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Hawaiian Bottomfish Assessment Update for 2008

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## Hawaiian Bottomfish Assessment Update for 2008

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#### Abstract

The Hawaiian bottomfish assessment was updated through 2007. The updated assessment was conducted using re-audited bottomfish catch and effort data from commercial logbook records collected during 1948-2007. A standardized bottomfish catch-per-unit effort (CPUE) data set was constructed for the main Hawaiian Islands using the re-audited catch and effort data and a multiplicative log-linear estimation model. The standardized series included fishing year, month, and fishing area as predictors of bottomfish CPUE on directed fishing trips ( $\geq 50 \%$ bottomfish catch by weight). Each of the predictors had a significant ( $\mathrm{P}<0.0001$ ) influence on observed CPUE by trip. For comparison, the CPUE filtering approach used in previous bottomfish assessments was also applied to the revised CPUE data set to obtain a status quo CPUE series. A Bayesian production model was used to estimate bottomfish biomass and harvest rate time series. This model was also used to conduct short-term projections of future catches and associated risks of overfishing. These projections explicitly included uncertainty in the posterior distribution of estimated bottomfish biomass in 2007 and population dynamics parameters. The production model was fit to catch and standardized CPUE data for each of the three Hawaiian fishing zones: the main Hawaiian Islands zone, the Mau zone, and the Hoomalu zone. The production model fitting incorporated uninformative priors for carrying capacity, process error, observation error, and catchability parameters and an informative prior for intrinsic growth rate. Results of the catch and CPUE analyses, production modeling, and projections are summarized below.


## INTRODUCTION

The Hawaiian bottomfish complex is a U.S. fishery management unit comprised primarily of several species of snappers and jacks and a grouper inhabiting waters of the Hawaiian Archipelago (Table 1). The archipelagic management unit includes three fishing zones: the main Hawaiian Islands (MHI) zone, the Mau zone, and the Hoomalu zone. Most fishing currently takes place in the MHI zone. The assessed bottomfish complex includes a subset of seven species, the "Deep 7", that are a particular focus of management (Table 1). The Hawaiian bottomfish fishery is a traditional deep-water fishery that targets these species primarily using deepwater handline gear. Hawaiian bottomfish were targeted by Hawaiians using deep handlines from canoes for hundreds of years before the advent of the modern fishery after World War II. The modern fishery employs similar deep handline gear, albeit with braided synthetic line, along with power reels to haul back gear, fish finders to locate schools of fish, and GPS units and other navigational aids to find fishing grounds. Although the efficiency of fishing vessels has improved through time (Moffitt et al., 2009), the Hawaiian bottomfish fishery still uses traditional deep handline capture methods for commercial and recreational harvest.

## Previous Assessment

The previous assessment of the Hawaiian bottomfish complex was conducted in 2005 using fishery data through calendar year 2004 (Moffitt et al., 2006 ). This assessment included surplus production model analyses of bottomfish catch and catch-per-unit effort data for the three fishing zones (Fig. 1). In these analyses, bottomfish CPUE was assumed to be proportional to relative abundance in each zone. Catchability of bottomfish in the MHI zone was assumed to be time varying and increasing through the assessment time horizon. Observed CPUE in the MHI was adjusted by an estimated catchability multiplier in four time periods to account for changes in catchability. Predicted CPUE was fitted to observed CPUE using nonlinear least squares to estimate model parameters. Biological reference points to achieve maximum sustainable yield (MSY) were estimated for each fishing zone. Stock status of the MHI bottomfish was assessed for calendar years 1948-2004 while status of the Mau and Hoomalu zones was assessed for 1988-2004. The status of the Archipelagic management unit was assessed based on a weighted average of the stock status of the three zones, where the weights were the fraction of total bottomfish habitat by zone. Assessment results indicated that the Archipelagic management unit was not depleted but was experiencing overfishing in 2004. Fishing mortality for the Archipelagic bottomfish management unit was estimated to be $24 \%$ above the overfishing threshold. As a result, management measures were crafted to reduce fishing mortality on MHI bottomfish where overfishing was occurring; these included seasonal fishery closures and total allowable commercial catch limits. Analyses of bottomfish total allowable catch limits were conducted by the PIFSC in 2007-2008 to address the issue of overfishing on the Archipelagic management unit and the MHI bottomfish stock.

## Previous Risk Analyses

Risk analyses for bottomfish were conducted to provide guidance on appropriate total allowable bottomfish catches in 2007 and again in 2008. In 2007, a deterministic projection analysis was conducted to estimate the magnitude of the reduction in the MHI fishing mortality and catch to cease overfishing on the Archipelagic management unit. It was expected that there might be a sufficient reduction in commercial landings ( $25 \%$ ) assuming the seasonal bottomfish closure (WPFMC, 2007) was perfectly implemented with no seasonal shifts in fishing effort and catch. Because the assumption that fishing effort would not shift seemed unlikely, alternative 2007 catch levels in the MHI that would meet the target reduction in fishing mortality (F) were projected (Brodziak et al., 2007). Three alternative total allowable catch levels in $2007\left(\mathrm{TAC}_{2007}\right)$ for the Deep 7 bottomfish species in the main Hawaiian Islands were developed. These were: (1) $\mathrm{TAC}_{2007}$ set at $76 \%$ of reported 2004 Deep 7 landings; (2) TAC 2007 set with $\mathrm{F}_{2007}$ at $76 \%$ of $\mathrm{F}_{2006}$ which was assumed to be equal to $\mathrm{F}_{2005}$; (3) $\mathrm{TAC}_{2007}$ set with F at $76 \%$ of $\mathrm{F}_{2004}$. Under each alternative it was assumed that bottomfish fishing mortality had remained constant in the Northwestern Hawaiian Islands (NWHI) since 2004. For each alternative, the basic input data were the estimated bottomfish stock status in 2004 from Moffitt et al. (2006).

Results of the deterministic TAC calculations for the Deep 7 bottomfish in 2007 ranged from $\mathrm{TAC}_{2007}=72$ thousand pounds (klb) under alternative 3 , to $\mathrm{TAC}_{2007}=110 \mathrm{klb}$ under alternative 2 , to $\mathrm{TAC}_{2007}=178 \mathrm{klb}$ under alternative 1 (Brodziak et al., 2007). It is important to note that all of the alternative TACs in 2007 were calculated to just meet the overfishing threshold and did not provide any precautionary reductions to account for uncertainty in the input data, process dynamics, or assessment estimates. These considerations were incorporated into a subsequent stochastic TAC risk analysis in 2008.

In 2008, another assessment of the risk of overfishing the bottomfish complex in the Hawaiian Archipelago was conducted (Brodziak, 2008). The purpose of this analysis was to quantify the probabilities that overfishing would occur on the archipelagic bottomfish complex for a range of total allowable commercial catches of Deep 7 bottomfish species in the main Hawaiian Islands during the 2008 fishing year. This analysis was conditioned on the results of the 2005 stock assessment of the Hawaiian bottomfish complex (Moffitt et al., 2006) which provided estimates of bottomfish biomass and fishing mortality through 2004. Thus, one component of the assessment was to project the biomasses and fishing mortalities that likely occurred in 2005-2007 in order to compute probabilities of overfishing in 2008 under various TACs. This projection was accomplished by simulating the impacts of reported commercial catches during 2005-2007 using the biomass dynamics model from the most recent assessment (Moffitt et al., 2006) and by accounting for uncertainty in the estimates of key model parameters. In particular, the simulation analysis estimated the TACs for the Deep 7 bottomfish species that would produce risks of archipelagic overfishing in 2008 of $0 \%, 5 \%, 10 \%, \ldots, 100 \%$, conditioned on the baseline model assumptions.

Results of the stochastic risk analyses (Brodziak, 2008) indicated that the largest $\mathrm{TAC}_{2008}$ that would produce approximately $0 \%$ chance of overfishing in 2008 (i.e., exceeding $\mathrm{F}_{\text {TARGET }}$ ) was 24 thousand pounds (klb). In contrast, the smallest TAC that would lead to a roughly $100 \%$
chance of overfishing was 273 klb . Total allowable commercial catches in 2008 ranging from 24 to 99 thousand pounds, corresponded to risks of archipelagic overfishing ranging from $0 \%$ to $50 \%$. The TAC to achieve a low risk of overfishing ( $25 \%$ ) in 2008 was estimated to be $\mathrm{TAC}_{25 \%}=61 \mathrm{klb}$ and the TAC to achieve a neutral risk of overfishing (50\%) in 2008 was estimated to be $\mathrm{TAC}_{50 \%}=99 \mathrm{klb}$. The probability of exceeding $\mathrm{F}_{\text {TARGET }}$ was a concave function of $\mathrm{TAC}_{2008}$ over most of the TACs examined. This indicated that risk of overfishing increased less than proportionally with increasing $\mathrm{TAC}_{2008}$ values. Sensitivity analyses showed that the estimates of overfishing risk were highly sensitive to the estimates of biomass in 2004, intrinsic growth rate, and the proportion of Deep 7 bottomfish in the catch. In contrast, estimates of overfishing risk were moderately sensitive to the estimate of carrying capacity, the assumed bottomfish catches in 2007, and the coefficients of variation of key model parameters.

## Current Assessment

The current assessment was conducted in 2008 using fishery data through 2007 and provides an update of the previous Hawaiian bottomfish assessment through 2007. In this context, the updated assessment was constrained to using the same data sources as the preceding assessment and also was constrained to using a modeling approach that was similar or identical to that used in the preceding assessment. However, it is important to note that the updated assessment addressed two concerns about the 2005 bottomfish assessment. The first concern was that the bottomfish catch and effort data from the MHI during the early portion of the time series may have contained multi-day trips with larger catches than the single-day trips used to compute a standardized catch-per-unit effort time series. The presence of multi-day bottomfish trips in the 1950s and 1960s could skew the average MHI CPUE to higher values than were representative of the single-day trips used to calculate the standard CPUE time series. This would lead to an overestimate of bottomfish relative abundance in the 1950s-1960s when MHI CPUE was used as an abundance index in the surplus production model. Thus, an intensive effort was made to re-audit the MHI catch and effort data to eliminate multi-day trip records using criteria developed from interviews with long-time bottomfish fishermen. The second concern about the 2005 bottomfish assessment was that it employed a least-squares estimation approach that provided no direct measure of the variability of parameter estimates for conducting risk analyses. To address this concern, a Bayesian surplus production modeling approach was applied to estimate bottomfish biomass and fishing mortality rates in the current assessment, assuming similar model structure and assumptions.

Data for the updated assessment included re-audited bottomfish catch and effort data from Hawaii Division of Aquatic Resources (HDAR) records collected during 1948-2007. A revised bottomfish catch-per-unit effort data set was constructed using the re-audited catch and effort data. The CPUE filtering approach used in previous bottomfish assessments was applied to the revised CPUE data set to obtain the status quo CPUE series used in the previous assessment (Moffitt et al., 2006). A statistical log-linear model described by Gavaris (1980) was applied to develop a spatially standardized CPUE series for the main Hawaiian Islands. This series included fishing year, month, and fishing area as predictors of bottomfish CPUE on directed fishing trips ( $>50 \%$ bottomfish catch by weight). Each of the predictors had a significant ( $\mathrm{P}<0.0001$ ) influence on observed CPUE by trip. A Bayesian production model was used to
estimate bottomfish biomass and harvest rate time series (Brodziak, 2007). This model was also used to conduct short-term projections of future catches and associated risks of overfishing. These projections explicitly included uncertainty in the posterior distribution of estimated bottomfish biomass in 2007 and population dynamics parameters. The production model was fit to catch and standardized CPUE data for each of the three Hawaiian fishing zones: the main Hawaiian Islands zone, the Mau zone, and the Hoomalu zone. The production model fitting incorporated uninformative priors for carrying capacity, process error, observation error, and catchability parameters and an informative prior for intrinsic growth rate. The mean of the prior for intrinsic growth rate was equal to the estimate of the parameter used in the previous bottomfish assessment (Moffitt et al., 2006). Methods and results of the catch and CPUE analyses, production modeling analyses, and stochastic projection analyses for estimating 2009 TACs and associated probabilities of overfishing are described in detail below.

## MATERIALS AND METHODS

In this section, the bottomfish fishery data, the CPUE standardization method and results, and the production model used to estimate biomass and fishing mortality for the Hawaiian bottomfish assessment update are described.

## Fishing Year

In the previous assessment (Moffitt et al., 2006), the annual time period for reporting bottomfish catches (Table 2) was the calendar year from 1-January to 31-December. The time period for calculating CPUE in the MHI was from 1-July of the previous year through 30-June of the current year (the fishing year, FY) during 1948-1992. This time period was chosen because it corresponded to the time period used for renewal of commercial marine licenses (CML) for the State of Hawaii. In particular, one CML number was associated with one vessel permit during the FY, while on a calendar year basis the CML would change at the end of June halfway through each year. Beginning on 1-July-1993, the CML became permanent and did not change from fishing year to fishing year. This made it possible to track vessel permit numbers beginning in 1993. In the previous assessment, the calendar year was used as the time period for calculating CPUE from 1993 to 2005. This enabled the catches of individual vessels to be tallied across years in the hope of developing time series of individual vessel CPUE in the MHI. In the Mau and Hoomalu fishing zones, CPUE used in the previous assessment was calculated on a fishing year basis.

In the current assessment, the annual time period for reporting bottomfish catch and CPUE is the fishing year from 1-July of the previous year through 30-June of the current year. There were two primary reasons for making the change to fishing year. First, the fishing year corresponds to the annual biological cycle of the deepwater snapper and grouper bottomfish complex in which spawning occurs in late spring to early summer. Thus, estimates of annual production biomass starting in July coincided with the settlement of juvenile bottomfish through mid-summer. Second, the commercial fishery catch of bottomfish is typically highest
during the winter months when there is strong market demand for snapper during New Years and other holidays. The use of calendar year split the primary bottomfish season into two separate years. In contrast, the use of fishing year included the primary bottomfish season in one annual assessment time period. In addition, the use of fishing year might make it possible to track individual fishing vessel catches and CPUE prior to 1993 using annual records of the CML associated with individual vessel permit numbers. Overall, the treatment of MHI CPUE in the current assessment differs slightly from the use of CPUE from a mixture of fishing year and calendar year in the previous assessment.

## Fishery Catch

Fishery catch data for the updated assessment included commercial bottomfish catch and effort data extracted from a total of 4,066,464 HDAR logbook catch records submitted by commercial fishers during fishing years 1949-2007 (M. Quach, PIFSC, pers. comm. 14-July2008). Bottomfish catch data by trip were assigned to the three fishing zones based on the reported HDAR fishing area in the logbook (Fig. 1). Some bottomfish catch trip records had an unknown area and these small catches were apportioned to the three areas based on the annual catch proportion by area. As in the previous bottomfish assessment (Moffitt et al., 2006), the bottomfish catch used in the assessment update consisted of all reported bottomfish landings (Table 2 and Fig. 2) excluding the catches of kahala (Seriola dumerili) and taape (Lutjanus kasmira). Catches by trip were assigned to fishing year (the fishing year extends from July $1^{\text {st }}$ of the previous year to June $30^{\text {th }}$ of the current year). The fishing year corresponds to the fiscal year basis on which commercial fishing licenses are allocated by the state of Hawaii and also includes the entire winter season, when bottomfish catches typically are highest, in a single fishing year. Estimates of recreational bottomfish catch were not included in the current assessment, similar to the previous assessment. Overall, the total commercial bottomfish catch used in the assessment included fishing year catches from the MHI during 1949-2007 and the fishing year catches from the Mau and Hoomalu zones during 1988-2007 when fishery CPUE data were available from these two zones.

## Standardized Commercial Fishery Catch-Per-Unit Effort

Bottomfish catch per trip was calculated in the MHI for directed deep handline trips for which information was reported on fishing date, commercial marine license, area fished, and fishing gear code during FY 1948-2007. In particular, logbook catch records were audited to remove records that were missing either fishing date, CML number, or fishing area as well as records that did not have a fishing gear code equal to deep handline fishing gear. This led to a total of $\mathrm{n}=214,981$ directed commercial bottomfish trips using deep handline gear with a total catch of roughly 33.2 million pounds.

In an independent review of the previous Hawaiian bottomfish assessment, it was reported that there were a few logbook trip records for the main Hawaiian Islands fishing zone that exceeded 5,000 pounds of bottomfish catch per trip in the 1950s and 1960s. Such trips were expected to represent multiday trips in the MHI by larger vessels, and not the single-day trip that was
representative of the primary fishing fleet operating in the MHI during FY 1948-2007. Since there were no screening procedures to eliminate possible multi-days trips from the data set used for calculating MHI CPUE in the previous assessment, there was a concern that the status quo CPUE series might overestimate the long-term decline in commercial fishery CPUE. This, in turn, would be expected to lead to an overestimate of the decline of relative bottomfish biomass in the MHI.

To address the concern about the possible inclusion of multi-day trips in the CPUE standardization, several interviews with long-term bottomfish fishermen were conducted by the PIFSC to gather information on the historic operation of the fishery. In August 2008, the results of these interviews were reviewed at a CPUE standardization workshop that included participants from the PIFSC, HDAR, and the Western Pacific Regional Fishery Management Council (Moffitt et al., 2009). The workshop was also convened to address the issue of changes in bottomfish fishing technology through time and to address the issues of what data were appropriate and what qualifying criteria for including a trip in the bottomfish CPUE analyses were appropriate, e.g., include a trip only if the ratio of bottomfish catch weight to total catch weight was greater than $90 \%$, as in Moffitt et al. (2006).

The CPUE standardization workshop participants agreed to a number of points regarding the treatment of CPUE data from the MHI for the current assessment update (Moffitt et al. 2009). The workshop participants agreed that the exclusion of catch data for kahala, a species with decreased market demand since the mid-1980s due to possible ciguatera poisoning, and for taape, an introduced species, was appropriate and should be continued in the assessment update. The workshop also agreed that consideration of different trip qualifying criteria would be appropriate. In particular, there was a consensus that trips landing over 1,500 pounds per day would not be appropriate for single-day trip operations. The workshop participants also agreed that it would be useful to consider less restrictive qualifying criteria to increase the spatial coverage of the fishing operations used to compute standardized CPUE. Workshop participants also agreed that changes in the fishery catchability coefficient (Fig. 3) corresponding to major changes in technology and fleet dynamics should continue to be used in future assessments. Workshop participants also thought it would be useful to investigate the effects of vessel characteristics, oceanographic conditions, and socioeconomic factors on bottomfish CPUE in a medium to long-term research effort.

Based on the CPUE workshop recommendations, bottomfish CPUE time series for the main Hawaiian Islands were calculated using a standardization approach that included the spatial effect of fishing area as well as the status quo approach from Moffitt et al. (2006) for comparison. For each fishing year in 1949-2007, the status quo CPUE in the MHI was calculated using the selection algorithm from Moffitt et al. (2006) with the additional trip qualifying criterion that trips reporting over 1,500 pounds of bottomfish catch were not included in the CPUE analysis of single-day trips. Another trip qualifying criterion for the status quo approach was that bottomfish comprise at least $90 \%$ of the total trip catch by weight. Further, qualifying trips had to have annual catches that were $30 \%$ or more of the annual median of the top ten ranked Maui Nui fishing area catches by CML. The value of status quo CPUE in each year was the arithmetic average of the CPUE values for qualifying trips. On average, the status quo CPUE time series was based on $25 \%$ of the total deep handline catch of

MHI bottomfish during 1949-2007 (Fig. 4). Similarly, the status quo CPUE time series was based on $10 \%$ of the total deep handline trips that captured MHI bottomfish, on average, during 1949-2007 (Fig. 5). Applying the status quo CPUE approach to the re-audited HDAR data produced a similar trend to that observed in the status quo CPUE series from the previous assessment (Spearman $\mathrm{R}=0.83, \mathrm{P}<0.001$ ). Overall, the status quo CPUE from the current and the previous assessments were significantly positively correlated.

For the current assessment, bottomfish CPUE time series in the Mau and Hoomalu zones during 1988-2004 were taken from the previous assessment (Moffitt et al., 2006) and CPUE values during 2005-2007 were updated using the same approach as in the previous assessment (Fig. 7). Overall, the CPUE series from the Mau and Hoomalu zones were negatively correlated (Spearman $\mathrm{R}=-0.32, \mathrm{P}=0.17$ ) during 1988-2007 but this association was not significant.

For the current assessment, bottomfish CPUE for the main Hawaiian Islands was standardized using a multiplicative loglinear model (e.g., a special case of a generalized linear model, GLM) applied to the re-audited HDAR data. This enabled us to evaluate the statistical significance of the available predictors and to calculate the standardized CPUE from the estimated year coefficients of the fitted GLM. The format of the HDAR logbook has changed since 2000 to include more detailed information on fishing operations, including line hours fished for deep handline fishing, but these detailed data were not recorded throughout most of the logbook time series from 1949 to 2007. However, two potential predictors of bottomfish CPUE were consistently recorded in the HDAR data during 1949-2007; these were area fished and month fished for each trip. In the context of CPUE standardization, each of these factors could have an effect on bottomfish CPUE that varied on an annual basis due to changes in the fishery and distribution of fish. The area factor represented potential differences in the spatial distribution of bottomfish and their catchability. The month factor represented potential differences in the seasonal distribution of bottomfish and associated catchability. Overall, the goal of the standardization analysis was to remove the impact of the spatial and seasonal factors on the annual relative abundance of bottomfish, as indexed by the year effect estimated in the GLM.

The HDAR logbook reporting areas were used to develop a set of area strata for the directed bottomfish fishery in the MHI. In particular, the logbook CPUE data were constrained to be single-day trips (bottomfish catch less than $1,500 \mathrm{lbs}$ ) using deep handline gear. There were a large number of individual HDAR reporting areas in the nearshore ( $<3$ miles) and offshore ( $>3$ miles) around the main Hawaiian Islands (Fig. 8). To avoid estimating area effect coefficients for each of these areas, many of which had little or no reported bottomfish catch, we tabulated the reported bottomfish catch by area during 1949-2007 to discern the key bottomfish areas. In this case, the key 3-digit areas were those areas that accounted for at least $1 \%$ of the total bottomfish handline catch during 1949-2007 (Fig. 8, red shading 3-digit areas). The remaining three digit areas were grouped by island region to define an aggregate area for estimating the spatial effect of area on CPUE in the GLM. In particular, the six aggregate areas were (Fig. 8): Other Offshore (light tan shading), Other Kauai (lavender shading), Other Oahu (light orange shading), Other Maui, Molokai, and Lanai (light green shading), and Other Hawaii (pink shading). Overall, there were twenty six key areas and six aggregate areas used to estimate the spatial effect in the GLM.

Potential seasonal effects on bottomfish CPUE in the MHI were accounted for by considering each month as a factor level. In this case, the significance of the seasonal effect was assessed using the analysis of deviance explained by the month effect in the GLM.

The selection of a qualifying percentage of bottomfish catch by weight per trip was revisited for the GLM analyses of MHI CPUE as suggested by the CPUE Standardization Workshop. Roughly one-half of the deep handline trips that landed some bottomfish during 1949-2007 also landed some other non-bottomfish species. This suggested that a ratio of at least $50 \%$ bottomfish catch weight per trip would be a reasonable cutoff to define a fishing trip that had probably targeted bottomfish for a substantial fraction, if not a majority, of the fishing time expended. The annual distribution of bottomfish catch ratio for deep handline gear (Fig. 9) also indicated that the median trip caught a high percentage of bottomfish, fluctuating around $80 \%$ during the 1950s-1980s and increasing to nearly $100 \%$ in the 1990 s. This implied that the use of a $50 \%$ bottomfish ratio would provide better coverage of the fishery than a $90 \%$ cutoff. As a result, a $50 \%$ bottomfish ratio per trip was chosen as the qualifying criterion for the GLM analyses to standardize CPUE. This choice implied that, on average, roughly $85 \%$ of the annual bottomfish catch (Fig. 4) and $73 \%$ of deep handline trips (Fig. 5) were used in the CPUE standardization analysis. GLM sensitivity analyses were conducted for a bottomfish ratio of $90 \%$ and for all deep handline trips that caught at least 1 pound of bottomfish.

A statistical approach described by Gavaris (1980) was applied to develop a spatially standardized CPUE series for the main Hawaiian Islands. The multiplicative loglinear model to standardize bottomfish catch rate per trip (CPUE) in the MHI had three explanatory variables. These were fishing year (Y), fishing area (A), and fishing month (M). These were the only factors that were available for the entire time series of HDAR logbook data besides the CML, which changed each fishing year until 1993, the fishing gear, and the catch by species. For an individual deep handline trip, the multiplicative model took the form

$$
\begin{equation*}
C P U E_{i j k}=U_{R} \cdot \prod_{i} Y_{i}^{X_{y, i}} \prod_{j} A_{j}^{X_{a, j}} \prod_{k} M_{k}^{X_{m, k}} \cdot \exp \left(\varepsilon_{i j k}\right) \tag{1}
\end{equation*}
$$

where $U_{R}$ was the reference mean catch for a particular combination of year, area, and month, $Y_{i}$ were the year effect coefficients with indicator variable $X_{y, i}=1$ if the trip occurred in year i and $X_{y, i}=0$ otherwise, $A_{k}$ were the area effect coefficients with indicator variable $X_{a, j}=1$ if the trip occurred in area $j$ and $X_{a, j}=0$ otherwise, $\mathrm{M}_{\mathrm{k}}$ were the month effect coefficients with indicator variable $X_{m, k}=1$ if the trip occurred in month k and $\mathrm{X}_{\mathrm{m}, \mathrm{k}}=0$ otherwise, and the iid normal error term $\varepsilon_{i \mathrm{ijk}}$. In this case, the multiplicative model predicts CPUE for an individual trip as a mean catch rate times a proportional effect for year (interpreted as a relative abundance index for bottomfish), area, and month along with a multiplicative lognormally-distributed error term. The properties of this loglinear model are well established. In particular, if the catch rate is lognormally distributed then the estimator of catch rate determined by this model is the minimum variance unbiased estimator. The linear form of the model is typically used for estimation purposes

$$
\begin{equation*}
\ln \left(C P U E_{i j k}\right)=U_{R}+\sum_{i} X_{y, i} \ln \left(Y_{i}\right)+\sum_{j} X_{a, j} \ln \left(A_{j}\right)+\sum_{k} X_{m, k} \ln \left(M_{k}\right)+\varepsilon_{i j k} \tag{2}
\end{equation*}
$$

In this case, maximum likelihood estimates $(\theta)$ were calculated from the normal equations to solve this linear regression model

$$
\begin{equation*}
\theta=\left(X^{T} X\right)^{-1} X^{T} y \tag{3}
\end{equation*}
$$

using SAS Proc GLM (SAS, 1990), where $\theta$ is the vector of log-transformed parameter estimates, X is the design matrix of indicator variable values and y is the vector of logtransformed CPUE observations. Alternatively, model parameters could be estimated using iteratively reweighted least squares, as in a GLM. The estimated parameters within each factor were linearly dependent and treatment contrasts were used with the reference cell year equal to 1980 (the start of the standard catchability time period, see Fig. 3), the area equal to 331 (i.e., Penguin Bank), and the month equal to December. The choice of the reference cell was arbitrary and it did not affect the parameter estimates.

Back-transformed year coefficients with bias adjustment to account for the natural logarithmic transformation (Gavaris, 1980; eqn 7 with $g_{m}(t)=\exp (t)$ ) were used to compute the mean standardized CPUE in each year $t$ as

$$
\begin{equation*}
\operatorname{CPUE}(t)=U_{R} \cdot \exp \left(\ln \left(Y_{t}\right)+0.5\left(\sigma^{2}-\operatorname{Var}\left[\ln \left(Y_{t}\right)\right]\right)\right) \tag{4}
\end{equation*}
$$

where $\sigma^{2}$ was the residual variance from the regression and $\operatorname{Var}\left[\ln \left(\mathrm{Y}_{\mathrm{t}}\right)\right]$ was the variance of the log-scale year coefficient. The variance of $\operatorname{CPUE}(\mathrm{t})$ was computed from the associated expression for the unbiased estimator of variance (Gavaris, 1980; eqn 8 with $g_{m}(t)=\exp (t)$ )

$$
\begin{equation*}
\operatorname{Var}[\operatorname{CPUE}(t)]=U_{R}^{2} \cdot \exp \left(2 \ln \left(Y_{t}\right)\right)\left\{\exp \left(0.5\left(\sigma^{2}-\operatorname{Var}\left[\ln \left(Y_{t}\right)\right]\right)\right)^{2}-\exp \left(\sigma^{2}-2 \cdot \operatorname{Var}\left[\ln \left(Y_{t}\right)\right]\right)\right\} \tag{5}
\end{equation*}
$$

Bottomfish catches were log-transformed for the standardization analysis. The resulting logscale catch rate distribution appeared to be approximately normal (Fig. 10) although a hypothesis test would reject that assumption due to the large number of CPUE observations. The individual 3-digit areas used in the spatial standardization included the individual HDAR reporting areas that accounted for at least $1 \%$ of the total bottomfish catch reported during 1949-2007 fishing years along with the six aggregate areas (Fig. 5).

The fit of the GLM was highly significant ( $\mathrm{P}<0.0001$ ), in part due to the large number of CPUE observations ( $\mathrm{n} \approx 141,000$ ) included in the analysis (Appendix, Table A1). A diagnostic plot of residuals versus fitted CPUE did not indicate substantial model misspecification (Fig. A1) although the upper bound of 1500 pounds per trip was apparent. In particular, the CPUE residuals and fitted values were not correlated (Spearman rank correlation $\mathrm{R}=0.001, \mathrm{P}=0.59$ ). A diagnostic plot to assess whether the variance of CPUE changed as a function of the fitted value suggested there was a moderate increase in variance with increasing CPUE (Fig. A2). In comparison, the multiplicative lognormal error assumption of the GLM implied that variance in CPUE would be expected to increase in proportion to mean CPUE. A diagnostic plot to assess whether the three explanatory variables reduced the observed variability in CPUE suggested that some of the variability was being explained by the combination of the year, area, and
month effects (Fig. A3) but that there was a substantial amount of unexplained variation. Nonetheless, a linear regression of observed log-scale CPUE on fitted values was highly significant $(\mathrm{P}<0.001)$ with an estimated slope coefficient of $\mathrm{B}_{1}=1.000( \pm 0.006 \mathrm{SE})$ for fitted values. Overall, the GLM explained approximately $18 \%$ of the observed variation in the CPUE data. A Q-Q plot of the standardized CPUE residuals suggested that there was a moderate departure from normality (Fig. A4) and this was likely due to the tendency for variability in CPUE to increase slightly more than in proportion with mean CPUE. Last, a histogram of the standardized CPUE residuals from the fitted GLM suggested that the residuals were approximately normal (Fig. A5), although there was a moderate negative skew in the residual distribution. Overall, the GLM appeared to fit the CPUE data reasonably well, although there was some evidence of heteroscedastcity and non-normality of the residuals.

Standardized mean annual catch rates for MHI bottomfish exhibited a decreasing trend since the 1960s (Fig. 11). The standardized CPUE was significantly positively correlated with the status quo CPUE (Spearman $\mathrm{R}=0.75, \mathrm{P}<0.001$ ) but exhibited much less interannual variation (Fig. 11). The estimated area coefficients and associated mean CPUE values by fishing area indicated that there was substantial heterogeneity in the spatial pattern of bottomfish CPUE in the main Hawaiian Islands (Fig. 12). The estimated month coefficients and associated standardized CPUE values suggested that there were moderate but statistically significant seasonal differences in bottomfish CPUE in the main Hawaiian Islands (Fig. 13).

The GLM sensitivity analysis with a bottomfish ratio of $90 \%$ had a similar declining CPUE trend as the baseline model. However, the $90 \%$ ratio GLM provided a poorer overall fit to the selected data $\left(\mathrm{R}^{2}=0.135\right)$ and exhibited a slightly higher residual variance. Similarly, the GLM sensitivity analysis using all deep handline trips that caught at least 1 pound of bottomfish had a similar qualitative trend in CPUE but had a poorer fit $\left(\mathrm{R}^{2}=0.166\right)$ and had a higher residual variance than the baseline GLM.

## Assessment Model

In the previous assessment of the Hawaiian bottomfish complex, the status of the multispecies complex in each fishing zone was assessed using a Schaefer surplus production model (Moffitt et al., 2006). The previous assessment assumed that the intrinsic growth rate of Hawaiian bottomfish was equal across zones and a single intrinsic growth rate was estimated for all zones in the surplus production models. The previous assessment also assumed that the relative amount of bottomfish habitat in each zone was proportional to the linear extent of its 100fathom contour. In this case, the carrying capacity of the MHI was a freely estimated parameter while the carrying capacities of the Mau and Hoomalu zones were set based on the ratio of MHI habitat to Mau or Hoomalu habitat multiplied by the carrying capacity estimate for the MHI. Overall, the bottomfish surplus production model for the previous assessment included linkages between the intrinsic growth rates and carrying capacities estimated for each fishing zone. These linkages were maintained in the current assessment.

Status determination of the Archipelagic bottomfish management unit in the previous assessment was based on the habitat-weighted average of the model results by fishing zone.

That is, the status of the complex was assessed for the entire Hawaiian Archipelago as specified in the Fishery Management Plan. This status determination was based on the assumption that there was sufficient interchange among adult and larval fishes of the three zones to treat them as a single management unit. Status determination criteria in the current assessment were calculated in the same manner as in the previous assessment.

In the previous assessment, the estimation procedure was nonlinear least squares with no assumptions regarding the error structure for fitting parameters. As a result, no estimates of parameter uncertainty were available to quantify the risks associated with alternative fishery management actions. In the current assessment, a Bayesian statistical framework was applied to directly estimate production model parameters and parameter uncertainty for risk analyses.

## Production Model

The Schaefer production model for bottomfish was formulated as a Bayesian state space model with explicit process and observation error terms (see, for example, Meyer and Millar, 1999). The unobserved biomass states were estimated from the observed relative abundance indices (CPUE) and catches based on an observation error likelihood function and prior distributions for each model parameter $(\theta)$. In this case, the observation error likelihood measured the discrepancy between observed and model predictions of CPUE.

The process dynamics were based on a quadratic surplus production model with an annual time step. In this model, biomass in year $\mathrm{T}\left(B_{T}\right)$ depends on the previous biomass, catch $\left(C_{T-1}\right)$, the intrinsic growth rate parameter $(R)$, and the carrying capacity parameter ( $K$ ), for years $T=2, \ldots$, $N$.

$$
\begin{equation*}
B_{T}=B_{T-1}+R \cdot B_{T-1}\left(1-\left(\frac{B_{T-1}}{K}\right)\right)-C_{T-1} \tag{6}
\end{equation*}
$$

The values of biomass and harvest rate that maximize surplus production were relevant for fishery management under the Magnuson Fishery Conservation and Management Act as reauthorized in 1996. For the Schaefer model, the biomass that maximizes surplus production ( $B_{\mathrm{MSY}}$ ) was

$$
\begin{equation*}
B_{M S Y}=\frac{K}{2} \tag{7}
\end{equation*}
$$

The corresponding harvest rate that maximizes surplus production $\left(\mathrm{H}_{\mathrm{MSY}}\right)$ was

$$
\begin{equation*}
H_{M S Y}=\frac{R}{2} \tag{8}
\end{equation*}
$$

and the maximum surplus production (MSY) was

$$
\begin{equation*}
M S Y=\frac{R \cdot K}{4} \tag{9}
\end{equation*}
$$

The biomass dynamics were reparameterized in terms of the proportion of carrying capacity ( $P$ $=B / K)$ to improve the efficiency of the Markov Chain Monte Carlo sampling algorithm. Based on this parameterization, the process dynamics became

$$
\begin{equation*}
P_{T}=P_{T-1}+R \cdot P_{T-1}\left(1-P_{T-1}\right)-\frac{C_{T-1}}{K} \tag{10}
\end{equation*}
$$

Annual fluctuations in life history parameters, trophic interactions, environmental conditions and other factors were expected to change the process dynamics through an annual process error term. In this context, the process error represented the joint effect of a large number of random multiplicative events which combined to form a multiplicative lognormal random variable under the Central Limit Theorem. Given this, the process error terms were independent and lognormally distributed random variables $\eta_{T}=e^{U_{T}}$ where the $U_{T}$ are normal random variables with mean 0 and variance $\sigma^{2}$.

State equations described how the stochastic process dynamics related the unobserved biomass states to the observed catches and the model parameters. Given the lognormal process error assumption, the state equations for the initial time period $T=1$ and subsequent periods $T>1$ of the Schaefer model were

$$
\begin{align*}
& P_{1}=\eta_{1} \\
& P_{T}=\left(P_{T-1}+R \cdot P_{T-1}\left(1-P_{T-1}\right)-\frac{C_{T-1}}{K}\right) \cdot \eta_{T} \tag{11}
\end{align*}
$$

These equations were used to set the prior distribution for the proportion of carrying capacity, $p\left(P_{T}\right)$, in each time period $T$, conditioned on the previous proportion.

## Changes in Catchability

In the previous assessment, trends in relative abundance were based solely on fishery catch-per-unit-effort (CPUE) data. To address the fact that fishing technology used to locate and capture bottomfish has improved over time (e.g., Moffitt et al., 2009), it was assumed that the fishery catchability coefficient Q increased in each of four successive time periods with catchability being constant in each period (Fig. 3). In the previous assessment, the relative catchabilities of the four periods were set based on limited field observations, anecdotal knowledge, and subjective judgment. The fishery catchability coefficient for the standard time period ( $\mathrm{Q}_{\text {STD }}$ ) during 1980-1991 was a freely estimated parameter in the production and the other values were set based on their relative catchabilities (c). The relative catchabilities during 1949-1967, 19681979, 1980-1991, and 1992-2007 were $c=0.7, c=0.8, c=1$, and $c=1.2$, respectively, and the
resulting catchability coefficients were calculated as $\mathrm{Q}=\mathrm{c} \cdot \mathrm{Q}_{\text {STD }}$. Thus, the current assessment applied the same approach to dealing with long-term improvements in fishing technology as was used in the previous assessment.

## Likelihood Model

The likelihood model related the observed standardized fishery CPUE to the exploitable bottomfish biomass. It was assumed that the CPUE index (I) was proportional to biomass with catchability coefficient $Q$

$$
\begin{equation*}
I_{T}=Q B_{T}=Q K P_{T} \tag{12}
\end{equation*}
$$

The CPUE values were observed with sampling error and it was assumed that the observation errors were lognormally distributed. In particular, the observation errors were $v_{T}=e^{V_{T}}$ where the $V_{T}$ were iid normal random variables with zero mean and variance $\tau^{2}$. Given these assumptions, the observation equations for the state space model in years $T=1, \ldots, N$ were

$$
\begin{equation*}
I_{T}=Q K P_{T} \cdot v_{T} \tag{13}
\end{equation*}
$$

This equation specified the likelihood function $p\left(I_{T} \mid \theta\right)$ for each period.

## Prior Distributions

Under the Bayesian paradigm, prior distributions were employed to quantify existing knowledge (or the lack thereof) of the likely value of model parameters and the unobserved biomass states. In this context, the model parameters were the carrying capacity parameter, intrinsic growth rate parameter, catchability parameter, the process and observation error variance parameters, and the initial biomass as a proportion of carrying capacity parameters. Unobserved biomass states were estimated from the proportion of carrying capacity, conditioned on the previous proportion, and the catchability parameter.

## Prior for Carrying Capacity

The prior distribution for the carrying capacity $p(K)$ of MHI was chosen to be a diffuse normal distribution with mean $\left(\mu_{K}\right)$ and variance $\left(\sigma_{K}^{2}\right)$ parameters

$$
\begin{equation*}
p(K)=\frac{1}{\sqrt{2 \pi} \sigma_{K}} \exp \left(-\frac{\left(K-\mu_{K}\right)^{2}}{2 \sigma_{K}^{2}}\right) \tag{14}
\end{equation*}
$$

The mean parameter was set to be 2 million pounds based on the numerical scale of biomass estimates in Moffitt et al. (2006).The variance parameter was set to be 100 million pounds to
allow for a wide range of probable carrying capacity values. In effect, this was an uninformative prior for $K$. The values of carrying capacity for the Mau and Hoomalu zones were set based on the estimate of $K$ and the relative habitat weights for the three bottomfish management zones $\left(W_{\mathrm{MHI}}=0.447, W_{\text {Mau }}=0.124, \mathrm{~W}_{\text {Hoomalu }}=0.429\right)$ as in the previous assessment (Moffitt et al., 2006). Thus, the carrying capacities of the Mau and Hoomalu zones calculated as a fraction of the habitat weighted carrying capacity of the main Hawaiian Islands were 0.277 K and 0.960 K , respectively.

## Prior for Intrinsic Growth Rate

The prior distribution for intrinsic growth rate $p(R)$ was a beta distribution with parameters $c$ and $d$ :

$$
\begin{equation*}
p(R)=\frac{\Gamma(c+d)}{\Gamma(c) \Gamma(d)} \cdot x^{c-1}(1-x)^{d-1} \tag{15}
\end{equation*}
$$

This choice constrained the intrinsic growth rate parameter to be within the interval $[0,1]$. The mean of the prior for R was set to be $\mu_{\mathrm{R}}=0.46$. This choice was based on the estimate from Moffitt et al. (2006) of $R=0.46$. This range was consistent with the fact that the bottomfish complex included snappers with relatively high values of natural mortality ( $M>0.25$ ) and moderate values of Brody growth coefficients (Martinez-Andrade, 2003).Values of the beta distribution parameters were set to $c=7.67$ and $d=9$ to set the coefficient of variation (CV) of R to be $26 \%$. The sensitivity of model results to the choice of the informative prior distribution for $R$ was examined below.

## Prior for Catchability

The prior for the inverse of catchability $p\left(Q^{-1}\right)$ was chosen to be a diffuse gamma distribution with scale parameter $\lambda$ and shape parameter $k$. This prior was used to approximate an improper prior for catchability $p(Q)=Q^{-1}$ (see Meyer and Millar, 1999). In particular the prior for the inverse of Q was

$$
\begin{equation*}
p\left(Q^{-1}\right)=\frac{\lambda^{k} Q^{-(k-1)}}{\Gamma(k)} \exp \left(-\lambda Q^{-1}\right) \tag{16}
\end{equation*}
$$

The scale and shape parameters were set to be $\lambda=k=0.001$. This choice of parameters implied that $1 / Q$ had a mean of 1 and a variance of 1000 . As a result, the prior for catchability was an inverse gamma distribution that was approximately proportional to $1 / \mathrm{Q}$, e.g., $p(Q) \propto Q^{-1}$. Since $1 / Q$ is unbounded at $Q=0$, an additional numerical constraint that $Q$ lie within the interval $[0.0001,10]$ was imposed.

## Priors for Error Variances

Priors for the process error variance $p\left(\sigma^{2}\right)$ and observation error variance $p\left(\tau^{2}\right)$ were also chosen to be inverse-gamma distributions, a natural choice for dispersion priors (Congdon, 2001). For the process error variance prior, the scale parameter was set to $\lambda=4$ and the shape parameter was $\mathrm{k}=0.01$. This choice of parameters produced an $80 \%$ confidence interval of approximately [ $0.04,0.08]$ for $\sigma$. Similarly, for the observation error variance prior, the scale parameter was set to $\lambda=2$ and the shape parameter was $\mathrm{k}=0.01$. This choice of parameters gave an $80 \%$ confidence interval of approximately $[0.05,0.14]$ for $\tau$. The ratio of the observation error prior mean to the process error prior mean was $\mathrm{E}[\tau] / \mathrm{E}[\sigma]=0.0707 / 0.05 \approx 1.41$. Thus, the observation error variance was assumed to be about $40 \%$ greater than the process error variance. The sensitivity of model results to this assumption was examined below.

## Priors for Proportions of Carrying Capacity

Prior distributions for the time series of biomass in proportion to carrying capacity, $p\left(P_{T}\right)$, were determined by the lognormal distributions and the process dynamics. The prior mean for the initial proportion of carrying capacity in the main Hawaiian Islands in 1949 ( $\mathrm{P}_{\mathrm{MHI}, 1}$ ) was set to be 0.6 based on the minimum root-mean square error of the fit to the MHI CPUE over alternative values of $\mathrm{P}_{\mathrm{MHI}, 1}$ from 0.1 to 1.0 (Fig. A.6). Prior means for the initial proportions of carrying capacity in the Mau and Hoomalu zones in 1988 were set to 0.8 .

## Posterior Distribution

The posterior distribution was numerically sampled to make inferences about model parameters. From Bayes' theorem, the posterior distribution given catch and CPUE data $D, p(\theta \mid D)$, was proportional to the product of the priors and the likelihood of the CPUE data.

$$
\begin{equation*}
p(\theta \mid D) \propto p(K) p(R) p(M) p(Q) p\left(\sigma^{2}\right) p\left(\tau^{2}\right) \prod_{T=1}^{N} p\left(P_{T}\right) \prod_{T=1}^{N} p\left(I_{T} \mid \theta\right) \tag{17}
\end{equation*}
$$

There was no analytic solution to determine parameter estimates from the posterior distribution in (17).

Under the Bayesian paradigm, parameter estimation for nonlinear models like the bottomfish production model is typically based on simulating a large number of independent samples from the posterior distribution. In this case, Markov Chain Monte Carlo (MCMC) simulation (Gilks et al., 1996) was applied to numerically generate a sequence of samples from the posterior distribution. The WINBUGS software (Lunn et al., 2000; Spiegelhalter et al., 2003) was used to program the model, set the initial conditions, perform the MCMC calculations, and summarize the production model results (Appendix, Table A2). The baseline production model used the standardized CPUE in the MHI along with Mau and Hoomalu CPUE series to estimate model parameters.

MCMC simulations were conducted in an identical manner for the baseline model and sensitivity analyses described below. Two chains of 210,000 samples from the posterior distribution were simulated in each model run. The first 60,000 samples of each chain were excluded from the estimation process. This burn-in period removed dependence of the MCMC samples on the initial conditions. Next, each chain was thinned by 3 to remove autocorrelation, e.g., every third sample from the posterior distribution was used for inference. As a result, 100,000 samples from the posterior distribution were used to summarize model results. Convergence of the MCMC chains to the posterior distribution was checked using the Brooks-Gelman-Rubin (BGR) convergence diagnostic (Brooks and Gelman, 1998). This diagnostic assessed convergence by comparing the variance ratio (V) of within- to between-chain variability and was monitored for key model parameters (intrinsic growth rate, carrying capacity, catchability, initial proportion of carrying capacity) and the root-mean squared error (RMSE). Values of the variance ratio V that were close to unity and no greater than 1.2 indicated convergence was likely to have occurred (Gelman et al., 1995).

Model residuals were used to measure the goodness of fit of the production model. Residuals for the CPUE series were the $\log$-scale observation errors $\varepsilon_{T}$.

$$
\begin{equation*}
\varepsilon_{T}=\ln \left(I_{T}\right)-\ln \left(Q K P_{T}\right) \tag{18}
\end{equation*}
$$

Non-random patterns in the residuals indicated that the observed CPUE did not conform to one or more model assumptions. The RMSE of the CPUE fit provided another diagnostic of the model goodness of fit with lower RMSE indicating a better fit.

Production model results included the status of the entire Archipelagic management unit as well as the individual fishing zones relative to MSY reference points. Time series of the relative harvest rate (for example, in 2007 the relative harvest rate is the ratio $\mathrm{H}_{2007} / \mathrm{H}_{\text {MSY }}$ ) and relative biomass (e.g., the ratio $\mathrm{B}_{2007} / \mathrm{B}_{\mathrm{MSY}}$ ) were calculated for the Archipelagic unit and individual zones using the mean values from the joint posterior distribution of model parameters.

## Sensitivity Analyses

The sensitivity of model biomass estimates to the CPUE index was evaluated by fitting the production model using the status quo CPUE as the relative abundance index for the MHI with all other input data and assumptions remaining the same. This sensitivity analysis addressed the question of whether model results were sensitive to the use of the status quo CPUE series instead of the standardized CPUE series.

The sensitivity of model results to the prior mean for intrinsic growth rate was also evaluated by fitting the model using different prior means for R . In this case, the prior mean for R ranged from 0.25 to 0.75 in increments of 0.05 . This sensitivity analysis addressed whether the choice of a prior mean for $\mathrm{R}=0.46$ had a strong influence on model results.

Similarly, the sensitivity of model results to the prior mean for carrying capacity was also evaluated by fitting the model using different prior means for K. For this analysis, the prior mean for K ranged from 1500 to 2500 klbs in increments of 100 klbs . This sensitivity analysis addressed whether the choice of a prior mean for $\mathrm{K}=2000$ had a strong influence on model results.

Sensitivity analyses to the prior means for the observation and process error terms were conducted by fitting the model using different prior means for $\tau^{2}$ and $\sigma^{2}$. For the observation error variance, the prior mean for $\tau^{2}$ was multiplied by $10,100,1000$, and 10000 , and the production model was fit in each of the four cases with all other inputs and assumptions remaining the same. Similarly, for the process error variance, the prior mean for $\sigma^{2}$ was multiplied by $10,100,1000$, and 10000 , and the production model was fit in each case with all other inputs and assumptions constant. These analyses addressed the question of whether the choice of the prior mean for the observation or process error had a strong influence on model results.

## Projections for 2008-2010

Bottomfish biomass, catch, and relative biomass were projected for fishing years 2008-2010 under alternative fishing mortality and catch assumptions for the main Hawaiian Islands. In these projections, status quo fishing mortality rates were assumed for the Mau and Hoomalu zones in 2008. With the exception of the $\mathrm{F}_{\text {MSY }}$ projection, fishing mortality rates for the Hoomalu and Mau zones in fishing years 2009-2010 were also set equal to the mean of the 2007 harvest rate estimate. These status quo rates were $\mathrm{H}_{\mathrm{SQ}, \mathrm{Mau}}=0.03$, and $\mathrm{H}_{\mathrm{SQ}, \mathrm{Hoomalu}}=0.07$. In comparison, the mean estimate of $\mathrm{H}_{\mathrm{MSY}}$ for both zones was $\mathrm{H}_{\mathrm{MSY}}=0.29$. In the MHI the bottomfish catch in fishing year 2008 was estimated using an estimate of the total Deep 7 bottomfish catch in FY 2008, which was C DEEP7, $2008=192,614 \mathrm{lbs}$, and the average ratio of Deep 7 to total bottomfish catch in 2005-2007, which was $\mathrm{R}_{\text {DEEP } 7,2005-2007}=0.723$ (Table A3). The resulting estimate of total MHI bottomfish catch in 2008 was $\mathrm{C}_{2008}=266,409 \mathrm{lbs}$. This 2008 catch was used to estimate the MHI fishing mortality rate in fishing year 2008 in each of the alternative projections.

Three sets of projections were conducted using alternative fishing mortality or catch assumptions for the MHI. The first projection was the $\mathrm{F}_{\text {MSY }}$ scenario which provided an estimate of the upper bound of catch for each of the zones. In the $\mathrm{F}_{\text {MSY }}$ scenario, harvest rates in the MHI, Mau, and Hoomalu zones were set equal to $\mathrm{F}_{\text {MSY }}$ in 2009-2010.

The second projection was the status quo fishing mortality scenario. In this projection, fishing mortality rates during 2009-2010 in the main Hawaiian Island, Mau zone, and Hoomalu zone were set equal to the estimate of the mean harvest rate in 2008.

The third projection was the constant total allowable catch for the main Hawaiian Islands scenario in fishing year 2009-2010. For this scenario, the set of median catches of MHI bottomfish in 2009 that would produce probabilities of overfishing in the MHI ranging from
$0 \%$ to $100 \%$ by $5 \%$ intervals were calculated. These TACs were then applied to the MHI in 2009-2010 while status quo fishing mortality rates were assumed in the Mau and Hoomalu zones.

## RESULTS

In this section, production model results are described. These include: model diagnostics, biomass and fishing mortality estimates to assess stock status, sensitivity analyses, and projection analyses.

## Model Diagnostics

Model residual diagnostics indicated that the Bayesian production model provided a good fit to the bottomfish CPUE time series from the MHI (Fig. 14), the Mau zone (Fig. 15) and the Hoomalu zone (Fig. 16). The BGR convergence diagnostic indicated that the MCMC chains had converged after the burn-in period for all the monitored parameters including: MHI carrying capacity $\mathrm{V}=1.08$; initial proportion of carrying capacity for MHI $\mathrm{V}=0.98$, for Mau $\mathrm{V}=1.01$, and for Hoomalu $\mathrm{V}=1.00$; catchability for MHI $\mathrm{V}=1.01$, for Mau $\mathrm{V}=1.00$, and for Hoomalu $V=1.01$; and for intrinsic growth rate $\mathrm{V}=1.04$. Overall, the residual and convergence diagnostics indicated that there were no problems with convergence of the baseline model.

## Stock Status

Bottomfish biomass and harvest rate estimates and MSY-based reference points from the production model were summarized for the Archipelago and the MHI, Mau, and Hoomalu zones (Tables 3, 4, and 5 and Table A4). Estimates of biological reference points for bottomfish were similar to the estimates from the previous assessment (Moffitt et al., 2006). Relative biomass and harvest rate estimates for the Archipelagic bottomfish stock indicated that the stock was not overfished $\left(\mathrm{B}_{2007} / \mathrm{B}_{\mathrm{MSY}}=1.13\right.$, Fig. 17) and was not currently experiencing overfishing $\left(\mathrm{H}_{2007} / \mathrm{H}_{\mathrm{MSY}}=0.62\right)$. In fishing year 2007, there was a $97 \%$ probability that Archipelagic biomass exceeded $\mathrm{B}_{\mathrm{MSY}}$ and a $0 \%$ chance that the harvest rate exceeded $\mathrm{H}_{\mathrm{MSY}}$. Biomass of the Archipelagic stock declined from roughly $150 \%$ of $B_{\text {MSY }}$ in the late-1980s to slightly above $\mathrm{B}_{\mathrm{MSY}}$ in recent years. Harvest rates for the Archipelagic stock declined from roughly $90 \%-100 \%$ of $\mathrm{H}_{\text {MSY }}$ in the late-1980s to range from $60 \%-80 \%$ of $\mathrm{H}_{\text {MSY }}$ since 2000.

Results for the Hoomalu zone bottomfish stock indicated that the stock was not overfished $\left(\mathrm{B}_{2007} / \mathrm{B}_{\mathrm{MSY}}=1.54\right.$, Table 3 and Fig. 18) and was not currently experiencing overfishing $\left(\mathrm{H}_{2007} / \mathrm{H}_{\mathrm{MSY}}=0.25\right.$, Table 4 and Fig. 18). In fishing year 2007, there was a $99.9 \%$ probability that Hoomalu biomass exceeded $\mathrm{B}_{\mathrm{MSY}}$ and a $0 \%$ chance that the harvest rate exceeded $\mathrm{H}_{\mathrm{MSY}}$. Biomass in the Hoomalu zone declined from about $200 \%$ of $\mathrm{B}_{\mathrm{MSY}}$ in the late-1980s to range from $120 \%$ to $150 \%$ of $\mathrm{B}_{\mathrm{MSY}}$ since 2000. Relative harvest rates in the Hoomalu zone fluctuated from $10 \%$ to $50 \%$ of $\mathrm{H}_{\text {MSY }}$ since 1988.

Similarly, results for the Mau zone bottomfish stock indicated that the stock was not overfished $\left(\mathrm{B}_{2007} / \mathrm{B}_{\mathrm{MSY}}=1.55\right.$, Table 3 and Fig. 19) and was not currently experiencing overfishing $\left(\mathrm{H}_{2007} / \mathrm{H}_{\mathrm{MSY}}=0.12\right.$, Table 4 and Fig. 19). In fishing year 2007, there was a $99.9 \%$ probability that Mau biomass exceeded $\mathrm{B}_{\mathrm{MSY}}$ and a $0 \%$ chance that the harvest rate exceeded $\mathrm{H}_{\mathrm{MSY}}$. Biomass in the Mau zone declined from $180 \%$ of $\mathrm{B}_{\text {MSY }}$ in the late 1980 s to about $120 \%$ of $\mathrm{B}_{\text {MSY }}$ in the late 1990s and has increased to $150 \%-160 \%$ of BMSY since 2000. Harvest rates in the Mau zone exceeded $\mathrm{H}_{\text {MSY }}$ in 1991 but have since declined to range from roughly $10 \%$ to $60 \%$ of $\mathrm{H}_{\mathrm{MSY}}$ since 2000.

In contrast, results for the main Hawaiian Islands bottomfish stock indicated that the stock was depleted ( $\mathrm{B}_{2007} / \mathrm{B}_{\mathrm{MSY}}=0.62$, Table 3 and Fig. 20) and was currently experiencing overfishing $\left(\mathrm{H}_{2007} / \mathrm{H}_{\mathrm{MSY}}=1.11\right.$, Table 4 and Fig. 20). In fishing year 2007, there was a $0 \%$ probability that MHI biomass exceeded $\mathrm{B}_{\text {MSY }}$ and an $87 \%$ chance that the harvest rate exceeded $\mathrm{H}_{\text {MSY }}$. Bottomfish biomass exhibited a long-term decline from high values in the 1960s-1970s to relatively low values since the mid-1990s (Fig. 21). Biomass in the MHI zone declined from roughly $\mathrm{B}_{\text {MSY }}$ in the late-1980s to about $60 \%$ of $\mathrm{B}_{\text {MSY }}$ since 2000 . Harvest rates were relatively low in the 1960s to mid-1970s, increased to peak in 1989, and have declined gradually since then (Fig. 22). Harvest rates in the MHI declined from roughly $200 \%$ of $\mathrm{H}_{\text {MSY }}$ in the late-1980s to range from $94 \%$ to $140 \%$ of $\mathrm{H}_{\text {MSY }}$ since 2000.

## Sensitivity Analyses

The sensitivity analysis on the effect of using the status quo MHI CPUE series indicated that using the status quo CPUE series instead of the standardized CPUE series would have decreased estimates of relative biomass and increased estimates of relative harvest rate across zones (Table 6). In particular, the status quo CPUE indicated a much more abrupt decline in bottomfish biomass over time (Fig. 11) decreasing from an average of $462 \mathrm{lbs} /$ trip in the 1950s and 1960s to roughly $181 \mathrm{lbs} /$ trip in the 2000s ( $-61 \%$ ). In comparison, the standardized CPUE declined from an average of $304 \mathrm{lbs} /$ trip in the 1950s-1960s to an average of $170 \mathrm{lbs} /$ trip since $2000(-44 \%)$. As a result, the Archipelagic bottomfish stock would be estimated to be approaching an overfished condition in 2007 and would be estimated to have been experiencing overfishing since 1988 (Table 6, Fig. 23) if relative abundance was assumed to be proportional to status quo CPUE.

The sensitivity analysis on the effect of using different prior means for the intrinsic growth rate parameter showed that changing the mean R would not have much of an effect on estimates of relative biomass in the MHI (Fig. 24). In comparison to the mean $\mathrm{R}=0.46$ used in the assessment, changing the mean value from $\mathrm{R}=0.25(-46 \%)$ to $\mathrm{R}=0.75$ ( $+63 \%$ ) did not alter the trend in relative biomass. Overall, assessment results did not appear to be sensitive to the prior mean for intrinsic growth rate.

The sensitivity analysis to examine the effect of using different prior means for carrying capacity in the MHI indicated that there was no practical impact for the range of K investigated (Fig. 25). This suggested that there was sufficient information in the CPUE and fishery catch
trends to estimate K. Overall, the assessment results were not sensitive to the prior mean for carrying capacity.

The sensitivity analysis on the effect of increasing the mean of the prior for the observation error variance suggested that a 10-fold increase in observation error would have no practical effect (Fig. 26) while an increase of 100 -fold in the observation error variance would have a moderate effect on relative MHI biomass. Increases on the order of 1000 -fold or more would have an important effect on the estimated trends in relative biomass and in particular, would produce a sharply increasing trend in recent years. Overall, the assessment results appeared to be sensitive to increases of 1000 -fold or more in the mean of the observation error variance prior.

The sensitivity analysis on the effect of increasing the process error variance indicated that increasing the process error variance would tend to increase the relative MHI biomass (Fig. 27). A 10 -fold increase in the process error mean produced a nearly identical trend in relative biomass as estimated in the assessment and a 100 -fold increase produced a similar trend. Increases in process error on the order of 1000 -fold to 10000 -fold increased the relative biomass in the MHI but did not affect the estimated trend. Overall, results were moderately sensitive to the mean of the prior for process error variance.

## Projections

Under the $\mathrm{F}_{\text {MSY }}$ scenario, projected bottomfish catches, probabilities of overfishing, and relative biomasses (Table 7) showed the probable distribution of outcomes if fishing mortality was set equal to the overfishing threshold in 2009-2010. In this case, fishing mortality was set to the overfishing threshold in each fishing zone. The resulting yields in 2009-2010 were higher in the Mau and Hoomalu zones because the current fishing mortality in these zones was much lower than $\mathrm{F}_{\mathrm{MSY}}$ (Table 4). Relative biomass of the Archipelagic stock would be projected to decrease by 2010 under this scenario but would still exceed $\mathrm{B}_{\mathrm{MSY}}$. In comparison, relative biomass in the MHI would be projected to increase moderately by 2010 but would still remain below the estimate of $\mathrm{B}_{\mathrm{MSY}}$.

Under the status quo fishing effort scenario, projected bottomfish catches, probabilities of overfishing, and relative biomasses (Table 8) indicated the probable distribution of outcomes if fishing effort and mortality rates in the Hoomalu and Mau zones remained at fishing year 2007 levels in 2009-2010 while fishing mortality in the MHI in 2009-2010 was equal to the value in 2008 as estimated from the MHI catch in 2008. The status quo results led to a $0 \%$ probability of overfishing the Archipelagic stock in 2009-2010 (Table 8). The catch biomasses in the Hoomalu zone were projected to increase slightly (Fig. 28) under the status quo scenario while catch in the Mau zone would remain relatively constant (Fig. 29). The MHI catch biomass was projected to increase in 2009-2010 (Fig. 30) following a decrease in 2008. Stock biomasses in the Hoomalu and Mau zones were projected to increase moderately in 2009-2010 (Figs. 31 and 32). Similarly, biomass in the MHI was projected to increase under the $\mathrm{F}_{2008}$ fishing mortality scenario (Fig. 33) although the probability that MHI biomass exceeded $\mathrm{B}_{\text {MSY }}$ in 2010 would be relatively low (2\%).

Under the constant TAC scenarios, projected probabilities of overfishing, relative biomasses, and probabilities of depletion of MHI bottomfish (Table 9) showed the distribution of outcomes that would likely occur if constant TACs were applied in the MHI during 2009-2010. The probability of overfishing the MHI bottomfish stock as a function of the Deep 7 bottomfish TAC in 2009 showed that there was a range of Deep 7 bottomfish TAC values that would produce a less than $50 \%$ chance of overfishing in the MHI (Fig. 34). Results of the stochastic projections indicated that the largest Deep $7 \mathrm{TAC}_{2009}$ that would produce approximately $0 \%$ chance of overfishing in 2009 (i.e., exceeding $\mathrm{F}_{\mathrm{MSY}}$ ) was 172 thousand pounds. For comparison, the smallest Deep 7 TAC that would lead to a roughly $100 \%$ chance of overfishing the Archipelagic stock was 343 klb . Total allowable commercial catches of Deep 7 bottomfish in 2009 ranging from 172 to 249 thousand pounds corresponded to risks of Archipelagic overfishing ranging from $0 \%$ to $50 \%$. The Deep 7 TAC to achieve a low risk of overfishing $(25 \%)$ in 2009 was estimated to be $\mathrm{TAC}_{25 \%}=227 \mathrm{klb}$ and the Deep 7 TAC to achieve a neutral risk of overfishing ( $50 \%$ ) in 2008 was estimated to be TAC $_{50 \%}=249 \mathrm{klb}$. Sensitivity analyses of a similar projection model suggested that the estimates of overfishing risk would be sensitive to the estimates of biomass, intrinsic growth rate, and the proportion of Deep 7 bottomfish in the catch (Brodziak, 2008). In contrast, estimates of overfishing risk were unlikely to be sensitive to the estimate of carrying capacity, the assumed bottomfish catches in 2007, and the coefficients of variation of these model parameters.

## SUMMARY

Assessment results indicated that the Archipelagic bottomfish management unit was not overfished during 1988-2007 but had experienced overfishing in 1989 when the record bottomfish catch of roughly 1.144 million pounds was harvested. Archipelagic biomass in 2007 was estimated to be about $13 \%$ above $\mathrm{B}_{\text {MSY }}$ with a harvest rate of roughly $60 \%$ of $\mathrm{F}_{\text {MSY. }}$. Thus, the Archipelagic stock was not overfished and was not experiencing overfishing in 2007.

Results for bottomfish in the Hoomalu and Mau fishing zones were generally similar. Bottomfish biomass in both zones ranged from roughly $120 \%$ to $200 \%$ of $\mathrm{B}_{\mathrm{MSY}}$ since 1988 . Current biomass in both zones was estimated to be about $150 \%$ of $\mathrm{B}_{\mathrm{MSY}}$ in 2007. Harvest rates in the Hoomalu zone ranged from $12 \%$ to $49 \%$ of $\mathrm{F}_{\text {MSY }}$ since 1988. In comparison, harvest rates in the Mau zone were more variable and ranged from $12 \%$ to $102 \%$ of $\mathrm{F}_{\text {MSY }}$ during 1988-2007. The 2007 harvest rate in the Hoomalu zone was estimated to be about $25 \%$ of $\mathrm{F}_{\text {MSY }}$ while in the Mau zone, the current harvest rate was $12 \%$ of $\mathrm{F}_{\text {MSY }}$. Thus, both Hoomalu and Mau zone bottomfish stocks were not overfished and not experiencing overfishing in 2007.

For the main Hawaiian Islands bottomfish, assessment results indicated a long-term decline in exploitable biomass from high levels in the 1950s-1960s to relatively low levels since the 1990s. In the late 1980s, bottomfish biomass in the MHI was roughly at $\mathrm{B}_{\text {MSY }}$. Since then, biomass has declined and has fluctuated around $60 \%$ of $\mathrm{B}_{\mathrm{MSY}}$ since the mid-1990s. Current MHI biomass was estimated to be $62 \%$ of $\mathrm{B}_{\text {MSY }}$ in 2007. Bottomfish harvest rates in the MHI were relatively low in the 1960 s at roughly $50 \%$ of $\mathrm{F}_{\text {MSY }}$. MHI harvest rates increased in the late-1970s and peaked at roughly $200 \%$ of $\mathrm{F}_{\text {MSY }}$ in the late-1980s. Since then, MHI bottomfish
harvest rates have declined and have ranged from $94 \%$ to $140 \%$ of $\mathrm{F}_{\mathrm{MSY}}$ since 2000. The current MHI harvest rate was $111 \%$ of $\mathrm{F}_{\text {MSY }}$ in 2007. Thus, the MHI bottomfish stock was overfished and was experiencing overfishing in 2007.

Projection results indicated that it was highly unlikely that the Archipelagic management unit would be overfished or experience overfishing at current levels of fishing effort. Similarly, bottomfish biomass in the Hoomalu and Mau zones would be likely to remain above $\mathrm{B}_{\text {MSY }}$ in 2009-2010 at current levels of fishing effort. Projection results for the main Hawaiian Islands indicated that a range of total allowable catches of Deep 7 bottomfish in the MHI of roughly 172 klb to 249 klb in 2009 would produce probabilities of overfishing in the MHI ranging from $0 \%$ to $50 \%$. Applying a TAC from this range in both 2009 and 2010 would be projected to lead to a probability of $10 \%$ or less that MHI bottomfish would be considered to be overfished in 2010.

The primary difference between the results of the current and previous bottomfish assessments was the use of the standardized CPUE to index the relative abundance of bottomfish in the main Hawaiian Islands. In this case, the standardized CPUE index was developed using a larger fraction of directed bottomfish catches ( $85 \%$ ) than the CPUE calculations in the previous assessment ( $25 \%$ ). The standardized CPUE index also incorporated a larger subset of deep handline fishing trips (73\%) than the previous approach (10\%). Thus, the standardized CPUE index was based on a fuller coverage of the directed bottomfish fishery in the MHI.

In the future, assessments of individual bottomfish species will likely be considered. However, such attempts to assess individual bottomfish species will be constrained by the amount and quality of fishery data. If fishery-dependent and fishery-independent data collection systems for Hawaiian bottomfish were augmented, age- or length-structured assessment models could be more readily applied to assess individual bottomfish species. In this context, one priority would be to sample the recreational fishery to estimate total recreational catch as well as catch at length by species. Collection of bottomfish length composition and length-weight data from the Honolulu fish auction could provide valuable information on commercial fishery catches and some auction sampling was conducted in 2008. Such fishery-dependent data could be used to evaluate the catch at age by species given sufficient age-length keys and ongoing sampling effort. Last, the development of a consistent fishery-independent survey in the main Hawaiian Islands and Northwestern Hawaiian Islands would greatly enhance the capacity to assess and to effectively manage the bottomfish resources.

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Table 1.--List of Hawaiian bottomfish species used for reporting fishery catch and CPUE and those included in production model analyses in the current bottomfish assessment update and in the previous stock assessment (Moffitt et al., 2006).

| Common name | Local name | Scientific name | Deep 7 <br> species | Species included <br> in assessment |
| :--- | :--- | :--- | :---: | :---: |
| Pink snapper | Opakapaka | Pristipomoides filamentosus | X | X |
| Longtail snapper | Onaga | Etelis coruscans | X | X |
| Squirrelfish snapper | Ehu | Etelis carbunculus | X | X |
| Sea bass | Hapuupuu | Epinephelus quernus | X | X |
| Grey jobfish | Uku | Aprion virescens |  | X |
| Snapper | Gindai | Pristipomoides zonatus | X | X |
| Snapper | Kalekale | Pristipomoides seiboldii | X | X |
| Blue stripe snapper | Taape | Lutjanus kasmira |  |  |
| Yellowtail snapper | Yellowtail kalekale | Pristipomoides auricilla |  | X |
| Silver jaw jobfish | Lehi | Aphareus rutilans | X | X |
| Amberjack | Kahala | Seriola dumerili |  |  |
| Thick lipped trevally | Butaguchi | Pseudocaranx dentex |  | X |
| Giant trevally | White ulua | Caranx ignobilis |  | X |
| Black jack | Black ulua | Caranx lugubris | X |  |
| Armorhead |  | Pseudopentaceros wheeleri |  |  |

Table 2.--Total Hawaiian bottomfish catch used in the assessment update by zone and fishing year and fishery zone. As in the previous assessment, fishery catches in the Mau and Hoomalu zones prior to 1988 are not used in the assessment update because there are no CPUE data available for these zones prior to 1988 and the initial proportion of carrying capacity in 1988 was an estimated parameter for the Mau and Hoomalu zones.

|  | Main Hawaiian Is lands |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Fishing Year | Bottom fish Catch Used in this <br> Asses sment ( 000 lbs ) | Mau Zone Bottomfish Catch Used in this Assessment (000 lbs) | $\begin{aligned} & \text { Hoomalu Zone } \\ & \text { Bottom fish Catch } \\ & \text { Used in this } \\ & \text { Assessment } \\ & (000 \mathrm{lbs}) \end{aligned}$ | Total Catch by Fishing Year during 19882007 (000 lbs) |
| 1949 | 512.812 |  |  |  |
| 1950 | 431.817 |  |  |  |
| 1951 | 416.819 |  |  |  |
| 1952 | 389.113 |  |  |  |
| 1953 | 375.470 |  |  |  |
| 1954 | 370.070 |  |  |  |
| 1955 | 318.004 |  |  |  |
| 1956 | 382.184 |  |  |  |
| 1957 | 434.718 |  |  |  |
| 1958 | 312.884 |  |  |  |
| 1959 | 293.832 |  |  |  |
| 1960 | 226.944 |  |  |  |
| 1961 | 189.962 |  |  |  |
| 1962 | 237.813 |  |  |  |
| 1963 | 299.509 |  |  |  |
| 1964 | 307.018 |  |  |  |
| 1965 | 317.130 |  |  |  |
| 1966 | 249.043 |  |  |  |
| 1967 | 328.351 |  |  |  |
| 1968 | 273.108 |  |  |  |
| 1969 | 264.002 |  |  |  |
| 1970 | 233.280 |  |  |  |
| 1971 | 203.334 |  |  |  |
| 1972 | 303.987 |  |  |  |
| 1973 | 233.679 |  |  |  |
| 1974 | 326.603 |  |  |  |
| 1975 | 324.690 |  |  |  |
| 1976 | 366.530 |  |  |  |
| 1977 | 363.726 |  |  |  |
| 1978 | 436.206 |  |  |  |
| 1979 | 400.264 |  |  |  |
| 1980 | 343.842 |  |  |  |
| 1981 | 450.492 |  |  |  |
| 1982 | 464.614 |  |  |  |
| 1983 | 579.104 |  |  |  |
| 1984 | 555.910 |  |  |  |
| 1985 | 619.434 |  |  |  |
| 1986 | 621.324 |  |  |  |
| 1987 | 725.632 |  |  |  |
| 1988 | 804.011 | 32.897 | 194.618 | 1031.526 |
| 1989 | 964.785 | 77.349 | 101.816 | 1143.950 |
| 1990 | 647.051 | 90.941 | 119.346 | 857.338 |
| 1991 | 497.024 | 195.536 | 205.428 | 897.988 |
| 1992 | 493.009 | 79.843 | 196.154 | 769.006 |
| 1993 | 348.334 | 60.781 | 237.251 | 646.365 |
| 1994 | 407.289 | 115.097 | 288.539 | 810.925 |
| 1995 | 458.570 | 162.243 | 252.011 | 872.824 |
| 1996 | 368.267 | 135.961 | 173.606 | 677.834 |
| 1997 | 397.395 | 146.442 | 199.328 | 743.166 |
| 1998 | 381.278 | 63.616 | 249.348 | 694.242 |
| 1999 | 313.286 | 47.047 | 314.988 | 675.322 |
| 2000 | 419.407 | 57.963 | 250.293 | 727.662 |
| 2001 | 348.517 | 53.195 | 226.287 | 627.999 |
| 2002 | 294.996 | 76.305 | 154.183 | 525.484 |
| 2003 | 302.622 | 125.516 | 144.175 | 572.313 |
| 2004 | 279.466 | 95.415 | 142.609 | 517.490 |
| 2005 | 336.920 | 81.741 | 170.777 | 589.438 |
| 2006 | 258.497 | 103.217 | 124.158 | 485.872 |
| 2007 | 309.522 | 22.929 | 168.229 | 500.680 |
| Average |  |  |  |  |
| 1988-2007 | 431.512 | 91.202 | 195.657 | 718.371 |

Table 3.--Production model estimates of bottomfish biomass and relative biomass ( $\mathrm{B} / \mathrm{B}_{\mathrm{MSY}}$ ) status by fishery zone, 1988-2007.

Bottomfish Biomass and Status Estimates, 1988-2007

|  | Mean <br> MHI <br> Biomass <br> (klb) | Mean <br> Mau <br> Biomass <br> (klb) | Mean <br> Hoomalu <br> Biomass <br> (klb) | Mean <br> Relative <br> MHI <br> Biomass | Mean <br> Relative <br> Mau <br> Biomass | Mean <br> Relative <br> Hoomalu <br> Biomass | Mean <br> Relative |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 1622 | 698 | 3152 | 1.02 | 1.59 | 2.07 | 1.54 |
| 1989 | 1579 | 801 | 3121 | 0.99 | 1.82 | 2.05 | 1.55 |
| 1990 | 1276 | 763 | 2819 | 0.80 | 1.74 | 1.85 | 1.37 |
| 1991 | 1232 | 678 | 2773 | 0.78 | 1.54 | 1.82 | 1.32 |
| 1992 | 1029 | 541 | 2680 | 0.65 | 1.23 | 1.76 | 1.20 |
| 1993 | 972 | 568 | 2844 | 0.61 | 1.29 | 1.87 | 1.23 |
| 1994 | 1018 | 624 | 2599 | 0.64 | 1.42 | 1.71 | 1.19 |
| 1995 | 1044 | 608 | 2416 | 0.66 | 1.38 | 1.59 | 1.15 |
| 1996 | 914 | 562 | 2340 | 0.58 | 1.28 | 1.54 | 1.08 |
| 1997 | 961 | 568 | 2361 | 0.61 | 1.29 | 1.55 | 1.10 |
| 1998 | 905 | 533 | 2253 | 0.57 | 1.21 | 1.48 | 1.04 |
| 1999 | 953 | 568 | 2263 | 0.60 | 1.29 | 1.49 | 1.07 |
| 2000 | 1053 | 602 | 2358 | 0.66 | 1.37 | 1.55 | 1.13 |
| 2001 | 1010 | 642 | 2184 | 0.64 | 1.46 | 1.43 | 1.08 |
| 2002 | 970 | 712 | 1858 | 0.61 | 1.62 | 1.22 | 1.00 |
| 2003 | 938 | 739 | 1988 | 0.59 | 1.68 | 1.31 | 1.03 |
| 2004 | 911 | 701 | 1906 | 0.57 | 1.59 | 1.25 | 0.99 |
| 2005 | 1019 | 711 | 1965 | 0.64 | 1.62 | 1.29 | 1.04 |
| 2006 | 962 | 716 | 2006 | 0.61 | 1.63 | 1.32 | 1.04 |
| 2007 | 982 | 682 | 2344 | 0.62 | 1.55 | 1.54 | 1.13 |

Table 4.--Production model estimates of bottomfish biomass and relative harvest rate $\left(\mathrm{H} / \mathrm{H}_{\text {MSY }}\right)$ status by fishery zone, 1988-2007.

Bottomfish Harvest Rate and Status Estimates, 1988-2007

|  | Mean |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | MHI <br> Harvest <br> Rate | Mean <br> Mau <br> Harvest <br> Rate | Mean <br> Hoomalu <br> Harvest <br> Rate | Mean <br> Relative <br> MHI <br> Harvest <br> Rate | Mean <br> Relative <br> Mau <br> Harvest <br> Rate | Mean <br> Relative <br> Hoomalu <br> Harvest <br> Rate | Mean <br> Relative <br> Harvest Rate |
| 1988 | $50 \%$ | $5 \%$ | $6 \%$ | 1.74 | 0.17 | 0.22 | 0.89 |
| 1989 | $62 \%$ | $10 \%$ | $3 \%$ | 2.15 | 0.34 | 0.12 | 1.05 |
| 1990 | $51 \%$ | $12 \%$ | $4 \%$ | 1.78 | 0.42 | 0.15 | 0.91 |
| 1991 | $41 \%$ | $29 \%$ | $7 \%$ | 1.42 | 1.02 | 0.26 | 0.87 |
| 1992 | $48 \%$ | $15 \%$ | $7 \%$ | 1.68 | 0.52 | 0.26 | 0.93 |
| 1993 | $36 \%$ | $11 \%$ | $8 \%$ | 1.26 | 0.38 | 0.29 | 0.74 |
| 1994 | $40 \%$ | $19 \%$ | $11 \%$ | 1.41 | 0.65 | 0.39 | 0.88 |
| 1995 | $44 \%$ | $27 \%$ | $11 \%$ | 1.54 | 0.95 | 0.37 | 0.97 |
| 1996 | $41 \%$ | $25 \%$ | $8 \%$ | 1.42 | 0.86 | 0.26 | 0.85 |
| 1997 | $42 \%$ | $26 \%$ | $9 \%$ | 1.45 | 0.92 | 0.30 | 0.89 |
| 1998 | $43 \%$ | $12 \%$ | $11 \%$ | 1.48 | 0.43 | 0.39 | 0.88 |
| 1999 | $33 \%$ | $8 \%$ | $14 \%$ | 1.16 | 0.29 | 0.49 | 0.76 |
| 2000 | $40 \%$ | $10 \%$ | $11 \%$ | 1.40 | 0.34 | 0.38 | 0.83 |
| 2001 | $35 \%$ | $8 \%$ | $10 \%$ | 1.21 | 0.30 | 0.37 | 0.74 |
| 2002 | $31 \%$ | $11 \%$ | $8 \%$ | 1.07 | 0.38 | 0.29 | 0.65 |
| 2003 | $33 \%$ | $17 \%$ | $7 \%$ | 1.13 | 0.60 | 0.26 | 0.69 |
| 2004 | $31 \%$ | $14 \%$ | $8 \%$ | 1.08 | 0.48 | 0.27 | 0.66 |
| 2005 | $33 \%$ | $12 \%$ | $9 \%$ | 1.16 | 0.41 | 0.31 | 0.70 |
| 2006 | $27 \%$ | $15 \%$ | $6 \%$ | 0.94 | 0.51 | 0.22 | 0.58 |
| 2007 | $32 \%$ | $3 \%$ | $7 \%$ | 1.11 | 0.12 | 0.25 | 0.62 |

Table 5.--Production model estimates of bottomfish maximum sustainable yield (MSY) reference points by fishery zone, 1988-2007, including the posterior mean biomass (thousands of pounds) to produce MSY ( $\mathrm{B}_{\mathrm{MSY}}$ ) and its $80 \%$ credibility interval ( $\left[\mathrm{P}_{10}, \mathrm{P}_{90}\right]$ ), the posterior mean harvest rate to produce MSY $\left(\mathrm{H}_{\mathrm{MSY}}\right)$ and its $80 \%$ credibility interval, and the annual instantaneous fishing mortality rate to produce MSY ( $\mathrm{F}_{\mathrm{MSY}}$ ) and its $80 \%$ credibility interval.

|  | $\mathrm{B}_{\text {MSY }}$ |  |  |
| :--- | :---: | :---: | :---: |
| Fishery Zone | $\mathrm{P}_{10}$ | Mean | $\mathrm{P}_{90}$ |
| Main Hawaiian Islands | 1451 | 1588 | 1729 |
| Mau Zone | 402 | 440 | 479 |
| Hoomalu Zone | 1392 | 1524 | 1659 |


|  | $\mathrm{H}_{\mathrm{MSY}}$ |  |  |
| :--- | :---: | :---: | :---: |
| Fishery Zone | $\mathrm{P}_{10}$ | Mean | $\mathrm{P}_{90}$ |
| Main Hawaiian Islands, <br> Mau Zone, Hoomalu Zone | $25 \%$ | $29 \%$ | $33 \%$ |
|  |  | $\mathrm{~F}_{\mathrm{MSY}}$ |  |
| Fishery Zone | $\mathrm{P}_{10}$ | Mean | $\mathrm{P}_{90}$ |
| Main Hawaiian Islands, <br> Mau Zone, Hoomalu Zone | 0.29 | 0.34 | 0.39 |

Table 6.--Sensitivity analysis using the status quo MHI CPUE to assess bottomfish relative biomass $\left(\mathrm{B} / \mathrm{B}_{\mathrm{MSY}}\right)$ and relative harvest rate $\left(\mathrm{H} / \mathrm{H}_{\mathrm{MSY}}\right)$ status by fishery zone, 1988-2007.

## Sensitivity Analysis Using Status Quo MHI CPUE

| Year | Mean <br> Relative <br> MHI <br> Biomass | Mean <br> Relative <br> Mau <br> Biomass | Mean <br> Relative <br> Hoomalu <br> Biomass | Mean <br> Relative Archipelagic Biomass | Mean <br> Relative <br> MHI <br> Harvest <br> Rate | Mean <br> Relative <br> Mau <br> Harvest <br> Rate | Mean <br> Relative <br> Hoomalu <br> Harvest <br> Rate | Mean <br> Relative <br> Archipelagic Harvest Rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 1.05 | 1.09 | 1.11 | 1.08 | 2.54 | 0.36 | 0.63 | 1.45 |
| 1989 | 1.29 | 1.63 | 1.07 | 1.24 | 2.46 | 0.59 | 0.35 | 1.32 |
| 1990 | 1.09 | 1.54 | 0.96 | 1.09 | 1.97 | 0.72 | 0.45 | 1.17 |
| 1991 | 0.82 | 1.21 | 0.95 | 0.92 | 1.99 | 1.95 | 0.79 | 1.47 |
| 1992 | 0.68 | 0.90 | 0.91 | 0.81 | 2.39 | 1.08 | 0.78 | 1.54 |
| 1993 | 0.62 | 0.97 | 0.96 | 0.81 | 1.86 | 0.76 | 0.90 | 1.31 |
| 1994 | 0.67 | 1.14 | 0.88 | 0.82 | 2.00 | 1.21 | 1.19 | 1.56 |
| 1995 | 0.79 | 1.10 | 0.81 | 0.84 | 1.93 | 1.78 | 1.14 | 1.57 |
| 1996 | 0.60 | 1.03 | 0.78 | 0.73 | 2.03 | 1.59 | 0.82 | 1.45 |
| 1997 | 0.57 | 1.14 | 0.79 | 0.74 | 2.29 | 1.58 | 0.93 | 1.61 |
| 1998 | 0.56 | 1.01 | 0.76 | 0.70 | 2.24 | 0.78 | 1.20 | 1.61 |
| 1999 | 0.57 | 1.01 | 0.76 | 0.71 | 1.82 | 0.57 | 1.51 | 1.53 |
| 2000 | 0.64 | 0.97 | 0.78 | 0.74 | 2.16 | 0.73 | 1.18 | 1.56 |
| 2001 | 0.53 | 1.06 | 0.72 | 0.68 | 2.19 | 0.61 | 1.16 | 1.55 |
| 2002 | 0.57 | 1.33 | 0.60 | 0.68 | 1.72 | 0.69 | 0.93 | 1.25 |
| 2003 | 0.54 | 1.45 | 0.65 | 0.70 | 1.84 | 1.05 | 0.82 | 1.30 |
| 2004 | 0.50 | 1.38 | 0.62 | 0.66 | 1.86 | 0.84 | 0.84 | 1.29 |
| 2005 | 0.54 | 1.44 | 0.65 | 0.70 | 2.05 | 0.69 | 0.97 | 1.42 |
| 2006 | 0.50 | 1.43 | 0.66 | 0.68 | 1.70 | 0.88 | 0.69 | 1.17 |
| 2007 | 0.54 | 1.30 | 0.78 | 0.74 | 1.89 | 0.21 | 0.80 | 1.21 |

Table 7.--Projected bottomfish catches, probabilities of overfishing, and relative biomasses under the $\mathrm{F}=\mathrm{F}_{\text {MSY }}$ alternative in fishing years 2009-2010 in the main Hawaiian Islands, Mau zone, and Hoomalu zone.

F=FMSY Projection for 2009-2010

|  | Median |  |  | Median |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Median | Potential |  |  |  | Median | Median |  |  |  |
|  | Potential | Deep 7 |  |  |  | Potential Bottomfish | Deep 7 |  |  |  |
|  | Bottomfish | Bottomfish |  |  |  |  | Bottomfish |  |  | $\operatorname{Pr}$ (Biomass |
|  | Catch at | Catch at |  | Mean Relative | $\operatorname{Pr}$ (Biomass | Catch at | Catch at |  | Mean Relative | exceeds |
| Bottomfish | FMSY in | FMSY in |  | Biomass | exceeds | FMSY in | FMSY in |  | Biomass | BMSY at |
| Management | FY 2009 | FY 2009 | $\operatorname{Pr}$ (Overfishing | (B/BMSY) at | BMSY at Start | FY 2010 | FY 2010 | Pr(Overfishing | (B/BMSY) at | Start of |
| Unit | (klb) | (klb) | in 2009) | Start of 2009 | of 2009) | (klb) | (klb) | in 2010) | Start of 2010 | 2010) |
| Archipelago |  |  | 1.00 | 1.28 | 1.00 |  |  | 1.00 | 1.11 | 1.00 |
| MHI | 343.6 | 248.4 | 1.00 | 0.76 | 0.00 | 368.1 | 266.1 | 1.00 | 0.81 | 0.00 |
| Mau | 226.1 | 163.5 | 1.00 | 1.78 | 1.00 | 174.1 | 125.9 | 1.00 | 1.38 | 1.00 |
| Hoomalu | 738.3 | 533.8 | 1.00 | 1.68 | 1.00 | 590.6 | 427.0 | 1.00 | 1.35 | 1.00 |

Table 8.--Projected bottomfish catches, probabilities of overfishing, and relative biomasses under the $\mathrm{F}=\mathrm{F}_{\text {STATUS }}$ QUO alternative in fishing years 2009-2010 in the main Hawaiian Islands, Mau zone, and Hoomalu zone.

## F=F2008 Status Quo Projection for 2009-2010



Table 9.--Projected TACs, probabilities of overfishing, relative biomasses, and probabilities of depletion of MHI bottomfish under the constant MHI TAC alternative in fishing years 20092010.

Constant MHI TAC Projections for 2009-2010

| Main |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Main | Hawaiian Islands Deep | Probability of | Probability of | Probability of | Mean | Probability That Bottomfish in the Main |
| Hawaiian | 7 Bottomfish |  | Overfishing | Overfishing | Relative |  |
| Islands | Annual Total | Overfishing | Bottomfish | Bottomfish | Biomass | Hawaiian |
| Bottomfish | Allowable | Bottomfish in | in the Main | in the Main | (B/BMSY) | Islands Are |
| Annual Total | Catch | the Hawaiian | Hawaiian | Hawaiian | in the MHI | Depleted |
| Allowable | Equivalent | Archipelago | Islands in | Islands in | at Start of | (B<0.7*BMSY) |
| Catch (klb) | (klb) | in FY 2009 | FY 2009 | FY 2010 | FY 2010 | in FY 2010 |
| 238 | 172 | 0.000 | 0.00 | 0.00 | 0.88 | 0.02 |
| 274 | 198 | 0.000 | 0.05 | 0.02 | 0.85 | 0.04 |
| 288 | 208 | 0.000 | 0.10 | 0.04 | 0.84 | 0.05 |
| 299 | 216 | 0.000 | 0.15 | 0.07 | 0.84 | 0.05 |
| 307 | 222 | 0.000 | 0.20 | 0.10 | 0.83 | 0.06 |
| 314 | 227 | 0.000 | 0.25 | 0.13 | 0.83 | 0.07 |
| 321 | 232 | 0.000 | 0.30 | 0.17 | 0.82 | 0.07 |
| 327 | 236 | 0.000 | 0.35 | 0.21 | 0.82 | 0.08 |
| 333 | 241 | 0.000 | 0.40 | 0.25 | 0.82 | 0.09 |
| 338 | 244 | 0.000 | 0.45 | 0.29 | 0.81 | 0.09 |
| 344 | 249 | 0.000 | 0.50 | 0.34 | 0.81 | 0.10 |
| 350 | 253 | 0.000 | 0.55 | 0.39 | 0.80 | 0.11 |
| 356 | 257 | 0.000 | 0.60 | 0.44 | 0.80 | 0.12 |
| 362 | 262 | 0.000 | 0.65 | 0.50 | 0.80 | 0.13 |
| 368 | 266 | 0.000 | 0.70 | 0.55 | 0.79 | 0.14 |
| 375 | 271 | 0.000 | 0.75 | 0.61 | 0.79 | 0.15 |
| 383 | 277 | 0.000 | 0.80 | 0.68 | 0.78 | 0.16 |
| 392 | 283 | 0.000 | 0.85 | 0.75 | 0.78 | 0.18 |
| 405 | 293 | 0.000 | 0.90 | 0.83 | 0.77 | 0.21 |
| 423 | 306 | 0.001 | 0.95 | 0.91 | 0.76 | 0.25 |
| 475 | 343 | 0.007 | 1.00 | 0.99 | 0.73 | 0.38 |



Figure 1.--Location of the three Hawaiian bottomfish fishing zones: MHI, the Mau zone, and the Hoomalu zone.


Figure 2.--Total Hawaiian bottomfish catch used in the assessment update by zone and fishing year (e.g., fishing 1949 corresponds to July 1st 1948 to June 30th 1949) and fishery zone. As in the previous assessment, fishery catches in the Mau and Hoomalu zones prior to 1988 were not used in the assessment update because there were no CPUE data available to provide relative abundance indices for these zones prior to 1988.


Figure 3.--Bottomfish CPUE adjustment coefficients to account for increases in Hawaiian bottomfish fishing technology during 1949-2007 adapted from Moffitt et al. (2006). Observed CPUE input to the assessment model was the product of the catchability parameter (Q) for the standard time period (1980-1991) times the technology coefficient (c) times the model estimate of fishable biomass (B).


Figure 4.--Total bottomfish catch on deep handline trips by fishing year used for status quo and GLM standardization approaches to compute CPUE for the bottomfish fishery in the main Hawaiian Islands along with total bottomfish catch on all deep handline trips.


Figure 5.--Total bottomfish deep handline trips by fishing year used for status quo and GLM standardization approaches to compute CPUE for the bottomfish fishery in the main Hawaiian Islands along with total bottomfish deep handline trips in the MHI.


Figure 6.--Comparison of trends in status quo bottomfish CPUE in the MHI as calculated in this assessment and in the previous assessment. The status quo CPUE for the MHI is calculated using re-audited HDAR logbook data and the same approach as in Moffitt et al. (2006) except that a single-day trip qualifying criterion of no more than 1,500 pounds of bottomfish is applied.


Figure 7.--Trends in bottomfish CPUE (pounds per day fished) by fishing year in the Mau and Hoomalu zones. The CPUE for the Mau and Hoomalu zones during 1988-2004 were taken from the previous assessment (Moffitt et al., 2006) and CPUE in 2005-2007 were calculated using the same approach as in the previous assessment.


Figure 8.--HDAR fishery reporting areas used for the area effect in the GLM approach to standardize bottomfish CPUE in the main Hawaiian Islands bottomfish fishery.

## Bottomfish Catch Ratio by Fishing Year



Figure 9.--Annual percentiles of the ratio of bottomfish catch (Table 2) to total catch weight per trip for fishing trips reporting deep handline gear in the main Hawaiian Islands during 19492007 including the $90^{\text {th }}$ percentile ( p 90 ), the median and the $10^{\text {th }}$ percentile ( p 10 ).


Figure 10.--Histogram of natural log-transformed bottomfish catch rates for directed deep handline trips from the main Hawaiian Islands during fishing years 1949-2007.


Figure 11.--Annual trends in mean standardized bottomfish CPUE in the main Hawaiian Islands (solid circle) and $95 \%$ confidence interval as estimated by the log-linear model along with the status quo CPUE (open circle) for comparison.


Figure 12.--Estimated mean CPUE by fishing area for the directed main Hawaiian Islands bottomfish fishery.


Figure 13.--Estimated standardized CPUE by month for the spatial CPUE standardization approach applied to the main Hawaiian Islands bottomfish fishery where January is month 1, February is month $2, \ldots$, and December is month 12 .

Observed standardized bottomfish CPUE versus predicted CPUE in the Main Hawaiian Islands by fishing year, 1949-2007


Standardized log-scale residuals of the production model fit to standardized CPUE in the Main Hawaiian Islands by fishing year, 1949-2007


Figure 14.--Production model fit to the observed standardized bottomfish CPUE for the main Hawaiian Islands along with standardized log-scale CPUE residuals.

Observed standardized bottomfish CPUE versus predicted CPUE in the Mau zone by fishing year, 1988-2007


Standardized log-scale residuals of the model fit to standardized CPUE in the Mau zone by fishing year, 1988-2007


Figure 15.--Production model fit to the observed standardized bottomfish CPUE for the Mau zone along with standardized log-scale CPUE residuals.

Observed standardized bottomfish CPUE versus predicted CPUE in the Hoomalu zone by fishing year, 1988-2007


Standardized log-scale residuals of the production model fit to standardized CPUE in the Hoomalu zone by fishing year, 1988-2007


Figure 16.--Production model fit to the observed standardized bottomfish CPUE for the Hoomalu zone along with standardized log-scale CPUE residuals.


Figure 17.--Stock status of the Archipelagic bottomfish stock during 1988-2007. The stock was experiencing overfishing in a fishing year if the harvest rate fraction exceeded 1. The stock was overfished in a fishing year if the biomass fraction was less than 0.7.


Figure 18.--Stock status of the Hoomalu zone bottomfish stock during 1988-2007. The stock was experiencing overfishing in a fishing year if the harvest rate fraction exceeded 1 . The stock was overfished in a fishing year if the biomass fraction was less than 0.7 .


Figure 19.--Stock status of the Mau zone bottomfish stock during 1988-2007. The stock was experiencing overfishing in a fishing year if the harvest rate fraction exceeded 1 . The stock was overfished in a fishing year if the biomass fraction was less than 0.7.


Figure 20.--Stock status of the main Hawaiian Islands bottomfish stock during 1949-2007. The stock was experiencing overfishing in a fishing year if the harvest rate fraction exceeded 1 . The stock was overfished in a fishing year if the biomass fraction was less than 0.7 .


Figure 21.--Mean bottomfish biomass in the main Hawaiian Islands and $80 \%$ confidence interval from the baseline assessment model, 1949-2007.

## Estimated bottomfish exploitation rate in the Main Hawaiian Islands

 from the Baseline Assessment Model, 1949-2007

Figure 22.--Mean exploitation rate for bottomfish in the main Hawaiian Islands and 80\% confidence interval from the baseline assessment model, 1949-2007.


Figure 23.--Sensitivity analysis of stock status of the Archipelagic bottomfish stock during 1988-2007 if status quo main Hawaiian Islands CPUE was used to fit the model instead of standardized CPUE. The stock was experiencing overfishing in a fishing year if the harvest rate fraction exceeded 1. The stock was overfished in a fishing year if the biomass fraction was less than 0.7.


Figure 24.--Sensitivity analysis of relative bottomfish biomass estimates for the main Hawaiian Islands to assumed values of prior means for the intrinsic growth rate parameter from $\mathrm{R}=0.25$ to 0.75 . The mean prior value of R used in the baseline model was $\mathrm{R}=0.46$.


Figure 25.--Sensitivity analysis of relative bottomfish biomass estimates for the main Hawaiian Islands to assumed values of prior means for the MHI carrying capacity parameter from $\mathrm{K}=1500$ to 2500 klbs . The mean prior value of K used in the baseline model was $\mathrm{K}=2000 \mathrm{klbs}$.


Fishing year

Figure 26.--Sensitivity analysis of relative bottomfish biomass estimates for the main Hawaiian Islands to assumed values of prior means of the observation error variance parameter $\left(\tau^{2}\right)$ by fishing zone for values $10 \tau^{2}$., $100 \tau^{2}, 1000 \tau^{2}$, and $10000 \tau^{2}$.


Figure 27.--Sensitivity analysis of relative bottomfish biomass estimates for the main Hawaiian Islands to assumed values of prior means of the process error variance parameter $\left(\sigma^{2}\right)$ by fishing zone for values $10 \sigma^{2}$., $100 \sigma^{2}, 1000 \sigma^{2}$, and $10000 \sigma^{2}$.


Figure 28.--Projected bottomfish catch biomass in the Hoomalu zone during 2008-2010 assuming status quo fishing effort and mortality.


Figure 29.--Projected bottomfish catch biomass in the Mau zone during 2008-2010 assuming status quo fishing effort and mortality.

Projected Main Hawaiian Islands Bottomfish Biomass in 2008-2010 Assuming Status Quo Fishing Mortality from 2008


Figure 30.--Projected bottomfish catch biomass in the main Hawaiian Islands during 2009-2010 assuming status quo fishing effort and mortality.

Projected Hoomalu Zone Bottomfish Biomass in 2008-2010
Assuming Status Quo Fishing Mortality ( $77 \%$ Below $\mathrm{F}_{\text {MSY }}$ )


Figure 31.--Projected bottomfish biomass in the Hoomalu zone during 2008-2010 assuming status quo fishing effort and mortality.

Projected Mau Zone Bottomfish Biomass in 2008-2010 Assuming Status Quo Fishing Mortality ( $89 \%$ Below $\mathrm{F}_{\text {MSY }}$ )


Figure 32.--Projected bottomfish biomass in the Mau zone during 2008-2010 assuming status quo fishing effort and mortality.

Projected Main Hawaiian Islands Bottomfish Biomass in 2008-2010 Assuming Status Quo Fishing Mortality from 2008


Figure 33.--Projected bottomfish biomass in the main Hawaiian Islands during 2009-2010 assuming status quo fishing effort and mortality.


Figure 34.--Risk of overfishing the main Hawaiian Islands bottomfish stock complex in 2009 as a function of the total allowable catch (TAC) in 2009 of bottomfish or the deep 7 bottomfish species.

## Appendix.

Table A1.--Results of the multiplicative loglinear model (Gavaris, 1980) used to standardize bottomfish CPUE for the main Hawaiian Islands. The CPUE predictors included fishing year (fishyear), month, and fishing area (garea) as predictors of bottomfish CPUE on directed fishing trips (defined as trips that had at least $50 \%$ bottomfish catch by weight). Estimates of log-scale parameter factor level along with their standard errors and P-values were listed for each predictor. Inferences about the significance of predictors was judged using the Type III sums of squares which are appropriate for unbalanced data sets (e.g., Searle, 1987) which do not contain the same number of observations in each cell.

```
Bottomfish directed CPUE standardization, 1949-2007
log(bfish) = fishyear, area, month
12:19 Monday, September 22, 2008
The GLM Procedure
Class Level Information
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|}
\hline Class & Levels & \multicolumn{9}{|l|}{Values} \\
\hline \multirow[t]{5}{*}{fishyear} & & 194919501951 & 1952 & 1953 & 19541955 & 1956 & 1957 & 1958 & 1959 & 1960 \\
\hline & \multirow{3}{*}{59} & 196119621963 & 1964 & 1965 & 19661967 & 1968 & 1969 & 1970 & 1971 & 1972 \\
\hline & & 197319741975 & 1976 & 1977 & 19781979 & 1981 & 1982 & 1983 & 1984 & 1985 \\
\hline & & 198619871988 & 1989 & 1990 & 19911992 & 1993 & 1994 & 1995 & 1996 & 1997 \\
\hline & & 199819992000 & 2001 & 2002 & 20032004 & 2005 & 2006 & 2007 & 91980 & \\
\hline \multirow[t]{3}{*}{garea} & \multirow[t]{3}{*}{32} & 101102105120 & 121 & 1221 & 125126128 & 305 & 3203 & 321 & 23324 & 4325 \\
\hline & & 327328332333 & 423 & 4295 & 521523525 & 527 & 5281 & 1000 & 3000 & 4000 \\
\hline & & 500060009331 & & & & & & & & \\
\hline month & 12 & 12345678 & 910 & 111 & & & & & & \\
\hline
\end{tabular}
```

[^0]Table A1.--continued.

$$
\begin{aligned}
& \text { Bottomfish directed CPUE standardization, 1949-2007 } \\
& \log (\text { bfish })=\text { fishyear, area, month } \\
& 12: 19 \text { Monday, September 22, } 2008
\end{aligned}
$$

The GLM Procedure
Dependent Variable: logbfish

| Source | DF | Sum of Squares | Mean Square | F Value | $\mathrm{Pr}>\mathrm{F}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Model | 100 | 33308.4659 | 333.0847 | 300.34 | <. 0001 |
| Error | 1410221 | 156396.8501 | 1.1090 |  |  |
| Corrected | 1411221 | 189705.3160 |  |  |  |
|  | Coeff Var25.29915 | $r$ Root MSE | logbfish Mean |  |  |
|  |  | 51.053102 | 2 4.16 | 2599 |  |
| Source | DF | Type I SS | Mean Square | F Value | $\mathrm{Pr}>\mathrm{F}$ |
| fishyear | 58 | 6884.13761 | 118.69203 | 107.02 | <. 0001 |
| garea | $31 \quad 2$ | 26325.00996 | 849.19387 | 765.71 | <. 0001 |
| month | 11 | 99.31834 | 9.02894 | 8.14 | <. 0001 |
| Source | DF T | Type III SS | Mean Square | F Value | $\mathrm{Pr}>\mathrm{F}$ |
| fishyear | 58 | 5550.80400 | 95.70352 | 86.30 | <. 0001 |
| garea | $31 \quad 2$ | 26243.88126 | 846.57681 | 763.35 | <. 0001 |
| month | 11 | 99.31834 | 9.02894 | 8.14 | <. 0001 |


| Parameter |  | Standard |  |  |  | $\operatorname{Pr}>\|t\|$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Intercept |  | 4.721213986 | B | 0.02396156 | 197.03 | <. 0001 |
| fishyear | 1949 | -0.040312401 | B | 0.03056248 | -1.32 | 0.1872 |
| fishyear | 1950 | 0.000001308 | B | 0.03169828 | 0.00 | 1.0000 |
| fishyear | 1951 | 0.112011992 | B | 0.03343181 | 3.35 | 0.0008 |
| fishyear | 1952 | 0.294633125 | B | 0.03729274 | 7.90 | <. 0001 |
| fishyear | 1953 | 0.368273500 | B | 0.04164092 | 8.84 | <. 0001 |
| fishyear | 1954 | 0.385474478 | B | 0.04215206 | 9.14 | <. 0001 |
| fishyear | 1955 | 0.683201259 | B | 0.05120106 | 13.34 | <. 0001 |
| fishyear | 1956 | 0.515491905 | B | 0.04531814 | 11.37 | <. 0001 |
| fishyear | 1957 | 0.622372072 | B | 0.04246508 | 14.66 | <. 0001 |
| fishyear | 1958 | 0.280701852 | B | 0.04584210 | 6.12 | <. 0001 |
| fishyear | 1959 | 0.082593782 | B | 0.04628571 | 1.78 | 0.0744 |
| fishyear | 1960 | 0.323732640 | B | 0.04368901 | 7.41 | <. 0001 |
| fishyear | 1961 | 0.663261680 | B | 0.05936434 | 11.17 | <. 0001 |
| fishyear | 1962 | 0.703328701 | B | 0.05114267 | 13.75 | <. 0001 |
| fishyear | 1963 | 0.395829033 | B | 0.04570463 | 8.66 | <. 0001 |
| fishyear | 1964 | 0.407243969 | B | 0.04409104 | 9.24 | <. 0001 |
| fishyear | 1965 | 0.604452587 | B | 0.04108083 | 14.71 | <. 0001 |

Table A1.--continued.

> Bottomfish directed CPUE standardization, 1949-2007

> $$
> \begin{array}{c}\log (\text { bfish })=\text { fishyear, area, month } \\ 12: 19 \text { Monday, September 22, } 2008 \\ \text { The GLM Procedure }\end{array}
>
$$

Dependent Variable: logbfish

| Parameter |  | Estimate | Standard Error | t Value | $\operatorname{Pr}>\|t\|$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| fishyear | 1966 | 0.572696922 B | 0.04280716 | 13.38 | $<.0001$ |
| fishyear | 1967 | 0.386446453 B | 0.03669147 | 10.53 | <. 0001 |
| fishyear | 1968 | 0.481069688 B | 0.04031231 | 11.93 | <. 0001 |
| fishyear | 1969 | 0.301167251 B | 0.03860226 | 7.80 | <. 0001 |
| fishyear | 1970 | 0.220276025 B | 0.03937677 | 5.59 | <. 0001 |
| fishyear | 1971 | 0.197326482 В | 0.03952253 | 4.99 | <. 0001 |
| fishyear | 1972 | 0.296050329 В | 0.03535004 | 8.37 | <. 0001 |
| fishyear | 1973 | 0.224969475 В | 0.03717966 | 6.05 | <. 0001 |
| fishyear | 1974 | 0.241788471 B | 0.03241352 | 7.46 | <. 0001 |
| fishyear | 1975 | 0.154955574 B | 0.03285501 | 4.72 | <. 0001 |
| fishyear | 1976 | 0.266222004 B | 0.03322212 | 8.01 | <. 0001 |
| fishyear | 1977 | -0.011470396 B | 0.03196104 | -0.36 | 0.7197 |
| fishyear | 1978 | 0.052504268 В | 0.03244660 | 1.62 | 0.1056 |
| fishyear | 1979 | 0.109998194 В | 0.03458804 | 3.18 | 0.0015 |
| fishyear | 1981 | -0.001729916 B | 0.02804660 | -0.06 | 0.9508 |
| fishyear | 1982 | -0.103224912 B | 0.02841338 | -3.63 | 0.0003 |
| fishyear | 1983 | -0.102117053 B | 0.02673888 | -3.82 | 0.0001 |
| fishyear | 1984 | -0.217175984 B | 0.02750504 | -7.90 | <. 0001 |
| fishyear | 1985 | -0.193410087 B | 0.02632161 | -7.35 | <. 0001 |
| fishyear | 1986 | -0.140542397 B | 0.02633684 | -5.34 | <. 0001 |
| fishyear | 1987 | 0.009187137 B | 0.02634475 | 0.35 | 0.7273 |
| fishyear | 1988 | 0.172509603 В | 0.02594805 | 6.65 | <. 0001 |
| fishyear | 1989 | 0.130704923 В | 0.02543734 | 5.14 | <. 0001 |
| fishyear | 1990 | -0.077829378 B | 0.02605490 | -2.99 | 0.0028 |
| fishyear | 1991 | -0.077223417 B | 0.02731937 | -2.83 | 0.0047 |
| fishyear | 1992 | -0.129073171 B | 0.02727384 | -4.73 | <. 0001 |
| fishyear | 1993 | -0.163120179 B | 0.02872986 | -5.68 | <. 0001 |
| fishyear | 1994 | -0.123346948 B | 0.02796375 | -4.41 | <. 0001 |
| fishyear | 1995 | -0.079349172 B | 0.02772992 | -2.86 | 0.0042 |
| fishyear | 1996 | -0.242271411 B | 0.02793067 | -8.67 | <. 0001 |
| fishyear | 1997 | -0.165683338 B | 0.02776633 | -5.97 | <. 0001 |
| fishyear | 1998 | -0.251648126 B | 0.02781503 | -9.05 | <. 0001 |
| fishyear | 1999 | -0.182509575 B | 0.02937734 | -6.21 | <. 0001 |
| fishyear | 2000 | -0.078669349 B | 0.02793479 | -2.82 | 0.0049 |
| fishyear | 2001 | -0.118673073 В | 0.02867416 | -4.14 | <. 0001 |
| fishyear | 2002 | -0.161782953 В | 0.02999149 | -5.39 | <. 0001 |
| fishyear | 2003 | -0.202209728 B | 0.02937616 | -6.88 | <. 0001 |
| fishyear | 2004 | -0.244808567 B | 0.02982776 | -8.21 | <. 0001 |
| fishyear | 2005 | -0.101512305 B | 0.02983524 | -3.40 | 0.0007 |
| fishyear | 2006 | -0.176096393 В | 0.03109526 | -5.66 | <. 0001 |
| fishyear | 2007 | -0.168124788 B | 0.02970689 | -5.66 | <. 0001 |
| fishyear | 91980 | 0.000000000 B | . | . |  |
| garea | 101 | -0.974282250 B | 0.02049550 | -47.54 | <. 0001 |
| garea | 102 | -0.894392706 В | 0.02123207 | -42.12 | <. 0001 |
| garea | 105 | -1.372208164 B | 0.01808404 | -75.88 | <. 0001 |
| garea | 120 | -0.464481003 В | 0.02999957 | -15.48 | <. 0001 |
| garea | 121 | -0.935436701 B | 0.01565290 | -59.76 | <. 0001 |

Table A1.--continued.

$$
\begin{gathered}
\text { Bottomfish directed CPUE standardization, 1949-2007 } \\
\begin{array}{c}
\text { log(bfish) }=\text { fishyear, area, month } \\
12: 19 \text { Monday, September 22, } 2008 \\
\text { The GLM Procedure }
\end{array} \text { } 4
\end{gathered}
$$

Dependent Variable: logbfish

| Parameter |  | Estimate |  | Standard Error | t Value | $\operatorname{Pr}>\|t\|$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| garea | 122 | -0.801311706 | B | 0.01307570 | -61.28 | <. 0001 |
| garea | 125 | -1.139865966 | B | 0.01812828 | -62.88 | <. 0001 |
| garea | 126 | -1.005327470 | B | 0.02116473 | -47.50 | <. 0001 |
| garea | 128 | 0.194495262 | B | 0.04386945 | 4.43 | <. 0001 |
| garea | 305 | 0.150410225 | B | 0.03101746 | 4.85 | <. 0001 |
| garea | 320 | -0.637359395 | B | 0.01145319 | -55.65 | <. 0001 |
| garea | 321 | -0.370246381 | B | 0.01619020 | -22.87 | <. 0001 |
| garea | 323 | -0.327256054 | B | 0.02653792 | -12.33 | <. 0001 |
| garea | 324 | 0.131016949 | B | 0.02933380 | 4.47 | <. 0001 |
| garea | 325 | 0.321392382 | B | 0.02974878 | 10.80 | <. 0001 |
| garea | 327 | 0.065727983 | B | 0.01889979 | 3.48 | 0.0005 |
| garea | 328 | 0.353773485 | B | 0.01866461 | 18.95 | <. 0001 |
| garea | 332 | 0.121565082 | B | 0.04347714 | 2.80 | 0.0052 |
| garea | 333 | 0.320841166 | B | 0.03978457 | 8.06 | <. 0001 |
| garea | 423 | -0.686846311 | B | 0.02151546 | -31.92 | <. 0001 |
| garea | 429 | -1.024662815 | B | 0.02382235 | -43.01 | <. 0001 |
| garea | 521 | -0.576986493 | B | 0.03546451 | -16.27 | <. 0001 |
| garea | 523 | -0.476038630 | B | 0.03073719 | -15.49 | <. 0001 |
| garea | 525 | 0.454329896 | B | 0.04129522 | 11.00 | <. 0001 |
| garea | 527 | 0.650410241 | B | 0.06056900 | 10.74 | <. 0001 |
| garea | 528 | 0.989596210 | B | 0.04678633 | 21.15 | <. 0001 |
| garea | 1000 | -0.717895280 | B | 0.01563849 | -45.91 | <. 0001 |
| garea | 3000 | -0.410445640 | B | 0.01320285 | -31.09 | <. 0001 |
| garea | 4000 | -0.918167951 | B | 0.01316849 | -69.72 | <. 0001 |
| garea | 5000 | -0.657986570 | B | 0.01356081 | -48.52 | <. 0001 |
| garea | 6000 | 0.014392663 | B | 0.03576605 | 0.40 | 0.6874 |
| garea | 9331 | 0.000000000 | B | . | . | . |
| month | 1 | 0.038091677 | B | 0.01116962 | 3.41 | 0.0006 |
| month | 2 | 0.014676044 | B | 0.01168437 | 1.26 | 0.2091 |
| month | 3 | -0.013873985 | B | 0.01222211 | -1.14 | 0.2563 |
| month | 4 | -0.059030186 | B | 0.01356812 | -4.35 | <. 0001 |
| month | 5 | 0.020992007 | B | 0.01340970 | 1.57 | 0.1175 |
| month | 6 | 0.043253451 | B | 0.01411368 | 3.06 | 0.0022 |
| month | 7 | 0.014489902 | B | 0.01438072 | 1.01 | 0.3137 |
| month | 8 | -0.003327236 | B | 0.01324220 | -0.25 | 0.8016 |
| month | 9 | -0.027394116 | B | 0.01207241 | -2.27 | 0.0233 |
| month | 10 | -0.015012321 | B | 0.01182470 | -1.27 | 0.2042 |
| month | 11 | -0.029830086 | B | 0.01163992 | -2.56 | 0.0104 |
| month | 12 | 0.000000000 | B |  |  |  |

Table A2.--WINBUGS source code used to fit baseline assessment model for bottomfish using standardized MHI CPUE during 1949-2007 and to determine TACs that produce probabilities of overfishing the MHI in 2009 of $0 \%, 5 \%, \ldots, 100 \%$.

## Hawaiian Bottomfish Archipelagic Assessment 2008 Bayesian State-Space Implementation of Schaefer Production Model Fishing Year 1949-2007 PROJECTION MODEL bfish_baseline_TAC_grid MHI C2008 Estimate=266.409 FIND TAC TO SET P(MHI 2009 Overfishing) $=0.05, \ldots, 1.00$

\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\# \# Jon Brodziak, PIFSC, September 2008<br>\# Catch units are thousands of pounds<br>\# CPUE units are thousands of pounds per trip<br>\# This program analyzes MHI catch \& cpue data for 1949-2007<br>\# This program analyzes Mau and Hoomalu catch \& cpue data for 1988-2007<br>\# This program uses CPUE calculated with GLM standardization approach.<br>\# This program assumes 4 periods for MHI catchability based on interviews similar to AR H0601<br>\# This program assumes a common intrinsic growth rate across regions, r<br>\# This program estimates carrying capacity for MHI only<br>\# This program assumes carrying capacities for Mau and Hoomalu are proportional to habitat<br>\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#

```
model bfish_baseline_TAC_test
{
# Prior distributions
#########################################################
# Diffuse normal prior for carrying capacity parameter, K
#(1)########################################################
K_MHI ~ dnorm(2000,0.00001)
K_Mau <- 0.277*K_MHI
K_Hoomalu <- 0.960*K_MHI
# Beta prior for intrinsic growth rate parameter, r
# with mean=0.46 and CV=26%
#(2)#######################################################
r ~ dbeta(7.67,9)
# Gamma priors for CPUE catchability coefficients
# within interval (0.0001,10), q
#(3)#######################################################
iq_MHI ~ dgamma(0.001,0.001)I(0.1,10000)
q_MHI <- 1/iq_MHI
iq_Mau ~ dgamma(0.001,0.001)!(0.1,10000)
q_Mau <- 1/iq_Mau
iq_Hoomalu ~ dgamma(0.001,0.001)!(0.1,10000)
q_Hoomalu <- 1/iq_Hoomalu
# Gamma prior for process error variances, sigma2
#(4)#######################################################
isigma2_MHI ~ dgamma(aO_MHI,b0_MHI)
sigma2_MHI <- 1/isigma2_M
isigma2_Mau ~ dgamma(a0_Mau,b0_Mau)
sigma2_Mau <- 1/isigma2_Mau
isigma2_Hoomalu ~ dgamma(a0_Hoomalu,b0_Hoomalu)
sigma2_Hoomalu <- 1/isigma2_Hoomalu
```

Table A2.--continued.
\# Gamma priors for observation error variances, tau2
\#(5) \#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#
itau2_MHI ~dgamma(c0_MHI,d0_MHI)
tau2_MHI <-1/itau2_MHI
itau2_Mau ~dgamma(c0_Mau,d0_Mau)
tau2_Mau <-1/itau2_Mau
itau2_Hoomalu ~dgamma(c0_Hoomalu,d0_Hoomalu)
tau2_Hoomalu <-1/itau2_Hoomalu
\# Lognormal priors for time series of proportions of $\mathrm{K}, \mathrm{P}$
\#(6)\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#
\# MHI time series starts in 1949 and ends in 2007, n=59
\# Start at 60\% of K
Pmean_MHI[1] <- - 0.511
P_MHI[1] ~ dlnorm(Pmean_MHI[1],isigma2_MHI) I(0.001,10)
\# Process dynamics
for (i in 2:N_MHI) \{
Pmean_MHI[i] <- log(max(P_MHI[i-1] + r*P_MHI[i-1]*(1-P_MHI[i-1]) - Catch_MHI[i-1]/K_MHI,0.001))
P_MHI[i] ~ dlnorm(Pmean_MHI[i],isigma2_MHI)I(0.001,10)
\}
\# Mau time series starts in 1988 and ends in 2007, $n=20, P[1]=0.8$
Pmean_Mau[1] <- -0.223
P_Mau[1] ~ dlnorm(Pmean_Mau[1],isigma2_Mau) I(0.001,10)
\# Process dynamics
for (i in 2:N_Mau) \{
Pmean_Mau[i] <- log(max(P_Mau[i-1] + r*P_Mau[i-1]*(1-P_Mau[i-1]) - Catch_Mau[i-1]/K_Mau,0.001))
P_Mau[i] ~ dlnorm(Pmean_Mau[i],isigma2_Mau)l( $0.001,10$ )
\}
\# Hoomalu time series starts in 1988 and ends in 2007, $n=20, P[1]=0.8$
Pmean_Hoomalu[1] <- -0.223
P_Hoomalu[1] ~ dlnorm(Pmean_Hoomalu[1],isigma2_Hoomalu) I(0.001,10)
\# Process dynamics
for (i in 2:N_Hoomalu) \{
Pmean_Hoomalu[i] <- $\log \left(\max \left(P \_H o o m a l u[i-1]+r^{*} P \_H o o m a l u[i-1] *\left(1-P \_H o o m a l u[i-1]\right)\right.\right.$ - Catch_Hoomalu[i-
1]/K_Hoomalu,0.001))
P_Hoomalu[i] ~ dlnorm(Pmean_Hoomalu[i],isigma2_Hoomalu)I(0.001,10)
\}
\# Lognormal likelihood for observed CPUE indices
\#(7) \#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#
\# MHI CPUE LIKELIHOOD 1949-2007 P_MHI[1:59], 4 Periods for Q
\# Period 1: 1949--1967 relative Q=0.7
for (i in 1:19) \{
CPUEmean_MHI[i] <- log(0.7*q_MHI*K_MHI*P_MHI[i])
CPUE_MHI[i] ~ dlnorm(CPUEmean_MEI[i],itau2_MHI)
RESID_MHI[i] <- log(CPUE_MHI[i]) - $\log \left(0.7^{*} \mathrm{q}_{-}\right.$MHI*K_MHI*P_MHI[i])
$\overline{\}}$
\# Period 2: 1968--1979 relative Q=0.8
for (i in 20:31) \{
CPUEmean_MHI[i] <- log(0.8*q_MHI*K_MHI*P_MHI[i])
CPUE_MHI[i] ~ dlnorm(CPUEmean_MHI[i],itau2_MHI)
RESID_MHI[i] <- log(CPUE_MHI[i]) - $\log \left(0.8^{*} q \_M H I * K \_M H I * P \_M H I[i]\right)$
\}
\# Period 3: 1980--1991 relative Q=1
for (i in 32:43) \{
CPUEmean_MHI[i] <- log(q_MHI*K_MHI*P_MHI[i])
CPUE_MHI[i] ~ dlnorm(CPUEmean_MHI[i],itau2_MHI)
RESID_MHI[i] <- $\log \left(C P U E \_M H I[i]\right)-\log \left(q \_M H I * K \_M H I * P \_M H I[i]\right)$
\}
\# Period 4: 1992--2007 relative $\mathrm{Q}=1.2$
for (i in 44:59) \{
CPUEmean_MHI[i] <- log(1.2*q_MHI*K_MHI*P_MHI[i])

Table A2.--continued.

```
    CPUE_MHI[i] ~ dlnorm(CPUEmean_MHI[i],itau2_MHI)
    RESID_MHI[i] <- log(CPUE_MHI[i]) - log(1.2*q_MHI*K_MHI*P_MHI[i])
        }
# Compute RMSE for MHI CPUE
RSS_MHI <- inprod(RESID_MHI[], RESID_MHI[])
RMSE_MHI <- sqrt(RSS_MHI/N_MHI)
AIC_MHI <- N_MHI*log(RSS_MHI/N_MHI)+2*NPAR_MHI
AICC_MHI <- AIC_MHI+2*NPAR_MHI*(NPAR_MHI+1)/(N_MHI-NPAR_MHI-1)
# Mau CPUE LIKELIHOOD 1988-2007 P_Mau[1:20], qMultiplier=1
for (i in 1:N_Mau) {
    CPUEmēean_Mau[i] <- log(q_Mau*K_Mau*P_Mau[i])
    CPUE_Mau[i] ~ dlnorm(CPUEmean_Mau[i],itau2_Mau)
    RESID_Mau[i] <- log(CPUE_Mau[i])- - log(q_Mau*K_Mau*P_Mau[i])
        }
# Compute RMSE for Mau CPUE
RSS_Mau <- inprod(RESID_Mau[], RESID_Mau[])
RMSE_Mau <- sqrt(RSS_Mau/N_Mau)
# Hoomalu CPUE LIKELIHOOD 1988-2007 P_Hoomalu[1:20], qMultiplier=1
for (i in 1:N_Hoomalu) {
    CPUEmean_Hoomalu[i]<- log(q_Hoomalu*K_Hoomalu*P_Hoomalu[i])
    CPUE_Hoomalu[i] ~ dlnorm(CPUEmean_Hoomalu[i],itau2_Hoomalu)
    RESID_Hoomalu[i] <- log(CPUE_Hoomalu[i]) - log(q_Hoomalu*K_Hoomalu*P_Hoomalu[i])
        }
# Compute RMSE for Hoomalu CPUE
RSS_Hoomalu <- inprod(RESID_Hoomalu[], RESID_Hoomalu[])
RMSE_Hoomalu <- sqrt(RSS_Hoomalu/N_Hoomalu)
# Use total likelihood for overall AIC calculation
N_TOT <- N_MHI+N_Mau+N_Hoomalu
AIC TOT<-
N_M
AICC_TOT <- AIC_TOT+2*NPAR_TOT*(NPAR_TOT+1)/(N_TOT-NPAR_TOT-1)
# Compute exploitation rate and biomass time series
#(8)#######################################################
# MHI 1949-2007 P_MHI[1:59]
for (i in 1:N_MHI) {
    B_MHI[i] <- P_MHI[i]*K_MHI
    H_MHI[i] <- Catch_MHI[i]/B_MHI[i]
    }
# Mau 1988-2007 P_Mau[1:20]
for (i in 1:N_Mau) {
    B_Mau[i] <- P_Mau[i]*K_Mau
    H_Mau[i] <- Catch_Mau[i]/B_Mau[i]
    }
# Hoomalu 1988-2007 P_Hoomalu[1:20]
for (i in 1:N_Hoomalu) {
    B_Hoomalu[i] <- P_Hoomalu[i]*K_Hoomalu
    H_Hoomalu[i] <- Catch_Hoomalu[i]/B_Hoomalu[i]
    }
# Compute reference points
#(9)######################################################
# MHI Reference points
BMSP_MHI <- K_MHI/2
PMSP_MHI <- BMSP_MHI/K_MHI
MSP_M
HMSP`_MHI <- r/2
INDEXMSP_MHI <- q_MHI*BMSP_MHI
# MHI 1949-2007 BSTATUS_MHI and HSTATUS_MHI
for (i in 1:N_MHI) {
    BSTATUS_MHI[i] < - B_MHI[i]/BMSP_MHI
```

Table A2.--continued.

```
    HSTATUS_MHI[i] <- H_MHI[i]/HMSP_MHI
    }
# Mau Reference points
BMSP_Mau <- K_Mau/2
PMSP_Mau <- BMSP_Mau/K_Mau
MSP_Mau <- r*K_Mau/4
HMS\overline{P}MMau <- r/2
INDEXMSP_Mau <- q_Mau*BMSP_Mau
# Mau 1988-2007 BSTATUS_Mau ānd HSTATUS_Mau
for (i in 1:N_Mau) {
    BSTATUS_Mau[i] <- B_Mau[i]/BMSP_Mau
    HSTATUS_Mau[i] <- H_Mau[i]/HMSP_Mau
    }
```

\# Hoomalu Reference points
BMSP_Hoomalu <- K_Hoomalu/2
PMSP_Hoomalu <- BMSP_Hoomalu/K_Hoomalu
MSP_Hoomalu <- r*K_Hoomalu/4
HMSP_Hoomalu <-r/2
INDEXMSP_Hoomalu <- q_Hoomalu*BMSP_Hoomalu
\# Hoomalu 1988-2004 BSTATUS_Hoomalu ānd HSTATUS_Hoomalu
for (i in 1:N Hoomalu) \{
BSTATUS_Hoomalu[i] <- B_Hoomalu[i]/BMSP_Hoomalu
HSTATUS_Hoomalu[i] <- H_Hoomalu[i]/HMSP_Hoomalu
\}
\# Archipelago Reference points
BMSP_Archipelago <- BMSP_MHI + BMSP_Mau + BMSP_Hoomalu
MSP_Archipelago <- MSP_MHI + MSP_Mau + MSP_Hoomalu
HMS $\bar{P}$ _Archipelago <-r/2
K_Archipelago <- K_MHI + K_Mau + K_Hoomalu
\# Archipelago 1988-2007 BSTATUS_Archipelago and HSTATUS_Archipelago
for (i in 1:N_Mau) \{
BSTATUS_Archipelago[i] <-
weight_MHI*BSTATUS_MHI[i+39]+weight_Mau*BSTATUS_Mau[i]+weight_Hoomalu*BSTATUS_Hoomalu[i]
HSTATUS_Archipelago [i] <-
weight_MHI*HSTATUS_MHI[i+39]+weight_Mau*HSTATUS_Mau[i]+weight_Hoomalu*HSTATUS_Hoomalu[i]
\}
\# Compute probabilities of overfishing and overfished in 2007
pH MHIO7<- step(HSTATUS MHI[N MHI]-1.0)
pH_Mau07 <-step(HSTATUS_Mau[N_Mau] - 1.0)
pH_Hoomalu07 <- step(HSTATUS_Hoomalu[N_Hoomalu] - 1.0)
pH_Archipelago07<- step(HSTATŪS_Archipelägo[N_Mau] - 1.0)
pB_MHIO7 <- step(BSTATUS_MHI[N_MHI] - 1.0)
pB_Mau07 <- step(BSTATUS_Mau[N_Mau] - 1.0)
pB_Hoomalu07<-step(BSTATUS_Hoomalu[N_Hoomalu] - 1.0)
pB_Archipelago07 <- step(BSTATUS_Archipelago[N_Mau] - 1.0)
pB7̄_MHIO7 <- step(BSTATUS_MHI[N_MHI] - 0.7)
\# F-Based Projections with MHI Catch Estimate for FY 2008
\#(11)\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#
P_P_MHI[1] <- max(P_MHI[N_MHI]+r*P_MHI[N_MHI]*(1-P_MHI[N_MHI])-Catch_MHI[N_MHI]/K_MHI,0.001)
P_P_-Mau[1] <- $\max \left(\bar{P} \_M a u\left[\bar{N} \_M a u\right]+r * \bar{P} \_M a u\left[\bar{N} \_M a u\right] *\left(1-\bar{P} \_M a u\left[N_{-}\right.\right.\right.$Mau] $\left.)-C a t c h \_M a u\left[N \_M a u\right] / K \_M a u, 0.001\right)$
P_P_Hoomalu[1] <- max (P_Hoomalu[N_Hoomalu] $+r^{*} P_{\text {_ }}$ Hoomalu[N_Hoomalu]*(1-P_Hoomalu[N_Hoomalu])-
Catch_Hoomalu[N_Hoomalu]/K_Hoomalu,0.001)
\# Use estimate of C2008 for MHI projection
P_B_MHI[1] <- P_P_MHI[1]*K_MHI
P_B_Mau[1] <- P_P_Mau[1]*K_Mau
$P_{\text {_B_Hoomalu[1] }}^{\text {- }}-\overline{\text { P_P_Hoomalu[1]*K_Hoomalu }}$
P_C_MHI[1] <- C2008_MHI
P_C_Mau[1] <- P_B_Mau[1]*HPROJ_Mau[1]
P_C_Hoomalu[1] <- P_B_Hoomalu[1]*HPROJ_Hoomalu[1]

Table A2.--continued.

```
    P_H2008_MHI <- C2008_MHI/P_B_MHI[1]
    P HSTAT MHI[1] <- P H2O08 MHH/HMSP MHI
    P_HSTAT_Mau[1] <- HPROJ_Mau[1]/HMSP__Mau
    P_HSTAT_Hoomalu[1] <- HP\overline{ROJ_Hoomalu[1]/HMSP_Hoomalu}
    P_HSTAT_Archipelago[1] <-
weight_MHI*P_HSTAT_MHI[1]+weight_Mau*P_HSTAT_Mau[1]+weight_Hoomalu*P_HSTAT_Hoomalu[1]
    pH_MHI[1] <- step(P_HSTAT_MHI[1] - 1.0)
    pH_Mau[1] <- step(P_HSTAT_Mau[1] - 1.0)
    pH_Hoomalu[1] <- step(P_HSTAT_Hoomalu[1] - 1.0)
    pH_Archipelago[1] <- step(P_HSTAT_Archipelago[1] - 1.0)
    P_BSTAT_MHI[1] <- P_B_MHI[1]/BMSP_MHI
    P_BSTAT_Mau[1] <- P_B_Mau[1]/BMSP_Mau
    P_BSTAT_Hoomalu[1]<--P_B_Hoomalu[1]/BMSP_Hoomalu
    P_BSTAT_Archipelago[1] <-
weight_MHI*P_BSTAT_MHI[1]+weight_Mau*P_BSTAT_Mau[1]+weight_Hoomalu*P_BSTAT_Hoomalu[1]
    pB_MHI[1] <- step(P_BSTAT_MHI[1] - 1.0)
    pB_Mau[1] <- step(P_BSTAT_Mau[1] - 1.0)
    pB_Hoomalu[1] <- step(P_BSTAT_Hoomalu[1] - 1.0)
    pB_Archipelago[1] <- step(P_BSTAT_Archipelago[1] - 1.0)
    pB70_MHI[1] <- step(P_BSTAT_MHI[1] - 0.7)
    P_P_MHI[2] <- P_P_MHI[1]*(1.0+r*(1.0-P_P_MHI[1])-P_H2008_MHI)
    P_P_Mau[2] <- P_P_Mau[1]*(1.0+r*(1.0-P_P_Mau[1])-MPROJ_Mau[1])
    P_P_Hoomalu[2] <- P_P_Hoomalu[1]*(1.0+r**(1.0-P_P_Hoomalu[1])-HPROJ_Hoomalu[1])
# F Status Quo 2007 Based Projections for 2009 and Beyond
# Grid of TAC in 2009 from 50 to 550 by 5
for (j in 1:NTAC)
{
    P_B2009_MHI[j] <- P_P_MHI[2]*K_MHI
    P_C2009_MHI[j] <- TAC_MHI[2]+j-1
    P_H_MHI[j] <- P_C2009_MHI[j]/P_B2009_MHI[j]
    P_HSTAT2009_MHI[j] <- P_H_MHI[j]/HMSP_MHI
    pH2009_MHI[j] <- step(P_HSTAT2009_MHI[j] - 1.0)
}
# END OF CODE
#####################################################################
}
```

Table A3.--Ratio of Deep 7 bottomfish catch to total bottomfish catch in the main Hawaiian Islands by fishing year, 1949-2007 along with recent average ratios.

| MHI | MHI Deep 7 Ratio by Fishing Year, 1949-2007 <br> MHI FY BMUS Catch |  |  |
| :---: | :---: | :---: | :---: |
| Fishing | Kahala = Catch Used in | MHI Deep 7 | Catch Used in |
| Year | the Assessment | Catch (lbs) | Assessment |
| 1949 | 512812 | 354117 | 0.691 |
| 1950 | 431817 | 302676 | 0.701 |
| 1951 | 416819 | 320154 | 0.768 |
| 1952 | 389113 | 299094 | 0.769 |
| 1953 | 375470 | 256895 | 0.684 |
| 1954 | 370070 | 273287 | 0.738 |
| 1955 | 318004 | 223141 | 0.702 |
| 1956 | 382184 | 276124 | 0.722 |
| 1957 | 434718 | 317814 | 0.731 |
| 1958 | 312884 | 211031 | 0.674 |
| 1959 | 293832 | 204004 | 0.694 |
| 1960 | 226944 | 163861 | 0.722 |
| 1961 | 189962 | 129041 | 0.679 |
| 1962 | 237813 | 167839 | 0.706 |
| 1963 | 299509 | 210809 | 0.704 |
| 1964 | 307018 | 201581 | 0.657 |
| 1965 | 317130 | 223807 | 0.706 |
| 1966 | 249043 | 181868 | 0.730 |
| 1967 | 328351 | 231315 | 0.704 |
| 1968 | 273108 | 195039 | 0.714 |
| 1969 | 264002 | 177495 | 0.672 |
| 1970 | 233280 | 158195 | 0.678 |
| 1971 | 203334 | 135189 | 0.665 |
| 1972 | 303987 | 228375 | 0.751 |
| 1973 | 233679 | 169273 | 0.724 |
| 1974 | 326603 | 225767 | 0.691 |
| 1975 | 324690 | 222114 | 0.684 |
| 1976 | 366530 | 258852 | 0.706 |
| 1977 | 363726 | 274882 | 0.756 |
| 1978 | 436206 | 306376 | 0.702 |
| 1979 | 400264 | 273846 | 0.684 |
| 1980 | 343842 | 244278 | 0.710 |
| 1981 | 450492 | 308296 | 0.684 |
| 1982 | 464614 | 329436 | 0.709 |
| 1983 | 579104 | 409948 | 0.708 |
| 1984 | 555910 | 341576 | 0.614 |
| 1985 | 619434 | 485057 | 0.783 |
| 1986 | 621324 | 512075 | 0.824 |
| 1987 | 725632 | 579170 | 0.798 |
| 1988 | 804011 | 566724 | 0.705 |
| 1989 | 964785 | 559538 | 0.580 |
| 1990 | 647051 | 455802 | 0.704 |
| 1991 | 497024 | 324897 | 0.654 |
| 1992 | 493009 | 361617 | 0.733 |
| 1993 | 348334 | 254050 | 0.729 |
| 1994 | 407289 | 307305 | 0.755 |
| 1995 | 458570 | 356485 | 0.777 |
| 1996 | 368267 | 288231 | 0.783 |
| 1997 | 397395 | 299683 | 0.754 |
| 1998 | 381278 | 296755 | 0.778 |
| 1999 | 313286 | 214803 | 0.686 |
| 2000 | 419407 | 309747 | 0.739 |
| 2001 | 348517 | 260267 | 0.747 |
| 2002 | 294996 | 215492 | 0.730 |
| 2003 | 302622 | 244322 | 0.807 |
| 2004 | 279466 | 205129 | 0.734 |
| 2005 | 336920 | 245737 | 0.729 |
| 2006 | 258497 | 186731 | 0.722 |
| 2007 | 309522 | 221576 | 0.716 |


| Recent Averages of MHI Deep 7 Ratio |  |  |  |
| :---: | :---: | :---: | :---: |
| Period | Average | Standard |  |
| (Fishing | MHI Deep | Error of | Standard |
| Year) | 7 Ratio | Mean | Deviation |
| 2000-2007 | 0.741 | 0.0101 | 0.0286 |
| 2001-2007 | 0.741 | 0.0117 | 0.0308 |
| 2002-2007 | 0.740 | 0.0137 | 0.0337 |
| 2003-2007 | 0.742 | 0.0167 | 0.0373 |
| 2004-2007 | 0.725 | 0.0040 | 0.0080 |
| 2005-2007 | 0.723 | 0.0039 | 0.0068 |
| 2006-2007 | 0.719 | 0.0033 | 0.0046 |
| 2007 | 0.716 |  |  |
| 1949-2007 | 0.717 | 0.0057 | 0.0439 |

Table A4.--Distributions of parameter estimates for the baseline bottomfish assessment model fit using standardized MHI CPUE during 1949-2007.

| Variable | Mean | Standard Deviation | MCMC Error | $\begin{gathered} \hline \text { Percentile } \\ 2.5 \% \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 10 \% \\ \hline \end{gathered}$ | Median | $\begin{gathered} \text { Percentile } \\ 90 \% \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 97.5 \% \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BMSP_Hoomalu | 1524 | 104 | 2.81 | 1327 | 1392 | 1521 | 1659 | 1734 |
| BMSP_MHI | 1588 | 108.4 | 2.927 | 1383 | 1451 | 1585 | 1729 | 1806 |
| BMSP_Mau | 439.9 | 30.02 | 0.8109 | 383 | 401.8 | 439 | 478.8 | 500.3 |
| BSTATUS_Archipelago[1] | 1.541 | 0.0923 | $4.77 \mathrm{E}-04$ | 1.361 | 1.422 | 1.54 | 1.659 | 1.724 |
| BSTATUS_Archipelago[2] | 1.549 | 0.09 | 5.40E-04 | 1.378 | 1.435 | 1.547 | 1.664 | 1.731 |
| BSTATUS_Archipelago[3] | 1.368 | 0.07683 | 4.99E-04 | 1.219 | 1.271 | 1.367 | 1.466 | 1.523 |
| BSTATUS_Archipelago[4] | 1.319 | 0.07521 | 5.03E-04 | 1.171 | 1.224 | 1.318 | 1.414 | 1.468 |
| BSTATUS_Archipelago[5] | 1.197 | 0.07192 | 4.77E-04 | 1.054 | 1.105 | 1.197 | 1.288 | 1.339 |
| BSTATUS_Archipelago[6] | 1.234 | 0.07612 | 4.92E-04 | 1.087 | 1.139 | 1.233 | 1.332 | 1.387 |
| BSTATUS_Archipelago[7] | 1.194 | 0.07075 | 4.77E-04 | 1.054 | 1.105 | 1.194 | 1.284 | 1.335 |
| BSTATUS_Archipelago[8] | 1.146 | 0.06789 | 4.65E-04 | 1.012 | 1.06 | 1.145 | 1.232 | 1.28 |
| BSTATUS_Archipelago[9] | 1.075 | 0.06653 | 4.58E-04 | 0.9457 | 0.9908 | 1.074 | 1.159 | 1.208 |
| BSTATUS_Archipelago[10] | 1.095 | 0.0689 | 4.94E-04 | 0.9634 | 1.009 | 1.094 | 1.184 | 1.234 |
| BSTATUS_Archipelago[11] | 1.039 | 0.06851 | 4.90E-04 | 0.9085 | 0.9542 | 1.037 | 1.128 | 1.18 |
| BSTATUS_Archipelago[12] | 1.066 | 0.06698 | 4.80E-04 | 0.9362 | 0.9816 | 1.064 | 1.152 | 1.2 |
| BSTATUS_Archipelago[13] | 1.13 | 0.06848 | 4.76E-04 | 0.996 | 1.044 | 1.129 | 1.217 | 1.267 |
| BSTATUS_Archipelago[14] | 1.08 | 0.0653 | 4.75E-04 | 0.9533 | 0.9979 | 1.08 | 1.164 | 1.21 |
| BSTATUS_Archipelago[15] | 0.997 | 0.06681 | 4.97E-04 | 0.8775 | 0.9167 | 0.9923 | 1.084 | 1.145 |
| BSTATUS_Archipelago[16] | 1.032 | 0.06494 | 4.87E-04 | 0.9122 | 0.9526 | 1.029 | 1.115 | 1.171 |
| BSTATUS_Archipelago[17] | 0.9908 | 0.06738 | 4.97E-04 | 0.8715 | 0.911 | 0.9859 | 1.077 | 1.144 |
| BSTATUS_Archipelago[18] | 1.041 | 0.06898 | 5.07E-04 | 0.9179 | 0.9583 | 1.036 | 1.129 | 1.195 |
| BSTATUS_Archipelago[19] | 1.037 | 0.06974 | 5.00E-04 | 0.9134 | 0.9533 | 1.032 | 1.128 | 1.192 |
| BSTATUS_Archipelago[20] | 1.129 | 0.06871 | 4.58E-04 | 0.9972 | 1.043 | 1.127 | 1.217 | 1.268 |
| BSTATUS_Hoomalu[1] | 2.069 | 0.2038 | 0.001087 | 1.677 | 1.804 | 2.067 | 2.331 | 2.473 |
| BSTATUS_Hoomalu[2] | 2.048 | 0.1907 | 0.001157 | 1.685 | 1.806 | 2.045 | 2.292 | 2.432 |
| BSTATUS_Hoomalu[3] | 1.85 | 0.164 | 0.001143 | 1.529 | 1.645 | 1.849 | 2.057 | 2.179 |
| BSTATUS_Hoomalu[4] | 1.82 | 0.1644 | 0.001196 | 1.497 | 1.614 | 1.82 | 2.027 | 2.147 |
| BSTATUS_Hoomalu[5] | 1.759 | 0.1591 | 0.001148 | 1.443 | 1.557 | 1.759 | 1.959 | 2.072 |
| BSTATUS_Hoomalu[6] | 1.866 | 0.1691 | 0.001188 | 1.538 | 1.656 | 1.863 | 2.083 | 2.207 |
| BSTATUS_Hoomalu[7] | 1.706 | 0.1551 | 0.001104 | 1.4 | 1.511 | 1.706 | 1.902 | 2.015 |
| BSTATUS_Hoomalu[8] | 1.586 | 0.1483 | 0.001088 | 1.297 | 1.399 | 1.585 | 1.773 | 1.881 |
| BSTATUS_Hoomalu[9] | 1.536 | 0.1466 | 0.001088 | 1.252 | 1.352 | 1.534 | 1.722 | 1.831 |
| BSTATUS_Hoomalu[10] | 1.55 | 0.149 | 0.001152 | 1.263 | 1.363 | 1.548 | 1.739 | 1.849 |
| BSTATUS_Hoomalu[11] | 1.479 | 0.1492 | 0.001142 | 1.195 | 1.294 | 1.474 | 1.671 | 1.786 |
| BSTATUS_Hoomalu[12] | 1.485 | 0.1451 | 0.001109 | 1.206 | 1.304 | 1.482 | 1.671 | 1.779 |
| BSTATUS_Hoomalu[13] | 1.548 | 0.1463 | 0.001079 | 1.263 | 1.364 | 1.546 | 1.733 | 1.841 |
| BSTATUS_Hoomalu[14] | 1.434 | 0.1388 | 0.001084 | 1.165 | 1.26 | 1.431 | 1.611 | 1.713 |
| BSTATUS_Hoomalu[15] | 1.219 | 0.144 | 0.001121 | 0.9687 | 1.049 | 1.207 | 1.406 | 1.547 |
| BSTATUS_Hoomalu[16] | 1.305 | 0.1392 | 0.001126 | 1.05 | 1.137 | 1.298 | 1.483 | 1.609 |
| BSTATUS_Hoomalu[17] | 1.252 | 0.1466 | 0.001179 | 0.9969 | 1.081 | 1.239 | 1.437 | 1.595 |
| BSTATUS_Hoomalu[18] | 1.29 | 0.1477 | 0.001186 | 1.031 | 1.117 | 1.279 | 1.478 | 1.633 |
| BSTATUS_Hoomalu[19] | 1.317 | 0.1513 | 0.001171 | 1.051 | 1.138 | 1.304 | 1.511 | 1.662 |
| BSTATUS_Hoomalu[20] | 1.539 | 0.1497 | 0.001135 | 1.252 | 1.352 | 1.535 | 1.73 | 1.845 |
| BSTATUS_MHI[1] | 1.2 | 0.08271 | 7.87E-04 | 1.047 | 1.097 | 1.196 | 1.306 | 1.373 |
| BSTATUS_MHI[2] | 1.242 | 0.08904 | 8.65E-04 | 1.077 | 1.131 | 1.239 | 1.357 | 1.428 |
| BSTATUS_MHI[3] | 1.39 | 0.09963 | 9.70E-04 | 1.205 | 1.266 | 1.386 | 1.518 | 1.598 |
| BSTATUS_MHI[4] | 1.643 | 0.1187 | 0.001124 | 1.422 | 1.495 | 1.639 | 1.797 | 1.889 |
| BSTATUS_MHI[5] | 1.776 | 0.1282 | 0.001198 | 1.538 | 1.617 | 1.771 | 1.94 | 2.043 |
| BSTATUS_MHI[6] | 1.837 | 0.1302 | 0.001223 | 1.595 | 1.676 | 1.832 | 2.006 | 2.108 |
| BSTATUS_MHI[7] | 2.36 | 0.1733 | 0.001592 | 2.032 | 2.142 | 2.355 | 2.582 | 2.716 |
| BSTATUS_MHI[8] | 2.076 | 0.1462 | 0.001406 | 1.804 | 1.894 | 2.07 | 2.264 | 2.379 |
| BSTATUS_MHI[9] | 2.236 | 0.1644 | 0.001523 | 1.925 | 2.031 | 2.232 | 2.446 | 2.573 |
| BSTATUS_MHI[10] | 1.64 | 0.1178 | 0.001123 | 1.423 | 1.494 | 1.635 | 1.792 | 1.886 |
| BSTATUS_MHI[11] | 1.389 | 0.1009 | 9.63E-04 | 1.205 | 1.265 | 1.384 | 1.52 | 1.602 |
| BSTATUS_MHI[12] | 1.714 | 0.1222 | 0.001161 | 1.488 | 1.562 | 1.709 | 1.872 | 1.968 |
| BSTATUS_MHI[13] | 2.329 | 0.1682 | 0.001597 | 2.011 | 2.118 | 2.325 | 2.545 | 2.673 |
| BSTATUS_MHI[14] | 2.435 | 0.1773 | 0.001582 | 2.101 | 2.214 | 2.43 | 2.662 | 2.799 |
| BSTATUS_MHI[15] | 1.86 | 0.1327 | 0.001294 | 1.614 | 1.696 | 1.854 | 2.032 | 2.138 |
| BSTATUS_MHI[16] | 1.872 | 0.133 | 0.001286 | 1.624 | 1.707 | 1.866 | 2.044 | 2.149 |
| BSTATUS_MHI[17] | 2.215 | 0.1595 | 0.001493 | 1.916 | 2.015 | 2.21 | 2.42 | 2.544 |

Table A4.--continued.

| Variable | Mean | Standard Deviation | MCMC Error | $\begin{gathered} \hline \text { Percentile } \\ 2.5 \% \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 10 \% \\ \hline \end{gathered}$ | Median | $\begin{gathered} \hline \text { Percentile } \\ 90 \% \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 97.5 \% \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BSTATUS_MHI[18] | 2.155 | 0.1562 | 0.001455 | 1.862 | 1.959 | 2.15 | 2.355 | 2.477 |
| BSTATUS_MHI[19] | 1.831 | 0.1311 | 0.001236 | 1.589 | 1.669 | 1.826 | 1.999 | 2.105 |
| BSTATUS_MHI[20] | 1.733 | 0.1259 | 0.001168 | 1.496 | 1.577 | 1.728 | 1.895 | 1.993 |
| BSTATUS_MHI[21] | 1.474 | 0.1066 | 0.001033 | 1.277 | 1.342 | 1.47 | 1.611 | 1.698 |
| BSTATUS_MHI[22] | 1.359 | 0.09824 | 9.62E-04 | 1.177 | 1.237 | 1.355 | 1.485 | 1.563 |
| BSTATUS_MHI[23] | 1.333 | 0.09637 | 9.30E-04 | 1.155 | 1.213 | 1.329 | 1.457 | 1.533 |
| BSTATUS_MHI[24] | 1.448 | 0.1049 | 9.95E-04 | 1.253 | 1.317 | 1.444 | 1.583 | 1.666 |
| BSTATUS_MHI[25] | 1.363 | 0.09889 | 9.34E-04 | 1.181 | 1.241 | 1.36 | 1.491 | 1.569 |
| BSTATUS_MHI[26] | 1.378 | 0.09973 | 9.52E-04 | 1.194 | 1.255 | 1.374 | 1.507 | 1.586 |
| BSTATUS_MHI[27] | 1.283 | 0.09304 | 9.06E-04 | 1.112 | 1.168 | 1.279 | 1.404 | 1.477 |
| BSTATUS_MHI[28] | 1.38 | 0.1013 | $9.47 \mathrm{E}-04$ | 1.19 | 1.254 | 1.377 | 1.511 | 1.59 |
| BSTATUS_MHI[29] | 1.101 | 0.07983 | 7.63E-04 | 0.9537 | 1.002 | 1.097 | 1.203 | 1.268 |
| BSTATUS_MHI[30] | 1.145 | 0.0825 | 7.96E-04 | 0.9931 | 1.043 | 1.142 | 1.252 | 1.317 |
| BSTATUS_MHI[31] | 1.17 | 0.08681 | 8.21E-04 | 1.008 | 1.062 | 1.168 | 1.282 | 1.35 |
| BSTATUS_MHI[32] | 0.8875 | 0.06413 | 6.27E-04 | 0.7692 | 0.8083 | 0.8845 | 0.9703 | 1.021 |
| BSTATUS_MHI[33] | 0.8664 | 0.06218 | 6.02E-04 | 0.7514 | 0.7891 | 0.8639 | 0.9468 | 0.9952 |
| BSTATUS_MHI[34] | 0.7918 | 0.05642 | 5.52E-04 | 0.6873 | 0.7223 | 0.7894 | 0.865 | 0.9103 |
| BSTATUS_MHI[35] | 0.7857 | 0.05545 | 5.41E-04 | 0.6832 | 0.7168 | 0.7834 | 0.8573 | 0.9013 |
| BSTATUS_MHI[36] | 0.7132 | 0.0502 | 4.88E-04 | 0.6209 | 0.6515 | 0.7107 | 0.778 | 0.819 |
| BSTATUS_MHI[37] | 0.7306 | 0.05107 | 4.91E-04 | 0.637 | 0.6676 | 0.7282 | 0.7966 | 0.8379 |
| BSTATUS_MHI[38] | 0.7705 | 0.05407 | 5.24E-04 | 0.6714 | 0.7035 | 0.768 | 0.8405 | 0.8843 |
| BSTATUS_MHI[39] | 0.8913 | 0.06182 | 5.91E-04 | 0.7774 | 0.815 | 0.8885 | 0.9709 | 1.022 |
| BSTATUS_MHI[40] | 1.021 | 0.07154 | 6.78E-04 | 0.8901 | 0.9327 | 1.019 | 1.114 | 1.171 |
| BSTATUS_MHI[41] | 0.9944 | 0.0677 | $6.38 \mathrm{E}-04$ | 0.8694 | 0.9105 | 0.9918 | 1.082 | 1.135 |
| BSTATUS_MHI[42] | 0.8034 | 0.05673 | 5.52E-04 | 0.6988 | 0.7329 | 0.8009 | 0.8767 | 0.9219 |
| BSTATUS_MHI[43] | 0.7757 | 0.05684 | 5.54E-04 | 0.6694 | 0.7047 | 0.7738 | 0.8488 | 0.8934 |
| BSTATUS_MHI[44] | 0.648 | 0.04591 | 4.43E-04 | 0.5631 | 0.5911 | 0.646 | 0.7073 | 0.7443 |
| BSTATUS_MHI[45] | 0.6119 | 0.04412 | 4.40E-04 | 0.5305 | 0.5573 | 0.61 | 0.6686 | 0.7045 |
| BSTATUS_MHI[46] | 0.6411 | 0.04586 | 4.39E-04 | 0.5564 | 0.5842 | 0.6393 | 0.7005 | 0.7363 |
| BSTATUS_MHI[47] | 0.6574 | 0.04682 | 4.50E-04 | 0.5709 | 0.5993 | 0.6557 | 0.7176 | 0.7548 |
| BSTATUS_MHI[48] | 0.5757 | 0.04108 | 4.06E-04 | 0.5005 | 0.5249 | 0.5737 | 0.6288 | 0.6613 |
| BSTATUS_MHI[49] | 0.6053 | 0.04328 | 4.21E-04 | 0.5251 | 0.5515 | 0.6035 | 0.6611 | 0.6956 |
| BSTATUS_MHI[50] | 0.57 | 0.04053 | 4.00E-04 | 0.4954 | 0.5198 | 0.5681 | 0.6221 | 0.6548 |
| BSTATUS_MHI[51] | 0.5998 | 0.04318 | 4.24E-04 | 0.5195 | 0.5461 | 0.5983 | 0.6554 | 0.6897 |
| BSTATUS_MHI[52] | 0.6629 | 0.04732 | 4.71E-04 | 0.5745 | 0.6041 | 0.6612 | 0.7239 | 0.761 |
| BSTATUS_MHI[53] | 0.636 | 0.04586 | 4.46E-04 | 0.551 | 0.5791 | 0.6342 | 0.6946 | 0.7318 |
| BSTATUS_MHI[54] | 0.6106 | 0.04425 | 4.37E-04 | 0.5288 | 0.5557 | 0.6089 | 0.6676 | 0.7028 |
| BSTATUS_MHI[55] | 0.5904 | 0.04236 | 4.14E-04 | 0.5115 | 0.538 | 0.5888 | 0.6451 | 0.6786 |
| BSTATUS_MHI[56] | 0.5737 | 0.04129 | 4.10E-04 | 0.4975 | 0.5227 | 0.5719 | 0.6269 | 0.6603 |
| BSTATUS_MHI[57] | 0.6419 | 0.04623 | 4.46E-04 | 0.5556 | 0.5841 | 0.6402 | 0.7012 | 0.7382 |
| BSTATUS_MHI[58] | 0.6058 | 0.04382 | 4.23E-04 | 0.5243 | 0.5515 | 0.6041 | 0.6622 | 0.6969 |
| BSTATUS_MHI[59] | 0.6183 | 0.04548 | 4.26E-04 | 0.5338 | 0.5617 | 0.6164 | 0.677 | 0.713 |
| BSTATUS_Mau[1] | 1.586 | 0.1119 | 5.49E-04 | 1.333 | 1.454 | 1.591 | 1.714 | 1.798 |
| BSTATUS_Mau[2] | 1.821 | 0.2253 | 0.001664 | 1.54 | 1.619 | 1.772 | 2.064 | 2.5 |
| BSTATUS_Mau[3] | 1.735 | 0.1862 | 0.001228 | 1.463 | 1.55 | 1.706 | 1.945 | 2.246 |
| BSTATUS_Mau[4] | 1.54 | 0.137 | 7.65E-04 | 1.24 | 1.362 | 1.549 | 1.702 | 1.791 |
| BSTATUS_Mau[5] | 1.23 | 0.1384 | 9.80E-04 | 0.9139 | 1.042 | 1.245 | 1.391 | 1.466 |
| BSTATUS_Mau[6] | 1.29 | 0.1436 | 9.43E-04 | 0.9754 | 1.098 | 1.302 | 1.462 | 1.544 |
| BSTATUS_Mau[7] | 1.418 | 0.1383 | 8.36E-04 | 1.131 | 1.24 | 1.424 | 1.588 | 1.679 |
| BSTATUS_Mau[8] | 1.382 | 0.1328 | 8.17E-04 | 1.106 | 1.212 | 1.387 | 1.545 | 1.633 |
| BSTATUS_Mau[9] | 1.278 | 0.1229 | 7.56E-04 | 1.036 | 1.123 | 1.279 | 1.431 | 1.523 |
| BSTATUS_Mau[10] | 1.29 | 0.1575 | 0.00111 | 1.028 | 1.113 | 1.275 | 1.478 | 1.676 |
| BSTATUS_Mau[11] | 1.211 | 0.1469 | 0.001019 | 0.9442 | 1.034 | 1.203 | 1.392 | 1.535 |
| BSTATUS_Mau[12] | 1.292 | 0.1423 | 9.21E-04 | 1.013 | 1.112 | 1.292 | 1.471 | 1.574 |
| BSTATUS_Mau[13] | 1.368 | 0.1641 | 0.001161 | 1.016 | 1.149 | 1.378 | 1.567 | 1.664 |
| BSTATUS_Mau[14] | 1.459 | 0.1645 | 0.001127 | 1.097 | 1.241 | 1.472 | 1.656 | 1.751 |
| BSTATUS_Mau[15] | 1.62 | 0.149 | 9.20E-04 | 1.325 | 1.434 | 1.621 | 1.802 | 1.92 |
| BSTATUS_Mau[16] | 1.679 | 0.1563 | 9.62E-04 | 1.399 | 1.497 | 1.669 | 1.869 | 2.03 |
| BSTATUS_Mau[17] | 1.593 | 0.1509 | 9.14E-04 | 1.33 | 1.42 | 1.579 | 1.78 | 1.941 |
| BSTATUS_Mau[18] | 1.616 | 0.1727 | 0.001143 | 1.341 | 1.429 | 1.593 | 1.829 | 2.049 |

Table A4.--continued.

| Variable | Mean | Standard Deviation | MCMC Error | $\begin{gathered} \hline \text { Percentile } \\ 2.5 \% \end{gathered}$ | $\begin{gathered} \text { Percentile } \\ 10 \% \end{gathered}$ | Median | $\begin{gathered} \text { Percentile } \\ 90 \% \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 97.5 \% \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BSTATUS_Mau[19] | 1.627 | 0.1606 | 9.91E-04 | 1.357 | 1.446 | 1.61 | 1.825 | 2.007 |
| BSTATUS_Mau[20] | 1.551 | 0.1375 | 6.63E-04 | 1.293 | 1.386 | 1.546 | 1.723 | 1.845 |
| B_Hoomalu[1] | 3152 | 367.9 | 5.46 | 2460 | 2689 | 3141 | 3628 | 3904 |
| B_Hoomalu[2] | 3121 | 350.4 | 5.464 | 2476 | 2682 | 3106 | 3576 | 3847 |
| B_Hoomalu[3] | 2819 | 305.5 | 4.906 | 2253 | 2439 | 2806 | 3215 | 3460 |
| B_Hoomalu[4] | 2773 | 301.8 | 4.771 | 2212 | 2398 | 2761 | 3165 | 3398 |
| B_Hoomalu[5] | 2680 | 294 | 4.696 | 2134 | 2313 | 2668 | 3060 | 3282 |
| B_Hoomalu[6] | 2844 | 308.9 | 4.751 | 2279 | 2461 | 2830 | 3245 | 3490 |
| B_Hoomalu[7] | 2599 | 283.7 | 4.454 | 2075 | 2246 | 2588 | 2966 | 3188 |
| B_Hoomalu[8] | 2416 | 269.1 | 4.158 | 1920 | 2082 | 2405 | 2765 | 2973 |
| B_Hoomalu[9] | 2340 | 263.1 | 4.023 | 1854 | 2013 | 2328 | 2681 | 2887 |
| B_Hoomalu[10] | 2361 | 264.3 | 3.97 | 1874 | 2032 | 2349 | 2704 | 2909 |
| B_Hoomalu[11] | 2253 | 261.6 | 3.82 | 1774 | 1928 | 2241 | 2592 | 2801 |
| B_Hoomalu[12] | 2263 | 256.8 | 3.803 | 1789 | 1943 | 2252 | 2597 | 2794 |
| B_Hoomalu[13] | 2358 | 261 | 3.895 | 1874 | 2032 | 2348 | 2695 | 2901 |
| B_Hoomalu[14] | 2184 | 245.1 | 3.609 | 1733 | 1879 | 2174 | 2502 | 2693 |
| B_Hoomalu[15] | 1858 | 243 | 3.258 | 1436 | 1565 | 1838 | 2176 | 2395 |
| B_Hoomalu[16] | 1988 | 237.2 | 3.259 | 1563 | 1697 | 1975 | 2295 | 2498 |
| B_Hoomalu[17] | 1906 | 245.9 | 3.259 | 1481 | 1613 | 1886 | 2226 | 2460 |
| B_Hoomalu[18] | 1965 | 248.4 | 3.307 | 1534 | 1668 | 1945 | 2289 | 2523 |
| B_Hoomalu[19] | 2006 | 256 | 3.447 | 1557 | 1698 | 1985 | 2341 | 2573 |
| B_Hoomalu[20] | 2344 | 262.3 | 3.804 | 1863 | 2017 | 2333 | 2685 | 2891 |
| B_MHI[1] | 1905 | 187.4 | 4.037 | 1565 | 1673 | 1895 | 2150 | 2302 |
| B_MHI[2] | 1973 | 198.1 | 4.238 | 1616 | 1728 | 1962 | 2232 | 2393 |
| B_MHI[3] | 2207 | 221.9 | 4.745 | 1807 | 1933 | 2194 | 2497 | 2677 |
| B_MHI[4] | 2610 | 261.8 | 5.542 | 2136 | 2288 | 2595 | 2952 | 3164 |
| B_MHI[5] | 2820 | 283 | 5.993 | 2312 | 2471 | 2803 | 3189 | 3419 |
| B_MHI[6] | 2918 | 293.6 | 6.289 | 2391 | 2556 | 2901 | 3302 | 3540 |
| B_MHI[7] | 3747 | 378.3 | 7.909 | 3063 | 3280 | 3728 | 4239 | 4546 |
| B_MHI[8] | 3297 | 331.8 | 7.147 | 2697 | 2887 | 3277 | 3731 | 4.00E+03 |
| B_MHI[9] | 3550 | 356.7 | 7.397 | 2903 | 3109 | 3533 | 4015 | 4304 |
| B_MHI[10] | 2605 | 263.3 | 5.617 | 2131 | 2281 | 2589 | 2949 | 3166 |
| B_MHI[11] | 2206 | 226 | 4.811 | 1802 | 1929 | 2193 | 2502 | 2690 |
| B_MHI[12] | 2722 | 273 | 5.834 | 2230 | 2385 | 2706 | 3077 | 3302 |
| B_MHI[13] | 3699 | 370.2 | 7.844 | 3027 | 3241 | 3680 | 4181 | 4483 |
| B_MHI[14] | 3867 | 387.5 | 8.081 | 3164 | 3390 | 3847 | 4369 | 4688 |
| B_MHI[15] | 2955 | 299.8 | 6.473 | 2415 | 2585 | 2937 | 3346 | 3587 |
| B_MHI[16] | 2973 | 299.3 | 6.416 | 2435 | 2603 | 2956 | 3362 | 3608 |
| B_MHI[17] | 3517 | 352.7 | 7.458 | 2880 | 3084 | 3497 | 3977 | 4264 |
| B_MHI[18] | 3422 | 344.9 | 7.272 | 2797 | 2995 | 3403 | 3872 | 4150 |
| B_MHI[19] | 2908 | 292.8 | 6.229 | 2380 | 2549 | 2891 | 3289 | 3530 |
| B_MHI[20] | 2752 | 276.5 | 5.827 | 2253 | 2411 | 2737 | 3113 | 3336 |
| B_MHI[21] | 2341 | 236.6 | 5.04 | 1916 | 2049 | 2328 | 2649 | 2843 |
| B_MHI[22] | 2158 | 217.3 | 4.657 | 1766 | 1890 | 2146 | 2440 | 2621 |
| B_MHI[23] | 2117 | 213.8 | 4.569 | 1733 | 1853 | 2105 | 2396 | 2569 |
| B_MHI[24] | 2299 | 230.3 | 4.82 | 1884 | 2015 | 2287 | 2599 | 2787 |
| B_MHI[25] | 2165 | 218.4 | 4.619 | 1771 | 1897 | 2152 | 2450 | 2629 |
| B_MHI[26] | 2189 | 219.4 | 4.651 | 1791 | 1918 | 2177 | 2475 | 2654 |
| B_MHI[27] | 2038 | 206.5 | 4.412 | 1668 | 1784 | 2026 | 2308 | 2477 |
| B_MHI[28] | 2192 | 220 | 4.574 | 1796 | 1921 | 2179 | 2479 | 2660 |
| B_MHI[29] | 1748 | 177.5 | 3.777 | 1430 | 1530 | 1737 | 1979 | 2126 |
| B_MHI[30] | 1819 | 182.5 | 3.874 | 1490 | 1593 | 1808 | 2057 | 2204 |
| B_MHI[31] | 1859 | 187.3 | 3.896 | 1520 | 1628 | 1849 | 2103 | 2255 |
| B_MHI[32] | 1409 | 142.9 | 3.047 | 1153 | 1234 | 1401 | 1596 | 1714 |
| B_MHI[33] | 1376 | 137.8 | 2.94 | 1128 | 1206 | 1368 | 1556 | 1668 |
| B_MHI[34] | 1257 | 125.9 | 2.709 | 1031 | 1103 | 1250 | 1421 | 1526 |
| B_MHI[35] | 1248 | 123.3 | 2.641 | 1025 | 1096 | 1240 | 1409 | 1510 |
| B_MHI[36] | 1133 | 112.4 | 2.417 | 930.5 | 994.2 | 1126 | 1279 | 1372 |
| B_MHI[37] | 1160 | 114.2 | 2.453 | 956 | 1020 | 1154 | 1308 | 1405 |
| B_MHI[38] | 1224 | 120.9 | 2.591 | 1007 | 1074 | 1217 | 1381 | 1481 |

Table A4.--continued.

| Variable | Mean | Standard Deviation | MCMC Error | $\begin{gathered} \hline \text { Percentile } \\ 2.5 \% \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 10 \% \\ \hline \end{gathered}$ | Median | $\begin{gathered} \hline \text { Percentile } \\ 90 \% \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 97.5 \% \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| B_MHI[39] | 1415 | 138.2 | 2.966 | 1167 | 1244 | 1407 | 1596 | 1708 |
| B_MHI[40] | 1622 | 158.4 | 3.37 | 1337 | 1427 | 1612 | 1829 | 1959 |
| B_MHI[41] | 1579 | 150.3 | 3.199 | 1309 | 1394 | 1570 | 1775 | 1899 |
| B_MHI[42] | 1276 | 127.2 | 2.736 | 1047 | 1119 | 1268 | 1442 | 1547 |
| B_MHI[43] | 1232 | 124.4 | 2.627 | 1007 | 1078 | 1225 | 1394 | 1495 |
| B_MHI[44] | 1029 | 102.9 | 2.218 | 843.7 | 901.9 | 1023 | 1163 | 1248 |
| B_MHI[45] | 971.9 | 98.28 | 2.105 | 795.7 | 850.8 | 966.1 | 1100 | 1181 |
| B_MHI[46] | 1018 | 101.6 | 2.153 | 835.5 | 892.9 | 1012 | 1151 | 1234 |
| B_MHI[47] | 1044 | 103.5 | 2.203 | 857 | 916.8 | 1037 | 1179 | 1263 |
| B_MHI[48] | 914.3 | 92.08 | 1.975 | 748.7 | 801 | 909 | 1034 | 1110 |
| B_MHI[49] | 961.2 | 95.77 | 2.041 | 788.2 | 842.8 | 956 | 1086 | 1164 |
| B_MHI[50] | 905.2 | 90.78 | 1.953 | 742.1 | 794 | 899.9 | 1024 | 1098 |
| B_MHI[51] | 952.6 | 96.01 | 2.051 | 780.3 | 834.3 | 946.9 | 1077 | 1158 |
| B_MHI[52] | 1053 | 104.6 | 2.239 | 864 | 923.4 | 1047 | 1188 | 1275 |
| B_MHI[53] | 1010 | 101.6 | 2.156 | 826.3 | 884.9 | 1004 | 1143 | 1226 |
| B_MHI[54] | 969.7 | 97.82 | 2.081 | 794.1 | 849.6 | 963.9 | 1097 | 1178 |
| B_MHI[55] | 937.6 | 94.1 | 2.013 | 768.8 | 821.3 | 932.6 | 1061 | 1137 |
| B_MHI[56] | 911.1 | 91.84 | 1.971 | 745.1 | 798.1 | 905.8 | 1031 | 1107 |
| B_MHI[57] | 1019 | 101.8 | 2.161 | 836.5 | 893.5 | 1014 | 1152 | 1235 |
| B_MHI[58] | 962.1 | 97.09 | 2.061 | 787.8 | 842.7 | 956.5 | 1089 | 1168 |
| B_MHI[59] | 981.8 | 99.55 | 2.087 | 803.2 | 858.9 | 975.7 | 1112 | 1193 |
| B_Mau[1] | 697.5 | 69.33 | 1.345 | 560.3 | 612.1 | 697 | 784.5 | 836.3 |
| B_Mau[2] | 800.5 | 110.6 | 1.495 | 643.9 | 688.7 | 781.8 | 927.1 | 1102 |
| B_Mau[3] | 763.1 | 95.8 | 1.417 | 612.9 | 658.2 | 750.7 | 878.3 | $1.00 \mathrm{E}+03$ |
| B_Mau[4] | 677.5 | 76.52 | 1.299 | 525.3 | 581 | 677.4 | 774.2 | 829 |
| B_Mau[5] | 541.3 | 74.93 | 1.238 | 388.2 | 444.6 | 543 | 635 | 686.5 |
| B_Mau[6] | 567.8 | 75.87 | 1.189 | 415.2 | 470 | 568.9 | 663.4 | 715 |
| B_Mau[7] | 623.8 | 73.77 | 1.159 | 481.5 | 530.8 | 623.1 | 717.6 | 772.5 |
| B_Mau[8] | 608 | 71.85 | 1.157 | 469.8 | 517.2 | 606.9 | 699.5 | 753.2 |
| B_Mau[9] | 562.4 | 68.24 | 1.141 | 436.3 | 477.2 | 559.8 | 650.2 | 704.1 |
| B_Mau[10] | 567.8 | 80.58 | 1.182 | 432.6 | 473.5 | 560.2 | 669.6 | 753.4 |
| B_Mau[11] | 532.8 | 76.2 | 1.144 | 397.8 | 440.2 | 528 | 630.5 | 698.5 |
| B_Mau[12] | 568.4 | 74.33 | 1.118 | 430.2 | 475.4 | 565.9 | 663.4 | 722.5 |
| B_Mau[13] | 601.5 | 82.64 | 1.17 | 433.8 | 494.9 | 602.9 | 704.9 | 761.2 |
| B_Mau[14] | 641.5 | 83.55 | 1.207 | 469.7 | 533.9 | 643.5 | 745.9 | 801.6 |
| B_Mau[15] | 712.4 | 79.6 | 1.25 | 563.5 | 614.4 | 709.3 | 813.6 | 878.2 |
| B_Mau[16] | 738.5 | 84.1 | 1.338 | 590.7 | 638.3 | 732.3 | 845.3 | 923.3 |
| B_Mau[17] | 700.6 | 83.24 | 1.388 | 556.8 | 601.5 | 693.9 | 807.7 | 886.1 |
| B_Mau[18] | 710.9 | 90.94 | 1.415 | 561.6 | 606.2 | 700.7 | 826.8 | 923.3 |
| B_Mau[19] | 715.6 | 86.48 | 1.387 | 568.7 | 614.1 | 707.5 | 826.4 | 911.3 |
| B_Mau[20] | 682.4 | 77.97 | 1.34 | 542.2 | 587.6 | 677.8 | 782.8 | 850.2 |
| HMSP_Hoomalu | 0.2885 | 0.02886 | 5.82E-04 | 0.233 | 0.2518 | 0.2883 | 0.3255 | 0.346 |
| HMSP_MHI | 0.2885 | 0.02886 | 5.82E-04 | 0.233 | 0.2518 | 0.2883 | 0.3255 | 0.346 |
| HMSP_Mau | 0.2885 | 0.02886 | 5.82E-04 | 0.233 | 0.2518 | 0.2883 | 0.3255 | 0.346 |
| HSTATUS_Archipelago[1] | 0.8925 | 0.0712 | 3.32E-04 | 0.7634 | 0.8048 | 0.8888 | 0.9849 | 1.043 |
| HSTATUS_Archipelago[2] | 1.051 | 0.08291 | 4.02E-04 | 0.9003 | 0.9484 | 1.046 | 1.159 | 1.226 |
| HSTATUS_Archipelago[3] | 0.9128 | 0.07296 | 3.60E-04 | 0.7795 | 0.8227 | 0.9087 | 1.008 | 1.067 |
| HSTATUS_Archipelago[4] | 0.8726 | 0.07061 | 3.73E-04 | 0.7459 | 0.7862 | 0.8681 | 0.9641 | 1.023 |
| HSTATUS_Archipelago[5] | 0.9284 | 0.07497 | $3.77 \mathrm{E}-04$ | 0.7931 | 0.8361 | 0.9243 | 1.026 | 1.087 |
| HSTATUS_Archipelago[6] | 0.7366 | 0.06044 | 3.10E-04 | 0.6287 | 0.663 | 0.733 | 0.8154 | 0.866 |
| HSTATUS_Archipelago[7] | 0.8776 | 0.07241 | $4.01 \mathrm{E}-04$ | 0.7495 | 0.7894 | 0.8727 | 0.972 | 1.033 |
| HSTATUS_Archipelago[8] | 0.9652 | 0.0793 | $4.42 \mathrm{E}-04$ | 0.8237 | 0.8689 | 0.96 | 1.068 | 1.136 |
| HSTATUS_Archipelago[9] | 0.8515 | 0.07065 | 3.83E-04 | 0.7252 | 0.7653 | 0.8471 | 0.9432 | 1.003 |
| HSTATUS_Archipelago[10] | 0.8913 | 0.07527 | $4.24 \mathrm{E}-04$ | 0.7576 | 0.7994 | 0.8863 | 0.9898 | 1.052 |
| HSTATUS_Archipelago[11] | 0.8825 | 0.07404 | 4.14E-04 | 0.7499 | 0.7925 | 0.8776 | 0.9792 | 1.041 |
| HSTATUS_Archipelago[12] | 0.7644 | 0.06572 | 3.98E-04 | 0.6481 | 0.6847 | 0.7597 | 0.8499 | 0.9062 |
| HSTATUS_Archipelago[13] | 0.8294 | 0.06946 | 3.79E-04 | 0.7055 | 0.7447 | 0.825 | 0.9197 | 0.9787 |
| HSTATUS_Archipelago[14] | 0.7358 | 0.06189 | $3.48 \mathrm{E}-04$ | 0.6258 | 0.6604 | 0.7317 | 0.8165 | 0.8688 |
| HSTATUS_Archipelago[15] | 0.6513 | 0.0549 | 3.17E-04 | 0.5532 | 0.5843 | 0.6479 | 0.7229 | 0.7689 |
| HSTATUS_Archipelago[16] | 0.6918 | 0.05804 | 3.19E-04 | 0.5888 | 0.6211 | 0.688 | 0.7671 | 0.8165 |

Table A4.--continued.

| Variable | Mean | Standard Deviation | MCMC Error | $\begin{gathered} \hline \text { Percentile } \\ 2.5 \% \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 10 \% \\ \hline \end{gathered}$ | Median | $\begin{gathered} \hline \text { Percentile } \\ 90 \% \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 97.5 \% \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| HSTATUS_Archipelago[17] | 0.6556 | 0.05513 | 3.07E-04 | 0.5568 | 0.5883 | 0.6522 | 0.7273 | 0.7733 |
| HSTATUS_Archipelago[18] | 0.702 | 0.05892 | 3.35E-04 | 0.5973 | 0.6302 | 0.6983 | 0.7784 | 0.8282 |
| HSTATUS_Archipelago[19] | 0.5797 | 0.04874 | 2.69E-04 | 0.4927 | 0.5202 | 0.5765 | 0.643 | 0.6838 |
| HSTATUS_Archipelago[20] | 0.6191 | 0.05354 | 2.73E-04 | 0.5233 | 0.5537 | 0.6156 | 0.6887 | 0.7339 |
| HSTATUS_Hoomalu[1] | 0.2185 | 0.02976 | 2.12E-04 | 0.1687 | 0.1829 | 0.2156 | 0.2579 | 0.284 |
| HSTATUS_Hoomalu[2] | 0.1154 | 0.01553 | 1.22E-04 | 0.08973 | 0.09709 | 0.1138 | 0.1355 | 0.1501 |
| HSTATUS_Hoomalu[3] | 0.1496 | 0.01964 | $1.69 \mathrm{E}-04$ | 0.1177 | 0.1268 | 0.1473 | 0.1749 | 0.1942 |
| HSTATUS_Hoomalu[4] | 0.2618 | 0.03497 | 3.08E-04 | 0.205 | 0.2214 | 0.2578 | 0.3067 | 0.3419 |
| HSTATUS_Hoomalu[5] | 0.2587 | 0.03459 | 3.03E-04 | 0.2028 | 0.2189 | 0.2545 | 0.3035 | 0.3376 |
| HSTATUS_Hoomalu[6] | 0.2949 | 0.03987 | 3.41E-04 | 0.23 | 0.2488 | 0.2903 | 0.3466 | 0.3856 |
| HSTATUS_Hoomalu[7] | 0.3924 | 0.05289 | $4.58 \mathrm{E}-04$ | 0.3065 | 0.3313 | 0.3863 | 0.4608 | 0.5126 |
| HSTATUS_Hoomalu[8] | 0.3689 | 0.05064 | $4.48 \mathrm{E}-04$ | 0.2866 | 0.3102 | 0.3632 | 0.4343 | 0.484 |
| HSTATUS_Hoomalu[9] | 0.2626 | 0.03659 | 3.24E-04 | 0.2031 | 0.22 | 0.2585 | 0.31 | 0.3455 |
| HSTATUS_Hoomalu[10] | 0.2988 | 0.04195 | 3.79E-04 | 0.2303 | 0.2499 | 0.2942 | 0.3531 | 0.3938 |
| HSTATUS_Hoomalu[11] | 0.392 | 0.05587 | 5.03E-04 | 0.2996 | 0.3268 | 0.3861 | 0.4642 | 0.5176 |
| HSTATUS_Hoomalu[12] | 0.4927 | 0.06938 | 6.22E-04 | 0.379 | 0.4114 | 0.4852 | 0.5823 | 0.6497 |
| HSTATUS_Hoomalu[13] | 0.3757 | 0.05296 | 4.62E-04 | 0.2899 | 0.3144 | 0.3695 | 0.444 | 0.4965 |
| HSTATUS_Hoomalu[14] | 0.3668 | 0.05216 | $4.71 \mathrm{E}-04$ | 0.2817 | 0.3063 | 0.3611 | 0.4341 | 0.4854 |
| HSTATUS_Hoomalu[15] | 0.2949 | 0.04501 | $4.04 \mathrm{E}-04$ | 0.2158 | 0.2414 | 0.2913 | 0.3527 | 0.3931 |
| HSTATUS_Hoomalu[16] | 0.2571 | 0.03804 | $3.48 \mathrm{E}-04$ | 0.1928 | 0.2127 | 0.2535 | 0.3063 | 0.342 |
| HSTATUS_Hoomalu[17] | 0.2658 | 0.0405 | $3.71 \mathrm{E}-04$ | 0.1937 | 0.218 | 0.2625 | 0.3173 | 0.355 |
| HSTATUS_Hoomalu[18] | 0.3085 | 0.04666 | 4.27E-04 | 0.2264 | 0.2536 | 0.3045 | 0.368 | 0.4119 |
| HSTATUS_Hoomalu[19] | 0.2198 | 0.03323 | 3.00E-04 | 0.1614 | 0.1805 | 0.2172 | 0.2621 | 0.2932 |
| HSTATUS_Hoomalu[20] | 0.2541 | 0.0361 | $3.21 \mathrm{E}-04$ | 0.1949 | 0.212 | 0.2501 | 0.3004 | 0.336 |
| HSTATUS_MHI[1] | 0.9454 | 0.07811 | $3.74 \mathrm{E}-04$ | 0.803 | 0.8484 | 0.9417 | 1.047 | 1.109 |
| HSTATUS_MHI[2] | 0.7691 | 0.06514 | $3.20 \mathrm{E}-04$ | 0.6507 | 0.6884 | 0.7661 | 0.8536 | 0.9067 |
| HSTATUS_MHI[3] | 0.6636 | 0.05618 | 2.71E-04 | 0.5611 | 0.5941 | 0.6606 | 0.7369 | 0.7812 |
| HSTATUS_MHI[4] | 0.5239 | 0.0445 | 2.15E-04 | 0.4434 | 0.4689 | 0.5215 | 0.5816 | 0.6177 |
| HSTATUS_MHI[5] | 0.4679 | 0.03955 | $1.90 \mathrm{E}-04$ | 0.3956 | 0.4192 | 0.4659 | 0.5192 | 0.5513 |
| HSTATUS_MHI[6] | 0.4456 | 0.03729 | $1.73 \mathrm{E}-04$ | 0.3778 | 0.3993 | 0.4437 | 0.4941 | 0.5244 |
| HSTATUS_MHI[7] | 0.2982 | 0.02553 | $1.22 \mathrm{E}-04$ | 0.2525 | 0.2668 | 0.2966 | 0.3315 | 0.3529 |
| HSTATUS_MHI[8] | 0.4073 | 0.03376 | $1.62 \mathrm{E}-04$ | 0.3462 | 0.3654 | 0.4057 | 0.4511 | 0.4786 |
| HSTATUS_MHI[9] | 0.4303 | 0.03704 | $1.74 \mathrm{E}-04$ | 0.3636 | 0.3848 | 0.4281 | 0.4785 | 0.5096 |
| HSTATUS_MHI[10] | 0.4221 | 0.03565 | 1.67E-04 | 0.3569 | 0.3781 | 0.4204 | 0.4685 | 0.4971 |
| HSTATUS_MHI[11] | 0.4681 | 0.04012 | $1.97 \mathrm{E}-04$ | 0.3946 | 0.4182 | 0.4662 | 0.5201 | 0.5524 |
| HSTATUS_MHI[12] | 0.293 | 0.02473 | 1.19E-04 | 0.2481 | 0.2623 | 0.2918 | 0.325 | 0.3451 |
| HSTATUS_MHI[13] | 0.1804 | 0.01527 | $7.41 \mathrm{E}-05$ | 0.1531 | 0.1616 | 0.1796 | 0.2003 | 0.2129 |
| HSTATUS_MHI[14] | 0.2161 | 0.01841 | 8.46E-05 | 0.1831 | 0.1934 | 0.2151 | 0.2402 | 0.2553 |
| HSTATUS_MHI[15] | 0.3562 | 0.02993 | $1.46 \mathrm{E}-04$ | 0.3013 | 0.3192 | 0.3547 | 0.3951 | 0.4191 |
| HSTATUS_MHI[16] | 0.3629 | 0.03047 | $1.45 \mathrm{E}-04$ | 0.3073 | 0.3252 | 0.3614 | 0.4025 | 0.4271 |
| HSTATUS_MHI[17] | 0.3168 | 0.02681 | $1.27 \mathrm{E}-04$ | 0.2686 | 0.2837 | 0.3154 | 0.3517 | 0.3739 |
| HSTATUS_MHI[18] | 0.2557 | 0.02177 | $1.03 \mathrm{E}-04$ | 0.2164 | 0.2289 | 0.2546 | 0.2839 | 0.3022 |
| HSTATUS_MHI[19] | 0.3967 | 0.03345 | $1.60 \mathrm{E}-04$ | 0.3357 | 0.3553 | 0.395 | 0.4402 | 0.4671 |
| HSTATUS_MHI[20] | 0.3487 | 0.02986 | $1.41 \mathrm{E}-04$ | 0.2948 | 0.3118 | 0.3471 | 0.3877 | 0.4118 |
| HSTATUS_MHI[21] | 0.3963 | 0.03368 | $1.65 \mathrm{E}-04$ | 0.3347 | 0.3548 | 0.3947 | 0.4397 | 0.4672 |
| HSTATUS_MHI[22] | 0.3798 | 0.0325 | $1.59 \mathrm{E}-04$ | 0.3206 | 0.3396 | 0.3782 | 0.422 | 0.4482 |
| HSTATUS_MHI[23] | 0.3375 | 0.02887 | $1.41 \mathrm{E}-04$ | 0.285 | 0.3019 | 0.3361 | 0.375 | 0.3985 |
| HSTATUS_MHI[24] | 0.4646 | 0.03972 | $1.91 \mathrm{E}-04$ | 0.3925 | 0.4158 | 0.4625 | 0.5162 | 0.5492 |
| HSTATUS_MHI[25] | 0.3793 | 0.03249 | $1.52 \mathrm{E}-04$ | 0.3203 | 0.3391 | 0.3775 | 0.4215 | 0.4479 |
| HSTATUS_MHI[26] | 0.5243 | 0.04482 | 2.20E-04 | 0.4427 | 0.4688 | 0.5221 | 0.5824 | 0.6191 |
| HSTATUS_MHI[27] | 0.5598 | 0.04789 | 2.37E-04 | 0.4725 | 0.5004 | 0.5574 | 0.6217 | 0.6613 |
| HSTATUS_MHI[28] | 0.5876 | 0.05049 | $2.38 \mathrm{E}-04$ | 0.4961 | 0.5256 | 0.5848 | 0.6532 | 0.6944 |
| HSTATUS_MHI[29] | 0.7313 | 0.06262 | 3.03E-04 | 0.6166 | 0.6539 | 0.7281 | 0.8122 | 0.8626 |
| HSTATUS_MHI[30] | 0.8428 | 0.07164 | $3.43 \mathrm{E}-04$ | 0.7127 | 0.7542 | 0.8393 | 0.9354 | 0.9937 |
| HSTATUS_MHI[31] | 0.7568 | 0.06505 | 3.07E-04 | 0.6395 | 0.6766 | 0.7532 | 0.8418 | 0.8937 |
| HSTATUS_MHI[32] | 0.8572 | 0.07292 | 3.62E-04 | 0.7242 | 0.7671 | 0.8535 | 0.9522 | 1.01 |
| HSTATUS_MHI[33] | 1.15 | 0.09668 | $4.71 \mathrm{E}-04$ | 0.9747 | 1.031 | 1.146 | 1.276 | 1.353 |
| HSTATUS_MHI[34] | 1.298 | 0.1088 | 5.42E-04 | 1.099 | 1.164 | 1.293 | 1.439 | 1.529 |
| HSTATUS_MHI[35] | 1.63 | 0.1353 | $6.50 \mathrm{E}-04$ | 1.384 | 1.463 | 1.623 | 1.807 | 1.916 |
| HSTATUS_MHI[36] | 1.724 | 0.1429 | 6.87E-04 | 1.461 | 1.547 | 1.717 | 1.909 | 2.024 |

Table A4.--continued.

| Variable | Mean | Standard Deviation | MCMC Error | $\begin{gathered} \hline \text { Percentile } \\ 2.5 \% \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 10 \% \end{gathered}$ | Median | $\begin{gathered} \text { Percentile } \\ 90 \% \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 97.5 \% \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| HSTATUS_MHI[37] | 1.875 | 0.1544 | 7.38E-04 | 1.592 | 1.684 | 1.867 | 2.076 | 2.2 |
| HSTATUS_MHI[38] | 1.784 | 0.1475 | 7.05E-04 | 1.513 | 1.601 | 1.777 | 1.974 | 2.095 |
| HSTATUS_MHI[39] | 1.801 | 0.1475 | 7.14E-04 | 1.53 | 1.619 | 1.794 | 1.991 | 2.113 |
| HSTATUS_MHI[40] | 1.741 | 0.1428 | 6.76E-04 | 1.479 | 1.564 | 1.734 | 1.926 | 2.042 |
| HSTATUS_MHI[41] | 2.145 | 0.1721 | 8.24E-04 | 1.831 | 1.932 | 2.136 | 2.368 | 2.506 |
| HSTATUS_MHI[42] | 1.781 | 0.1462 | 7.22E-04 | 1.513 | 1.6 | 1.774 | 1.971 | 2.089 |
| HSTATUS_MHI[43] | 1.418 | 0.1197 | 5.98E-04 | 1.2 | 1.27 | 1.412 | 1.573 | 1.671 |
| HSTATUS_MHI[44] | 1.683 | 0.1399 | 6.70E-04 | 1.426 | 1.509 | 1.677 | 1.864 | 1.975 |
| HSTATUS_MHI[45] | 1.259 | 0.1056 | 5.24E-04 | 1.066 | 1.129 | 1.254 | 1.397 | 1.48 |
| HSTATUS_MHI[46] | 1.405 | 0.1174 | 5.67E-04 | 1.191 | 1.26 | 1.399 | 1.557 | 1.652 |
| HSTATUS_MHI[47] | 1.543 | 0.128 | 6.18E-04 | 1.31 | 1.384 | 1.537 | 1.709 | 1.814 |
| HSTATUS_MHI[48] | 1.415 | 0.1184 | 5.85E-04 | 1.197 | 1.269 | 1.409 | 1.568 | 1.664 |
| HSTATUS_MHI[49] | 1.452 | 0.121 | 5.91E-04 | 1.232 | 1.303 | 1.447 | 1.609 | 1.707 |
| HSTATUS_MHI[50] | 1.48 | 0.1234 | 6.06E-04 | 1.254 | 1.327 | 1.474 | 1.64 | 1.738 |
| HSTATUS_MHI[51] | 1.155 | 0.09694 | 4.81E-04 | 0.9784 | 1.036 | 1.151 | 1.281 | 1.36 |
| HSTATUS_MHI[52] | 1.4 | 0.1168 | 5.87E-04 | 1.186 | 1.255 | 1.394 | 1.551 | 1.646 |
| HSTATUS_MHI[53] | 1.212 | 0.1017 | 5.01E-04 | 1.027 | 1.087 | 1.207 | 1.344 | 1.426 |
| HSTATUS_MHI[54] | 1.069 | 0.09 | 4.47E-04 | 0.9051 | 0.9577 | 1.064 | 1.186 | 1.26 |
| HSTATUS_MHI[55] | 1.134 | 0.09542 | 4.64E-04 | 0.9595 | 1.016 | 1.129 | 1.258 | 1.336 |
| HSTATUS_MHI[56] | 1.078 | 0.09104 | 4.51E-04 | 0.9107 | 0.9647 | 1.074 | 1.196 | 1.269 |
| HSTATUS_MHI[57] | 1.161 | 0.0974 | 4.76E-04 | 0.9843 | 1.041 | 1.157 | 1.288 | 1.367 |
| HSTATUS_MHI[58] | 0.9441 | 0.08001 | 3.86E-04 | 0.798 | 0.8452 | 0.9402 | 1.048 | 1.113 |
| HSTATUS_MHI[59] | 1.108 | 0.09783 | 4.61E-04 | 0.9305 | 0.9873 | 1.103 | 1.235 | 1.316 |
| HSTATUS_Mau[1] | 0.166 | 0.01789 | 1.22E-04 | 0.1364 | 0.1454 | 0.1642 | 0.1886 | 0.2064 |
| HSTATUS_Mau[2] | 0.343 | 0.04773 | 3.72E-04 | 0.2376 | 0.2846 | 0.3437 | 0.401 | 0.4365 |
| HSTATUS_Mau[3] | 0.4221 | 0.05694 | 4.25E-04 | 0.308 | 0.3534 | 0.4204 | 0.4933 | 0.5405 |
| HSTATUS_Mau[4] | 1.02 | 0.1336 | 9.61E-04 | 0.8059 | 0.8674 | 1.004 | 1.193 | 1.326 |
| HSTATUS_Mau[5] | 0.5249 | 0.08184 | 6.12E-04 | 0.4015 | 0.4341 | 0.5122 | 0.6317 | 0.7181 |
| HSTATUS_Mau[6] | 0.3807 | 0.05934 | 4.44E-04 | 0.2887 | 0.3136 | 0.3723 | 0.4581 | 0.5193 |
| HSTATUS_Mau[7] | 0.6538 | 0.09399 | 7.16E-04 | 0.5019 | 0.5447 | 0.6428 | 0.7771 | 0.8673 |
| HSTATUS_Mau[8] | 0.9453 | 0.134 | 0.001024 | 0.7273 | 0.7894 | 0.9296 | 1.121 | 1.249 |
| HSTATUS_Mau[9] | 0.857 | 0.1232 | 9.45E-04 | 0.6554 | 0.7129 | 0.843 | 1.018 | 1.135 |
| HSTATUS_Mau[10] | 0.9191 | 0.1489 | 0.001172 | 0.6577 | 0.7428 | 0.9063 | 1.112 | 1.248 |
| HSTATUS_Mau[11] | 0.426 | 0.07228 | 5.69E-04 | 0.3081 | 0.342 | 0.4178 | 0.5201 | 0.5921 |
| HSTATUS_Mau[12] | 0.2946 | 0.04821 | 3.78E-04 | 0.2183 | 0.2394 | 0.2883 | 0.3579 | 0.4063 |
| HSTATUS_Mau[13] | 0.3437 | 0.05876 | 4.64E-04 | 0.2536 | 0.2773 | 0.3351 | 0.4217 | 0.4811 |
| HSTATUS_Mau[14] | 0.2951 | 0.04812 | 3.77E-04 | 0.2216 | 0.2408 | 0.288 | 0.3582 | 0.4079 |
| HSTATUS_Mau[15] | 0.3791 | 0.05407 | 4.29E-04 | 0.2926 | 0.3173 | 0.3724 | 0.4497 | 0.5044 |
| HSTATUS_Mau[16] | 0.6014 | 0.08305 | 6.46E-04 | 0.4609 | 0.5051 | 0.5928 | 0.7084 | 0.7893 |
| HSTATUS_Mau[17] | 0.4821 | 0.06638 | 5.01E-04 | 0.3678 | 0.4048 | 0.476 | 0.5668 | 0.6314 |
| HSTATUS_Mau[18] | 0.4078 | 0.05875 | 4.47E-04 | 0.3012 | 0.3379 | 0.4037 | 0.4822 | 0.5368 |
| HSTATUS_Mau[19] | 0.5109 | 0.07145 | 5.35E-04 | 0.3861 | 0.4266 | 0.5049 | 0.6022 | 0.669 |
| HSTATUS_Mau[20] | 0.1188 | 0.01604 | $1.12 \mathrm{E}-04$ | 0.09231 | 0.1002 | 0.1171 | 0.1395 | 0.1551 |
| H_Hoomalu[1] | 0.06259 | 0.007446 | 1.10E-04 | 0.04986 | 0.05364 | 0.06195 | 0.07239 | 0.07912 |
| H_Hoomalu[2] | 0.03304 | 0.003752 | 5.81E-05 | 0.02647 | 0.02847 | 0.03278 | 0.03796 | 0.04112 |
| H_Hoomalu[3] | 0.04284 | 0.004694 | 7.49E-05 | 0.03449 | 0.03713 | 0.04253 | 0.04893 | 0.05298 |
| H_Hoomalu[4] | 0.07496 | 0.008264 | 1.30E-04 | 0.06046 | 0.06491 | 0.0744 | 0.08566 | 0.09288 |
| H_Hoomalu[5] | 0.07409 | 0.008247 | 1.30E-04 | 0.05976 | 0.06411 | 0.07351 | 0.08479 | 0.09193 |
| H_Hoomalu[6] | 0.08443 | 0.009249 | $1.41 \mathrm{E}-04$ | 0.06798 | 0.07311 | 0.08384 | 0.0964 | 0.1041 |
| H_Hoomalu[7] | 0.1123 | 0.01242 | 1.94E-04 | 0.09051 | 0.09729 | 0.1115 | 0.1284 | 0.1391 |
| H_Hoomalu[8] | 0.1056 | 0.01191 | 1.83E-04 | 0.08478 | 0.09113 | 0.1048 | 0.1211 | 0.1313 |
| H_Hoomalu[9] | 0.07515 | 0.008563 | 1.30E-04 | 0.06014 | 0.06475 | 0.07457 | 0.08626 | 0.09363 |
| H_Hoomalu[10] | 0.0855 | 0.009692 | $1.45 \mathrm{E}-04$ | 0.06853 | 0.07372 | 0.08485 | 0.09808 | 0.1064 |
| H_Hoomalu[11] | 0.1122 | 0.01317 | 1.92E-04 | 0.08902 | 0.0962 | 0.1113 | 0.1293 | 0.1406 |
| H_Hoomalu[12] | 0.141 | 0.01619 | 2.39E-04 | 0.1127 | 0.1213 | 0.1399 | 0.1621 | 0.176 |
| H_Hoomalu[13] | 0.1075 | 0.01204 | $1.78 \mathrm{E}-04$ | 0.08628 | 0.09287 | 0.1066 | 0.1232 | 0.1336 |
| H_Hoomalu[14] | 0.1049 | 0.01193 | 1.75E-04 | 0.08401 | 0.09044 | 0.1041 | 0.1204 | 0.1306 |
| H_Hoomalu[15] | 0.08441 | 0.01095 | 1.49E-04 | 0.06438 | 0.07086 | 0.0839 | 0.09855 | 0.1074 |
| H_Hoomalu[16] | 0.07356 | 0.008794 | 1.22E-04 | 0.05773 | 0.06282 | 0.07301 | 0.08497 | 0.09224 |
| H_Hoomalu[17] | 0.07604 | 0.009677 | $1.31 \mathrm{E}-04$ | 0.05797 | 0.06407 | 0.0756 | 0.08842 | 0.09628 |

Table A4.--continued.

| Variable | Mean | Standard Deviation | MCMC Error | $\begin{gathered} \hline \text { Percentile } \\ 2.5 \% \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 10 \% \\ \hline \end{gathered}$ | Median | $\begin{gathered} \hline \text { Percentile } \\ 90 \% \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 97.5 \% \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| H_Hoomalu[18] | 0.08826 | 0.01103 | 1.50E-04 | 0.0677 | 0.07462 | 0.08778 | 0.1024 | 0.1113 |
| H_Hoomalu[19] | 0.06291 | 0.007974 | 1.09E-04 | 0.04826 | 0.05305 | 0.06254 | 0.07313 | 0.07972 |
| H_Hoomalu[20] | 0.07269 | 0.008234 | 1.19E-04 | 0.0582 | 0.06266 | 0.07211 | 0.08339 | 0.09032 |
| H_MHI[1] | 0.2718 | 0.02666 | 5.73E-04 | 0.2228 | 0.2385 | 0.2706 | 0.3065 | 0.3278 |
| H_MHI[2] | 0.2211 | 0.02212 | 4.71E-04 | 0.1805 | 0.1935 | 0.2201 | 0.2499 | 0.2672 |
| H_MHI[3] | 0.1908 | 0.01911 | 4.07E-04 | 0.1557 | 0.167 | 0.19 | 0.2156 | 0.2306 |
| H_MHI[4] | 0.1506 | 0.01507 | 3.17E-04 | 0.123 | 0.1318 | 0.15 | 0.1701 | 0.1822 |
| H_MHI[5] | 0.1345 | 0.01344 | 2.83E-04 | 0.1098 | 0.1177 | 0.134 | 0.152 | 0.1624 |
| H_MHI[6] | 0.1281 | 0.01283 | $2.74 \mathrm{E}-04$ | 0.1045 | 0.1121 | 0.1276 | 0.1448 | 0.1548 |
| H_MHI[7] | 0.08573 | 0.00865 | 1.80E-04 | 0.06996 | 0.07501 | 0.0853 | 0.09695 | 0.1038 |
| H_MHI[8] | 0.1171 | 0.01175 | 2.52E-04 | 0.09556 | 0.1024 | 0.1166 | 0.1324 | 0.1417 |
| H_MHI[9] | 0.1237 | 0.01242 | 2.56E-04 | 0.101 | 0.1083 | 0.123 | 0.1398 | 0.1498 |
| H_MHI[10] | 0.1213 | 0.01221 | 2.60E-04 | 0.09882 | 0.1061 | 0.1208 | 0.1372 | 0.1468 |
| H_MHI[11] | 0.1346 | 0.0137 | 2.91E-04 | 0.1092 | 0.1174 | 0.134 | 0.1523 | 0.1631 |
| H_MHI[12] | 0.08422 | 0.008409 | 1.79E-04 | 0.06872 | 0.07376 | 0.08388 | 0.09514 | 0.1018 |
| H_MHI[13] | 0.05187 | 0.005186 | 1.09E-04 | 0.04237 | 0.04543 | 0.05162 | 0.05861 | 0.06276 |
| H_MHI[14] | 0.06212 | 0.006217 | $1.29 \mathrm{E}-04$ | 0.05073 | 0.05443 | 0.06182 | 0.07016 | 0.07515 |
| H_MHI[15] | 0.1024 | 0.01034 | 2.22E-04 | 0.08349 | 0.08951 | 0.102 | 0.1158 | 0.124 |
| H_MHI[16] | 0.1043 | 0.01046 | 2.23E-04 | 0.08509 | 0.09131 | 0.1039 | 0.118 | 0.1261 |
| H_MHI[17] | 0.09107 | 0.009112 | $1.91 \mathrm{E}-04$ | 0.07438 | 0.07974 | 0.09069 | 0.1028 | 0.1101 |
| H_MHI[18] | 0.07352 | 0.007395 | 1.55E-04 | 0.06001 | 0.06432 | 0.07319 | 0.08316 | 0.08904 |
| H_MHI[19] | 0.114 | 0.01144 | 2.42E-04 | 0.09302 | 0.09983 | 0.1136 | 0.1288 | 0.138 |
| H_MHI[20] | 0.1002 | 0.01005 | $2.11 \mathrm{E}-04$ | 0.08187 | 0.08774 | 0.09979 | 0.1133 | 0.1212 |
| H_MHI[21] | 0.1139 | 0.01146 | 2.43E-04 | 0.09285 | 0.09965 | 0.1134 | 0.1289 | 0.1378 |
| H_MHI[22] | 0.1092 | 0.01095 | 2.34E-04 | 0.089 | 0.09559 | 0.1087 | 0.1234 | 0.1321 |
| H_MHI[23] | 0.09703 | 0.009768 | 2.08E-04 | 0.07914 | 0.08486 | 0.09661 | 0.1097 | 0.1173 |
| H_MHI[24] | 0.1335 | 0.01332 | $2.78 \mathrm{E}-04$ | 0.1091 | 0.117 | 0.1329 | 0.1509 | 0.1614 |
| H_MHI[25] | 0.109 | 0.01095 | 2.31E-04 | 0.08887 | 0.09538 | 0.1086 | 0.1232 | 0.1319 |
| H_MHI[26] | 0.1507 | 0.01506 | $3.18 \mathrm{E}-04$ | 0.1231 | 0.1319 | 0.1501 | 0.1703 | 0.1823 |
| H_MHI[27] | 0.1609 | 0.01622 | 3.46E-04 | 0.1311 | 0.1407 | 0.1603 | 0.182 | 0.1947 |
| H_MHI[28] | 0.1689 | 0.01689 | $3.49 \mathrm{E}-04$ | 0.1378 | 0.1479 | 0.1682 | 0.1908 | 0.2041 |
| H_MHI[29] | 0.2102 | 0.02123 | 4.51E-04 | 0.1711 | 0.1838 | 0.2093 | 0.2378 | 0.2544 |
| H_MHI[30] | 0.2423 | 0.0242 | 5.12E-04 | 0.1979 | 0.212 | 0.2413 | 0.2737 | 0.2928 |
| H_MHI[31] | 0.2175 | 0.0219 | 4.52E-04 | 0.1775 | 0.1903 | 0.2165 | 0.2459 | 0.2633 |
| H_MHI[32] | 0.2464 | 0.02488 | 5.29E-04 | 0.2006 | 0.2154 | 0.2455 | 0.2787 | 0.2983 |
| H_MHI[33] | 0.3307 | 0.03296 | 7.01E-04 | 0.2701 | 0.2896 | 0.3292 | 0.3734 | 0.3992 |
| H_MHI[34] | 0.3732 | 0.03719 | 7.98E-04 | 0.3044 | 0.327 | 0.3716 | 0.4214 | 0.4506 |
| H_MHI[35] | 0.4687 | 0.0461 | 9.84E-04 | 0.3836 | 0.4111 | 0.4669 | 0.5285 | 0.5649 |
| H_MHI[36] | 0.4956 | 0.04889 | 0.001049 | 0.4051 | 0.4345 | 0.4937 | 0.5592 | 0.5975 |
| H_MHI[37] | 0.539 | 0.05269 | 0.001129 | 0.441 | 0.4735 | 0.537 | 0.6074 | 0.648 |
| H_MHI[38] | 0.5127 | 0.05037 | 0.001077 | 0.4194 | 0.4498 | 0.5106 | 0.5783 | 0.6172 |
| H_MHI[39] | 0.5176 | 0.05027 | 0.001076 | 0.4248 | 0.4547 | 0.5157 | 0.5831 | 0.6218 |
| H_MHI[40] | 0.5004 | 0.0486 | 0.001031 | 0.4105 | 0.4397 | 0.4987 | 0.5633 | 0.6015 |
| H_MHI[41] | 0.6166 | 0.05833 | 0.001239 | 0.5082 | 0.5435 | 0.6146 | 0.6921 | 0.7369 |
| H_MHI[42] | 0.5122 | 0.05078 | 0.001088 | 0.4182 | 0.4486 | 0.5102 | 0.5781 | 0.618 |
| H_MHI[43] | 0.4076 | 0.04108 | 8.62E-04 | 0.3325 | 0.3566 | 0.4057 | 0.461 | 0.4936 |
| H_MHI[44] | 0.4838 | 0.04813 | 0.001035 | 0.3951 | 0.4238 | 0.4817 | 0.5466 | 0.5843 |
| H_MHI[45] | 0.3621 | 0.03643 | 7.77E-04 | 0.295 | 0.3167 | 0.3606 | 0.4094 | 0.4378 |
| H_MHI[46] | 0.404 | 0.04014 | 8.47E-04 | 0.33 | 0.354 | 0.4023 | 0.4562 | 0.4875 |
| H_MHI[47] | 0.4436 | 0.04378 | 9.28E-04 | 0.363 | 0.3889 | 0.442 | 0.5002 | 0.5351 |
| H_MHI[48] | 0.4069 | 0.04076 | 8.71E-04 | 0.3318 | 0.3561 | 0.4051 | 0.4597 | 0.4919 |
| H_MHI[49] | 0.4175 | 0.04145 | 8.80E-04 | 0.3414 | 0.366 | 0.4157 | 0.4715 | 0.5042 |
| H_MHI[50] | 0.4254 | 0.04245 | 9.11E-04 | 0.3474 | 0.3725 | 0.4237 | 0.4802 | 0.5138 |
| H_MHI[51] | 0.3322 | 0.03333 | 7.09E-04 | 0.2706 | 0.2908 | 0.3309 | 0.3755 | 0.4015 |
| H_MHI[52] | 0.4023 | 0.03986 | 8.49E-04 | 0.329 | 0.353 | 0.4006 | 0.4542 | 0.4854 |
| H_MHI[53] | 0.3485 | 0.03494 | 7.38E-04 | 0.2844 | 0.305 | 0.347 | 0.3939 | 0.4218 |
| H_MHI[54] | 0.3073 | 0.03086 | 6.54E-04 | 0.2504 | 0.2689 | 0.306 | 0.3472 | 0.3715 |
| H_MHI[55] | 0.326 | 0.03259 | 6.94E-04 | 0.2662 | 0.2853 | 0.3245 | 0.3685 | 0.3936 |
| H_MHI[56] | 0.3099 | 0.0311 | 6.65E-04 | 0.2524 | 0.2712 | 0.3085 | 0.3502 | 0.3751 |
| H_MHI[57] | 0.3338 | 0.03325 | 7.02E-04 | 0.2728 | 0.2925 | 0.3324 | 0.3771 | 0.4028 |

Table A4.--continued.

| Variable | Mean | Standard Deviation | MCMC Error | $\begin{gathered} \hline \text { Percentile } \\ 2.5 \% \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 10 \% \\ \hline \end{gathered}$ | Median | $\begin{gathered} \text { Percentile } \\ 90 \% \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 97.5 \% \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| H_MHI[58] | 0.2714 | 0.02729 | 5.77E-04 | 0.2213 | 0.2374 | 0.2703 | 0.3067 | 0.3281 |
| H_MHI[59] | 0.3185 | 0.03213 | 6.71E-04 | 0.2595 | 0.2783 | 0.3172 | 0.3604 | 0.3854 |
| H_Mau[1] | 0.04764 | 0.004928 | 9.26E-05 | 0.03934 | 0.04194 | 0.0472 | 0.05374 | 0.05871 |
| H_Mau[2] | 0.09824 | 0.01194 | $1.77 \mathrm{E}-04$ | 0.07018 | 0.08344 | 0.09894 | 0.1123 | 0.1201 |
| H_Mau[3] | 0.1209 | 0.01409 | 2.21E-04 | 0.0909 | 0.1035 | 0.1211 | 0.1382 | 0.1484 |
| H_Mau[4] | 0.2924 | 0.03454 | 5.66E-04 | 0.2359 | 0.2526 | 0.2886 | 0.3366 | 0.3722 |
| H_Mau[5] | 0.1505 | 0.02256 | 3.52E-04 | 0.1163 | 0.1257 | 0.147 | 0.1796 | 0.2057 |
| H_Mau[6] | 0.1091 | 0.01557 | 2.32E-04 | 0.08501 | 0.09162 | 0.1068 | 0.1293 | 0.1464 |
| H_Mau[7] | 0.1872 | 0.02294 | 3.50E-04 | 0.149 | 0.1604 | 0.1847 | 0.2169 | 0.239 |
| H_Mau[8] | 0.2707 | 0.03311 | 5.19E-04 | 0.2154 | 0.2319 | 0.2673 | 0.3137 | 0.3454 |
| H_Mau[9] | 0.2454 | 0.03031 | 4.99E-04 | 0.1931 | 0.2091 | 0.2429 | 0.2849 | 0.3116 |
| H_Mau[10] | 0.2629 | 0.03619 | 5.46E-04 | 0.1944 | 0.2187 | 0.2614 | 0.3092 | 0.3385 |
| H_Mau[11] | 0.1218 | 0.01748 | 2.64E-04 | 0.09107 | 0.1009 | 0.1205 | 0.1445 | 0.1599 |
| H_Mau[12] | 0.08423 | 0.01131 | $1.67 \mathrm{E}-04$ | 0.06512 | 0.07092 | 0.08313 | 0.09896 | 0.1094 |
| H_Mau[13] | 0.09832 | 0.01457 | $1.97 \mathrm{E}-04$ | 0.07615 | 0.08223 | 0.09614 | 0.1171 | 0.1336 |
| H_Mau[14] | 0.08443 | 0.01186 | $1.63 \mathrm{E}-04$ | 0.06636 | 0.07131 | 0.08267 | 0.09964 | 0.1132 |
| H_Mau[15] | 0.1085 | 0.01229 | $1.90 \mathrm{E}-04$ | 0.08689 | 0.09378 | 0.1076 | 0.1242 | 0.1354 |
| H_Mau[16] | 0.1721 | 0.0193 | 3.10E-04 | 0.136 | 0.1485 | 0.1714 | 0.1966 | 0.2125 |
| H_Mau[17] | 0.1381 | 0.01607 | 2.72E-04 | 0.1077 | 0.1181 | 0.1375 | 0.1586 | 0.1714 |
| H_Mau[18] | 0.1168 | 0.0143 | 2.31E-04 | 0.08854 | 0.09886 | 0.1167 | 0.1348 | 0.1456 |
| H_Mau[19] | 0.1463 | 0.01717 | 2.82E-04 | 0.1133 | 0.1249 | 0.1459 | 0.1681 | 0.1815 |
| H_Mau[20] | 0.03404 | 0.003881 | 6.68E-05 | 0.02697 | 0.02929 | 0.03383 | 0.03902 | 0.04229 |
| K_Archipelago | 7104 | 484.8 | 13.1 | 6186 | 6490 | 7090 | 7734 | 8081 |
| K_Hoomalu | 3049 | 208.1 | 5.621 | 2655 | 2785 | 3043 | 3319 | 3468 |
| K_MHI | 3176 | 216.7 | 5.855 | 2765 | 2901 | 3170 | 3457 | 3613 |
| K_Mau | 879.7 | 60.03 | 1.622 | 766 | 803.6 | 878 | 957.7 | 1001 |
| MSP_Hoomalu | 437.8 | 32.03 | 0.2463 | 373.4 | 396.6 | 438.3 | 478.1 | 499.4 |
| MSP_MHI | 456 | 33.37 | 0.2565 | 388.9 | 413.1 | 456.6 | 498.1 | 520.2 |
| MSP_Mau | 126.3 | 9.243 | 0.07106 | 107.7 | 114.4 | 126.5 | 138 | 144.1 |
| RESID_Hoomalu[1] | 0.1255 | 0.1031 | $5.85 \mathrm{E}-04$ | -0.03012 | 0.01193 | 0.1076 | 0.266 | 0.3823 |
| RESID_Hoomalu[2] | 0.06573 | 0.08431 | $4.39 \mathrm{E}-04$ | -0.06968 | -0.02821 | 0.0541 | 0.1755 | 0.2721 |
| RESID_Hoomalu[3] | -0.01292 | 0.06334 | 2.52E-04 | -0.1374 | -0.09168 | -0.01345 | 0.06599 | 0.1159 |
| RESID_Hoomalu[4] | -0.002422 | 0.06195 | 2.32E-04 | -0.1251 | -0.07998 | -0.002405 | 0.0751 | 0.1214 |
| RESID_Hoomalu[5] | -0.01688 | 0.061 | $2.11 \mathrm{E}-04$ | -0.1387 | -0.09346 | -0.01661 | 0.05909 | 0.1041 |
| RESID_Hoomalu[6] | 0.04716 | 0.06783 | 2.76E-04 | -0.07667 | -0.03417 | 0.04297 | 0.135 | 0.1937 |
| RESID_Hoomalu[7] | -0.002189 | 0.06132 | $2.06 \mathrm{E}-04$ | -0.1226 | -0.07829 | -0.002829 | 0.07481 | 0.1215 |
| RESID_Hoomalu[8] | -0.00653 | 0.06066 | $1.96 \mathrm{E}-04$ | -0.1281 | -0.08226 | -0.006553 | 0.06899 | 0.1137 |
| RESID_Hoomalu[9] | -0.00743 | 0.06073 | $1.91 \mathrm{E}-04$ | -0.1295 | -0.08394 | -0.006713 | 0.06803 | 0.1115 |
| RESID_Hoomalu[10] | 0.002742 | 0.06039 | $1.98 \mathrm{E}-04$ | -0.1191 | -0.07329 | 0.003228 | 0.07793 | 0.1215 |
| RESID_Hoomalu[11] | -0.03534 | 0.06281 | 2.15E-04 | -0.1665 | -0.116 | -0.03269 | 0.04132 | 0.08239 |
| RESID_Hoomalu[12] | -0.02688 | 0.06053 | 1.89E-04 | -0.1503 | -0.1043 | -0.0252 | 0.04801 | 0.08906 |
| RESID_Hoomalu[13] | 0.04998 | 0.06521 | 2.37E-04 | -0.07166 | -0.02969 | 0.04698 | 0.1342 | 0.1868 |
| RESID_Hoomalu[14] | 0.0253 | 0.06152 | 2.01E-04 | -0.09545 | -0.05063 | 0.0243 | 0.103 | 0.1488 |
| RESID_Hoomalu[15] | -0.08691 | 0.07524 | 3.44E-04 | -0.2574 | -0.1866 | -0.07912 | 0.001159 | 0.04051 |
| RESID_Hoomalu[16] | 0.01316 | 0.06527 | 2.76E-04 | -0.1268 | -0.06844 | 0.01547 | 0.09198 | 0.1352 |
| RESID_Hoomalu[17] | -0.05186 | 0.07344 | 3.58E-04 | -0.225 | -0.1457 | -0.04438 | 0.03261 | 0.07239 |
| RESID_Hoomalu[18] | -0.0294 | 0.0714 | 3.30E-04 | -0.196 | -0.1203 | -0.02266 | 0.05296 | 0.09333 |
| RESID_Hoomalu[19] | -0.06251 | 0.07294 | 3.20E-04 | -0.2296 | -0.1579 | -0.05505 | 0.02231 | 0.06209 |
| RESID_Hoomalu[20] | 0.01184 | 0.06438 | 2.21E-04 | -0.1178 | -0.06829 | 0.01173 | 0.0923 | 0.1391 |
| RESID_MHI[1] | -0.008822 | 0.0478 | $1.47 \mathrm{E}-04$ | -0.1054 | -0.06922 | -0.00832 | 0.05097 | 0.08401 |
| RESID_MHI[2] | -0.003239 | 0.04785 | $1.49 \mathrm{E}-04$ | -0.09875 | -0.06405 | -0.00297 | 0.0568 | 0.09084 |
| RESID_MHI[3] | -0.003486 | 0.04805 | 1.65E-04 | -0.09958 | -0.06416 | -0.003331 | 0.05688 | 0.09077 |
| RESID_MHI[4] | 0.01141 | 0.0487 | $1.60 \mathrm{E}-04$ | -0.08322 | -0.04927 | 0.01078 | 0.07295 | 0.1091 |
| RESID_MHI[5] | 0.007518 | 0.04875 | $1.52 \mathrm{E}-04$ | -0.08829 | -0.05343 | 0.007106 | 0.06881 | 0.1051 |
| RESID_MHI[6] | -0.009536 | 0.04821 | $1.55 \mathrm{E}-04$ | -0.1063 | -0.07027 | -0.009157 | 0.05067 | 0.0852 |
| RESID_MHI[7] | 0.03768 | 0.05171 | $1.69 \mathrm{E}-04$ | -0.05855 | -0.02543 | 0.03537 | 0.1039 | 0.1466 |
| RESID_MHI[8] | -0.001677 | 0.04873 | 1.68E-04 | -0.09812 | -0.06296 | -0.001749 | 0.05985 | 0.09477 |
| RESID_MHI[9] | 0.03119 | 0.05141 | 1.72E-04 | -0.06515 | -0.03193 | 0.02919 | 0.0968 | 0.1392 |
| RESID_MHI[10] | -8.59E-04 | 0.049 | 1.59E-04 | -0.09871 | -0.06194 | -8.46E-04 | 0.06052 | 0.09664 |
| RESID_MHI[11] | -0.03288 | 0.05002 | $1.70 \mathrm{E}-04$ | -0.1372 | -0.0966 | -0.03107 | 0.02833 | 0.06042 |

Table A4.--continued.

| Variable | Mean | Standard Deviation | MCMC Error | $\begin{gathered} \hline \text { Percentile } \\ 2.5 \% \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 10 \% \end{gathered}$ | Median | $\begin{gathered} \text { Percentile } \\ 90 \% \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 97.5 \% \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| RESID_MHI[12] | -0.00175 | 0.04834 | 1.58E-04 | -0.09828 | -0.06246 | -0.001555 | 0.05881 | 0.09378 |
| RESID_MHI[13] | 0.03021 | 0.05042 | $1.68 \mathrm{E}-04$ | -0.0647 | -0.03205 | 0.02841 | 0.0943 | 0.1354 |
| RESID_MHI[14] | 0.02631 | 0.0514 | 1.73E-04 | -0.07001 | -0.03666 | 0.02433 | 0.09197 | 0.134 |
| RESID_MHI[15] | -0.0118 | 0.04941 | $1.74 \mathrm{E}-04$ | -0.1125 | -0.07417 | -0.01105 | 0.04966 | 0.08445 |
| RESID_MHI[16] | -0.006451 | 0.04847 | $1.61 \mathrm{E}-04$ | -0.1043 | -0.06726 | -0.006113 | 0.05434 | 0.08783 |
| RESID_MHI[17] | 0.02265 | 0.04984 | 1.64E-04 | -0.07192 | -0.03905 | 0.02129 | 0.08602 | 0.1249 |
| RESID_MHI[18] | 0.0184 | 0.05026 | $1.54 \mathrm{E}-04$ | -0.07784 | -0.04403 | 0.01716 | 0.08198 | 0.1216 |
| RESID_MHI[19] | -0.005027 | 0.0489 | 1.55E-04 | -0.1028 | -0.06626 | -0.004719 | 0.0561 | 0.09128 |
| RESID_MHI[20] | 0.01117 | 0.04939 | $1.54 \mathrm{E}-04$ | -0.08519 | -0.05048 | 0.0107 | 0.07349 | 0.111 |
| RESID_MHI[21] | -0.006907 | 0.04861 | 1.55E-04 | -0.105 | -0.06815 | -0.006464 | 0.05371 | 0.08823 |
| RESID_MHI[22] | -0.006557 | 0.04864 | 1.52E-04 | -0.1038 | -0.06799 | -0.006058 | 0.05403 | 0.08854 |
| RESID_MHI[23] | -0.01015 | 0.04868 | $1.48 \mathrm{E}-04$ | -0.1078 | -0.07152 | -0.009696 | 0.05075 | 0.08469 |
| RESID_MHI[24] | 0.006027 | 0.04868 | 1.61E-04 | -0.08985 | -0.05482 | 0.005796 | 0.06722 | 0.103 |
| RESID_MHI[25] | -0.004995 | 0.04863 | 1.52E-04 | -0.103 | -0.06598 | -0.00455 | 0.05567 | 0.09047 |
| RESID_MHI[26] | 0.001148 | 0.04866 | $1.61 \mathrm{E}-04$ | -0.09594 | -0.05992 | 0.001382 | 0.06221 | 0.09682 |
| RESID_MHI[27] | -0.01438 | 0.04861 | 1.56E-04 | -0.1127 | -0.07606 | -0.01364 | 0.04617 | 0.07957 |
| RESID_MHI[28] | 0.02414 | 0.04934 | 1.60E-04 | -0.07071 | -0.03715 | 0.02318 | 0.08674 | 0.1254 |
| RESID_MHI[29] | -0.02716 | 0.04887 | $1.49 \mathrm{E}-04$ | -0.1269 | -0.08944 | -0.02566 | 0.03315 | 0.06583 |
| RESID_MHI[30] | -0.002888 | 0.04796 | 1.52E-04 | -0.09887 | -0.06331 | -0.002682 | 0.05723 | 0.09108 |
| RESID_MHI[31] | 0.03279 | 0.05004 | $1.69 \mathrm{E}-04$ | -0.06171 | -0.02858 | 0.03105 | 0.09666 | 0.1366 |
| RESID_MHI[32] | -0.02367 | 0.04825 | $1.57 \mathrm{E}-04$ | -0.1226 | -0.08486 | -0.02263 | 0.03613 | 0.06829 |
| RESID_MHI[33] | -0.00124 | 0.04732 | $1.49 \mathrm{E}-04$ | -0.0959 | -0.06097 | -8.44E-04 | 0.05812 | 0.0917 |
| RESID_MHI[34] | -0.0127 | 0.04694 | 1.43E-04 | -0.1081 | -0.07214 | -0.0118 | 0.0458 | 0.0778 |
| RESID_MHI[35] | -0.003829 | 0.04633 | 1.51E-04 | -0.09724 | -0.06275 | -0.003266 | 0.05412 | 0.08657 |
| RESID_MHI[36] | -0.02213 | 0.04677 | $1.45 \mathrm{E}-04$ | -0.1185 | -0.08154 | -0.02083 | 0.03603 | 0.0665 |
| RESID_MHI[37] | -0.02254 | 0.04641 | 1.43E-04 | -0.1183 | -0.08188 | -0.02116 | 0.03496 | 0.06522 |
| RESID_MHI[38] | -0.02275 | 0.04647 | 1.41E-04 | -0.1192 | -0.08236 | -0.02115 | 0.03469 | 0.06459 |
| RESID_MHI[39] | -0.01871 | 0.04566 | 1.51E-04 | -0.1122 | -0.07683 | -0.01768 | 0.03804 | 0.06822 |
| RESID_MHI[40] | 0.008278 | 0.04579 | 1.56E-04 | -0.08216 | -0.04941 | 0.008317 | 0.06614 | 0.09882 |
| RESID_MHI[41] | -0.006765 | 0.04461 | $1.79 \mathrm{E}-04$ | -0.09607 | -0.06332 | -0.00647 | 0.04938 | 0.08065 |
| RESID_MHI[42] | -0.001821 | 0.04572 | $1.44 \mathrm{E}-04$ | -0.09258 | -0.05964 | -0.001673 | 0.05581 | 0.08826 |
| RESID_MHI[43] | 0.03406 | 0.04818 | $1.66 \mathrm{E}-04$ | -0.05725 | -0.0258 | 0.03263 | 0.0958 | 0.1336 |
| RESID_MHI[44] | -0.02045 | 0.0463 | $1.48 \mathrm{E}-04$ | -0.1148 | -0.07925 | -0.01966 | 0.03741 | 0.06853 |
| RESID_MHI[45] | 0.002771 | 0.04691 | 1.43E-04 | -0.09075 | -0.05629 | 0.002842 | 0.06173 | 0.09499 |
| RESID_MHI[46] | -0.004063 | 0.04663 | $1.49 \mathrm{E}-04$ | -0.09744 | -0.0629 | -0.003822 | 0.05449 | 0.08731 |
| RESID_MHI[47] | 0.01484 | 0.0466 | 1.50E-04 | -0.07617 | -0.04342 | 0.0143 | 0.07399 | 0.1086 |
| RESID_MHI[48] | -0.0153 | 0.04666 | 1.52E-04 | -0.1093 | -0.07446 | -0.01457 | 0.0432 | 0.07536 |
| RESID_MHI[49] | 0.01111 | 0.04664 | 1.54E-04 | -0.08097 | -0.04738 | 0.01109 | 0.06998 | 0.1042 |
| RESID_MHI[50] | -0.01471 | 0.0465 | 1.52E-04 | -0.1094 | -0.07392 | -0.01387 | 0.04338 | 0.07493 |
| RESID_MHI[51] | 0.003345 | 0.04682 | $1.49 \mathrm{E}-04$ | -0.08933 | -0.0554 | 0.003383 | 0.06213 | 0.09613 |
| RESID_MHI[52] | 0.007217 | 0.04667 | 1.56E-04 | -0.08436 | -0.05133 | 0.006924 | 0.06614 | 0.1007 |
| RESID_MHI[53] | 0.00863 | 0.04717 | 1.59E-04 | -0.08312 | -0.05034 | 0.008057 | 0.06851 | 0.1032 |
| RESID_MHI[54] | 0.006319 | 0.04728 | 1.62E-04 | -0.08665 | -0.05312 | 0.00586 | 0.06594 | 0.1016 |
| RESID_MHI[55] | -4.71E-04 | 0.04709 | 1.57E-04 | -0.09415 | -0.0595 | -3.37E-04 | 0.05866 | 0.09282 |
| RESID_MHI[56] | -0.01434 | 0.04737 | $1.56 \mathrm{E}-04$ | -0.1106 | -0.07416 | -0.01367 | 0.04465 | 0.07735 |
| RESID_MHI[57] | 0.01662 | 0.04726 | $1.48 \mathrm{E}-04$ | -0.07507 | -0.04245 | 0.01602 | 0.07636 | 0.112 |
| RESID_MHI[58] | -1.43E-04 | 0.04733 | $1.49 \mathrm{E}-04$ | -0.09456 | -0.05965 | -4.14E-05 | 0.05929 | 0.09363 |
| RESID_MHI[59] | -0.01241 | 0.05051 | 1.57E-04 | -0.1151 | -0.07598 | -0.01164 | 0.05013 | 0.08597 |
| RESID_Mau[1] | -0.2306 | 0.09181 | 5.11E-04 | -0.4097 | -0.3451 | -0.2321 | -0.114 | -0.04124 |
| RESID_Mau[2] | 0.3785 | 0.1174 | 8.06E-04 | 0.07386 | 0.2273 | 0.3943 | 0.5088 | 0.5665 |
| RESID_Mau[3] | 0.2583 | 0.1021 | 6.13E-04 | 0.01777 | 0.1253 | 0.2677 | 0.3778 | 0.4357 |
| RESID_Mau[4] | -0.1664 | 0.08995 | 4.53E-04 | -0.3342 | -0.2774 | -0.1702 | -0.04913 | 0.02288 |
| RESID_Mau[5] | -0.2702 | 0.1113 | 7.46E-04 | -0.4587 | -0.3989 | -0.2827 | -0.1195 | -0.006903 |
| RESID_Mau[6] | -0.2078 | 0.1051 | 6.12E-04 | -0.3911 | -0.3325 | -0.2177 | -0.06496 | 0.02874 |
| RESID_Mau[7] | -0.02493 | 0.08652 | 3.63E-04 | -0.1902 | -0.1323 | -0.0268 | 0.08531 | 0.1533 |
| RESID_Mau[8] | -0.1421 | 0.08541 | $3.77 \mathrm{E}-04$ | -0.3032 | -0.2481 | -0.145 | -0.03133 | 0.03532 |
| RESID_Mau[9] | -0.09026 | 0.08059 | 3.11E-04 | -0.2498 | -0.192 | -0.09027 | 0.01201 | 0.0681 |
| RESID_Mau[10] | 0.2668 | 0.09958 | 5.76E-04 | 0.03356 | 0.1387 | 0.2752 | 0.3844 | 0.4424 |
| RESID_Mau[11] | 0.1666 | 0.09763 | 5.33E-04 | -0.03676 | 0.0392 | 0.1703 | 0.2869 | 0.3516 |
| RESID_Mau[12] | 0.02332 | 0.08736 | 3.51E-04 | -0.1477 | -0.08621 | 0.02251 | 0.1341 | 0.1975 |

Table A4.--continued.

| Variable | Mean | Standard Deviation | MCMC Error | $\begin{gathered} \hline \text { Percentile } \\ 2.5 \% \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 10 \% \end{gathered}$ | Median | $\begin{gathered} \hline \text { Percentile } \\ 90 \% \end{gathered}$ | $\begin{gathered} \hline \text { Percentile } \\ 97.5 \% \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| RESID_Mau[13] | -0.2915 | 0.1082 | 6.71E-04 | -0.4755 | -0.4169 | -0.3029 | -0.1489 | -0.03168 |
| RESID_Mau[14] | -0.2722 | 0.1043 | 6.30E-04 | -0.4513 | -0.3933 | -0.283 | -0.1348 | -0.02298 |
| RESID_Mau[15] | 0.05723 | 0.08102 | 3.24E-04 | -0.1019 | -0.0452 | 0.05681 | 0.1596 | 0.2179 |
| RESID_Mau[16] | 0.115 | 0.0856 | 4.03E-04 | -0.06011 | 0.004729 | 0.1175 | 0.2214 | 0.2773 |
| RESID_Mau[17] | 0.09712 | 0.08932 | 4.36E-04 | -0.08763 | -0.01828 | 0.1002 | 0.2081 | 0.2645 |
| RESID_Mau[18] | 0.1933 | 0.09977 | 5.65E-04 | -0.02591 | 0.06048 | 0.2009 | 0.313 | 0.3725 |
| RESID_Mau[19] | 0.1368 | 0.09455 | 4.80E-04 | -0.05949 | 0.01341 | 0.1406 | 0.254 | 0.3121 |
| RESID_Mau[20] | -0.002929 | 0.08989 | 3.35E-04 | -0.1852 | -0.1174 | -0.001491 | 0.1099 | 0.1712 |
| RMSE_Hoomalu | 0.07737 | 0.02981 | 1.83E-04 | 0.03734 | 0.04635 | 0.07124 | 0.1159 | 0.1574 |
| RMSE_MHI | 0.04961 | 0.01063 | 4.50E-05 | 0.03228 | 0.03705 | 0.04839 | 0.0636 | 0.0738 |
| RMSE_Mau | 0.2151 | 0.04188 | 3.33E-04 | 0.08357 | 0.1649 | 0.225 | 0.2526 | 0.268 |
| pB_Archipelago | 0.9724 | 0.1638 | 9.53E-04 | 0 | 1 | 1 | 1 | 1 |
| pB_Hoomalu | 0.9998 | 0.01414 | 5.94E-05 | 1 | 1 | 1 | 1 | 1 |
| pB_MHI | 0 | 0 | 2.24E-13 | 0 | 0 | 0 | 0 | 0 |
| pB_Mau | 0.9999 | 0.01183 | 4.20E-05 | 1 | 1 | 1 | 1 | 1 |
| pH_Archipelago | $1.00 \mathrm{E}-05$ | 0.003162 | $1.00 \mathrm{E}-05$ | 0 | 0 | 0 | 0 | 0 |
| pH_Hoomalu | 0 | 0 | 2.24E-13 | 0 | 0 | 0 | 0 | 0 |
| pH_MHI | 0.8722 | 0.3339 | 0.001267 | 0 | 0 | 1 | 1 | 1 |
| pH_Mau | 0 | 0 | $2.24 \mathrm{E}-13$ | 0 | 0 | 0 | 0 | 0 |
| q_Hoomalu | 0.2451 | 0.02397 | 4.14E-04 | 0.2035 | 0.2162 | 0.2433 | 0.2762 | 0.2972 |
| q_MHI | 0.1452 | 0.01277 | 3.07E-04 | 0.1214 | 0.1293 | 0.1448 | 0.1618 | 0.1716 |
| q_Mau | 0.5875 | 0.06183 | 0.001125 | 0.4786 | 0.5119 | 0.5832 | 0.6683 | 0.7207 |
| r | 0.577 | 0.05771 | 0.001164 | 0.4661 | 0.5036 | 0.5765 | 0.6509 | 0.6919 |
| sigma2_Hoomalu | 0.01268 | 0.005544 | $2.60 \mathrm{E}-05$ | 0.003806 | 0.006414 | 0.01196 | 0.01976 | 0.02567 |
| sigma2_MHI | 0.02395 | 0.005135 | 4.31E-05 | 0.01573 | 0.01798 | 0.02333 | 0.03073 | 0.03571 |
| sigma2_Mau | 0.006231 | 0.008022 | 6.61E-05 | 0.001261 | 0.001698 | 0.00354 | 0.01327 | 0.03202 |
| tau2_Hoomalu | 0.007161 | 0.006123 | $3.65 \mathrm{E}-05$ | 0.001761 | 0.00249 | 0.005336 | 0.01371 | 0.02443 |
| tau2_MHI | 0.002816 | 0.001238 | 5.02E-06 | 0.0012 | 0.001541 | 0.002555 | 0.004404 | 0.005927 |
| tau2_Mau | 0.0446 | 0.0202 | $1.11 \mathrm{E}-04$ | 0.006804 | 0.02161 | 0.04276 | 0.06944 | 0.09059 |



Figure A1.--Residual plot of standardized CPUE residual as a function of fitted CPUE values to evaluate whether model misspecification was occurring.


Figure A2.--Residual plot of the square root of the absolute value of standardized CPUE residuals as a function of fitted CPUE values to evaluate whether variance changed with fitted value.


Figure A3.--Diagnostic plot of observed log-scale CPUE as a function of the fitted value along with the estimated regression slope for observed CPUE as a function of the predicted value.


Figure A4.--A Q-Q plot of CPUE residuals versus the standard normal distribution.


Figure A5.--Histogram of standardized CPUE residuals from the fitted GLM.

Goodness-of-fit of predicted MHI CPUE as a function of the prior mean for the initial proportion of carrying capacity in the MHI


Posterior mean estimate of initial MHI proportion of carrying capacity in 1949 as a function of the prior mean for the initial proportion of carrying capacity in the Main Hawaiian Islands


Figure A6.--Goodness-of-fit of alternative values for the mean of the prior distribution of the initial proportion of carrying capacity in the main Hawaiian Islands in 1949.


[^0]:    Number of observations 141123

