survey biomass estimates, and catch estimates from the input data files.

The plot of retrospective estimates of spawning biomass is shown in Figure 11. The 2016 model run shows the largest biomass than any of the retrospective runs, as new data in 2016 allows improved fit to the recent high AI trawl survey biomass index. Changes in the retrospective pattern occur in 2012 and 2010. The 2014 survey contains similar information as in the 2012 survey and thus has relatively little effect on the retrospective pattern.

The change in estimated spawning biomass from the 2009 to 2010 end years was particularly large, as the 2010 survey biomass estimate was substantially increased from the 2006 estimate. A series of exploratory models runs conducted in the 2010 assessment revealed that a combination of the high survey biomass and new observations of strong 1994-2000 year classes observed in both the fishery and survey age and length composition data lowered the estimates of survey catchability and increased estimated biomass.

Mohn's rho can be used to evaluate the severity of any retrospective pattern, and compares an estimated quantity (in this case, spawning stock biomass) in the terminal year of each retrospective model run with the estimated quantity in the same year of the model using the full data set. The absence of any retrospective pattern would result in a Mohn's rho of 0 , and would result from either identical estimates in the model runs, or from positive deviations from the reference model being offset by negative deviations. The Mohn's rho for this retrospective runs was -0.348 , similar to the value of -0.343 obtained in the 2014 assessment.

## Time series results

In this assessment, spawning biomass is defined as the biomass estimate of mature females age 3 and older. Total biomass is defined as the biomass estimate of POP age 3 and older. Recruitment is defined as the number of age 3 POP.

## Prior and Posterior Distributions

Posterior distributions for $M$, $q$, total 2016 biomass, and median recruitment, based upon the MCMC integrations, are shown in Figure 12.

## Biomass Trends

The AI survey biomass index begins with 1,163,030 t in 1960, declines to $238,384 \mathrm{t}$ in 1981, increases to $831,448 \mathrm{t}$ in 2010, and declines to $755,437 \mathrm{t}$ in 2016 (Figure 13). The relative proportion of the stock in the AI survey area between 1991 and 2016 ranges between 0.79 and 0.84 (Figure 14). The product of the survey catchability and the proportion available in the Aleutian Islands has ranged between 1.08 and 1.15 over these years, averaging 1.12. This is a decrease from the estimate of 1.28 in the 2014 model, but similar to the estimate of 1.12 from the 2012 model. One factor that may result in survey catchability being above 1 is the expansion of survey trawl estimates to untrawlable areas (Kreiger and Sigler 1996).

The predicted EBS survey biomass generally matches the observed data, although the high biomass in 2016 is not fit well due to its high CV (Figure 15). The estimate of EBS survey catchability was 1.88.

The total biomass showed a similar trend as the survey biomass, with the 2016 total biomass estimated as $783,492 \mathrm{t}$. The estimated time series of total biomass and spawning biomass, with $95 \%$ credibility bounds obtained from MCMC integration, are shown in Figure 16. Total biomass, spawning biomass, and recruitment (and their CVs from the Hessian approximation) are given in Table 13, and numbers at age are shown in Table 14.

## Age/size compositions

The fits to the fishery age and length composition is shown in Figures 17-18. The observed proportion in the binned length group of 39+ cm for 1964 and 1965 was lower than the estimated proportion, reflecting the modeling of the initial numbers at age as an equilibrium population. However, by 1966 reasonable fits were observed for the binned length group in the fishery length composition (Figure 18). Some of the lack of fit in the mid- to late-1980s is attributable to the low sample size of lengths observed from a reduced fishery. Good fits are obtained for most age groups in the 1991-2012 AI surveys (Figure 19), although 2004 and 2005 year classes are not well estimated in the 2014 AI survey age composition. The model does not fit a good fit to the 2016 length composition from the AI survey (Figure 20), resulting in a low weight for this data component.

The model captures the general pattern of the EBS survey age compositions, although the 1990 year class is underestimated across several years (Figure 21). The model provides a reasonable fit the 2016 EBS survey length composition (Figure 22).

## Fishing and Survey Selectivity

Younger fish show higher survey selection in the AI survey than in the EBS survey, with the ages at $50 \%$ selection estimated as 6.08 and 11.92 , respectively (Figure 23). The estimated fishery selectivity by age and year is shown in Figure 24, and shows pattern consistent with the empirical data in fishery catch examined above. Strong dome-shaped selectivity is estimated in the early 1960s to allow fish of age 20 older from this period to survive the large fully-selected fishing rates in the 1960s and early 1970s and be available for capture in the fishery and survey in the early 1980s (by which time they have entered the $40+$ group). The model estimates that dome-shaped selectivity has gradually become less peaked over time, and in recent year a slight reduction in selectivity exists for fish greater than age 35.

## Fishing Mortality

The estimates of instantaneous fishing mortality for POP range from highs during the 1970's to low levels in the 1980's (Figure 25). Fishing mortality rates since the early 1980's, however, have moderated considerably due to the phase out of the foreign fleets and quota limitations imposed by the North Pacific Fishery Management Council. Note that because of the change in the fishery selectivity over time, the fully-selected rates are not completely comparable over time with respect to the degree to which the stock
has been harvested. Nonetheless, the average fully-selected fishing mortality from 1965 to 1980 was 0.42 , whereas the average from 1981 to 2015 was 0.04 .

The plot of estimated fishing mortality rates and spawning stock biomass relative to the harvest control rules (Figure 26) indicate that BSAI POP would be considered overfished (using current definitions) during much of the period from the mid-1960s to the mid-1980s, although it should be noted the current definitions of $B_{35 \%}$ are based on the estimated recruitment of the post-1977 year classes and the average fishery selectivity from the most recent 5 years.

## Recruitment

Year-class strength varies widely for BSAI POP (Figure 27; Table 13). The relationship between spawning stock and recruitment also displays a high degree of variability (Figure 28). The 1957 and 1962 year classes are particularly large and sustained the heavy fishing in the 1960s. The rebuilding of the stock in the 1980s and 1990s was based upon recruitments for the 1981, 1984, 1986, and 1988-89 year classes. Recruitment appears to be lower in early 1990s, but several cohorts from 1994 to 2008 generally show relatively strong recruitment (with the exception the 1997 and 1999 year classes), which is consistent with the increasing trend of biomass and the fishery and AI survey age compositions shown in Figures 17 and 19. In particular, the largest estimated year class occurred in 2000, at 388 million. The 2004-05, and 2008 year classes are estimated as much stronger relative to the 2014 assessment model, although most of these cohorts have only been partially selected in the trawl surveys.

## Harvest recommendations

## Amendment 56 reference points

The reference fishing mortality rate for Pacific ocean perch is determined by the amount of reliable population information available (Amendment 56 of the Fishery Management Plan for the groundfish fishery of the Bering Sea/Aleutian Islands). Estimates of $F_{0.40}, F_{0.35}$, and $S P R_{0.40}$ were obtained from a spawner-per-recruit analysis. Assuming that the average recruitment from the 1977-2010 year classes estimated in this assessment represents a reliable estimate of equilibrium recruitment, then an estimate of $B_{0.40}$ is calculated as the product of $S P R_{0.40}$ * equilibrium recruits, and this quantity is $214,685 \mathrm{t}$. The year 2017 estimated spawning stock biomass is $314,489 \mathrm{t}$.

## Specification of OFL and maximum permissible ABC

Since reliable estimates of the 2017 spawning biomass ( $B$ ), $B_{0.40}, F_{0.40}$, and $F_{0.35}$ exist and $B>B_{0.40}$ ( $314,489 \mathrm{t}>214,685 \mathrm{t}$ ), POP reference fishing mortality have been classified in tier 3a. For this tier, $F_{A B C}$ maximum permissible $F_{A B C}$ is $F_{0.40}$, and $F_{O F L}$ is equal to $F_{0.35}$. The values of $F_{0.40}$ and $F_{0.35}$ are 0.082 and 0.101 , respectively.

The 2017 ABC associated with the $\boldsymbol{F}_{0.40}$ level of $\mathbf{0 . 0 8 3}$ is $\mathbf{4 3 , 7 2 3} \mathbf{t}$.
The estimated catch level for year 2017 associated with the overfishing level of $F=0.101$ is $53,152 \mathrm{t}$. A summary of these values is below.

$$
\begin{array}{clr}
2017 \text { SSB estimate (B) } & =314,489 \mathbf{t} \\
B_{0.40} & = & 214,685 \mathrm{t} \\
F_{\text {ABC }}=F_{0.40} & = & 0.082 \\
F_{\text {OFL }}=F_{0.35} & = & 0.101 \\
\text { MaxPermABC } & = & 43,723 \mathrm{t} \\
\text { OFL } & = & 43,723 \mathrm{t}
\end{array}
$$

## $A B C$ recommendation

We recommend the maximum permissible ABC 43,723 t in 2017.

## Projections

A standard set of projections is conducted for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with the vector of 2016 numbers at age estimated in the assessment. This vector is then projected forward to the beginning of 2017 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end) catch for 2016. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TAC for 2017, are as follow (" $m a x F_{A B C}$ " refers to the maximum permissible value of $F_{A B C}$ under Amendment 56):

Scenario 1: In all future years, $F$ is set equal to $\max F_{A B C}$. (Rationale: Historically, TAC has been constrained by ABC , so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, $F$ is set equal to a constant fraction of $\max F_{A B C}$, where this fraction is equal to the ratio of the $F_{A B C}$ value for 2017 recommended in the assessment to the max $F_{A B C}$ for 2015. (Rationale: When $F_{A B C}$ is set at a value below $\max F_{A B C}$, it is often set at the value recommended in the stock assessment.)

Scenario 3: In all future years, $F$ is set equal to $50 \%$ of $\max F_{A B C}$. (Rationale: This scenario provides a likely lower bound on $F_{A B C}$ that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 4: In all future years, $F$ is set equal to the 2011-2015 average $F$. (Rationale: For some stocks, TAC can be well below ABC, and recent average $F$ may provide a better indicator of $F_{\text {TAC }}$ than $F_{A B C}$.)

Scenario 5: In all future years, $F$ is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether the Pacific
ocean perch stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follow (for Tier 3 stocks, the MSY level is defined as $B_{35 \%}$ ):

Scenario 6: In all future years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be above 1) above its MSY level in 2016 or 2 ) above $1 / 2$ of its MSY level in 2016 and above its MSY level in 2016 under this scenario, then the stock is not overfished.)

Scenario 7: In 2017 and 2018, $F$ is set equal to $\max F_{A B C}$, and in all subsequent years $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2029 under this scenario, then the stock is not approaching an overfished condition.)

The recommended $F_{A B C}$ and the maximum $F_{A B C}$ are equivalent in this assessment, and projections of the mean harvest and spawning stock biomass for the remaining six scenarios are shown in Table 15.

## Status Determination

In addition to the seven standard harvest scenarios, Amendments $48 / 48$ to the BSAI and GOA Groundfish Fishery Management Plans require projections of the likely OFL two years into the future. While Scenario 6 gives the best estimate of OFL for 2017, it does not provide the best estimate of OFL for 2018, because the mean 2017 catch under Scenario 6 is predicated on the 2017 catch being equal to the 2017 OFL, whereas the actual 2017 catch will likely be less than the 2017 OFL. The executive summary contains the appropriate one- and two-year ahead projections for both ABC and OFL. Catches for 2017 and 2018 were obtained by setting the $F$ rate for these years to the estimated $F$ rate for 2016.

Under the MSFCMA, the Secretary of Commerce is required to report on the status of each U.S. fishery with respect to overfishing. This report involves the answers to three questions: 1 ) Is the stock being subjected to overfishing? 2) Is the stock currently overfished? 3) Is the stock approaching an overfished condition?

Is the stock being subjected to overfishing? The official BSAI catch estimate for the most recent complete year (2015) is $31,425 \mathrm{t}$. This is less than the 2015 BSAI OFL of $42,588 \mathrm{t}$. Therefore, the stock is not being subjected to overfishing.

Harvest Scenarios \#6 and \#7 are intended to permit determination of the status of a stock with respect to its minimum stock size threshold (MSST). Any stock that is below its MSST is defined to be overfished. Any stock that is expected to fall below its MSST in the next two years is defined to be approaching an overfished condition. Harvest Scenarios \#6 and \#7 are used in these determinations as follows:

Is the stock currently overfished? This depends on the stock's estimated spawning biomass in 2016:
a. If spawning biomass for 2016 is estimated to be below $1 / 2 B_{35 \%}$, the stock is below its MSST.
b. If spawning biomass for 2016 is estimated to be above $B_{35 \%}$ the stock is above its MSST.
c. If spawning biomass for 2016 is estimated to be above $1 / 2 B_{35 \%}$ but below $B_{35 \%}$, the stock's status relative to MSST is determined by referring to harvest Scenario \#6 (Table 15). If the mean spawning biomass for 2026 is below $B 35 \%$, the stock is below its MSST. Otherwise, the stock is above its MSST.

Is the stock approaching an overfished condition? This is determined by referring to harvest Scenario \#7:
a. If the mean spawning biomass for 2019 is below $1 / 2 B_{35 \%}$, the stock is approaching an overfished condition.
b. If the mean spawning biomass for 2019 is above $B 35 \%$, the stock is not approaching an overfished
condition.
c. If the mean spawning biomass for 2019 is above $1 / 2$ B35\% but below $B 35 \%$, the determination depends on the mean spawning biomass for 2029. If the mean spawning biomass for 2029 is below B35\%, the stock is approaching an overfished condition. Otherwise, the stock is not approaching an overfished condition.

The results of these two scenarios indicate that the BSAI POP stock is neither overfished nor approaching an overfished condition. With regard whether the stock is currently overfished, the expected stock size in the year 2016 of Scenario 6 is 1.72 times its $B_{35 \%}$ value of $187,849 \mathrm{t}$. With regard to whether the BSAI POP stock is likely to be overfished in the future, the expected stock size in 2019 of Scenario 7 is 1.52 times the $B_{35 \%}$ value.

## Area Allocation of Harvests

The ABC of BSAI POP is currently partitioned into subarea ABCs based on the relative biomass from research surveys. A random effects model is used to smooth subarea survey biomass estimates to obtain the proportions. This procedure assumes equivalent survey catchability and selectivity across subareas, such that any difference in survey biomass between areas can be attributed to true changes in biomass rather than differences in catchability and selectivity. In previous years this assumption was reasonable because the selectivity and catchability of the EBS slope survey had not been estimated. Estimates of these quantities are now available from this assessment, and indicate that the EBS slope survey has a higher catchability and a lower selectivity for young fish relative to the AI survey.

In order to use the survey biomass estimates to partition the ABC , we propose the following equation to produce an adjusted EBS survey biomass estimate in year $t\left(B_{a d j, t}\right)$ that is in comparable units to the AI survey:

$$
B_{a d j, t}=B_{t}\left(\frac{\sum_{a} q_{A I} s_{A I, a} w_{a} N_{a, t}}{\sum_{a} q_{E B S} s_{E B S, a} w_{a} N_{a, t}}\right)
$$

where $N_{a, t}$ is the estimated numbers at age, $s$ is selectivity, and $q$ is catchability, and $B_{t}$ is the smoothed unadjusted EBS survey slope estimate. The adjustment factor has varied since 1992, reaching a maximum of 1.1 in 2007 and declining to 0.94 in 2016 (Figure 29). The unadjusted smoothed EBS biomass of $245,905 \mathrm{t}$ in 2016 is lowered to an adjusted smoothed biomass of $230,736 \mathrm{t}$ :

|  |  | Area |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
|  | WAI | CAI | EAI | SBS | EBS slope |
| Unadjusted smoothed biomass | 356,896 | 216,425 | 278,507 | 83,742 | 245,905 |
| percentage | $30.21 \%$ | $18.32 \%$ | $23.57 \%$ | $7.09 \%$ | $20.81 \%$ |
|  |  |  |  |  |  |
| Adjusted smoothed biomass | 356,896 | 216,425 | 278,507 | 83,742 | 230,736 |
| percentage | $30.60 \%$ | $18.56 \%$ | $23.88 \%$ | $7.18 \%$ | $19.78 \%$ |

The apportioned ABCs for 2017 and 2018 from the two methods are as follows:

|  | Area |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
|  | WAI | CAI | EAI | EBS | Total |
|  | 13,208 | 8,009 | 10,307 | 12,199 | 43,723 |
| 2017 ABCs, unadjusted | 12,909 | 7,828 | 10,074 | 11,924 | 42,735 |
| 2018 ABCs, unadjusted |  |  |  |  |  |
|  | 13,380 | 8,113 | 10,441 | 11,789 | 43,723 |
| 2017 ABCs, adjusted | 13,077 | 7,930 | 10,205 | 11,523 | 42,735 |

## Ecosystem Considerations

## Ecosystem Effects on the stock

1) Prey availability/abundance trends

POP feed upon calanoid copepods, euphausids, myctophids, and other miscellaneous prey (Yang 2003). From a sample of 292 Aleutian Island specimens collected in 1997, calanoid copepods, euphausids, and myctophids contributed $70 \%$ of the total diet by weight. The diet of small POP was composed primarily of calanoid copepods ( $89 \%$ by weight), with euphausids and myctophids contributing approximately $35 \%$ and $10 \%$ of the diet, respectively, of larger POP. The availability and abundance trends of these prey species are unknown.

## 2) Predator population trends

POP are not commonly observed in field samples of stomach contents, although previous studies have identified sablefish, Pacific halibut, and sperm whales as predators (Major and Shippen 1970). The population trends of these predators can be found in separate chapters within this SAFE document.

## 3) Changes in habitat quality

POP appear to exhibit ontogenetic shifts in habitat use. Carlson and Straty (1981) used a submersible off southeast Alaska to observe juvenile red rockfish they believed to be POP at approximately $90-100 \mathrm{~m}$ in rugged habitat including boulder fields and rocky pinnacles. Kreiger (1993) also used a submersible to observe that the highest densities of small red rockfish in untrawlable rough habitat. As POP mature, they move into deeper and less rough habitats. Length frequencies of the Aleutian Islands survey data indicate that large POP (> 25 cm ) are generally found at depths greater than 150 m . Brodeur (2001) also found that POP was associated with epibenthic sea pens and sea whips along the Bering Sea slope. There has been little information identifying how rockfish habitat quality has changed over time.

## Fishery Effects on the ecosystem

Catch of prohibited species from 2003-2008 by fishery are available from the NMFS Regional Office. The rockfish fishery in the BSAI area, which consists only of the AI POP target fishery, contributed approximately $2 \%$ of the gold/brown king crab catch and approximately $1 \%$ of the halibut bycatch. For other prohibited species, the BSAI rockfish fisheries contributed much lower that $1 \%$ of the bycatch.

Estimates of non-target catches in the rockfish fishery are also available from the Catch Accounting System database maintained by the NMFS Regional Office. BSAI rockfish fisheries contribute mostly to the bycatch of coral, sponge, and polychaetes. From 2003 to 2008, the BSAI rockfish fisheries
contributed $31 \%$ of the coral and bryozoan bycatch, $18 \%$ of the sponge bycatch, $8 \%$ of the red tree coral bycatch, and $7 \%$ of the polychaete bycatch. The relative contribution was variable between years; for example, the annual relative contribution corals and bryozoans ranged from 5\% in 2004 to $53 \%$ in 2003, and the other groups listed above show similar levels of variability.

The POP fishery is not likely to diminish the amount of POP available as prey due to its low selectivity for fish less than 27 cm . Additionally, the fishery is not suspected of affecting the size-structure of the population due to the relatively light fishing mortality, averaging 0.04 over the last 5 years. It is not known what effects the fishery may have on the maturity-at-age of POP.

## Data Gaps and Research Priorities

Although Pacific ocean perch may be considered a "data-rich" species relative to other rockfish, little information is known regarding most aspects of their biology, including reproductive biology and the distribution, duration, and habitat requirements of various life-history stages. Given the relatively unusual reproductive biology of rockfish and its importance in establishing management reference points, data on reproductive capacity should be collected on a periodic basis.

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Table 1. Total allowable catch (TAC), acceptable biological catch (ABC), and catch of the species groups used to manage Pacific ocean perch from 1977 to 2001 in the Aleutian Islands and the eastern Bering Sea. The "POP complex" includes the other red rockfish species (shortraker rockfish, rougheye rockfish, northern rockfish, and sharpchin rockfish) plus POP.


Table 2. Total allowable catch (TAC), acceptable biological catch (ABC), and catch for BSAI POP from 2002 to present. Catch data is through October 10, 2016, from NMFS Alaska Regional Office.

Bering Sea/Aleutian Islands

| Management <br> Group | OFL (t) | ABC (t) | TAC $(\mathrm{t})$ | Catch $(\mathrm{t})$ |
| :---: | :---: | :---: | :---: | :---: |
| 2002 POP | 17500 | 14800 | 14800 | 11215 |
| 2003 POP | 18000 | 15100 | 14100 | 14744 |
| 2004 POP | 15800 | 13300 | 12580 | 11896 |
| 2005 POP | 17300 | 14600 | 12600 | 10427 |
| 2006 POP | 17600 | 14800 | 12600 | 12867 |
| 2007 POP | 26100 | 21900 | 19900 | 18451 |
| 2008 POP | 25700 | 21700 | 21700 | 17436 |
| 2009 POP | 22300 | 18800 | 18800 | 15347 |
| 2010 POP | 22400 | 18860 | 18860 | 17852 |
| 2011 POP | 36300 | 24700 | 24700 | 24004 |
| 2012 POP | 35000 | 24700 | 24700 | 24161 |
| 2013 POP | 41900 | 35100 | 35100 | 31362 |
| 2014 POP | 39585 | 33122 | 33122 | 32381 |
| 2015 POP | 42588 | 34988 | 32021 | 31425 |
| 2016 POP | 40529 | 33320 | 31900 | 24796 |

Table 3. Total allowable catch (TAC), acceptable biological catch (ABC), and catch of POP by area and management group from 1977 to 2016.

| Year | Eastern Bering Sea |  |  |  | Aleutian Islands |  |  |  | BSAI <br> Total catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Foreign | JVP |  | DAP | Foreign | JVP |  | DAP |  |
| 1977 | 2,406 |  | 0 |  | 7,927 |  | 0 |  | 10,333 |
| 1978 | 2,230 |  | 0 |  | 5,286 |  | 0 |  | 7,516 |
| 1979 | 1,722 |  | 0 |  | 5,486 |  | 0 |  | 7,208 |
| 1980 | 907 |  | 52 |  | 4,010 |  | 0 |  | 4,969 |
| 1981 | 1,185 |  | 1 |  | 3,668 |  | 0 |  | 4,854 |
| 1982 | 186 |  | 19 |  | 977 |  | 2 |  | 1,183 |
| 1983 | 99 |  | 93 |  | 463 |  | 8 |  | 663 |
| 1984 | 172 |  | 142 |  | 324 |  | 241 |  | 879 |
| 1985 | 30 |  | 31 |  | 0 |  | 216 |  | 277 |
| 1986 | 18 |  | 103 | 549 | 0 |  | 163 | 139 | 972 |
| 1987 | 5 |  | 49 | 1,123 | 0 |  | 502 | 554 | 2,233 |
| 1988 | 0 |  | 46 | 1,280 | 0 |  | 1,512 | 512 | 3,350 |
| 1989 | 0 |  | 26 | 2,507 | 0 |  | 0 | 2,963 | 5,496 |
| 1990 |  |  |  | 6,499 |  |  |  | 11,826 | 18,324 |
| 1991 |  |  |  | 5,099 |  |  |  | 2,785 | 7,884 |
| 1992 |  |  |  | 3,255 |  |  |  | 10,280 | 13,534 |
| 1993 |  |  |  | 3,764 |  |  |  | 13,376 | 17,139 |
| 1994 |  |  |  | 1,688 |  |  |  | 10,866 | 12,554 |
| 1995 |  |  |  | 1,208 |  |  |  | 10,304 | 11,511 |
| 1996 |  |  |  | 2,855 |  |  |  | 12,827 | 15,681 |
| 1997 |  |  |  | 681 |  |  |  | 12,648 | 13,329 |
| 1998 |  |  |  | 956 |  |  |  | 9,047 | 10,003 |
| 1999 |  |  |  | 421 |  |  |  | 12,484 | 12,905 |
| 2000 |  |  |  | 451 |  |  |  | 9,328 | 9,780 |
| 2001 |  |  |  | 896 |  |  |  | 8,557 | 9,453 |
| 2002 |  |  |  | 639 |  |  |  | 10,575 | 11,215 |
| 2003 |  |  |  | 1,145 |  |  |  | 13,600 | 14,744 |
| 2004 |  |  |  | 731 |  |  |  | 11,165 | 11,896 |
| 2005 |  |  |  | 879 |  |  |  | 9,548 | 10,427 |
| 2006 |  |  |  | 1,041 |  |  |  | 11,826 | 12,867 |
| 2007 |  |  |  | 870 |  |  |  | 17,581 | 18,451 |
| 2008 |  |  |  | 513 |  |  |  | 16,923 | 17,436 |
| 2009 |  |  |  | 623 |  |  |  | 14,725 | 15,347 |
| 2010 |  |  |  | 3,547 |  |  |  | 14,304 | 17,852 |
| 2011 |  |  |  | 5,601 |  |  |  | 18,403 | 24,004 |
| 2012 |  |  |  | 5,591 |  |  |  | 18,570 | 24,161 |
| 2013 |  |  |  | 5,051 |  |  |  | 26,311 | 31,362 |
| 2014 |  |  |  | 7,437 |  |  |  | 24,944 | 32,381 |
| 2015 |  |  |  | 7,918 |  |  |  | 23,507 | 31,425 |
| 2016* |  |  |  | 3,744 |  |  |  | 21,052 | 24,796 |

${ }^{*}$ Estimated removals through October 10, 2016.

Table 4. Estimated retained and discarded catch ( t ), and percent discarded, of Pacific ocean perch from the eastern Bering Sea (EBS) and Aleutian Islands (AI) regions.

| Year | EBS |  |  | AI |  |  | BSAI |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Retained | Discarded | Percent <br> Discarded | Retained | Discarded | Percent <br> Discarded | Retained | Discard | Percent <br> Discarded |
| 1990 | 5,069 | 1,275 | 20.1 | 10,288 | 1,551 | 13.1 | 15,357 | 2,826 | 15.54 |
| 1991 | 4,126 | 972 | 19.07 | 1,815 | 970 | 34.82 | 5,942 | 1,942 | 24.63 |
| 1992 | 2,732 | 522 | 16.05 | 8,666 | 1,614 | 15.7 | 11,398 | 2,136 | 15.78 |
| 1993 | 2,601 | 1,163 | 30.9 | 11,479 | 1,896 | 14.18 | 14,080 | 3,059 | 17.85 |
| 1994 | 1,187 | 501 | 29.69 | 9,491 | 1,375 | 12.65 | 10,678 | 1,876 | 14.94 |
| 1995 | 839 | 368 | 30.49 | 8,603 | 1,701 | 16.51 | 9,442 | 2,069 | 17.97 |
| 1996 | 2,522 | 333 | 11.66 | 9,831 | 2,995 | 23.35 | 12,353 | 3,328 | 21.22 |
| 1997 | 420 | 261 | 38.35 | 10,854 | 1,794 | 14.18 | 11,274 | 2,055 | 15.42 |
| 1998 | 813 | 143 | 19.62 | 8,041 | 1,006 | 10.93 | 8,854 | 1,149 | 11.79 |
| 1999 | 277 | 144 | 34.28 | 10,985 | 1,499 | 12.01 | 11,261 | 1,644 | 12.73 |
| 2000 | 230 | 221 | 49.01 | 8,586 | 743 | 7.96 | 8,816 | 964 | 9.85 |
| 2001 | 399 | 497 | 55.45 | 7,195 | 1,362 | 15.92 | 7,594 | 1,859 | 19.66 |
| 2002 | 286 | 354 | 55.44 | 9,315 | 1,260 | 11.91 | 9,601 | 1,614 | 14.4 |
| 2003 | 564 | 581 | 53.31 | 11,558 | 2,042 | 16 | 12,122 | 2,622 | 19.14 |
| 2004 | 536 | $196{ }^{\prime \prime}$ | - 26.75 | 9,286 | 1,879 | 16.83 | 9,822 | 2,074 | 17.44 |
| 2005 | 627 | 253 | - 28.74 | 8,100 | 1,448 | 15.16 | 8,727 | 1,700 | 16.31 |
| 2006 | 751 | $290{ }^{\prime \prime}$ | - 27.83 | 9,869 | 1,957 | 16.55 | 10,620 | 2,246 | 17.46 |
| 2007 | 508 | 363 " | - 41.68 | 15,051 | 2,530 | 14.39 | 15,558 | 2,893 | 15.68 |
| 2008 | 318 | $195{ }^{\prime \prime}$ | - 37.94 | 16,640 | 283 | 1.67 | 16,959 | 477 | 2.74 |
| 2009 | 463 | $160{ }^{\prime \prime}$ | - 25.67 | 14,011 | 713 | 4.84 | 14,474 | 873 | 5.69 |
| 2010 | 3,347 | $200{ }^{\circ}$ | - 5.64 | 13,988 | 316 | 2.21 | 17,335 | 516 | 2.89 |
| 2011 | 5,249 | 353 " | - 6.30 | 18,021 | 382 | 2.08 | 23,269 | 735 | 3.06 |
| 2012 | 5,182 | $408{ }^{\text {² }}$ | - 7.31 | 18,169 | 401 | 2.16 | 23,352 | 810 | 3.35 |
| 2013 | 4,746 | 304 | - 6.03 | 26,063 | 249 | 0.94 | 30,809 | 553 | 1.76 |
| 2014 | 6,614 | 823 " | - 11.07 | 24,770 | 174 | 0.70 | 31,384 | 997 | 3.08 |
| 2015 | 6,749 | 1,169 ${ }^{\prime}$ | - 14.77 | 23,267 | 240 | 1.02 | 30,016 | 1,409 | 4.48 |
| 2016* | 3,085 | 659 | 17.59 | 20,903 | 149 | 0.71 | 23,988 | 807 | 3.26 |

*Estimated removals through October 10, 2016.
Source: NMFS Alaska Regional Office

Table 5. Percentage catch (by weight) of Aleutians Islands POP in the foreign/joint venture fisheries and the domestic fishery by depth.

| Year | 0 | Depth Zone (m) |  |  |  |  |  | Observed Estimated |  | Percent sampled |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  | Observed catch (t) | stimated total |  |
|  |  | 100 | 200 | 300 | 400 | 500 | 501 |  | catch |  |
| 1977 | 25 | 23 | 39 | 11 | 2 | 1 | 0 | 173 | 7,927 | 2 |
| 1978 | 0 | 40 | 36 | 19 | 3 | 1 | 1 | 145 | 5,286 | 3 |
| 1979 | 0 | 13 | 60 | 23 | 4 | 0 | 0 | 311 | 5,486 | 6 |
| 1980 | 0 | 7 | 45 | 49 | 0 | 0 | 0 | 108 | 4,010 | 3 |
| 1981 | 0 | 9 | 67 | 23 | 0 | 0 | 0 | 138 | 3,668 | 4 |
| 1982 | 0 | 34 | 56 | 5 | 2 | 1 | 2 | 115 | 979 | 12 |
| 1983 | 0 | 11 | 85 | 0 | 1 | 1 | 1 | 54 | 471 | 11 |
| 1984 | 0 | 53 | 42 | 5 | 0 | 1 | 0 | 85 | 565 | 15 |
| 1985 | 0 | 87 | 13 | 0 | 0 | 0 | 0 | 109 | 216 | 50 |
| 1986 | 0 | 74 | 25 | 2 | 0 | 0 | 0 | 66 | 163 | 40 |
| 1987 | 0 | 39 | 61 | 0 | 0 | 0 | 0 | 258 | 502 | 51 |
| 1988 | 0 | 78 | 21 | 1 | 0 | 0 | 0 | 76 | 1,512 | 5 |
| 1989 |  |  |  |  |  |  |  |  |  |  |
| 1990 | 2 | 23 | 58 | 14 | 2 | 1 | 0 | 7,726 | 11,826 | 65 |
| 1991 | 0 | 23 | 70 | 5 | 1 | 1 | 0 | 1,588 | 2,785 | 57 |
| 1992 | 0 | 21 | 71 | 8 | 0 | 0 | 0 | 6,785 | 10,280 | 66 |
| 1993 | 0 | 20 | 77 | 3 | 0 | 0 | 0 | 8,867 | 13,376 | 66 |
| 1994 | 0 | 20 | 69 | 11 | 0 | 0 | 0 | 7,562 | 10,866 | 70 |
| 1995 | 0 | 15 | 68 | 14 | 2 | 0 | 0 | 6,154 | 10,304 | 60 |
| 1996 | 0 | 17 | 54 | 26 | 2 | 1 | 0 | 8,547 | 12,827 | 67 |
| 1997 | 0 | 13 | 66 | 21 | 0 | 0 | 0 | 9,320 | 12,648 | 74 |
| 1998 | 0 | 21 | 72 | 7 | 0 | 0 | 0 | 7,380 | 9,047 | 82 |
| 1999 | 0 | 30 | 63 | 7 | 0 | 0 | 0 | 10,369 | 12,484 | 83 |
| 2000 | 0 | 21 | 63 | 15 | 0 | 0 | 0 | 7,456 | 9,328 | 80 |
| 2001 | 0 | 29 | 61 | 10 | 0 | 0 | 0 | 5,679 | 8,557 | 66 |
| 2002 | 2 | 36 | 57 | 5 | 1 | 0 | 0 | 8,124 | 10,575 | 77 |
| 2003 | 0 | 26 | 70 | 3 | 0 | 0 | 0 | 11,266 | 13,600 | 83 |
| 2004 | 1 | 26 | 65 | 7 | 1 | 0 | 0 | 10,083 | 11,165 | 90 |
| 2005 | 2 | 36 | 55 | 6 | 1 | 0 | 0 | 7,403 | 9,548 | 78 |
| 2006 | 1 | 33 | 61 | 5 | 0 | 0 | 0 | 9,895 | 11,826 | 84 |
| 2007 | 0 | 23 | 68 | 7 | 1 | 0 | 0 | 15,551 | 17,581 | 88 |
| 2008 | 1 | 20 | 74 | 5 | 0 | 0 | 0 | 16,685 | 16,923 | 99 |
| 2009 | 1 | 26 | 65 | 8 | 1 | 0 | 1 | 14,495 | 14,725 | 98 |
| 2010 | 1 | 21 | 71 | 7 | 1 | 0 | 0 | 14,299 | 14,304 | 100 |
| 2011 | 0 | 13 | 78 | 7 | 1 | 0 | 0 | 18,391 | 18,403 | 100 |
| 2012 | 0 | 22 | 67 | 11 | 1 | 0 | 0 | 18,569 | 18,570 | 100 |
| 2013 | 0 | 12 | 76 | 11 | 1 | 0 | 0 | 26,297 | 26,311 | 100 |
| 2014 | 0 | 12 | 79 | 8 | 0 | 0 | 0 | 24,882 | 24,944 | 100 |
| 2015 | 1 | 21 | 73 | 4 | 0 | 0 | 0 | 23,421 | 23,507 | 100 |

Table 6. Proportional catch (by weight) of Aleutians Islands POP in the foreign and joint venture fisheries and the domestic fishery by management area.

| Area |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 541 | 542 | 543 | Observed catch ( t ) | Estimated total catch | Percent sampled |
| 1977 | 17 | 22 | 61 | 173 | 7,927 | 2 |
| 1978 | 30 | 36 | 35 | 145 | 5,286 | 3 |
| 1979 | 21 | 25 | 55 | 311 | 5,486 | 6 |
| 1980 | 11 | 42 | 47 | 108 | 4,010 | 3 |
| 1981 | 42 | 40 | 17 | 138 | 3,668 | 4 |
| 1982 | 42 | 38 | 20 | 115 | 979 | 12 |
| 1983 | 85 | 8 | 7 | 54 | 471 | 11 |
| 1984 | 84 | 8 | 7 | 85 | 565 | 15 |
| 1985 | 66 | 34 | 0 | 109 | 216 | 50 |
| 1986 | 99 | 1 | 0 | 66 | 163 | 40 |
| 1987 | 94 | 6 | 0 | 258 | 502 | 51 |
| 1988 | 6 | 94 | 0 | 76 | 1,512 | 5 |
| 1989 |  |  |  |  |  |  |
| 1990 | 63 | 16 | 21 | 7,726 | 11,826 | 65 |
| 1991 | 27 | 57 | 16 | 1,588 | 2,785 | 57.0276 |
| 1992 | 81 | 15 | 3 | 6,785 | 10,280 | 66.0025 |
| 1993 | 67 | 22 | 11 | 8,867 | 13,376 | 66.2949 |
| 1994 | 64 | 31 | 5 | 7,562 | 10,866 | 69.5992 |
| 1995 | 70 | 25 | 5 | 6,154 | 10,304 | 59.7296 |
| 1996 | 27 | 20 | 54 | 8,547 | 12,827 | 66.6318 |
| 1997 | 20 | 23 | 57 | 9,320 | 12,648 | 73.6868 |
| 1998 | 21 | 27 | 52 | 7,380 | 9,047 | 81.5705 |
| 1999 | 22 | 23 | 56 | 10,369 | 12,484 | 83.0618 |
| 2000 | 22 | 24 | 54 | 7,456 | 9,328 | 79.9303 |
| 2001 | 27 | 25 | 48 | 5,679 | 8,557 | 66.3676 |
| 2002 | 24 | 28 | 48 | 8,124 | 10,575 | 76.8175 |
| 2003 | 30 | 22 | 48 | 11,266 | 13,600 | 82.841 |
| 2004 | 24 | 27 | 49 | 10,083 | 11,165 | 90.3064 |
| 2005 | 23 | 24 | 52 | 7,403 | 9,548 | 77.5385 |
| 2006 | 24 | 28 | 48 | 9,895 | 11,826 | 83.67 |
| 2007 | 30 | 26 | 45 | 15,551 | 17,581 | 88.455 |
| 2008 | 28 | 28 | 44 | 16,685 | 16,923 | 98.5931 |
| 2009 | 27 | 28 | 44 | 14,495 | 14,725 | 98.4395 |
| 2010 | 28 | 28 | 44 | 14,299 | 14,304 | 99.9622 |
| 2011 | 30 | 26 | 44 | 18,391 | 18,403 | 99.935 |
| 2012 | 30 | 26 | 44 | 18,569 | 18,570 | 99.9942 |
| 2013 | 36 | 26 | 38 | 26,297 | 26,311 | 99.9441 |
| 2014 | 36 | 26 | 38 | 24,882 | 24,944 | 99.7533 |
| 2015 | 33 | 29 | 38 | 23,421 | 23,507 | 99.6354 |

Table 7. Length measurements from the EBS and AI POP fisheries during 1964-1972, from Chikuni (1975)

| Year | EBS | AI | Total |
| ---: | ---: | ---: | ---: |
| 1964 | 24,150 | 55,599 | 79,749 |
| 1965 | 14,935 | 66,120 | 81,055 |
| 1966 | 26,458 | 25,502 | 51,960 |
| 1967 | 48,027 | 59,576 | 107,603 |
| 1968 | 38,370 | 36,734 | 75,104 |
| 1969 | 28,774 | 27,206 | 55,980 |
| 1970 | 11,299 | 27,508 | 38,807 |
| 1971 | 14,045 | 18,926 | 32,971 |
| 1972 | 10,996 | 18,926 | 29,922 |

Table 8. Length measurements and otoliths read from the EBS and AI POP fisheries, from the NORPAC Observer database.

*Used to create age composition. ${ }^{* *}$ Not used.

Table 9. Pacific ocean perch biomass estimates ( t ) from the 1991-2016 triennial trawl surveys for the three management sub-areas in the Aleutian Islands region, and the 2002-2016 EBS slope survey.

| Aleutian Islands Survey |  |  |  |  |  |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: |
| Year | Western | Central |  |  |  |  |  | Eastern | southern BS | Total AI survey | EBS slope survey |
| 1991 | $208,465(0.31)$ | $78,776(0.25)$ | $55,545(0.40)$ | $1,501(0.51)$ | $344,286(0.21)$ |  |  |  |  |  |  |
| 1994 | $184,703(0.39)$ | $84,411(0.33)$ | $100,585(0.42)$ | $18,217(0.64)$ | $387,916(0.23)$ |  |  |  |  |  |  |
| 1997 | $178,437(0.19)$ | $166,816(0.28)$ | $220,633(0.29)$ | $12,099(0.58)$ | $577,984(0.15)$ |  |  |  |  |  |  |
| 2000 | $229,850(0.32)$ | $129,740(0.32)$ | $140,528(0.25)$ | $18,870(0.54)$ | $518,988(0.18)$ | $72,665(0.53)$ |  |  |  |  |  |
| 2002 | $196,704(0.26)$ | $140,361(0.41)$ | $109,795(0.14)$ | $16,311(0.41)$ | $463,171(0.17)$ | $112,273(0.38)$ |  |  |  |  |  |
| 2004 | $212,639(0.21)$ | $153,477(0.17)$ | $137,112(0.29)$ | $74,208(0.45)$ | $577,436(0.13)$ |  |  |  |  |  |  |
| 2006 | $278,990(0.16)$ | $170,942(0.23)$ | $190,752(0.37)$ | $23,701(0.47)$ | $664,384(0.14)$ | $107,886(0.41)$ |  |  |  |  |  |
| 2008 |  |  |  |  |  |  |  |  |  |  |  |
| 2010 | $395,944(0.21)$ | $221,700(0.17)$ | $266,607(0.18)$ | $87,794(0.55)$ | $972,046(0.12)$ | $203,421(0.38)$ |  |  |  |  |  |
| 2012 | $263,661(0.23)$ | $233,666(0.17)$ | $366,413(0.36)$ | $38,658(0.63)$ | $902,398(0.17)$ | $231,046(0.38)$ |  |  |  |  |  |
| 2014 | $338,455(0.21)$ | $315,544(0.49)$ | $233,560(0.28)$ | $83,409(0.50)$ | $970,968(0.19)$ |  |  |  |  |  |  |
| 2016 | $403,049(0.19)$ | $206,593(0.19)$ | $284,909(0.17)$ | $87,952(0.47)$ | $982,503(0.11)$ | $357,369(0.68)$ |  |  |  |  |  |

Table 10. Length measurements and otoliths read from the Aleutian Islands surveys.

| Aleutian Islands survey |  | Eastern Bering Sea slope <br> survey |  |  |
| :--- | ---: | ---: | ---: | ---: |
| Year | Length | Otoliths read | Length Otoliths read |  |
| 1980 | 20,796 | 890 |  |  |
| 1983 | 22,873 | 2,495 |  |  |
| 1986 | 14,804 | 1,860 |  |  |
| 1991 | 14,262 | 1,015 |  |  |
| 1994 | 18,922 | 849 |  | 425 |
| 1997 | 22,823 | 1,224 |  | 413 |
| 2000 | 21,972 | 1,238 |  | 415 |
| 2002 | 20,284 | 337 | 2,040 | 4,084 |
| 2004 | 24,949 | 1,031 |  | 472 |
| 2006 | 19,737 | 462 |  | 2,818 |
| 2008 |  |  | 3,348 |  |
| 2010 | 22,725 | 951 | 3,459 | 4,398 |
| 2012 | 31,450 | 1,140 | 3,398 |  |
| 2014 | 30,204 | 1,078 |  |  |
| 2016 | 36,277 |  |  |  |

Table 11. Negative log likelihoods, root mean squared errors, and estimates and CV for key model quantities, for BSAI POP models.

| Model 0 |  | Model 14 | Model 16.1 | Model 16.2 | Model 16.3 | Model 16.4 | Model 16.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Negative log-likelihood |  |  |  |  |  |  |  |
| Data components |  |  |  |  |  |  |  |
| AI survey biomass | 8.52 | 8.71 | 8.80 | 8.13 | 8.65 | 9.75 | 3.72 |
| CPUE | 26.28 | 26.02 | 26.26 |  |  |  |  |
|  |  |  | 1.46 | 1.41 | 1.40 | 1.48 | 1.17 |
| Catch biomass | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Fishery age comp | 226.03 | 232.47 | 232.34 | 233.10 | 206.52 | 283.44 | 152.21 |
| Fishery length comp | 358.98 | 363.90 | 363.97 | 363.98 | 248.72 | 387.19 | 81.60 |
| AI survey age comp | 150.23 | 177.95 | 177.62 | 176.63 | 131.25 | 203.03 | 30.26 |
| AI survey lengths comp | 10.54 | 12.91 | 12.97 | 13.21 | 4.43 | 11.22 | 7.30 |
| EBS survey age comp |  |  | 14.96 | 14.98 | 64.64 | 94.56 | 45.22 |
| EBS survey lengths comp |  |  | 4.13 | 4.14 | 7.19 | 14.71 | 10.81 |
| Maturity | 2.71 | 2.71 | 2.71 | 2.71 | 2.71 | 2.71 | 2.71 |
| Priors and penalties |  |  |  |  |  |  |  |
| Recruitment | 11.95 | 11.16 | 11.42 | 11.42 | 9.62 | 11.51 | 8.61 |
| Prior on survey q | 9.89 | 9.74 | 9.06 | 8.62 | 5.00 | 7.86 | 0.55 |
| Prior on M | 0.30 | 0.21 | 0.49 | 0.37 | 0.42 | 0.39 | 1.42 |
| Fishery selectivity | 142.86 | 125.90 | 127.94 | 128.53 | 108.78 | 139.22 | 43.62 |
| Total negative log-likelihood | 942.41 | 965.77 | 988.09 | 974.36 | 806.38 | 1174.07 | 395.59 |
| Parameters | 137 | 141 | 144 | 143 | 143 | 143 | 143 |
| Root mean square error |  |  |  |  |  |  |  |
| AI survey biomass | 0.222 | 0.196 | 0.191 | 0.184 | 0.188 | 0.201 | 0.122 |
| EBS survey biomass |  |  | 0.360 | 0.353 | 0.353 | 0.363 | 0.316 |
| CPUE | 0.804 | 0.802 | 0.804 |  |  |  |  |
| Recruitment | 0.813 | 0.797 | 0.800 | 0.800 | 0.777 | 0.802 | 0.763 |
| Fishery age comp | 0.014 | 0.013 | 0.013 | 0.013 | 0.013 | 0.013 | 0.012 |
| Fishery length comp | 0.023 | 0.021 | 0.021 | 0.021 | 0.022 | 0.021 | 0.027 |
| AI survey age comp | 0.011 | 0.011 | 0.011 | 0.011 | 0.011 | 0.011 | 0.013 |
| AI survey lengths comp | 0.026 | 0.023 | 0.023 | 0.023 | 0.026 | 0.026 | 0.016 |
| EBS survey age comp |  |  | 0.016 | 0.016 | 0.015 | 0.015 | 0.016 |
| EBS survey lengths comp |  |  | 0.017 | 0.017 | 0.017 | 0.017 | 0.015 |
| Estimated key quantities |  |  |  |  |  |  |  |
| M | 0.062 | 0.062 | 0.062 | 0.061 | 0.058 | 0.061 | 0.053 |
| CV | 0.030 | 0.027 | 0.027 | 0.028 | 0.031 | 0.026 | 0.046 |
| AI survey $q$ | 1.280 | 1.211 | 1.408 | 1.328 | 1.367 | 1.342 | 1.921 |
| CV | 0.140 | 0.133 | 0.134 | 0.142 | 0.147 | 0.135 | 0.153 |
| 2016 total biomass( $t$ ) |  | 719,310 | 766,840 | 821,540 | 783,490 | 783,730 | 630,290 |
| CV |  | 0.159 | 0.158 | 0.167 | 0.171 | 0.161 | 0.171 |

Table 12. Estimated parameter values and standard deviations for the BSAI POP assessment model.

| Parameter | Standard |  |  | Standard |  |  | Standard |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Estimate | Deviation | Parameter | Estimate | Deviation | Parameter | Estimate | Deviation |
| sel_par | -2.5797 | 0.1997 | fmort_dev | -1.5094 | 0.3020 | rec_dev | -0.4250 | 0.3614 |
| sel_par | -1.1072 | 0.1363 | fmort_dev | -2.1591 | 0.3014 | rec_dev | -0.6369 | 0.4190 |
| sel_par | -2.9869 | 0.1329 | fmort_dev | -1.9563 | 0.3008 | rec_dev | -0.4730 | 0.3055 |
| sel_par | -2.5295 | 0.2862 | fmort_dev | -3.1935 | 0.3004 | rec_dev | -1.1767 | 0.4054 |
| sel_par | 1.9985 | 0.1110 | fmort_dev | -2.0248 | 0.3002 | rec_dev | -1.0137 | 0.3175 |
| sel_par | 1.0326 | 0.0747 | fmort_dev | -1.2924 | 0.3000 | rec_dev | -1.0667 | 0.3531 |
| sel_par | 0.7676 | 0.0619 | fmort_dev | -0.9828 | 0.3000 | rec_dev | -0.1261 | 0.2208 |
| sel_par | -0.0251 | 0.1155 | fmort_dev | -0.5862 | 0.3000 | rec_dev | -0.2502 | 0.2929 |
| sel_par | 0.3278 | 0.1151 | fmort_dev | 0.5409 | 0.2999 | rec_dev | -0.5684 | 0.4452 |
| sel_par | 0.0807 | 0.0780 | fmort_dev | -0.3851 | 0.3000 | rec_dev | -0.0528 | 0.4396 |
| sel_par | 0.3415 | 0.0618 | fmort_dev | 0.0625 | 0.3001 | rec_dev | 0.2855 | 0.4212 |
| sel_par | 0.5328 | 0.1252 | fmort_dev | 0.2191 | 0.3002 | rec_dev | 0.6582 | 0.3059 |
| sel_par | -0.6935 | 0.1277 | fmort_dev | -0.1702 | 0.3004 | rec_dev | 0.0882 | 0.4177 |
| sel_par | -0.3308 | 0.0790 | fmort_dev | -0.3358 | 0.3006 | rec_dev | -0.2067 | 0.4604 |
| sel_par | 0.1959 | 0.0765 | fmort_dev | -0.0919 | 0.3009 | rec_dev | 1.4813 | 0.1207 |
| sel_par | 0.5189 | 0.1256 | fmort_dev | -0.3039 | 0.3011 | rec_dev | -0.2442 | 0.4891 |
| sel_par | -1.4230 | 0.2061 | fmort_dev | -0.6314 | 0.3013 | rec_dev | 0.6878 | 0.2035 |
| sel_par | -0.8472 | 0.1103 | fmort_dev | -0.4020 | 0.3014 | rec_dev | -0.1067 | 0.3906 |
| sel_par | -0.1748 | 0.1149 | fmort_dev | -0.6895 | 0.3015 | rec_dev | 1.0669 | 0.1542 |
| sel_par | 0.1675 | 0.2125 | fmort_dev | -0.7240 | 0.3016 | rec_dev | 0.4537 | 0.2566 |
| sel_aslope_ai | 0.8504 | 0.0756 | fmort_dev | -0.5416 | 0.3016 | rec_dev | -0.0799 | 0.3043 |
| sel_a50_ai | 6.0783 | 0.1920 | fmort_dev | -0.2489 | 0.3017 | rec_dev | -0.8087 | 0.3938 |
| sel_aslope_srv_ebs | 0.6269 | 0.0758 | fmort_dev | -0.4477 | 0.3019 | rec_dev | -0.3519 | 0.2644 |
| sel_a50_srv_ebs | 11.9220 | 0.4836 | fmort_dev | -0.5794 | 0.3021 | rec_dev | -0.5252 | 0.3491 |
| $\log M$ | -2.8388 | 0.0309 | fmort_dev | -0.3822 | 0.3024 | rec_dev | 0.6121 | 0.1738 |
| log_avg_fmort | -3.8507 | 0.3149 | fmort_dev | -0.0407 | 0.3028 | rec_dev | 0.3573 | 0.2587 |
| fmort_dev | -2.2005 | 0.3136 | fmort_dev | -0.1200 | 0.3035 | rec_dev | 1.1371 | 0.1434 |
| fmort_dev | -0.1078 | 0.3134 | fmort_dev | -0.2753 | 0.3042 | rec_dev | -0.3236 | 0.4170 |
| fmort_dev | -0.8900 | 0.3132 | fmort_dev | -0.1502 | 0.3051 | rec_dev | 1.0699 | 0.1512 |
| fmort_dev | -0.0182 | 0.3129 | fmort_dev | 0.1304 | 0.3062 | rec_dev | -0.0597 | 0.4094 |
| fmort_dev | 1.0938 | 0.3119 | fmort_dev | 0.1325 | 0.3076 | rec_dev | 1.6876 | 0.1232 |
| fmort_dev | 1.5117 | 0.3097 | fmort_dev | 0.4016 | 0.3094 | rec_dev | -0.2958 | 0.5057 |
| fmort_dev | 1.6843 | 0.3081 | fmort_dev | 0.4542 | 0.3116 | rec_dev | 0.5797 | 0.2446 |
| fmort_dev | 1.6121 | 0.3073 | fmort_dev | 0.4503 | 0.3141 | rec_dev | -0.4284 | 0.5014 |
| fmort_dev | 1.8261 | 0.3068 | fmort_dev | 0.4797 | 0.3170 | rec_dev | 0.9102 | 0.2317 |
| fmort_dev | 1.5804 | 0.3061 | rec_dev | 1.1181 | 0.2556 | rec_dev | 0.7435 | 0.2864 |
| fmort_dev | 2.0406 | 0.3053 | rec_dev | -0.2803 | 0.6184 | rec_dev | -0.0808 | 0.4460 |
| fmort_dev | 1.2139 | 0.3052 | rec_dev | -0.3551 | 0.5818 | rec_dev | -0.0615 | 0.4283 |
| fmort_dev | 1.4500 | 0.3052 | rec_dev | -0.1285 | 0.6645 | rec_dev | 0.6705 | 0.3046 |
| fmort_dev | 0.5682 | 0.3053 | rec_dev | 0.8930 | 0.5445 | rec_dev | -0.4077 | 0.5320 |
| fmort_dev | 1.5181 | 0.3050 | rec_dev | 1.6181 | 0.2644 | rec_dev | -0.4392 | 0.5441 |
| fmort_dev | 1.3216 | 0.3045 | rec_dev | -0.3867 | 0.6155 | mean_log_rec | 4.2721 | 0.0927 |
| fmort_dev | 1.6348 | 0.3043 | rec_dev | -0.6163 | 0.5352 | log_rinit | 4.3551 | 0.0720 |
| fmort_dev | 0.7240 | 0.3045 | rec_dev | -0.5506 | 0.4912 | logq_ai | 0.3127 | 0.1466 |
| fmort_dev | 0.4340 | 0.3040 | rec_dev | -0.5807 | 0.4588 | logq_srv_ebs | 0.6320 | 0.2326 |
| fmort_dev | 0.3954 | 0.3035 | rec_dev | -0.9198 | 0.4490 | mat_beta 1 | -6.6118 | 3.6559 |
| fmort_dev | 0.0075 | 0.3031 | rec_dev | -1.1847 | 0.4509 | mat_beta2 | 0.7270 | 0.4473 |
| fmort_dev | -0.0466 | 0.3026 | rec_dev | -0.9071 | 0.4342 |  |  |  |

Table 13. Estimated time series of POP total biomass ( t ), spawning biomass ( t ), and recruitment (thousands).

|  | Total Biomass (ages 3+) |  |  |  | Spawner Biomass (ages 3+) |  |  |  | Recruitment (age 3) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Assessment Year |  |  |  | Assessment Year |  |  |  | Assessment Year |  |  |  |
|  | 2016 |  | 2014 |  | 2016 |  | 2014 |  | 2016 |  | 2014 |  |
| Year | Est | CV | Est | CV | Est | CV | Est | CV | Est | CV | Est | CV |
| 1977 | 274,888 | 0.100 | 238,630 | 0.092 | 116,033 | 0.113 | 99,038 | 0.108 | 26,008 | 0.326 | 25,581 | 0.305 |
| 1978 | 262,984 | 0.102 | 226,774 | 0.094 | 110,276 | 0.117 | 93,438 | 0.113 | 24,665 | 0.364 | 23,814 | 0.352 |
| 1979 | 257,861 | 0.104 | 221,931 | 0.095 | 106,198 | 0.122 | 89,477 | 0.118 | 63,180 | 0.231 | 65,542 | 0.210 |
| 1980 | 254,398 | 0.105 | 218,921 | 0.096 | 103,135 | 0.127 | 86,494 | 0.124 | 55,807 | 0.302 | 57,894 | 0.279 |
| 1981 | 253,569 | 0.105 | 218,423 | 0.096 | 101,289 | 0.131 | 84,678 | 0.127 | 40,599 | 0.454 | 38,725 | 0.474 |
| 1982 | 256,237 | 0.105 | 221,874 | 0.096 | 100,247 | 0.133 | 83,672 | 0.130 | 67,986 | 0.444 | 72,287 | 0.412 |
| 1983 | 267,115 | 0.104 | 231,810 | 0.093 | 100,882 | 0.134 | 84,372 | 0.132 | 95,357 | 0.430 | 78,598 | 0.476 |
| 1984 | 285,325 | 0.101 | 250,389 | 0.091 | 102,464 | 0.138 | 86,102 | 0.138 | 138,426 | 0.313 | 145,384 | 0.277 |
| 1985 | 302,560 | 0.099 | 267,617 | 0.089 | 105,161 | 0.144 | 89,001 | 0.148 | 78,283 | 0.426 | 76,490 | 0.412 |
| 1986 | 320,399 | 0.098 | 285,047 | 0.087 | 109,328 | 0.152 | 93,376 | 0.158 | 58,289 | 0.473 | 51,875 | 0.477 |
| 1987 | 360,821 | 0.096 | 324,368 | 0.085 | 114,775 | 0.160 | 98,946 | 0.167 | 315,248 | 0.141 | 311,366 | 0.126 |
| 1988 | 387,848 | 0.096 | 349,846 | 0.085 | 121,516 | 0.171 | 105,631 | 0.177 | 56,146 | 0.506 | 45,770 | 0.511 |
| 1989 | 420,835 | 0.095 | 381,769 | 0.085 | 129,852 | 0.183 | 113,713 | 0.189 | 142,581 | 0.217 | 140,206 | 0.191 |
| 1990 | 447,206 | 0.096 | 406,537 | 0.086 | 138,411 | 0.193 | 121,809 | 0.200 | 64,421 | 0.407 | 59,069 | 0.383 |
| 1991 | 471,337 | 0.100 | 427,526 | 0.090 | 145,982 | 0.202 | 128,713 | 0.209 | 208,307 | 0.179 | 192,205 | 0.164 |
| 1992 | 501,836 | 0.101 | 454,119 | 0.092 | 157,974 | 0.209 | 139,920 | 0.216 | 112,821 | 0.277 | 92,298 | 0.270 |
| 1993 | 522,102 | 0.104 | 470,393 | 0.094 | 169,576 | 0.214 | 150,480 | 0.221 | 66,169 | 0.319 | 52,015 | 0.314 |
| 1994 | 532,982 | 0.107 | 477,652 | 0.098 | 181,230 | 0.213 | 160,910 | 0.218 | 31,925 | 0.410 | 26,405 | 0.392 |
| 1995 | 545,964 | 0.109 | 486,685 | 0.101 | 194,101 | 0.204 | 172,343 | 0.207 | 50,410 | 0.282 | 43,833 | 0.264 |
| 1996 | 555,839 | 0.111 | 492,300 | 0.103 | 206,095 | 0.196 | 182,568 | 0.196 | 42,388 | 0.368 | 33,146 | 0.376 |
| 1997 | 565,460 | 0.114 | 498,090 | 0.107 | 215,875 | 0.190 | 190,338 | 0.188 | 132,189 | 0.202 | 127,572 | 0.187 |
| 1998 | 575,365 | 0.117 | 503,929 | 0.110 | 225,248 | 0.182 | 197,447 | 0.178 | 102,451 | 0.284 | 92,073 | 0.274 |
| 1999 | 599,003 | 0.119 | 522,414 | 0.112 | 233,016 | 0.170 | 202,870 | 0.164 | 223,462 | 0.178 | 201,659 | 0.175 |
| 2000 | 609,566 | 0.121 | 529,217 | 0.115 | 237,412 | 0.158 | 205,025 | 0.152 | 51,858 | 0.438 | 48,880 | 0.414 |
| 2001 | 635,219 | 0.123 | 549,735 | 0.118 | 241,135 | 0.151 | 206,720 | 0.145 | 208,919 | 0.186 | 188,650 | 0.185 |
| 2002 | 653,000 | 0.125 | 563,361 | 0.120 | 244,435 | 0.152 | 208,269 | 0.147 | 67,519 | 0.431 | 59,844 | 0.416 |
| 2003 | 695,586 | 0.127 | 597,042 | 0.124 | 248,170 | 0.159 | 210,198 | 0.157 | 387,507 | 0.167 | 331,236 | 0.169 |
| 2004 | 716,500 | 0.129 | 611,396 | 0.127 | 253,395 | 0.170 | 213,591 | 0.170 | 53,321 | 0.526 | 38,212 | 0.523 |
| 2005 | 744,056 | 0.131 | 629,571 | 0.130 | 262,357 | 0.180 | 220,422 | 0.182 | 127,975 | 0.270 | 81,642 | 0.284 |
| 2006 | 765,602 | 0.133 | 643,005 | 0.133 | 273,625 | 0.186 | 229,350 | 0.189 | 46,697 | 0.521 | 29,979 | 0.498 |
| 2007 | 791,935 | 0.135 | 654,833 | 0.136 | 285,167 | 0.191 | 238,001 | 0.195 | 178,087 | 0.261 | 94,791 | 0.290 |
| 2008 | 811,004 | 0.139 | 657,981 | 0.140 | 296,483 | 0.197 | 245,839 | 0.201 | 150,752 | 0.313 | 78,861 | 0.348 |
| 2009 | 823,258 | 0.142 | 656,210 | 0.144 | 309,251 | 0.201 | 254,326 | 0.204 | 66,111 | 0.465 | 36,897 | 0.485 |
| 2010 | 834,091 | 0.144 | 653,665 | 0.148 | 321,869 | 0.198 | 261,815 | 0.201 | 67,396 | 0.449 | 45,917 | 0.459 |
| 2011 | 845,050 | 0.147 | 646,482 | 0.153 | 330,686 | 0.193 | 264,704 | 0.194 | 140,127 | 0.330 | 59,397 | 0.455 |
| 2012 | 841,177 | 0.151 | 633,007 | 0.158 | 334,878 | 0.189 | 262,190 | 0.189 | 47,677 | 0.552 |  |  |
| 2013 | 832,801 | 0.156 | 618,851 | 0.163 | 336,258 | 0.189 | 256,200 | 0.190 | 46,194 | 0.566 |  |  |
| 2014 | 817,728 | 0.161 | 597,506 | 0.170 | 333,615 | 0.193 | 246,104 | 0.196 |  |  |  |  |
| 2015 | 800,496 | 0.166 | 577,967 | 0.177 | 329,214 | 0.196 |  |  |  |  |  |  |
| 2016 | 783,492 | 0.171 |  |  | 323,393 | 0.199 |  |  |  |  |  |  |
| 2017 | 767,767 |  |  |  | 316,117 |  |  |  |  |  |  |  |
| Mean recru of post-197 |  |  |  |  |  |  |  |  | 109,512 |  | 94,787 |  |

Table 14. Estimated numbers at age for POP (millions).

|  | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| 1960 | 219.2 | 97.3 | 91.8 | 86.6 | 81.6 | 77.0 | 72.6 | 68.5 | 64.6 | 60.9 | 57.5 | 54.2 | 51.1 | 48.2 | 45.5 | 42.9 | 40.5 | 38.2 |
| 1961 | 54.2 | 206.7 | 91.7 | 86.5 | 81.5 | 76.8 | 72.3 | 68.1 | 64.0 | 60.1 | 56.5 | 53.2 | 50.2 | 47.4 | 44.9 | 42.4 | 40.1 | 37.9 |
| 1962 | 50.2 | 51.0 | 194.4 | 86.1 | 80.7 | 75.4 | 70.1 | 64.7 | 59.2 | 54.0 | 49.5 | 45.9 | 43.3 | 41.4 | 40.0 | 38.7 | 37.4 | 35.9 |
| 1963 | 63.0 | 47.4 | 48.0 | 182.9 | 80.8 | 75.5 | 70.1 | 64.5 | 58.8 | 53.1 | 47.9 | 43.6 | 40.5 | 38.4 | 37.1 | 36.2 | 35.4 | 34.4 |
| 1964 | 175.1 | 59.4 | 44.5 | 45.0 | 170.6 | 74.6 | 68.7 | 62.3 | 55.7 | 49.1 | 43.2 | 38.4 | 35.0 | 33.0 | 32.0 | 31.7 | 31.7 | 31.5 |
| 1965 | 361.5 | 164.3 | 55.5 | 41.2 | 41.0 | 150.9 | 63.0 | 54.2 | 45.0 | 36.4 | 29.5 | 24.8 | 22.1 | 21.1 | 21.4 | 22.4 | 23.8 | 25.1 |
| 1966 | 48.7 | 338.1 | 152.5 | 50.7 | 36.8 | 35.0 | 120.4 | 45.6 | 34.5 | 25.0 | 18.1 | 13.8 | 11.7 | 11.1 | 11.7 | 13.2 | 15.2 | 17.5 |
| 1967 | 38.7 | 45.4 | 312.1 | 138.3 | 44.6 | 30.7 | 27.0 | 83.3 | 27.5 | 18.0 | 11.5 | 7.8 | 6.1 | 5.5 | 5.8 | 6.9 | 8.6 | 10.8 |
| 1968 | 41.3 | 36.0 | 41.9 | 282.9 | 121.7 | 37.4 | 24.0 | 19.2 | 52.6 | 15.4 | 9.1 | 5.6 | 3.8 | 3.1 | 3.1 | 3.6 | 4.7 | 6.3 |
| 1969 | 40.1 | 38.3 | 33.0 | 37.4 | 243.8 | 99.0 | 28.0 | 16.1 | 11.3 | 26.9 | 7.0 | 4.0 | 2.5 | 1.8 | 1.6 | 1.8 | 2.3 | 3.3 |
| 1970 | 28.6 | 37.2 | 35.2 | 29.7 | 32.8 | 204.2 | 77.9 | 20.3 | 10.6 | 6.7 | 14.9 | 3.8 | 2.2 | 1.4 | 1.1 | 1.1 | 1.3 | 1.8 |
| 1971 | 21.9 | 26.2 | 33.5 | 30.7 | 24.8 | 25.5 | 144.6 | 48.9 | 11.1 | 5.0 | 2.9 | 6.1 | 1.6 | 1.0 | 0.7 | 0.6 | 0.7 | 0.9 |
| 1972 | 28.9 | 20.4 | 24.2 | 30.5 | 27.4 | 21.5 | 21.2 | 114.5 | 36.6 | 7.9 | 3.4 | 1.9 | 4.1 | 1.1 | 0.7 | 0.5 | 0.5 | 0.5 |
| 1973 | 46.9 | 26.8 | 18.6 | 21.7 | 26.7 | 23.1 | 17.3 | 16.1 | 81.6 | 24.5 | 5.0 | 2.2 | 1.2 | 2.7 | 0.8 | 0.5 | 0.4 | 0.4 |
| 1974 | 37.9 | 43.8 | 24.9 | 17.2 | 19.8 | 24.0 | 20.4 | 14.9 | 13.6 | 67.2 | 19.9 | 4.1 | 1.7 | 1.0 | 2.3 | 0.6 | 0.4 | 0.4 |
| 1975 | 44.7 | 34.8 | 39.7 | 22.2 | 14.9 | 16.6 | 19.2 | 15.5 | 10.7 | 9.2 | 43.8 | 12.8 | 2.6 | 1.2 | 0.7 | 1.7 | 0.5 | 0.3 |
| 1976 | 22.1 | 41.2 | 31.8 | 35.7 | 19.5 | 12.7 | 13.7 | 15.2 | 11.7 | 7.7 | 6.5 | 30.6 | 9.0 | 1.9 | 0.9 | 0.5 | 1.3 | 0.4 |
| 1977 | 26.0 | 20.2 | 37.0 | 28.0 | 30.5 | 16.0 | 10.0 | 10.2 | 10.7 | 7.8 | 5.0 | 4.2 | 19.9 | 6.0 | 1.3 | 0.6 | 0.4 | 1.0 |
| 1978 | 24.7 | 24.2 | 18.6 | 33.9 | 25.3 | 27.2 | 14.1 | 8.6 | 8.6 | 8.9 | 6.4 | 4.1 | 3.4 | 16.4 | 5.1 | 1.1 | 0.6 | 0.4 |
| 1979 | 63.2 | 23.0 | 22.5 | 17.2 | 31.0 | 23.0 | 24.4 | 12.4 | 7.5 | 7.4 | 7.5 | 5.4 | 3.5 | 2.9 | 14.2 | 4.4 | 1.0 | 0.5 |
| 1980 | 55.8 | 59.0 | 21.4 | 20.7 | 15.8 | 28.2 | 20.6 | 21.6 | 10.8 | 6.5 | 6.3 | 6.4 | 4.7 | 3.0 | 2.5 | 12.5 | 3.9 | 0.9 |
| 1981 | 40.6 | 52.3 | 55.0 | 19.9 | 19.2 | 14.5 | 25.7 | 18.6 | 19.4 | 9.7 | 5.7 | 5.6 | 5.7 | 4.2 | 2.7 | 2.3 | 11.4 | 3.6 |
| 1982 | 68.0 | 38.0 | 48.8 | 51.3 | 18.4 | 17.7 | 13.3 | 23.4 | 16.8 | 17.4 | 8.6 | 5.1 | 5.0 | 5.1 | 3.7 | 2.4 | 2.1 | 10.4 |
| 1983 | 95.4 | 64.0 | 35.8 | 45.9 | 48.2 | 17.3 | 16.6 | 12.4 | 21.8 | 15.7 | 16.2 | 8.0 | 4.8 | 4.6 | 4.8 | 3.5 | 2.3 | 2.0 |
| 1984 | 138.4 | 89.9 | 60.3 | 33.7 | 43.2 | 45.3 | 16.3 | 15.6 | 11.6 | 20.5 | 14.7 | 15.1 | 7.5 | 4.5 | 4.4 | 4.5 | 3.3 | 2.1 |
| 1985 | 78.3 | 130.5 | 84.7 | 56.8 | 31.7 | 40.6 | 42.6 | 15.3 | 14.6 | 10.9 | 19.1 | 13.7 | 14.2 | 7.0 | 4.2 | 4.1 | 4.2 | 3.1 |
| 1986 | 58.3 | 73.8 | 123.0 | 79.8 | 53.5 | 29.9 | 38.3 | 40.1 | 14.4 | 13.7 | 10.3 | 18.0 | 12.9 | 13.3 | 6.6 | 3.9 | 3.8 | 4.0 |
| 1987 | 315.2 | 54.9 | 69.6 | 115.9 | 75.1 | 50.4 | 28.1 | 36.0 | 37.6 | 13.5 | 12.9 | 9.6 | 16.9 | 12.1 | 12.5 | 6.2 | 3.7 | 3.6 |
| 1988 | 56.1 | 297.0 | 51.7 | 65.4 | 108.9 | 70.5 | 47.2 | 26.3 | 33.6 | 35.0 | 12.5 | 12.0 | 8.9 | 15.7 | 11.3 | 11.6 | 5.8 | 3.4 |
| 1989 | 142.6 | 52.9 | 279.5 | 48.6 | 61.5 | 102.1 | 66.0 | 44.0 | 24.4 | 31.1 | 32.4 | 11.6 | 11.0 | 8.3 | 14.5 | 10.5 | 10.8 | 5.4 |
| 1990 | 64.4 | 134.2 | 49.7 | 262.5 | 45.6 | 57.5 | 95.1 | 61.2 | 40.7 | 22.5 | 28.5 | 29.6 | 10.6 | 10.1 | 7.6 | 13.4 | 9.7 | 10.0 |
| 1991 | 208.3 | 60.5 | 125.6 | 46.4 | 243.4 | 42.0 | 52.3 | 85.6 | 54.3 | 35.6 | 19.5 | 24.5 | 25.4 | 9.1 | 8.8 | 6.6 | 11.8 | 8.6 |
| 1992 | 112.8 | 196.1 | 56.9 | 118.0 | 43.5 | 227.5 | 39.1 | 48.5 | 78.9 | 49.8 | 32.5 | 17.7 | 22.3 | 23.1 | 8.3 | 8.0 | 6.1 | 10.9 |
| 1993 | 66.2 | 106.2 | 184.3 | 53.3 | 110.3 | 40.5 | 210.5 | 35.9 | 44.1 | 71.1 | 44.6 | 28.9 | 15.7 | 19.8 | 20.7 | 7.5 | 7.2 | 5.5 |
| 1994 | 31.9 | 62.3 | 99.7 | 172.8 | 49.8 | 102.5 | 37.4 | 192.7 | 32.5 | 39.6 | 63.2 | 39.3 | 25.4 | 13.9 | 17.5 | 18.4 | 6.7 | 6.5 |
| 1995 | 50.4 | 30.1 | 58.6 | 93.7 | 162.0 | 46.6 | 95.4 | 34.6 | 177.0 | 29.6 | 35.8 | 56.9 | 35.4 | 22.9 | 12.5 | 15.9 | 16.7 | 6.1 |
| 1996 | 42.4 | 47.5 | 28.3 | 55.1 | 88.0 | 151.7 | 43.4 | 88.6 | 31.9 | 162.3 | 27.0 | 32.5 | 51.6 | 32.1 | 20.8 | 11.4 | 14.6 | 15.4 |
| 1997 | 132.2 | 39.9 | 44.7 | 26.6 | 51.7 | 82.3 | 141.2 | 40.2 | 81.4 | 29.1 | 146.8 | 24.3 | 29.2 | 46.3 | 28.8 | 18.8 | 10.3 | 13.3 |
| 1998 | 102.5 | 124.6 | 37.6 | 42.1 | 25.0 | 48.4 | 76.8 | 131.2 | 37.1 | 74.7 | 26.5 | 133.2 | 22.0 | 26.4 | 42.0 | 26.3 | 17.2 | 9.5 |
| 1999 | 223.5 | 96.6 | 117.4 | 35.4 | 39.6 | 23.5 | 45.4 | 71.7 | 122.0 | 34.3 | 68.8 | 24.3 | 122.1 | 20.2 | 24.2 | 38.6 | 24.2 | 15.9 |
| 2000 | 51.9 | 210.6 | 91.0 | 110.5 | 33.3 | 37.1 | 22.0 | 42.2 | 66.4 | 112.4 | 31.5 | 62.7 | 22.1 | 111.0 | 18.4 | 22.1 | 35.4 | 22.2 |
| 2001 | 208.9 | 48.9 | 198.5 | 85.7 | 104.0 | 31.3 | 34.8 | 20.5 | 39.3 | 61.6 | 103.7 | 29.0 | 57.6 | 20.3 | 102.0 | 16.9 | 20.4 | 32.7 |
| 2002 | 67.5 | 197.0 | 46.1 | 187.0 | 80.6 | 97.7 | 29.3 | 32.5 | 19.1 | 36.5 | 57.0 | 95.6 | 26.6 | 53.0 | 18.7 | 94.0 | 15.6 | 18.9 |
| 2003 | 387.5 | 63.6 | 185.6 | 43.4 | 175.9 | 75.7 | 91.6 | 27.4 | 30.3 | 17.7 | 33.7 | 52.3 | 87.6 | 24.4 | 48.5 | 17.1 | 86.4 | 14.4 |
| 2004 | 53.3 | 365.2 | 60.0 | 174.7 | 40.8 | 165.0 | 70.9 | 85.3 | 25.4 | 27.9 | 16.2 | 30.7 | 47.5 | 79.4 | 22.1 | 44.1 | 15.6 | 79.0 |
| 2005 | 128.0 | 50.3 | 344.1 | 56.5 | 164.3 | 38.3 | 154.6 | 66.2 | 79.3 | 23.5 | 25.7 | 14.9 | 28.1 | 43.4 | 72.6 | 20.3 | 40.5 | 14.4 |
| 2006 | 46.7 | 120.6 | 47.4 | 324.1 | 53.1 | 154.4 | 35.9 | 144.6 | 61.7 | 73.7 | 21.7 | 23.7 | 13.7 | 25.8 | 39.9 | 66.8 | 18.7 | 37.3 |
| 2007 | 178.1 | 44.0 | 113.7 | 44.6 | 304.9 | 49.9 | 144.7 | 33.6 | 134.5 | 57.1 | 67.9 | 20.0 | 21.7 | 12.5 | 23.6 | 36.5 | 61.2 | 17.1 |
| 2008 | 150.8 | 167.8 | 41.5 | 106.9 | 41.9 | 285.8 | 46.6 | 134.7 | 31.1 | 123.8 | 52.2 | 61.8 | 18.1 | 19.6 | 11.3 | 21.3 | 33.1 | 55.6 |
| 2009 | 66.1 | 142.1 | 158.1 | 39.0 | 100.5 | 39.3 | 267.4 | 43.5 | 125.0 | 28.7 | 113.7 | 47.7 | 56.2 | 16.4 | 17.8 | 10.3 | 19.4 | 30.2 |
| 2010 | 67.4 | 62.3 | 133.8 | 148.8 | 36.7 | 94.4 | 36.8 | 249.7 | 40.4 | 115.8 | 26.5 | 104.5 | 43.7 | 51.4 | 15.0 | 16.3 | 9.4 | 17.8 |
| 2011 | 140.1 | 63.5 | 58.7 | 125.9 | 139.8 | 34.4 | 88.3 | 34.4 | 232.1 | 37.4 | 106.6 | 24.3 | 95.5 | 39.9 | 46.9 | 13.7 | 14.8 | 8.6 |
| 2012 | 47.7 | 132.0 | 59.8 | 55.2 | 118.2 | 131.0 | 32.1 | 82.2 | 31.8 | 213.7 | 34.3 | 97.1 | 22.0 | 86.3 | 36.0 | 42.3 | 12.4 | 13.4 |
| 2013 | 46.2 | 44.9 | 124.2 | 56.2 | 51.8 | 110.7 | 122.4 | 29.9 | 76.1 | 29.3 | 195.9 | 31.3 | 88.2 | 19.9 | 78.1 | 32.5 | 38.2 | 11.2 |
| 2014 | 94.9 | 43.5 | 42.2 | 116.7 | 52.7 | 48.4 | 103.2 | 113.5 | 27.6 | 69.8 | 26.7 | 177.3 | 28.1 | 79.1 | 17.8 | 69.7 | 29.1 | 34.1 |
| 2015 | 94.9 | 89.3 | 40.9 | 39.6 | 109.3 | 49.2 | 45.1 | 95.6 | 104.6 | 25.3 | 63.5 | 24.2 | 159.7 | 25.2 | 70.7 | 15.9 | 62.1 | 25.9 |
| 2016 | 94.9 | 89.3 | 84.0 | 38.4 | 37.1 | 102.1 | 45.8 | 41.8 | 88.3 | 96.0 | 23.1 | 57.7 | 21.8 | 143.6 | 22.6 | 63.2 | 14.2 | 55.4 |

Table 14 (continued). Estimated numbers at age for POP (millions).

|  | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 34 | 35 | 36 | 37 | 38 | 39 | 40+ |
| 1960 | 36.0 | 34.0 | 32.0 | 30.2 | 28.5 | 26.9 | 25.3 | 23.9 | 22.5 | 21.3 | 20.1 | 18.9 | 17.8 | 16.8 | 15.9 | 15.0 | 14.1 | 13.3 | 12.6 | 208.5 |
| 1961 | 35.8 | 33.8 | 31.9 | 30.1 | 28.4 | 26.8 | 25.3 | 23.9 | 22.5 | 21.2 | 20.0 | 18.9 | 17.8 | 16.8 | 15.9 | 15.0 | 14.1 | 13.3 | 12.6 | 208.4 |
| 1962 | 34.4 | 32.8 | 31.2 | 29.5 | 27.9 | 26.4 | 25.0 | 23.6 | 22.3 | 21.0 | 19.8 | 18.7 | 17.7 | 16.7 | 15.7 | 14.9 | 14.0 | 13.2 | 12.5 | 207.4 |
| 1963 | 33.3 | 32.0 | 30.6 | 29.1 | 27.6 | 26.2 | 24.8 | 23.4 | 22.1 | 20.9 | 19.7 | 18.6 | 17.6 | 16.6 | 15.7 | 14.8 | 14.0 | 13.2 | 12.4 | 207.0 |
| 1964 | 31.1 | 30.4 | 29.4 | 28.2 | 27.0 | 25.7 | 24.3 | 23.1 | 21.8 | 20.6 | 19.5 | 18.4 | 17.4 | 16.4 | 15.5 | 14.7 | 13.9 | 13.1 | 12.4 | 205.9 |
| 1965 | 26.1 | 26.5 | 26.5 | 26.0 | 25.2 | 24.2 | 23.2 | 22.1 | 20.9 | 19.9 | 18.8 | 17.8 | 16.9 | 16.0 | 15.1 | 14.3 | 13.5 | 12.8 | 12.1 | 202.7 |
| 1966 | 19.6 | 21.2 | 22.3 | 22.7 | 22.5 | 22.1 | 21.4 | 20.5 | 19.6 | 18.7 | 17.8 | 16.9 | 16.1 | 15.3 | 14.5 | 13.7 | 13.0 | 12.4 | 11.7 | 197.8 |
| 1967 | 13.3 | 15.6 | 17.5 | 18.7 | 19.4 | 19.5 | 19.2 | 18.7 | 18.1 | 17.4 | 16.6 | 15.8 | 15.1 | 14.4 | 13.7 | 13.1 | 12.4 | 11.8 | 11.3 | 191.9 |
| 1968 | 8.4 | 10.7 | 13.0 | 14.8 | 16.1 | 16.8 | 17.0 | 16.9 | 16.5 | 16.0 | 15.4 | 14.8 | 14.2 | 13.6 | 13.0 | 12.4 | 11.8 | 11.3 | 10.8 | 186.3 |
| 1969 | 4.7 | 6.6 | 8.7 | 10.7 | 12.5 | 13.7 | 14.5 | 14.7 | 14.7 | 14.4 | 14.0 | 13.6 | 13.1 | 12.6 | 12.1 | 11.6 | 11.1 | 10.7 | 10.2 | 179.3 |
| 1970 | 2.6 | 3.8 | 5.5 | 7.4 | 9.3 | 10.9 | 12.0 | 12.7 | 13.0 | 13.0 | 12.8 | 12.5 | 12.1 | 11.7 | 11.3 | 10.9 | 10.5 | 10.1 | 9.7 | 173.6 |
| 1971 | 1.3 | 2.0 | 3.0 | 4.4 | 6.1 | 7.7 | 9.1 | 10.2 | 10.8 | 11.1 | 11.2 | 11.1 | 10.9 | 10.6 | 10.3 | 10.0 | 9.6 | 9.3 | 9.0 | 164.9 |
| 1972 | 0.7 | 1.1 | 1.7 | 2.7 | 3.9 | 5.4 | 6.9 | 8.2 | 9.2 | 9.8 | 10.1 | 10.1 | 10.1 | 9.9 | 9.6 | 9.4 | 9.1 | 8.8 | 8.6 | 160.5 |
| 1973 | 0.4 | 0.6 | 0.9 | 1.5 | 2.3 | 3.5 | 4.8 | 6.1 | 7.3 | 8.2 | 8.7 | 9.0 | 9.1 | 9.1 | 8.9 | 8.7 | 8.5 | 8.3 | 8.1 | 155.1 |
| 1974 | 0.3 | 0.4 | 0.6 | 0.8 | 1.4 | 2.1 | 3.2 | 4.4 | 5.6 | 6.7 | 7.5 | 8.1 | 8.3 | 8.4 | 8.4 | 8.3 | 8.1 | 7.9 | 7.7 | 152.1 |
| 1975 | 0.3 | 0.3 | 0.3 | 0.5 | 0.7 | 1.2 | 1.9 | 2.8 | 3.9 | 5.0 | 6.0 | 6.7 | 7.2 | 7.5 | 7.6 | 7.6 | 7.5 | 7.3 | 7.2 | 145.9 |
| 1976 | 0.3 | 0.2 | 0.2 | 0.3 | 0.4 | 0.7 | 1.1 | 1.7 | 2.5 | 3.5 | 4.5 | 5.4 | 6.0 | 6.5 | 6.7 | 6.9 | 6.9 | 6.8 | 6.7 | 140.5 |
| 1977 | 0.3 | 0.2 | 0.2 | 0.2 | 0.3 | 0.4 | 0.6 | 0.9 | 1.5 | 2.2 | 3.1 | 3.9 | 4.7 | 5.4 | 5.8 | 6.0 | 6.1 | 6.2 | 6.1 | 133.6 |
| 1978 | 0.9 | 0.3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.3 | 0.5 | 0.8 | 1.3 | 2.0 | 2.8 | 3.6 | 4.4 | 4.9 | 5.3 | 5.6 | 5.7 | 5.7 | 129.7 |
| 1979 | 0.3 | 0.8 | 0.3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.3 | 0.5 | 0.8 | 1.2 | 1.9 | 2.6 | 3.4 | 4.0 | 4.6 | 4.9 | 5.2 | 5.3 | 126.2 |
| 1980 | 0.4 | 0.3 | 0.8 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.3 | 0.4 | 0.7 | 1.1 | 1.7 | 2.4 | 3.1 | 3.7 | 4.2 | 4.6 | 4.8 | 122.4 |
| 1981 | 0.8 | 0.4 | 0.3 | 0.7 | 0.2 | 0.2 | 0.1 | 0.2 | 0.2 | 0.3 | 0.4 | 0.7 | 1.1 | 1.6 | 2.2 | 2.9 | 3.5 | 4.0 | 4.3 | 119.0 |
| 1982 | 3.3 | 0.8 | 0.4 | 0.3 | 0.7 | 0.2 | 0.2 | 0.1 | 0.1 | 0.2 | 0.2 | 0.4 | 0.6 | 1.0 | 1.5 | 2.1 | 2.7 | 3.3 | 3.7 | 115.3 |
| 1983 | 9.7 | 3.1 | 0.7 | 0.4 | 0.2 | 0.6 | 0.2 | 0.1 | 0.1 | 0.1 | 0.2 | 0.2 | 0.4 | 0.6 | 0.9 | 1.4 | 2.0 | 2.5 | 3.1 | 112.0 |
| 1984 | 1.8 | 9.2 | 2.9 | 0.7 | 0.3 | 0.2 | 0.6 | 0.2 | 0.1 | 0.1 | 0.1 | 0.2 | 0.2 | 0.3 | 0.5 | 0.9 | 1.3 | 1.8 | 2.4 | 108.4 |
| 1985 | 2.0 | 1.7 | 8.6 | 2.7 | 0.6 | 0.3 | 0.2 | 0.5 | 0.2 | 0.1 | 0.1 | 0.1 | 0.1 | 0.2 | 0.3 | 0.5 | 0.8 | 1.2 | 1.7 | 104.3 |
| 1986 | 2.9 | 1.9 | 1.6 | 8.1 | 2.6 | 0.6 | 0.3 | 0.2 | 0.5 | 0.2 | 0.1 | 0.1 | 0.1 | 0.1 | 0.2 | 0.3 | 0.5 | 0.8 | 1.2 | 100.0 |
| 1987 | 3.7 | 2.7 | 1.8 | 1.5 | 7.6 | 2.4 | 0.6 | 0.3 | 0.2 | 0.5 | 0.2 | 0.1 | 0.1 | 0.1 | 0.1 | 0.2 | 0.3 | 0.5 | 0.7 | 95.3 |
| 1988 | 3.4 | 3.5 | 2.6 | 1.7 | 1.4 | 7.2 | 2.3 | 0.5 | 0.3 | 0.2 | 0.5 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.2 | 0.3 | 0.4 | 90.3 |
| 1989 | 3.2 | 3.2 | 3.3 | 2.4 | 1.6 | 1.3 | 6.7 | 2.1 | 0.5 | 0.2 | 0.2 | 0.4 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.2 | 0.2 | 85.1 |
| 1990 | 5.0 | 3.0 | 2.9 | 3.0 | 2.2 | 1.5 | 1.3 | 6.3 | 2.0 | 0.5 | 0.2 | 0.2 | 0.4 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 80.0 |
| 1991 | 9.0 | 4.5 | 2.7 | 2.7 | 2.8 | 2.0 | 1.3 | 1.1 | 5.7 | 1.8 | 0.4 | 0.2 | 0.1 | 0.4 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 73.9 |
| 1992 | 7.9 | 8.3 | 4.2 | 2.5 | 2.5 | 2.6 | 1.9 | 1.2 | 1.1 | 5.3 | 1.7 | 0.4 | 0.2 | 0.1 | 0.3 | 0.1 | 0.1 | 0.1 | 0.1 | 69.2 |
| 1993 | 9.9 | 7.3 | 7.6 | 3.8 | 2.3 | 2.3 | 2.4 | 1.7 | 1.1 | 1.0 | 4.9 | 1.6 | 0.4 | 0.2 | 0.1 | 0.3 | 0.1 | 0.1 | 0.1 | 64.3 |
| 1994 | 5.0 | 9.0 | 6.6 | 7.0 | 3.5 | 2.1 | 2.1 | 2.2 | 1.6 | 1.0 | 0.9 | 4.5 | 1.4 | 0.3 | 0.2 | 0.1 | 0.3 | 0.1 | 0.1 | 59.6 |
| 1995 | 6.0 | 4.6 | 8.3 | 6.1 | 6.5 | 3.3 | 2.0 | 1.9 | 2.0 | 1.5 | 1.0 | 0.8 | 4.2 | 1.3 | 0.3 | 0.2 | 0.1 | 0.3 | 0.1 | 55.6 |
| 1996 | 5.6 | 5.5 | 4.3 | 7.7 | 5.7 | 6.0 | 3.0 | 1.8 | 1.8 | 1.9 | 1.4 | 0.9 | 0.8 | 3.9 | 1.2 | 0.3 | 0.1 | 0.1 | 0.2 | 51.9 |
| 1997 | 14.1 | 5.2 | 5.1 | 3.9 | 7.1 | 5.2 | 5.5 | 2.8 | 1.7 | 1.7 | 1.7 | 1.3 | 0.8 | 0.7 | 3.6 | 1.1 | 0.3 | 0.1 | 0.1 | 48.5 |
| 1998 | 12.2 | 13.0 | 4.8 | 4.7 | 3.6 | 6.6 | 4.9 | 5.1 | 2.6 | 1.6 | 1.5 | 1.6 | 1.2 | 0.8 | 0.7 | 3.3 | 1.1 | 0.2 | 0.1 | 45.3 |
| 1999 | 8.8 | 11.3 | 12.1 | 4.4 | 4.4 | 3.4 | 6.1 | 4.5 | 4.8 | 2.4 | 1.4 | 1.4 | 1.5 | 1.1 | 0.7 | 0.6 | 3.1 | 1.0 | 0.2 | 42.4 |
| 2000 | 14.6 | 8.1 | 10.5 | 11.2 | 4.1 | 4.1 | 3.1 | 5.7 | 4.2 | 4.4 | 2.2 | 1.3 | 1.3 | 1.4 | 1.0 | 0.7 | 0.6 | 2.9 | 0.9 | 39.7 |
| 2001 | 20.6 | 13.6 | 7.5 | 9.7 | 10.4 | 3.8 | 3.8 | 2.9 | 5.3 | 3.9 | 4.1 | 2.1 | 1.2 | 1.2 | 1.3 | 0.9 | 0.6 | 0.5 | 2.7 | 38.0 |
| 2002 | 30.3 | 19.1 | 12.6 | 7.0 | 9.1 | 9.7 | 3.6 | 3.5 | 2.7 | 4.9 | 3.6 | 3.8 | 1.9 | 1.2 | 1.1 | 1.2 | 0.9 | 0.6 | 0.5 | 38.0 |
| 2003 | 17.5 | 28.1 | 17.7 | 11.7 | 6.5 | 8.4 | 9.0 | 3.3 | 3.3 | 2.5 | 4.6 | 3.4 | 3.6 | 1.8 | 1.1 | 1.1 | 1.1 | 0.8 | 0.5 | 35.9 |
| 2004 | 13.2 | 16.0 | 25.8 | 16.3 | 10.8 | 6.0 | 7.8 | 8.3 | 3.1 | 3.0 | 2.3 | 4.2 | 3.1 | 3.3 | 1.7 | 1.0 | 1.0 | 1.0 | 0.8 | 33.8 |
| 2005 | 72.8 | 12.2 | 14.8 | 23.9 | 15.1 | 10.0 | 5.6 | 7.2 | 7.7 | 2.8 | 2.8 | 2.2 | 3.9 | 2.9 | 3.0 | 1.5 | 0.9 | 0.9 | 1.0 | 32.2 |
| 2006 | 13.3 | 67.4 | 11.3 | 13.7 | 22.2 | 14.0 | 9.3 | 5.2 | 6.7 | 7.1 | 2.6 | 2.6 | 2.0 | 3.6 | 2.7 | 2.8 | 1.4 | 0.9 | 0.9 | 30.9 |
| 2007 | 34.4 | 12.2 | 62.2 | 10.4 | 12.7 | 20.5 | 13.0 | 8.6 | 4.8 | 6.2 | 6.6 | 2.4 | 2.4 | 1.8 | 3.3 | 2.5 | 2.6 | 1.3 | 0.8 | 29.5 |
| 2008 | 15.6 | 31.4 | 11.2 | 56.9 | 9.5 | 11.6 | 18.8 | 11.9 | 7.8 | 4.4 | 5.6 | 6.0 | 2.2 | 2.2 | 1.7 | 3.1 | 2.3 | 2.4 | 1.2 | 28.0 |
| 2009 | 50.8 | 14.3 | 28.7 | 10.3 | 52.2 | 8.7 | 10.7 | 17.2 | 10.9 | 7.2 | 4.0 | 5.2 | 5.5 | 2.0 | 2.0 | 1.6 | 2.8 | 2.1 | 2.2 | 27.0 |
| 2010 | 27.7 | 46.7 | 13.1 | 26.4 | 9.4 | 48.0 | 8.0 | 9.8 | 15.8 | 10.0 | 6.6 | 3.7 | 4.8 | 5.1 | 1.9 | 1.9 | 1.4 | 2.6 | 1.9 | 27.1 |
| 2011 | 16.2 | 25.3 | 42.7 | 12.0 | 24.2 | 8.7 | 44.0 | 7.4 | 9.0 | 14.5 | 9.2 | 6.1 | 3.4 | 4.4 | 4.7 | 1.7 | 1.7 | 1.3 | 2.4 | 26.9 |
| 2012 | 7.8 | 14.7 | 23.0 | 38.8 | 10.9 | 22.0 | 7.9 | 40.0 | 6.7 | 8.2 | 13.2 | 8.3 | 5.5 | 3.1 | 4.0 | 4.3 | 1.6 | 1.6 | 1.2 | 26.9 |
| 2013 | 12.1 | 7.0 | 13.3 | 20.8 | 35.2 | 9.9 | 19.9 | 7.1 | 36.3 | 6.1 | 7.4 | 12.0 | 7.6 | 5.0 | 2.8 | 3.6 | 3.9 | 1.4 | 1.4 | 25.8 |
| 2014 | 10.0 | 10.9 | 6.3 | 11.9 | 18.6 | 31.5 | 8.9 | 17.9 | 6.4 | 32.5 | 5.4 | 6.6 | 10.7 | 6.8 | 4.5 | 2.5 | 3.3 | 3.5 | 1.3 | 24.8 |
| 2015 | 30.4 | 8.9 | 9.7 | 5.6 | 10.7 | 16.6 | 28.1 | 7.9 | 15.9 | 5.7 | 29.0 | 4.9 | 5.9 | 9.6 | 6.1 | 4.1 | 2.3 | 3.0 | 3.2 | 23.7 |
| 2016 | 23.1 | 27.1 | 7.9 | 8.6 | 5.0 | 9.5 | 14.8 | 25.0 | 7.0 | 14.2 | 5.1 | 25.9 | 4.3 | 5.3 | 8.6 | 5.5 | 3.7 | 2.0 | 2.7 | 24.4 |

Table 15. Projections of BSAI spawning biomass (t), catch ( t ), and fishing mortality rate for each of the several scenarios. The values of $B_{40 \%}$ and $B_{35 \%}$ are $214,685 t$ and $187,849 t$, respectively.

| Catch | Scenario 1 | Scenario 2 | Scenario 3 | Scenario 4 | Scenario 5 | Scenario 6 | Scenario 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | 31,411 | 31,411 | 31,411 | 31,411 | 31,411 | 31,411 | 31,411 |
| 2017 | 43,723 | 43,723 | 22,241 | 15,974 | 0 | 53,152 | 43,723 |
| 2018 | 41,865 | 41,865 | 22,036 | 15,982 | 0 | 50,120 | 41,865 |
| 2019 | 40,000 | 40,000 | 21,760 | 15,932 | 0 | 47,187 | 48,626 |
| 2020 | 38,194 | 38,194 | 21,442 | 15,843 | 0 | 44,431 | 45,727 |
| 2021 | 36,511 | 36,511 | 21,116 | 15,737 | 0 | 41,924 | 43,082 |
| 2022 | 35,008 | 35,008 | 20,811 | 15,636 | 0 | 39,723 | 40,752 |
| 2023 | 33,727 | 33,727 | 20,556 | 15,561 | 0 | 37,868 | 38,778 |
| 2024 | 32,678 | 32,678 | 20,362 | 15,519 | 0 | 36,082 | 37,086 |
| 2025 | 31,868 | 31,868 | 20,248 | 15,527 | 0 | 34,343 | 35,335 |
| 2026 | 31,177 | 31,177 | 20,195 | 15,570 | 0 | 33,033 | 33,889 |
| 2027 | 30,596 | 30,596 | 20,201 | 15,650 | 0 | 32,125 | 32,843 |
| 2028 | 30,143 | 30,143 | 20,243 | 15,751 | 0 | 31,510 | 32,101 |
| 2029 | 29,818 | 29,818 | 20,315 | 15,868 | 0 | 31,112 | 31,594 |

Sp. Scenario 1 Scenario 2 Scenario 3 Scenario 4 Scenario 5 Scenario 6 Scenario 7 Biomass

| 2016 | 323,395 | 323,395 | 323,395 | 323,395 | 323,395 | 323,395 | 323,395 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2017 | 314,489 | 314,489 | 317,188 | 317,963 | 319,912 | 313,282 | 314,489 |
| 2018 | 300,167 | 300,167 | 312,928 | 316,674 | 326,272 | 294,607 | 300,167 |
| 2019 | 286,070 | 286,070 | 307,915 | 314,475 | 331,580 | 276,791 | 284,966 |
| 2020 | 272,960 | 272,960 | 302,887 | 312,074 | 336,457 | 260,566 | 267,897 |
| 2021 | 261,563 | 261,563 | 298,603 | 310,223 | 341,601 | 246,601 | 253,140 |
| 2022 | 252,193 | 252,193 | 295,449 | 309,306 | 347,371 | 235,138 | 240,945 |
| 2023 | 244,888 | 244,888 | 293,575 | 309,492 | 353,949 | 226,135 | 231,273 |
| 2024 | 239,393 | 239,393 | 292,838 | 310,652 | 361,219 | 219,300 | 223,804 |
| 2025 | 235,325 | 235,325 | 292,953 | 312,520 | 368,935 | 214,314 | 218,162 |
| 2026 | 232,272 | 232,272 | 293,588 | 314,778 | 376,798 | 210,778 | 214,004 |
| 2027 | 229,962 | 229,962 | 294,525 | 317,228 | 384,642 | 208,264 | 210,948 |
| 2028 | 228,189 | 228,189 | 295,594 | 319,708 | 392,312 | 206,441 | 208,661 |
| 2029 | 226,841 | 226,841 | 296,744 | 322,183 | 399,795 | 205,122 | 206,951 |

F Scenario 1 Scenario 2 Scenario 3 Scenario 4 Scenario 5 Scenario 6 Scenario 7

| 2016 | 0.057 | 0.057 | 0.057 | 0.057 | 0.057 | 0.057 | 0.057 |
| :--- | :--- | :--- | :--- | :--- | ---: | ---: | :--- |
| 2017 | 0.082 | 0.082 | 0.041 | 0.029 | 0 | 0.101 | 0.082 |
| 2018 | 0.082 | 0.082 | 0.041 | 0.029 | 0 | 0.101 | 0.082 |
| 2019 | 0.082 | 0.082 | 0.041 | 0.029 | 0 | 0.101 | 0.101 |
| 2020 | 0.082 | 0.082 | 0.041 | 0.029 | 0 | 0.101 | 0.101 |
| 2021 | 0.082 | 0.082 | 0.041 | 0.029 | 0 | 0.101 | 0.101 |
| 2022 | 0.082 | 0.082 | 0.041 | 0.029 | 0 | 0.101 | 0.101 |
| 2023 | 0.082 | 0.082 | 0.041 | 0.029 | 0 | 0.101 | 0.101 |
| 2024 | 0.082 | 0.082 | 0.041 | 0.029 | 0 | 0.100 | 0.101 |
| 2025 | 0.082 | 0.082 | 0.041 | 0.029 | 0 | 0.098 | 0.099 |
| 2026 | 0.082 | 0.082 | 0.041 | 0.029 | 0 | 0.097 | 0.097 |
| 2027 | 0.082 | 0.082 | 0.041 | 0.029 | 0 | 0.095 | 0.096 |
| 2028 | 0.081 | 0.081 | 0.041 | 0.029 | 0 | 0.094 | 0.095 |
| 2029 | 0.081 | 0.081 | 0.041 | 0.029 | 0 | 0.094 | 0.094 |



Figure 1. Distribution of observed Aleutian Islands Pacific ocean perch catch (from North Pacific Groundfish Observer Program) by depth zone (top panel) and AI subarea (bottom panel) from 1977 to 2015.


Figure 2. Fishery age composition data for the BSAI POP; The diameter of the circles are scaled within each year of samples, and dashed lines denote cohorts.


Figure 3. AI survey POP CPUE (kg/km²) from 1992-2016; the symbol $\times$ denotes tows with no catch. The red lines indicate boundaries between the WAI, CAI, EAI, and EBS areas.


Figure 4. Age composition data from the Aleutian Islands trawl survey; bubbles are scaled within each year of samples; and dashed lines denote cohorts.


2012 EBS Survey POP CPUE (wgt/km²)


Figure 5. EBS slope survey POP CPUE (kg/km²) from 2010-2016; the symbol $\times$ denotes tows with no catch.


Figure 6. Age composition data from the eastern Bering Sea trawl survey; bubbles are scaled within each year of samples; and dashed lines denote cohorts.


Figure 7. Estimated time series of total biomass across the models.


Figure 8. Fit to Aleutian Islands survey biomass indices across the models.


Figure 9. Fit to eastern Bering Sea survey biomass indices across the models.


Figure 10. Data weights for the age and length composition data across the models.


Figure 11. Retrospective estimates of spawning stock biomass for model runs with end years of 2006 to 2016.


Figure 12. Posterior distributions for key model quantities $M$, survey catchability, median recruitment, and 2016 total biomass. For $M$ and survey catchability, the prior distributions are also shown in the solid lines. The MLE estimates are indicated by the vertical lines.


Figure 13. Observed AI survey biomass (data points, +/- 2 standard deviations), predicted survey biomass(solid line), and BSAI harvest (dashed line).


Figure 14. Smoothed proportion of BSAI biomass in the AI surveu area (lower line, from time series of survey biomass estimates) and product of the smoothed proportion and estimated AI survey catchability (top line).


Figure 15. Observed EBS survey biomass (data points, +/- 2 standard deviations) and predicted survey biomass (solid line.


Figure 16. Total and spawner biomass for BSAI Pacific ocean perch, with 95\% confidence intervals from MCMC integration.

Fishery age composition data


Figure 17. Model fits (dots) to fishery age composition data (columns) for Aleutian Islands Pacific ocean perch, 1981-2013. Colors correspond to cohorts (except for the 40+ group).

Fishery length composition data


Figure 18. Model fits (dots) to fishery length composition data (columns) for Aleutian Islands Pacific ocean perch, 1964-2015.

AI Survey age composition data


Figure 19. Model fits (dots) to survey age composition data (columns) for Aleutian Islands Pacific ocean perch, 1991-2014. Colors correspond to cohorts (except for the 40+ group).

## AI Survey length composition data



Length (cm)

Figure 20. Model fits (dots) to 2016 AI survey length composition data (columns) for Pacific ocean perch.

EBS Survey age composition data


Figure 21. Model fits (dots) to EBS slope survey age composition data (columns) for Pacific ocean perch, 2002-2012. Colors correspond to cohorts (except for the 40+ group).

## EBS Survey length composition data



Length (cm)

Figure 22. Model fits (dots) to 2016 EBS survey length composition data (columns) for Pacific ocean perch.


Figure 23. Estimated AI (black line) and EBS (red line) survey selectivity curve for BSAI POP.


Figure 24. Estimated fishery selectivity from 1960-2016.


Figure 25. Estimated fully selected fishing mortality for BSAI POP.


Figure 26. (Top panel) Estimated fishing mortality and SSB in reference to OFL (upper line) and ABC (lower line) harvest control rules, with 2016 shown in red. The bottom panel shows a reduced vertical scale, and the projected F and stock size for 2017 and 2018.


Figure 27. Estimated recruitment (age 3) of BSAI POP, with 95\% CI limits obtained from MCMC integration.


Figure 28. Scatterplot of BSAI POP spawner-recruit data; label is year class.


Figure 29. Estimated adjustment ratio to convert the EBS survey biomass into comparable units to the AI survey biomass, accounting for differences in the catchability and selectivity between the surveys, and changes in age composition over time.

## Appendix A. Supplemental Catch Data.

In order to comply with the Annual Catch Limit (ACL) requirements, non-commercial removals that do not occur during directed groundfish fishing activities are reported (Table A1). In these datasets, blackspotted /rougheye rockfish are often reported as rougheye rockfish. This includes removals incurred during research, subsistence, personal use, recreational, and exempted fishing permit activities, but does not include removals taken in fisheries other than those managed under the groundfish FMP. These estimates represent additional sources of removals to the existing Catch Accounting System estimates. For BSAI blackspotted/rougheye rockfish, these estimates can be compared to the trawl research removals reported in previous assessments. BSAI blackspotted/rougheye rockfish research removals are small relative to the fishery catch. The majority of removals are taken by the Alaska Fisheries Science Center's (AFSC) biennial bottom trawl survey which is the primary research survey used for assessing the population status of BSAI blackspotted/rougheye rockfish. The annual amount of blackspotted/rougheye rockfish captured in research longline gear not exceeded 0.5 t . Total removals ranged between 2010 and 2015 ranged between 0.016 t and 0.6 t , which were less than $1.0 \%$ of the ABC in these years.

## Appendix A. Supplemental Catch Data

In order to comply with the Annual Catch Limit (ACL) requirements, non-commercial removals that do not occur during directed groundfish fishing activities are reported (Table A1). This includes removals incurred during research, subsistence, personal use, recreational, and exempted fishing permit activities, but does not include removals taken in fisheries other than those managed under the groundfish FMP. These estimates represent additional sources of removals to the existing Catch Accounting System estimates. For BSAI POP, these estimates can be compared to the trawl research removals reported in previous assessments. POP research removals are small relative to the fishery catch. The majority of removals are taken by the Alaska Fisheries Science Center's (AFSC) biennial bottom trawl survey which is the primary research survey used for assessing the population status of BSAI POP. The amount of POP captured in research longline gear has typically been less than 0.1 t . There was no recorded recreational harvest or harvest that was non-research related in 2010 and 2011. Total removals of POP ranged between 3 and 286 t between 2010 and 2015, and did not exceed 1.4 of the ABC for these years.

Appendix Table A1. Removals of BSAI POP from activities other than groundfish fishing ( t ). Trawl and longline include research survey and occasional short-term projects.

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