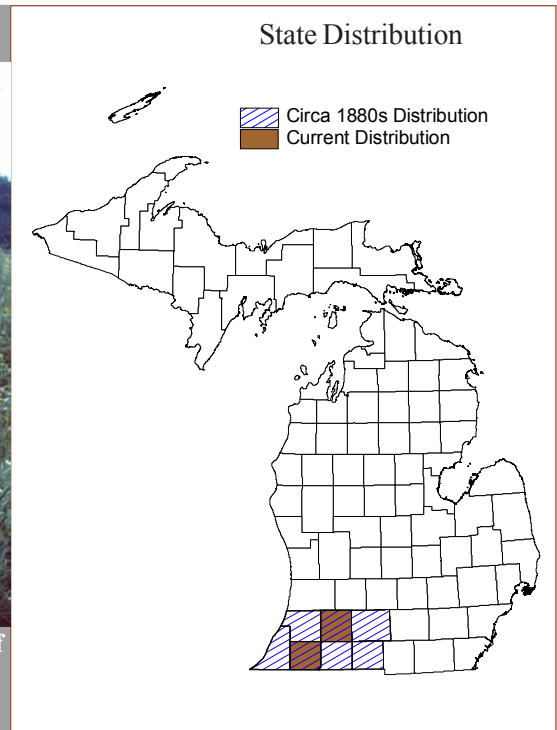




Photo by MNFI staff



Overview: Mesic prairie is a native grassland community dominated by big bluestem (*Andropogon gerardii*), little bluestem (*Andropogon scoparius*), and Indian grass (*Sorghastrum nutans*) that occurs on loam, sandy loam or silt loam soils on level or slightly undulating glacial outwash. Mesic prairie may also be referred to as black soil, tall grass prairie. Areas dominated by native grasses with less than one mature tree per acre (0.4 ha) are considered prairie (Curtis 1959).

Global and State Rank: G2/S1

Range: Mesic prairie occurs in IA, IL, IN, MI, MN, MO, NE, OH, and WI. (Faber-Langendoen 2001). In Michigan, the community is restricted to the southwestern portion of the Lower Peninsula, corresponding with the “Prairie Peninsula” described by Transeau (1935) (Chapman 1984). Historically, Kalamazoo County had the largest contiguous mesic prairie, Prairie Rhonde, which stretched over 12,600 acres (5,100 ha). Less than one mile west of Prairie Rhonde stood Gourdneck Prairie, another large mesic prairie (>2,100 acres, 850 ha). St. Joseph and Cass County also harbored several large patches of mesic prairies greater than 1,000 acres (400 ha) with several in St. Joseph County over 5,000 acres (2,000 ha). Smaller patches occurred in Branch, Calhoun, Berrien, and Van Buren counties. Lakeplain prairies, which occur on the glacial

lakeplains of the southern Lower Peninsula and Saginaw Bay, share many species in common with mesic prairie. However, the soil type and hydrologic processes supporting lakeplain prairie are significantly different from those of the mesic prairies in southwest Lower Michigan (Albert and Kost 1998, Kost et al. 2007). Today, mesic prairie is known to occur in only a few, very small, degraded sites in Kalamazoo and Cass counties.

Rank Justification: In the early to mid 1800s, the southern Lower Peninsula supported approximately 60,500 acres (24,500 ha) of upland prairie, which included pockets of mesic prairie, mesic sand prairie, dry-mesic prairie, dry sand prairie, and hillside prairie (Comer et al. 1995). The Michigan Natural Features Inventory database currently includes three element occurrences of mesic prairie, two are located in Kalamazoo County and the other in Cass County. The acreage of the remaining mesic prairies ranges from 0.6 to 3.7 acres (0.2 to 1.5 ha) and totals 6.2 acres (2.5 ha) in all. By contrast, in the early to mid 1800s grasslands in these two counties, much of which were mesic prairie (Holder et al. 1981, Brewer et al. 1984), totaled nearly 34,000 acres (13,750 ha) (Comer et al. 1995). In the counties believed to support mesic prairie in the 1800s, grasslands once totaled nearly 60,000 acres (24,000 ha) (Comer et al. 1995). It is difficult to reliably determine the total acreage of mesic prairie in Michigan in



the 1800s. However, based on comparisons of the total acreage of all upland prairie element occurrences in southern Lower Michigan today (480 acres, 194 ha) with that found in the early to mid 1800s (provided above), it appears that less than 1% of the original upland prairie remains intact.

The earliest European settlers erroneously assumed that the lack of trees within mesic prairies was a result of poor soils (Curtis 1959). However, following the introduction of the steel moldboard plow, which allowed for cultivation of the thick prairie sod, mesic prairies quickly became favored locations for homesteads because of their productive soils (Curtis 1959). Mesic prairies that were not plowed or grazed soon converted to forest in the absence of the annual fires that once maintained their open character (Curtis 1959).

Mesic prairie has been essentially eliminated from the Michigan landscape. The few, tiny remnants that remain are severely disturbed and their prospects for long-term viability are bleak. The status of mesic prairie in other Midwestern states is similarly dismal.

Landscape and Abiotic Context: In Michigan, mesic prairie occurs exclusively on glacial outwash on nearly level to slightly undulating sites (Chapman 1984, Albert 1995). The vast majority of historical mesic prairies occur on the Battle Creek Outwash Plain Sub-subsection within the Kalamazoo Interlobate Subsection (Albert 1995). This level outwash plain is the northernmost portion of the “Prairie Peninsula” described by Transeau (1935).

Soils supporting mesic prairie are loam or sandy loam and occasionally silt loam with pH ranging from 4.9 to 7.5 (ave. pH 6.2) and water retaining capacity ranging from 37 to 106% (ave. 64%) (Chapman 1984). The soil profile typically contains a B horizon dominated by clay (Chapman 1984). Mesic prairies in Michigan occur on both mollisols and udic alfisols. Mollisols are considered true prairie soils and udic alfisols (e.g., udalfs or Gray-Brown Podzols), which cover much of southern Lower Michigan, are often considered gray to brown forest soils (Chapman 1984).

In the 1800s, mesic prairie in Michigan frequently bordered beech-maple forest or graded into bur oak plains. These communities all occupied level outwash plains and grew on similarly rich soils. Where level outwash plains met end moraines and ground moraines with sandy, drier

soils, mesic prairie gave way to dry-mesic prairie and oak openings, or oak woodlands. Today the community is restricted to railroad right-of-ways, cemeteries, and other small remnants, which often border agricultural fields.

Natural Processes: Fire played a critical role in maintaining open conditions in Michigan prairie and oak savanna ecosystems. The frequency and intensity of fire depended on a variety of factors including the type and volume of fuel, topography, presence of natural firebreaks, and density of Native Americans (Chapman 1984). The rich soils of mesic prairie promoted very high volumes of fine fuels (e.g., grasses), which enabled fire to rapidly spread throughout the community. In general, the probability of a wide-ranging fire increases in level topography like large outwash plains (Chapman 1984). Thus, on the broad, level outwash plains of southwestern Lower Michigan, annual, wind-swept fires once spread easily through the mesic prairies and bur oak plains. Carried by wind, these fires moved across the outwash plains and up slopes of end moraines and ground moraines, carving oak forests into dry-mesic prairies and oak openings.

While occasional lightning strikes resulted in fires that spread across the landscape, Native Americans were the main sources of ignition. There are many early accounts of Native Americans intentionally setting fires to accomplish specific objectives (see Day 1953, Curtis 1959, Thompson and Smith 1970, Chapman 1984, Denevan 1992, Kay 1995). Native Americans intentionally set fires in the fall to clear briars and brush and make the land more easily passable. Frequent fires kept the land open, increasing both short- and long-range visibility, which facilitated large game hunting and provided a measure of safety from surprise attacks by neighboring tribes. Fire was used to increase productivity of berry crops and agricultural fields. As a habitat management tool, fires were used to maintain high quality forage for deer, elk, woodland caribou, bison and other game species. It was also used as a hunting tool to both drive and encircle game. During warfare, fire was strategically employed to drive away advancing enemies, create cover for escape, and for waging attacks.

In addition to maintaining open conditions, fire plays a critical role in maintaining species diversity. A re-census of 54 prairie remnants in Wisconsin found that 8 to 60% of the original plant species recorded at the sites had been lost over time (32 to 52 years) even though the sites ap-



peared relatively undisturbed (Leach and Givnish 1996). The authors suggest that taller vegetation outcompeted species with small stature, small seeds (e.g., orchids), and nitrogen-fixing symbioses such as members of the legume family (*Fabaceae*), including lupine (*Lupinus perennis*), wild indigo (*Baptisia* spp.), bush clover (*Lespedeza* spp.), and tick-trefoil (*Desmodium* spp.). Because fire maintains open conditions and burns off accumulated leaf litter, small species and those with small seeds that require open microsites are able to garner enough space and light to remain viable. In the absence of frequent fires, small species are outcompeted by taller and denser types of vegetation. As fire volatilizes much of the nitrogen stored in combustible vegetation, frequent burning also favors species that form nitrogen-fixing symbioses (e.g., legumes and rhizobium bacteria) and thus provides these plants with a competitive edge not found in unburned sites (Leach and Givnish 1996).

Fire also helps maintain species diversity by facilitating expression of the soil seed bank and promoting seed germination and establishment. By consuming accumulated and standing leaf litter, fire increases light availability to the soil surface and increases diurnal temperature fluctuations, both of which trigger seed germination. In addition, the removal of litter by fire creates critical microsites for seed germination and fosters seedling establishment.

The removal of litter by fire also increases the availability of many important plant nutrients (e.g., N, P, K, Ca and Mg), which are thought to contribute to higher plant biomass, increased flowering and seed production, and greater palatability to herbivores (Vogl 1964, Daubenmire 1968, Viro 1974, Vogl 1974, Smith and Kadlec 1985, Abrams et al. 1986, Collins and Gibson 1990, Reich et al. 1990, Schmalzer and Hinkle 1992, Timmins 1992, Laubhan 1995, Warners 1997).

While this discussion has focused on plants it is important to note that these species serve as host plants for a variety of insects and the structure of open grasslands is critical to a wide variety of animal species, many of which are considered rare or declining today (see Other Noteworthy Species section).

Ants, particularly the genus *Formica*, play an important role in mixing and aerating prairie soils (Curtis 1959, Trager 1998). Large ant mounds, which may measure half a meter in height and over one meter wide and number 40

to 50 per acre are especially conspicuous following a prairie fire (Curtis 1959). Because of their abundance and frequent habit of abandoning old mounds and building new ones, ants overturn large portions of prairies in a relatively short time (Curtis 1959). Other important species contributing to soil mixing and aeration include moles, mice, skunks, and badgers (Curtis 1959).

Historically, large herbivores such as bison significantly influenced plant species diversity in Michigan prairie and oak savanna ecosystems. The diet of bison consists of 90 to 95% grasses and sedges (Steuter 1997). As bison selectively forage on grasses and sedges, they reduce the dominance of graminoids and provide a competitive advantage to forb species. The activities of bison, which includes wallowing and trampling, promotes plant species diversity by creating microsites for seed germination and seedling establishment and reducing the dominance of robust perennials (Steuter 1997).

Mesic prairies are thought to result from the catastrophic destruction of mesic southern forest (beech-sugar maple forest or sugar maple-basswood forest) by fire (Curtis 1959, Grimm 1984). Evidence for its origin from mesic forest is based on soil type, landscape position, vegetation, and observation (Curtis 1959, Chapman 1984, Caton 1870 in Curtis 1959). In southern Michigan both mesic prairie and mesic southern forest frequently occurred on the same types of rich forest soils and both occupy level to gently rolling terrain (Curtis 1959, Chapman 1984). A variety of mesic forest plants occur within mesic prairies and are thought to represent relict populations from a previously forested environment rather than having colonized the open prairie from nearby forests (Curtis 1959, Chapman 1984). In addition, mesic prairie typically lacks oak grubs, which were very abundant in bur oak plains, oak openings, and dry-mesic prairie and were tolerant of the annual fires that swept through these communities. Thus, while dry-mesic prairie and oak openings are strongly related to the demise of an oak forest canopy by annual fires, the lack of oak grubs in mesic prairie may indicate its origin from a different type of forest, one that occurs on rich, mesic soils and lacks abundant oak trees. Lastly, Curtis provides detailed accounts of windswept prairie fires moving through adjacent hardwood forests and resulting in complete destruction of the forest canopy and conversion of these former forests to prairies (see Curtis 1959, pp. 303-304).



Vegetation Description: Unfortunately, no detailed ecological study of mesic prairie was completed in Michigan before the nearly total demise of the community. What information is available comes from written descriptions of the community by early European settlers and from studies of small prairie remnants.

In 1881, while recalling Prairie Rhonde and Gourdneck Prairie, both in southwest Kalamazoo County, Brown eulogizes “Early in March the rank growth of last year’s grass, dried by the sun and wind, was set on fire and the whole prairie burned over, leaving it bare and black as midnight. Then in a few days came the beautiful flowers, covering the whole prairie with one uniform kind of color; first the blue violet [*Viola pedatifida*]; then purple phlox [*Phlox pilosa*], and this succeeded by some other color. In July and August a tall yellow flower, the name of which I do not know [*Silphium* spp. or *Ratibida pinnata*?], mixed profusely with the tall grass, gave yellow as the predominating color” (Brown 1981 in Chapman 1984).

Another early account (Taylor 1855), eloquently describes Grand Prairie near Kalamazoo in 1855: “Beneath, about, and beyond me, as far as the eye could reach, was spread out, in undulating elegance, an emerald carpet of nature’s choicest fabric, inlaid profusely with flowers of every imaginable variety of name and tint – gorgeous and fascinating as the most brilliant hues of the rainbow” (Taylor 1855 in Chapman 1984).

Several early accounts of Michigan mesic prairies attest to their open, treeless character. Wheating and Bergquist write in 1923, “the rich black, open prairies which required no clearing were undoubtedly the choice of the first settlers. The Sturgis prairie and the White Pigeon prairie [both in St. Joseph County] show evidence of having been the homes of these pioneers” (Wheating and Bergquist 1923 in Chapman 1984). Brown describes Prairie Rhonde as he found it in 1831 as “...covered with a pretty rank growth of grass, then dry and sere, no tree except the Big Island grove and one or two other small groves” (Brown 1981 in Chapman 1984).

The transition from the open bur oak plains to prairie was apparently clearly demarcated based on Coffinberry’s 1880 account of the area around Nottawa-Sippi Prairie in St. Joseph County. “On the one hand stretched bur-oak plains, spread with a verdant carpet, variegated with dazling wild flowers, without an obstacle to intercept the

view for miles, save somber trunks of low oaks, sparsely spreading their shadows across the lawn; on the other hand arose the undulations of the white oak openings, with picturesque outlines of swells and slopes gracefully sweeping and sharply defined in the distance. Then, there lay the majestic prairie, grand in expansive solitude, its fringe of timber, as seen in the distance, resembling a diligently trained and well-trimmed garden parterre” (Coffinberry 1880 in Chapman 1984).

Chapman (1984) completed a study of 66 prairie and savanna remnants in southern Lower Michigan, four of which he classified as mesic prairie. In addition, Curtis (1959) and Curtis and Green (1949) collected detailed information on 45 mesic prairie stands in Wisconsin and much of their data may be applicable to Michigan’s mesic prairies.

Mesic prairie supports a dense to moderately dense growth of medium to tall vegetation (Chapman 1984). The community is dominated by big bluestem, little bluestem, and Indian grass, which can occur in varying degrees of dominance (Chapman 1984). In addition to the three grasses mentioned above, Curtis (1959) lists three additional grass species as dominant in mesic prairie in Wisconsin, porcupine grass (*Stipa spartea*), prairie dropseed (*Sporobolus heterolepis*), and Leigberg’s panic grass (*Panicum leibergii*). While all three are native to southwestern Michigan, it is important to note that both prairie dropseed and Leigberg’s panic grass are now rare in Michigan and listed as State Threatened. Switch grass (*Panicum virgatum*) is mentioned within the text by Curtis (1959) but does not appear within his list of prevalent species for the community and the species occurs sparsely (< 5% frequency) in only one site studied by Chapman (1984).

The following table of mesic prairie plants was compiled from Chapman’s (1984) study of four mesic prairie sites and includes native species occurring in at least 50% of the sites he studied.

SCIENTIFIC NAME	COMMON NAME
Grasses and sedges	
<i>Andropogon gerardii</i>	big bluestem
<i>Andropogon scoparius</i>	little bluestem grass
<i>Carex bicknellii</i>	sedge
<i>Panicum oligoanthos</i>	panic grass
<i>Sorghastrum nutans</i>	Indian grass
<i>Spartina pectinata</i>	prairie cordgrass



Forbs

<i>Anemone cylindrica</i>	thimbleweed
<i>Antennaria parlinii</i>	smooth pussytoes
<i>Apocynum androsaemifolium</i>	spreading dogbane
<i>Asclepias syriaca</i>	common milkweed
<i>Asclepias tuberosa</i>	butterfly weed
<i>Aster laevis</i>	smooth aster
<i>Aster oolentangiensis</i>	sky-blue aster
<i>Cacalia atriplicifolia</i>	pale Indian plantain
<i>Comandra umbellata</i>	bastard toadflax
<i>Coreopsis palmata</i>	prairie coreopsis
<i>Coreopsis tripteris</i>	tall coreopsis
<i>Desmodium illinoense</i>	prairie tick-trefoil
<i>Dioscorea villosa</i>	wild yam
<i>Erigeron annuus</i>	annual fleabane
<i>Erigeron strigosus</i>	daisy fleabane
<i>Eryngium yuccifolium</i>	rattlesnake master
<i>Euphorbia corollata</i>	flowering spurge
<i>Euthamia graminifolia</i>	grass-leaved goldenrod
<i>Fragaria virginiana</i>	wild strawberry
<i>Frasera caroliniensis</i>	American columbo
<i>Galium boreale</i>	northern bedstraw
<i>Galium pilosum</i>	hairy bedstraw
<i>Gentiana flavida</i> (E)	yellowish gentian
<i>Geranium maculatum</i>	wild geranium
<i>Helianthemum canadense</i>	common rockrose
<i>Helianthus divaricatus</i>	woodland sunflower
<i>Helianthus occidentalis</i>	western sunflower
<i>Helianthus strumosus</i>	pale-leaved sunflower
<i>Heuchera americana</i>	alum root
<i>Kuhnia eupatorioides</i> (SC)	false boneset
<i>Lactuca canadensis</i>	tall lettuce
<i>Lespedeza capitata</i>	round-headed bush clover
<i>Lespedeza hirta</i>	hairy bush clover
<i>Lithospermum canescens</i>	hoary puccoon
<i>Luzula multiflora</i>	common wood rush
<i>Monarda fistulosa</i>	wild bergamot
<i>Phlox pilosa</i>	prairie phlox
<i>Potentilla simplex</i>	old field cinquefoil
<i>Ranunculus fascicularis</i>	early buttercup
<i>Ratibida pinnata</i>	yellow coneflower
<i>Rudbeckia hirta</i>	black-eyed susan
<i>Silphium integrifolium</i> (T)	rosin weed
<i>Silphium terebinthinaceum</i>	prairie dock
<i>Smilacina racemosa</i>	false spikenard
<i>Solidago altissima</i>	tall goldenrod
<i>Solidago nemoralis</i>	old field goldenrod
<i>Solidago rigida</i>	stiff goldenrod

<i>Solidago speciosa</i>	showy goldenrod
<i>Taenidia integerrima</i>	yellow pimpernel
<i>Thalictrum dasycarpum</i>	purple meadow rue
<i>Tradescantia ohiensis</i>	common spiderwort
<i>Verbena stricta</i>	hoary vervain
<i>Veronicastrum virginicum</i>	Culver's root
<i>Vicia americana</i>	American vetch
<i>Viola sororia</i>	common blue violet
<i>Zizia aurea</i>	golden alexanders

Ferns and Fern Allies

<i>Equisetum laevigatum</i>	smooth scouring rush
<i>Pteridium aquilinum</i>	bracken fern

Shrubs

<i>Ceanothus americanus</i>	New Jersey tea
<i>Cornus foemina</i>	gray dogwood
<i>Corylus americana</i>	hazelnut
<i>Rhus glabra</i>	smooth sumac
<i>Rhus typhina</i>	staghorn sumac
<i>Rosa carolina</i>	pasture rose
<i>Rubus hispidus</i>	swamp dewberry
<i>Salix humilis</i>	prairie willow

Michigan Indicator Species: Chapman (1984) lists the following native species as indicators of mesic prairie: pale Indian plantain (*Cacalia atriplicifolia*), sedge (*Carex bicknellii*), prairie coreopsis (*Coreopsis palmata*), American columbo (*Frasera caroliniensis*), yellowish gentian (*Gentiana flavida*), wild geranium (*Geranium maculatum*), pale-leaved sunflower (*Helianthus strumosus*), false boneset (*Kuhnia eupatorioides*), bur oak (*Quercus macrocarpa*), smooth sumac (*Rhus glabra*), yellow pimpernel (*Taenidia integerrima*), hoary vervain (*Verbena stricta*), American vetch (*Vicia americana*), prairie violet (*Viola pedatifida*) and common blue violet (*Viola sororia*). Species more common in wet-mesic or lowland sites but which can serve as indicators of mesic prairie when found in upland sites include the following: wild yam (*Dioscorea villosa*), daisy fleabane (*Erigeron annuus*), rattlesnake master (*Eryngium yuccifolium*), swamp dewberry (*Rubus hispidus*), rosin weed (*Silphium integrifolium*), prairie dock (*S. terebinthinaceum*), tall goldenrod (*Solidago altissima*), prairie cord grass (*Spartina pectinata*), tall meadow rue (*Thalictrum dasycarpum*), and golden alexanders (*Zizia aurea*) (Chapman 1984). It is noteworthy that none of the indicators listed by Curtis (1959) for mesic prairie in Wisconsin are considered mesic prairie indicators in Michigan by Chapman (1984).



Other Noteworthy Species: Rare plant species associated with mesic prairie are listed below along with their status, which is indicated by the following abbreviations: X, extirpated from state; E, State Endangered; T, State Threatened; SC, State Species of Special Concern.

Scientific Name	Common Name	Status
<i>Amorpha canescens</i>	lead-plant	SC
<i>Baptisia lactea</i>	white false indigo	SC
<i>Baptisia leucophaea</i>	cream wild indigo	E
<i>Coreopsis palmata</i>	prairie coreopsis	T
<i>Dodecatheon meadia</i>	shooting-star	E
<i>Echinacea purpurea</i>	purple coneflower	X
<i>Eryngium yuccifolium</i>	rattlesnake-master	T
<i>Gentiana flavida</i>	white gentian	E
<i>Panicum leibergii</i>	Leigberg's panic grass	T
<i>Polygala incarnata</i>	pink milkwort	X
<i>Silphium integrifolium</i>	rosinweed	T
<i>Silphium laciniatum</i>	compass-plant	T
<i>Sisyrinchium strictum</i>	blue-eyed-grass	SC
<i>Spiranthes ovalis</i>	lesser ladies'-tresses	T
<i>Sporobolus heterolepis</i>	prairie dropseed	SC
<i>Viola pedatifida</i>	prairie birdfoot violet	T

Rare animal species associated with mesic prairie include the following:

Grassland Birds: Henslow's sparrow (*Ammodramus henslowii*) (SC), grasshopper sparrow (*Ammodramus savannarum*) (SC), short-eared owl (*Asio flammeus*) (E), long-eared owl (*Asio otus*) (T), northern harrier (*Circus cyaneus*) (SC), migrant loggerhead shrike (*Lanius ludovicianus migrans*) (E), Dickcissel (*Spiza americana*) (SC), western meadowlark (*Sturnella neglecta*) (SC), and barn owl (*Tyto alba*) (E).

Insects: blazing star borer (*Papaipema beeriana*) (SC), Culver's root borer (*Papaipema sciata*) (SC), Silphium borer (*Papaipema silphii*) (T), leadplant flower moth (*Schinia lucens*) (E), red-legged spittlebug (*Prosapia ignipectus*) (SC), Sprague's pygarcia (*Pygarcia spraguei*) (SC), and Spartina moth (*Spartiniphaga inops*) (SC).

Mammals: prairie vole (*Microtus ochrogaster*) (E)

Reptiles: eastern massasauga (*Sistrurus c. catenatus*) (SC and Federal Candidate Species), black rat snake (*Elaphe o. obsoleta*) (SC), Kirtland's snake (*Clonophis kirtlandii*) (T), and eastern box turtle (*Terrapene c. carolina*) (SC). Spotted turtle (*Clemmys*

guttata) (T) and Blanding's turtle (*Emydoidea blandingii*) (SC) may nest in mesic prairie when it occurs adjacent to wetlands.

Conservation and Management: Efforts should be made to identify, protect, and manage remnants of mesic prairie. Several studies to identify prairie remnants in Michigan have been undertaken and most remnants are very small and/or occur as narrow strips adjacent to railroads (Scharrer 1972, Thompson 1970, 1975, and 1983, Chapman 1984). The small size and poor landscape context of most remnant mesic prairies make large-scale restoration of existing prairies nearly impossible. Prairie plantings located in areas of former mesic prairie, especially on the Battle Creek Outwash Plain (Albert 1995), are particularly needed.

Managing mesic prairie requires frequent burning, from annual to every two to three years. Longer burn intervals will result in tree and tall shrub encroachment. Prescribed burning is required to protect and enhance plant species diversity and prevent encroachment of trees and tall shrubs, which outcompete light-demanding prairie plants. In prairie remnants where fire has been excluded for long periods (e.g., decades), local extinctions of plant species are common (Leach and Givnish 1996).

In addition to prescribed fire, brush cutting accompanied by stump application of herbicide is an important component of prairie restoration. While fires frequently kill woody seedlings, long established trees and tall shrubs like black cherry (*Prunus serotina*) and dogwoods (*Cornus* spp.) typically resprout and can reach former levels of dominance within two to three years. Herbicide application to cut stumps will prevent resprouting.

To reduce the impacts of management on fire-intolerant species it will be important to consider a rotating schedule of prescribed burning in which adjacent management units are burned in alternate years. This is especially important when planning burns in open grasslands such as mesic prairie. Insect species that are restricted to these habitats have already experienced severe losses in the amount of available habitat due to forest succession brought on by years of fire suppression. By burning adjacent management units in alternate years, insect species



from unburned units may be able to recolonize burned areas (Panzer et al. 1995). Avian species diversity is also thought to be enhanced by managing large areas as a mosaic of burned and unburned patches (Herkert et al. 1993).

Prairie ants (*Formica*) are an extremely important component of grassland communities and research indicates that they respond with population increases to restoration activities, especially prescribed fire (Trager 1998). Prescribed burning precipitates changes in the dominance of ant species from carpenter and woodland ants (*Camponotus* and *Aphaenogaster*) to prairie ants because it reduces woody vegetation and detritus used by the arboreal and litter- and twig-nesting species in favor of species restricted to grassland habitats (Trager 1998). Restorations involving prairie plantings near old fields or remnant prairies are typically colonized by several species of prairie ants within a few years (Trager 1990).

Controlling invasive species is a critical step in restoring and managing mesic prairie. By outcompeting native species, invasives alter vegetation structure, reduce species diversity, and upset delicately balanced ecological processes such as trophic relationships, interspecific competition, nutrient cycling, soil erosion, hydrologic balance, and solar insolation (Bratton 1982, Harty 1986). At present some of the most aggressive invasive species that threaten biodiversity of grassland communities include reed canary grass (*Phalaris arundinacea*), white and yellow sweet clover (*Melilotus alba* and *M. officinalis*), spotted knapweed (*Centaurea maculosa*), autumn olive (*Elaeagnus umbellata*), multiflora rose (*Rosa multiflora*), common buckthorn (*Rhamnus cathartica*), Eurasian honeysuckles (*Lonicera maackii*, *L. morrowii*, *L. tatarica*, *L. x bella*.), and black locust (*Robinia pseudoacacia*).

In addition to reestablishing ecological processes such as fire, most restoration sites will require the reintroduction of appropriate native species and genotypes. Plants can be reintroduced both through seeding or seedling transplants. Small, isolated prairie remnants may harbor plant populations that have suffered from reduced gene flow. Restoration efforts at isolated prairie remnants should consider introducing seeds collected from nearby stocks to augment and maintain genetic diversity of remnant plant populations. The Michigan Native Plant Producers Association may be a helpful resource for locating sources of Michigan genotypes (<http://www.nohlc.org/MNPPA.htm>).

Several helpful guides are available for restoring prairies and starting prairie plants from seed (Packard and Mutel 1997, Nuzzo 1976, Schulenberg 1972). See Packard and Mutel (1997) for a comprehensive treatment of the subject and additional references

Restoration and management of grasslands such as mesic prairie are critically important to grassland birds, which have suffered precipitous population declines due to habitat loss and changing agricultural practices (e.g., early mowing of hay fields). Detailed habitat management guidelines for grassland birds have been developed by Herkert et al. (1993) and Sample and Mossman (1997). Listed below are several of the recommendations suggested by Herkert et al. (1993) (see publication for complete list of management guidelines).

1. Avoid fragmentation of existing grasslands.
2. Grassland restorations aimed at supporting populations of the most area-sensitive grassland birds should be at least 125 acres and preferably more than 250 acres in size. Area sensitive species requiring large patches of grassland (>100 acres) include northern harrier (SC), bobolink (*Dolichonyx oryzivorus*), savannah sparrow (*Passerculus sandwichensis*), Henslow's sparrow (SC), grasshopper sparrow (SC), eastern meadowlark (*Sturnella magna*), western meadowlark, sedge wren (*Cistothorus platensis*), sharp-tailed grouse (*Pedioecetes phasianellus*), upland sandpiper (*Bartramia longicauda*), short-eared owl (E), and barn owl (E) (Herkert et al. 1993, Sample and Mossman 1997). Patches of grassland less than 50 acres will benefit the least area-sensitive grassland birds such as northern bobwhite (*Colinus virginianus*), red-winged black bird (*Agelaius phoeniceus*), American goldfinch (*Carduelis tristis*), Vesper sparrow (*Pooecetes gramineus*), field sparrow (*Spizella pusilla*), song sparrow (*Melospiza melodia*), dickcissel (SC), and common yellowthroat (*Geothlypis trichas*) (Herkert et al. 1993).
3. Maximize interior grassland habitat by establishing circular (best) or square grassland plantings and avoiding long, narrow plantings, which increase edge habitat.
4. Where grassland habitats border forests, strive to create a feathered edge by allowing prescribed fires to burn through adjacent forests as opposed to installing firebreaks along the forest edge. Grasslands with feathered edges experience lower rates of nest predation than those with sharply contrasting edges (Ratti and Reese 1988).



Research Needs: Remaining remnants of mesic prairie need to be identified, protected, and managed. Further research on the historical plant species composition of mesic prairie in Michigan would be useful for developing seed mixes for restoration. Genetic studies of the effects of small, isolated populations on plant species genetic diversity will provide information on managing remnants of mesic prairie. Research on the utilization of restored and remnant prairies by grassland birds and insects will provide useful information for understanding how mesic prairies contribute to biodiversity. Studies on methods of prairie establishment and management, including controlling invasive species, will benefit both ongoing and new efforts to restore mesic prairie. Conservation and management efforts will benefit from further study of how species composition is influenced by fire frequency, intensity, and periodicity.

Similar Communities: Dry-mesic prairie, bur oak plains, oak openings, mesic sand prairie, and wet-mesic prairie

Other Classifications:

Michigan Natural Features Inventory Circa 1800s Vegetation (MNFI): Grassland

The Nature Conservancy U.S. National Vegetation Classification and International Classification of Ecological Communities (Faber-Langendoen 2001, NatureServe 2004):

CODE; ALLIANCE; ASSOCIATION; COMMON NAME

V.A.5.N.a;

Andropogon gerardii – (*Sorghastrum nutans*)

Herbaceous Alliance;

Andropogon gerardii – *Sorghastrum nutans* – (*Sporobolus heterolepis*) – *Liatris* spp. – *Ratibida pinnata* Herbaceous Vegetation;

Big Bluestem – Yellow Indiangrass – (Prairie Dropseed) – Blazingstar Species – Grey-head Prairie Coneflower Herbaceous Vegetation

Related Abstracts: bur oak plains, oak openings, oak barrens, dry-mesic prairie, Culver's root borer, eastern box turtle, eastern massasauga, Henslow's sparrow, migrant loggerhead shrike, northern harrier, red-legged spittlebug, and prairie dropseed.

Literature Cited:

- Abrams, M.D., A.K. Knapp and L.C. Hulbert. 1986. A ten year record of aboveground biomass in a Kansas tallgrass prairie: Effects of fire and topographic position. *American Journal of Botany* 73: 1509-15.
- Albert, D.A. 1995. Regional landscape ecosystems of Michigan, Minnesota, and Wisconsin: A working map and classification. North Central Forest Experiment Station, Forest Service – U.S. Department of Agriculture. St. Paul, MN.
- Albert, D.A. and M.A. Kost. 1998. Natural community abstract for lakeplain wet-mesic prairie. Michigan Natural Features Inventory, Lansing, MI. 6 pp.
- Bratton, S. P. 1982. The Effects of Exotic Plant and Animal Species on Nature Preserves. *Natural Areas Journal*. 2(3):3-13.
- Brewer, L.R. H.A. Raup and T.W. Holder. 1984. Presettlement vegetation of southwest Michigan (map). Western Michigan University, Department of Geology, Kalamazoo, Michigan.
- Brown, E.L. 1981. Speech at the ninth annual reunion of the pioneers of Kalamazoo County, Michigan Pioneer Historical Collections 3:523-526.
- Chapman, K.A. 1984. An ecological investigation of native grassland in southern Lower Michigan. M.S. Thesis, Western Michigan University, Kalamazoo, MI. 235 pp.
- Collins, S.L. and D.J. Gibson. 1990. Effects of fire on community structure in tallgrass and mixed grass prairie. Pp. 81-98 in S.L. Collins and L.L. Wallace (eds.), *Fire in North American tallgrass prairies*, University of Oklahoma Press, Norman, OK.
- Comer, P.J., D.A. Albert, H.A. Wells, B.L. Hart, J.B. Raab, D.L. Price, D.M. Kashian, R.A. Corner and D.W. Schuen. 1995. Michigan's presettlement vegetation, as interpreted from the General Land Office Surveys 1816-1856. Michigan Natural Features Inventory, Lansing, MI. Digital Map.
- Curtis, J.T. 1959. *Vegetation of Wisconsin*. The University of Wisconsin Press, Madison, WI. 657 pp.
- Curtis, J.T. and H.C. Greene. 1949. A study of relic Wisconsin prairies by the species-presence method. *Ecology* 30:152-55.
- Daubenmire, R. 1968. Ecology of fire in grasslands. *Advances in Ecological Research* 5:209-266.
- Day, G.M. 1953. The Indian as an ecological factor in the northeast forest. *Ecology* 34:329-346



- Denevan, W.M. 1992. The pristine myth: The landscape of the Americas in 1492. *Annals of the Association of American Geographers* 83:369-385.
- Faber-Langendoen, D. editor. 2001. Plant communities of the Midwest: Classification in an ecological context. Association for Biodiversity Information, Arlington, VA. 61 pp. + appendix (705 pp.).
- Grimm, E.C. 1984. Fire and other factors controlling the big woods vegetation of Minnesota in the mid-nineteenth century. *Ecological Monographs* 54:291-311.
- Harty, F. M. 1986. Exotics and their ecological ramifications. *Natural Areas Journal*. 6(4):20-26.
- Herkert, J.R., R.E. Szafoni, V.M. Kleen, and J.E. Schwegman. 1993. Habitat establishment, enhancement and management for forest and grassland birds in Illinois. Division of Natural Heritage, Illinois Department of Conservation, Natural Heritage Technical Publication #1, Springfield, IL, 20pp.
- Holder, T.W., R. Brewer, L.G. Brewer and H.A. Raup. 1981. Presettlement vegetation of Kalamazoo County (map). Department of Geography, Western Michigan University, Kalamazoo MI.
- Kay, C.E. 1995. Aboriginal overkill and Native burning: Implications for modern ecosystem management. *Western Journal of Applied Forestry* 10:121-126.
- Kost, M.A., D.A. Albert, J.G. Cohen, B.S. Slaughter, R.K. Schillo, C.R. Weber, and K.A. Chapman. 2007. Natural Communities of Michigan: Classification and Description. Michigan Natural Features Inventory, Report Number 2007-21, Lansing, MI. 314 pp.
- Laubhan, M.K. 1995. Effects of prescribed fire on moist-soil vegetation and macronutrients. *Wetlands* 15:159-66.
- Leach, M.K. and T.J. Givnish. 1996. Ecological determinants of species loss in remnant prairies. *Science* 273:1555-1558.
- NatureServe. 2004. NatureServe Explorer: An online encyclopedia of life [web application]. Version 1.8. NatureServe, Arlington, Virginia. Available <http://www.natureserve.org/explorer>. (Accessed: September 20, 2004).
- Nuzzo, V. 1976. Propagation and planting of prairie forbs and grasses in southern Wisconsin. Pp. 182-189 in *Proceedings of the Fifth Midwest Prairie Conference*. Iowa State University, Ames, Iowa.
- Packard, S. and C.F. Mutel. 1997. The tallgrass restoration handbook for prairies savannas and woodlands. Island Press, Washington D.C. 463pp.
- Panzer, R.D., D. Stillwaugh, R. Gnaedinger, and G. Derkowitz. 1995. Prevalence of remnant dependence among prairie-and savanna-inhabiting insects of the Chicago region. *Natural Areas Journal* 15:101-116.
- Ratti, J.T. and K.P. Reese. 1988. Preliminary test of the ecological trap hypothesis. *Journal of Wildlife Management* 52:484-491.
- Reich, P.B., M.D. Abrams, D.S. Ellsworth E. L. Kruger and T. J. Tabone. 1990. Fire affects ecophysiology and community dynamics of Central Wisconsin oak forest regeneration. *Ecology* 71:2179-90.
- Sample, D.W. and M.J. Mossman. 1997. Managing habitat for grassland birds: A guide for Wisconsin. Bureau of Integrated Science Services, Department of Natural Resources, Madison, WI. 154pp.
- Scharrer, E.M. 1972. Relict prairie flora of southwestern Michigan. Pp. 9-12 in J.H. Zimmerman (ed.) *Proceedings of the Second Midwest Prairie Conference*, Madison, WI. 242 pp.
- Schmalzer, P.A. and C. R. Hinkle. 1992. Soil dynamics following fire in *Juncus* and *Spartina* marshes. *Wetlands* 12:8-21.
- Schulenberg, R. 1972. Notes on the propagation of prairie plants. The Morton Arboretum, Lisle IL. 15p.
- Smith, L.M. and J.A. Kadlec. 1985. Fire and herbivory in a Great Salt Lake marsh. *Ecology* 66:259-65.
- Steuter, A.A. 1997. Bison. Pp. 339-347 in Packard, S. and C.F. Mutel (eds.), *The Tallgrass Restoration Handbook for Prairies Savannas and Woodlands*. Island Press, Washington D.C. 463 pp.
- Taylor, H.S. 1855. Ladies Library Association quarter centennial celebration of the settlement of Kalamazoo, Michigan. Kalamazoo, MI: Gazette Printers.
- Thompson, D.Q. and R.H. Smith. 1970. The forest primeval in the Northeast - a great myth? in *Proceedings of the Tall Timbers Fires Ecology Conference*. 10:255-265.
- Thompson, P.W. 1970. The preservation of prairie stands in Michigan. Pp. 13-14 in J.H. Zimmerman (ed.) *Proceedings of the Second Midwest Prairie Conference*, Madison, WI. 242 pp.
- Thompson, P.W. 1975. The floristic composition of prairie stands in southern Michigan. pp. 317-331 in M.K. Wali (ed.), *Prairie: A multiple view*. The University of North Dakota, Grand Fork, N.D.



- Thompson, P.W. 1983. Composition of prairie stands in southern Michigan and adjoining areas. Pp. 105-111 in R. Brewer (ed.), Proceedings of the Eighth North American Prairie Conference.
- Timmins, S.M. 1992. Wetland vegetation recovery after fire: Eweburn Bog, Te Anau, New Zealand. New Zealand Journal of Botany 30:383-99.
- Trager, J.C. 1990. Restored prairies colonized by native prairie ants (Missouri, Illinois). Restoration and Management Notes 8:104-105.
- Trager, J.C. 1998. An introduction to ants (*Formicidae*) of the tallgrass prairie. Missouri Prairie Journal 18:4-8.
- Transeau, E.N. 1935. The prairie peninsula. Ecology 16:423-437.
- Viro, P.J. 1974. Effects of forest fire on soil. Pp. 7-45 in T. T. Kozlowski and C. E. Ahlgren (eds.), Fire and Ecosystems. Academic Press, New York, NY.
- Vogl, R.J. 1964. The effects of fire on a muskeg in northern Wisconsin. Journal of Wildlife Management 28:317-29.
- Vogl, R.J. 1974. Effects of fire on grasslands. Pp. 139-94 in T. T. Kozlowski and C. E. Ahlgren (eds.), Fire and Ecosystems. Academic Press, New York, NY.
- Warners, D.P. 1997. Plant diversity in sedge meadows: Effects of groundwater and fire. Ph.D. dissertation, University of Michigan, Ann Arbor, MI. 231 pp.

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