

Prepared in cooperation with the Prairie Band Potawatomi Nation

Water Quality on the Prairie Band Potawatomi Reservation, Northeastern Kansas, June 1996 Through August 2006

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Scientific Investigations Report 2007–5201

U.S. Department of the Interior U.S. Geological Survey

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By Heather C. Ross Schmidt, Heidi E. Mehl, and Larry M. Pope

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Suggested citation:

Schmidt, H.C.R., Mehl, H.E., and Pope, L.M., 2007, Water quality on the Prairie Band Potawatomi Reservation, northeastern Kansas, June 1996 through August 2006: U.S. Geological Survey Scientific Investigations Report 2007–5201, 75 p.

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Conversion Factors, Abbreviations, and Datums

Temperature in degrees Celsius (°C) may be converted to degrees Fahrenheit (°F) as follows:

 ${}^{\circ}$ F= (1.8× ${}^{\circ}$ C) +32.

Temperature in degrees Fahrenheit (°F) may be converted to degrees Celsius (°C) as follows:

 $^{\circ}$ C= ($^{\circ}$ F-32)/1.8.

Horizontal coordinate information is referenced to the North American Datum of 1983 (NAD 83).

Specific conductance is given in microsiemens per centimeter at 25 degrees Celsius (µS/cm at 25°C).

Concentrations of chemical constituents in water are given either in milligrams per liter (mg/L) or micrograms per liter (µg/L).

Water-Quality Abbreviations

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Water Quality on the Prairie Band Potawatomi Reservation, Northeastern Kansas, June 1996 through August 2006

By Heather C. Ross Schmidt, Heidi E. Mehl, and Larry M. Pope

Abstract

This report describes surface- and ground-water-quality data collected on the Prairie Band Potawatomi Reservation in northeastern Kansas from November 2003 through August 2006 (hereinafter referred to as the "current study period"). Data from this study period are compared to results from June 1996 through August 2003, which are published in previous reports as part of a multiyear cooperative study with the Prairie Band Potawatomi Nation. Surface and ground water are valuable resources to the Prairie Band Potawatomi Nation as tribal members currently (2007) use area streams to fulfill subsistence hunting and fishing needs and because ground water potentially could support expanding commercial enterprise and development.

Surface-water-quality samples collected from November 2003 through August 2006 were analyzed for physical properties, dissolved solids, major ions, nutrients, trace elements, pesticides, fecal-indicator bacteria, suspended-sediment concentration, and total suspended solids. Ground-water samples were analyzed for physical properties, dissolved solids, major ions, nutrients, trace elements, pesticides, and fecal-indicator bacteria. Chemical oxygen demand and volatile organic compounds were analyzed in all three samples from one monitoring well located near a construction and demolition landfill on the reservation, and in one sample from another well in the Soldier Creek drainage basin.

Previous reports published as a part of this ongoing study identified total phosphorus, triazine herbicides, and fecal coliform bacteria as exceeding their respective water-quality criteria in surface water on the reservation. Previous groundwater assessments identified occasional sample concentrations of dissolved solids, sodium, sulfate, boron, iron, and manganese as exceeding their respective water-quality criteria.

Fifty-six percent of the 55 surface-water samples collected during the current study period and analyzed for total phosphorus exceeded the goal of 0.1 mg/L (milligram per liter) established by the U.S. Environmental Protection Agency (USEPA) to limit cultural eutrophication in flowing water. Concentrations of dissolved solids frequently exceeded

the USEPA Secondary Drinking-Water Regulation (SDWR) of 500 mg/L in samples from two sites. Concentrations of sodium exceeded the Drinking-Water Advisory of 20 mg/L set by USEPA in almost 50 percent of the surface-water samples. All four samples analyzed for atrazine concentrations showed some concentration of the pesticide, but none exceeded the Maximum Contaminant Level (MCL) established for drinking water by USEPA of 3.0 µg/L (micrograms per liter) as an annual average. A triazine herbicide screen was used on 55 surface-water samples, and triazine compounds were frequently detected. Triazine herbicides and their degradates are listed on the USEPA Contaminant Candidate List. In 41 percent of surface-water samples, densities of *Escherichia coli (E. coli)* bacteria exceeded the primary contact, single-sample maximum in public-access bodies of water (1,198 colonies per 100 milliliters of water for samples collected between April 1 and October 31) set by the Kansas Department of Health and Environment (KDHE).

Nitrite plus nitrate concentrations in all three water samples from 1 of 10 monitoring wells exceeded the MCL of 10 mg/L established by USEPA for drinking water. Arsenic concentrations in all three samples from one well exceeded the proposed MCL of 10 µg/L established by USEPA for drinking water. Boron also exceeded the drinking-water advisory in three samples from one well, and iron concentrations were higher than the SDWR in water from four wells. There was some detection of pesticides in ground-water samples from three of the wells, and one detection of the volatile organic compound diethyl ether in one well. Concentrations of dissolved solids exceeded the SDWR in 20 percent of groundwater samples collected during the current study period, and concentrations of sulfate and chloride exceeded their respective SDWR in 10 percent of the ground-water samples. Concentrations exceeded the Drinking-Water Advisory Level set by USEPA for sodium in 50 percent of the ground-water samples.

Results from the current study period remained similar to results from previous study periods. The median triazine herbicide concentration (triazine screen by ELISA) for the current study period decreased slightly compared to past study periods. In the event that ground water on the reservation is to be used as a drinking-water source, additional treatment may be necessary to remove excess dissolved solids, sulfate, sodium, and chloride.

Introduction

Water quality is important to the Prairie Band Potawatomi Nation. Three of the four creeks draining the reservation (Soldier, Little Soldier, and South Cedar Creeks) in northeastern Kansas provide fishing and hunting resources for tribal members (fig. 1). Ground water is used as a domestic water supply within the reservation boundaries, and the potential of developing a stable ground-water supply is important as the tribe expands commercial enterprise and development. Establishing a baseline for ground-water quality on the reservation will help identify any potential treatment concerns for this prospective water source (Trombley, 2001). Presently, all drinking water on the reservation is purchased from Rural Water District #3 in Jackson County (Sharon Bosse, Prairie Band Potawatomi Nation Department of Planning and Environmental Protection, oral commun., 2007).

Nonpoint sources such as runoff from agricultural lands and point-source seepage from septic systems or sewage lagoons servicing a rural population are considered to be the primary sources of potential water-quality problems on the reservation (Trombley, 2001). Three documented wastewater facilities (including a sewage-treatment lagoon) are operated by the tribe on the reservation (fig. 1) and are potential sources of point-source wastewater discharge (Sharon Bosse, Prairie Band Potawatomi Nation Department of Planning and Environmental Protection, oral commun., 2006).

Under the Federal Clean Water Act of 1972, an Indian tribe is sovereign for the purposes of delegating the authority to regulate water within reservation boundaries. The Prairie Band Potawatomi Nation, in cooperation with the U.S. Geological Survey (USGS), began a water-quality monitoring program in 1996 with the purpose of establishing waterquality standards for the reservation. The program began as a surface-water assessment focused on nutrients, pesticides, bacteria, and total suspended solids and now includes major ions, dissolved solids, trace elements, and suspended-sediment analysis of surface-water samples as well as the addition of a ground-water-quality monitoring component using dedicated ground-water monitoring wells.

Purpose and Scope

The purpose of this report is to describe surface- and ground-water quality on the Prairie Band Potawatomi Reservation from November 2003 through August 2006 (hereinafter referred to as the "current study period") and to compare results from this study period to results from June 1996 through August 2003, which are published in previous reports

as part of this ongoing study. Surface-water samples were collected approximately every 3 months from November 2003 through August 2006, and ground-water samples were collected annually in July 2004, July 2005, and August 2006 to describe the distribution of inorganic and organic constituents and bacteria in water on the reservation.

This report is the fourth in a series of reports resulting from the ongoing cooperative study with the Prairie Band Potawatomi Nation. Previous reports describe the surfacewater quality on the reservation from June 1996 through November 1998 (Trombley, 1999), the surface- and groundwater quality on the reservation from February 1999 through February 2001 (Trombley, 2001), and the surface- and groundwater quality on the reservation from May 2001 through August 2003 (Schmidt, 2004).

Principal Results from Previous Study Periods

Total phosphorus and triazine herbicide concentrations, and fecal coliform bacteria densities in surface water were identified as exceeding water-quality criteria established by the U.S. Environmental Protection Agency (USEPA) and the Kansas Department of Health and Environment (KDHE) during the June 1996 through November 1998 study period (Trombley, 1999). The same constituents in surface water exceeded water-quality criteria established by USEPA and KDHE during the study period February 1999 through February 2001 (Trombley, 2001). Dissolved solids, sodium, sulfate, boron, iron, and manganese concentrations occasionally exceeded their respective water-quality criteria established by USEPA for drinking water during the February 1999 through February 2001 ground-water assessment and were related to large mineral concentrations resulting from dissolution of the surrounding sedimentary rocks (Trombley, 2001).

During the study period May 2001 through August 2003, 40 percent of surface-water samples exceeded the water-quality goal of 0.10 mg/L established by USEPA for total phosphorus. Triazine herbicides, specifically atrazine, were detected frequently in surface-water samples but did not exceed the USEPA Maximum Contaminant Level (MCL) of 3.0 µg/L for atrazine as an annual average (Schmidt, 2004). Fecal-indicator bacteria were detected in the majority of surface-water samples and exceeded water-quality criteria in 45 samples (12 of 65 exceeded USEPA recommended criterion for *E. coli*; 33 of 36 exceeded USEPA recommended criterion for enterococci). Dissolved-solids concentrations in some surface-water samples exceeded water-quality criteria established by USEPA and were greatest in water samples from Big Elm Creek, at a site located downstream from a sewagetreatment lagoon. Dissolved-solids concentrations exceeded USEPA Secondary Drinking-Water Regulations (SDWRs) in 20 percent of ground-water samples (Schmidt, 2004). Sodium, sulfate, nitrite plus nitrate, and boron concentrations in ground-water samples occasionally exceeded their respective water-quality criteria for drinking water as established by

Figure 1. Ecoregions and location of water-quality monitoring sites on and near Prairie Band Potawatomi Reservation, northeastern Kansas.

USEPA. Arsenic concentrations in two ground-water samples exceeded the MCL for arsenic; the largest arsenic concentration also corresponded to the largest iron concentration. Arsenic was detected in other ground-water samples as well but in concentrations that were substantially less than the MCL. Three volatile organic compounds (VOCs) were detected in one ground-water sample; these were carbon disulfide, ethyl methyl ketone, and tetrahydrofuran. The presence of these compounds may be explained by a nearby construction and demolition landfill (fig. 1; Schmidt, 2004).

Description of Study Area

The Prairie Band Potawatomi Reservation is located approximately 20 mi north of the city of Topeka in northeastern Kansas and encompasses approximately 121 mi² of Jackson County (fig. 1). Primary land uses in the two major drainage basins (Soldier and Little Soldier Creek Basins), which comprise about 86 percent of the reservation, are cropland (26 and 27 percent, respectively) and pastureland (68 and 64 percent, respectively) (Kansas Applied Remote Sensing Program, 1993). Cropland is mostly along the streams, and pastureland is in upland areas (table 1, fig. 2).

The Prairie Band Potawatomi Reservation is located within the Kansas Drift Plains section of the Central Lowland Province of the Interior Plains (Fenneman, 1938; Fenneman and Johnson, 1946; Schoewe, 1949). The topography of the Kansas Drift Plains reflects a gently undulating, glacial-driftdominated, erosional surface. The broad, smooth, wellrounded, interstream areas are remnants of the uneroded original basal till deposited by the last retreating glacial ice sheet. Glacial deposits of boulder till, sand and gravel, silt, and clay lie beneath the erosional surface. A preglacial bedrock valley, buried by glacial deposits, lies along the course of Soldier Creek within the reservation (Dreeszen and Burchett, 1971).

Geologically, the Prairie Band Potawatomi Reservation lies within the stable interior of the North American Continent. Since Precambrian time (about 600 million years ago), most of this part of the continent has undergone gentle upwarp and downwarp of the Earth's crust over large areas. Structurally, this part of the continent is characterized by broad basins and arches, with subtle folding of sedimentary rocks and few major fault zones (Jorgensen and others, 1993). Uppermost bedrock within the reservation consists mostly of limestone and shale of Permian and Pennsylvanian age (about 245 to 320 million years ago) (Walters, 1953). Unconsolidated glacial and stream (alluvial) deposits overlie the erosional surface of sedimentary rocks (fig. 3). The composition of these unconsolidated deposits varies vertically and horizontally and ranges from fine-grained sediment consisting of till, silt, and clay to coarse-grained sediment consisting of sand, gravel, pebbles, cobbles, and boulders (Trombley and others, 1996). The unconsolidated glacial and alluvial deposits are the primary source of ground water on the reservation. Well yields from

Table 1. Drainage area and land-use percentages in Soldier and Little Soldier Creek Basins within the Prairie Band Potawatomi Reservation, northeastern Kansas, 1993.

these deposits are largely variable and are generally less than 300 gal/min (Trombley and others, 1996).

The average monthly precipitation received on the Prairie Band Potawatomi Reservation over a 30-year period (1971 through 2000) and the monthly divergence from the average during the current study period (September 2003–August 2006) is shown in figure 4. During the current study period, the amount of precipitation received on the reservation was below the monthly average for 23 out of 36 months (fig. 4). The basin experienced precipitation significantly above average (3 in. or more above the monthly average) in June and October 2005, and August 2006. Precipitation data for the reservation are estimated from data collected near Holton, Kansas (fig. 1) by the High Plains Regional Climate Center (long-term average monthly precipitation) and the National Climatic Data Center (monthly precipitation for September 2003–August 2006). Precipitation data are available on the Web at *<http://www.hprcc.unl.edu/>* and *[http://cdo.ncdc.noaa.](http://cdo.ncdc.noaa.gov) [gov](http://cdo.ncdc.noaa.gov)*, respectively.

The average monthly streamflow in Soldier Creek near Delia (station 06889200, fig. 1) over a 30-year period (1971 through 2000) and the monthly divergence from the average during the current study period (September 2003–August 2006) is shown in figure 5. Streamflow was below average at this Soldier Creek site for most of the current study period (32 out of 36 months). Streamflow was markedly higher than the 30-year average in June and October 2005 (275 and 786 ft³/s, respectively). Low streamflow during the current study period probably affected some water-quality constituents such as suspended-sediment concentration, total suspended solids, and bacteria densities. Streamflow data for the Soldier Creek near Delia gaging station are available on the Web at *<http://waterdata.usgs.gov/nwis>* .

Months with above-average precipitation amounts generally increased streamflow in Soldier Creek, at least relative to preceding months. However, large precipitation months did not necessarily produce large mean monthly streamflow, relative

Figure 2. Land use in Soldier and Little Soldier Creek Basins within Prairie Band Potawatomi Reservation, 1993.

to the 30-year average. For example, in August 2006, the total precipitation was 9.32 in. for the month but was split between 11 days. The largest daily total was 2.83 in. on August 26th. Antecedent dry conditions for the four preceding months produced extreme low flow at the site for the first 2 weeks in August. These conditions resulted in an average streamflow for the month that was less than the 30-year average. Similar conditions contributed to lower-than-average streamflow despite above-average precipitation in August 2005.

Methods of Sample Collection and Analysis

Sample Collection

Water samples were collected from surface- and groundwater sites on and near the reservation to assess the quality of the Prairie Band Potawatomi Nation's water resources. Analyses were performed to identify potential contaminants and treatment concerns. This section describes the sampling sites, methods used to collect samples, analyses performed on the water-quality samples, and quality-assurance and qualitycontrol measures.

Surface Water

A network of 20 monitoring sites (referred to herein as "synoptic sites") was established in 1996 to evaluate the spatial and temporal variability of water quality on and near the reservation (fig. 1, table 2). Eighteen of the 20 sites were located within the two major drainage basins on the reservation, Soldier Creek and Little Soldier Creek, which eventually drain into the Kansas River at Topeka (fig. 1). Surface-waterquality sampling sites were spatially distributed across the reservation, representing surface water flowing into and out of the reservation, as well as surface water downstream from potential sources of contamination. Two sites (BC01 and SCC01) were located on streams that flow eastward into the Delaware River. All 20 sites were sampled at least annually

from 1996 to 2001 as part of previous study periods (results from these previous study periods are published in Trombley 1999, 2001; Schmidt, 2004). Six synoptic sampling events were conducted from 1996–2001. The last synoptic sampling event was conducted in June 2001 (Schmidt, 2004). Synoptic sampling is the collection of samples from many locations during a short period of time, typically a few hours.

Since November 1996, a subset of 5 of the 20 surfacewater-quality monitoring sites have been sampled approximately quarterly (referred to herein as "quarterly sites") (fig. 1). Two sites are located on Soldier Creek (sites SC03 and SC07), two are located on Little Soldier Creek (sites LSC03 and LSC04), and one is located on Big Elm Creek (site LSC06) (fig. 1) downstream from a sewage-treatment lagoon in operation on the reservation (Sharon Bosse, Prairie

Figure 4. Average monthly precipitation at Holton, Kansas, and departure from average during current study period, September 2003 through August 2006.

Band Potawatomi Nation Department of Planning and Environmental Protection, oral commun., 2006) and approximately 1 mi upstream from its confluence with Little Soldier Creek.

Fifty-five surface-water samples were collected from November 2003 through August 2006. These samples were analyzed for physical properties, dissolved solids, major ions, nutrients, trace elements, pesticides, fecal-indicator bacteria, suspended sediment, and total suspended solids (table 3). Streamflow was measured concurrently with each surfacewater-quality sample. Surface-water sampling and streamflow measurement techniques are described in Wilde and others (1999).

Ground Water

Ground-water samples were collected from 11 wells installed in August 2002 and July 2003 and dedicated solely for water-quality-monitoring purposes (fig. 1, table 2). A previous ground-water assessment on the reservation used water collected from domestic and pumping wells (Trombley, 2001). For the current study period, a total of 10 ground-water monitoring wells within the reservation were sampled in July 2004, July 2005, and August 2006. Monitoring well MW05 had been sampled during a previous study period (2003) but was omitted for the current study period because of low water production. Monitoring wells were located to provide a spatial assessment of the ground-water quality across the reservation. The monitoring wells, installed by a private contractor through the Prairie Band Potawatomi Nation, consist of 2-in. polyvinyl chloride (PVC) pipe and are completed at various depths in the local water-bearing formation (table 2) (Gregory Wold,

Prairie Band Potawatomi Nation, Department of Planning and Environmental Protection, oral commun., 2004). Water in these wells is derived from alluvial and glacial deposits of Pleistocene age as well as carbonate deposits of Pennsylvanian age (Walters, 1953).

Ground-water-quality samples were analyzed for physical properties, dissolved solids, major ions, nutrients, trace elements, pesticides, and fecal-indicator bacteria (table 3). Additional analyses for chemical oxygen demand (COD) and volatile organic compounds (VOCs) were completed on all three annual samples from monitoring well MW04 and one annual sample collected from well MW07 in July 2004. Ground-water-quality samples were collected and processed in a mobile water-quality laboratory following ground-watersampling protocols described in detail in Gibs and Wilde (1999). Each well was purged of at least three well volumes before samples were collected. Well purging was necessary to assure that water samples were representative of aquifer conditions. During the purging period, physical properties of the water (specific conductance, pH, water temperature, dissolved oxygen, oxidation-reduction potential, and turbidity) were observed and recorded every 5 minutes. Water-quality samples were collected after the three-volume purge was complete and physical properties stabilized (Gibs and Wilde, 1999).

After onsite processing, all water-quality samples were chilled to a temperature of 4° C during transport to the USGS laboratory in Lawrence, Kansas. Samples then were processed, preserved, and shipped according to the desired analysis to the USGS National Water-Quality Laboratory (NWQL), Lakewood, Colorado; the Organic Geochemistry Research

Figure 5. Average monthly streamflow at Soldier Creek near Delia streamflow-gaging station and departure from average during current study period, September 2003 through August 2006.

Laboratory (OGRL), Lawrence, Kansas; and the USGS Sediment Laboratory, Iowa City, Iowa.

Sample Analysis

Water-quality samples were analyzed at NWQL for physical properties, major ions, nutrients, trace elements, COD, and VOCs using methods described in Fishman and Friedman (1989), Zaugg and others (1995), Connor and others (1998), and Jones and Garbarino (1999). Dissolved pesticides, including herbicides and insecticides, were analyzed at OGRL using two different methods. Enzyme-linked immunosorbent assay (ELISA) was used for all surface-water samples to screen for triazine herbicides using procedures described in Thurman and others (1990). Selected surface-water and all ground-water samples were analyzed using gas chromatography/ mass spectrometry (GC/MS) for concentrations of selected pesticides using methods described in Kish and others (2000). Suspended-sediment concentrations in surface-water samples were determined at the USGS Sediment Laboratory in Iowa City, Iowa, using methods described in Guy (1977).

Fecal coliform, *Escherichia coli* (*E. coli*), fecal streptococcus, and enterococci indicator bacteria in water samples were processed onsite by USGS or tribal personnel. Bacteria samples were returned to the USGS laboratory in Lawrence, Kansas, to complete incubation and subsequent enumeration of bacteria colonies in each water sample. All bacteria (samples) were processed within 6 hours of collection using

membrane-filtration methods presented in Myers and Wilde (2003). Fecal coliform bacteria were cultured using the mFC method (American Public Health Association and others, 1992), *E. coli* bacteria were cultured using a modified mTec agar (U.S. Environmental Protection Agency, 2002a), and enterococci bacteria were cultured using the mEI method EPA 1600 (U.S. Environmental Protection Agency, 2000a).

Constituents that were detected at concentrations less than minimum laboratory reporting levels and that could be identified are reported as estimated. Estimated concentrations are noted with a remark code of "E" preceding the value in the data tables and should be used with the knowledge that there are increased uncertainties associated with these data (Childress and others, 1999). The remark code "E" also is used in the reporting of bacterial data where calculations to determine concentrations were based on nonideal counts of colonies. Ideal counts are ranges of colony counts that are defined for the individual bacteria types. The maximum limit of the range is established because overcrowding of the colonies inhibits the full growth of individual colonies causing atypical growth and increased errors in counting colonies. The lower limit of the ideal range is set as a number below which statistical validity becomes questionable (Britton and Greeson, 1989).

Quality Assurance and Quality Control

Field blank and replicate samples were collected and analyzed to identify variability in the collection and processing

Table 2. Surface- and ground-water-quality monitoring sites on Prairie Band Potawatomi Reservation, northeastern Kansas. **Table 2.** Surface- and ground-water-quality monitoring sites on Prairie Band Potawatomi Reservation, northeastern Kansas.

Table 3. Constituents analyzed as part of water-quality assessment on Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.

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Table 3. Constituents analyzed as part of water-quality assessment on Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued

[Unless otherwise noted, constituents were analyzed in both surface- and ground-water samples; ELISA, enzyme-linked immunosorbent assay]

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Table 3. Constituents analyzed as part of water-quality assessment on Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued

[Unless otherwise noted, constituents were analyzed in both surface- and ground-water samples; ELISA, enzyme-linked immunosorbent assay]

2 Sampled in ground water only.

of water-quality data. Water-quality samples were collected and processed in the field; therefore, field blank samples were collected throughout the study. Field blank samples were collected by passing blank water through the same sampling equipment as the environmental samples and processing the blank sample in the same manner as an environmental sample would be processed. The inorganic and organic blank water used are quality assured by the USGS National Field Supply Services Unit (Denver, Colorado). Quality-assurance data are on file with the USGS office in Lawrence, Kansas, and are available upon request.

 The purpose of a field blank sample is to determine the concentrations of the analytes of interest that could be present in the environment in which the samples are collected and processed (Wilde and others, 1999). Field blanks incorporate contamination from all potential sources, including environmental contamination, sample equipment, and laboratory analysis. For the current study period (November 2003 through August 2006) reported herein, one ground-water field blank was collected and analyzed in July 2004 for major ions, nutrients, and selected trace elements. Analytical results of this sample detected small concentrations of calcium (0.02 mg/L), boron (12 µg/L), and iron (estimated at 5 µg/L). All other analyzed constituents were less than reporting levels. These concentrations are small relative to environmental sample results; therefore, equipment and procedures used in collection of ground-water samples for the study period described in this report are environmentally insignificant sources of sample contamination. The field blank sample was not analyzed for pesticides in the 2004 sample; however, a blank sample collected in 2000, during the previous reporting period, was analyzed for pesticides and had no detection of the pesticides listed in table 3. Results from the previous study period, May 2001 through August 2003, indicated similar analytical results in field blank samples relative to the major ions, nutrients, and selected trace elements (Schmidt, 2004).

Replicate samples are a set of samples collected close in time and space. Replicate samples are collected to provide information on the variability introduced from collection, processing, shipping, and laboratory handling and analysis of the sample (Wilde and others, 1999). The current study period used concurrent replicate samples, meaning that two or more samples were collected consecutively from the same place using the same methods and analyzed identically.

The target goal for variability among analyses of replicate samples was a relative percentage difference (RPD) of ±20 percent except when constituent concentrations were at or near analytical detection limits. RPD was calculated as the difference in replicate analyses divided by the mean and expressed as a percentage.

 Thirteen surface-water and three ground-water replicate samples were collected and analyzed during the current study period described in this report. In the analyses of these 16 replicate-sample pairs, only three physical property analyses exceeded the ±20 percent RPD qualityassurance goal; the largest was a turbidity pair with an RPD of 29 percent collected from well MW09 on August 17, 2006. All replicate analyses for major ions were within the established quality-assurance goal except for replicate surface-water samples collected on May 19, 2004, from site LSC03 (fig. 1). This replicate-sample pair had RPDs for dissolved fluoride, silica, and sulfate, as well as dissolved solids and total suspended solids that exceeded the ± 20 percent RPD goal. These excesses ranged from 24 percent (dissolved solids and total suspended solids) to 33 percent (chloride). The results for this one replicate-sample pair from site LSC03 were examined, but no apparent reason for the relative inconsistent RPDs could be determined.

Nutrient and trace element constituents had replicatesample RPDs within the ± 20 percent quality-assurance goal except for two instances of dissolved ammonia plus organic nitrogen (23 and 26 percent, respectively) and one instance of dissolved iron (48 percent). Bacteria analyses showed the largest variability between replicate samples of all constituent groups. Among the four bacterial groups or species listed in table 3, a total of 27 paired-replicate analyses of surfacewater samples were performed. RPDs for these paired analyses ranged from 0 to 39 percent with 10 (37 percent of the replicates) exceeding the ±20-percent quality-assurance goal. This relatively large variability probably is a combination of factors that may include (1) nonrandom distribution of bacteria in surface water and samples of surface water, (2) nonrandom distribution in sample bottles, (3) interferences with colony development during the incubation period (turbidity, nontarget bacteria, water/chemical matrix), and (4) analyst inconsistency in distinguishing and counting target bacteria. However, with the possible exception of bacterial results and the sample of major ions from site LSC03 on May 19, 2004, variability in sample collection and analysis was within acceptable limits.

During the current reporting period, replicate analysis samples for pesticides listed in table 3 were not collected. However, during previous reporting periods (1996–2003), five surface-water replicate samples were collected and analyzed for pesticides. Most of the pesticides listed in table 3 had concentrations less than laboratory reporting levels. Replicate analyses of surface-water samples with concentrations of pesticides greater than laboratory reporting levels generally did not exceed the ± 20 -percent quality-assurance goal, with the exception of one paired analyses of metolachlor and simazine. RPDs for these paired analyses were 23 and 21 percent, respectively. For sampling sites or wells with replicate analyses, analytical results of these paired samples were averaged prior to assessment, presentation, and discussion in this report.

Water Quality on the Prairie Band Potawatomi Reservation

Surface-water-quality samples collected during November 2003 through August 2006 were analyzed for physical properties, dissolved solids, major ions, nutrients, trace

elements, pesticides, fecal-indicator bacteria, suspendedsediment concentration, and total suspended solids. Groundwater samples were analyzed for physical properties, dissolved solids, major ions, nutrients, trace elements, pesticides, and fecal-indicator bacteria. COD and VOCs were analyzed in selected ground-water samples. This section describes the constituents analyzed and the results obtained from sample analysis.

Physical Properties

Physical properties of water include specific conductance, pH, water temperature, turbidity, dissolved oxygen, hardness, oxidation-reduction potential, and acid neutralizing capacity or alkalinity (table 3). Physical properties of surface water were measured onsite with a multiparameter water-quality instrument that determined specific conductance, pH, water temperature, turbidity (YSI 6136 sensor, Yellow Springs Instruments, Yellow Springs, Ohio), and dissolved oxygen. Specific conductance, pH, water temperature, and dissolved oxygen also were measured in ground water with a multiparameter water-quality instrument. Ground-water turbidity was measured with a portable benchtop turbidimeter (HACH 2100AN, Hach Company, Loveland, Colorado). Specific conductance, pH, and turbidity also were measured in all water samples submitted to NWQL. The acid neutralizing capacity of surface-water samples and the alkalinity of ground-water samples were determined onsite through digital titration techniques (HACH digital titrator, Hach Company, Loveland, Colorado), and the acid neutralizing capacity for each sample also was determined in the laboratory.

The health of aquatic life in a stream may be affected by pH, water temperature, dissolved oxygen, and turbidity. Fish and macroinvertebrate species, which some fish depend on as a food source, can be sensitive to changes in pH, dissolved oxygen, and turbidity. The physical properties of a water sample also can provide information about the major ion chemistry of a sample. For example, dissolved major ions (calcium, sodium, chloride, and sulfate) have a tendency to increase specific conductance values (Hem, 1992), which can have implications for drinking-water treatment.

Specific conductance is a measure of the capacity of a sample of water to conduct an electrical current and is associated with the concentration of ionized substances in the water. Therefore, the specific conductance becomes greater as the amount of ionized substances in the water increases (Hem, 1992). The reporting units for specific conductance in this report are microsiemens per centimeter at 25° C (μ S/cm).

Onsite specific conductance measurements made at the five quarterly surface-water sites ranged from 172 to 1,170 µS/cm (table 4). The smallest specific conductance measurement of 172 µS/cm was recorded at site LSC04 on May 11, 2005, when streamflow was from runoff conditions $(189 \text{ ft}^3/\text{s})$. The largest specific conductance measurement, 1,170 µS/cm, was recorded at site LSC06 on November 19,

 2003 , when streamflow was 0.02 ft³/s. This site is downstream from a discharging sewage-treatment lagoon. Specific conductance measurements may be larger in streams affected by wastewater discharge (Pope and Putnam, 1997).

Specific conductance measurements made onsite in ground water ranged from 460 to 3,030 µS/cm (table 5). The median onsite specific conductance measurement in ground water was 692 µS/cm. The largest specific conductance was detected in monitoring well MW03, which is the deepest well sampled during the current study period (table 2). Monitoring well MW03 is located next to a pumping well (S28–4, fig. 1) that was sampled during the 1999–2001 study period when water sampled on September 22 and 24, 2000, had onsite specific conductance values of 2,650 and 2,560 µS/cm, respectively (table 7 in Trombley, 2001). These large specific conductance measurements also correspond to the largest concentrations of dissolved solids. The reason for these large values currently (2007) is not known.

The pH of water is a measure of the activity of the hydrogen ion and indicates the degree of acidity or alkalinity of a solution. A pH of less than 7.0 standard units is considered acidic, and a pH greater than 7.0 standard units is considered alkaline. Generally, the pH of natural stream water ranges from 6.0 to 8.5 standard units (Hem, 1992). In water, extremes on the pH scale can stress aquatic organisms in addition to having effects on surface- or ground-water chemistry because water with a smaller pH can have an increased capacity for erosion (Hem, 1992).

The median onsite pH in all surface-water determinations was 8.0 standard units and ranged from 6.8 to 8.3 standard units (table 4). This range is within KDHE freshwater aquaticlife criterion of 6.5 to 8.5 standard units (Kansas Department of Health and Environment, 2001). Onsite pH measurements of ground-water samples from the Prairie Band Potawatomi Reservation ranged from 6.6 to 7.6 standard units, with a median of 7.0 standard units (table 5) and were within the USEPA primary drinking-water criterion (6.5–8.5 standard units) (U.S. Environmental Protection Agency, 2006).

The specific conductance of water, pH, density, solubility of dissolved constituents, rate of chemical reactions, and biological activity in water are all temperature dependent. Water temperature in streams on the reservation ranged from 0.1 to 27.2 °C, with a median temperature of 13.0 °C, and varied seasonally (table 4). Water temperature in ground water ranged from 13.2 to 17.0 \degree C, with a median temperature of 14.0 \degree C (table 5). The highest ground-water temperature was recorded in well MW11 on July 26, 2004. An open, hand-dug well, with a water-surface depth of about 10 ft below land surface, is located 10 ft from well MW11, and the warmer air temperatures may heat the water in the open well and possibly account for the higher water temperatures in well MW11. The other two samples collected from well MW11 in July 2005 and August 2006 were closer to the median value of the samples collected from all wells.

Turbidity refers to the murkiness of water and is caused by particles or "suspended matter." Turbidity is measured by Table 4. Streamflow measurements and results of analysis of physical properities in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November **Table 4**. Streamflow measurements and results of analysis of physical properities in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006. 2003 through August 2006.

[ft3/s, cubic feet per second; µS/cm, microsiemens per centimeter at 25 degrees Celsius; oC, degrees Celsius; YSI, Yellow Springs Instruments Incorporated, Yellow Springs, Ohio; Hach 2100 AN, HACH

[ft3/s, cubic feet per second; µS/cm, microsiemens per centimeter at 25 degrees Celsius; °C, degrees Celsius; YSI, Yellow Springs Instruments Incorporated, Yellow Springs, Ohio; Hach 2100 AN, HACH

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Table 4. Streamflow measurements and results of analysis of physical properities in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November **Table 4**. Streamflow measurements and results of analysis of physical properities in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.-Continued 2003 through August 2006.—Continued [ft³/s, cubic feet per second; µS/cm, microsiemens per centimeter at 25 degrees Celsius; °C, degrees Celsius; YSI, Yellow Springs Instruments Incorporated, Yellow Springs, Ohio; Hach 2100 AN, HACH
Company, Loveland, Colo Company, Loveland, Colorado; FNU, formazin nephelometric units; NTRU, nephelometric units; NTU, nephelometric turbidity unit; mg/L, milligrams per liter; mg/L as CaCO₃, milligrams per [ft3/s, cubic feet per second; µS/cm, microsiemens per centimeter at 25 degrees Celsius; oC, degrees Celsius; YSI, Yellow Springs Instruments Incorporated, Yellow Springs, Ohio; Hach 2100 AN, HACH liter as calcium carbonate; --, not determined or not applicable; E, estimated; <, less than; >, greater than] liter as calcium carbonate; --, not determined or not applicable; E, estimated; <, less than; >, greater than]

Table 4. Streamflow measurements and results of analysis of physical properities in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November Table 4. Streamflow measurements and results of analysis of physical properities in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006. - Continued 2003 through August 2006.—Continued

Company, Loveland, Colorado; FNU, formazin nephelometric units; NTRU, nephelometric turbidity ratio units; NTU, nephelometric turbidity unit; mg/L, milligrams per liter; mg/L as CaCO3, milligrams per [ft^ys, cubic feet per second; µS/cm, microsiemens per centimeter at 25 degrees Celsius; °C, degrees Celsius; YSL, Yellow Springs Instruments Incorporated, Yellow Springs, Ohio; Hach 2100 AN, HACH
Commons I avaland Colora [ft3/s, cubic feet per second; µS/cm, microsiemens per centimeter at 25 degrees Celsius; oC, degrees Celsius; YSI, Yellow Springs Instruments Incorporated, Yellow Springs, Ohio; Hach 2100 AN, HACH

18 Water Quality on the Prairie Band Potawatomi Reservation, Northeastern Kansas, June 1996 through August 2006

Table 4. Streamflow measurements and results of analysis of physical properities in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November Table 4. Streamflow measurements and results of analysis of physical properities in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006. - Continued 2003 through August 2006.—Continued

Company, Loveland, Colorado; FNU, formazin nephelometric units; NTRU, nephelometric units; NTU, nephelometric turbidity unit; mg/L, milligrams per liter; mg/L as CaCO₃, milligrams per Company, Loveland, Colorado; FNU, formazin nephelometric units; NTRU, nephelometric units; NTU, nephelometric turbidity unit; mg/L, milligrams per liter; mg/L as CaCO₃, milligrams per (ft³/s, cubic feet per second; µS/cm, microsiemens per centimeter at 25 degrees Celsius; °C, degrees Celsius; YSI, Yellow Springs Instruments Incorporated, Yellow Springs, Ohio; Hach 2100 AN, HACH [ft3/s, cubic feet per second; µS/cm, microsiemens per centimeter at 25 degrees Celsius; oC, degrees Celsius; YSI, Yellow Springs Instruments Incorporated, Yellow Springs, Ohio; Hach 2100 AN, HACH liter as calcium carbonate; --, not determined or not applicable; E, estimated; <, less than; >, greater than] liter as calcium carbonate; --, not determined or not applicable; E, estimated; <, less than; >, greater than]

Monitoring-

Date of sample

Specific

Specific conduc-

pH,

Water

Dis-

Acid neutralizing

Acid neutral-

3 Water-quality criteria established by U.S. Environmental Protection Agency (1986) in support of freshwater aquatic life for warm-water species.

³ Water-quality criteria established by U.S. Environmental Protection Agency (1986) in support of freshwater aquatic life for warm-water species.

4 Value was determined to be erroneous due to equipment malfunction and was deleted from this data set.

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Table 5. Results of analysis of physical properties in ground-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and August 2006. **Table 5.** Results of analysis of physical properties in ground-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and August 2006.— Continued

[µS/cm, microsiemens per centimeter at 25 degrees Celsius; °C, degrees Celsius; NTRU, nephelometric units; NTU, nephelometric turbidity unit; mg/L, milligrams per liter; mg/L as CaCO₃, milligrams per liter as calcomate: [µS/cm, microsiemens per centimeter at 25 degrees Celsius; °C, degrees Celsius; NTRU, nephelometric turbidity unit; mg/L, milligrams per liter; mg/L as CaCO₃, milligrams per liter as calcium carbonate; mV, millivolts; --, not determined or not applicable; ≤, less than or equal to; <, less than; E, estimated]

quantifying the amount of light scattered by particles in the water. In this report, turbidity is described using three different units of measurement—formazin nephelometric units (FNUs), nephelometric turbidity units (NTUs), and nephelometric turbidity ratio units (NTRUs). These three turbidity units are not comparable and should be differentiated (Ziegler, 2002). Some suspended matter that causes turbidity in water is known to be harmful and includes lead, bacteria, viruses, and protozoan cysts such as *Giardia* and *Cryptosporidium*. Turbidity also can interfere with the disinfection of drinking water (Anderson, 2005).

The water-quality criteria for turbidity in finished drinking water established by USEPA state that at no time can turbidity in finished drinking water exceed 5 NTU, and filtration systems must ensure that turbidity in finished drinking water does not exceed 1 NTU in at least 95 percent of daily turbidity samples (U.S. Environmental Protection Agency, 2006). USEPA also has established water-quality criteria for onsite turbidity in streams and rivers according to ecoregions. An ecoregion is defined as an area of similar climate, landform, soil, potential natural vegetation, hydrology, or other ecologically relevant variables (U.S. Environmental Protection Agency, 2001). The Prairie Band Potawatomi Reservation is located just north of the convergence of three of USEPA's defined Level III ecoregions. These are ecoregion IV, Great Plains Grass and Shrublands; ecoregion VI, Corn Belt and Northern Great Plains; and ecoregion IX, Southeastern Temperate Forested Plains and Hills (U.S. Environmental Protection Agency, 2000 b,c; 2001). Ecoregions IV and VI converge within reservation boundaries (fig. 1). Each ecoregion is assigned a unique turbidity criterion; the largest turbidity criterion on the reservation is assigned to ecoregion IV and is 6.4 NTU (U.S. Environmental Protection Agency, 2001). Turbidity was measured onsite with a YSI 6136 turbidity sensor (Yellow Springs Instruments Incorporated, Yellow Springs, Ohio).

Onsite turbidity measurements in the streams ranged from less than 0.10 to greater than 1,500 FNU (table 4). The largest onsite turbidity was measured in Little Soldier Creek at 134 Road near Mayetta (site LSC04) on May 11, 2005. This sample was collected during runoff when streamflow was 189 ft³/s, the largest streamflow measured during sample collection (1996–2006) since a sample collected at $3,050$ ft $\frac{3}{s}$ from site SC03 on June 22, 1999 (Trombley, 2001). Turbidity was not measured during the 1999 sampling period. The median turbidity concentration for all sites was 13 FNU, and 70 percent of turbidity measurements exceeded the water-quality criterion of 6.4 NTU for ecoregion IV. Turbidity samples also were collected and analyzed by NWQL using a HACH 2100 AN in differential mode (HACH Company, Loveland, Colorado). These measurements ranged from less than 2 to 3,600 NTRU (table 4). The largest laboratory turbidity for a surface-water sample also was collected at site LSC04 on May 11, 2005. The difference between the two values demonstrates the limits of the onsite sensors to record high turbidity values. The maximum detection limit for a YSI 6136 is

1,500 NTU, but the turbidity for this sample was much larger and would have been underestimated if not for the laboratory sample.

Onsite turbidity measurements in ground water were made using a HACH 2100 P turbidimeter (HACH Company, Loveland, Colorado). Turbidity also was measured in the laboratory by a HACH 2100 AN turbidimeter. Onsite turbidity concentrations in ground-water samples ranged from 0.22 to greater than 1,000 NTRU. The largest onsite turbidity was measured in a sample from well MW09 collected on August 17, 2006 (table 5). However, the corresponding laboratory value is much less (170 NTRU). Larger turbidity values occurred in water from wells that had slow recovery rates during pumping and were difficult to sample with the submersible pump. Water sampled from wells MW04 and MW11 also had large turbidity values.

The concentration of dissolved oxygen in surface water is related to atmospheric reaeration, photosynthetic activity of aquatic plants, temperature, and salinity of the water (Radtke and others, 1998). Dissolved oxygen contributes to chemical reactions in water and in the survival of aquatic organisms. USEPA has established a freshwater aquatic-life support standard of not less than 5.0 mg/L and a mean of 6.5 mg/L for cold water (defined as water containing one or more species of the family Salmonidae or other cold-water species), and a minimum of 4.0 mg/L and a mean of 5.5 mg/L for warm water (level set to protect early-life stages of warm-water fish species as sensitive as channel catfish and other life stages of species such as the largemouth bass) (U.S. Environmental Protection Agency, 1986). Surface water on the reservation is considered to be warm water.

Dissolved oxygen concentrations in ground water are one indicator of the oxidation-reduction (redox) environment of the aquifer in the vicinity of the monitoring well. Oxygen is supplied to ground water through recharge and by movement of air through the unsaturated zone above the water table. As the oxygen moves through the system it reacts with oxidizable material. Measurable amounts of dissolved oxygen in ground water may persist for long distances if little reactive material is available along the flow path (Hem, 1992).

Onsite dissolved oxygen measurements in surface water collected during the current study period ranged from 5.0 to 18.4 mg/L (table 4), which are above the minimum waterquality criterion (4.0 mg/L for warm-water aquatic life). The median dissolved oxygen concentration for surface water was 9.4 mg/L. Dissolved oxygen concentrations in ground water ranged from 0.10 to 11.2, with a median concentration of 1.2 mg/L. The largest dissolved oxygen concentrations in ground water were detected in monitoring wells MW04 and MW09 (table 5). About one-half of the wells showed consistently anoxic (oxygen-depleted) conditions, which may affect certain activities such as the rate of biodegration of VOCs.

Hardness is a characteristic of water that causes an increase in soap consumption in water used for cleaning. The presence of calcium as well as magnesium (discussed in the section on "Major Ions") contributes to the "hardness"

of water. Hard water is typical in regions largely dominated by carbonate rocks (limestone and shale) at the land surface and causes increased expense in water treatment as well as sometimes requiring further treatment by consumers in the form of water softeners. The hardness of water is computed by multiplying the sum of milliequivalents per liter (meq/L) of calcium and magnesium by 50. The resulting hardness value is expressed as "hardness as calcium carbonate $(CaCO₃)$ " (Hem, 1992). Milligrams-per-liter values may be converted to milliequivalents per liter by multiplying the milligrams per liter by the reciprocals of the combined weights of the appropriate ions (Hem, 1992). Water with total calculated hardness concentrations of 0 to 60 mg/L is considered "soft," from 61 to 120 mg/L it is considered "moderately hard," from 121 to 180 mg/L it is considered "hard," and greater than 180 mg/L, "very hard" (Hem, 1992).

Total calculated hardness in surface-water samples ranged from 61 to 460 mg/L as $CaCO$, (table 4), with a median concentration of 300 mg/L. Total calculated hardness in ground-water samples ranged from 230 to 1,100 mg/L as CaCO₃ (table 5), with a median concentration of 330 mg/L. All ground-water sampled during the current study period (November 2003–August 2006) would be classified as "very hard," and may require further treatment with water softeners before it is used as a domestic supply.

The oxidation-reduction (redox) potential (Eh) is a numerical index of the intensity of oxidizing or reducing conditions within a system (Hem, 1992). Eh was measured as a part of the ground-water-quality sampling to determine the redox conditions of the ground-water environment in the vicinity of the well. The redox potential of a system is important because the amount of oxygen in an environmental system is one of the factors affecting the chemical precipitation and dissolution of rocks and minerals. Some water-quality constituents such as arsenic, iron, and selenium are sensitive to the redox environment, and any change to the redox potential in an environment potentially could cause these constituents to dissolve into solution from surrounding aquifer material or to precipitate out of solution. Both situations eventually could affect the permeability of the aquifer in addition to releasing potentially harmful constituents such as arsenic. Positive potentials indicate that the system is relatively oxidizing, and negative potentials indicate that it is relatively reducing (Hem, 1992).

Eh measurements in samples from the monitoring wells ranged from 30 to 510 millivolts (mV) relative to a standard hydrogen electrode, with a median value of 250 mV (table 5). The smallest redox potential was measured in a sample from monitoring well MW03 collected on July 12, 2005, and the largest potential was measured in a sample from monitoring well MW04 collected on July 14, 2005.

The acid neutralizing capacity (ANC) or alkalinity of a solution is defined as the capacity of solutes it contains to react with and neutralize acid (Hem, 1992). The properties of alkalinity evaluate the potential of the solution for some kinds of water-rock interaction or interaction with other material the water may contact (Hem, 1992). ANC is determined by titration of an unfiltered sample, whereas alkalinity is determined by titration of a filtered water sample.

The range of laboratory measurements of ANC in surface-water samples collected as a part of the current study period was 82 to 340 mg/L as $CaCO₃$, with a median concentration of 220 mg/L as $CaCO₃$ (table 4). Laboratory ANC measurements for ground-water samples ranged from 140 to 410 mg/L as $CaCO₃$, with a median concentration of 250 mg/L as $CaCO₃$ (table 5).

Dissolved Solids

Dissolved solids in natural water consist primarily of the cations calcium, magnesium, sodium, and potassium and the anions bicarbonate, sulfate, and chloride. Dissolved solids in surface water are a result of natural dissolution of rocks and minerals or of discharges from municipal or industrial sources (Hem, 1992). Dissolved-solids concentrations are used extensively as an evaluation of water quality. Freshwater has dissolved-solids concentrations less than 1,000 mg/L, whereas slightly saline water ranges from 1,000 to 3,000 mg/L (Hem, 1992). According to USEPA, excess dissolved solids and major ions are objectionable in drinking water because of unpalatable mineral tastes and higher costs because of corrosion or the necessity for additional treatment (U.S. Environmental Protection Agency, 1986). Consequently, USEPA established a Secondary Drinking-Water Regulation (SDWR) for dissolved solids of 500 mg/L (U.S. Environmental Protection Agency, 2006).

Dissolved-solids concentrations in surface-water samples collected for this study ranged from 120 to 696 mg/L, with the largest concentration occurring in a sample from site LSC06 (table 6, fig. 6*A*). The median concentration for dissolved solids in surface water was 400 mg/L. Twelve of the 55 surface-water samples (22 percent) collected during the current study period exceeded the USEPA SDWR of 500 mg/L for dissolved solids. Eight of 11 samples from site LSC06 exceeded 500 mg/L. Site LSC06 is located downstream from a discharging sewage-treatment lagoon, and larger concentrations of dissolved solids generally are found downstream from treated sewage when compared to upstream concentrations (Pope and Putnam, 1997). The remaining four samples exceeding 500 mg/L were from site SC03 and may reflect more mineralized water during low-flow conditions.

The range of dissolved-solids concentrations in groundwater samples was 186 to 2,360 mg/L. The largest concentration occurred in a sample from well MW03 collected on August 15, 2006. This concentration was approximately 5.7 times larger than the median concentration of 412 mg/L for all the ground-water samples (table 7, fig. 6*B*), and corresponds to large concentrations of major ions found in this sample. Currently (2007), the exact reason for the unusually large concentration of dissolved solids in water from well MW03 is unknown. Twenty percent of all ground-water samples had

Table 6. Results of analysis of dissolved solids and major ions in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August Table 6. Results of analysis of dissolved solids and major ions in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.

[mg/L, milligrams per liter; mg/L as HCO3, milligrams per liter as bicarbonate; mg/L as CaCO3, milligrams per liter as calcium carbonate; SDWR, Secondary Drinking-Water Regulation; DWA, Drinking-

[mg/L, milligrams per liter; mg/L as HCO₃, milligrams per liter as bicarbonate; mg/L as CaCO₃, milligrams per liter as calcium carbonate; SDWR, Secondary Drinking-Water Regulation; DWA, Drinking-

dissolved dissolved **(mg/L) Silica,** 9.6 7.5 8.5 6.2 9.1 8.4 9.7 8.4 9.7 7.1 10.02/25/25/25/04 263 57 13 13 14 14 14 15 10 10 13 14 14 14 15 16 16 17 18 18 19 19 19 19 19 19 19 19 19 19 1 05/19/04 209 48 11 7.9 4.1 -- -- 18 3.0 .3 9.6 08/25/04 182 -- 182 182 -- 12 12 13 13 13 13 13 13 13 13 142 -- 18 142 18 142 -- 18 142 18 142 18 14 02/10/05 325 72 18 13 -- - - 38 7.6 2.7 9.7 06/01/05 401 96 26 19 3.4 350 0 40 12 .4 7.5 02/25/04 226 48 11 7.7 4.6 150 0 35 6.6 .3 8.4 08/25/04 165 36 7.0 6.2 4.0 -- -- 16 3.1 .3 8.5 02/10/05 318 70 17 14 3.7 -- -- 39 8.9 .2 9.7 05/11/05 120 18 4.0 6.2 3.7 -- -- 14 3.1 .4 6.2 08/16/05 278 66 14 3.1 260 0 260 261 8.1 3.1 \overline{C} \overline{c} Ť \subseteq $\overline{13}$ $\overline{4}$ \mathbf{I} $\overline{16}$ $\overline{2}$ \equiv $\frac{1}{2}$ \Box Water-quality criteria1 500 -- -- 20 -- -- -- 250 250 2.0 -- 11/19/03 -- -- -- -- -- -- -- -- -- -- -- 12/08/04 424 424 -- -- 3.5 13 13 -- -- -- 58 58 38 13 08/16/05 442 100 24 20 3.1 420 0 41 20 .3 13 11 11 11 11 11 120 141 142 16 16 17 17 18 18 18 18 18 18 18 18 18 19 19 19 19 19 19 19 19 19 19 19 1 02/22/06 399 130 31 22 2.2 490 0 64 13 .3 12 06/06/06 347 76 18 15 3.0 320 0 32 8.5 .3 11 08/02/06 -- -- -- -- -- -- -- -- -- -- -- 11/19/03 -- -- -- -- -- -- -- -- -- -- -- 05/19/04 192 39 7.8 6.8 4.4 -- -- 18 3.3 .2 10 12/08/04 314 69 14 15 3.6 -- -- 34 12 .3 13 11/09/05 454 110 24 19 4.6 460 0 45 12 .3 14 02/22/06 386 386 386 396 310 310 310 0 0 63 15 31 **dissolved** dissolved **Fluoride,** (SDWR) (SDWR) (DWA) (SDWR) (SDWR) (SDWR) **(mg/L)** $0.\overline{3}$ $\ddot{\omega}$ $\ddot{\omega}$ $\ddot{\Omega}$ Ĵ. ن ن \dot{c} $\dot{\omega}$ ω ω 2.0 $\dot{\omega}$ $\mathcal{L}_{\mathcal{A}}$ $\overrightarrow{4}$ $\dot{\omega}$ $\ddot{}$ $\ddot{}$ $\dot{\mathcal{L}}$ $\overline{4}$ \mathbf{I} Ť **Chloride,** dissolved **dissolved** (SDWR) **(mg/L)** 250 3.0 7.6 6.6 8.2 3.2 8.5 3.3 3.1 8.9 3.1 8.1 \mathbf{I} $\frac{1}{2}$ \mathfrak{f} \overline{C} 28 \overline{C} $\overline{20}$ \equiv $\overline{13}$ \overline{C} $\overline{15}$ (SDWR) **Sulfate, solved (mg/L) dis-**250 18 $18\,$ 58 38 \overline{Q} $\overline{4}$ 48 \mathcal{Q} 32 35 18 16 34 39 14.49 $\frac{4}{5}$ 63 $\overline{1}$ Ť **ate, total, (mg/L as Carbononsite** $\overline{1}$ **CaCO3)** ÷ \mathbb{T} Ť. \circ \mathbf{I} $\overline{1}$ \mathbf{I} \mathbf{I} \mathcal{A} $\overline{1}$ $\overline{1}$ **(mg/L as** Bicarbon**ate, total, Bicarbononsite HCO3)** $\mathop{!}\nolimits$ \mathbb{I} $\frac{1}{2}$ 440 $\frac{1}{2}$ \mathbf{I} 180 350 320 $\mathord{\uparrow}$ $\mathop{!}\nolimits$ $\frac{1}{2}$ $\mathop{!}\nolimits$ 460 420 490 $\frac{1}{4}$ ł 150 $\overline{1}$ 260 470 **dissolved** dissolved **(mg/L)** 4.2 $3.\overline{5}$ 3.6 4.7 3.0 3.6 3.7 4.6 **Potas-** 4.7 3.4 3.1 2.2 4.6 4.4 4.0 3.7 3.1 2.5 **sium,** $\frac{1}{4}$ $\mathop{!}\rule{0pt}{.1ex}\hspace{-0.2ex}\mathop{!}\hspace{-0.2ex}$ Little Soldier Creek Basin Little Soldier Creek Basin **solved** (DWA) 7.9 7.2 $7.7\,$ 6.8 6.2 **(mg/L)** 6.2 **Sodium,** 23 $\overline{15}$ $\overline{4}$ $\overline{4}$ 24 **dis-** $\overline{10}$ $\overline{0}$ $20\,$ $\frac{8}{22}$ 15 $\overline{13}$ $\overline{20}$ $\overline{1}$ dissolved **dissolved nesium, (mg/L) Mag-** 7.8 $7.0\,$ 4.0 8.9 $\overline{4}$ 13 \overline{c} $\overline{18}$ \Box \overline{C} $\overline{16}$ $\frac{4}{3}$ $\frac{8}{2}$ $\frac{1}{2}$ \Box $18\,$ 8337 Ť $\overline{1}$ **dissolved Calcium, (mg/L)** $\overline{20}$ $\overline{30}$ 110 48 42 83 72 ∞ 76 48 39 36 $\sqrt{2}$ 18 66 $\mathop{!}\nolimits$ 57 96 $\frac{1}{4}$ ł. \overline{c} ÷ **Dissolved Dissolved** (SDWR) **solids (mg/L)** 500 182 470 226 165 314 318 278 209 424 399 192 120 454 386 263 325 $\overline{401}$ 442 347 $\frac{1}{2}$ $\frac{1}{4}$ \mathbf{I} **collection** collection **sample day/year) (month/** 02/25/04 05/19/04 02/25/04 **Date of** Little Soldier Creek, O Road 11/19/03 08/25/04 12/08/04 02/10/05 06/01/05 08/16/05 11/09/05 02/22/06 06/06/06 08/02/06 11/19/03 05/19/04 08/25/04 12/08/04 02/10/05 05/11/05 08/16/05 11/09/05 02/22/06 Water Advisory; --, not determined or not applicable] Water Advisory; --, not determined or not applicable] LSC03 Little Soldier Creek, O Road Little Soldier Creek, 134 LSC04 Little Soldier Creek, 134 Road near Mayetta Water-quality criteria¹ Road near Mayetta Site name **Site name** near Mayetta near Mayetta **Monitoring-Monitoring identifier** identifier **(fig. 1) site** LSC03 LSC04

Table 6. Results of analysis of dissolved solids and major ions in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August **Table 6.** Results of analysis of dissolved solids and major ions in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.

—Continued 2006.—Continued

[mg/L, milligrams per liter; mg/L as HCO₃, milligrams per liter as bicarbonate; mg/L as CaCO₃, milligrams per liter as calcium carbonate; SDWR, Secondary Drinking-Water Regulation; DWA, Drinking-[mg/L, milligrams per liter; mg/L as HCO3, milligrams per liter as bicarbonate; mg/L as CaCO3, milligrams per liter as calcium carbonate; SDWR, Secondary Drinking-Water Regulation; DWA, Drinking-

Table 6. Results of analysis of dissolved solids and major ions in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August Table 6. Results of analysis of dissolved solids and major ions in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued 2006.—Continued

[mg/L, milligrams per liter; mg/L as HCO₃, milligrams per liter as bicarbonate; mg/L as CaCO₃, milligrams per liter as calcium carbonate; SDWR, Secondary Drinking-Water Regulation; DWA, Drinking-
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Figure 6. Concentrations of dissolved solids in *(A)* surface- and *(B)* ground-water-quality samples, November 2003 through August 2006.
Table 7. Results of analysis of dissolved solids and major ions in ground-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and Table 7. Results of analysis of dissolved solids and major ions in ground-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and August 2006. August 2006. [mg/L, milligrams per liter; mg/L as HCO3, milligrams per liter as bicarbonate; mg/L as CaCO3, milligrams per liter as calcium carbonate; SDWR, Secondary Drinking-Water Regulation; DWA, Drinking-

mg/L, milligrams per liter; mg/L as HCO,, milligrams per liter as bicarbonate; mg/L as CaCO,, milligrams per liter as calcium carbonate; SDWR, Secondary Drinking-Water Regulation; DWA, Drinking-

solved (mg/L) Silica, dis- 858 299 $\frac{17}{18}$ $rac{66}{20}$ 277 2237 $\frac{22}{22}$ 122 $\frac{7}{17}$ 32.5 92 Monitoring well as the control of the control and the controller and controller and controller and contro 7/11/2005 437 95 1.6 3.10 0 82 1.2 3.28 1.9 2.7 8/16/2006 435 86 17 1.3 330 10 82 10 10 36 26 MW02 Monitoring well #2 7/27/2004 280 71 13 14 .58 270 0 29 3.2 .4 18 7/11/2005 285 72 14 13 .45 300 0 28 2.4 .4 19 8/15/2006 315 71 14 14 .54 300 0 28 3.2 .4 19 MW03 Monitoring well #3 7/26/2004 2,240 270 100 270 8.3 390 0 940 290 .3 32 7/12/2005 2,310 260 100 260 8.8 400 0 970 290 .3 32 022 Monitoring Monitoring Monitoring Destates and De 7/14/2005 401 68 18 53 .81 340 0 24 6.2 .5 22 8/16/2006 412 63 18 52 .85 380 0 25 6.5 .4 22 LT T. 9.1 9.6 OS O 081 06. DT 2.15 2.15 2.00 11 2.015 1002/05/212 941Tex attissing.com/NX 7/13/2005 50 176 176 1976 1976 1976 1976 1976 1981 1992 1993 1994 1994 1995 1992 1994 1994 1994 1994 1994 1994 8/14/2006 478 100 31 18 .92 470 0 49 4.3 .5 18 MW07 Monitoring well #7 7/27/2004 284 85 24 10 .84 380 0 17 2.9 .4 12 7/13/2005 338 8.8 8.8 8.8 8.8 8.9 .87 0 17 2.7 .4 .4 .4 8/15/2006 332 132 132 134 8.4 .81 2.4 .81 12 14 .4 .4 .4 .132 MW08 Monitoring well #8 7/27/2004 434 130 29 24 .82 460 0 47 25 .4 16 7/12/2005 530 530 470 470 30 30 470 470 33 33 41 41 42 530 530 530 531 41 531 42 8/14/2006 517 130 30 23 .82 490 0 48 23 .4 16 MW09 Monitoring well #9 7/29/2004 186 88 19 2.6 .57 340 0 9.5 1.5 .4 10 MW10 Monitoring well #10 7/28/2004 380 110 24 17 1.0 370 0 68 3.0 .3 17 7/14/2005 430 100 24 17 .68 350 0 66 2.8 .3 17 8/16/2006 429 94 24 16 .73 370 0 62 2.9 .3 17 7/14/2005 472 93 33 25 1.3 390 0 18 13 .8 32 8/18/2006 509 98 30 300 98 30 36 2006 309 98 36 21 440 0 32 14 .9 9 32 14 .9 32 32 33 33 33 33 33 33 33 33 33
Water-quality criteria established by the U.S. Environmental Protection Agency (2006). $\overline{1}$ 7.15 2,360 2,360 2,360 2,360 2,360 2,360 2,360 2,360 2,360 2,360 2,360 110 120 2,360 2,360 2,360 2,3 7/14/2005 304 85 20 2.9 .60 320 0 9.6 <.2 .5 11 8/17/2006 305 83 19 2.8 .65 340 0 9.2 1.5 .4 11 The mean of the state of th Water-quality criteria1 500 -- -- 20 -- -- -- 250 250 2.0 -- **Fluoride, dissolved** dissolved $\frac{2.0}{\text{(SDWR)}}$ (SDWR) (DWA) (DWA) (DWA) (SDWR) (SDWR) (SDWR) (SDWR) **(mg/L)** ن
تا نا تا 4 v v σ ∞ σ 4 4 $\overline{4}$ $m \, m \, m$ 4 N $\overline{4}$ 4 नं 4 4 4 $\overline{4}$ $4N4$ $m \, m \, m$ **dissolved** dissolved **Chloride,** SDWR) **(mg/L)** 2.7
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2142 4.3 $\frac{15}{10}$ 250 $\frac{4}{12}$ $\frac{1}{2}$ ลิลีลี 333 $\frac{6}{2}$ **dissolved** dissolved $\frac{250}{87}$
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82 **Sulfate, (mg/L)** 9.5 9.2 288 97800 235 នន $\frac{8}{3}$ $17/2$ 47.84 888 282 **ate, total, (mg/L as onsite Carbon-CaCO3)** $\circ \circ \circ$ $\circ \circ \circ$ \circ \circ $\circ \circ \circ$ $\overline{1}$ **Bicarbonate, total, (mg/L as** Bicarbon **onsite HCO3)** $\frac{90}{330}$ 370
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ماريخ \mathbf{I} **dissolved** dissolved 28.4 **Sodium, (mg/L)** $\frac{20}{(DWA)}$ 1780 2827 121 27000 538 ನನ ನ 272 <u>ິ</u>ລິສົ **dissolved** dissolved **Mag-nesium, (mg/L)** 991 882 288 $32₅$ 222 282 88
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 -130 284 888 888 727 883 883 $\overline{1}$ **Dissolved** Dissolved $\frac{500}{50 \text{N} \cdot \text{R}}$
 $\frac{500}{437}$
 435 **solids (mg/L)** Water Advisory; --, not determined or not applicable; <, less than] 2,240
2,310
2,360 Water Advisory; --, not determined or not applicable; <, less than] 315 380
 412 307
476
478 284
3332 434 517 888 888 375 285 509 **tion (month/** Date of sam-**Date of sam**ion (month/ **ple collecday/year)** 7/28/2004
7/11/2005 7/27/2004
7/11/2005 7/26/2004
7/12/2005
8/15/2006 7/29/2004
7/14/2005 7/29/2004
7/13/2005 7/27/2004
7/13/2005 7/27/2004
7/12/2005 7/29/2004
7/14/2005 7/28/2004
7/14/2005 7/26/2004
7/14/2005 8/16/2006 8/16/2006 8/14/2006 8/15/2006 8/14/2006 8/17/2006 8/16/2006 8/18/2006 8/15/2006 Monitoring well #11 Monitoring well #10 Monitoring well #3 Monitoring well #4 Monitoring well #6 Monitoring well #7 Monitoring well #8 Monitoring well #9 Monitoring well #2 Monitoring well #1 Site name **Site name** Water-quality criteria **identifier ing-site Monitor**identifier **(fig. 1)** MW07 MW11 MW01 MW06 **MW08 MW10** MW₀₂ MW₀₃ MW04 MW₀₉

¹Water-quality criteria established by the U.S. Environmental Protection Agency (2006).

dissolved-solids concentrations equal to or greater than the SDWR of 500 mg/L (table 7, fig. 6*B*). Dissolved solids were identified as exceeding the SDWR by Trombley (2001) and continue to exceed the SDWR in the current ground-water assessment in three of the monitoring wells (MW03, MW08, and MW11). Treatment to remove the dissolved solids may be needed before ground-water resources on the reservation are used as a source for public drinking-water supply.

Major Ions

Major ions in natural water typically are a result of the dissolution of rocks and minerals but also can occur in discharge from municipal or industrial sources (Hem, 1992). The major ions analyzed for the current study period included calcium, magnesium, sodium, potassium, bicarbonate, carbonate, sulfate, chloride, fluoride, and silica. SDWRs have been established by USEPA for sulfate (250 mg/L), chloride (250 mg/L), and fluoride (2.0 mg/L) concentrations in finished drinking water, and a Drinking-Water Advisory (DWA) has been established by USEPA for sodium concentrations (20 mg/L) in finished drinking water (U.S. Environmental Protection Agency, 2006). Results of analyses for major ions in surface- and ground-water samples are reported in tables 6 and 7, respectively.

Calcium is a major constituent of carbonate rocks such as limestone and dolomite, and it dissolves readily in water; therefore, the calcium concentration in water from areas with carbonate rocks and associated deposits tends to be larger than in other areas (Hem, 1992). Carbonate rocks dominate the surface and near-surface geology in Kansas including the Prairie Band Potawatomi Reservation (fig. 3). Large concentrations of calcium are objectionable in domestic water supplies because they contribute to the hardness of the water and tend to cause encrustations on cooking utensils, pipes, and in water heaters.

Calcium concentrations in surface-water samples ranged from 18 to 140 mg/L, with a median concentration of 86 mg/L (table 6). In ground-water samples, calcium concentrations ranged from 63 to 280 mg/L, with a median concentration of 95 mg/L (table 7).

Magnesium is a common alkaline-earth metal and is essential for plant and animal nutrition (Hem, 1992). The principal sources for magnesium on the reservation are carbonate rocks. Magnesium also contributes to the total hardness of water and may cause encrustation and increase soap or detergent consumption in water used for cleaning.

Magnesium concentrations in the surface-water samples collected for this study ranged from 4.0 to 31 mg/L, with a median concentration of 20 mg/L (table 6). Concentrations in ground-water samples ranged from 13 to 110 mg/L, with a median concentration of 25 mg/L (table 7).

Sodium is the most abundant member of the alkali-metal group of elements in the Earth and, once dissolved, tends to remain in solution. Natural sources of sodium include the weathering of plagioclase feldspar and the dissolution of

sodium salts from sedimentary rocks (Hem, 1992). Humanrelated sources include road salt, seepage from septic systems, and a by-product of water treatment (it is discharged by water softeners and reverse-osmosis units) (U.S. Environmental Protection Agency, 2003). Sodium in drinking water may impart a salty taste and could be harmful to persons suffering from heart, kidney, and circulatory diseases and women with toxemias of pregnancy (U.S. Environmental Protection Agency, 2003). Therefore, USEPA has established a DWA of 20 mg/L in finished drinking water for people who are on restricted sodium diets (500 mg/d) and a taste threshold of 30 to 60 mg/L in finished drinking water (U.S. Environmental Protection Agency, 2006).

Sodium concentrations in surface-water samples ranged from 6.2 to 140 mg/L, with a median concentration of 20 mg/L (table 6). The largest sodium concentrations were detected in samples from site LSC06. Six of 11 samples from this site equaled or exceeded the maximum taste threshold of 60 mg/L. The relatively large sodium concentrations in water at site LSC06 may be the result of discharge from an upstream sewage-treatment lagoon. Although sodium concentrations exceeded the USEPA DWA of 20 mg/L in almost 50 percent of surface-water samples, the concentrations at each of the other four sites generally were about 20 mg/L or less and never exceeded 30 mg/L (fig. 7*A*).

Sodium concentrations in ground-water samples ranged from 2.6 to 300 mg/L (table 7). The largest concentration was detected in a water sample from well MW03 (fig. 7*B*) and was 15 times larger than the median sodium concentration of 20 mg/L from all samples collected from ground-water monitoring wells. Sodium concentrations exceeded the USEPA DWA in 15 ground-water samples from wells MW01, MW03, MW04, MW08, and MW11 (a total of 50 percent of groundwater samples).

Potassium is an essential element for both plants and animals. Maintenance of optimal soil fertility depends on a supply of available potassium. Potassium is present in plant material and is lost from agricultural soil by crop harvesting and by leaching and runoff acting on organic residues. Although potassium is an abundant element and its common salts are highly soluble, it seldom occurs in large concentrations in natural water (Hem, 1992).

Concentrations of potassium in surface-water samples collected during the current study period ranged from 2.0 to 9.4 mg/L, with a median concentration of 3.8 mg/L (table 6). The largest concentration of 9.4 mg/L occurred in a water sample from site LSC06. Potassium concentrations from site LSC06 generally were larger than results from the other surface-water sites. This site had a median concentration of 6.0 mg/L, compared with a median of about 3.5 mg/L in samples from the other four sites.

Potassium concentrations in ground-water samples ranged from 0.45 to 8.8 mg/L, with a median concentration of 0.89 mg/L. The largest concentration in ground water was detected in water from monitoring well MW03 (table 7).

Figure 7. Concentrations of dissolved sodium in *(A)* surface- and *(B)* ground-water-quality samples, November 2003 through August 2006.

In surface water, bicarbonate concentrations generally are less than 200 mg/L, but larger concentrations in ground water are not uncommon (Hem, 1992). The primary source for bicarbonate on the Prairie Band Potawatomi Reservation is the dissolution of carbonate rocks and deposits.

Bicarbonate concentrations in surface-water samples ranged from 150 to 550 mg/L, with a median concentration of 315 mg/L (table 6). The largest concentration of bicarbonate occurred in water sampled from site SC07 on February 22, 2006.

Bicarbonate concentrations in ground-water samples ranged from 270 to 490 mg/L, with a median concentration of 370 mg/L (table 7). The largest bicarbonate concentrations in ground water were detected in samples from wells MW06 and MW08.

Natural sources of sulfate in water include the weathering of sulfur-bearing minerals, such as pyrite and gypsum, volcanic discharges to the atmosphere, and biologic and biochemical processes (Hem, 1992). Human-related sources include industrial discharges to both streams and the atmosphere and the combustion of fossil fuels, such as coal and gasoline. USEPA established a SDWR of 250 mg/L in drinking water to avoid laxative effects that can result from excessive sulfate concentrations in drinking water (U.S. Environmental Protection Agency, 2006).

Sulfate concentrations in surface-water samples ranged from 14 to 190 mg/L, with a median concentration of 67 mg/L. Sulfate concentrations were less than the SDWR of 250 mg/L (table *6,* fig. 8*A*).

Sulfate concentrations in ground-water samples ranged from 9.2 to 1,000 mg/L, with a median concentration of 32 mg/L (table 7, fig. 8*B*). Sulfate concentrations exceeded the SDWR in 10 percent of ground-water samples. The largest sulfate concentration was detected in water from well MW03 and was approximately 31 times the median concentration from all ground-water samples and four times the SDWR of 250 mg/L. The large sulfate concentration in water from well MW03 corresponds to large specific conductance values and large concentrations of other dissolved ions in the water from this well. Treatment to remove excess dissolved ion concentrations would be needed in the event this water should become a drinking-water source.

Chloride is present in all natural water but generally is found in small concentrations in surface water (Hem, 1992). The probable natural source of chloride on the reservation is dissolution of halite from sedimentary rocks. The discharge of human, agricultural, animal, or industrial wastes also may add substantial quantities of chloride to surface and ground water. Chloride can impart a salty taste to drinking water and may accelerate the corrosion of metals used in water-supply systems. On the basis of taste, USEPA established a SDWR of 250 mg/L in finished drinking water (U.S. Environmental Protection Agency, 2006). KDHE established an aquatic-life criterion of 352 mg/L for chronic exposure to chloride (Kansas Department of Health and Environment, 2001).

Chloride concentrations in surface-water samples collected during the current study period ranged from 3.0 to 200 mg/L and were less than the SDWR (table 6, fig. 9*A*). The median concentration of chloride in the surface-water samples was 17 mg/L. The concentrations of chloride from site LSC06 were substantially larger than from the other four sites, but these concentrations were still less than the SDWR. Relatively large chloride concentrations at this site probably are attributable to its location downstream from a discharging sewagetreatment lagoon.

Chloride concentrations in ground-water samples ranged from less than 0.2 to 300 mg/L, with a median concentration of 6.4 mg/L (table 7, fig. 9*B*). Chloride concentrations exceeded the SDWR in 10 percent of ground-water samples. The largest chloride concentration was detected in water from well MW03, and all samples from this well exceeded the SDWR.

Nutrients

Nitrogen and phosphorus are essential for the growth and reproduction of plants (Hem, 1992). Rooted aquatic plants and algae require dissolved forms of nitrogen and phosphorus as nutrients. Compounds of nitrogen, such as nitrite, nitrate, and ammonia, are the basic building blocks for protein synthesis, and phosphorus serves as an energy source in cellular chemical reactions. However, large inputs of nitrogen and phosphorus compounds to bodies of water can cause excessive algal growth (Hem, 1992). This may result in taste-and-odor problems in drinking water, reduce the aesthetic and recreational value of water, and stress aquatic organisms due to decreased dissolved oxygen concentrations during decomposition. Therefore, it is desirable to prevent or mitigate the introduction of excessive nutrient concentrations into surface water used as public supply or where sensitive aquatic organisms may be present (Trombley, 1999, 2001).

Major sources of nutrients in and around the reservation include agricultural activities such as the application of fertilizers to enhance crop production and the pasturing and confined feeding of livestock. Common fertilizers used include, among others, anhydrous ammonia, ammonium nitrate, urea, and mono- and diammonium phosphates. The amount of fertilizer sold in Kansas has increased substantially during the last four decades. In 1950, about 180,000 tons of fertilizer were sold in Kansas, whereas by 2003 sales increased to more than 1,100,000 tons during a 6-month period (January through June 2003) (Kansas Department of Agriculture, 2004). It is likely that this statewide upward trend in fertilizer use also occurred on and near the reservation. Additionally, farm livestock such as cattle and buffalo can produce considerable amounts of nitrogenous waste (urine and manure) that can concentrate in areas where large numbers of animals are pastured or confined. Decomposition of large amounts of fertilizers and manure can release nutrients to surface runoff

Figure 8. Concentrations of dissolved sulfate in *(A)* surface- and *(B)* ground-water-quality samples, November 2003 through August 2006.

Figure 9. Concentrations of dissolved chloride in *(A)* surface- and *(B)* ground-water-quality samples, November 2003 through August 2006.

or to shallow ground water with the potential for discharge to nearby streams (Trombley, 1999, 2001).

Another potential source of nutrients on the reservation is bacterial decomposition of plant and animal protein and seepage from septic systems or sewage lagoons. Nutrients, particularly nitrate and ammonia, also may be components of precipitation (Christensen and Pope, 1997); however, because of dominant agricultural land use in the area, precipitation probably is a relatively minor contributor of nutrients to surface water and ground water on the reservation.

Nitrate is formed by complete oxidation of ammonium ions by microorganisms found in soil, water, sewage, and the digestive tract (U.S. Environmental Protection Agency, 1986). In most oxygenated surface water, nitrate is by far the dominant nitrogen ion due to rapid oxidation of nitrite. Nitrate nitrogen is the form of nitrogen most easily used by most rooted green plants and algae. Nitrate generally occurs in uncontaminated surface water, with a worldwide mean concentration of 0.30 mg/L (Reid and Wood, 1976). Larger nitrate nitrogen concentrations may stimulate growth of rooted plants or accelerate algal production to an extent that may result in taste-or-odor problems in finished drinking water. Because most aquatic organisms can tolerate nitrite plus nitrate concentrations far in excess of what normally might be found even in contaminated surface water, no water-quality criteria have been established for protection of aquatic life. However, USEPA has established recommended criteria on the basis of aggregate Level III nutrient ecoregions as well as subecoregions; the recommended criteria for aggregate ecoregion VI are 0.633 mg/L for nitrite plus nitrate and 1.22 mg/L for total nitrogen in surface water (U.S. Environmental Protection Agency, 2000b). A MCL in drinking water of 10 mg/L nitrate as nitrogen was established by USEPA because of possible toxic effects to infants and is applicable to any supply that may be used for drinking water (U.S. Environmental Protection Agency, 2006).

Nitrite plus nitrate concentrations ranged from an estimated 0.03 to 3.0 mg/L in the surface-water samples collected from the reservation (table 8, fig. 10*A*). The median surface-water concentration was 0.18 mg/L. Nitrite plus nitrate concentrations exceeded the USEPA recommended criterion for ecoregion VI (0.633 mg/L) in 25 percent of surface-water samples. There was some annual variability in nitrite plus nitrate concentrations, with the largest concentrations detected in winter and spring samples and the smallest concentrations in summer and fall. Total nitrogen also was compared to the recommended criterion for aggregate ecoregion VI and exceeded the recommendation (1.22 mg/L) in 25 percent of surface-water samples. Total nitrogen for each site is calculated as the sum of the corresponding ammonia plus organic nitrogen and nitrite plus nitrate concentrations.

Nitrite plus nitrate concentrations in ground-water samples ranged from less than 0.06 to 15 mg/L (table 9, fig. 10*B*), with a median concentration of 0.34 mg/L. The largest concentration was in water from well MW11; the well had similarly large concentrations in all three rounds of testing. The

nitrite plus nitrate concentrations in water from well MW11 exceeded the USEPA MCL. The well is located adjacent to an open hand-dug well that could provide a conduit for contaminants into the ground water in that localized area. Primary land use near the well is pastureland and cropland, and agricultural activities in the area could contribute to elevated nitrite plus nitrate concentrations in water from this well.

Ammonia toxicity is dependent on the pH and temperature of the water; as the pH and temperature increase, the chronic exposure criterion for ammonia decreases (U.S. Environmental Protection Agency, 1999). Acute toxicity of ammonia in fish causes increased respiration, oxygen uptake, and heart rate; reduction in hatching success and growth and morphologic development; and injuries to gills, liver, and kidneys. At larger concentrations, fish may experience convulsions, coma, and death. Small ammonia concentrations can cause reduction in hatching success, reduction in growth rate or morphological development, and pathologic changes in tissues of the gills, livers, and kidneys (U.S. Environmental Protection Agency, 1986). The pH in surface water on the reservation ranged from 6.8 to 8.3 standard units, and the water temperature ranged from 0.1 to 27.2 °C (table 4). For these pH and water temperature conditions, the range of the chronic ammonia criteria for streams that have fish early-life stages present is 0.7 to 6.7 mg/L (U.S. Environmental Protection Agency, 1999). The most likely sources of ammonia on the reservation probably are related to agricultural land use or wastewater.

Dissolved ammonia concentrations in surface-water samples ranged from less than 0.01 to 1.8 mg/L, with a median concentration of less than 0.04 mg/L (table 8). The largest concentration of 1.8 mg/L occurred on February 25, 2004, in a water sample from site LSC06, which is located downstream from a discharging sewage-treatment lagoon. This ammonia concentration was substantially larger than other samples from that site as well as all the other surface-water sites. The chronic exposure criterion for this sample (pH of 6.8 and temperature of $0.3\textdegree$ C) is 6.3 mg/L (U.S. Environmental Protection Agency, 1999).

Ammonia concentrations in ground-water samples ranged from an estimated 0.01 to 1.8 mg/L, with a median concentration of less than 0.04 mg/L (table 9). None of these concentrations exceeded USEPA (1999) chronic exposure criteria for early-life stages of fish, which are not directly applicable to ground water; however, in cases where ground water provides the base flow to streams on the reservation, these criteria could be relevant.

Total phosphorus includes all forms of phosphorus (Hem, 1992). Excessive concentrations of phosphorus in water may contribute to eutrophication of water bodies. Eutrophication (nutrient enrichment) is characterized by excessive nutrient concentrations, decreased dissolved oxygen concentrations, and dense growths of algae (Reid and Wood, 1976). Domestic, municipal, and industrial sewage effluents are substantial sources of phosphorus in surface water because phosphorus is essential in metabolism and is present in human and animal

Table 8. Results of analysis of nutrients in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006. **Table 8.** Results of analysis of nutrients in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.

[mg/L as N, milligrams per year as nitrogen; mg/L, milligrams per literas, partiligrams per liter as phosphorus; MCL, Maximum Contaminant Level; --, not determined or not applicable; <, less than;
E, estimated] [mg/L as N, milligrams per year as nitrogen; mg/L, milligrams per liter; mg/L as P, milligrams per liter as phosphorus; MCL, Maximum Contaminant Level; --, not determined or not applicable; <, less than; E, estimated]

35

Table 8. Results of analysis of nutrients in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006. - Continued **Table 8.** Results of analysis of nutrients in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued

Table 8. Results of analysis of nutrients in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006. - Continued **Table 8.** Results of analysis of nutrients in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued

 $\max_{\Pi} L$ as N, milligrams per year as nitrogen; mg/L, milligrams per liter; mg/L as P, milligrams per liter as phosphorus; MCL, Maximum Contaminant Level; --, not determined or not applicable; <, less than; [mg/L as N, milligrams per year as nitrogen; mg/L, milligrams per liter; mg/L as P, milligrams per liter as phosphorus; MCL, Maximum Contaminant Level; --, not determined or not applicable; <, less than; E, estimated]

2 U.S. Environmental Protection Agency (1986).

Figure 10. Concentrations of dissolved nitrite plus nitrate in *(A)* surface- and *(B)* ground-water-quality samples, November 2003 through August 2006.

[mg/L as N, milligrams per liter as nitrogen; mg/L, milligrams per liter, mg/L as P. milligrams per liter as phosphorus; MCL, Maximum Contaminant Level; --, not determined or not applicable; <, less than;
E, estimated] [mg/L as N, milligrams per liter as nitrogen; mg/L, milligrams per liter; mg/L as P, milligrams per liter as phosphorus; MCL, Maximum Contaminant Level; --, not determined or not applicable; <, less than; E, estimated]

Table 9. Results of analysis of nutrients in ground-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and August 2006. -- Continued **Table 9.** Results of analysis of nutrients in ground-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and August 2006.—Continued

metabolic waste. Potential sources for large concentrations of phosphorus in streams on the Prairie Band Potawatomi Reservation include human or animal waste and fertilizers applied to agricultural lands.

USEPA established a goal for total phosphorus concentration of 0.10 mg/L to limit cultural eutrophication of flowing water (U.S. Environmental Protection Agency, 1986). Larger concentrations also may interfere with coagulation in watertreatment plants. USEPA also has recommended water-quality criteria for total phosphorus on the basis of aggregate nutrient ecoregions. A total phosphorus criterion of 0.076 mg/L has been established by USEPA for ecoregion VI (U.S. Environmental Protection Agency, 2000b). However, previous reports compared stream concentrations of total phosphorus to the recommended goal of 0.10 mg/L (Trombley, 1999, 2001; Schmidt, 2004). Therefore, this criterion also will be used for total phosphorus evaluation in this report. Total phosphorus was identified as exceeding water-quality criteria on the reservation in three previous reports published as part of this cooperative study (Trombley, 1999, 2001; Schmidt, 2004).

Concentrations of total phosphorus in surface-water samples ranged from an estimated 0.02 to 3.5 mg/L, with a median concentration of 0.13 mg/L (table 8, fig. 11). All 11 samples from site LSC06 exceeded the water-quality goal of 0.10 mg/L; however, the largest concentration of 3.5 mg/L occurred in a sample from site LSC04 on May 11, 2005, during runoff conditions (streamflow of 189 ft³/s, table 4). Concentrations of other nutrients also were large in the May 11, 2005, sample from site LSC04. Overall, 56 percent

of the 55 surface-water samples that were analyzed for total phosphorus exceeded the goal of 0.10 mg/L to prevent cultural eutrophication (U.S. Environmental Protection Agency, 1986).

Total phosphorus concentrations in ground-water samples ranged from an estimated 0.02 to 0.62 mg/L, with a median concentration of 0.07 mg/L (table 9). Although the cultural eutrophication goal does not apply to ground water, it should be noted for comparison purposes that 33 percent of groundwater samples exceeded the 0.10-mg/L goal.

Trace Elements

Trace elements are found in small quantities (less than 1 percent) in or adsorbed to mineral surfaces (Bates and Jackson, 1980). Trace elements generally are found in the environment at relatively small concentrations (less than 1.0 mg/L), but some are vital for human and plant health. Concentrations of trace elements in surface and ground water can be increased as a result of human activities such as mining and industrial activities. Water samples for this study were analyzed for the trace elements arsenic, boron, iron, and selenium; all of which have applicable water-quality criteria.

Arsenic is a trace element that occurs naturally in the environment, but the presence of arsenic in water also can be a result of mining activity or application of pesticides. Arsenic also can occur in water as a result of an arsenic-bearing formation dissolving naturally over time (Haack and Rachol, 2000). Large concentrations of arsenic have shown to be detrimental to human health (Hem, 1992). In 2001, USEPA revised the

Figure 11. Concentrations of total phosphorus in surface-water-quality samples, November 2003 through August 2006.

drinking-water MCL for arsenic from 50 to 10 µg/L, which became effective in 2006 (U.S. Environmental Protection Agency, 2006).

Arsenic concentrations in the surface-water samples collected during the current study period ranged from an estimated 0.9 to 7.8 μ g/L (table 10) and were less than the drinking water MCL for arsenic. The median arsenic concentration in surface-water samples was less than 1.9 µg/L.

Ground-water arsenic concentrations ranged from 0.2 to 30 µg/L (table 11, fig. 12), with a median concentration of less than 1.9 µg/L. The largest concentration came from well MW06 and was three times the MCL for arsenic. The other two concentrations from this well were about twice the MCL. The reason for the unusually large concentrations of arsenic in water from well MW06 currently (2007) is not known. Arsenic was detected in water from the nine other monitoring wells but at concentrations substantially less than the MCL.

Boron is a minor constituent in ground water and usually is found in evaporate deposits as sodium or calcium borate salt (Hem, 1992). It is an essential element for the growth of plants, but there is no evidence that it is required by animals (Pais and Jones, 1997). Total boron essentially is conserved during wastewater treatment and may be used to determine whether a water sample contains wastewater effluent (Quanrud and others, 2004). USEPA established a lifetime DWA of 600 µg/L (concentration in drinking water that is not expected to cause any adverse effects for a lifetime of exposure) (U.S. Environmental Protection Agency, 2006).

Boron concentrations in surface-water samples collected during the current study period ranged from 30 to 80 µg/L, with a median concentration of 50 µg/L (table 10). The largest detection of 80 µg/L occurred on November 19, 2003, in a water sample from site LSC06 and may be attributable to the upstream sewage-treatment lagoon.

Concentrations of boron in ground-water samples ranged from 20 to 880 µg/L, with a median concentration of 70 µg/L (table 11). The largest concentration was in a sample from well MW03 collected on July 26, 2004, and corresponds to the large concentrations of calcium, magnesium, and sodium and several anions contributing to the large dissolved-solids concentration in water from this well (table 7).

Iron is an abundant element in the Earth's outer crust, but it typically occurs in small concentrations in water (Hem, 1992). Iron's chemical activity and its solubility in water depend largely on the oxygen in the system as well as the pH of the water (Hem, 1992). Large amounts of iron are undesirable in water supplies because it forms a red oxyhydroxide precipitate that stains laundry and plumbing fixtures (Hem, 1992). USEPA (2006) has established a SDWR of 300 µg/L for iron in treated drinking water.

Iron concentrations in surface-water samples ranged from an estimated 3 to 70 µg/L, with a median concentration of 10 µg/L (table 10). All concentrations in surface-water samples were substantially less than the USEPA SDWR.

Iron concentrations in ground-water samples ranged from an estimated 2 to 5,700 µg/L, with a median concentration of 45 µg/L (table 11, fig. 13). The three largest iron

Table 10. Results of analysis of trace elements in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.

[µg/L, micrograms per liter; MCL, Maximum Contaminant Level; DWA, Drinking-Water Advisory; SDWR, Secondary Drinking-Water Regulation; --, not determined or not applicable; E, estimated; <, less than]

Table 10. Results of analysis of trace elements in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued

[µg/L, micrograms per liter; MCL, Maximum Contaminant Level; DWA, Drinking-Water Advisory; SDWR, Secondary Drinking-Water Regulation; --, not determined or not applicable; E, estimated; <, less than]

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Table 10. Results of analysis of trace elements in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued

[µg/L, micrograms per liter; MCL, Maximum Contaminant Level; DWA, Drinking-Water Advisory; SDWR, Secondary Drinking-Water Regulation; --, not determined or not applicable; E, estimated; <, less than]

¹Water-quality criteria established by the U.S. Environmental Protection Agency (2006).

Table 11. Results of analysis of trace elements in ground-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and August 2006.

[µg/L, micrograms per liter; MCL, Maximum Contaminant Level; DWA, Drinking-Water Advisory; SDWR, Secondary Drinking-Water Regulation; E, estimated; <, less than]

Monitoring-site identifier (fig. 1)	Site name	Date of sample collection (month/day/ year)	Arsenic, dissolved $(\mu g/L)$	Boron, dissolved $(\mu g/L)$	Iron, dissolved $(\mu g/L)$	Selenium, dissolved $(\mu g/L)$
	Water-quality criteria ¹		$10\,$	600	300	50
			(MCL)	(DWA)	(SDWR)	(MCL)
MW01	Monitoring well #1	7/28/2004	E1.4	$70\,$	1,000	2.6
		7/11/2005	E1.4	70	1,100	2.6
		8/16/2006	1.4	60	880	< 1
MW02	Monitoring well #2	7/27/2004	<1.9	110	30	< 2.6
		7/11/2005	< 1.9	100	60	< 2.6
		8/15/2006	\cdot	90	30	$\mathop{\hbox{\rm E}}.1$
MW03	Monitoring well #3	7/26/2004	< 1.9	880	3,600	< 2.6
		7/12/2005	1.9	870	3,400	2.6
		8/15/2006	1.6	860	3,600	E.1
M W04	Monitoring well #4	7/29/2004	< 1.9	60	<6	2.6
		7/14/2005	E1.3	50	<6	2.6
		8/16/2006	\cdot .2	50	<6	.6
MW06	Monitoring well #6	7/29/2004	19	100	4,100	< 2.6
		7/13/2005	$22\,$	100	3,800	< 2.6
		8/14/2006	$30\,$	90 	5,700	< 1
MW07	Monitoring well #7	7/27/2004	< 1.9	90	40	< 2.6
		7/13/2005	< 1.9	$80\,$	50	2.6
		8/15/2006	$.5\,$	$70\,$	50	.3
MW08	Monitoring well #8	7/27/2004	<1.9	100	120	< 2.6
		7/12/2005	1.9	90 	340	2.6
		8/14/2006	1.2	70	330	$.2\,$
MW09	Monitoring well #9	7/29/2004	< 1.9	30	E ₆	< 2.6
		7/14/2005	< 1.9	$30\,$	<6	< 2.6
		8/17/2006	\cdot .2	$20\,$	$<\!\!6$.5
MW10	Monitoring well #10	7/28/2004	< 1.9	40	$\mathrm{E}4$	E1.3
		7/14/2005	< 1.9	$40\,$	E2	E1.7
		8/16/2006	$.3\,$	30	<6	1.9
MW11	Monitoring well #11	7/26/2004	E1.3	60	E4	2.6
		7/14/2005	< 1.9	$50\,$	<6	< 2.6
		8/18/2006	1.5	60	<6	$1.0\,$

¹Water-quality criteria established by the U.S. Environmental Protection Agency (2006).

concentrations were detected in samples from well MW06, which corresponds to larger arsenic concentrations. Iron concentrations in water from wells MW01, MW03, MW06, and MW08 (two samples) exceeded the USEPA SDWR of 300 µg/L. Iron concentrations in samples from the remainder of the wells were substantially less than the SDWR.

Selenium is a naturally occurring mineral element that is relatively rare, and geochemical factors tend to limit its aqueous mobility (Hem, 1992). Selenium was detected in extremely small concentrations, less than the reporting level in most surface- and ground-water samples (tables 10 and 11). The largest concentration was 4.1 µg/L, detected in a surface-water sample from site SC07 on August 25, 2004. The MCL for selenium is 50 µg/L (U.S. Environmental Protection Agency, 2006).

Pesticides

The use of pesticides on the reservation is characteristic of pesticide use in all agricultural regions throughout Kansas. Atrazine, one of the triazine herbicides, has been used since the 1950s in the production of corn and grain sorghum in the area. Another potential source of atrazine may be its use in controlling weeds along railroad rights-of-way and along roads and highways. It is the most frequently detected herbicide in Kansas (Stamer and Zelt, 1994).

Triazine herbicides were analyzed in both surface- and ground-water samples. Triazine herbicides and their degradates are listed on the USEPA Contaminant Candidate List (U.S. Environmental Protection Agency, 2006). The herbicides are applied pre- and post-emergence and are subject to runoff if applied before large rainfalls. Fifty-five surface-water samples collected during November 2003 through August 2006 were analyzed using the ELISA (enzyme-linked immunosorbent assay) triazine herbicide screen. The triazine herbicide screen ELISA is a reliable indicator of atrazine (fig. 14). Christensen and Ziegler (1998) found that at least 80 percent of the triazine herbicide concentration determined by ELISA is atrazine.

During June 1996 through August 2006, 282 surfacewater samples collected from the reservation were analyzed for triazine herbicides by ELISA; 78 of these had atrazine concentrations confirmed by GC/MS (gas chromatography/ mass spectrometry) analysis (see previous "Methods" section). Data from June 1996 through November 1998 are presented in Trombley (1999, table 5); data from February 1999 through February 2001 are presented in Trombley (2001, table 11); and data from May 2001 through August 2003 are presented in Schmidt (2004, table 12).

The relation between the triazine herbicide screen and atrazine is presented in figure 14; the correlation between results obtained from ELISA analysis and GC/MS analysis is apparent and has a coefficient of determination (R^2) of 0.91. Therefore, results from the triazine screen can be compared to the drinking-water MCL of 3.0 µg/L as an annual average established by USEPA for atrazine, with the understanding that, by regression analysis, 86 percent (slope of the regression) of the ELISA triazine screen concentration is atrazine.

Of 55 surface-water samples collected from November 2003 through August 2006 and analyzed by ELISA, concentrations of triazine herbicides ranged from less than 0.1 to 3.4 µg/L, with a median concentration of 0.2 µg/L (table 12, fig. 15). Triazine herbicides were detected in one surfacewater sample at a concentration greater than the MCL of

Figure 13. Concentrations of dissolved iron in ground-water-quality samples, July 2004, July 2005, and August 2006.

Figure 14. Relation between triazine herbicide concentrations determined by enzyme-linked immunosorbent assay (ELISA) and atrazine concentrations determined by gas chromatography/mass spectrometry (GC/MS) for

Table 12. Results of analysis of pesticides in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006. **Table 12.** Results of analysis of pesticides in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.

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Table 12. Results of analysis of pesticides in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued **Table 12.** Results of analysis of pesticides in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued

[ELISA, enzyme-linked immunosorbent assay; µg/L, micrograms per liter; --, not determined or not applicable; <, less than; E, estimated; MCL, Maximum Contaminant Level] [ELISA, enzyme-linked immunosorbent assay; µg/L, micrograms per liter; --, not determined or not applicable; <, less than; E, estimated; MCL, Maximum Contaminant Level]

Table 12. Results of analysis of pesticides in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006. —Continued **Table 12.** Results of analysis of pesticides in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued

Table 12. Results of analysis of pesticides in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued **Table 12.** Results of analysis of pesticides in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued

[ELISA, enzyme-linked immunosorbent assay; µg/L, micrograms per liter; --, not determined or not applicable; <, less than; E, estimated; MCL, Maximum Contaminant Level] [ELISA, enzyme-linked immunosorbent assay; µg/L, micrograms per liter; --, not determined or not applicable; <, less than; E, estimated; MCL, Maximum Contaminant Level] \overline{a}

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Table 12. Results of analysis of pesticides in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued **Table 12.** Results of analysis of pesticides in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued

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Figure 15. Concentrations of triazine herbicide screen determined using enyme-linked immunosorbent assay (ELISA) in surface-water-quality samples, November 2003 through August 2006.

3.0 µg/L for atrazine. That sample was collected from a site on Soldier Creek (site SC07) on February 25, 2004. However, the actual atrazine concentration in this sample probably was less than the MCL when the slope of the regression (fig. 14) is considered. Also, the 3.4 μ g/L concentration was not verified as atrazine as a result of GC/MS analysis of this sample, which determined an atrazine concentration of 2.7 µg/L (table 12).

Four surface-water samples collected between February 2004 and May 2004 were analyzed for specific triazine herbicides and other pesticides by GC/MS (table 12). Four of the triazine herbicides and other pesticides analyzed were detected in these surface-water samples. Atrazine was detected in the four surface-water samples analyzed by GC/MS at concentrations ranging from 0.79 to 2.7 µg/L, with a median concentration of 2.4 µg/L. All four surface-water samples also had detections of metolachlor, a general-use pesticide, with concentrations that ranged from 0.05 to 0.71 µg/L.

Atrazine, one atrazine metabolite (2-chloro-4-isopropylamino-6-amino-s-triazine), and (or) cyanazine amide were detected in the ground-water samples collected August 16–18, 2006, from monitoring wells MW04, MW10, and MW11 (table 13). There were no other detections of pesticides in water from these wells in previous samples during the current study period. All detected concentrations were small relative to laboratory reporting levels and established water-quality criteria.

Chemical Oxygen Demand and Volatile Organic Compounds

The chemical oxygen demand (COD) of a water sample is a measure of the amount of oxygen needed to oxidize the reactive chemicals in a water system (Hem, 1992). If the COD of a sample is high as that water moves through a system, it will consume dissolved oxygen from the surrounding water to complete chemical reactions. Decreases in dissolved oxygen in a system can destroy aerobic organisms and cause changes in the redox environment of a system, which could release redox sensitive chemicals such as arsenic and iron into the water. COD also is used as a quantitative measure of organic pollution in a water sample (Hem, 1992). The COD in the water sample from well MW04 was 10 mg/L (equivalent to the laboratory reporting level) for the sample collected on July 14, 2005 (table 14). In a sample collected a year later from the same well, the COD was less than 10 mg/L. As this water moves through the ground-water system, it probably will not consume a substantial amount of dissolved oxygen in the aquifer downgradient from the well.

Volatile organic compounds (VOCs) are compounds that have a high vapor pressure and low water solubility. Many VOCs are synthetic chemicals that are used and produced in the manufacture of paints, pharmaceuticals, and refrigerants. VOCs typically are industrial solvents, fuel oxygenates, or byproducts produced by chlorination in water treatment. VOCs often are components of petroleum fuels, hydraulic fluids,

[µg/L, micrograms per liter; --, not determined or not applicable; <, less than; MCL, Maximum Contaminant Level]

Table 13. Results of analysis of pesticides in ground-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and August **Table 13.** Results of analysis of pesticides in ground-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and August 2006—Continued 2006—Continued

micrograms per liter; --, not determined or not applicable; <, less than; MCL, Maximum Contaminant Level] [µg/L, micrograms per liter; --, not determined or not applicable; <, less than; MCL, Maximum Contaminant Level] $\lbrack \mu g/L,$

 \overline{a} [µg/L, micrograms per liter; --, not determined or not applicable; <, less than; MCL, Maximum Contaminant Level] $\overline{}$ ÷ ntar C_o \cdot M_{2} · MCT tha $\frac{d}{\sqrt{2}}$ ahle: ric $\ddot{ }$ ्र ÷ not deter liter ric

Table 14. Results of analysis of chemical oxygen demand and volatile organic compounds in ground-water-quality samples from monitoring wells MW04 and MW07 on Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and August 2006.

[[]mg/L, milligrams per liter; µg/L, micrograms per liter; <, less than; E, estimated; --, not determined]

	Concentration for indicate sample date					
Constituent and unit of measurement	Monitoring well MW04		Monitoring well MW07			
	7/29/04	7/14/05	8/16/06	7/27/04		
Chemical oxygen demand, total, mg/L	$\overline{}$	10	<10	$\overline{}$		
1,1,1,2-Tetrachloroethane, total, µg/L	< 0.03	< .03	< 0.03	< 0.03		
1,1,1-Trichloroethane, total, µg/L	< .032	< .032	< .032	< .032		
1,1,2,2-Tetrachloroethane, total, µg/L	< 16	&0.08	&0.08	16		
1,1,2-Trichloro-1,2,2-trifluoroethane, total, µg/L	< .04	< 0.04	< .04	< .04		
1,1,2-Trichloroethane, total, µg/L	< .06	< 0.04	< .04	< .06		
1,1-Dichloroethane, total, µg/L	< .035	< .035	< .035	< .035		
1,1-Dichloroethene, total, µg/L	$-.02$	< .02	< .02	< .02		
1,1-Dichloropropene, total, µg/L	< .03	< .03	< .03	< .03		
1,2,3,4-Tetramethylbenzene, total, µg/L	< 1	< 1	< 1	< 1		
1,2,3,5-Tetramethylbenzene, total, µg/L	< 1	< 1	$\langle .2 \rangle$	\lt .1		
1,2,3-Trichlorobenzene, total, µg/L	< .3	$\langle .2 \rangle$	$\langle .2 \rangle$	< .3		
1,2,3-Trichloropropane, total, µg/L	< 18	< 18	< 18	< 18		
1,2,3-Trimethylbenzene, total, µg/L	< 1	< 1	< 1	< 1		
1,2,4-Trichlorobenzene, total, µg/L	< 1	< 1	< 1	< 1		
1,2,4-Trimethylbenzene, total, µg/L	< .056	< .056	< .056	< .056		
1,2-Dibromo-3-chloropropane, total, µg/L	< .5	< .5	< .5	< .5		
1,2-Dibromoethane, total, µg/L	< .036	< .036	< .036	< .036		
1,2-Dichlorobenzene, total, µg/L	< .05	< .05	< .05	< .05		
1,2-Dichloroethane, total, µg/L	< 1	< 1	< 1	< 1		
1,2-Dichloropropane, total, µg/L	< .029	< .029	< .029	< .029		
1,3,5-Trimethylbenzene, total, µg/L	$-.044$	$-.044$	$-.044$	$-.044$		
1,3-Dichlorobenzene, total, µg/L	< .03	< .03	< .03	< .03		
1,3-Dichloropropane, total, µg/L	< 12	12	12	$-.12$		
1,4-Dichlorobenzene, total, µg/L	< .03	< .03	< .03	< .03		
2,2-Dichloropropane, total, µg/L	< .05	< .05	< 0.05	< .05		
2-Chlorotoluene, total, µg/L	$-.04$	< 04	< .04	< .04		
2-Ethyltoluene, total, µg/L	< .06	< .06	<06	< .06		
3-Chloropropene, total, µg/L	< .50	< 50	< .50	< .50		
4-Chlorotoluene, total, µg/L	$-.05$	< .05	< .05	< .05		

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Table 14. Results of analysis of chemical oxygen demand and volatile organic compounds in ground-water-quality samples from monitoring wells MW04 and MW07 on Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and August 2006.—Continued

[mg/L, milligrams per liter; µg/L, micrograms per liter; <, less than; E, estimated; --, not determined]

Table 14. Results of analysis of chemical oxygen demand and volatile organic compounds in ground-water-quality samples from monitoring wells MW04 and MW07 on Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and August 2006.—Continued

[[]mg/L, milligrams per liter; µg/L, micrograms per liter; <, less than; E, estimated; --, not determined]

	Concentration for indicate sample date					
Constituent and unit of measurement	Monitoring well MW04	Monitoring well MW07				
	7/29/04	7/14/05	8/16/06	7/27/04		
Methyl acrylate, total, µg/L	2.0	< 1.0	< 1.0	2.0		
Methyl methacrylate, total, µg/L	< .3	\leq .2	$\langle .2 \rangle$	< .3		
Methyl tert-pentyl ether, total, µg/L	&0.08	< 0.04	< 0.04	&0.08		
m-Xylene plus p-xylene, total, µg/L	< .06	< .06	< .06	< .06		
Naphthalene, total, µg/L	< .5	< .5	< .5	< .5		
n-Butyl methyl ketone, total, µg/L	ϵ .7	< .4	< .4	< 7		
n-Butylbenzene, total, µg/L	< 1	< 1	< 1	< 1		
n-Propylbenzene, total, µg/L	< .042	< .042	< .042	< .042		
o-Xylene, total, µg/L	${-.04}$	< 0.04	< .04	< .04		
sec-Butylbenzene, total, µg/L	< .06	< 06	< .06	< .06		
Styrene, total, μ g/L	< .042	< 042	< 0.042	< 0.042		
tert-Butyl ethyl ether, total, µg/L	< .05	< .03	< .03	< .05		
Methyl tert-butyl ether, total, µg/L	\leq .2	< 1	$\lt.1$	\leq .2		
tert-Butylbenzene, total, µg/L	< .06	< 0.06	&0.08	< .06		
Tetrachloroethene, total, µg/L	< .06	< .03	< .03	< .06		
Tetrachloromethane, total, µg/L	< .06	< 06	< .06	< .06		
Tetrahydrofuran, total, µg/L	\leq 2	<1	<1	\leq 2		
Toluene, total, µg/L	< .05	< .02	$-.02$	< .05		
trans-1,2-Dichloroethene, total, µg/L	< .032	< .032	< .032	< .032		
trans-1,3-Dichloropropene, total, µg/L	< .09	< .09	< .09	< .09		
trans-1,4-Dichloro-2-butene, total, μ g/L	ϵ .7	< 7	ϵ .7	< 7		
Tribromomethane, total, µg/L	10	10	10	10		
Trichloroethene, total, µg/L	< .038	< .038	< .038	< .038		
Trichlorofluoromethane, total, µg/L	16	&0.5	&0.5	16		
Trichloromethane, total, µg/L	< .024	< .024	< .024	< .024		
Vinyl chloride, total, µg/L	< 11	< 11	< 11	< 11		

paint thinners, and dry cleaning agents (Daly and Lindsey, 1996).

During the current study period, all three samples from well MW04 and the July 27, 2004, water sample from well MW07 were analyzed for the presence of VOCs. The samples from well MW04 were analyzed because of its proximity to a construction and demolition landfill (fig. 1); the sample from well MW07 was analyzed as a reference site for comparison to results from well MW04. During the previous study period (Schmidt, 2004, table 14), carbon disulfide, ethyl methyl ketone, and tetrahydrofuran were detected in a water sample from well MW04. Of the 85 VOCs analyzed in water samples from the two wells for the current study period, only one compound was detected—diethyl ether, found in the sample from well MW07 at an estimated concentration of 0.1 µg/L (table 14). Diethyl ether is used as a solvent, general anesthetic, and as a starting fluid for diesel and gasoline engines. No drinking-water criteria have been established for exposure to diethyl ether.

Bacteria

Fecal coliform, *Escherichia coli (E. coli)*, fecal streptococcus, and enterococci bacteria are indigenous to the intestinal tract of warmblooded animals (Holt and others, 1993). The fecal coliform bacteria group can include any combination of *E. coli* and species of the *Klebsiella, Enterobacter*, and *Citrobacter* genera (Gleeson and Gray, 1997). The presence of large concentrations of these organisms in surface water indicates fecal contamination and also may indicate the presence of disease-causing organisms. Potential sources of these bacteria on the reservation include sewage lagoons, runoff, and seepage from livestock areas, such as pastures and confined feedlots, and from wildlife populations.

USEPA has established water-quality criteria on the basis of full-body contact and single-sample allowable density for water bodies defined as moderately used, lightly used, and infrequently used (U.S. Environmental Protection Agency, 2002b). Primary contact use is defined as activities where the body is immersed in water to the extent that some inadvertent ingestion of the water is probable. These activities include boating, mussel harvesting, swimming, skin diving, water skiing, and wind surfing (Kansas Department of Health and Environment, 2001). These criteria are presented in table 15. USEPA currently (2007) has no recommended criteria for secondary contact (U.S. Environmental Protection Agency, 2002b); however, the Kansas Department of Health and Environment has set maximums for both primary and secondary contact recreation in public-access streams for the State of Kansas (Kansas Department of Health and Environment, 2004). Secondary contact activities include wading, fishing, trapping, and hunting where the ingestion of water is not probable (Kansas Department of Health and Environment, 2001). Tribal members primarily use the streams on the reservation

for hunting and fishing; therefore, ingestion of water is not probable during these activities.

Fecal coliform bacteria densities ranged from less than 3 to 32,000 col/100 mL in surface-water samples from the reservation (table 16). The largest density of fecal coliform colonies was in a sample collected on May 11, 2005, from site LSC04. The median density of fecal coliform bacteria in all samples from the surface-water sites was 230 col/100 mL. Fecal coliform bacteria were detected six times in groundwater samples from monitoring wells (table 17). The two largest densities were 34 col/100 mL from monitoring well MW02 on July 11, 2005, and an estimated 10 col/100 mL from monitoring well MW11 on August 18, 2006.

E. coli is the dominant bacteria of the fecal coliform group (Rasmussen and Ziegler, 2003). Rasmussen and Ziegler (2003) found that the ratio of *E. coli* to fecal coliform in the Soldier Creek Basin is 0.73, meaning that out of every 100 colonies of fecal coliform in a surface-water sample, 73 of the colonies are *E. coli* bacteria. This relation is important when comparing historic fecal coliform bacteria densities to the current (2007) USEPA bacteria water-quality criteria, which are based on *E. coli* densities.

The densities of *E. coli* colonies in surface-water samples ranged from less than 3 to 31,000 col/100 mL, with a median density of 170 col/100 mL (table 16). The largest density of *E. coli* colonies was detected in the sample collected from site LSC04 on May 11, 2005. Of 55 surface-water samples analyzed for *E. coli* bacteria, 22 samples (40 percent) exceeded the criterion established by the U.S. Environmental Protection Agency (2002b) for primary contact (illness rate of 8 per 1,000 swimmers, table 15) in moderately used water bodies (298 col/100 mL), 18 samples (33 percent) exceeded the criterion for lightly used water bodies (406 col/100 mL), and 16 samples (29 percent) exceeded the criterion for infrequently used water bodies (576 col/100 mL). In addition, 11 of 27 samples (41 percent) exceeded the KDHE (Kansas Department of Health and Environment, 2004) single-sample maximum for primary contact in public access bodies of water (1,198 col/100 mL for samples collected between April 1 and October 31), and 4 of 55 samples (7 percent) exceeded the KDHE single-sample maximum for secondary contact in public access bodies of water (9,760 col/100 mL; applicable year round).

E. coli was detected in eight samples (27 percent) of ground water (table 17). Most were at small estimated densities. The largest density of 23 col/100 mL was detected in a sample from well MW02 on July 11, 2005. This corresponds to the largest detected density of fecal coliform bacteria.

Fecal streptococci bacteria densities in surface-water samples ranged from 130 to 140,000 col/100 mL, with a median density of 1,100 col/100 mL (table 16). The largest density occurred in the May 19, 2004, sample from site LSC06. All 13 samples analyzed during the current study period had detectable densities of fecal streptococci bacteria. Ground-water samples were not analyzed for fecal streptococci bacteria.

Table 15. U.S. Environmental Protection Agency (2002b) recommended water-quality criteria for fecal-indicator bacteria.

[All values are in colonies per 100 milliliters of water]

1 Recreation during which the body is immersed in surface water to the extent that some inadvertent ingestion of water is probable.

Enterococci bacteria are a subgroup of the fecal streptococci bacteria and are a valuable bacterial indicator for determining the extent of fecal contamination of recreational surface water. Studies at marine and freshwater bathing beaches indicated that swimming-associated gastroenteritis is related directly to the quality of the bathing water, and enterococci are the most efficient bacterial indicator of water quality (American Public Health Association and others, 1992). USEPA has established water-quality criteria (table 15) that are based on enterococci bacteria densities in recreational water (U.S. Environmental Protection Agency, 2002b).

Enterococci bacteria densities in surface-water samples from the reservation ranged from an estimated 8 to 290,000 col/100 mL, with a median density of 1,500 col/100 mL (table 16). The largest density was detected in a sample from site LSC04 on May 11, 2005, and corresponds with the largest densities of fecal coliform and *E. coli* bacteria. USEPA criteria that are based on an illness rate of 8 per 1,000 swimmers and the single-sample maximum allowable density were used for comparison purposes (table 15). Of 55 surface-water samples analyzed for enterococci bacteria, 50 samples (91 percent) exceeded the criteria for primary

contact in both moderately used water bodies and lightly used water bodies (78 and 107 col/100 mL, respectively), and 48 samples (87 percent) exceeded the primary contact criterion for infrequently used water bodies (151 col/100 mL).

Enterococci bacteria were the most frequently detected bacteria in ground-water samples (table 17). Enterococci were detected in 16 of 30 ground-water samples (53 percent), and densities in all 30 samples ranged from less than 1 to an estimated 230 col/100 mL. The largest density of enterococci bacteria was estimated in a water sample from monitoring well MW02 on August 15, 2006.

Figure 16 shows median densities of fecal-indicator bacteria in surface-water samples during spring, summer, fall, and winter as well as median streamflow measured at the time of sample collection. Only quarterly samples were used in comparison, with the May and June samples defined as "spring," August samples defined as "summer," November and December samples defined as "fall," and February samples defined as "winter." Median densities of all indicator bacteria were largest during the summer months, except for fecal streptococci, which was not measured in any summer samples for the current study period because analyses of samples for fecal
Table 16. Results of analysis of fecal-indicator bacteria in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.

[Other water-quality criteria listed in table 15. col/100 mL, colonies per 100 milliliter of water; --, not determined or not applicable; E, estimated; <, less than]

Table 16. Results of analysis of fecal-indicator bacteria in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued

[Other water-quality criteria listed in table 15. col/100 mL, colonies per 100 milliliter of water; --, not determined or not applicable; E, estimated; <, less than]

1 Single-sample maximum set by the Kansas Department of Health and Environment (2004) for secondary contact recreation in public-access bodies of water.

Table 17. Results of analysis of fecal-indicator bacteria in ground-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, July 2004, July 2005, and August 2006.

Monitoring-site identifier (fig. 1)	Site name	Date of sample collection (month/ day/year)	Fecal-indicator bacteria		
			Fecal coliform (col/100 mL)	Escherichia coli (col/100 mL)	Enterococci (col/100 mL)
MW01	Monitoring well #1	7/28/2004	<1	<1	<1
		7/11/2005	$<\!\!1$	$<\!1$	<1
		8/16/2006	$\rm E1$	E1	E10
MW02	Monitoring well #2	7/27/2004	<1	<1	E16
		7/11/2005	34	23	E150
		8/15/2006	E3	$\rm E7$	E230
MW03	Monitoring well #3	7/26/2004	<1	<1	$<\!\!1$
		7/12/2005	<1	<1	<1
		8/15/2006	$<\!\!1$	$<\!1$	<1
MW04	Monitoring well #4	7/29/2004	<1	<1	43
		7/14/2005	<3	E ₃	$<3\,$
		8/16/2006	$<\!\!1$	<1	$<\!\!1$
MW06	Monitoring well #6	7/29/2004	<1	<1	$\mathrm{E}4$
		7/13/2005	<1	<1	E1
		8/14/2006	<1	<1	E1
MW07	Monitoring well #7	7/27/2004	E1	<1	<1
		7/13/2005	$<\!\!1$	$\rm E1$	<1
		8/15/2006	<1	<1	${<}1\,$
MW08	Monitoring well #8	7/27/2004	<1	<1	E3
		7/12/2005	<1	<1	$\rm E1$
		8/14/2006	<1	<1	$<\!\!1$
MW09	Monitoring well #9	7/29/2004	<1	<1	<1
		7/14/2005	$<\!1$	$<\!1$	$\rm E1$
		8/17/2006	<3	$<$ 3	$<$ 3
MW10	Monitoring well #10	7/28/2004	$<\!\!1$	$<\!\!1$	$\rm E2$
		7/14/2005	$<\!\!1$	$\rm E1$	$\rm E1$
		8/16/2006	$<\!1$	$<\!1$	$<\!1$
MW11	Monitoring well #11	7/26/2004	$<3\,$	$<3\,$	E42
		7/14/2005	$\rm E8$	$\rm E1$	140
		8/18/2006	$\rm E10$	E2	$\rm E27$

[Water-quality criteria listed in table 15. col/100 mL, colonies per 100 milliliter of water; E, estimated; <, less than]

Figure 16. Seasonal distribution of median streamflow and median densities of fecal-indicator bacteria in surfacewater samples, November 2003 through August 2006.

streptococci bacteria were discontinued after the spring 2004 sampling period. The bacteria densities were not necessarily larger during times of larger median streamflow.

Sediment

Fluvial sediment is defined as fragmentary material that originates mostly from weathering rocks and is transported by, suspended in, or deposited from water; it includes chemical and biological precipitates and decomposed organic material, such as humus (Federal Inter-Agency Sedimentation Project, 1963).

Fluvial sediment in surface-water samples collected for this study was measured by two different analytical methods suspended-sediment concentration (SSC) and total suspended solids (TSS). SSC data were collected by measuring the dry weight of all the sediment from a known volume of a watersediment mixture (Gray and others, 2000). TSS data were determined by measuring the dry weight of sediment from a known volume of a subsample of the original sample (Gray and others, 2000).

SSCs in surface-water samples ranged from 3 to 6,500 mg/L, with a median concentration of 47 mg/L (table 18). The largest SSC was detected in a sample from site LSC04 collected on May 11, 2005, and corresponds to the largest TSS concentration. Suspended sediment was measured in all of the surface-water samples.

Concentrations of TSS were analyzed in all surface-water samples collected during this study and ranged from less than 10 to an estimated 5,200 mg/L, with a median concentration of 16 mg/L (table 18). The largest concentration occurred in a water sample collected on May 11, 2005, from site LSC04 and corresponds to the largest densities of fecal coliform bacteria, *E. coli*, and enterococci bacteria detected during the current study period. TSS concentrations were greater than 10 mg/L in 62 percent of surface-water samples.

The median concentration of TSS in the 24 samples from the two Soldier Creek sites (SC03 and SC07, fig. 1) was larger than the median concentration in the 20 samples from two of the three Little Soldier Creek sites (LSC03 and LSC04, fig. 1), 28 and 18 mg/L, respectively. Median concentrations of suspended sediment were larger in samples collected from the two Little Soldier Creek sites than the two Soldier Creek sites (73 and 56 mg/L, respectively). Median concentrations were small relative to the potential range in concentrations (table 18, fig. 17) and probably are a reflection of the low streamflow conditions under which most surface-waterquality samples were collected during the current study period. Concentrations of SSC and TSS were much smaller in samples from the monitoring site on Big Elm Creek (LSC06, fig. 1), 8 and 10 mg/L respectively, which probably is due to the small amount of sediment discharge from the upstream wastewatertreatment plant (Schmidt, 2004).

Table 18. Results of analysis of suspended sediment and total suspended solids in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.

[mg/L, milligrams per liter; --, not determined; <, less than; E, estimated]

Table 18. Results of analysis of suspended sediment and total suspended solids in surface-water-quality samples from Prairie Band Potawatomi Reservation, northeastern Kansas, November 2003 through August 2006.—Continued

[mg/L, milligrams per liter; --, not determined; <, less than; E, estimated]

Surface-water-quality monitoring site and basin (fig. 1)

Figure 17. Median concentrations of suspended sediment and total suspended solids in surface-water-quality samples, November 2003 through August 2006.

Comparison of Water-Quality Results to Previous Study Periods

This report describes surface- and ground-water quality on the Prairie Band Potawatomi Reservation from November 2003 through August 2006, and a comparison of results from this study period to results in previous water-quality reports published as a part of this ongoing study is presented in this section. During previous study periods, surface-water samples were collected for analysis of physical properties, nutrients, pesticides, and fecal-indicator bacteria (Trombley, 1999, 2001). Results from these study periods found total phosphorus, triazine herbicides, and fecal-indicator bacteria exceeding their respective water-quality criteria. Previous study-period assessments collected ground-water samples from domestic supply wells and a pumping well installed for a pump test conducted as a part of exploration for development of the tribe's water supply. These ground-water samples were analyzed for physical properties, major ions, dissolved solids, nutrients, trace elements, pesticides, and bacteria, and results identified dissolved solids, sodium, sulfate, boron, iron, and manganese as exceeding their respective water-quality criteria (Trombley, 1999, 2001).

During the May 2001 through August 2003 study period, additional analyses for dissolved solids, major ions, trace elements, and suspended-sediment concentration were conducted (Schmidt, 2004). Results from this study period found that total phosphorus and triazine herbicide concentrations and fecal-indicator bacteria densities continued to occasionally exceed their respective water-quality criteria at some surfacewater monitoring sites on the reservation. Ranges of total phosphorus concentrations generally remained similar to the previous study periods with occasional samples exceeding the water-quality criteria. However, total phosphorus concentrations did increase at site LSC06 from the June 1996 through November 1998 study period (fig. 10 in Trombley, 1999) to the February 1999 through February 2001 study period (fig. 9 in Trombley, 2001). This increase in total phosphorus could be attributed to installation of a sewage-treatment lagoon upstream from the monitoring site in 1998. The range of total phosphorus concentrations for the current study period reported herein was similar to the May 2001 through August 2003 study period (Schmidt, 2004).

The median triazine herbicide concentration (triazine screen by ELISA) for the current study period was equal to or slightly less than past study periods—0.20 µg/L reported in Trombley (1999), 0.50 µg/L reported in Trombley (2001), 0.40 µg/L reported in Schmidt (2004), and 0.20 µg/L during the current study period. In addition, surface-water samples during the current study period did not exceed the USEPA MCL of 3.0 μ g/L for atrazine (by GC/MS or regression-predicted ELISA results) as an annual average. Four surfacewater samples collected during the first study period (June

1996–November 1998), nine surface-water samples collected during the second study period (February 1999–February 2001), and two surface-water samples collected during the third study period (May 2001–August 2003) exceeded the USEPA MCL for atrazine.

The median density of fecal coliform bacteria in surfacewater samples during the current study period was similar to the third study period (Schmidt, 2004) but less than the first two periods (Trombley, 1999, 2001). The median concentration for the June 1996 through November 1998 study period was 290 col/100 mL; the median concentration for the February 1999 through February 2001 study period was 570 col/100 mL; and the median concentration for the May 2001 through August 2003 study period was 190 col/100 mL. For the current study period, the median concentration for fecal coliform bacteria was 230 col/100 mL.

Trombley (2001) found dissolved solids, sodium, sulfate, boron, iron, and manganese concentrations in selected ground-water samples exceeding their respective water-quality criteria. It is not possible to infer trends in ground-water quality across all study periods because the first two study periods consisted of ground-water samples collected from domestic wells and one pumping well. Ground-water sample results presented herein and in the third study (Schmidt, 2004) were collected from dedicated monitoring wells installed across the reservation. The May 2001 through August 2003 study period (Schmidt, 2004) found that dissolved solids, sodium, and sulfate were present in ground-water samples at concentrations greater than their respective water-quality criteria. The current study period found that dissolved solids, sodium, sulfate, and chloride were again present in concentrations that exceeded their respective water-quality criteria and were larger than the previous study period. Results from these two most recent study periods show that arsenic concentrations continue to exceed the USEPA MCL in samples from monitoring well MW06 but do not show a substantial increase in concentration from the 2003 (Schmidt, 2004) to the current study period.

Summary and Conclusions

This report describes surface- and ground-water quality data collected on the Prairie Band Potawatomi Reservation in northeastern Kansas from November 2003 through August 2006 (hereinafter referred to as the "current study period"). Data from this study period are compared to results from June 1996 through August 2003, which are published in previous reports as part of a multiyear cooperative study with the Prairie Band Potawatomi Nation. Surface and ground water are important to the Prairie Band Potawatomi Nation to serve the hunting and fishing needs of tribal members and as a possible ground-water supply to support expanding commercial enterprise and development. Under the Federal Clean Water Act of 1977, an Indian tribe is sovereign for the purposes of delegating the authority to regulate water within

the reservation boundaries. To establish water-quality baseline concentrations with a goal of developing water-quality criteria to help manage the water resources on the reservation, the tribe, in cooperation with the U.S. Geological Survey, began a water-quality monitoring program in 1996.

A subset of 5 of the original 20 surface-water sites was sampled approximately quarterly from November 2003 through August 2006. All surface-water samples were analyzed for physical properties, dissolved solids, major ions, nutrients, trace elements, pesticides, fecal-indicator bacteria, suspended-sediment concentration, and total suspended solids. Ground water was sampled three times during this current study period from 10 dedicated ground-water monitoring wells completed on the reservation in August 2002 and July 2003. All ground-water samples were analyzed for physical properties, dissolved solids, major ions, nutrients, trace elements, pesticides, and fecal-indicator bacteria. Chemical oxygen demand and volatile organic compounds were analyzed in all three samples from one ground-water monitoring well located near a construction and demolition landfill on the reservation (monitoring well MW04) and in one sample from another well in the Soldier Creek drainage basin (well MW07).

Previous reports published as a part of this ongoing study identified total phosphorus, triazine herbicides, and fecal coliform bacteria as exceeding their respective water-quality criteria in surface water on the reservation. Previous groundwater assessments identified occasional sample concentrations of dissolved solids, sodium, sulfate, boron, iron, and manganese as exceeding their respective water-quality criteria.

During the current study period, dissolved-solids concentrations exceeded the Secondary Drinking-Water Regulation of 500 mg/L established by the U.S. Environmental Protection Agency (USEPA) in surface-water samples from site LSC06 located on Big Elm Creek downstream from a sewagetreatment lagoon and in samples from site SC03 on Soldier Creek. In general, larger concentrations of dissolved solids may occur downstream from wastewater-treatment facilities when compared to upstream concentrations. Twenty percent of dissolved-solids concentrations in ground-water samples were equal to or greater than the Secondary Drinking-Water Regulation.

Sodium concentrations exceeded the Drinking-Water Advisory of 20 mg/L established by USEPA in some surface-water samples from every site (almost 50 percent of all surface-water samples) and in ground-water samples from wells MW01, MW03, MW04, MW08, and MW11 (total of 50 percent of ground-water samples). Sulfate and chloride concentrations exceeded the Secondary Drinking-Water Regulation of 250 mg/L established by USEPA in 10 percent of ground-water samples, including all three samples from well MW03.

Detections of nitrite plus nitrate in ground-water samples from well MW11 exceeded the USEPA Maximum Contaminant Level of 10 mg/L. Fifty-six percent of the 55 surfacewater samples analyzed for total phosphorus exceeded the aquatic-life goal of 0.10 mg/L established by USEPA.

Three detections of arsenic in ground-water samples from well MW06 exceeded the USEPA Maximum Contaminant Level of 10 μ g/L. Boron also exceeded the USEPA Drinking-Water Advisory of 600 μ g/L in three samples from one well (MW03), and iron concentrations were larger than the Secondary Drinking-Water Regulation (300 µg/L) in four wells.

A triazine herbicide screen was used on 55 surface-water samples, and triazine herbicides, specifically, were detected frequently, although concentrations did not exceed the USEPA Maximum Contaminant Level of 3.0 µg/L as an annual average. Triazine herbicide compounds were not detected frequently in samples collected from ground-water monitoring wells. Triazine herbicides and their degradates are listed on the USEPA Contaminant Candidate List.

One volatile organic compound, diethyl ether, was detected in a water sample collected from well MW07. There are no USEPA drinking-water regulations for this compound.

Fecal-indicator bacteria (fecal coliform, *E. coli*, fecal streptococci, and enterococci) were detected in the majority of surface-water samples. Water-quality criterion have been established by USEPA and the Kansas Department of Health and Environment (KDHE) for densities of *E. coli*. Densities of *E. coli* in surface-water samples exceeded USEPA primarycontact water-quality criterion for moderately used water bodies (illness rate of 8 per 1,000 swimmers; 298 col/100 mL) in 40 percent of samples and exceeded KDHE single-sample maximum criterion for primary contact in public-access bodies of water (1,198 col/100mL for samples collected between April 1 and October 31) in 41 percent of samples. Median concentrations of both suspended sediment and total suspended solids in water samples from Soldier and Little Soldier Creek sites were small and reflect the low streamflow conditions under which most samples were collected.

Water quality on the Prairie Band Potawatomi Reservation generally meets water-quality criteria established by USEPA. Surface-water quality on the reservation is affected by agricultural runoff (triazine herbicides and fecal-indicator bacteria), seepage from septic systems or sewage lagoons serving a rural population, and wastewater discharge from documented point sources (dissolved solids and sodium). Sources of total phosphorus in streams can be attributed to both agricultural activities and human and animal waste. Ground-water quality on the reservation probably is affected by agricultural practices with the detections of nitrite plus nitrate, and elevated dissolved-solids and arsenic concentrations in ground water likely are from dissolution of sedimentary rocks. Concentrations of dissolved solids and sulfate in some ground-water samples exceeded their respective Secondary Drinking-Water Regulations, and concentrations exceeded the taste threshold of the USEPA's Drinking-Water Advisory Level for sodium. Consequently, in the event that ground water on the reservation is to be used as a drinking-water source, additional treatment may be necessary to remove excess dissolved solids, sulfate, sodium, chloride, nitrite plus nitrate, and arsenic.

References Cited

American Public Health Association, American Water Works Association, and Water Environment Federation, 1992, Standard methods for the examination of water and wastewater (18th ed.): Washington, D.C., American Public Health Association, Part 9000, p. 63–65.

Anderson, C.W., 2005, Turbidity (version 2.1): U.S. Geological Survey Techniques of Water-Resources Investigations, book 9, chap. A6, section 6.7, information available on Web, accessed May 10, 2007, at *http://pubs.water.usgs.gov/ twri9A6/*

Bates, R.L., and Jackson, J.A., eds., 1980, Glossary of geology (2nd ed.): Falls Church, Virginia, American Geological Institute, p. 658.

- Britton, L.J., and Greeson, P.E., eds., 1989, Methods for collection and analysis of aquatic biological and microbiological samples: U.S. Geological Survey Techniques of Water-Resources Investigations, book 5, chap A4, 363 p.
- Childress, C.J.O., Foreman, W.T., Connor, B.F., and Maloney, T.J., 1999, New reporting procedures based on longterm method detection levels and some considerations for interpretations of water-quality data provided by the U.S. Geological Survey National Water Quality Laboratory: U.S. Geological Survey Open-File Report 99–193, 19 p.
- Christensen, V.G., and Pope, L.M., 1997, Occurrence of dissolved solids, nutrients, atrazine, and fecal coliform bacteria during low flow in the Cheney Reservoir watershed, south-central Kansas, 1996: U.S. Geological Survey Water-Resources Investigations Report 97–4153, 13 p.
- Christensen, V.G., and Ziegler, A.C., 1998, Characterization of atrazine in source water—a case study from Kansas, *in* Water Quality Technical Conference, San Diego, California, November 1–4, 1998, Proceedings: Denver, Colorado, American Water Works Association, CD-ROM Catalog 20400, 9 p.
- Conner, B.F., Rose, D.L., Noriega, M.C., Murtagh, L.K., and Abney, S.R., 1998, Methods of analysis by the U.S. Geological Survey National Water Quality Laboratory—determination of 86 volatile organic compounds in water by gas chromatography/mass spectrometry, including detections less than reporting limits: U.S. Geological Survey Open-File Report 97–829, 78 p.
- Daly, M.H., and Lindsey, B.D., 1996, Occurrence and concentrations of volatile organic compounds in shallow ground water in the lower Sesquehanna River basin, Pennsylvania and Maryland: U.S. Geological Survey Water-Resources Investigations Report 96–4141, information available on Web, accessed May 14, 2004, at *[http://pa.water.usgs.gov/](http://pa.water.usgs.gov/reports/wrir_96-4141/report.html#HDR11) [reports/wrir_96-4141/report.html#HDR11](http://pa.water.usgs.gov/reports/wrir_96-4141/report.html#HDR11)*

 Dreeszen, V.H., and Burchett, R.R., 1971, Buried valleys in the lower part of the Missouri River Basin, *in* Pleistocene stratigraphy of the Missouri River Valley along the Kansas-Missouri border, Guidebook, 20th Annual Meeting of the Midwest Friends of the Pleistocene: Kansas Geological Survey Special Distribution Publication 53, p. 21–25.

Federal Inter-Agency Sedimentation Project, 1963, Determination of fluvial sediment discharge: Minneapolis, Minnesota, St. Anthony Falls Hydraulics Laboratory, Inter-agency Report M, 16 p.

Fenneman, N.M., 1938, Physiography of the Eastern United States: New York, McGraw-Hill, Inc., p. 588–605.

Fenneman, N.M., and Johnson, D.G., 1946, Physical divisions of the United States [map]: U.S. Geological Survey, 1 sheet, scale 1:7,000,000.

Fishman, M.J., and Friedman, L.C., 1989, Methods for determination of inorganic substances in water and fluvial sediments: U.S. Geological Survey Techniques of Water-Resources Investigations, book 5, chap. A1, 545 p.

Gibs, J., and Wilde, F.D., 1999, Ground-water sampling preparations and purging methods at water-supply wells and monitoring wells, *in* Wilde, F.D., Radtke, D.B., Gibs, J., and Iwatsubo, R.T., eds., National field manual for the collection of water-quality data—collection of water samples: U.S. Geological Survey Techniques of Water-Resources Investigations, book 9, chap. A4.2, 30 p.

Gleeson, C., and Gray, N.F., 1997, The coliform index and waterborne disease—problems of microbial drinking water assessment: London, E. and F.N. Spon, LTD., 210 p.

Gray, J.R., Glysson, G.D., Turcios, L.M., and Schwarz, G.E., 2000, Comparability of suspended-sediment concentration and total suspended solids data: U.S. Geological Survey Water-Resources Investigations Report 00–4191, 14 p.

Guy, H.P., 1977, Laboratory theory and methods for sediment analysis: U.S. Geological Survey Techniques of Water-Resources Investigations, book 5, chap. C1, 58 p.

Haack, S.K., and Rachol, C.M., 2000, Arsenic in ground water in Sanilac County, Michigan: U.S. Geological Survey Fact Sheet 132–00, 2 p.

Hem, J.D., 1992, Study and interpretation of the chemical characteristics of natural water (3rd ed.): U.S. Geological Survey Water-Supply Paper 2254, 263 p.

Holt, J.G., Krieg, N.R., Sneath, P.H.A., and others, eds., 1993, Bergey's manual of determinative bacteriology (9th ed.): Baltimore, Maryland, Williams and Wilkens, 787 p.

Jones, D.R., and Garbarino, J.R., 1999, Methods of analysis by the U.S. Geological Survey National Water Quality Laboratory—determination of arsenic and selenium in water and sediment by graphite furnace-atomic absorption spectrometry: U.S. Geological Survey Open-File Report 98–639, 39 p.

Jorgensen, D.G., Helgesen, J.O., and Imes, J.L., 1993, Regional aquifers in Kansas, Nebraska, and parts of Arkansas, Colorado, Missouri, New Mexico, Oklahoma, South Dakota, Texas, and Wyoming—geohydrologic framework: U.S. Geological Survey Professional Paper 1414–B, 72 p.

Kansas Applied Remote Sensing Program, 1993, Kansas land cover data base, 1:100,000 scale: Lawrence, Kansas, Data Access and Support Center, available on CD.

Kansas Department of Agriculture, 2004, Fertilizer tonnage reports totals, January 2003 through June 2003: Pesticide and Fertilizer Program, information available on Web, accessed May 24, 2004, at *[http://www.accesskansas.org/](http://www.accesskansas.org/kda/Pest&Fert/pest-ferttonnagetotal.htm) [kda/Pest&Fert/pest-ferttonnagetotal.htm](http://www.accesskansas.org/kda/Pest&Fert/pest-ferttonnagetotal.htm)*

Kansas Department of Health and Environment, 2001, Kansas register: Topeka, Kansas, Secretary of State, v. 20, no. 33, p. 1264–1277, information available on Web, accessed May 5, 2004, at *<http://www.kdhe.ks.us/pdf/regs/28-16-28.pdf>*

Kansas Department of Health and Environment, 2004, Kansas surface water quality standards: Information available on Web, accessed February 6, 2007, at *http://www.kdheks.gov/ water/download/swqs_numeric_criteria.pdf*

Kish, J.L., Thurman, E.M., Scribner, E.A., and Zimmerman, L.R., 2000, Methods of analysis by the U.S. Geological Survey Organic Geochemistry Research Group—determination of selected herbicides and their degradation products in water using solid-phase extraction and gas chromatography/ mass spectrometry: U.S. Geological Survey Open-File Report 00–385, 13 p.

Myers, D.N., and Wilde, F.D., eds., 2003, Biological indicators, *in* National field manual for the collection of water-quality data: U.S. Geological Survey Techniques of Water-Resources Investigations, book 9, chap. A7 (3rd ed.), various pagination, information available on Web, accessed April 1, 2004, at *[http://water.usgs.gov/owq/FieldManual/](http://water.usgs.gov/owq/FieldManual/Chapter7/NFMChap7.pdf) [Chapter7/NFMChap7.pdf](http://water.usgs.gov/owq/FieldManual/Chapter7/NFMChap7.pdf)*

Pais, I., and Jones, J.B., Jr., 1997, The handbook of trace elements: Boca Raton, Florida, St. Lucie Press, 223 p.

Pope, L.M., and Putnam, J.E., 1997, Effects of urbanization on water quality in the Kansas River, Shunganunga Creek Basin, and Soldier Creek, Topeka, Kansas, October 1993 through September 1995: U.S. Geological Survey Water-Resources Investigations Report 97–4045, 84 p.

Quanrud, D.M., Quast, K., Conroy, O., Karpiscak, M.M., Gerba, C.P., Lansey, K.E., Ela, W.P., and Arnold, R.G., 2004, Estrogenic activity and volume fraction of waste water origin in monitoring wells along the Santa Cruz River, Arizona: Ground Water Monitoring and Remediation, v. 24, no. 2, p. 86–93.

- Radtke, D.B., White, A.F., Davis, J.V., and Wilde, F.D., 1998, Dissolved oxygen, *in* Wilde, F.D., and Radtke, D.B., eds., 1998, National field manual for the collection of waterquality data—field measurements: U.S. Geological Survey Techniques of Water-Resources Investigations, book 9, chap. A6.2, 38 p.
- Rasmussen, P.P., and Ziegler, A.C., 2003, Comparison and continuous estimates of fecal coliform and *Escherichia coli* bacteria in selected Kansas streams, May 1999 through April 2002: U.S. Geological Survey Water-Resources Investigations Report 03–4056, 80 p.
- Reid, G.K., and Wood, R.D., 1976, Ecology of inland waters and estuaries: New York, D. Van Nostrand Co., p. 235 and 293.
- Schmidt, H.C. Ross, 2004, Quality of water on the Prairie Band Potawatomi Reservation, northeastern Kansas, May 2001 through August 2003: U.S. Geological Survey Scientific Investigations Report 2004–5243, 69 p.
- Schoewe, W.H., 1949, The geography of Kansas: Transactions of Kansas Academy of Science, v. 52, no. 3, p. 261–333.
- Stamer, J.K., and Zelt, R.B., 1994, Organo-nitrogen herbicides in the lower Kansas River Basin: Journal of American Water Works Association, January 1994, p. 93–104.
- Thurman, E.M., Meyer, Michael, Pomes, Michael, Perry, C.A., and Schwab, A.P., 1990, Enzyme-linked immumosorbent assay compared with gas chromatography/mass spectrometry for the determination of triazine herbicides in water: Analytical Chemistry, v. 62, p. 2043–2048.
- Trombley, T.J., 1999, Surface-water quality on the Prairie Band of Potawatomi Reservation, northeastern Kansas, June 1996 through November 1998: U.S. Geological Survey Water-Resources Investigations Report 99–4266, 67 p.
- Trombley, T.J., 2001, Quality of water on the Prairie Band Potawatomi Reservation, northeastern Kansas, February 1999 through February 2001: U.S. Geological Survey Water-Resources Investigations Report 01–4196, 51 p.
- Trombley, T.J., Wolf, R.J., Jordan, P.R., and Brewer, L.D., 1996, Overview of water resources in and near Indian lands in northeastern Kansas and southeastern Nebraska: U.S. Geological Survey Water-Resources Investigations Report 96–4070, 68 p.
- U.S. Environmental Protection Agency, 1986, Quality criteria for water, 1986: U.S. Environmental Agency Report 440/5–86–001, unnumbered pages.
- U.S. Environmental Protection Agency, 1999, 1999 Update of ambient water quality criteria for ammonia: Washington, D.C., EPA–822–R–99–014, December 1999, 147 p., information available on Web at *http://www.epa.gov/ waterscience/standards/ammonia/99update.pdf*
- U.S. Environmental Protection Agency, 2000a, Improved enumeration methods for the recreational water quality indicators—enterococci and *Escherichia coli*: Washington, D.C., EPA/821/R–97/004, March 2000, information available on Web, accessed February 2003 at *[http://www.epa.gov/](http://www.epa.gov/nerlcwww/RecManv.pdf) [nerlcwww/RecManv.pdf](http://www.epa.gov/nerlcwww/RecManv.pdf)*
- U.S. Environmental Protection Agency, 2000b, Ambient water quality criteria recommendations—information supporting the development of State and tribal nutrient criteria, rivers and streams in nutrient ecoregion VI: Washington, D.C., EPA 822–B–00–017, December 2000, information available on Web, accessed May 2004 at *[http://www.epa.](http://www.epa.gov/waterscience/criteria/nutrient/ecoregions/rivers/rivers_4.pdf) [gov/waterscience/criteria/nutrient/ecoregions/rivers/](http://www.epa.gov/waterscience/criteria/nutrient/ecoregions/rivers/rivers_4.pdf) [rivers_6.pdf](http://www.epa.gov/waterscience/criteria/nutrient/ecoregions/rivers/rivers_4.pdf)*
- U.S. Environmental Protection Agency, 2000c, Ambient water quality criteria recommendations—information supporting the development of State and tribal nutrient criteria, rivers and streams in nutrient ecoregion IX: Washington, D.C., EPA 822–B–00–019, December 2000, information available on Web, accessed May 2004 at *[http://www.epa.](http://www.epa.gov/waterscience/criteria/nutrient/ecoregions/rivers/rivers_9.pdf) [gov/waterscience/criteria/nutrient/ecoregions/rivers/](http://www.epa.gov/waterscience/criteria/nutrient/ecoregions/rivers/rivers_9.pdf) [rivers_9.pdf](http://www.epa.gov/waterscience/criteria/nutrient/ecoregions/rivers/rivers_9.pdf)*
- U.S. Environmental Protection Agency, 2001, Ambient water quality criteria recommendations—information supporting the development of State and tribal nutrient criteria, rivers and streams in nutrient ecoregion IV: Washington, D.C., EPA 822–B–01–013, December 2001, information available on Web, accessed May 2004 at *[http://www.epa.](http://www.epa.gov/waterscience/criteria/nutrient/ecoregions/rivers/rivers_4.pdf) [gov/waterscience/criteria/nutrient/ecoregions/rivers/](http://www.epa.gov/waterscience/criteria/nutrient/ecoregions/rivers/rivers_4.pdf) [rivers_4.pdf](http://www.epa.gov/waterscience/criteria/nutrient/ecoregions/rivers/rivers_4.pdf)*
- U.S. Environmental Protection Agency, 2002a, Method 16031 *Escherichia coli (E. coil)* in water by membrane using modified membrane-thermotolerant *Escherichia coli* agar (modified mTec): U.S. Environmental Protection Agency Report EPA–821–R–02–023, September 2002, 9 p.
- U.S. Environmental Protection Agency, 2002b, Implementation guidance for ambient water quality criteria for bacteria, May 2002 draft: Washington, D.C., EPA–823–B–02–003, information available on Web, accessed March 2004 at *[http://www.epa.gov/ostwater/standards/bacteria/](http://www.epa.gov/ostwater/standards/bacteria/bacteria.pdf) [bacteria.pdf](http://www.epa.gov/ostwater/standards/bacteria/bacteria.pdf)*
- U.S. Environmental Protection Agency, 2003, Contaminant candidate list regulatory determination support document for sodium: Washington, D.C., EPA 815–R–03–015, July 2003, information available on Web, accessed May 2004 at *http://www.epa.gov/safewater/ccl/pdf/sodium_rsd.pdf*
- U.S. Environmental Protection Agency 2006, 2006 Edition of the drinking water standards and health advisories: Washington, D.C., EPA 822–R–06–013, summer 2006, information available on Web, accessed February 13, 2007, at *http://www.epa.gov/waterscience/criteria/drinking/ dwstandards.pdf*
- Walters, K.L., 1953, Geology and ground-water resources of Jackson County, Kansas: Kansas Geological Survey Bulletin 101, 91 p.
- Wilde, F.D., Radtke, D.B., Gibs, Jacob, and Iwatsubo, R.T., eds., 1999, Collection of water samples: U.S. Geological Survey Techniques of Water-Resources Investigations, book 9, chap. A4, information available on Web, accessed March 3, 2004, at *http://pubs.water.usgs.gov/twri9A4/*
- Zaugg, S.D., Sandstrom, M.W., Smith, S.G., and Fehlberg, K.M., 1995, Methods of analysis by the U.S. Geological Survey National Water Quality Laboratory—determination of pesticides in water by C-18 solid-phase extraction and capillary column gas chromatography/mass spectrometry with selected-ion monitoring: U.S. Geological Survey Open-File Report 95–181, 60 p.
- Ziegler, A.C., 2002, Issues related to use of turbidity measurements as a surrogate for suspended sediment, *in* Turbidity and Other Sediment Surrogates Workshop, Reno, Nevada, April 30–May 2, 2002, Proceedings: Information available on Web, accessed August 6, 2007, at *http://water.usgs. gov/osw/techniques/TSS/ZieglerT.pdf*

Publishing support provided by: Rolla Publishing Service Center

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