Ambient Air Sampling for Particulate Matter

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Presentation Outline

- Background and statement of issues
- Characteristics of ambient aerosols (sources, size ranges, definitions of MMD, GSD, aerodynamic diameter)
- Size-selective measurement techniques (inertial impactors and cyclones, definitions of cutpoint and slope)
- Development of health-based PM standards physiological basis for PM sampling conventions
- EPA's method development for PM₁₀ and PM_{2.5}
- Review of TAMU's methodology for estimating "True" PM concentrations, and for estimating "oversampling" of EPA reference methods

Background

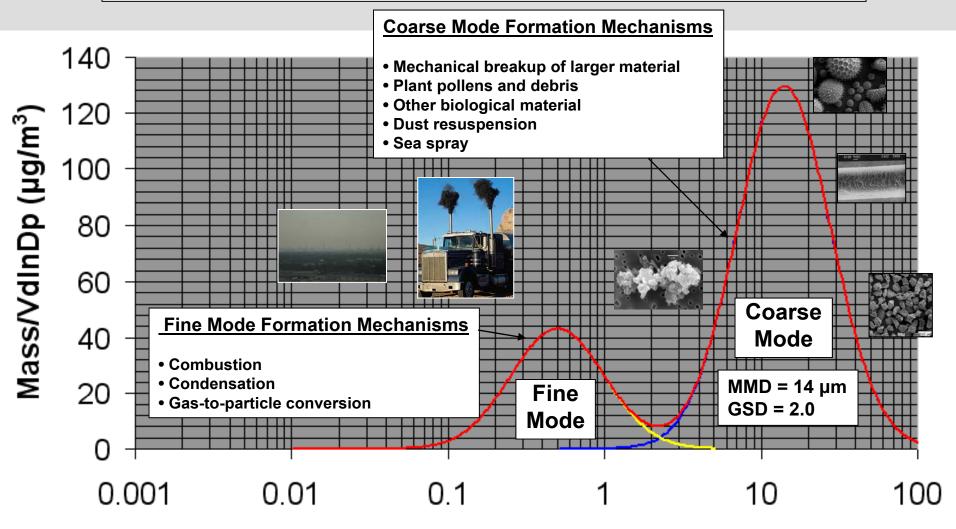
- To protect public health, the EPA has developed national ambient air quality standards (NAAQS) for airborne particulate matter (PM). Compliance with the PM NAAQS must be measured using EPA-approved samplers which were developed to accurately measure PM concentrations independent of wind speed and direction. FRM development was thoroughly peer-reviewed and independent evaluations of FRMs have validated their size-selective performance.
- The agricultural industry typically generates airborne PM with much larger mass median diameters (MMDs) than urban dusts, and has expressed the belief that certain agricultural operations are being over-regulated by EPA due to an over-estimation of PM emissions from these operations.
- In particular, the industry has stated that "...all EPA-approved federal reference method (FRM) samplers do not accurately measure PM concentrations in the presence of the large PM that is typical of PM emitted by agricultural operations. The term for this phenomenon is "over-sampling"."
- These "over-sampling" statements are directed towards source sampling methods (i.e., in-stack) as well as ambient sampling methods.

Background (cont)

- Representatives of the agricultural industry have conducted their own laboratory and field evaluations of EPA's PM_{2,5} and PM₁₀ FRM and have stated that these FRMs over-sample by a factor of 1.5 to 10 (i.e., 150% to 1000% over-sampling). Similar statements have been made regarding sampling of PM_{2.5} aerosols.
- The agricultural industry postulates that the mechanism for this over-sampling is the change in the sampler's size selective performance (i.e., cutpoint and slope) in the presence of large agricultural dusts.
- EPA has thoroughly examined these statements and does not agree with the fundamental basis for these statements, the methodology upon which the statements were based, nor the conclusions drawn from this research.
- During previous discussions, EPA and agricultural research staff have acknowledged that these are complex technical issues. EPA research staff appreciates the critical importance of accurately regulating the agricultural industry, and also for ensuring that public health is protected with an adequate margin of safety.

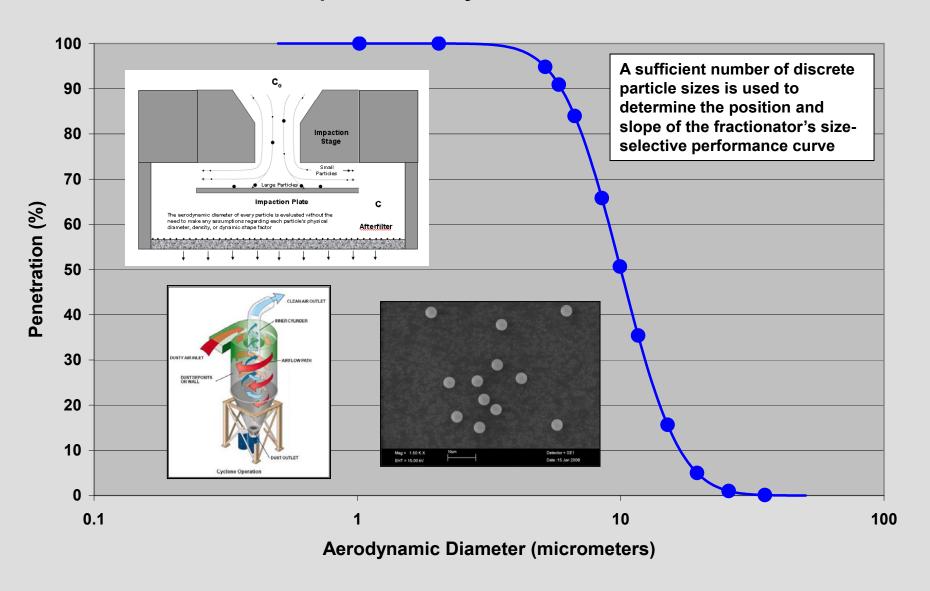
Characteristics of Ambient Particulate Matter

Ambient aerosols are bimodal in size and the relative modal concentrations can vary with site, season, and local activity. Modes are typically lognormal in shape.

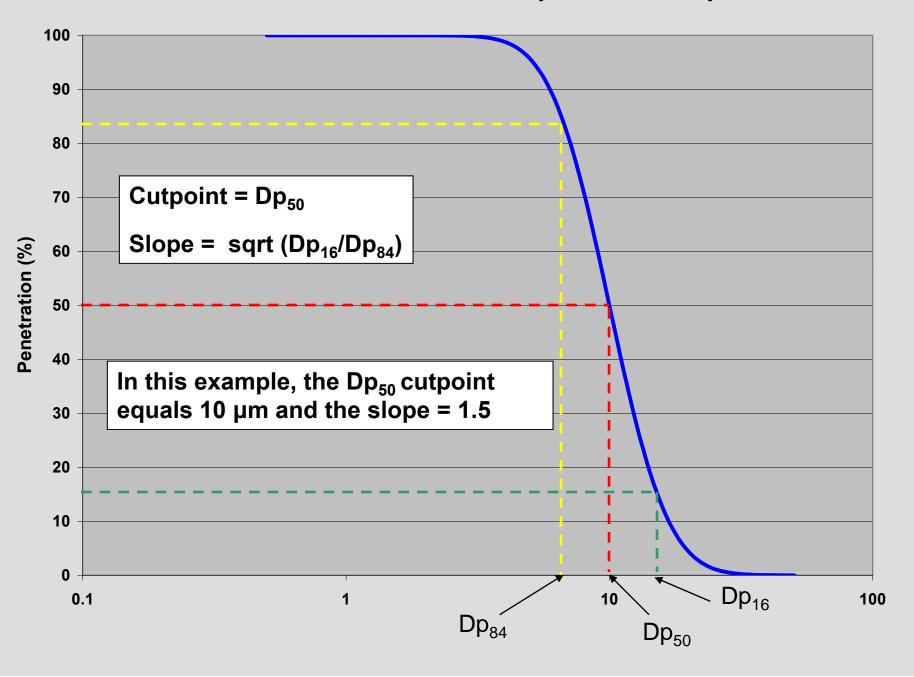


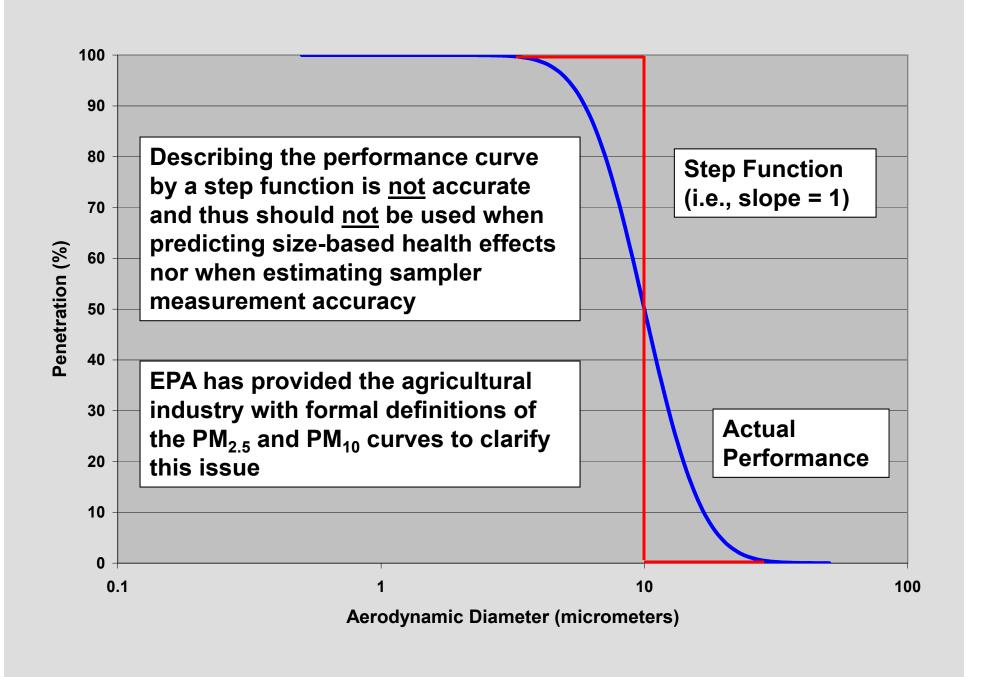
Aerodynamic Diameter (micrometers)

Laboratory Evaluation of Inertial Fractionators using Monodisperse, Primary Calibration Aerosols

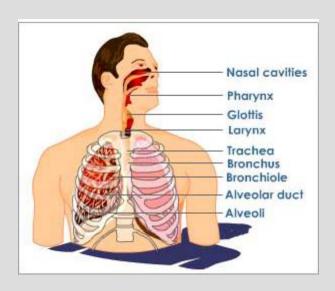


Definitions of Fractionator Cutpoint and Slope

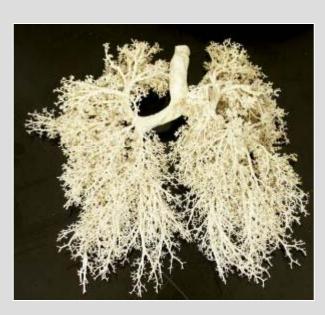


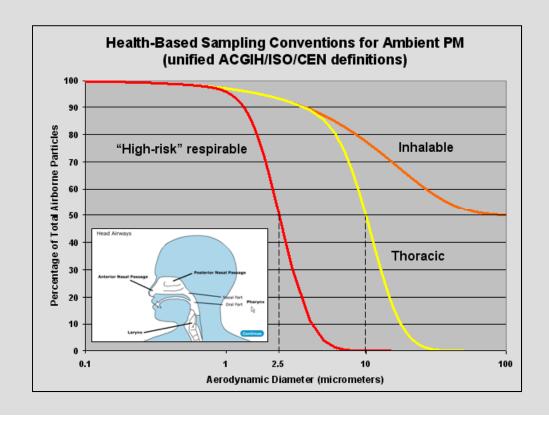


Physiological Basis for Health-Based PM NAAQS

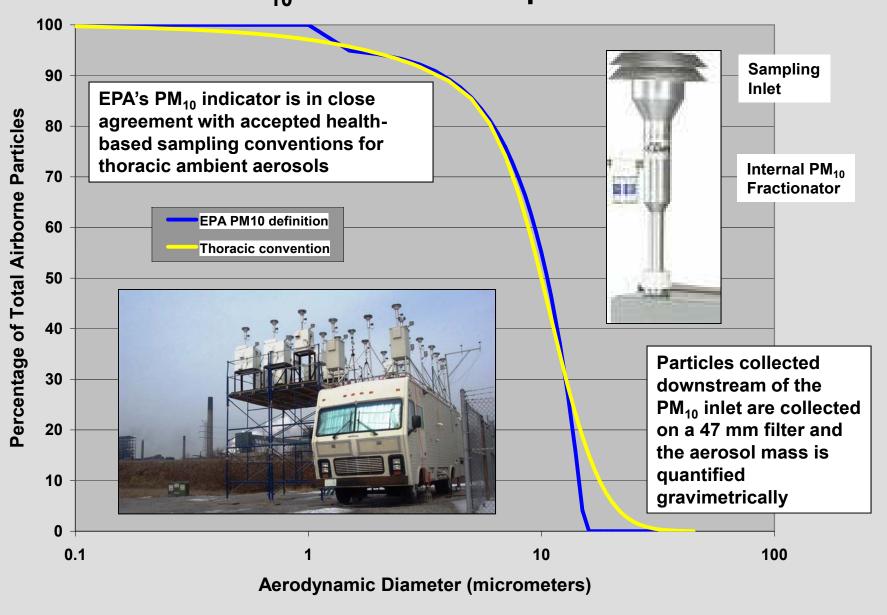


Since the 1970's, results from epidemiological studies, toxicological research, and deposition research have demonstrated that adverse health effects from exposure to airborne particles are primarily associated with those particles capable of entering the thoracic region of the human respiratory system (i.e., below the larynx)

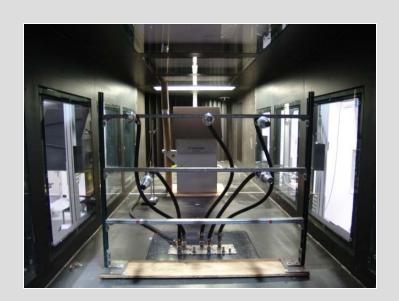


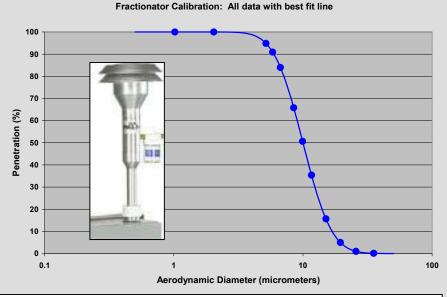


PM₁₀ Method Development



Wind Tunnel Evaluation of Size Selective Performance

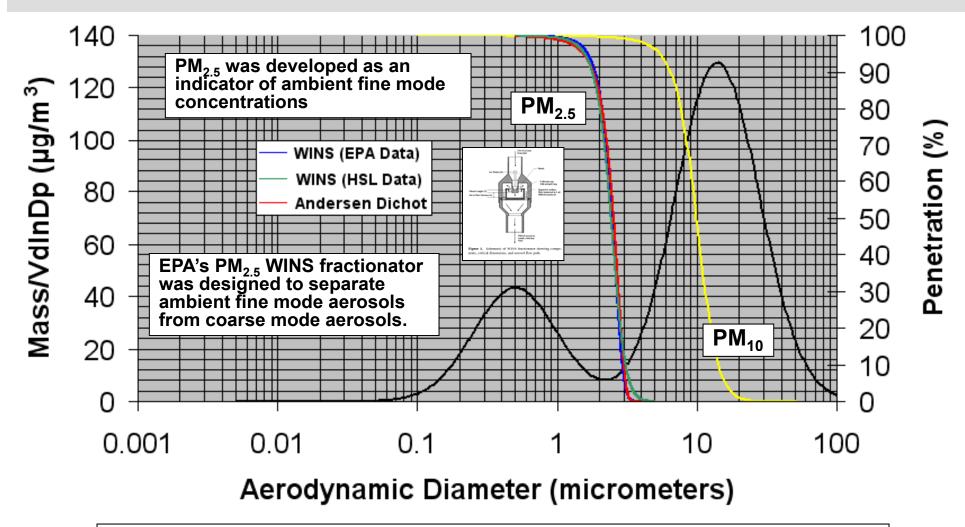




The size-selective performance of PM_{10} samplers must be demonstrated in an aerosol wind tunnel at wind speeds of 2, 8, and 24 km/hr, using monodisperse aerosols from 3 to 25 μ m diameter.

Performance of the Low-Volume PM ₁₀ Dichotomous Inlet								
			Cutpoint (μm)					
Reference #	Test Aerosol	Inlet Model	2 km/hr	8 km/hr	24 km/hr			
McFarland and Ortiz (1984)	monodispersed aerosol	SA 246B	9.9	10.2	10.0			
VanOsdell and Chen (1989)	monodispersed aerosol	SA 246B	9.8	10.0	9.9			
VanOsdell (1991)	monodispersed aerosol	R&P 10 µm inlet	9.8	-	9.6			
Tolocka et al. (2001)	monodispersed aerosol	Louvered dichot inlet	9.9	10.3	9.7			

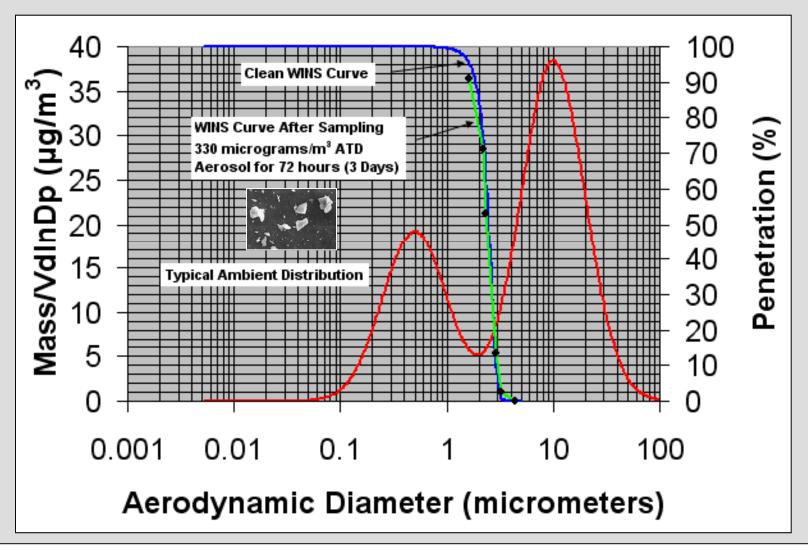
PM_{2.5} Method Development



During size-selective calibration of the WINS, very favorable inter-laboratory agreement was observed between EPA and England's Health and Safety Laboratory, for both cutpoint and slope.

The size-selective performance of the $PM_{2.5}$ WINS compares closely with the performance of the Andersen dichot's virtual impactor. The dichot had been used extensively during epidemiological studies and served as the basis for the position and shape of EPA's promulgated $PM_{2.5}$ performance curve.

Effect of WINS Loading on Size-Selective Performance



Sampling of high concentrations of Arizona Test Dust (ATD) results in a $\frac{\text{decrease}}{\text{decrease}}$ in WINS cutpoint and thus does $\frac{\text{not}}{\text{decrease}}$ result in positively biased PM_{2.5} mass concentrations

BGI's Very Sharp Cut Cyclone: An Alternative Fractionator for EPA's PM_{2.5} FRM



After loading the VSCC with ATD equivalent to 90 days sampling at a concentration of 150 µg/m³, no change in the VSCC's cutpoint or slope was noted

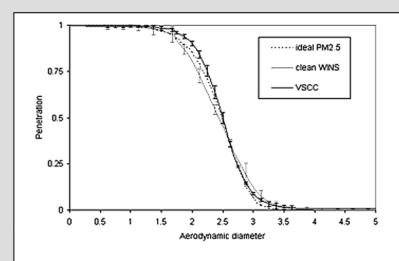


Figure 2. Comparison of WINS and VSCC showing curves fitted to data.

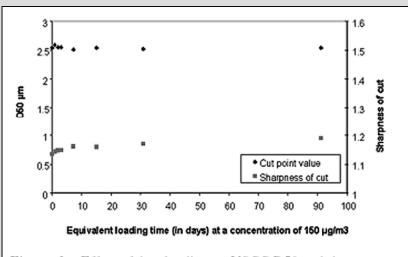


Figure 3. Effect of dust loading on VSCC D50 and sharpness of cut.

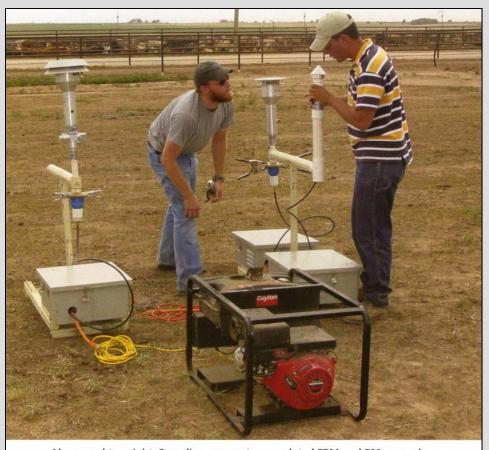
Definition of "Oversampling"

"True" PM_{10} is the mass fraction of the mass less than 10 μ m AED obtained from a particle size distribution of PM captured with a TSP sampler, times the measured TSP concentration.

- 1) Gravimetrically determine the mass concentrations of TSP and PM₁₀ using collocated low-vol TSP (LVTSP) samplers and PM₁₀ FRM samplers
- 2) Determine the particle size distribution (PSD) of collected PM on the LVTSP filter using Coulter counter analysis
- 3) Calculate the mass fraction of the collected LVTSP less than 10 µm AED from the measured PSD
- 4) Calculate the "true" PM₁₀ concentration by multiplying the mass fraction times the LVTSP mass concentration
- 5) Calculate oversampling as:

Oversampling =
$$\left[\frac{\text{FRM PM}_{10}}{\text{"True" PM}_{10}} - 1 \right] * 100 \%$$

Step 1: Gravimetrically determine mass concentrations of TSP and PM₁₀ using collocated LVTSP samplers and PM₁₀ FRM samplers



Above and top right: Sampling crew sets up co-lated FRM and PM₁₀ samplers.

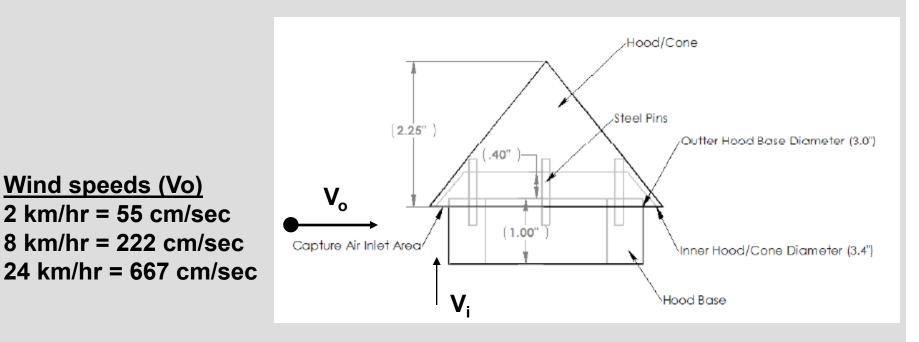
September/October 2010 RESOURCE

Problem:

TAMU's low-volume TSP (LVTSP) is fundamentally incapable of measuring total mass concentration in the ambient air. This will result in negatively biased mass concentrations, negatively biased "True PM₁₀", and thus will overestimate the calculated FRM "over-sampling".

Design of the TAMU Low-Vol TSP (LVTSP) Sampler Flow rate = 16.7 Lpm

Wind speeds (Vo) 2 km/hr = 55 cm/sec8 km/hr = 222 cm/sec

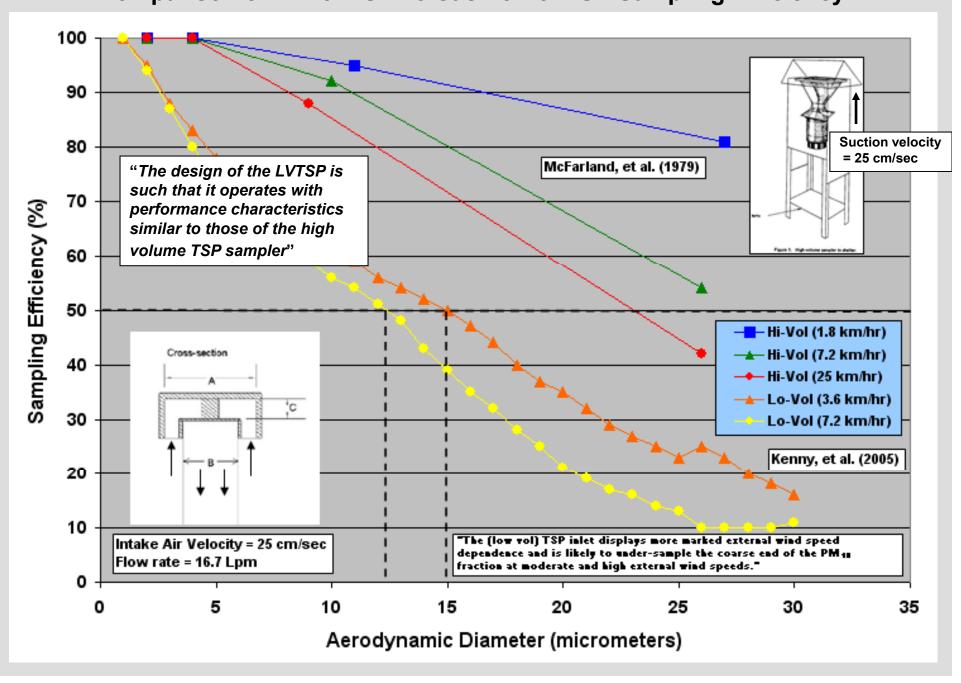


Design intake air velocity (Vi) = 21.5 cm/sec (i.e., designed to approximately match the settling velocity of 100 µm particles

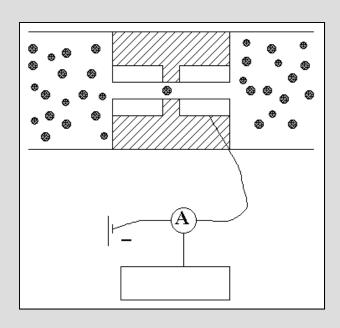
Aspiration Efficiency α 1/St α D_i/(ρ_p D_p² V_o)

Result: Small inlets with low sampling flow rates have poor aspiration efficiency particularly with respect to large particles in conjunction with high wind speeds.

Comparison of Hi-Vol TSP versus Lo-Vol TSP Sampling Efficiency



Step 2: Determine the particle size distribution (PSD) of collected PM on the LVTSP filter using Coulter counter analysis



The Coulter method measures the <u>volume</u> of individual particles suspended in a lithium chloride/methanol electrolyte.

Problem 1:

The Coulter method assumes that all particles (independent of particle size) are removed from a filter with 100% efficiency without any change in the original aerosol's size distribution or mass concentration

Step 2 (cont): Determine the particle size distribution (PSD) of collected PM on the LVTSP filter using Coulter counter analysis

$$D_a = D_p (\rho_p / K \rho_a)^{0.5}$$

Problem 2:

Conversion of each particle's measured volume to mass requires knowledge of each particle's density (ρ_p) . Because the composition of ambient aerosols varies widely as a function of particle size, measuring and applying an "average density" to all measured particles is not accurate.

Problem 3:

Conversion of physical size to aerodynamic size requires knowledge of each particle's dynamic shape factor (K). Because dynamic shape factors vary substantially as a function of particle size, apply an "average shape factor" to all measured particles is not accurate.

Step 2 (cont): Determine the particle size distribution (PSD) of collected PM on the TSP filter using Coulter counter analysis

Problem 4: Size Limitations of the Coulter Counter

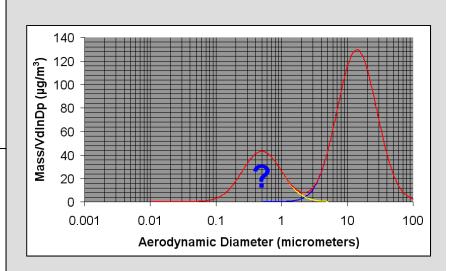
The Coulter counter is fundamentally incapable of detecting and measuring atmospheric fine mode particles, resulting in inaccurate size distribution determination. This will result in negatively biased mass concentrations and negatively biased "True PM_{10} "

"Coulter counter apertures can be used to measure particles within a size range of 2% to 60% of its nominal diameter. For a 100 μ m aperture, this translates to particle physical diameters between 2 μ m and 60 μ m."

Source: www.beckmancoulter.com

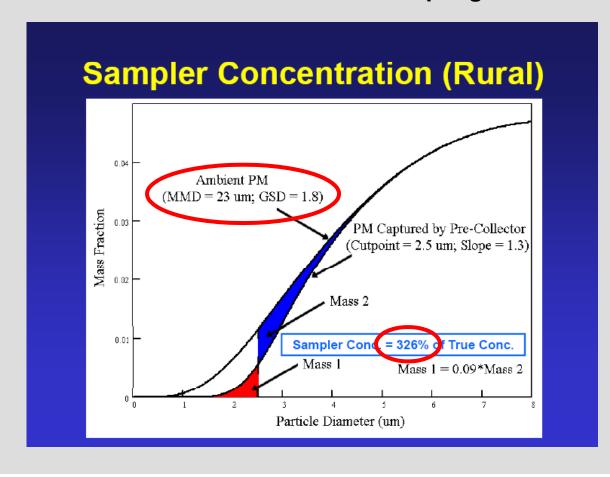
"TSP samplers quantified by a Coulter counter multisizer provide no information below an equivalent spherical diameter of 2 μ m and therefore underestimate respirable PM." Source: Park, et al. (TAMU) Atm. Env. 43 (2009) 280-289.

Use of the Coulter counter for ambient size distribution measurements thus <u>misses the entire fine mode</u> and can substantially underestimate total ambient mass concentrations.

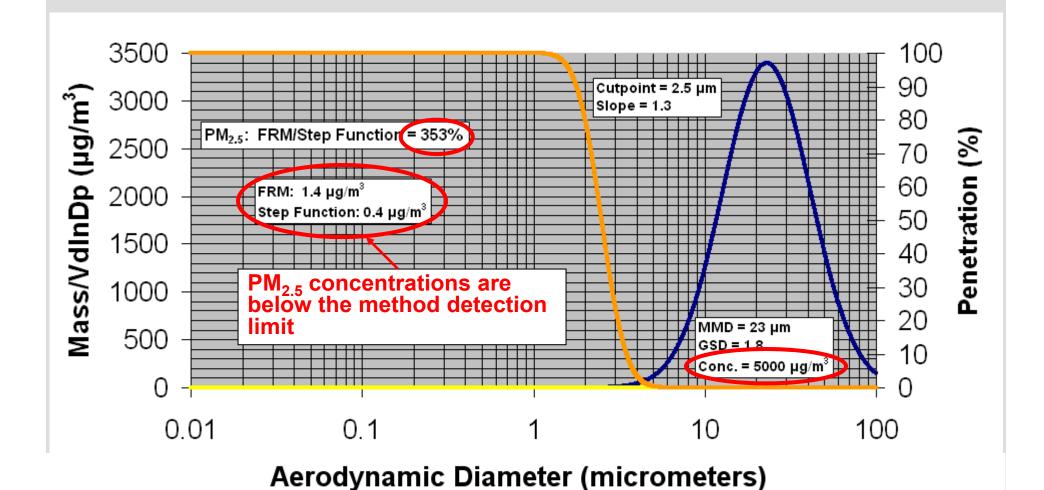


Step 3: Calculate the mass fraction of the collected TSP less than 10 µm AED from the measured PSD

<u>Problem</u>: TAMU's modeling of PM_{2.5} and PM₁₀ FRM performance as stepfunctions results in negatively biased mass concentrations, negatively biased "True PM", and thus over-estimates the calculated FRM "over-sampling".

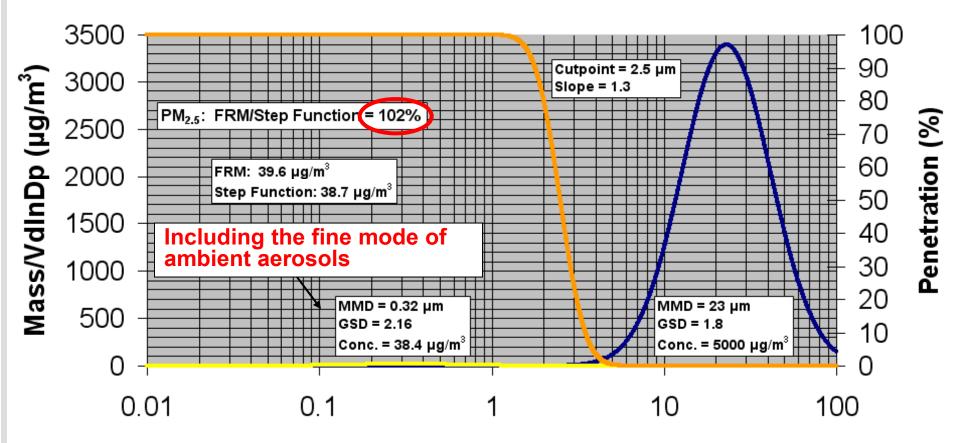


Example presented by Dr. Faulkner to EPA (Jan. 2010)



Even at extremely high coarse mode concentrations (5000 μ g/m³), predicted PM_{2.5} concentrations are well below the method detection limit. The use of percentages thus tends to over-emphasize what are actually minor differences in mass

concentrations.



Aerodynamic Diameter (micrometers)

Proper inclusion of an ambient atmosphere's fine mode virtually eliminates estimated differences between actual $PM_{2.5}$ concentrations versus those predicted using step functions. In this example, the fine mode concentration is less than 1% of the coarse mode concentration.

Example Publication



Estimating FRM PM₁₀ Sampler Performance Characteristics Using Particle Size Analysis and Collocated TSP and PM₁₀ Samplers: Cotton Gins, Buser, et al., 2008. Transactions of the ASABE, Vol. 51(2): 695-702.

Abstract

"Recent work at a south Texas cotton gin showed that ... the cutpoint and slope of the FRM PM $_{10}$ sampler shifted substantially and ranged from 13.8 to 34.5 μ m and from 1.7 to 5.6, respectively, when exposed to large PM as is characteristic of agricultural sources."

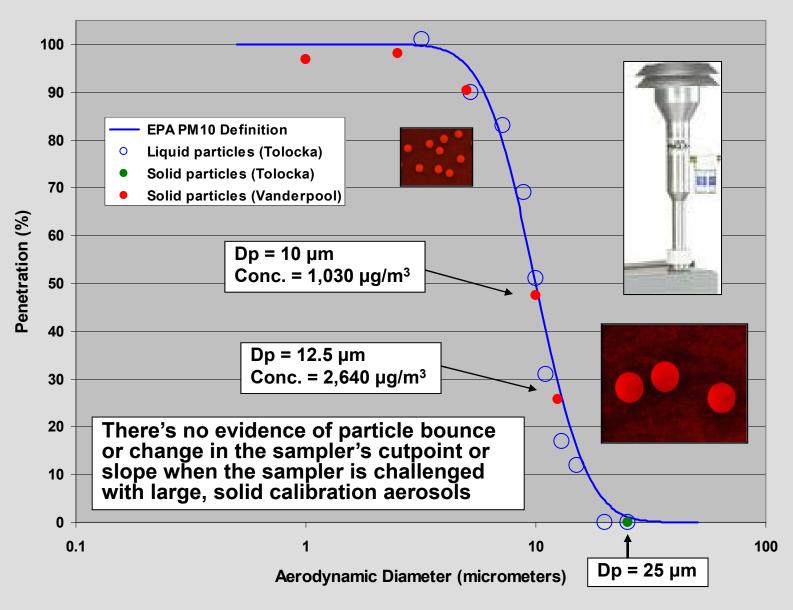
"These shifts in the cutpoint and slope of the FRM PM $_{10}$ sampler resulted in overestimation of true PM $_{10}$ concentrations by 145% to 287%."

MMD (μm)	GSD	Dust Conc. (μg/m³)	"True" PM ₁₀ (μg/m³)	FRM PM ₁₀ (µg/m³)	Estimated "Oversampling"	Estimated PM ₁₀ Cutpoint (µm)
13.6	2.3	1,385	494	1,099	122%	32.6

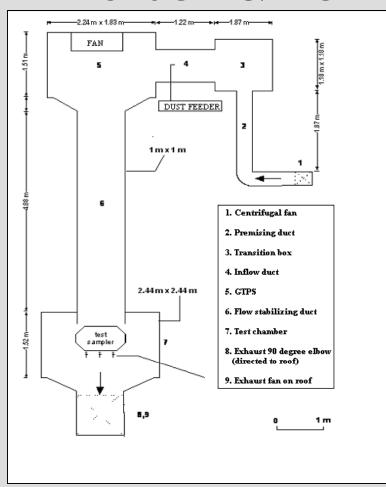
"The assumption is that <u>particle bounce</u> in effect increases the cutpoint of the inlet" (Parnell and McGee, 2010)

Size-Selective Performance of EPA's Louvered PM₁₀ Inlet as a Function of Solid and Liquid Monodisperse Calibration Aerosols

Source: Tolocka et al. (2001), Vanderpool (2008)



Texas A&M's Aerosol Wind Tunnel



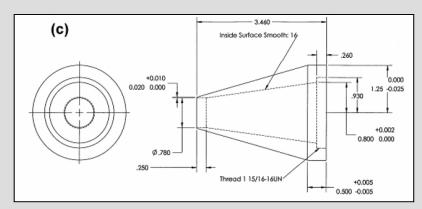
Schematic of TAMU's 1 m x 1m wind tunnel

Wind speeds = 2, 8, 24 km/hr

Tests showed that the tunnel met EPA's requirements for spatially uniform aerosol concentrations and air velocities in the test section



Wind tunnel test section showing arrangement of isokinetic nozzle and PM₁₀ test samplers

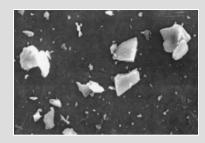


Schematic of 2 km/hr isokinetic nozzle used for representative aerosol collection independent of aerodynamic particle size

Test Aerosol: Polydisperse Arizona Test Dust (ADT)

- ATD is commercially available from Powder Technology Inc. in a range of sizes characteristic of agricultural dust emissions. The source material is collected in bulk from windblown topsoils in the Salt River Valley in Arizona.
- ATD is formulated under controlled conditions from naturally occurring materials to possess consistent particles sizes and chemical composition (NIST standards)
- ADT is large, dry, readily dispersed, and insoluble and nonreactive in the Coulter counter's electrolyte. ADT is thus a suitable test aerosol for Coulter counter analysis.
- Known properties (density and shape factor) allows computation of aerodynamic diameter from Coulter diameter
- ATD's primary components (68-76% SiO₂ and 10-15% Al₂O₃)
 are hard minerals and very suitable for conducting PM sampler
 bounce tests
- Sampler evaluation involves comparing the freestream particle size distribution (from the isokinetic sampler) to the size distribution of particles on the PM₁₀ FRM's after-filter



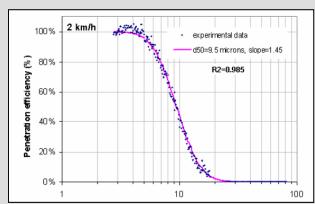


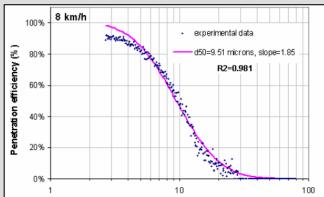
$$D_a = D_p (\rho_p / K \rho_a)^{0.5}$$

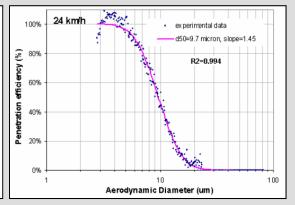
Density = 2.65 g/cm³ Shape factor = 1.4-1.5



Texas A&M's Wind Tunnel Evaluation of EPA's Louvered PM₁₀ FRM Inlet (Source: Chen and Shaw, 2007)







Performance of Low-Volume PM ₁₀ Dichotomous Inlets								
Reference #	Test Aerosol	Inlet Model	Cutpoint (µm)					
		iniet wodei	2 km/hr	8 km/hr	24 km/hr			
This study (TAMU)	polydisperse ATD	BGI PQ100/200	9.5	9.5	9.7			
McFarland and Ortiz	monodispersed aerosol	SA 246B	9.9	10.2	10.0			
VanOsdell and Chen	monodispersed aerosol	SA 246B	9.8	10.0	9.9			
VanOsdell	monodispersed aerosol	R&P 10 µm inlet	9.8	-	9.6			
Tolocka et al.	monodispersed aerosol	Louvered dichot	9.9	10.3	9.7			

"The results of dust wind tunnel testing clearly indicated that the cutpoint of BGI PQ100/200 louvered dichotomous PM_{10} inlet was within USEPA's requirement and the wind speed does not affect the cutpoint of the inlet."

Comparison of Measured to Predicted PM₁₀ FRM Cutpoint when Exposed to Large PM

Source: Chen and Shaw, 2007

Wind speed (km/hr)	Test Aerosol MMD (µm)	Test Aerosol GSD	ATD Dust Conc. (μg/m³)	Ratio of Dust Conc. to 24-hr PM ₁₀ NAAQS (Note: NAAQS = 150 μg/m³)	Measured Cutpoint (µm)
2	9.5	2.1	16,500	110	9.5
8	10.1	2.3	25,000	167	9.5
24	12.6	1.9	13,000	87	9.7

Source: Buser, et al. "Estimating FRM PM_{10} Sampling Performance Characteristics Using Particle Size Analysis and Collocated TSP and PM_{10} Samplers: Cotton Gins"

Sample No.	MMD (μm)	GSD	Dust Conc. (μg/m³)	"True" PM ₁₀ (μg/m³)	FRM PM ₁₀ (µg/m³)	Estimated "Oversampling"	Estimated PM ₁₀ Cutpoint (µm)
1	12.8	2.0	1,770	642	1,152	79%	23.1
2	13.4	2.1	852	294	687	134%	29.6
8	13.6	2.3	1,385	494	1,099	122%	32.6
11	10.4	1.8	603	284	557	96%	34.5
12	13.0	1.8	2,254	743	1,708	130%	22.9

Conclusion: Sampling of high concentrations of large aerosols characteristic of agricultural emissions clearly does <u>not</u> change the FRM sampler's size selective performance. Thus, no "oversampling" of agricultural emissions occurs.

Summary and Conclusions

- 1. EPA's FRM samplers for particulate matter were developed using strict design and performance criteria, have been thoroughly peer-reviewed, and their performance has been validated by independent researchers.
- 2. For the following reasons, the "True" method of estimating ambient concentrations is <u>inherently negatively biased</u> and should <u>not</u> be used for evaluating the accuracy of EPA's PM reference methods
 - Due to its design, the TAMU's low-vol TSP (LVTSP) <u>undermeasures</u> total mass concentrations, and its performance <u>decreases</u> with increasing particle size and increasing wind speed
 - The Coulter counter is incapable of accurately quantifying particles below approximately 2 µm and misses the entire fine mode of ambient aerosols, thus underestimating total mass concentration
 - Modeling PM_{2.5} and/or PM₁₀ performance curves using stepfunctions does not accurately reflect EPA's definition of these metrics, and results in an underprediction of actual mass concentration

Oversampling =
$$\left[\frac{\text{FRM PM}_{10}}{\text{"True" PM}_{10}} - 1 \right] * 100\%$$

Summary and Conclusions (cont)

- 3. The proposed mechanism of oversampling by particle bounce is not supported by actual laboratory or field tests for either $PM_{2.5}$ or PM_{10} FRM samplers.
- 4. Texas A&M's own wind tunnel tests of EPA's PM₁₀ FRM sampler conclusively prove that sampling of large, solid characteristic of the agricultural industry does <u>not</u> change the sampler's performance (i.e., cutpoint and slope), and thus does <u>not</u> result in oversampling of agricultural aerosols. These tests clearly demonstrate that the "True PM₁₀" method of assessing FRM performance is inherently inaccurate and should not be used.
- 5. EPA staff looks forward to continued discussions with agricultural industry representatives, and is committed to help resolve technical issues to help demonstrate that agricultural emissions are accurately measured and that agricultural operations are fairly regulated.