Draft

BRISTOL BAY RED KING CRAB STOCK ASSESSMENT IN Spring 2009

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EXECUTIVE SUMMARY

A length-based model was applied to eastern Bering Sea trawl survey, catch sampling, and commercial catch data to estimate stock abundance of Bristol Bay red king crab (*Paralithodes camtschaticus*) during 1968-2008. Three scenarios were compared for the model: (1) constant natural mortality (0.18), estimation of additional mortality for males during 1980-1984 and for females during 1976-1993, and with the Bering Sea Fisheries Research Foundation (BSFRF) survey data; (2) constant natural mortality (0.18) with BSFRF survey data; and (3) constant natural mortality (0.18), estimation of additional mortality for males during 1980-1984 and for females during 1976-1993, but without the BSFRF survey data. Scenarios (1) and (3) were compared to illustrate the value of the BSFRF surveys. Scenario (2) fits the data very poorly, and scenario (1) fits the data well and its results were used to estimate biological reference points and federal overfishing levels for 2009.

Average recruitments during three periods were used to estimate $B_{35\%}$: 1968-2008, 1985-2008, and 1995-2008. We recommend using the average recruitment during 1995-2008, which was used in 2008 to set the overfishing limits. There are several reasons for supporting our recommendation. First, estimated recruitment was higher after 1994 than during 1985-1994 and there was a potential regime shift after 1989 (Overland et al. 1999), which corresponded to recruitment in 1995 and later. Second, recruitments estimated

before 1985 came from a potentially higher natural mortality than we used to estimate $B_{35\%}$. Third, high recruitments during the late 1960s and 1970s generally occurred when the spawning stock was primarily located in southern Bristol Bay while the current spawning stock is mainly in the middle of Bristol Bay. The current flows favor larvae hatched in southern Bristol Bay (see the section on Ecosystem Considerations). The spawning stock has not moved back to southern Bristol Bay during the last 20 years. Finally, the $B_{35\%}$ based on average recruitment during 1969-2008 is higher than mature male biomasses in 34 out of 40 years and is close to 200% of the mean mature male biomass during 1968-2008. Even without fishing, the recruitment levels during the last 20 years could hardly produce a mature male biomass as high as the $B_{35\%}$ *based on the mean recruitment of 1969-2008*. Therefore, with this high $B_{35\%}$ we have to close the fishery and wait for some things to happen to rebuild the stock.

Based on the $B_{35\%}$ estimated from the average male recruitment during 1995-2008, the biological reference points were estimated as follows:

$B_{35\%}$ = 70.627 million lbs, or 32,036 t,

$F_{35\%}$ = 0.33,

$F_{40\%}$ = 0.26.

Based on $B_{35\%}$ and $F_{35\%}$, the retained catch and total male catch limits for 2009 were estimated to be:

Retained catch: XXX million lbs, or XXX t,

Total male catch: XXX million lbs, or XXX t,

MMB on 2/15/2010: XXX million lbs, or XXX t.

Summary of Major Changes in 2009

1. Pot fisheries data in 2008/2009 were included.
2. Areas-swept for the NMFS surveys have been re-estimated and trawl survey abundances have been re-estimated, which are generally lower than previous assessments.
3. The trawl survey data from the Bering Sea Fisheries Research Foundation (BSFRF) were included in the model.

4. Red king crab bycatch from the eastern Bering Sea Tanner crab fishery was included to estimate bycatch mortality.

5. The model was extended to the data during 1968 to 2009.

6. Three scenarios were compared: (1) constant M=0.18 with additional mortality during 1976-1993 and with BSFRF data, (2) constant M=0.18 with BSFRF data, and (3) constant M=0.18 with additional mortality during 1976-1993 but without BSFRF data.

7. Sizes at maturity for female red king crab were estimated annually, and different sizes at maturity during three periods (1968-1982, 1983-1993, and 1994-2008) were used to estimate growth increments per molt for females.

8. No weighting factors were used for survey biomass. All survey biomasses were weighted by CVs or variances.

9. A Bayesian approach was used to estimate NMFS survey catchability (Q).

10. Spatial and temporal distributions of females and tow numbers of groundfish fisheries were plotted.

11. High mortality during the early 1980s, pertaining to the "red bag" hypothesis, was discussed.

Response to CPT Comments (from September 2008)

"The authors are commended for updating the assessment per May CPT and June SSC recommendations. However, the team made several recommendations for improvements to the assessment. For example, to the extent possible, the model should incorporate data prior to 1985. It was also suggested that future assessments include some analysis of model sensitivity to different weightings (lambdas). The magnitude of lambdas has a direct effect on projected biomass and catch likelihood profiles because increasing lambdas artificially decreases the width of the profiles. It was also recommended that the authors consider parameter estimation in a Bayesian context. The authors noted that patterns seem to exist in the trawl survey residuals for female crab; the female maturity curve is currently knife-edged. It is requested that the authors examine scenarios with attempt to address the female trawl survey residuals patterns.

In addition, it is requested that when key parameters are fixed in the model, more justification, such a sensitivity analysis, should be included for estimating parameters external to the model. The CPT specifically recommended investigating the sensitivity of the survey q which is fixed in the model. The team recommends that these additional analyses be incorporated into the assessment for the Spring 2009 review."

"The Plan Team encourages the authors to work closely with NMFS survey staff to ensure consistency between ADF&G and NMFS survey estimates. A follow-up to this process would also be the inclusion of estimates of precision with survey values. Survey abundances should be tabulated in the assessment."

These comments have been addressed as described under **Summary of Major Changes in 2009**. above.

Response to SSC Comments specific to this assessment (from June 2008)

"The SSC suggests that the authors address ecosystem considerations beyond predation of groundfish on crab (which was well covered). This section should also address apex predators, such as seabirds that rely on juvenile crab during winter, that might be affected by changes in the crab population. Although data on crab predation from apex predators may not be specific to this stock, there are data available for the region."

This is a good suggestion and we will address it during the future reports. During this reporting period, the assessment authors were fully engaged with model extension and addressing the mortality issues during the early 1980s.

1. *"The period of recruitment that was selected for estimating $B_{35\%}$ was based on a presumed oceanographic regime shift in 1989. However, little evidence for a shift in mean recruitment or for an effect of the regime shift on red king crab was provided. Future analyses should include a more thorough evaluation of recruitment trends based on a model fit to the full time series. Absent a strong rationale to the contrary, the reference time period should include periods of both high and low recruitment to better represent the average reproductive potential of the stock."*

The mean recruitment is higher for brood period 1990-2003 than that during brood period 1985-1989. Brood year 1990 corresponds to recruitment in 1995 due to a

time lag of 5 years (from hatching to recruitment). After extending the model to the data before 1985, different reference periods can be evaluated in 2009.

INTRODUCTION

Stock Structure

Red king crab (RKC), *Paralithodes camtschaticus*, are found in several areas of the Aleutian Islands and eastern Bering Sea. The State of Alaska divides the Aleutian Islands and eastern Bering Sea into three management registration areas to manage RKC fisheries: Aleutian Islands, Bristol Bay, and Bering Sea (Alaska Department of Fish and Game (ADF&G) 2005). The Aleutian Islands area covers two stocks, Adak and Dutch Harbor, and the Bering Sea area contains two other stocks, the Pribilof Islands and Norton Sound. The largest stock is found in the Bristol Bay area, which includes all waters north of the latitude of Cape Sarichef (54°36' N lat.), east of 168° W long., and south of the latitude of Cape Newenham (58°39' N lat.) (ADF&G 2005). Besides these five stocks, RKC stocks elsewhere in the Aleutian Islands and eastern Bering Sea are currently too small to support a commercial fishery. This report summarizes the stock assessment results for the Bristol Bay RKC stock.

Fishery

The RKC stock in Bristol Bay, Alaska, supports one of the most valuable fisheries in the United States (Bowers et al. 2008). The Japanese fleet started the fishery in the early 1930s, stopped fishing from 1940 to 1952, and resumed the fishery from 1953 until 1974 (Bowers et al. 2008). The Russian fleet fished for RKC from 1959 through 1971. The Japanese fleet employed primarily tanglenets with a very small proportion of catch from trawls and pots. The Russian fleet used only tanglenets. United States trawlers started to fish for Bristol Bay RKC in 1947, and effort and catch declined in the 1950s (Bowers et al. 2008). The domestic RKC fishery began to expand in the late 1960s and peaked in 1980 with a catch of 129.95 million lbs (58,943 t), worth an estimated $115.3 million ex-vessel value (Bowers et al. 2008). The catch declined dramatically in the early 1980s and has stayed at low levels during the last two decades (Table 1). After the stock collapse in the early 1980s, the Bristol Bay RKC fishery took place during a short period in the fall (usually

lasting about a week), with the catch quota based on the stock assessment conducted in the previous summer (Zheng and Kruse 2002a). As a result of new regulations for crab rationalization, the fishery was open longer beginning with the 2005/2006 season from October 15, 2005 to January 15, 2006. With the implementation of crab rationalization, historical guideline harvest levels (GHL) were changed to a total allowable catch (TAC). The GHL/TAC and actual catch are compared in Table 2. The implementation errors are quite high for some years, and total actual catch from 1980 to 2007 is about 6% less than the sum of GHL/TAC over that period (Table 2).

Fisheries Management

King and Tanner crab stocks in the Bering Sea and Aleutian Islands are managed by the State of Alaska through a federal king and Tanner crab fishery management plan (FMP). Under the FMP, management measures are divided into three categories: (1) fixed in the FMP, (2) frameworked in the FMP, and (3) discretion of the State of Alaska. The State of Alaska is responsible for developing harvest strategies to determine GHL/TAC under the framework in the FMP.

Harvest strategies for the Bristol Bay RKC fishery have changed over time. Two major management objectives for the fishery are to maintain a healthy stock that ensures reproductive viability and to provide for sustained levels of harvest over the long term (ADF&G 2005). In attempting to meet these objectives, the GHL/TAC is coupled with size-sex-season restrictions. Only males ≥6.5-in carapace width (equivalent to 135-mm carapace length, CL) may be harvested and no fishing is allowed during molting and mating periods (ADF&G 2005). Specification of TAC is based on a harvest rate strategy. Before 1990, harvest rates on legal males were based on population size, abundance of prerecruits to the fishery, and postrecruit abundance, and rates varied from less than 20% to 60% (Schmidt and Pengilly 1990). In 1990, the harvest strategy was modified, and a 20% mature male harvest rate was applied to the abundance of mature-sized (≥120-mm CL) males with a maximum 60% harvest rate cap of legal (≥135-mm CL) males (Pengilly and Schmidt 1995). In addition, a minimum threshold of 8.4 million mature-sized females (≥90-mm CL) was added to existing

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management measures to avoid recruitment overfishing (Pengilly and Schmidt 1995). Based on a new assessment model and research findings (Zheng et al. 1995a, 1995b, 1997a, 1997b), the Alaska Board of Fisheries adopted a new harvest strategy in 1996. That strategy had two mature male harvest rates: 10% when effective spawning biomass (ESB) is between 14.5 and 55.0 million lbs and 15% when ESB is at or above 55.0 million lbs (Zheng el al. 1996). The maximum harvest rate cap of legal males was changed from 60% to 50%. An additional threshold of 14.5 million lbs of ESB was also added. In 1997, a minimum threshold of 4.0 million lbs was established as the minimum GHL for opening the fishery and maintaining fishery manageability when the stock abundance is low. In 2003, the Board modified the current harvest strategy by adding a mature harvest rate of 12.5% when the ESB is between 34.75 and 55.0 million lbs. The current harvest strategy is illustrated in Figure 1.

The purpose of this report is to document the stock assessments for Bristol Bay RKC. This report includes (1) all data used to conduct the stock assessments, (2) details of the analytic approach, (3) an evaluation of the assessment results, (4) estimates of biological reference points and federal overfishing limits for 2008, and (5) future projections and the near future outlook.

DATA

Catch Data

Data on landings of Bristol Bay RKC by length and year and catch per unit effort were obtained from annual reports of the International North Pacific Fisheries Commission from 1960 to 1973 (Hoopes et al. 1972; Jackson 1974; Phinney 1975) and from the ADF&G from 1974 to 2008 (Bowers et al. 2008). Bycatch data are available starting from 1990 and were obtained from the ADF&G observer database and reports (Bowers et al. 2008; Burt and Barnard 2006). Sample sizes for catch by length and shell condition are summarized in Table 3. Relatively large samples were taken from the retained catch each year. Sample sizes for trawl bycatch were the annual sums of length frequency samples in the National Marine Fisheries Service (NMFS) database.

Catch Biomass

Retained catch and estimated bycatch biomasses are summarized in Table 1. Retained catch and estimated bycatch from the directed fishery include both the general open access fishery (i.e., harvest not allocated to Community Development Quota [CDQ] groups) and the CDQ fishery. Starting in 1973, the fishery generally occurred during the late summer and fall. Before 1973, a small portion of retained catch in some years was caught from April to June. Because most crab bycatch from the groundfish trawl fisheries occurred during the spring, the years in Table 1 are one year less than those from the NMFS trawl bycatch database to approximate the annual bycatch for reporting years defined as June 1 to May 31; e.g., year 2002 in Table 1 corresponds to what is reported for year 2003 in the NMFS database. Catch biomass is shown in Figure 2.

Catch Size Composition

Retained catch by length and shell condition and bycatch by length, shell condition, and sex were obtained for stock assessments. From 1960 to 1966, only retained catch length compositions from the Japanese fishery were available. Retained catches from the Russian and U.S. fisheries were assumed to have the same length compositions as the Japanese fishery during this period. From 1967 to 1969, the length compositions from the Russian fishery were assumed to be the same as those from the Japanese and U.S. fisheries. After 1969, foreign catch declined sharply and only length compositions from the U.S. fishery were used to distribute catch by length.

Catch per Unit Effort

Catch per unit effort (CPUE) is defined as the number of retained crabs per tan (a unit fishing effort for tanglenets) for the Japanese and Russian fisheries and the number of retained crabs per potlift for the U.S. fishery (Table 4). Soak time, while an important factor influencing CPUE, is difficult to standardize. Furthermore, complete historical soak time data from the U.S. fishery are not available. Based on the approach of Balsiger (1974), all fishing effort from Japan, Russia, and U.S. were standardized to the Japanese tanglenet from 1960 to 1971, and the CPUE was standardized as crabs per tan. The U.S. CPUE

data have similar trends as survey legal abundance after 1971 (Figure 3). Due to the difficulty in estimating commercial fishing catchability and the ready availability of NMFS annual trawl survey data, commercial CPUE data were not used in the model.

NMFS Survey Data

The NMFS has performed annual trawl surveys of the eastern Bering Sea since 1968. Two vessels, each towing an eastern otter trawl with an 83 ft headrope and a 112 ft footrope, conduct this multispecies, crab-groundfish survey during the summer. Stations are sampled in the center of a systematic 20 X 20 nm grid overlaid in an area of ≈140,000 nm^2. Since 1972 the trawl survey has covered the full stock distribution. The survey in Bristol Bay occurs primarily during late May and June. Tow-by-tow trawl survey data for Bristol Bay RKC during 1975-2008 were provided by NMFS.

Abundance estimates by sex, carapace length, and shell condition were derived from survey data using an area-swept approach without post-stratification (Figures 4 and 5). If multiple tows were made for a single station in a given year, the average of the abundances from all tows was used as the estimate of abundance for that station. Until the late 1980s, NMFS used a post-stratification approach, but subsequently treated Bristol Bay as a single stratum. If more than one tow was conducted in a station because of high RKC abundance (i.e., the station is a "hot spot"), NMFS regards the station as a separate stratum. Due to poor documentation, it is difficult to duplicate past NMFS post-stratifications. A "hot spot" was not surveyed with multiple tows during the early years. Two such "hot spots" affected the survey abundance estimates greatly: station H13 in 1984 (mostly juvenile crabs 75-90 mm CL) and station F06 in 1991 (mostly newshell legal males). The tow at station F06 was discarded in the NMFS abundance estimates (Stevens et al. 1991). In this study, all tow data were used. NMFS re-estimated historic areas-swept in 2008 and re-estimated area-swept abundance as well. We used area-swept abundances estimated by NMFS in April 2008 in this report.

In additional to standard surveys, NMFS also conducted some surveys after the standard surveys to assess mature female abundance. Two surveys were conducted for

Bristol Bay RKC in 1999, 2000, 2006-2008: the standard survey that was performed in late May and early June (about two weeks earlier than historic surveys) in 1999 and 2000 and the standard survey that was performed in early June in 2006-2008 and resurveys of 31 stations (1999), 23 stations (2000), 31 stations (2006, 1 bad tow and 30 valid tows), and 32 stations (2007 and 2008) with high female density that was performed in late July, about six weeks after the standard survey. The resurveys were necessary because a high proportion of mature females had not yet molted or mated prior to the standard surveys (Figure 6). Differences in area-swept estimates of abundance between the standard surveys and resurveys of these same stations are attributed to survey measurement errors or to seasonal changes in distribution between survey and resurvey. More large females were observed in the resurveys than during the standard surveys in 1999 and 2000 because most mature females had not molted prior to the standard surveys. As in 2006, area-swept estimates of males >89 mm CL, mature males, and legal males within the 32 resurvey stations in 2007 were not significantly different between the standard survey and resurvey (P=0.74, 0.74 and 0.95) based on t-tests of paired two sample for means. However, similar to 2006, area-swept estimates of mature females within the 32 resurvey stations in 2007 are significantly different between the standard survey and resurvey (P=0.03) based on the t-test. All survey tows were included in NMFS estimates in 1999, 2000, 2006-2008. To maximize use of the survey data, we used data from both surveys to assess male abundance but only the resurvey data, plus the standard survey data outside the resurveyed stations, to assess female abundance during these five years.

For 1968-1970 and 1972-1974, abundance estimates were obtained from NMFS directly because the original survey data by tow were not available. There were spring and fall surveys in 1968 and 1969. The average of estimated abundances from spring and fall surveys was used for those two years. Different catchabilities were assumed for survey data before 1973 because of an apparent change in survey catchability. A footrope chain was added to the trawl gear starting in 1973, and the crab abundances in all length classes during 1973-1979 were much greater than those estimated prior to 1973 (Reeves et al. 1977).

Bering Sea Fisheries Research Foundation Survey Data

The BSFRF conducted trawl surveys for Bristol Bay red king crab in 2007 and 2008 with a small mesh-size net and 5-minute tows. The surveys occurred at a similar time with the NMFS standard surveys and cover about 97% of the Bristol Bay area. Few Bristol Bay red king crab are outside of the BSFRF survey area. Because of small mesh size, the BSFRF surveys are expected to catch nearly all red king crabs within the area-swept. Crab abundances of different size groups were estimated by the Kriging method.

ANALYTIC APPROACH

To reduce annual measurement errors associated with abundance estimates derived from the area-swept method, the ADF&G developed a length-based analysis (LBA) in 1994 that incorporates multiple years of data and multiple data sources in the estimation procedure (Zheng et al. 1995a). Annual abundance estimates of the Bristol Bay RKC stock from the LBA have been used to manage the directed crab fishery and to set crab bycatch limits in the groundfish fisheries since 1995 (Figure 1). An alternative LBA (research model) was developed in 2004 to include small size groups for federal overfishing limits. The crab abundance declined sharply during the early 1980s. The LBA estimated natural mortality for different periods of years, whereas the research model estimated additional mortality beyond a basic constant natural mortality during 1976-1993. In this report, we present only the research model that was fit to the data only from 1968 to 2008.

Model Scenarios

Three scenarios were examined in this report: (1) constant natural mortality (0.18), estimation of additional mortality for males during 1980-1984 and for females during 1976-1993, and with the Bering Sea Fisheries Research Foundation (BSFRF) survey data; (2) constant natural mortality (0.18) with BSFRF survey data; and (3) a basic constant natural mortality (0.18), estimation of additional mortality for males during 1980-1984 and for females during 1976-1993, but without the BSFRF survey data. Scenarios (1) and (3)

were compared to illustrate the value of the BSFRF surveys. As we will see in the Results section, scenario (2) fits the data very poorly, and scenario (1) fits the data well and its results were used to estimate biological reference points and federal overfishing levels for 2008. In all figures and tables, the results are for scenario (1) without any indication. If a figure or table contains the results from scenario (2) or (3), the caption will indicate it.

Different handling mortalities were run for the sensitivity study. The base handling mortality rate was assumed to be 0.2 for the directed pot fishery, and alternative handling mortality rates of 0.1 and 0.4 were used for comparison. Like the scenario results, the results in all tables or figures are for handling mortality of 0.2 unless indicated otherwise in the caption.

Main Assumptions for the Model

Many assumptions were made to develop the length-based model. The major assumptions are:

(1) The basic natural mortality is constant over shell condition and length and was estimated with a maximum age of 25 and the 1% rule (Zheng 2005).

(2) Survey and fisheries selectivities are a function of length and were constant over shell condition. Selectivities are a function of sex except for trawl bycatch selectivities, which are the same for both sexes. Four different survey selectivities were estimated: (1) 1968-69 (surveys at different times), (2) 1970-72 (surveys without a footrope chain), (3) 1973-1981, and (4) 1982-2008 (modifying approaches to surveys).

(3) Growth is a function of length and did not change over time for males. For females, three growth increments per molt as a function of length were estimated based on sizes at maturity (1968-1982, 1983-1993, and 1994-2008). Once mature, female red king crabs grow with a much small growth increment per molt.

(4) Molting probabilities are an inverse logistic function of length for males. Females molt annually.

(5) Annual fishing seasons for the directed fishery are short.

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(6) Survey catchability (Q) was estimated using a Bayesian approach based on a trawl experiment by Weinberg et al. (2004) and the BSFRF surveys. Q was assumed to be constant over time except during 1970-1972.

(7) Males mature at sizes ≥120 mm CL. For convenience, female abundance was summarized at sizes ≥90 mm CL as an index of mature females.

(8) For summer trawl survey data, shell ages of newshell crabs were 12 months or less, and shell ages of oldshell and very oldshell crabs were more than 12 months.

(9) Measurement errors were assumed to be normally distributed for length compositions and were log-normally distributed for biomasses.

Population Model

The original LBA model was described in detail by Zheng et al. (1995a, 1995b) and Zheng and Kruse (2002a). Pulse fishing was assumed for the model. Male crab abundances by carapace length and shell condition in any one year are modeled to result from abundances in the previous year minus catch and handling and natural mortalities, plus recruitment and additions to or losses from each length class due to growth:

$$N_{l+1,t+1} = \sum_{l'=1}^{l'=l+1} \{P_{l',l+1}[(N_{l',t}+O_{l',t})\,e^{-M_t} - (C_{l',t}+D_{l,t})\,e^{(y_t-1)M_t} - T_{l,t}e^{(j_t-1)M_t}]\,m_{l'}\} + R_{l+1,t+1},$$

$$O_{l+1,t+1} = [(N_{l+1,t}+O_{l+1,t})\,e^{-M_t} - (C_{l+1,t}+D_{l+1,t})\,e^{(y_t-1)M_t} - T_{l+1,t}e^{(j_t-1)M_t}](1-m_l),$$

(1)

where

$N_{l,t}$ is newshell crab abundance in length class l and year t,

$O_{l,t}$ is oldshell crab abundances in length class l and year t,

M is the instantaneous natural mortality,

m_l is the molting probability for length class l,

$R_{l,t}$ is recruitment into length class l in year t,

y_t is the lag in years between the assessment survey and the mid fishery time in year t,

j_t is the lag in years between the assessment survey and the mid Tanner

crab fishery time in year t,

$P_{l',l}$ is the proportion of molting crabs growing from length class l' to l after one molt,

$C_{l,t}$ is the retained catch of length class l in year t, and

$D_{l,t}$ is the discarded mortality catch of length class l in year t, including directed pot and trawl bycatch,

$T_{l,t}$ is the discarded mortality catch of length class l in year t from the Tanner crab fishery.

The minimum carapace length for males is set at 65 mm, and crab abundance is modeled with a length-class interval of 5 mm. The last length class includes all crabs $\geq$160-mm CL. There are 20 length classes/groups. $P_{l',l}$, m_l, $R_{l,t}$, $C_{l,t}$, and $D_{l,t}$ are computed as follows:

Mean growth increment per molt is assumed to be a linear function of pre-molt length:

$$G_l = a + b\,\iota, \tag{2}$$

where a and b are constants. Growth increment per molt is assumed to follow a gamma distribution:

$$g(x/\alpha_l, \beta) = x^{\alpha_l - 1}\, e^{-x/\beta} / [\beta^{\alpha_l}\, \Gamma(\alpha_l)]. \tag{3}$$

The expected proportion of molting individuals growing from length class l_1 to length class l_2 after one molt is equal to the sum of probabilities within length range $[\iota_1,\ \iota_2)$ of the receiving length class l_2 at the beginning of the next year:

$$P_{l_1,l_2} = \int_{\iota_1 - \iota}^{\iota_2 - \iota} g(x/\alpha_l, \beta)dx, \tag{4}$$

where ι is the mid-length of length class l_1. For the last length class L, $P_{L,L} = 1$.

The molting probability for a given length class l is modeled by an inverse logistic function:

$$m_l = 1 - \frac{1}{1 + e^{-\beta\ (l - L_{50})}}, \tag{5}$$

where

β, L_{50} are parameters, and

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ι is the mid-length of length class *l*.

Recruitment is defined as recruitment to the model and survey gear rather than recruitment to the fishery. Recruitment is separated into a time-dependent variable, R_t, and size-dependent variables, U_l, representing the proportion of recruits belonging to each length class. R_t was assumed to consist of crabs at the recruiting age with different lengths and thus represents year class strength for year *t*. $R_{l,t}$ is computed as

$$R_{l,t} = R_t U_l, \tag{6}$$

where U_l is described by a gamma distribution similar to equations (3) and (4) with a set of parameters α_r and β_r. Because of different growth rates, recruitment was estimated separately for males and females under a constraint of approximately equal sex ratios of recruitment over time.

Before 1990, no observed bycatch data were available in the directed pot fishery; the crabs that were discarded and died in those years were estimated as the product of handling mortality rate, legal harvest rates, and mean length-specific selectivities. It is difficult to estimate bycatches from the Tanner crab fishery before 1991. A reasonable index to estimate bycatch fishing mortalities is potlifts of the Tanner crab fishery within the distribution area of Bristol Bay red king crab. Thus, bycatch fishing mortalities from the Tanner crab fishery before 1991 were estimated to be proportional to the smoothing average of potlifts east of 163° W. The smoothing average is equal to $(P_{t-2}+2P_{t-1}+3P_t)/6$ for the potlift in year t. The smoothing process not smoothes the annual number of potlifts, it also indexes the effects of lost pots during the previous years. For bycatch, all fishery catch and discard mortality bycatch are estimated as:

$$C_{l,t} \ or \ D_{l,t} = (N_{l,t} + O_{l,t}) \, e^{-y_t M_t} (1 - e^{-s_l F_t}) \tag{7}$$

where

s_l is selectivity for retained, pot or trawl discarded mortality catch of length class *l*, and

F_t is full fishing mortality of retained, pot or trawl discarded mortality catch in year *t*.

For discarded mortality bycatch from the Tanner crab fishery, y_t is replaced by j_t in the right side of equation (7).

The female crab model is the same as the male crab model except that the retained catch equals zero and molting probability equals 1.0 to reflect annual molting (Powell 1967). The minimum carapace length for females is set at 65 mm, and the last length class includes all crabs ≥140-mm CL, resulting in length groups 1-16.

Fisheries Selectivities

Retained selectivity, female pot bycatch selectivity, and both male and female trawl bycatch selectivity are estimated as a function of length:

$$s_l = \frac{1}{1 + e^{-\beta \ (l - L_{50})}}, \tag{8}$$

Different sets of parameters (β, L_{50}) are estimated for retained males, female pot bycatch, male and female trawl bycatch, and discarded males and females from the Tanner crab fishery. Because some catches were from the foreign fisheries during 1968-1972, a different set of parameters (β, L_{50}) are estimated for retained males for this period and a third parameter, sel_62.5mm, is used to explain the high proportion of catches in the last length group.

Male pot bycatch selectivity is modeled by two linear functions:

$$
\begin{aligned}
s_l &= \varphi + \kappa\, l, \quad if \ l < 135 \, \text{mm CL}, \\
s_l &= s_{l-1} + 5\gamma, \quad if \ l > 134 \, \text{mm CL}
\end{aligned} \tag{9}
$$

Where

φ, κ, γ are parameters.

During 2005-2008, a portion of legal males were also discarded in the pot fishery. The selectivity for this highgrading was estimated to be the retained selectivity in each year times a highgrading parameter, hg_t.

Trawl Survey Selectivities/Catchability

Trawl survey selectivities/catchability are estimated as

$$s_l = \frac{Q}{1 + e^{-\beta \ (l - L_{50})}}, \tag{10}$$

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with different sets of parameters (β, L_{50}) estimated for males and females as well as four different periods (1968-69, 1970-72, 1973-81 and 1982-08). Survey selectivity for the first length group (67.5 mm) was assumed to be the same for both males and females, so only three parameters (β, L_{50} for females and L_{50} for males) were estimated in the model for each of the four periods.   Parameter Q was called the survey catchability that was estimated using a Bayesian approach based on a trawl experiment by Weinberg et al. (2004, Figure 7) and the BSFRF surveys (Figure 7). Q was assumed to be constant over time except during 1970-1972 when the survey catchability was small.

Assuming that the BSFRF survey caught all crabs within the area-swept, the ratio between NMFS abundance and BSFRF abundance is a capture probability for the NMFS survey net. The Delta method was used to estimate the variance for the capture probability.   A maximum likelihood method was used to estimate parameters for a logistic function as an estimated capture probability curve (Figure 7). For a given size, the estimated capture probability is smaller based on the BSFRF survey than from the trawl experiment. Because many small-sized crabs are in the shallow water areas that are not accessible for the trawl survey, NMFS survey selectivity consists of capture probability and crab availability.

Parameters Estimated Independently

Basic natural mortality, length-weight relationships, and mean growth increments per molt were estimated independently outside of the model. Mean length of recruits to the model depends on growth and was assumed to be 72.5 for both males and females. Highgrading parameters hg_t were estimated to be 0.2785 in 2005, 0.0440 in 2006, 0.0197 in 2007, and 0.0198 in 2008 based on the proportions of discarded legal males to total caught legal males. Handling mortality rates was set to 0.2 for the directed pot fishery, 0.25 for the Tanner crab fishery, and 0.8 for the trawl fisheries.

Natural Mortality

Based on an assumed maximum age of 25 years and the 1% rule (Zheng 2005), basic M was estimated to be 0.18 for both males and females. Natural mortality in a given year, M_t, equals to $M + Mm_t$ (for males) or $M + Mf_t$ (females). One value of Mm_t during 1980-1985 was estimated and two values of Mf_t during 1980-1984 and 1976-79, 1985-93 were estimated in the model.

Length-weight Relationship

Length-weight relationships for males and females were as follows:

Immature Females: $W = 0.010271 \, L^{2.388}$,

Ovigerous Females: $W = 0.02286 \, L^{2.234}$, (11)

Males: $W = 0.000361 \, L^{3.16}$,

where

W is weight in grams, and

L is CL in mm.

Growth Increment per Molt

A variety of data are available to estimate male mean growth increment per molt for Bristol Bay RKC. Tagging studies were conducted during the 1950s, 1960s and 1990s, and mean growth increment per molt data from these tagging studies in the 1950s and 1960s were analyzed by Weber and Miyahara (1962) and Balsiger (1974).

Modal analyses were conducted for the data during 1957-1961 and the 1990s (Weber 1967; Loher et al. 2001). Mean growth increment per molt may be a function of body size and shell condition and vary over time (Balsiger 1974; McCaughran and Powell 1977); however, for simplicity, mean growth increment per molt was assumed to be only a function of body size in the models. Tagging data were used to estimate mean growth increment per molt as a function of pre-molt length for males (Figure 8). The results from modal analyses of 1957-1961 and the 1990s were used to estimate mean growth increment per molt for immature females during 1968-1993 and 1994-2008, respectively, and the data presented in Gray (1963) were used to estimate those for mature females (Figure 8). To make a smooth transition of growth increment per molt from immature to mature females, weighted growth increment averages of 70% and 30% at 92.5 mm CL pre-molt length and 90% and 10% at 97.5 mm CL were used, respectively, for mature and immature females during 1983-1993. These percentages are roughly close to the composition of maturity. During 1968-1982, females matured at a smaller size, so the growth increment per molt as a function of length was shifted to smaller increments. Likewise, during 1994-2008, females matured at a slightly higher size, so the growth increment per molt was shifted to high increments for immature crabs (Figure 8). Once mature, the growth increment per molt for male crabs decreases slightly and annual molting probability decreases, whereas the growth increment for female crabs decreases dramatically but annual molting probability remains constant at 1.0 (Powell 1967).

Sizes at Maturity for Females

NMFS collected female reproductive condition data during the summer trawl surveys. Mature females are separated from immature females by a presence of egg clutches or egg cases. Proportions of mature females at 5-mm length intervals were summarized and a logistic curve was fit to the data each year to estimate sizes at 50% maturity. Sizes at 50% maturity are illustrated in Figure 9 with mean values for three different periods (1975-82, 1983-93 and 1994-08).

Sizes at Maturity for Males

Sizes at functional maturity for Bristol Bay male RKC have been assumed to be 120 mm CL (Schmidt and Pengilly 1990). This is based on mating pair data collected off Kodiak Island (Figure 10). Sizes at maturity for Bristol Bay female RKC are about 90 mm CL, about 15 mm CL less than Kodiak female RKC (Pengilly et al. 2002). The size ratio of mature males to females is 1.3333 at sizes at maturity for Bristol Bay RKC, and since mature males grow at much larger increments than mature females, the mean size ratio of mature males to females is most likely larger than this ratio. Size ratios of the large majority of Kodiak mating pairs were less than 1.3333, and in some bays, only a small proportion of mating pairs had size ratios above 1.3333 (Figure 10).

In the laboratory, male RKC as small as 80 mm CL from Kodiak and SE Alaska can successfully mate with females (Paul and Paul 1990). But few males less than 100 mm CL were observed to mate with females in the wild. Based on the size ratios of males to females in the Kodiak mating pair data, setting 120 mm CL as a minimum size of functional maturity for Bristol Bay male RKC is proper and conservative in terms of managing the fishery.

Parameters Estimated Conditionally

The following model parameters were estimated for male and female crabs: total recruits for each year (year class strength R_t for t = 1969 to 2008), total abundance in the first year (1968), growth parameter β and recruitment parameter β_r for males and females separately. Molting probability parameters β and L_{50} were also estimated for male crabs. Estimated parameters also include β and L_{50} for retained selectivity, β and L_{50} for pot-discarded female selectivity, β and L_{50} for pot-discarded male and female selectivities from the eastern Bering Sea Tanner crab fishery, β and L_{50} for groundfish trawl discarded selectivity, φ, κ and γ for pot-discarded male selectivity, and β for trawl survey selectivity and L_{50} for trawl survey male and females separately. NMFS survey catchabilities Q for 1968-69 and 1973-2008 and Q_m (for males) and Q_f (for females) for 1970-72 were also estimated. Annual fishing mortalities were also estimated for the directed pot fishery for males (1968-2008), pot-discarded females from the directed

fishery (1990-2008), pot-discarded males and females from the eastern Bering Sea Tanner crab fishery (1991-93), and groundfish trawl discarded males and females (1976-2007). Three additional mortality parameters for Mm_t and Mf_t were also estimated. The total number of parameters to be estimated was 223. Some estimated parameters were constrained in the model. For example, male and female recruitment estimates were forced to be close to each other for a given year.

To increase the efficiency of the parameter-estimation algorithm, we assumed that the smoothed relative frequencies of length and shell classes from survey year 1968 approximate the true relative frequencies within sexes. Thus, only total abundances of males and females for the first year were estimated; 3n unknown parameters for the abundances in the first year, where n is the number of length-classes, were reduced to one under this assumption.

A maximum likelihood approach was used to estimate parameters. For length compositions ($p_{l,t,s,sh}$), the likelihood functions are :

$$Rf = \prod_{l=1}^{L}\prod_{t=1}^{T}\prod_{s=1}^{2}\prod_{sh=1}^{2}\frac{\left\{\exp\left[-\frac{(p_{l,t,s,sh}-\hat{p}_{l,t,s,sh})^2}{2\sigma^2}\right]+0.01\right\}}{\sqrt{2\pi\sigma^2}}, \quad (12)$$

$$\sigma^2 = \left[\hat{p}_{l,t,s,sh}(1-\hat{p}_{l,t,s,sh})+0.1/L\right]/n,$$

where

    L is the number of length groups,

    T is the number of years, and

    n is the effective sample size, which was assumed to be 400 for retained males,

        200 for trawl survey and 100 for bycatch length composition data.

The weighted negative log-likelihood functions are:

$$Length\ compositions: \quad -\sum \ln(Rf_i),$$

$$Biomasses\ other\ than\ survey: \quad \lambda_j \sum \left[\ln(B_t / \hat{B}_t)^2\right],$$

$$NMFS\ survey\ biomass: \quad \sum \left[\ln(B_t / \hat{B}_t)^2 / (2\ln(CV_t^2 + 1))\right],$$

$$BSFRF\ mature\ males: \quad \sum [\ln(N_t / \hat{N}_t)^2 / (2\ln(CV_t^2 + 1))],$$

$$NMFS\ survey\ Q: \quad (Q - Q_{\exp})^2 / (2\sigma_{\exp}^2) + (Q - Q_{bsfrf})^2 / (2\sigma_{bsfrf}^2),$$

$$R\ variation: \quad \lambda_R \sum [\ln(R_t / \overline{R})^2],$$

$$R\ sex\ ratio: \quad \lambda_s [\ln(\overline{R}_M / \overline{R}_F)^2],$$

(13)

Where

Q is NMFS survey catchability during 1968-69 and 1973-2008,

Q_{exp} is 0.896, estimated capture probability at 162.5 mm CL from the experiment,

Q_{bsfrf} is 0.854, estimated capture probability at 162.5 mm CL from the BSFRF surveys,

σ_{exp} is 0.030, estimated standard deviation of Q_{exp},

σ_{bsfrf} is 0.1072, estimated standard deviation of Q_{bsfrf},

R_t is the recruitment in year t,

$\overline{R}$ is the mean recruitment,

$\overline{R}_M$ is the mean male recruitment,

$\overline{R}_F$ is the mean female recruitment.

Weighted λ_j are assumed to be 500 for retained catch biomass, and 100 for all bycatch biomasses, 2 for recruitment variation, and 20 for recruitment sex ratio. These λ_j values represent prior assumptions about the accuracy of the observed catch biomass data and about the variances of these random variables.

RESULTS

Population Abundance

The model (scenario 1) fit the fishery biomass data well and fit the survey biomass reasonably well (Figures 11 and 12). Because the model estimates annual fishing mortality for pot male catch, pot female bycatch, and trawl bycatch, the deviations of observed and predicted (estimated) fishery biomass are mainly due to size composition differences. The

model did not fit the mature crab abundance directly and depicted the trends of the mature abundance well (Figure 12). Estimated mature crabs abundance increased dramatically in the mid 1970s and decreased precipitously in the early 1980s. Estimated mature crab abundance has increased during the last 20 years with mature females being 6 times more abundant in 2008 than in 1985 and mature males being 2.9 times more abundant in 2008 than in 1985 (Figure 12).

The model also fit the length and shell composition data well (Figures 13-20). Model fit of length compositions in the trawl survey was better for newshell males and females than for oldshell males. The model predicted lower proportions of oldshell males in 1993, 1994, 2002 and 2007 and higher proportions of oldshell males in 1997, 2001, 2003, 2004 and 2006 than the area-swept estimates (Figure 14). In addition to size, molting probability may also be affected by age and environmental conditions. Tagging data show that molting probability changed over time (Basilger 1974). Therefore, the relatively poor fit to oldshell males may be due to use of a constant molting probability function as well as shell aging errors. It is surprising that the model fit the length proportions of the pot male bycatch well with two simple linear selectivity functions (Figure 17). We explored a logistic selectivity function, but due to the long left tail of the pot male bycatch selectivity, the logistic selectivity function did not fit the data well.

Modal progressions are tracked well in the trawl survey data, particularly beginning in the mid-1990s (Figures 13 and 15). Cohorts first seen in the trawl survey data in 1975, 1986, 1990, 1995, 1999, 2002 and 2005 can be tracked over time. Some cohorts can be tracked over time in the pot bycatch as well (Figure 17), but the bycatch data did not track the cohorts as well as the survey data. Groundfish trawl bycatch data provide little information to track modal progression (Figures 19 and 20).

Parameter Estimates

Negative log-likelihood values and parameter estimates are summarized in Tables 5 and 6, respectively. Length-specific fishing mortality is equal to its selectivity times the full fishing mortality. Estimated full pot fishing mortalities for females and full fishing mortalities for trawl bycatch were very low due to low bycatches as well as handling mortality rates

less than 1.0. Estimated recruits varied greatly from year to year (Table 6). Estimated low selectivities for male pot bycatch, relative to the retained catch, reflected the 20% handling mortality rate (Figure 21). Both selectivities were applied to the same level of full fishing mortality. Estimated selectivities for female pot bycatch were close to 1 for all mature females (Figure 21), and the estimated full fishing mortalities for female pot bycatch were much lower than for male retained catch and bycatch (Table 6).

One of the most important results is estimated trawl survey selectivity/catchability (Figure 21). Survey selectivity affects not only the fitting of the data but also the absolute abundance estimates. Estimated survey selectivities in Figure 21 are generally smaller than the capture probabilities in Figure 8 because survey selectivities include capture probabilities and crab availability. NMFS survey catchability was estimated to be 0.903, very close to that from the trawl experiment (0.896) and higher than that estimated from the BSFRF surveys (0.854). The reliability of estimated survey selectivities will greatly affect the application of the model to fisheries management. Under- or overestimates of survey selectivities will cause a systematical upward or downward bias of abundance estimates. Information about crab availability to the survey area at survey times will help estimate the survey selectivities.

Estimated molting probabilities during 1968-2008 (Figure 22) were generally lower than those estimated from the 1954-1961 and 1966-1969 tagging data (Balsiger 1974). Lower molting probabilities mean more oldshell crabs, possibly due to changes in molting probabilities over time or shell aging errors. Overestimates or underestimates of oldshell crabs will result in lower or higher estimates of male molting probabilities.

Residual Patterns

Residuals of total survey biomass and proportions of length and shell condition, calculated as observed minus predicted, were plotted to examine their patterns. Residuals of total survey biomass were standardized by the estimated standard deviation. The residuals of total survey biomass did not show any consistent patterns (Figure 23). Standardized residuals of proportions of survey newshell males appear to be random over length and year (Figure 24). Residuals of proportions of survey oldshell males were mostly

positive or negative for some years (Figure 25). This is expected since a constant molting probability function over time was used. Changes in molting probability over time or shell aging errors would create such residual patterns. There is an interesting pattern for residuals of proportions of survey females. Residuals were generally negative for large-sized mature females during 1969-1987 (Figure 26). Changes in growth over time or increased mortality may cause this pattern. The inadequateness of the model can be corrected by adding parameters to address these factors. However, the pattern appears to be minor, and the model appears to fit annual mature female abundance quite well (Figure 12).

Comparison of Scenarios

Scenario 1 fit the data very well, and scenario 2 did not fit the data and was difficult to converge (Figures 12 and 13). The differences of estimated abundance and biomass are huge between scenarios 1 and 2. The negative log likelihood values are -53908, -47292, and -53886 for scenarios 1, 2 and 3. The degrees of freedom are the same for scenarios 1 and 3, which are 3 less than scenario 2. Scenario 2 could not account for the big fall of abundance in the early 1980s.

Abundance and biomass estimates are very close between scenarios 1 and 3 except in the mid 1970s and recent years (Figures 12 and 13). Abundance estimates are slightly lower in the mid 1970s and higher during the recent years for scenario 3 than for scenario 1. As a comparison, estimated mature biomass and abundance are as follows for 2007 and 2008 (standard deviations are in parentheses):

	MMB215 (million lbs)		Mature male abundance (million)	
	2007	2008	2007	2008
Scenario 1:	77.64 (3.40)	90.25 (3.56)	22.23 (0.75)	25.93 (1.01)
Scenario 3:	99.74 (6.38)	116.43 (6.91)	27.56 (1.44)	33.34 (1.95)

Overall, BSFRF survey data increases the precision of abundance and biomass estimates. Without them, the model may have overestimated mature male abundance due to high mature female abundance because the objective function is total survey biomass as well as both length frequency data of males and females.

Retrospective Analyses

Two kinds of retrospective analyses were conducted for this report: (1) historical results and (2) the 2008 model results. The historical results are the trajectories of biomass and abundance from previous assessments that capture both new data and changes in methodology over time. Treating the 2009 estimates as the baseline values, we can also evaluate how well the model had done in the past. The 2008 model results are based on sequentially excluding one-year of data to evaluate the current model performance with less data.

Historical Results

The model first fit the data from 1985 to 2004 in 2004. Thus, five historical assessment results are available. The main differences of the 2004 model were weighting factors and effective sample sizes for the likelihood functions. In 2004, the weighting factors were 1000 for survey biomass, 2000 for retained catch biomass and 200 for bycatch biomasses. The effective sample sizes were set to be 200 for all proportion data but weighting factors of 5, 2, and 1 were also applied to retained catch proportions, survey proportions and bycatch proportions. Estimates of time series of abundance in 2004 were generally higher than those estimated after 2004 (Figure 27).

In 2005, to improve the fit for retained catch data, the weight for retained catch biomass increased to 3000 and the weight for retained catch proportions increased to 6. All other weights were not changed. In 2006, all weights were re-configured. No weights were used for proportion data, and instead, effective sample sizes were set to be 500 for retained catch, 200 for survey data, and 100 for bycatch data. Weights for biomasses were changed to 800 for retained catch, 300 for survey and 50 for bycatches. The weights in 2007 were the same as 2006. Generally, estimates of time series of abundance in 2005 were slightly lower than in 2006 and 2007, and there were few differences between estimates in 2006 and 2007 (Figure 27).

In 2008, estimated coefficients of variation for survey biomass were used to compute likelihood values as suggested by a Crab Plan Team member in 2007. Weights

were re-configured because of this change: 500 for retained catch biomass, 50 for survey biomass, and 20 for bycatch biomasses. Effective sample size was lowered to 400 for the retained catch data. These changes were necessary for the estimation to converge and for a relatively good balanced fit to both biomasses and proportion data. Also, sizes at 50% selectivities for all fisheries data were allowed to change annually, subject to a random walk pattern, for all assessments before 2008. The 2008 model does not allow annual changes in any fishery selectivities. Except for higher estimates of abundance during the late 1980s and early 1990s, estimates of time series of abundance in 2008 were generally close to those in 2006 and 2007 (Figure 27).

In 2009, the model was extended to the data through 1968. In addition, the BSFRF survey data were used to estimate the NMFS survey catchability. No weight factors were used for the NMFS survey biomass in 2009.

2008 Model Results

The performance of the 2008 model includes sequentially excluding one-year of data. The model performed well except estimates for the early 2000s made with terminal year 2002 (Figure 28). Lower estimates in the early 2000s were primarily due to extremely low survey estimates in 2001.

Overall, both historical results and the 2008 model results performed reasonably well. No great overestimates or underestimates occurred as observed in Pacific halibut (*Hippoglossus stenolepis*) (Parma 1993) or some eastern Bering Sea groundfish stocks (Zheng and Kruse 2002a; Ianelli et al. 2003). Since the most recent model has not been used to set TAC or overfishing limits, historical implications for management from these assessment errors can not be evaluated at the current time. However, management implications of the ADF&G stock assessment model were evaluated by Zheng and Kruse (2002a).

Effects of Handling Mortality Rate on Abundance Estimates

The baseline handling mortality rate for the directed pot fishery was set at 0.2. A 50% reduction and 100% increase resulted in 0.1 and 0.4 as alternatives. Overall, a higher

handling mortality rate resulted in slightly higher estimates of mature abundance, and a lower rate resulted in a minor reduction of estimated mature abundance (Figure 29). Differences of estimated legal abundance and mature male biomass were small among these handling mortality rates (Figure 30).

Potential Reasons for High Mortality during the early 1980s

Bristol Bay red king crab abundance had declined sharply during the early 1980s. Many factors have been speculated for this decline: (i) completely wiped out by fishing: directed pot fishery, other directed pot fishery (Tanner crab fishery), and botttom trawling; and (ii) high fishing and natural mortality. With the survey abundance, harvest rates in 1980 and 1981 were among the highest, thus the directed fishing definitely had a big impact on the stock decline, especially legal and mature males. However, for the sharp decline during 1980-1884 for males, 3 out of 5 years had low mature harvest rates. During 1981-1984 for females, 3 out of 4 years had low mature harvest rates. Also pot catchability for females and immature males are generally much lower than for legal males, so the directed pot fishing alone cannot explain the sharp decline for all segments of the stock during the early 1980s.

Red king crab bycatch in the eastern Bering Sea Tanner crab fishery is another potential factor. The main overlap between Tanner crab and Bristol Bay red king crab is east of 163° W. No absolute red king crab bycatch estimates are available until 1991. So there are insufficient data to fully evaluate the impact. Retained catch and potlifts from the eastern Bering Sea Tanner crab fishery are illustrated in Figure 31. The observed red king crab bycatches in the Tanner crab fishery during 1991-1993 and total potlifts east of 163° W during 1968 to 2005 were used estimate the bycatch mortality in the current model. Because winter sea surface temperatures and air temperatures were warmer (which means a lower handling mortality rate) and there were fewer potlifts during the early 1980s than during the early 1990s, bycatch in the Tanner crab fishery is unlikely to have been a main factor for the sharp decline of Bristol Bay red king crab.

Dew and Mcconnaughey (2005) speculated that bottom trawling in southern Bristol Bay wiped out the mature red king crab stock. The main data to support this

speculation are illustrated in Figure 32. The observed red king crab bycatch was very small relative to red king crab abundance, so the focus here is the unobserved bycatch, such as incomplete bycatch reporting. However, there are major flaws in the data analysis by Dew and Mcconnaughey (2005). The relationship established for trawling and mature female crab density was from log-transformed data. When plotting the data in a normal scale, the crab density declined more than 80% before any meaningful trawling occurred in the main crab habitat (Figure 32). From the spatial distributions of female red king crab and numbers of bottom trawling tows over time (See Appendix), we also see that there is a space and time mismatch of crab distribution and bottom trawling. The crab distribution started to shift in 1977, and a large majority of mature females were far away from the Unimak and Amak area since 1978, before the trawling was allowed into the area in 1982 (Witherell and Pautzke 1997). Therefore, the bottom trawling was unlikely to have had a great impact on the stock decline in southern Bristol Bay in the early 1980s. However, heavy trawling in the southern Bristol Bay may have habitat damage, which may affect the potential for crab stock rebuilding. Furthermore, trawling and the red king crab broodstock overlapped in the mid 1980s in the central Bristol Bay area, which may have impacted crab stock rebuilding as well.

Several factors may have caused increases in natural mortality. Crab diseases in the early 1980s were documented by Sparks and Morado (1985), but inadequate data were collected to examine their effects on the stock. Stevens (1990) speculated that senescence may be a factor because many crabs in the early 1980s were very old due to low temperatures in the 1960s and early 1970s. The biomass of the main crab predator, Pacific cod, increased about 10 times during the late 1970s and early 1980s. Yellowfin sole biomass also increased substantially during this period. Predation is primarily on juvenile and molting/softshell crabs. But we lack stomach samples in shallow waters (juvenile habitat) and during the period when red king crabs molt. Also cannibalism occurs during molting periods for red king crabs. High crab abundance in the late 1970s and early 1980s may have increased the occurrence of for cannibalism.

Overall, the likely causes for the sharp decline in the early 1980s are combinations of the above factors, such as pot fisheries on legal males, bycatch and predation on

females and juvenile and sublegal males, senescence on older crabs, and disease on all crabs. In our model, we estimated one mortality parameter for males and another for females during 1980-1984. We also estimated a mortality parameter for females during 1976-1979 and 1985-1993. These three mortality parameters are additional to the basic natural mortality of 0.18, all directed fishing mortality and non-directed fishing mortality. These three mortality parameters could be attributed to natural mortality as well as undocumented non-directed fishing mortality. The model fit the data much better with these three parameters (scenario 1) than without them (scenario 2).

Exploitation

Estimated full pot fishing mortalities ranged from 0.0 to 1.370 during 1968-2008 with estimated values over 0.4 during 1968-1981, 1986-1987, 1990, and 1993 (Table 6, Figure 34). Estimated fishing mortalities for pot female bycatch and trawl bycatch were less than 0.06.

The average of estimated male recruits from 1995 to 2008 (Figure 33) and mature male biomass per recruit was used to estimate $B_{35\%}$. Alternative periods of 1968-2008 and 1985-2008 were compared. The choice of this recruitment will be discussed in the "Biological Reference Points" section. The full fishing mortalities for the directed pot fishery at the time of fishing were plotted against mature male biomass on Feb. 15 (Figure 34). Before the current harvest strategy was adopted in 1996, many fishing mortalities were above $F_{35\%}$ (Figure 34). Under the current harvest strategy, estimated fishing mortalities were at or above the $F_{35\%}$ limits in 1998, 2005 and 2007 but below the $F_{35\%}$ limits in other years.

Stock-Recruitment Relationships

Estimated mature male biomass and recruitment were plotted to illustrate their relationships (Figure 35).

Egg clutch data collected during summer surveys may provide information about mature female reproductive conditions. Although egg clutch data are subject to rating errors as well as sampling errors, data trends over time may be useful. Proportions of

empty clutches for newshell mature females >89 mm CL were high in some years before 1990, but have been low since 1990 (Figure 36). The highest proportion of empty clutches (0.2) was in 1986, and they primarily involved soft shell females (shell condition 1). Clutch fullness fluctuated annually around average levels during two periods: before 1991 and after 1990 (Figure 36). The average clutch fullness was almost identical for these two periods (Figure 36).

BIOLOGICAL REFERENCE POINTS AND OVERFISHING LIMITS FOR 2008 and 2009

Bristol Bay RKC is currently placed in Tier 3 (NPFMC 2007). For Tier 3 stocks, estimated biological reference points include $B_{35\%}$, $F_{35\%}$ and $F_{40\%}$. Estimated model parameters were used to conduct mature male biomass per recruit analysis. Because trawl bycatch fishing mortality was not related to pot fishing mortality, average trawl bycatch fishing mortality during 1998 to 2007 was used for the per recruit analysis as well as for projections in the next section. Pot female bycatch fishing mortality was set equal to pot male fishing mortality times 0.02, an intermediate level during 1990-2008. Some discards of legal males occurred since the IFQ fishery started in 2005, but the discard rates were much lower during 2006-2008 than in 2005 after the fishing industry minimized discards of legal males. Thus, the average of retained selectivities and discard male selectivities during 2006-2008 were used to represent current trends for per recruit analysis and projections.

Average recruitments during three periods were used to estimate $B_{35\%}$: 1968-2008, 1985-2008, and 1995-2008 (Figure 33). Estimated $B_{35\%}$ are compared with historical mature male biomass in Figure 34. We recommend using the average recruitment during 1995-2008, which was used in 2008 to set the overfishing limits. There are several reasons for supporting our recommendation. First, estimated recruitment was higher after 1994 than during 1985-1994 and there was a potential regime shift after 1989 (Overland et al. 1999), which corresponded to recruitment in 1995 and later. Second, recruitments estimated before 1985 came from a potentially higher natural mortality than we used to estimate $B_{35\%}$. Third, high recruitments during the late 1960s and 1970s generally occurred when the spawning stock was primarily located in the southern Bristol Bay while the current

spawning stock is mainly in the middle of Bristol Bay. The current flows favor larva hatched in the southern Bristol Bay (see the section on Ecosystem Considerations). The spawning stock has not moved back to southern Bristol Bay during the last 20 years. Finally, the $B_{35\%}$ based on average recruitment during 1969-2008 is higher than mature male biomasses in 34 years out of 40 years and is close to 200% of the mean mature male biomass during 1968-2008. Even without fishing, the recruitment levels during the last 20 years could hardly produce a mature male biomass as high as it. Therefore, with this $B_{35\%}$ we have to close the fishery and wait for some things to happen to rebuild the stock.

Based on the $B_{35\%}$ estimated from the average male recruitment during 1995-2008, the biological reference points were estimated as follows:

$B_{35\%}$ = 70.627 million lbs, or 32,036 t,

$F_{35\%}$ = 0.33,

$F_{40\%}$ = 0.26.

Based on $B_{35\%}$ and $F_{35\%}$, the retained catch and total catch limits for 2009 are estimated to be:

Retained catch: XXX million lbs, or XXX t,

Total catch: XXX million lbs, or XXX t,

MMB on 2/15/2010: XXX million lbs, or XXX t.

Likelihood profiles of mature male biomass on February 15, retained catch and total catch for 2009 are illustrated in Figure 37. The confidence intervals are quite narrow for all three values.

PROJECTIONS AND FUTURE OUTLOOK

Projections

Future population projections primarily depend on future recruitment, but crab recruitment is difficult to predict. Therefore, annual recruitment for the projections was a random selection from estimated recruitments during 1995-2008. Besides recruitment, the other major uncertainty for the projections is estimated abundance in 2008. The 2008 abundance was randomly selected from the estimated normal distribution of the

assessment model for each replicate. Four scenarios of fishing mortality for the directed pot fishery were used in the projections:

(1) No directed fishery. This was used as a base projection.

(2) $F_{40\%}$. This fishing mortality creates a buffer between the limits and target levels.

(3) $F_{35\%}$. This is the maximum fishing mortality allowed under the current overfishing definitions.

(4) Current ADF&G harvest strategy with the $F_{35\%}$ constraint.

Each scenario was replicated 1000 times and projections made over 10 years beginning in 2008 (Table 8).

As expected, projected mature male biomasses were much higher without the directed fishing mortality than under the other scenarios. Among three scenarios with directed fishing, the ADF&G harvest strategy produced the most stable mature male biomass and catch over time (Table 8, Figures 38 and 39). With its forward looking feature, the ADF&G harvest strategy reduced fishing mortality one year or two years earlier than the $F_{40\%}$ and $F_{35\%}$ scenarios when recruitment was poor. At the end of 10 years, projected mature male biomass was above $B_{35\%}$ for the $F_{40\%}$ scenario and the ADF&G harvest strategy and similar to $B_{35\%}$ for the $F_{35\%}$ scenario (Figure 38).

Near Future Outlook

The near future outlook for the Bristol Bay RKC stock is stable. The three recent above-average year classes (hatching years 1990, 1994, and 1997) had largely entered the legal population by 2006 (Figure 40). Most individuals from the 1997 year class will continue to gain weight to offset loss of the legal biomass to fishing and natural mortalities. The above-average year class (hatching year 2000) with lengths centered around 87.5 mm CL for both males and females in 2006 and with lengths centered around 112.5-117.5 mm CL for males and around 107.5 mm CL for females in 2008 will continue to enter the mature male population next year and start to recruit to the legal population next year (Figure 40). However, no strong cohorts have been observed in the survey data after this cohort (Figure 40). Due to these above average year classes, mature and legal crabs should remain at relatively high abundance levels compared to the previous 20 years if

natural mortality does not increase greatly, as happened in the early 1980s for this stock and in 1999 for St. Matthew Island blue king crab (Zheng and Kruse 2002b). The mature and legal abundance may start to decline after next year. Current crab abundance is still low relative to the late 1970s, and without favorable environmental conditions, recovery to the high levels of the late 1970s is unlikely.

ECOSYSTEM CONSIDERATIONS

Three aspects of ecosystem considerations are reported in this report: impacts of changes in oceanographic conditions on RKC recruitment strength, predation by groundfish, and impacts of shifts of spatial distribution on crab recruitment success.

Impacts of Changes in Oceanographic Conditions on RKC Recruitment

Environmental factors may play important roles in determining recruitment strength. Climate variability, ocean temperature, surface winds, ocean currents and their ecological interactions may affect food availability and larval transport, growth and survival, thus affecting recruitment strength (Shepherd et al. 1984; Koslow et al. 1987). Changes in many of these oceanographic processes are associated with atmospheric pressure patterns in winter, such as the strength and position of the Aleutian Low Pressure System, which affects the direction and intensity of storms, and the Arctic Oscillation, which represents the spin up (or spin down) of the polar vortex and indexes the transfer of mass between high and mid latitudes (Overland et al., 1999). For instance, a climate regime shift in the late 1970s was manifested by increased winter storms and precipitation, faster alongshore currents, warmer sea surface temperatures, and higher coastal sea levels in the northeastern Pacific Ocean (Hollowed and Wooster 1992; Hare and Mantua 2000). Overland et al. (1999) found three shifts of wintertime climate forcing patterns that have been identified in the past three decades: 1967-1976 (positive Aleutian Low, mixed Arctic Oscillation), 1977-1988 (negative Aleutian Low, negative Arctic Oscillation), and 1989-1998 (mixed Aleutian Low, positive Arctic Oscillation).

The relationship between the recruitment strength of Bristol Bay RKC and the Aleutian Low Pressure index were examined by Zheng and Kruse (2000, 2006). They

found that the recruitment trends of Bristol Bay RKC may partly relate to decadal shifts in physical oceanography: all strong year classes occurred before 1977 when the Aleutian Low was weak. One of the largest year classes during the last 20 years, the 1990 year class, was also coincidental with the weak Aleutian Low index during 1989-1991 (Zheng and Kruse 2000, 2006). The mechanisms are uncertain, but food availability is hypothesized to be important to RKC (Zheng and Kruse 2000) because their larvae suffer reduced survival and feeding capability if they do not feed within the first 2-6 days after hatching (Paul and Paul 1980). Diatoms such as *Thalassiosira* are important food for first-feeding RKC larvae (Paul et al. 1989) and they are predominate in the spring bloom in years of light winds when the water column is stable (Ziemann et al. 1991; Bienfang and Ziemann 1995). One hypothesis is that years of strong wind mixing associated with intensified Aleutian Lows may depress RKC larval survival and subsequent recruitment (Zheng and Kruse 2000).

Predation by Groundfish

During the period from mating to recruitment, many events can modify crab year-class strength. This may explain the weak relationships between recruitment and spawning biomass as well as individual environmental factors. One such event is groundfish predation. Groundfish consume crabs from the pelagic larval to adult stages. Based on routine examination of stomach contents of some groundfish species (Alaska plaice, arrowtooth flounder, flathead sole, northern rock sole, Pacific cod, Pacific halibut, skates, walleye pollock, and yellowfin sole) in the eastern Bering Sea, a huge amount of early juvenile Tanner and snow crabs are consumed by groundfish each year during summer months, May to September (Lang et al. 2003). Predation on large crabs usually occurs during molting periods (Blau 1986), which are generally during spring. Few large crabs have been founded in groundfish stomachs during summer months when sampling occurs. Because female RKC molt later than males, sampling may bias against monitoring of predation on adult male RKC relative to females (Table 9). Likewise, juvenile RKC are usually found in nearshore, shallow waters, where hardly any samples of groundfish are taken. Thus, data are not available

to estimate groundfish predation on juvenile RKC. Overall, estimates of RKC biomass to be consumed by groundfish during summer months were low relative to the crab population abundance (Table 10).

Zheng and Kruse (2006) reported statistically significant correlations between Pacific cod biomass and Bristol Bay RKC recruitment with recruitment time lags from ages 0 to 3. Correlations between yellowfin sole biomass and log-transformed Bristol Bay RKC recruitment are also statistically significant with recruitment time lags from ages 0 to 2 (r = -0.85, -0.83, -0.79, and P = 0.03, 0.04, 0.04, respectively, Zheng and Kruse 2006). The spatial distribution of yellowfin sole mainly overlaps with Bristol Bay RKC and has not changed much over time. Higher Pacific cod and yellowfin sole biomass was associated with lower RKC recruitment (Zheng and Kruse 2006). Pacific cod is the main predator of red king crabs (Table 10).

Statistical significance does not necessarily imply biologically meaningful relationships. Multiple statistical tests increase the probability of Type I error. In a detailed study of predation and population trends, Livingston (1989) concluded that cod predation was not responsible for declines of RKC in Bristol Bay in the early 1980s. Estimates of RKC consumed by cod during 1981 and 1983-1996 (Livingston 1991; Livingston et al. 1993, Livingston & deReynier 1996; Lang et al. 2003) constitute only a very small proportion of the crab population. Most RKC in cod stomachs are softshell females >80 mm carapace length (Livingston 1989; Table 9) – well beyond the size at which year class strength is determined. However, as noted earlier, the lack of RKC in groundfish stomachs may also be due to sampling problems. Therefore, the lack of large numbers of early juvenile RKC in groundfish stomach data obtained during summer months in offshore waters does not necessarily invalidate the apparent negative relationships between RKC year-class strength and biomass of Pacific cod and yellowfin sole. Groundfish stomachs must be sampled at the appropriate spatial and temporal scales to resolve questions about groundfish predation on juvenile king crabs.

Spatial distributions of crabs and groundfish may also play an important role on groundfish predation on crabs. Like crab stocks, spatial distributions of groundfish stocks in the eastern Bering Sea changed over time (Figure 41). During recent years,

biomass distribution centers of Pacific cod, flathead sole and arrowtooth flounder shifted to the northwest, those of rock sole, skates and Alaska plaice shifted to the northeast, whereas spatial distributions of yellowfin sole remained relatively stable (Figure 41). The northward expansion for some groundfish seems to relate to warmer bottom temperatures, perhaps due to a northward extension of suitable habitat. With warmer temperatures, the center of groundfish spatial distributions moved farther to the north (Zheng and Kruse 2006).

Changes in spatial distributions of groundfish in the eastern Bering Sea are best illustrated by distributions of Pacific cod biomass from 1982 to 2004 (Figure 42). In the early 1980s, Pacific cod mainly occurred in shallow waters <50 m in the Bristol Bay area and in deep waters >100 m in the northwest of the eastern Bering Sea. However, during 1985-1988 and 1991-1996 the distribution of Pacific cod biomass was widespread across the shelf. In recent years, cod abundance concentrated in the north, around St. Matthew Island, and stayed at a relatively low density in Bristol Bay.

Other striking examples of changes in spatial distributions are provided by rock sole and skates (Figure 43). Rock sole mainly occurred in Bristol Bay and the Pribilof Islands in the 1980s. During the last 15 years, rock sole have expanded to the north up to St. Matthew Island. The biomass of skates has also increased greatly during the last 20 years and expanded northward. Among other commercially important species, biomass of arrowtooth flounder and flathead sole has also increased during the 1980s.

Impacts of Shifts of Spatial Distribution on Crab Recruitment Success

Spatial distributions of Bristol Bay RKC changed profoundly during the last three decades (Hsu 1987; Loher 2001; Zheng and Kruse 2006; Figure 44). Generally speaking, RKC abundance in southern Bristol Bay was high during the 1970s, declined, and was extremely low after 1979 (Zheng and Kruse 2006). Female RKC were found primarily in central Bristol Bay during 1980-1987 and 1992-2006 (Zheng and Kruse 2006). The distribution centers of mature females moved south slightly during 1988-1991 but did not reach the southern locations previously occupied in the 1970s. Loher (2001) hypothesized that changes in near bottom temperatures associated with the 1976/77

regime shift are causes for spatial shifts of RKC female distributions. Because small juvenile RKC are generally located downstream of the mature females (Zheng and Kruse 2006), larval advection appears to be an important process for RKC.

Zheng and Kruse (2008) used the ocean surface current simulator (OSCURS) to perform retrospective analyses of movements of Bristol Bay red king crab larvae from 1967 to 2002. Simulations started at the annual distribution centers of mature females >99 mm CL. The distribution centers were assumed to be the centers of larval hatching. Mature RKC females >99 mm CL are mostly multiparous females. The locations of larval settlements were taken to be the places where 325 degree-days were estimated to have been reached. To estimate larval durations, monthly sea surface temperatures for each year from 1967 to 2002 were estimated for grids of 1 degree longitude and 0.5 degree latitude in the eastern Bering Sea based on the Comprehensive Ocean-Atmosphere Dataset (COADS) from the National Climate Data Center (NCDC). To demonstrate the larval drift tracking for different locations and years, Zheng and Kruse (2008) also simulated the RKC larval drifts in 1975, 1987, and 2004 for two months starting at three locations −− south, middle and north − representing hatching locations of larvae from the southern, middle and northern range of the mature female distribution.

RKC larval drifts were similar among three years (1975, 1987 and 2004) but very different among different hatching locations (Figure 45). At southern and middle locations, larvae generally drifted to the northeast, and at the northern location, larvae drifted to the north or northwest. Larvae hatched in the southern location were estimated to reach central Bristol Bay, whereas larvae hatched in central Bristol Bay were estimated to settle in the northernmost reaches of Bristol Bay. Owing to prevailing currents, larvae hatched in central and northern Bristol Bay are very unlikely to settle in the southern portions of Bristol Bay (Figure 45).

Settling locations appear to have an important impact on resultant year-class strength for Bristol Bay RKC (Figure 46). For years with strong year classes, crab larvae were generally estimated to have settled in the central portion of Bristol Bay (Zheng and Kruse 2008). Because the simulations started at the centers of the annual distribution of the brood stock, larval settling locations from these years likely also represent the centers

of a broader distribution of settling larvae that are well dispersed from south to north along the shallow shelf of Bristol Bay. Larvae associated with weak year-classes generally settled farther downstream in northern Bristol Bay or to the northwest outside of Bristol Bay. Occasionally, larvae hatched in the southern Bristol Bay settled there. Larvae hatching in the middle or later portion of the hatching period may contribute disproportionately to subsequent recruitment; early hatching larvae had longer larval stages and were dispersed farther downstream from the hatching locations than those hatched late in a spawning season (Figure 46).

The simulation results by Zheng and Kruse (2008) show that the northward shifts in mature female distributions made it very difficult to supply larvae to the southern portions of their traditional nursery areas. This reduces the number of suitable habitats to which larvae are delivered (Armstrong et al. 1983; Loher 2001) and may affect recruitment strength. Perhaps this has contributed to long-term decline in recruitment and subsequent mature biomass of Bristol Bay RKC.

ACKNOWLEDGMENTS

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Table 1. Bristol Bay red king crab annual catch and bycatch mortality biomass (million lbs) from June 1 to May 31. A handling mortality rate of 20% for pot and 80% for trawl was assumed to estimate bycatch mortality biomass.

Year	Retained Catch U.S.	Cost-recovery	Foreign	Total	Pot Bycatch Males	Females	Trawl Bycatch	Total Catch
1960	0.600		26.898	27.498				27.498
1961	0.427		44.592	45.019				45.019
1962	0.068		54.275	54.343				54.343
1963	0.653		54.963	55.616				55.616
1964	0.823		58.170	58.993				58.993
1965	1.429		41.294	42.723				43.410
1966	0.997		42.356	43.353				44.732
1967	3.102		33.636	36.738				38.430
1968	8.686		27.469	36.155				34.523
1969	10.403		14.383	24.786				24.463
1970	8.559		12.984	21.543				20.516
1971	12.946		6.134	19.080				20.459
1972	21.745		4.720	26.465				27.296
1973	26.914		0.228	27.142				24.167
1974	42.266		0.476	42.742				42.742
1975	51.326		0.000	51.326				51.326
1976	63.920		0.000	63.920			1.426	65.346
1977	69.968		0.000	69.968			2.685	72.653
1978	87.618		0.000	87.618			2.757	90.375
1979	107.828		0.000	107.828			2.783	110.611
1980	129.948		0.000	129.948			2.135	132.083
1981	33.591		0.000	33.591			0.448	34.039
1982	3.001		0.000	3.001			1.201	4.202
1983	0.000		0.000	0.000			0.885	0.885
1984	4.182		0.000	4.182			2.316	6.498
1985	4.175		0.000	4.175			0.829	5.004
1986	11.394		0.000	11.394			0.432	11.825
1987	12.289		0.000	12.289			0.311	12.600
1988	7.388		0.000	7.388			1.174	8.561
1989	10.265		0.000	10.265			0.374	10.638
1990	20.362	0.081	0.000	20.443	1.139	1.154	0.501	23.237
1991	17.178	0.206	0.000	17.384	0.881	0.142	0.576	18.982
1992	8.043	0.074	0.000	8.117	1.191	0.780	0.571	10.659
1993	14.629	0.053	0.000	14.682	1.649	1.133	0.836	18.300
1994	0.000	0.093	0.000	0.093	0.000	0.000	0.180	0.274
1995	0.000	0.080	0.000	0.080	0.000	0.000	0.213	0.293
1996	8.406	0.108	0.000	8.514	0.356	0.002	0.238	9.109
1997	8.756	0.155	0.000	8.911	0.528	0.034	0.168	9.641
1998	14.757	0.188	0.000	14.946	2.074	1.547	0.355	18.922
1999	11.670	0.186	0.000	11.856	0.679	0.015	0.408	12.958
2000	8.154	0.086	0.000	8.241	0.779	0.078	0.230	9.328
2001	8.403	0.120	0.000	8.523	0.902	0.309	0.330	10.065
2002	9.570	0.096	0.000	9.666	0.956	0.013	0.245	10.881
2003	15.697	0.034	0.000	15.731	1.945	0.709	0.298	18.682
2004	15.245	0.202	0.000	15.447	0.746	0.338	0.277	16.807
2005	18.309	0.209	0.000	18.518	2.923	0.879	0.403	22.723
2006	15.444	0.304	0.000	15.748	1.199	0.067	0.205	17.220
2007	20.366	0.146	0.000	20.512	2.150	0.330	0.233	23.225
2008	20.318	0.000	0.000	20.318	2.518	0.264		23.100

Table 2. Comparison of GHL/TAC and actual catch (million lbs) of Bristol Bay red king crab.

Year	GHL Range	Mid-point	Actual Catch	Rel.Error	%Rel.Error
1980	70-120	95.00	129.95	34.95	36.79
1981	70-100	85.00	33.59	-51.41	-60.48
1982	10-20	15.00	3.00	-12.00	-79.99
1983	0	0.00	0.00	NA	NA
1984	2.5-6	4.25	4.18	-0.07	-1.59
1985	3-5	4.00	4.18	0.18	4.38
1986	6-13	9.50	11.39	1.89	19.94
1987	8.5-17.7	13.10	12.29	-0.81	-6.19
1988		7.50	7.39	-0.11	-1.50
1989		16.50	10.26	-6.24	-37.79
1990		17.10	20.36	3.26	19.08
1991		18.00	17.18	-0.82	-4.57
1992		10.30	8.04	-2.26	-21.91
1993		16.80	14.63	-2.17	-12.93
1994		0.00	0.00	0.00	
1995		0.00	0.00	0.00	
1996		5.00	8.41	3.41	68.11
1997		7.00	8.76	1.76	25.09
1998		16.40	14.76	-1.64	-10.02
1999		10.66	11.67	1.01	9.48
2000		8.35	8.15	-0.20	-2.34
2001		7.15	8.40	1.25	17.52
2002		9.27	9.57	0.30	3.24
2003		15.71	15.70	-0.01	-0.08
2004		15.40	15.25	-0.15	-1.00
2005		18.33	18.31	-0.02	-0.11
2006		15.53	15.44	-0.08	-0.53
2007		20.38	20.37	-0.02	-0.08
Total		461.23	431.38	-29.85	-6.47

Table 3. Annual sample sizes for catch by length and shell condition for retained catch and bycatch of Bristol Bay red king crab.

Year	Trawl Survey Males	Trawl Survey Females	Retained Catch	Pot Bycatch Males	Pot Bycatch Females	Trawl Bycatch Males	Trawl Bycatch Females
1968	3,684	2,165	18,044				
1969	6,144	4,992	22,812				
1970	1,546	1,216	3,394				
1971			10,340				
1972	1,106	767	15,046				
1973	1,783	1,888	11,848				
1974	2,505	1,800	27,067				
1975	2,943	2,139	29,570				
1976	4,724	2,956	26,450			2,327	676
1977	3,636	4,178	32,596			14,014	689
1978	4,132	3,948	27,529			8,983	1,456
1979	5,807	4,663	27,900			7,228	2,821
1980	2,412	1,387	34,747			47,463	39,689
1981	3,478	4,097	18,029			42,172	49,634
1982	2,063	2,051	11,466			84,240	47,229
1983	1,524	944	0			204,464	104,910
1984	2,679	1,942	4,404			357,981	147,134
1985	792	415	4,582			169,767	30,693
1986	1,962	367	5,773			62,023	20,800
1987	1,168	1,018	4,230			60,606	32,734
1988	1,834	546	9,833			102,037	57,564
1989	1,257	550	32,858			47,905	17,355
1990	858	603	7,218	873	699	5,876	2,665
1991	1,378	491	36,820	1,801	375	2,964	962
1992	513	360	23,552	3,248	2,389	1,157	2,678
1993	1,009	534	32,777	5,803	5,942		
1994	443	266	0	0	0	4,953	3,341
1995	2,154	1,718	0	0	0	1,729	6,006
1996	835	816	8,896	230	11	24,583	9,373
1997	1,282	707	15,747	4,102	906	9,035	5,759
1998	1,097	1,150	16,131	11,079	9,130	25,051	9,594
1999	820	540	17,666	1,048	36	16,653	5,187
2000	1,278	1,225	14,091	8,970	1,486	36,972	10,673
2001	611	743	12,854	9,102	4,567	56,070	32,745
2002	1,032	896	15,932	9,943	302	27,705	25,425
2003	1,669	1,311	16,212	17,998	10,327	281	307
2004	2,871	1,599	20,038	8,258	4,112	137	120
2005	1,283	1,682	21,938	55,019	26,775	186	124
2006	2,321	2,672	18,027	29,383	3,594	217	168
2007	2,252	2,499	22,387	58,097	12,411	1,981	2,880
2008	2,362	3,352	14,567	49,315	8,488		

Table 4. Annual catch (million crabs) and catch per unit effort of the Bristol Bay red king crab fishery.

| Year | Japanese Tanglenet | | Russian Tanglenet | | U.S. Pot/trawl | | Standardized |
	Catch	Crabs/tan	Catch	Crabs/tan	Catch	Crabs/potlift	Crabs/tan
1960	1.949	15.2	1.995	10.4	0.088		15.8
1961	3.031	11.8	3.441	8.9	0.062		12.9
1962	4.951	11.3	3.019	7.2	0.010		11.3
1963	5.476	8.5	3.019	5.6	0.101		8.6
1964	5.895	9.2	2.800	4.6	0.123		8.5
1965	4.216	9.3	2.226	3.6	0.223		7.7
1966	4.206	9.4	2.560	4.1	0.140	52	8.1
1967	3.764	8.3	1.592	2.4	0.397	37	6.3
1968	3.853	7.5	0.549	2.3	1.278	27	7.8
1969	2.073	7.2	0.369	1.5	1.749	18	5.6
1970	2.080	7.3	0.320	1.4	1.683	17	5.6
1971	0.886	6.7	0.265	1.3	2.405	20	5.8
1972	0.874	6.7			3.994	19	
1973	0.228				4.826	25	
1974	0.476				7.710	36	
1975					8.745	43	
1976					10.603	33	
1977					11.733	26	
1978					14.746	36	
1979					16.809	53	
1980					20.845	37	
1981					5.308	10	
1982					0.541	4	
1983					0.000		
1984					0.794	7	
1985					0.796	9	
1986					2.100	12	
1987					2.122	10	
1988					1.236	8	
1989					1.685	8	
1990					3.130	12	
1991					2.661	12	
1992					1.208	6	
1993					2.270	9	
1994					0.015		
1995					0.014		
1996					1.264	16	
1997					1.338	15	
1998					2.238	15	
1999					1.923	12	
2000					1.272	12	
2001					1.287	19	
2002					1.484	20	
2003					2.510	18	
2004					2.272	23	
2005					2.763	30	
2006					2.477	31	
2007					3.131	28	

Table 5. Summary of statistics for the model

Parameter counts

Fixed growth parameters	9
Fixed recruitment parameters	2
Fixed length-weight relationship parameters	6
Fixed mortality parameters	4
Fixed highgrading parameters	4
Fixed initial (1968) length composition parameters	56
Total number of fixed parameters	81
Free growth parameters	4
Initial abundance (1968)	1
Recruitment-distribution parameters	2
Mean recruitment parameters	1
Male recruitment deviations	40
Female recruitment deviations	40
Natural and fishing mortality parameters	6
Survey catchability parameters	3
Pot male fishing mortality deviations	44
Pot female bycatch fishing mortality deviations	22
Trawl bycatch fishing mortality deviations	32
Free selectivity parameters	28
Total number of free parameters	223
Total number of fixed and free parameters	304

Negative log likelihood components

Length compositions---retained catch	-992.286
Length compositions---pot male discard	-712.541
Length compositions---pot female discard	-1869.540
Length compositions---Tanner crab discards	-224.906
Length compositions---trawl discard	-1673.350
Length compositions---survey	-48883.800
Retained catch biomass	46.112
Pot discard male biomass	165.057
Pot discard female biomass	0.110
Trawl discard	5.957
Survey biomass	74.892
Recruitment variation	148.433
Sex ratio of recruitment	0.219
Others (BSFRF, Q, Tanner crab disc biomass)	6.737
Total	-53908.300

Table 6. Summary of model parameter estimates for Bristol Bay red king crab. Estimated values and standard deviations. All values are on a log scale.

Year	Females	S. dev.	Males	S.dev.	Males	S.dev.	Females	S.dev.	Est.	S.dev.
	Recruits				F for Directed Pot Fishery				F for Trawl	
Mean	16.253	0.023			-1.790	0.043	0.011	0.002	-4.760	0.062
1968					1.858	0.061				
1969	-0.312	0.108	0.905	0.065	1.812	0.075				
1970	0.538	0.115	0.877	0.094	1.520	0.076				
1971	-0.314	0.095	2.013	0.051	1.188	0.071				
1972	0.535	0.218	0.057	0.158	1.251	0.069				
1973	-0.442	0.116	1.469	0.057	0.998	0.070				
1974	0.187	0.089	1.445	0.059	1.183	0.065				
1975	0.336	0.059	2.326	0.043	1.030	0.061				
1976	-0.365	0.237	0.553	0.121	1.115	0.062			-0.399	0.078
1977	0.560	0.166	0.314	0.119	1.209	0.063			0.166	0.077
1978	0.568	0.131	0.794	0.091	1.396	0.056			0.110	0.076
1979	0.370	0.127	1.065	0.089	1.538	0.046			0.111	0.076
1980	0.074	0.123	1.266	0.094	2.099	0.004			0.115	0.077
1981	0.293	0.086	1.055	0.080	1.927	0.084			-0.460	0.077
1982	-0.125	0.049	1.973	0.049	-0.079	0.082			1.220	0.085
1983	-0.177	0.084	1.026	0.054	-10.199	0.455			1.158	0.081
1984	0.440	0.091	1.013	0.043	0.571	0.060			2.000	0.003
1985	0.081	0.210	-1.419	0.135	0.658	0.059			1.230	0.077
1986	0.332	0.059	0.298	0.046	1.273	0.057			0.210	0.076
1987	0.142	0.125	-0.445	0.080	0.973	0.053			-0.312	0.076
1988	-0.349	0.252	-1.528	0.156	0.101	0.049			0.875	0.075
1989	0.386	0.144	-0.839	0.108	0.239	0.047			-0.425	0.075
1990	-0.112	0.094	0.067	0.061	0.883	0.044	1.835	0.209	-0.181	0.075
1991	-0.245	0.112	-0.540	0.073	0.860	0.046	-0.288	0.209	0.058	0.075
1992	-0.289	0.324	-2.515	0.219	0.341	0.047	1.996	0.180	0.170	0.075
1993	-0.361	0.092	-0.607	0.056	1.003	0.049	1.833	0.209	0.559	0.075
1994	-0.370	0.352	-2.767	0.235	-10.896	0.442	1.647	6.468	-0.873	0.076
1995	0.015	0.038	0.888	0.035	-11.145	0.440	0.503	7.029	-0.861	0.076
1996	-0.061	0.102	-0.476	0.069	0.026	0.044	-3.799	0.245	-0.867	0.075
1997	-0.750	0.346	-2.853	0.239	0.148	0.044	-1.250	0.211	-1.228	0.075
1998	-0.243	0.104	-0.503	0.064	0.852	0.046	1.871	0.210	-0.532	0.074
1999	-0.084	0.058	0.521	0.043	0.402	0.046	-2.321	0.214	-0.407	0.075
2000	-0.114	0.169	-0.712	0.102	0.012	0.045	-0.393	0.211	-1.038	0.075
2001	0.929	0.179	-1.625	0.149	0.004	0.045	0.956	0.211	-0.731	0.075
2002	0.168	0.040	0.940	0.035	0.094	0.045	-2.329	0.215	-1.085	0.075
2003	-0.065	0.178	-0.905	0.115	0.597	0.045	1.024	0.211	-1.262	0.076
2004	0.001	0.085	0.368	0.075	0.423	0.047	0.298	0.211	-0.880	0.076
2005	0.155	0.050	0.924	0.047	0.814	0.049	0.766	0.212	-1.391	0.077
2006	-0.408	0.134	0.016	0.086	0.486	0.052	-1.605	0.213	-1.207	0.078
2007	-0.287	0.227	-0.818	0.140	0.738	0.057	-0.300	0.213	-1.305	0.079
2008	-0.065	0.347	-1.734	0.249	0.700	0.063	-0.446	0.213		

Table 6 (continue). Summary of model parameter estimates for Bristol Bay red king crab. Estimated values and standard deviations.

Parameter	Value	St.dev.	Parameter	Value	St.dev.
Mm80-84	0.5233	0.0163	log_srv_L50, m, 70-72	5.2000	0.0001
Mf80-84	0.8983	0.0237	srv_slope, f, 70-72	0.1475	0.0107
Mf76-79,85-93	0.0512	0.0064	log_srv_L50, f, 70-72	4.3838	0.0139
log_betal, females	0.1827	0.0536	log_srv_L50, m, 73-81	4.3560	0.0215
log_betal, males	0.6458	0.0713	srv_slope, f, 73-81	0.0644	0.0035
log_betar, females	-0.2276	0.0874	log_srv_L50, f, 73-81	4.4028	0.0157
log_betar, males	-0.2540	0.0583	log_srv_L50, m, 82-08	4.6407	0.0518
Q, females, 70-72	0.1725	0.0178	srv_slope, f, 82-08	0.0363	0.0018
Q, males, 70-72	0.8521	0.0963	log_srv_L50, f, 82-08	4.5883	0.0299
Q, 68-69, 73-08	0.9030	0.0226	log_srv_L50, m, 68-69	4.5104	0.0170
moltp_slope	0.0899	0.0030	srv_slope, f, 68-69	0.0187	0.0021
log_moltp_L50	4.9379	0.0028	log_srv_L50, f, 68-69	5.0403	0.0723
log_N68	18.9570	0.0318	TC_slope, females	0.2780	0.0646
log_avg_L50, 73-08	4.9256	0.0013	log_TC_L50, females	4.5423	0.0132
log_avg_L50, 68-72	4.8654	0.0062	TC_slope, males	0.0771	0.0094
ret_fish_slope, 73-08	0.5006	0.0219	log_TC_L50, males	4.7500	0.0005
ret_fish_slope, 68-72	0.3004	0.0382	log_TC_F, males, 91	-3.8884	0.0827
pot disc.males, φ	-0.2493	0.0117	log_TC_F, males, 92	-5.0014	0.0826
pot disc.males, κ	0.0028	0.0001	log_TC_F, males, 93	-6.2604	0.0865
pot disc.males, γ	-0.0123	0.0005	log_TC_F, females, 91	-2.9548	0.0851
sel_62.5mm, 68-72	1.4000	0.0001	log_TC_F, females, 92	-4.1387	0.0846
post disc.fema., slope	0.4768	0.1380	log_TC_F, females, 93	-4.7271	0.0841
log_pot disc.fema., L50	4.3969	0.0081			
trawl disc slope	0.0603	0.0030			
log_trawl disc L50	4.9508	0.0284			

Table 7. Annual abundance estimates (million crabs), mature male biomass (MMB, million lbs), and total survey biomass estimates (million lbs) for red king crab in Bristol Bay estimated by length-based analysis from 1968-2008. Mature male biomass for year *t* is on Feb. 15, year *t*+1. Size measurements are mm CL.

Year (t)	Males Mature (>119mm)	Males Legal (>134mm)	Males MMB (>119mm)	Males MMB SD	Females Mature (>89mm)	Total Survey Biomass Model Est. (>64mm)	Total Survey Biomass Area-swept (>64mm)
1968	14.906	8.754	34.000	1.750	61.277	81.178	80.070
1969	14.712	6.273	34.578	2.083	62.477	82.050	87.140
1970	17.963	7.004	46.772	2.779	65.379	35.811	43.040
1971	21.089	9.149	61.789	3.553	72.707	44.304	
1972	27.373	12.095	79.558	4.604	90.701	55.104	50.267
1973	34.469	15.319	107.277	5.501	107.753	196.924	159.504
1974	50.394	21.092	150.254	6.883	111.868	226.546	192.378
1975	55.518	28.626	174.013	8.031	117.598	271.416	209.197
1976	56.992	31.884	175.786	7.975	150.859	305.554	284.115
1977	63.014	32.288	187.306	7.467	184.011	318.271	362.950
1978	78.382	35.569	220.089	7.385	173.854	314.356	322.413
1979	76.945	41.401	204.840	8.069	154.290	294.859	243.342
1980	57.612	36.885	76.084	4.068	144.373	262.631	228.580
1981	20.431	11.744	31.296	3.025	58.159	113.981	112.143
1982	10.713	5.035	26.726	2.015	25.948	59.043	132.610
1983	8.767	4.265	24.309	1.378	16.720	46.159	47.235
1984	8.360	3.821	19.079	1.002	14.004	42.155	150.494
1985	8.850	3.220	28.436	1.206	11.937	30.487	33.005
1986	13.162	5.594	38.199	1.608	17.838	40.117	46.290
1987	15.778	7.388	49.418	1.906	21.564	45.100	66.139
1988	16.054	9.164	59.275	2.074	26.738	47.740	50.570
1989	17.310	10.571	65.214	2.161	25.742	50.828	58.735
1990	17.551	11.362	59.558	2.186	22.978	52.119	52.674
1991	14.243	10.086	48.961	2.136	22.672	47.684	82.835
1992	11.549	8.119	45.479	2.069	22.661	42.578	34.732
1993	12.347	7.583	40.926	2.048	20.062	41.423	47.159
1994	11.848	7.032	52.662	2.110	16.878	36.547	29.789
1995	12.188	8.768	58.303	2.057	16.003	43.666	35.927
1996	12.251	9.382	53.480	1.959	21.714	49.854	40.886
1997	11.729	8.416	49.937	1.883	31.170	52.172	78.993
1998	15.860	8.222	54.437	2.043	30.316	54.738	76.289
1999	17.621	9.575	64.698	2.321	26.633	55.787	56.086
2000	15.923	11.063	65.106	2.364	29.345	57.866	63.132
2001	15.057	10.792	63.346	2.288	33.153	59.875	47.621
2002	17.354	10.545	69.783	2.328	32.099	66.013	64.534
2003	18.267	11.697	68.053	2.375	39.065	70.198	87.428
2004	16.525	11.246	64.292	2.358	47.417	73.149	88.288
2005	19.560	10.909	67.681	2.576	47.718	81.019	96.177
2006	20.637	11.828	75.578	2.932	55.677	85.865	86.116
2007	22.235	13.092	77.651	3.398	62.816	91.498	93.625
2008	25.933	13.814	90.254	3.558	58.303	91.934	99.641

Table 8. Comparison of projected mature male biomass (million lbs) on Feb. 15, retained catch (million lbs), their 95% limits, and mean fishing mortality with no directed fishery, $F_{40\%}$, $F_{35\%}$, and ADF&G harvest strategy with $F_{35\%}$ constraint during 2008-2017. (This projection was conducted on August 2008 and will be updated in August 2009).

No directed fishery

Year	MMB	95% limits of MMB		Catch	95% limits of catch		Mean F
2008	116.202	112.204	120.406	0	0	0	0
2009	143.486	138.549	148.676	0	0	0	0
2010	153.376	148.101	158.919	0	0	0	0
2011	151.045	145.455	157.102	0	0	0	0
2012	150.017	136.958	171.109	0	0	0	0
2013	154.485	127.329	202.063	0	0	0	0
2014	161.403	118.720	220.974	0	0	0	0
2015	168.243	113.310	239.856	0	0	0	0
2016	173.957	111.725	254.114	0	0	0	0
2017	179.380	108.490	266.927	0	0	0	0

$F_{40\%}$

Year	MMB	95% limits of MMB		Catch	95% limits of catch		Mean F
2008	99.362	95.944	102.957	16.507	15.939	17.104	0.260
2009	107.966	104.251	111.871	19.403	18.736	20.105	0.260
2010	99.940	96.503	103.550	21.115	20.388	21.878	0.260
2011	83.671	80.170	87.774	19.968	19.284	20.697	0.260
2012	74.659	65.125	92.573	16.091	14.063	18.049	0.243
2013	74.590	54.391	114.145	13.850	9.605	18.559	0.232
2014	77.837	48.306	121.884	13.762	7.156	22.039	0.234
2015	80.965	47.160	131.341	14.409	6.180	23.812	0.236
2016	82.881	46.290	133.481	15.069	5.779	25.666	0.236
2017	84.565	46.198	138.211	15.547	5.685	26.375	0.237

$F_{35\%}$

Year	MMB	95% limits of MMB		Catch	95% limits of catch		Mean F
2008	95.434	92.150	98.886	20.328	19.629	21.064	0.330
2009	100.614	97.152	104.254	22.895	22.107	23.723	0.330
2010	90.080	86.983	93.334	24.115	23.286	24.988	0.330
2011	73.407	70.915	76.500	21.260	19.916	22.632	0.316
2012	65.684	57.272	81.145	15.790	13.410	19.307	0.276
2013	66.401	47.847	102.167	13.966	9.030	19.916	0.271
2014	69.908	43.003	109.457	14.310	6.815	24.173	0.278
2015	72.955	42.286	118.456	15.236	6.085	26.020	0.283
2016	74.608	41.936	120.746	16.059	5.757	28.229	0.286
2017	75.979	42.351	122.838	16.593	5.834	28.876	0.288

ADF&G harvest strategy

Year	MMB	95% limits of MMB		Catch	95% limits of catch		Mean F
2008	95.580	92.456	98.886	20.664	18.960	21.064	0.327
2009	102.379	99.399	105.266	22.195	19.915	22.852	0.303
2010	96.632	94.010	99.071	19.516	18.127	20.277	0.249
2011	84.318	81.955	87.172	16.584	15.313	17.349	0.215
2012	76.171	67.965	89.883	15.796	12.209	20.530	0.226
2013	74.583	57.140	108.066	15.853	8.853	22.126	0.253
2014	76.265	50.291	113.400	15.896	6.988	25.563	0.262
2015	78.410	47.550	121.062	16.029	5.389	27.014	0.260
2016	79.844	46.253	123.963	16.195	5.059	28.714	0.256
2017	81.313	47.453	126.263	16.386	4.833	29.367	0.254

Table 9. List of years, survey stations, dates and red king crab sizes founded in groundfish stomachs during NMFS summer trawl surveys. All identified crabs are females, mostly mature females. (Source: G.M. Lang, NMFS, Seattle).

YEAR	RLAT	RLONG	STATION	DATE	PRED_LEN	RKC CL(mm)
1984	57.99	-160.87	J-12	6/13/1984	92	110
1984	57.33	-162.16	H-10	6/14/1984	79	130
1981	57.34	-162.13	H-10	5/29/1981	67	121
1981	57.34	-162.13	H-10	5/29/1981	67	106
1981	56.69	-161.00	F-12	6/1/1981	66	100
1981	56.69	-161.00	F-12	6/1/1981	69	53
1981	57.01	-160.95	G-12	6/1/1981	69	160
1981	57.99	-160.87	J-12	6/21/1981	51	91
1981	57.99	-160.87	J-12	6/21/1981	62	95
1985	56.95	-159.85	G-14	10/29/1985	85	52
1986	57.67	-161.49	I-11	6/7/1986	89	91
1989	56.17	-161.52	D-11	6/4/1989	95	84
1989	56.17	-161.52	D-11	6/4/1989	95	99
1991	57.00	-159.12	G-15	6/8/1991	56	17
1992	57.32	-162.15	H-10	6/9/1992	98	101
1992	57.32	-162.15	H-10	6/9/1992	98	87
1992	57.32	-162.15	H-10	6/9/1992	98	95
1992	57.32	-162.15	H-10	6/9/1992	97	117
1992	56.67	-160.99	F-12	6/7/1992	89	144
1985	56.42	-161.58	E-11	4/25/1985	82	94
1992	56.67	-160.99	F-12	6/7/1992	89	144
1992	57.32	-162.15	H-10	6/9/1992	98	101
1992	57.32	-162.15	H-10	6/9/1992	98	87
1992	57.32	-162.15	H-10	6/9/1992	98	95
1992	57.32	-162.15	H-10	6/9/1992	97	117
2000	56.00	-162.25	D-10	5/28/2000	75	120
2002	57.68	-160.27	I-13	6/3/2002	70	125

Table 10. Summary of red king crab biomass (million lbs) in Bristol Bay that were consumed by groundfish during late May to September. Pacific cod is the main predator. (Source: G.M. Lang, NMFS, Seattle).

Year	Red king crab biomass
1984	3.719
1985	0.000
1986	14.457
1987	7.403
1988	0.000
1989	0.203
1990	1.853
1991	0.039
1992	4.488
1993	3.833
1994	1.545
1995	0.993
1996	0.000
1997	0.000
1998	2.192
1999	1.718
2000	1.199
2001	0.000
2002	2.008
2003	0.000
2004	0.000
2005	11.677

Mature Harvest Rate
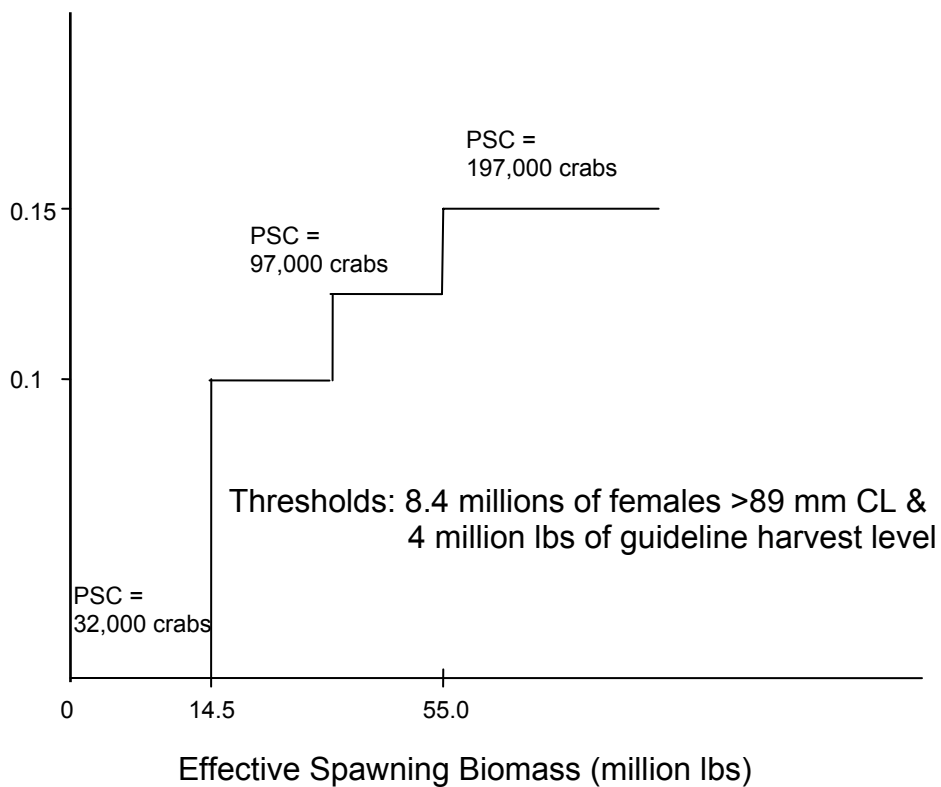
Figure 1. Current harvest rate strategy (line) for the Bristol Bay red king crab fishery and annual prohibited species catch (PSC) limits (numbers of crabs) of Bristol Bay red king crabs in the groundfish fisheries in zone 1 in the eastern Bering Sea. Harvest rates are based on current-year estimates of effective spawning biomass (ESB), whereas PSC limits apply to previous-year ESB.
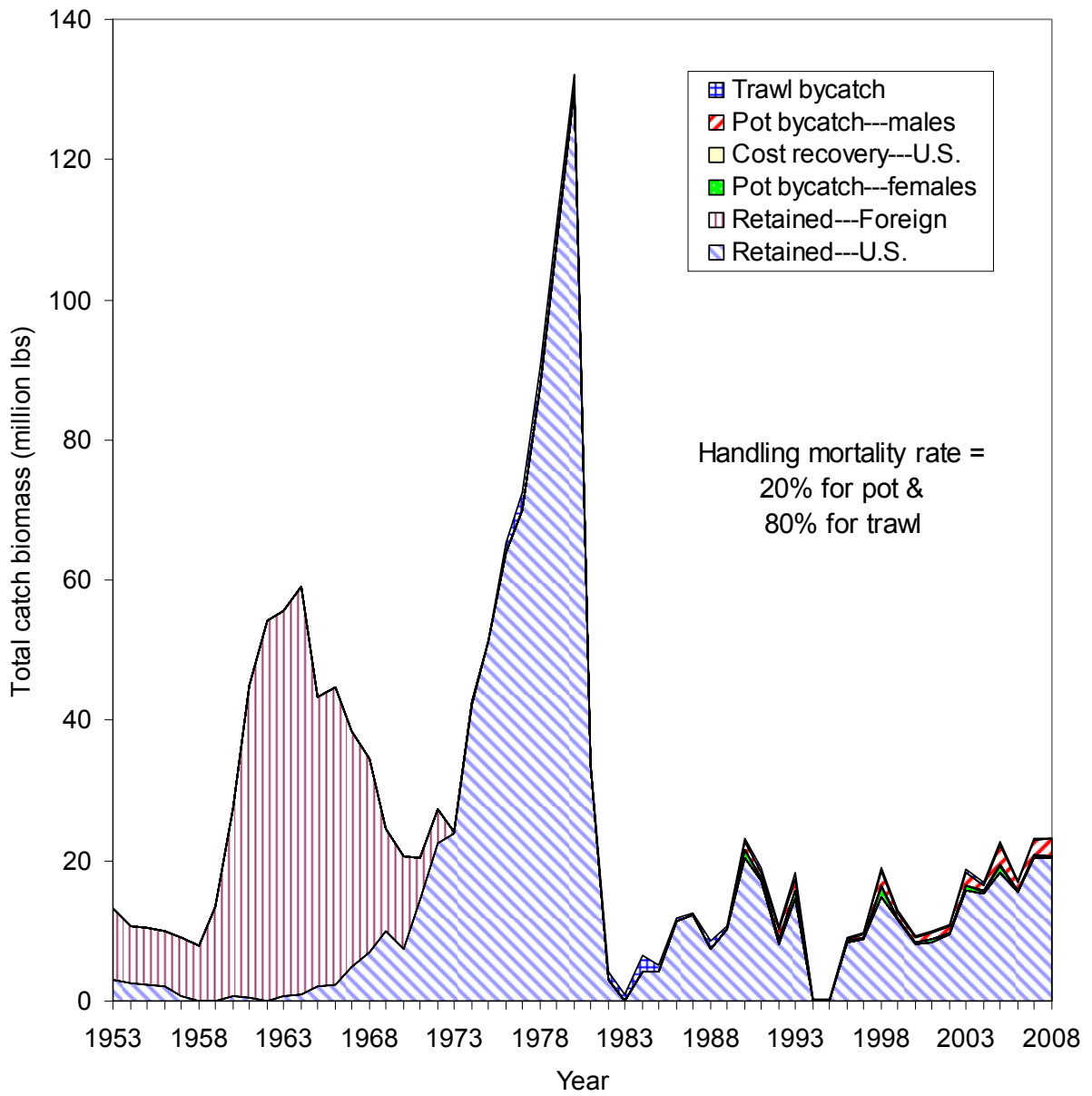
Figure 2. Retained catch biomass and bycatch mortality biomass (million lbs) for Bristol Bay red king crab from 1960 to 2008. Handling mortality rates were assumed to be 0.2 for the directed pot fishery and 0.8 for the trawl fisheries.
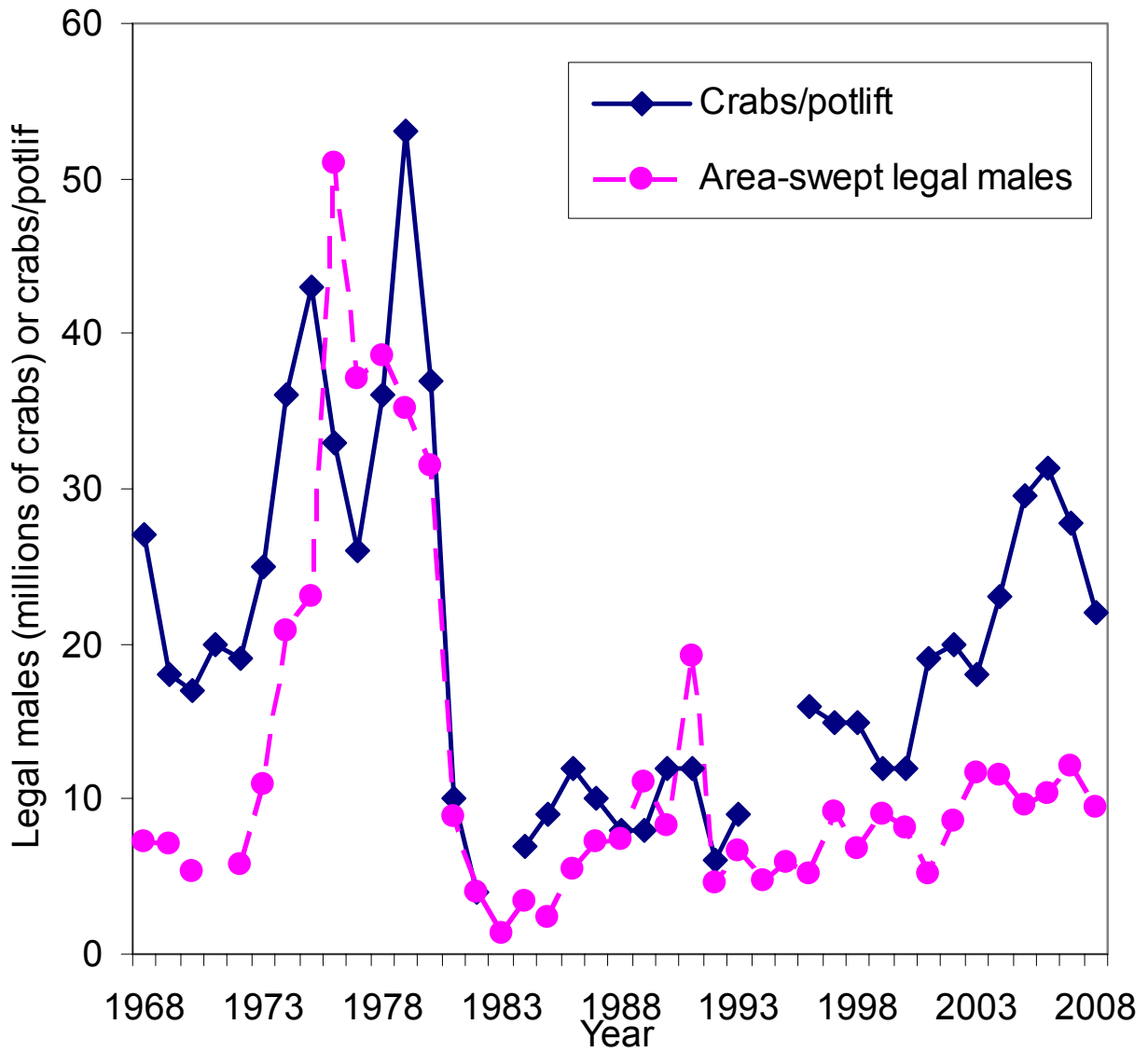
Figure 3. Comparison of survey legal male abundances and catches per unit effort for Bristol Bay red king crab from 1968 to 2008.
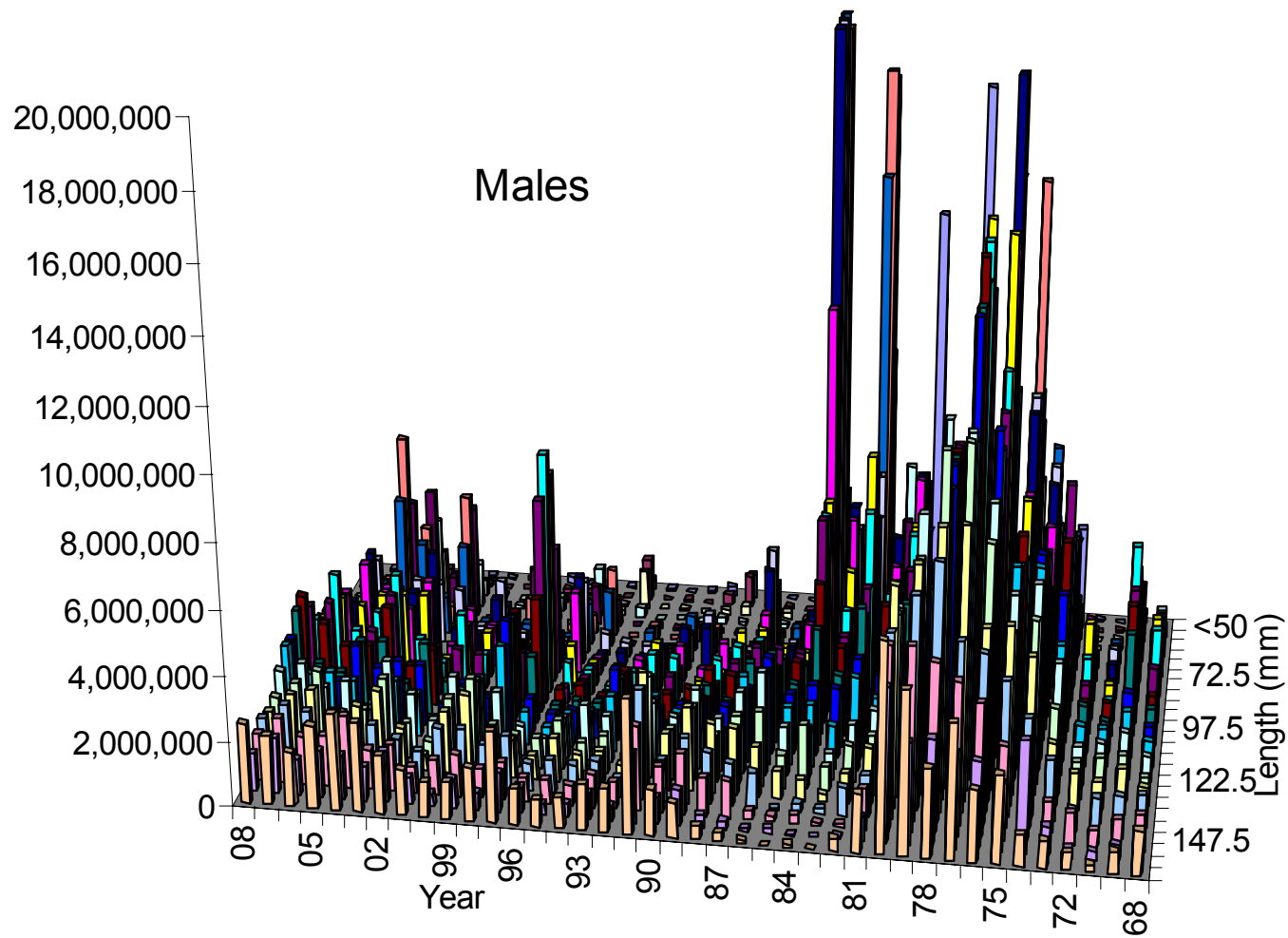
Figure 4. Survey abundances by length for male Bristol Bay red king crabs from 1968 to 2008.

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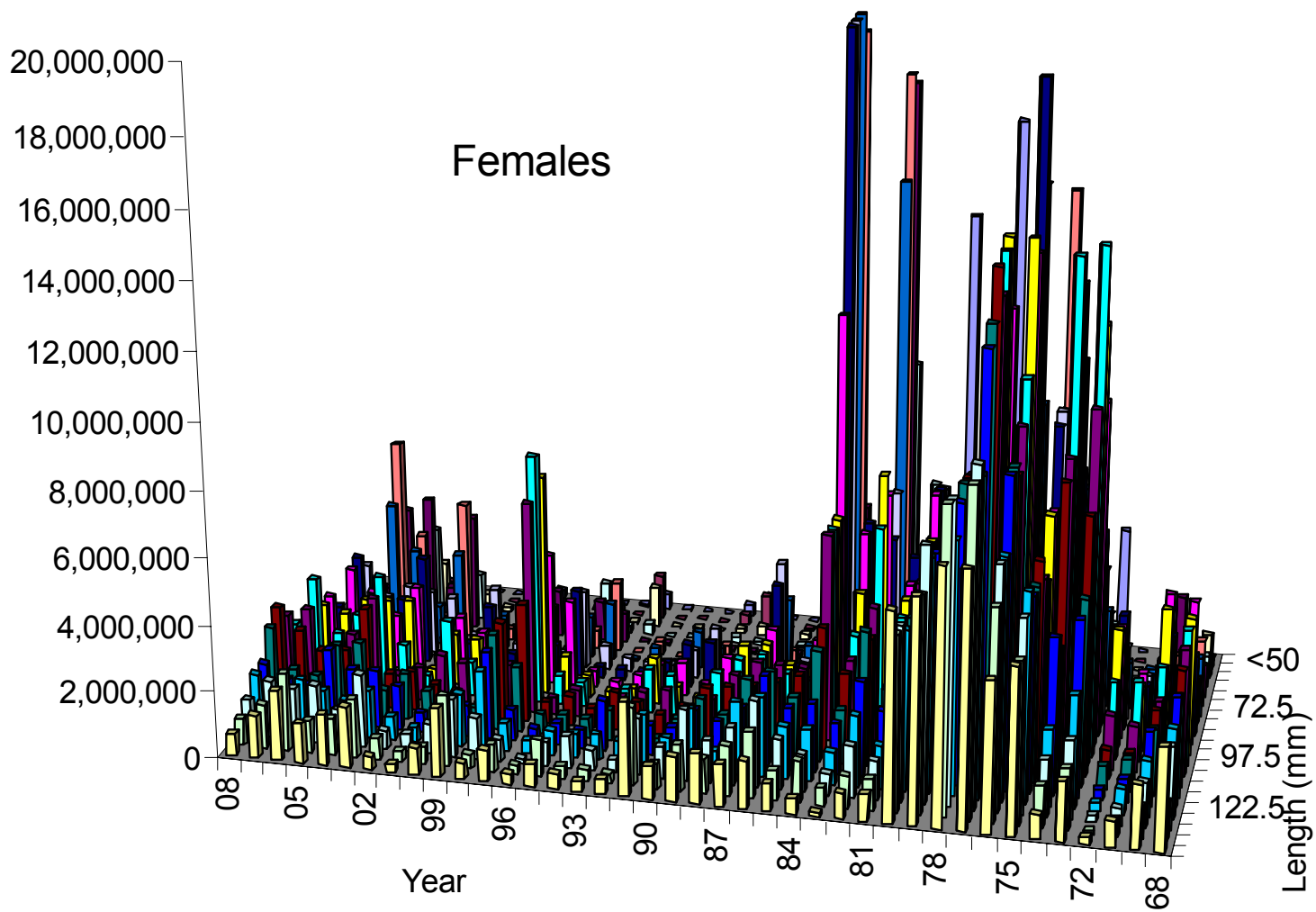
Figure 5. Survey abundances by length for female Bristol Bay red king crabs from 1968 to 2008.

61
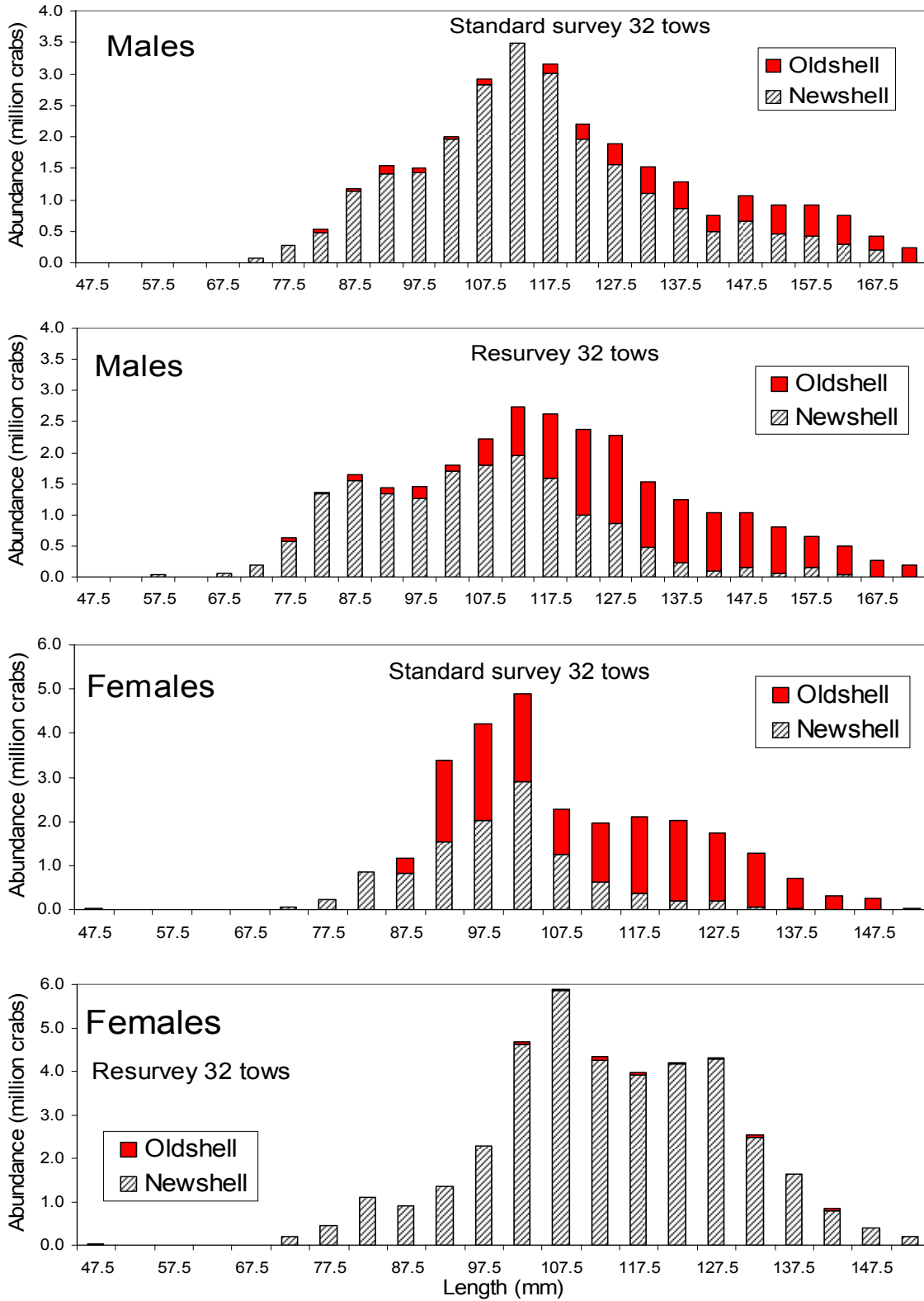
Figure 6. Comparison of area-swept estimates of abundance in 32 stations from the standard trawl survey and resurvey in 2008.
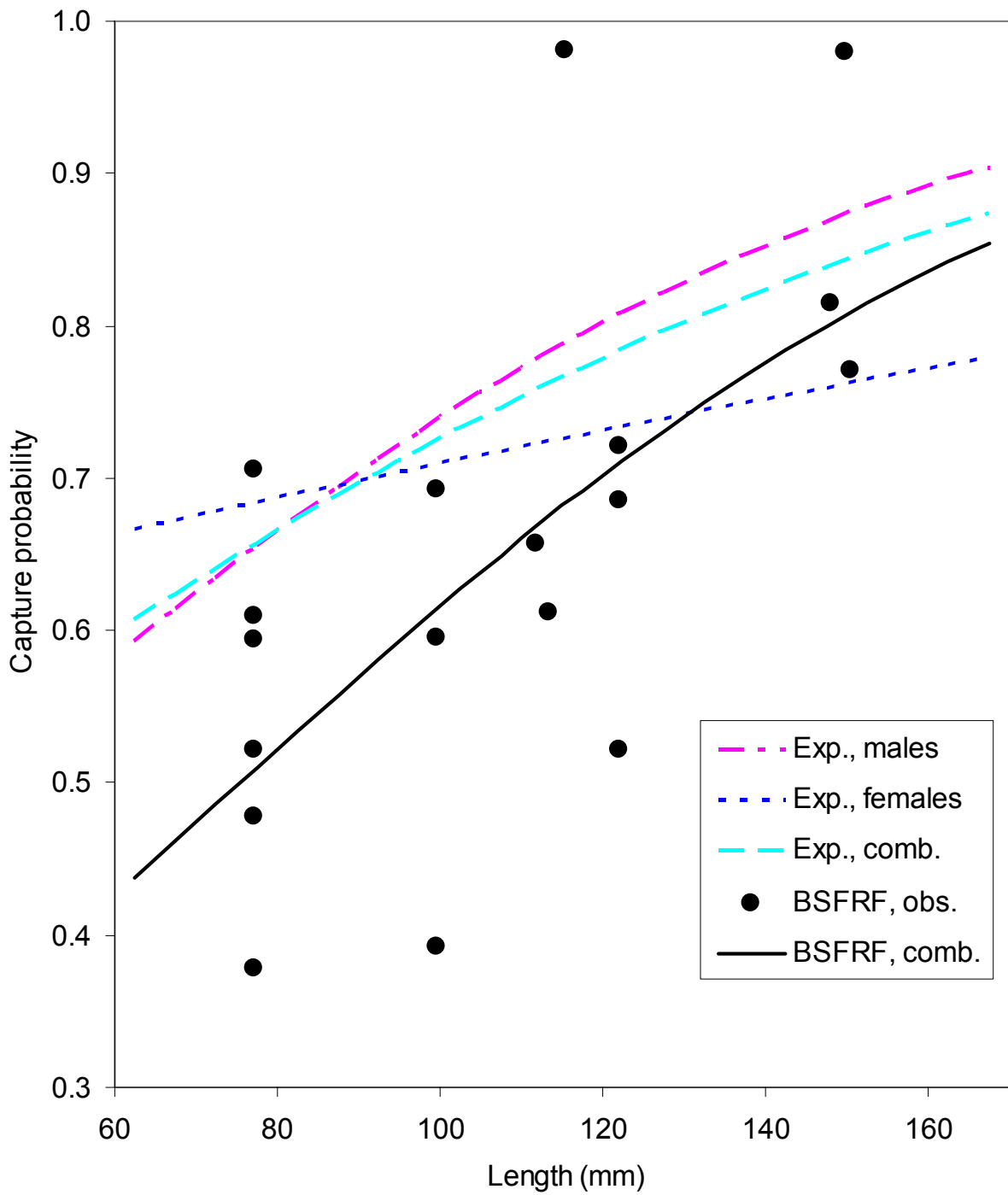
Figure 7. Estimated capture probabilities for NMFS Bristol Bay red king crab trawl surveys by Weinberg et al. (2004) and the Bering Sea Fisheries Research Foundation surveys.
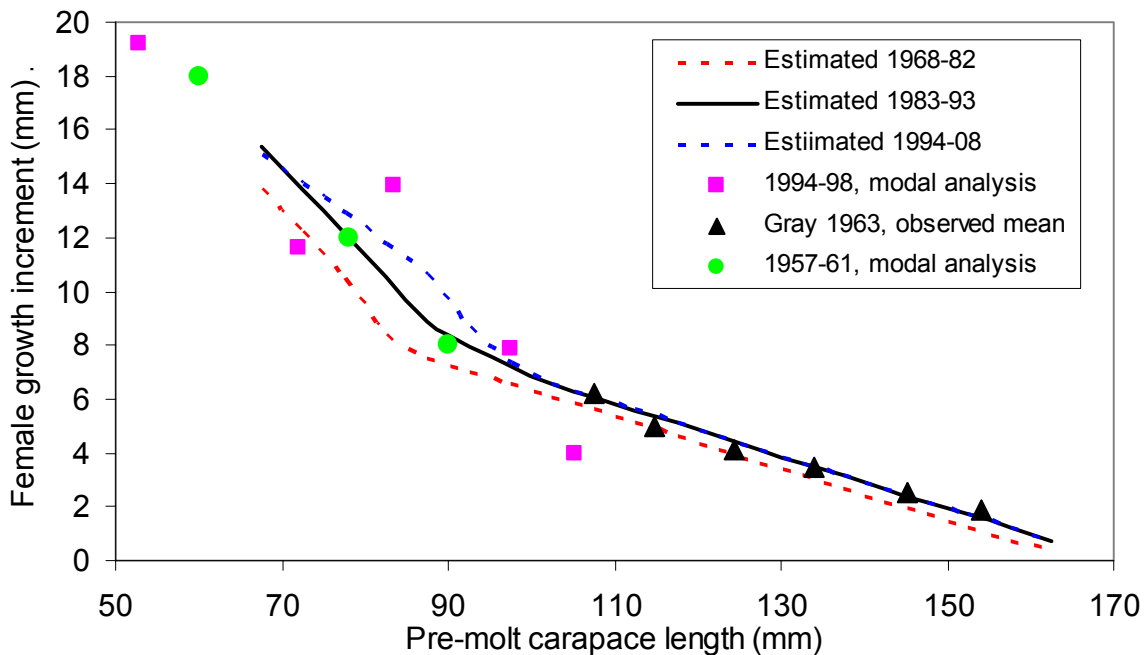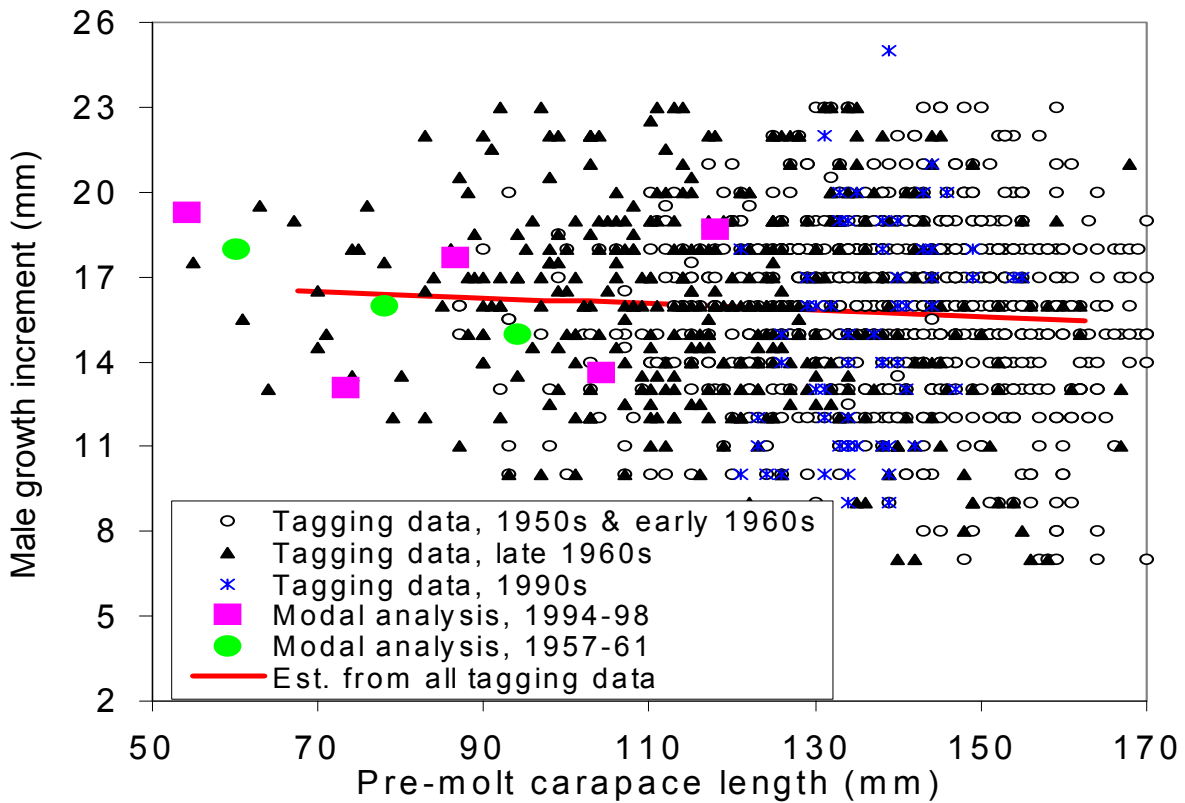
Figure 8. Mean growth increments per molt for Bristol Bay red king crab. Note: "tagging"---based on tagging data; "mode"---based on modal analysis.
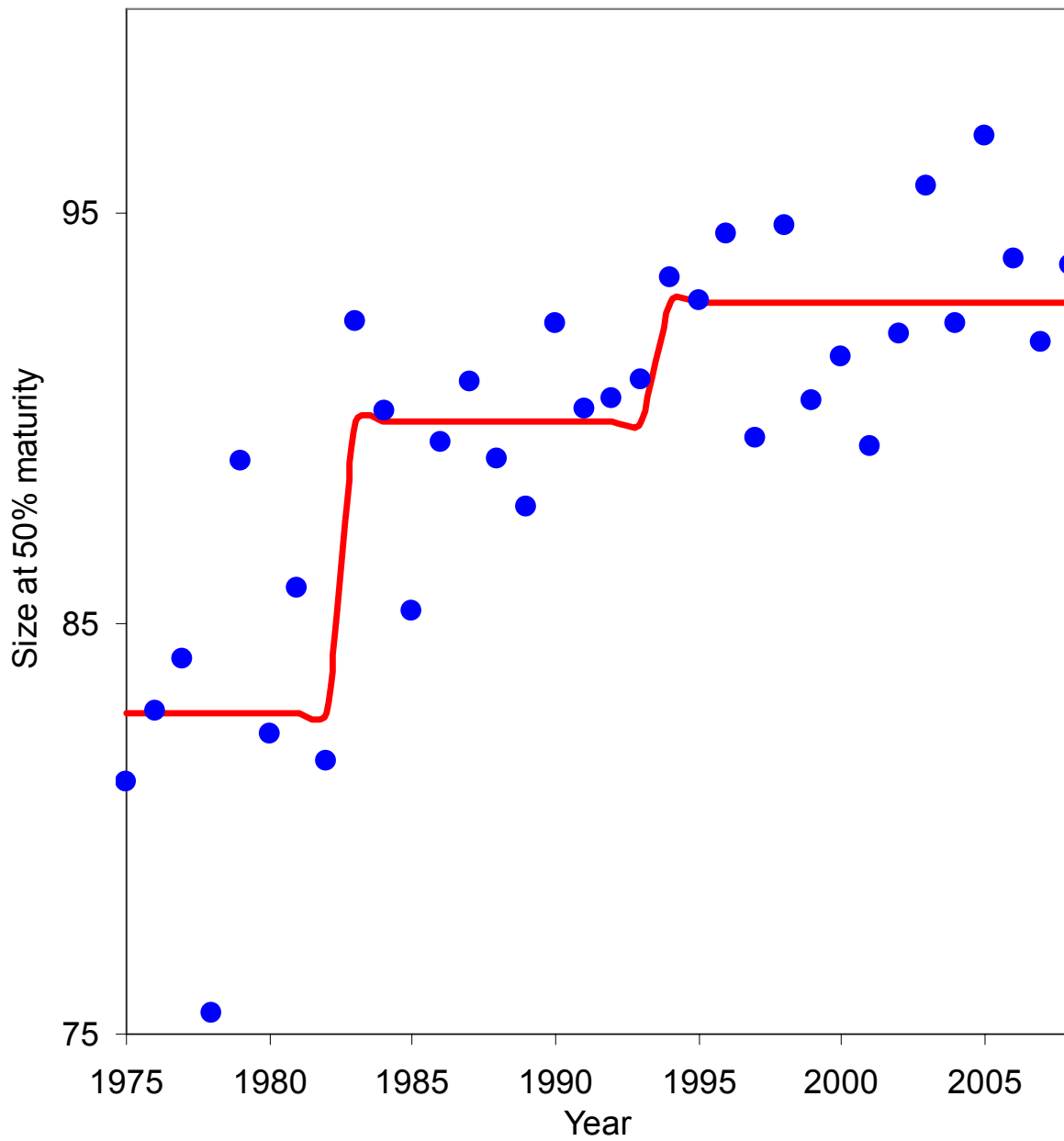
Figure 9. Estimated sizes at 50% maturity for Bristol Bay female red king crab from 1975 to 2008. Averages for three periods (1975-82, 1983-93, and 1994-08) are plotted with a line.
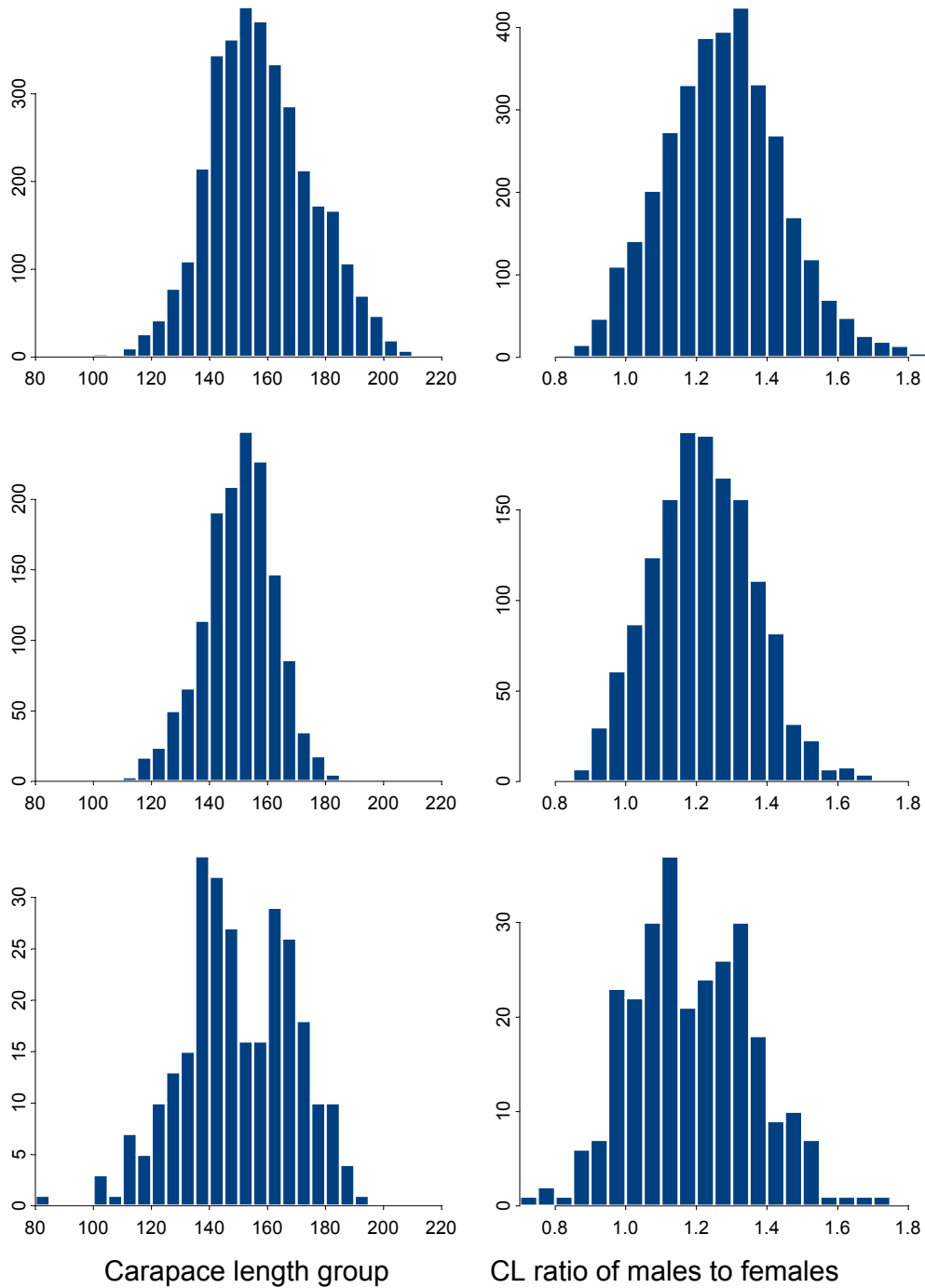
Figure 10. Histograms of carapace lengths (CL) and CL ratios of males to females for male shell ages ≤13 months of red king crab males in grasping pairs; Powell's Kodiak data. Upper plot: all locations and years pooled; middle plot: location 11; lower plot: locations 4 and 13. Sizes at maturity for Kodiak red king crab are about 15 mm larger than those for Bristol Bay red king crab. (Source: Doug Pengilly, ADF&G).
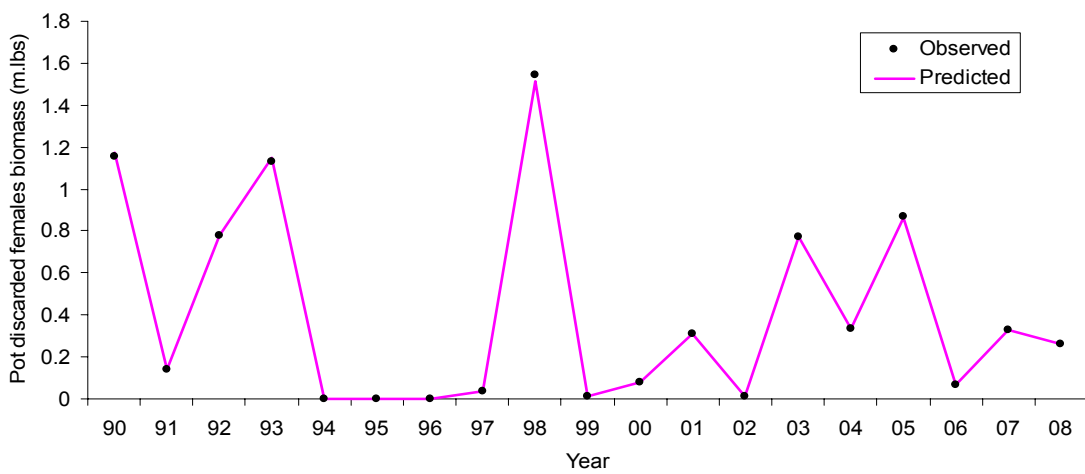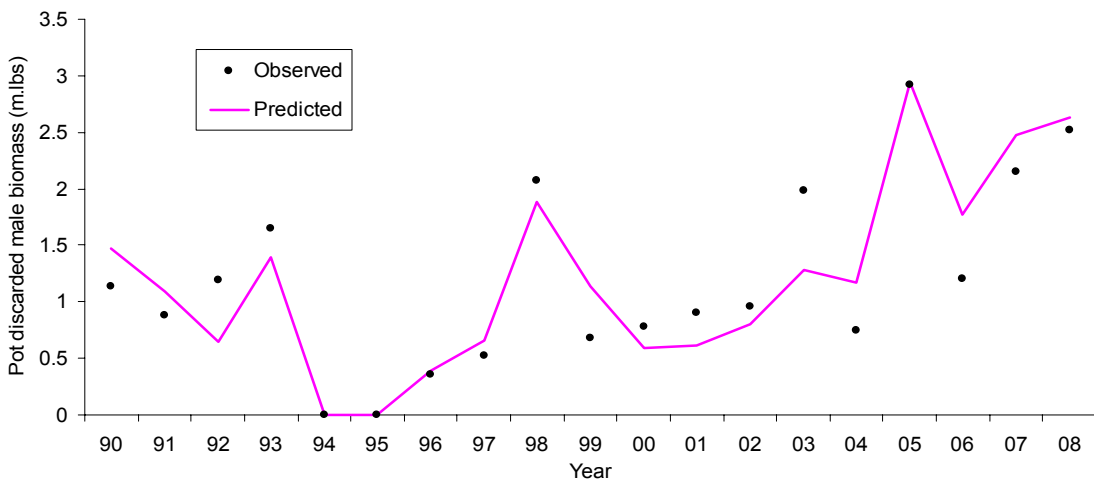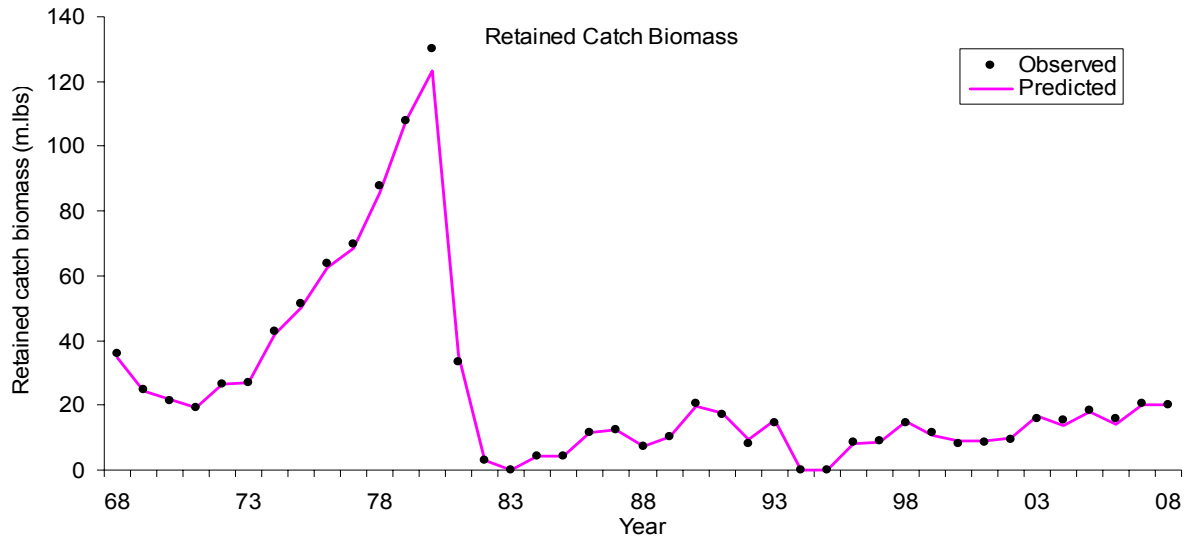
Figure 11a. Observed and predicted catch mortality biomass. Mortality biomass is equal to caught biomass times a handling mortality rate. Pot handling mortality rate is 0.2.

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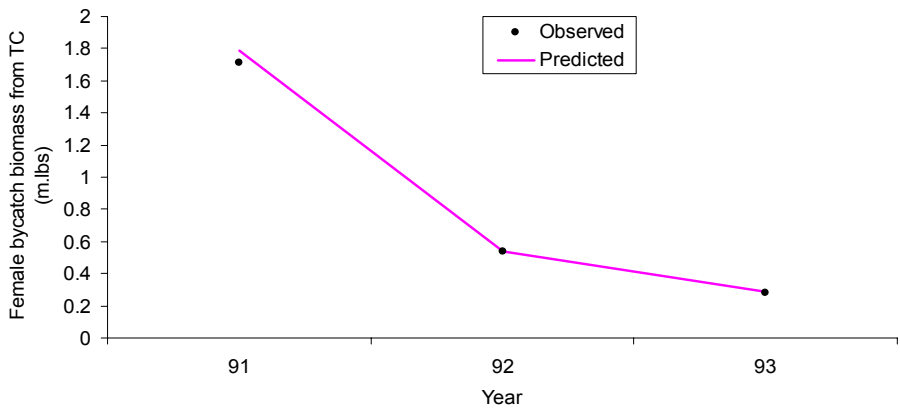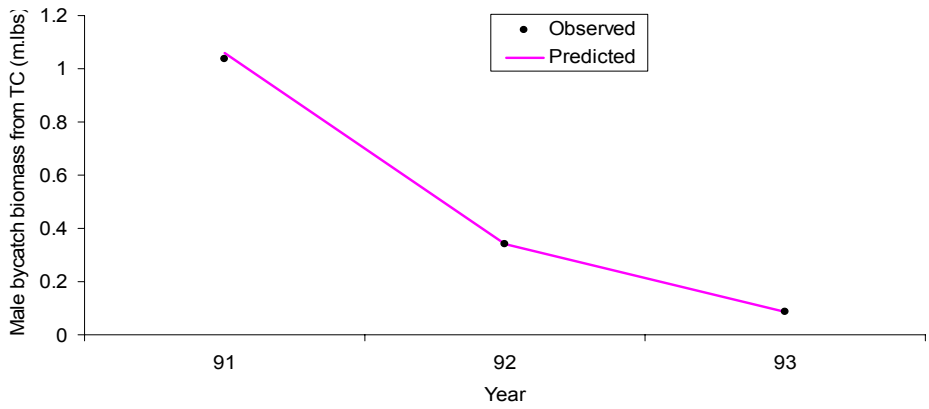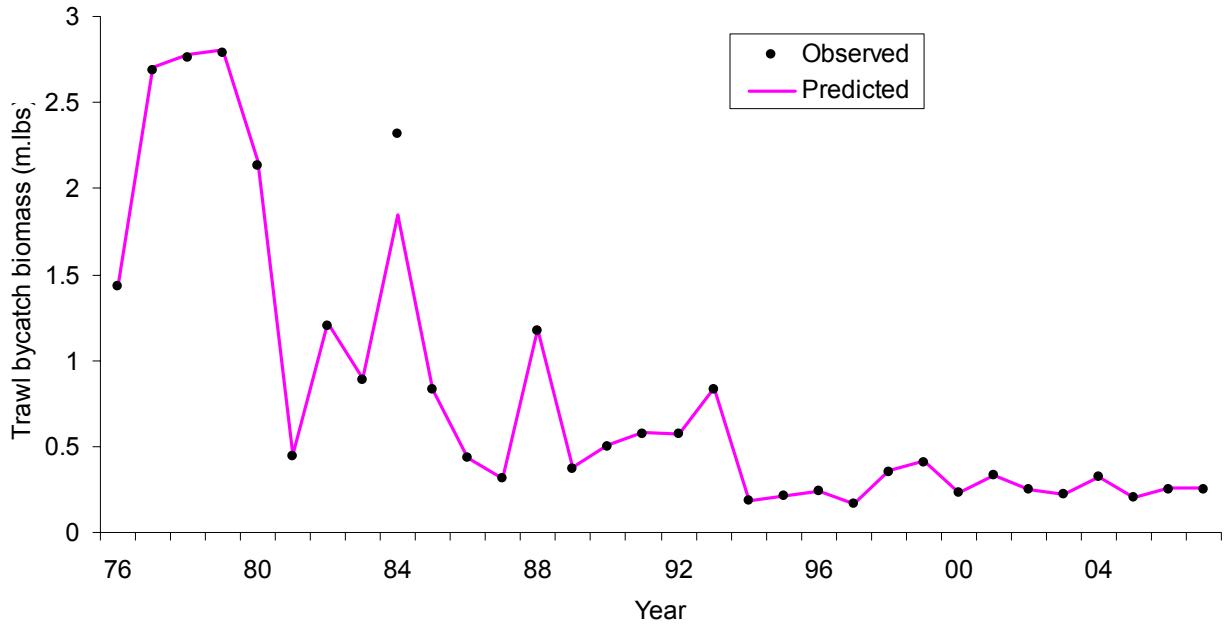
Figure 11(b). Observed and predicted catch mortality biomass from trawl fisheries and Tanner crab fisehry. Mortality biomass is equal to caught biomass times a handling mortality rate. Trawl handling mortality rate is 0.8, and Tanner crab pot handling mortality is 0.25.
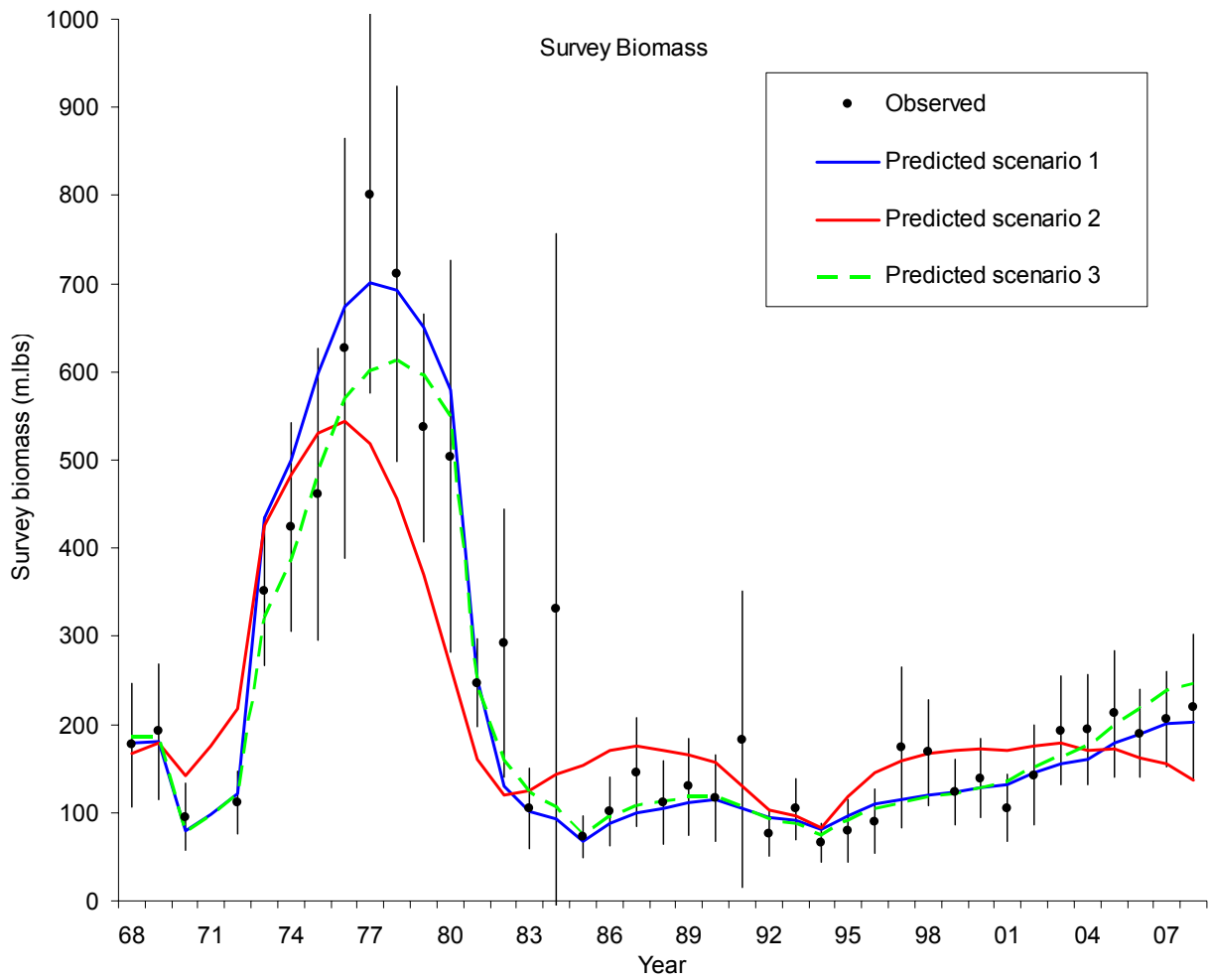
Figure 12a. Comparisons of area-swept estimates of total survey biomass and model prediction for three scenarios. Pot handling mortality rate is 0.2.
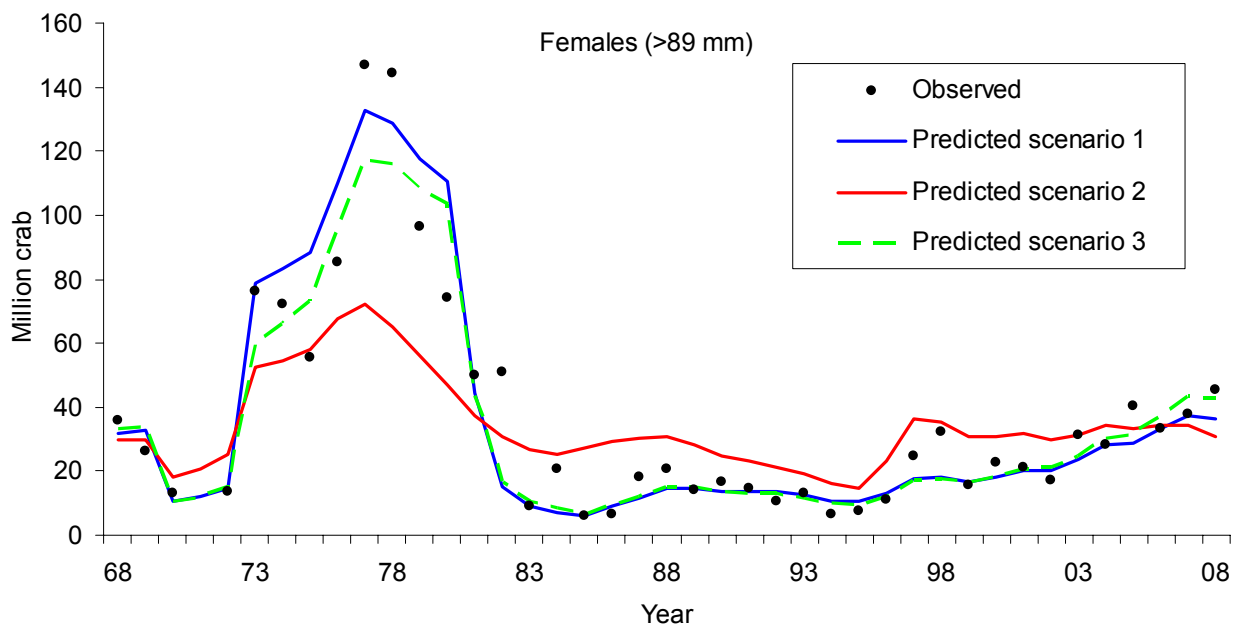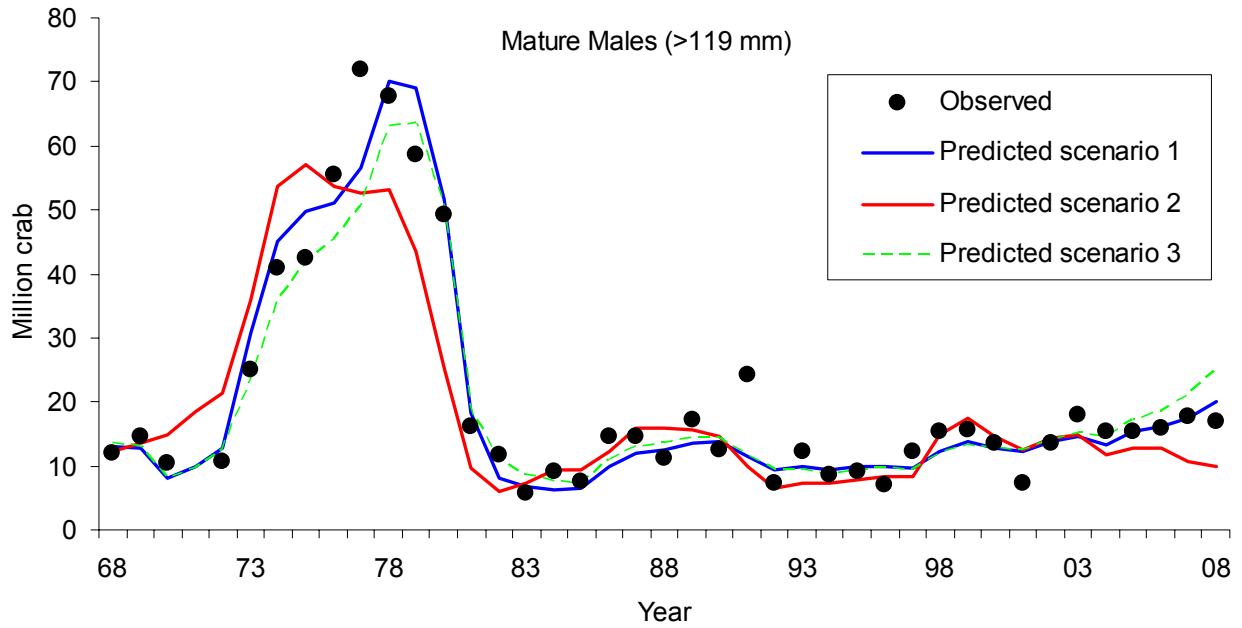
Figure 12b. Comparisons of area-swept estimates of mature male (>119 mm) and female (>89 mm) abundance and model prediction for three scenarios. Pot handling mortality rate is 0.2.
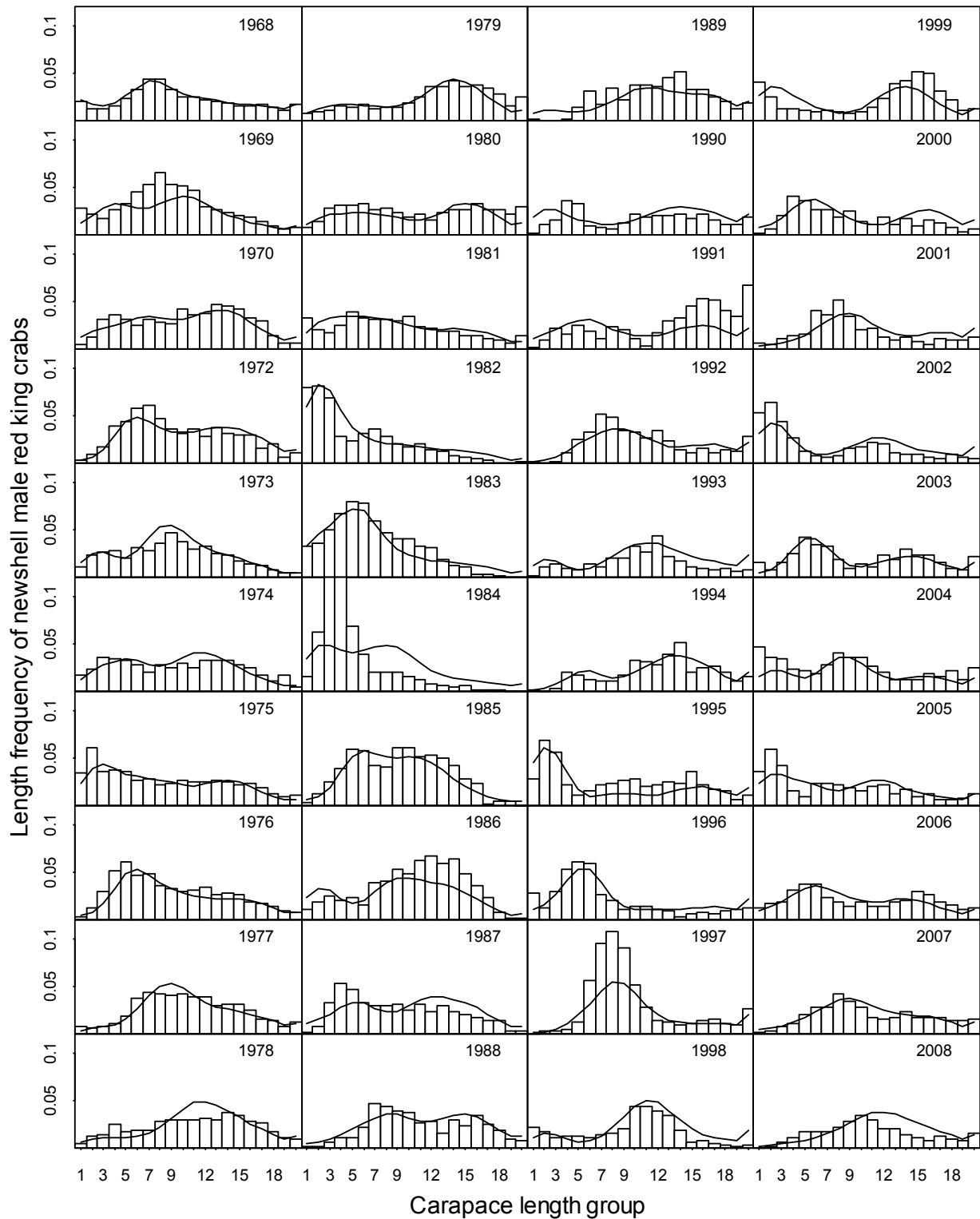
Figure 13. Comparison of area-swept and model estimated survey length frequencies of Bristol Bay all-shell (before 1986) and newshell (1986-2008) male red king crabs by year. Pot handling mortality rate is 0.2, and the first length group is 67.5 mm.

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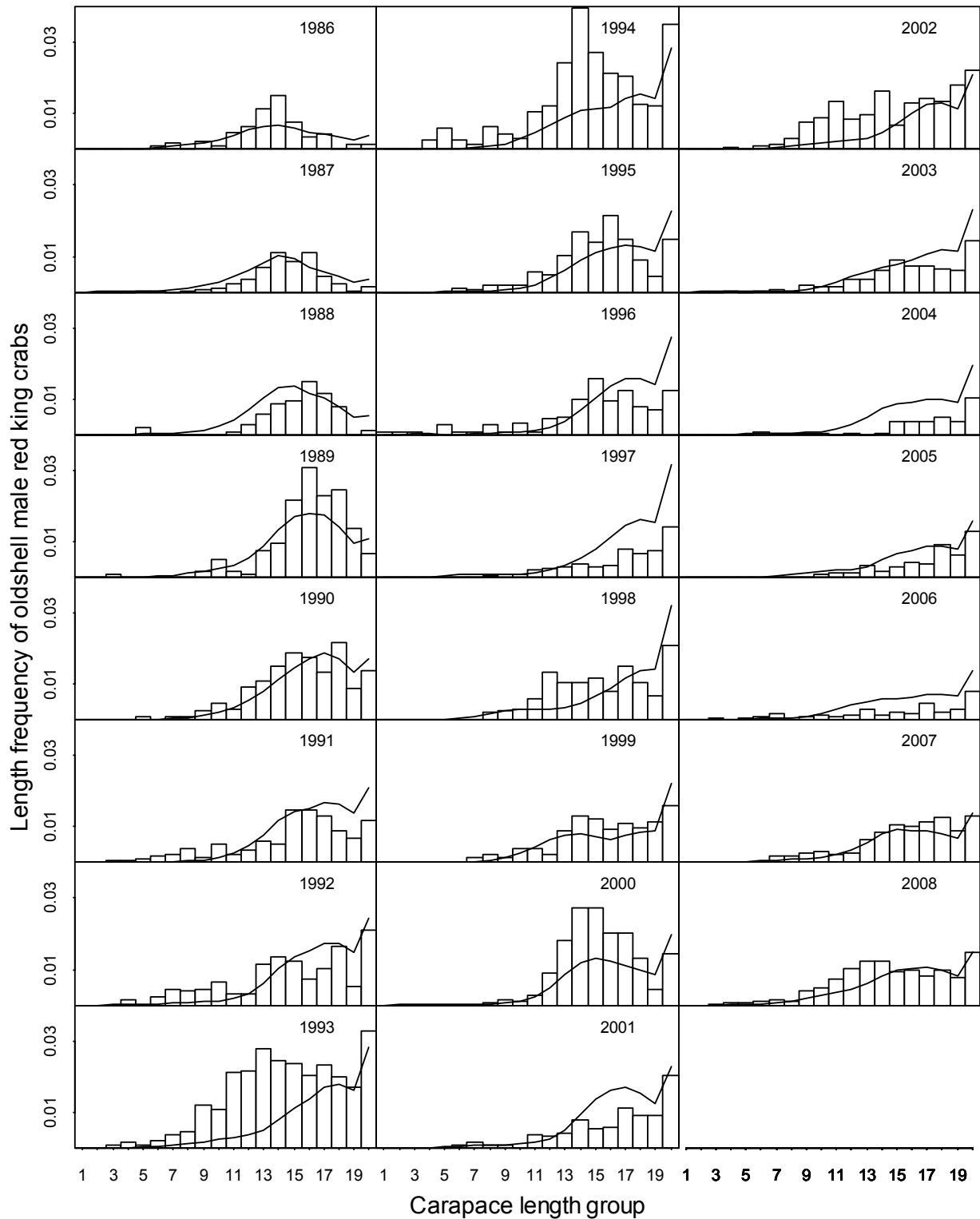
Figure 14. Comparison of area-swept and model estimated survey length frequencies of Bristol Bay oldshell male red king crabs by year. Pot handling mortality rate is 0.2, and the first length group is 67.5 mm.

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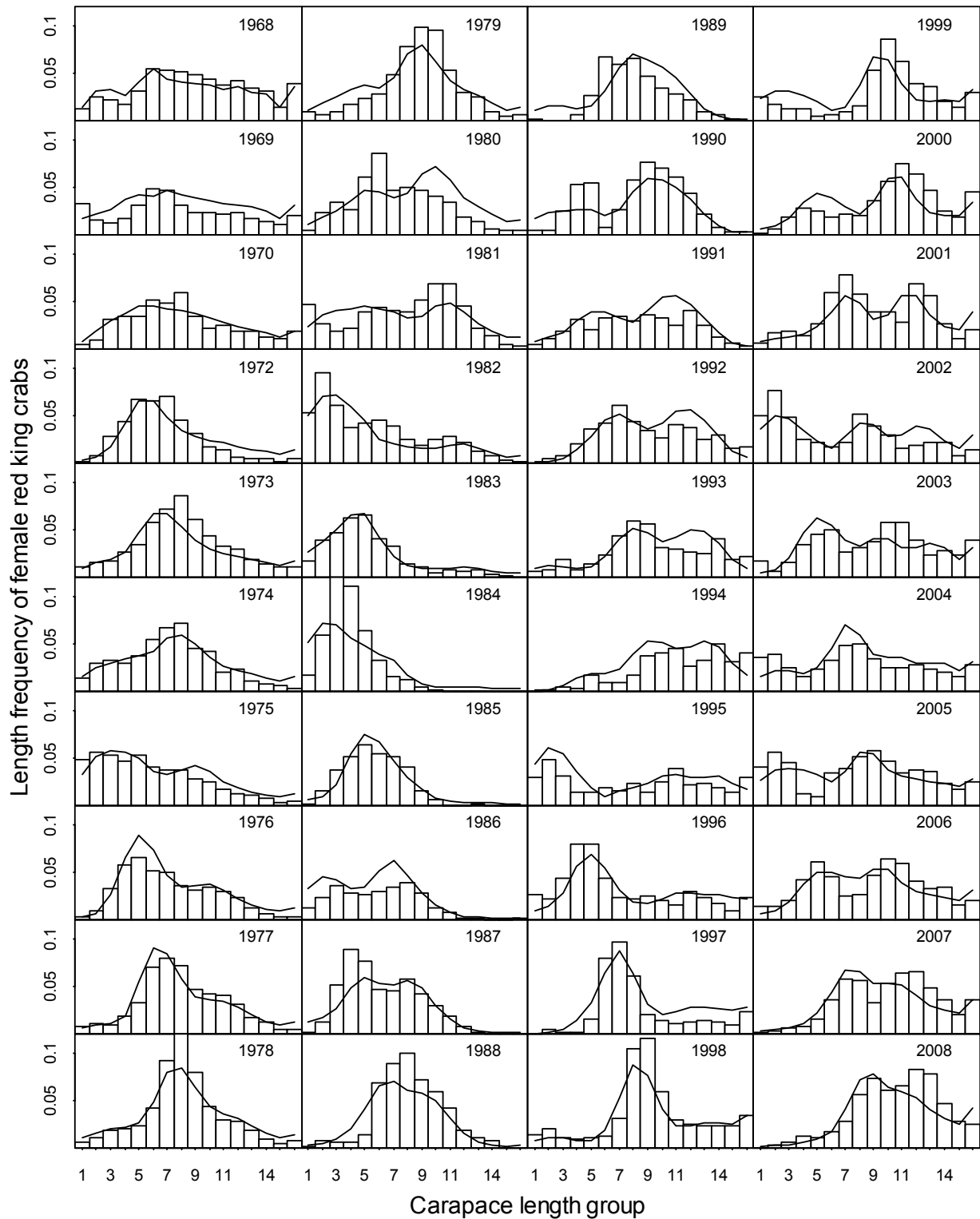
Figure 15. Comparison of area-swept and model estimated survey length frequencies of Bristol Bay female red king crabs by year. Pot handling mortality rate is 0.2, and the first length group is 67.5 mm.

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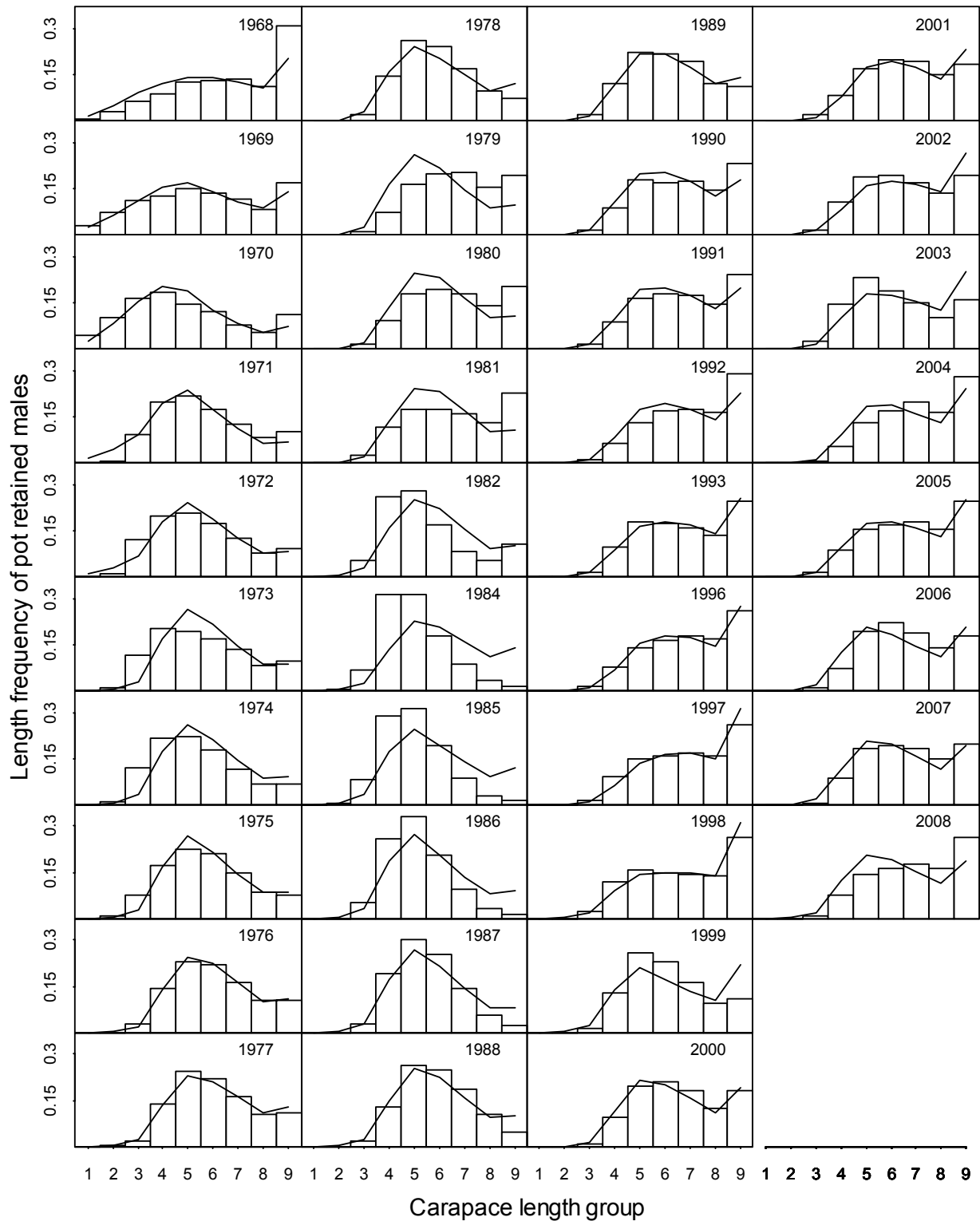
Figure 16. Comparison of observed and model estimated retained length frequencies of Bristol Bay male red king crabs by year in the directed pot fishery. Pot handling mortality rate is 0.2, and the first length group is 122.5 mm.

74
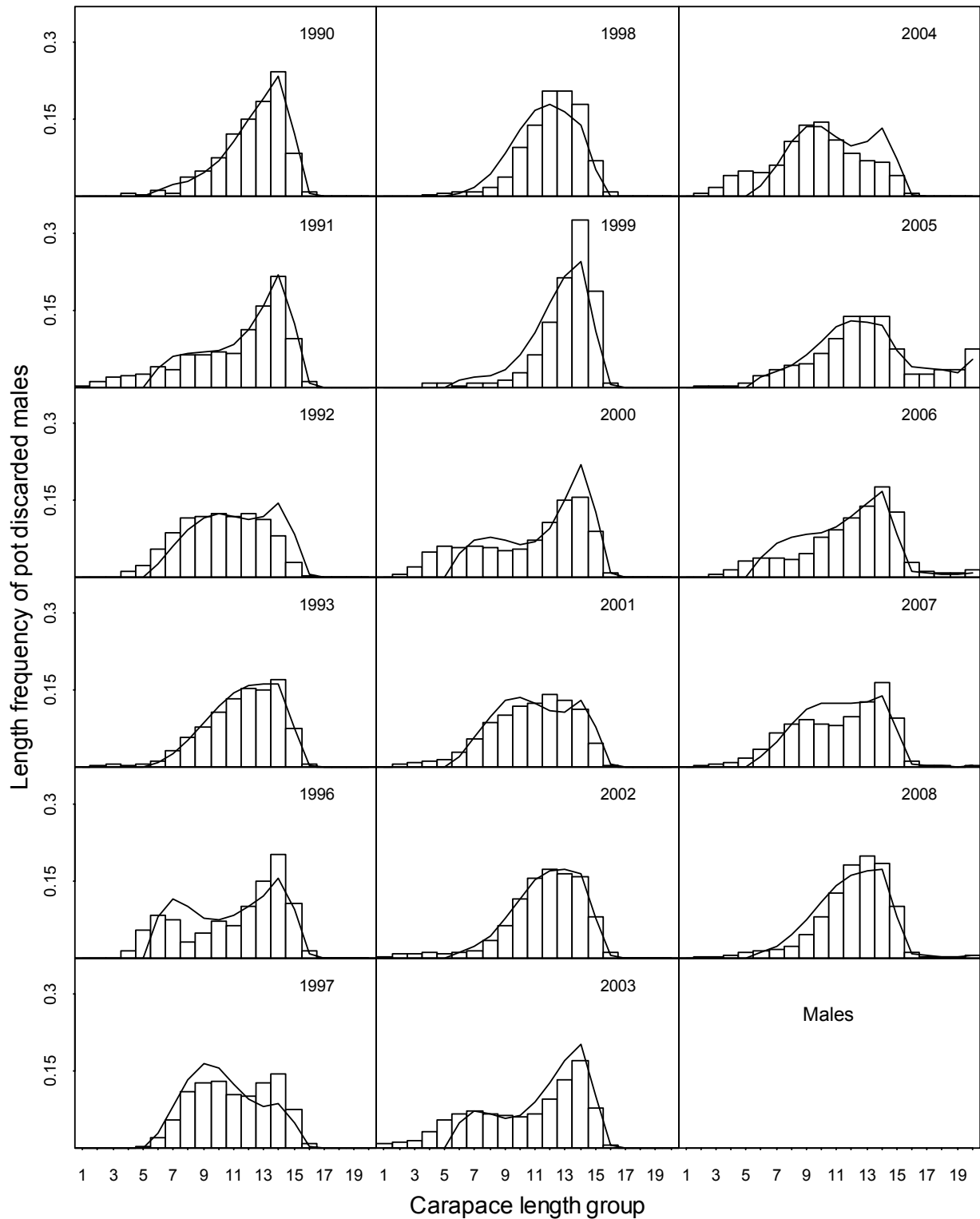
Figure 17. Comparison of observer and model estimated discarded length frequencies of Bristol Bay male red king crabs by year in the directed pot fishery. Pot handling mortality rate is 0.2, and the first length group is 67.5 mm.
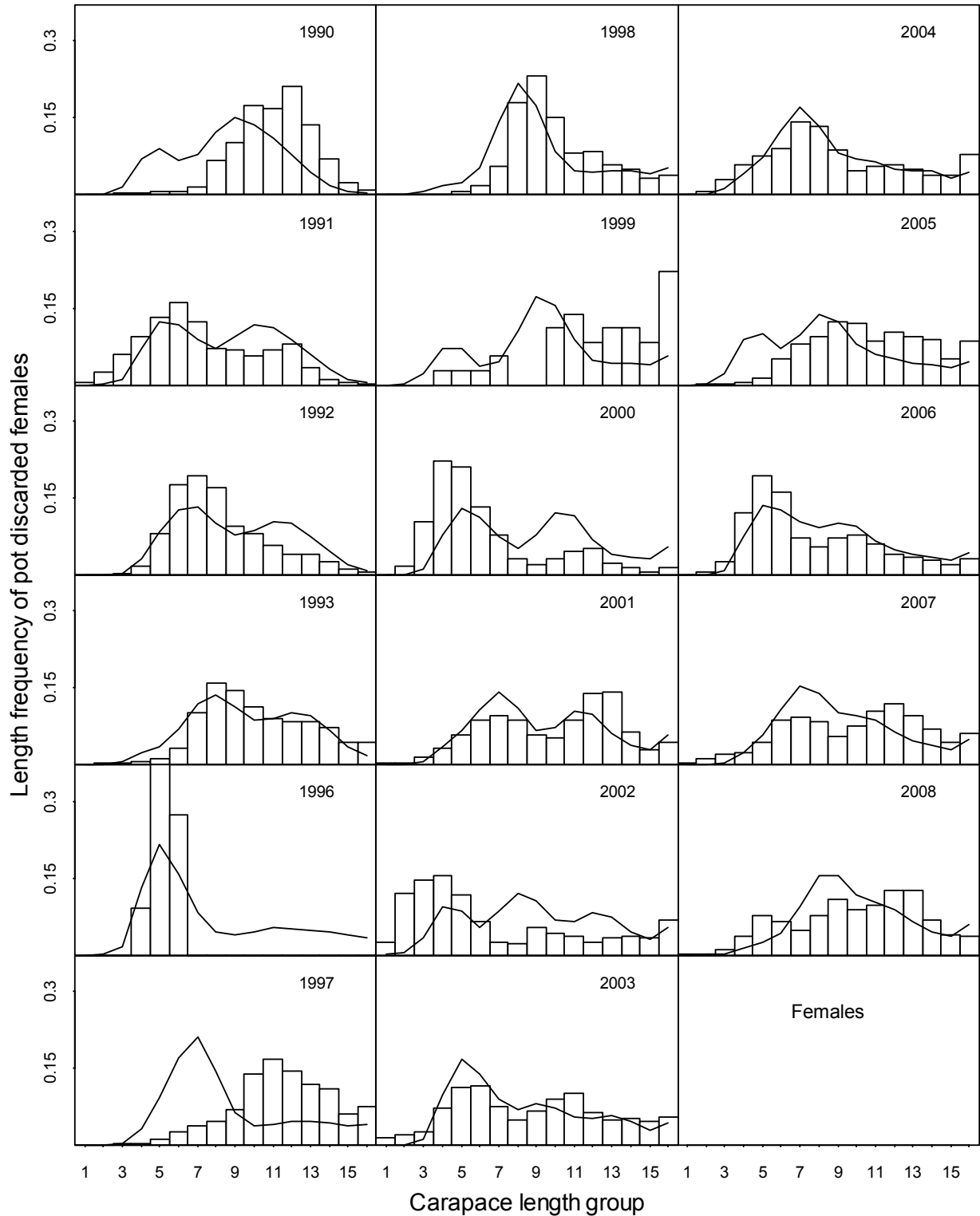
Figure 18. Comparison of observer and model estimated discarded length frequencies of Bristol Bay female red king crabs by year in the directed pot fishery. Pot handling mortality rate is 0.2, and the first length group is 67.5 mm.

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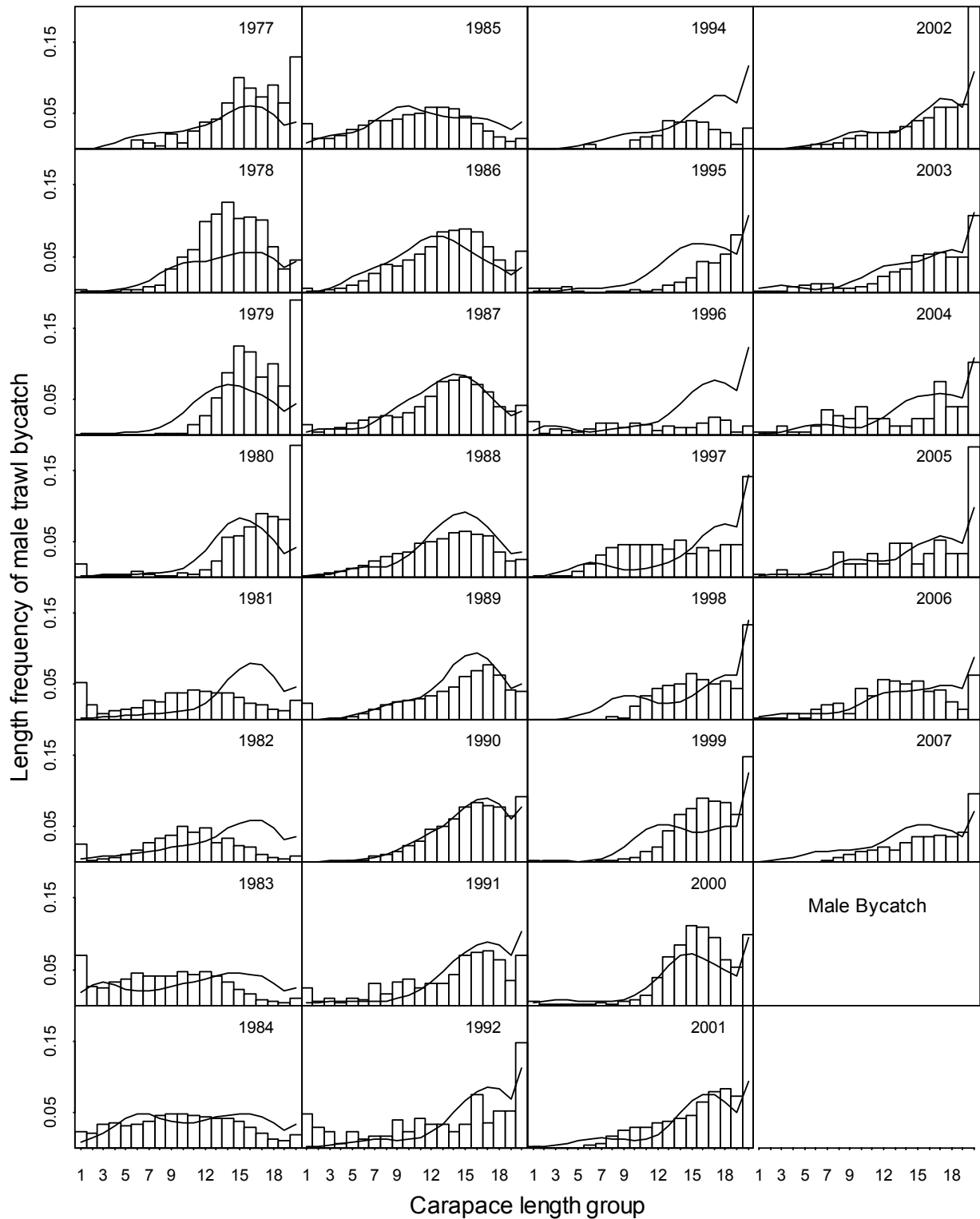
Figure 19. Comparison of observer and model estimated discarded length frequencies of Bristol Bay male red king crabs by year in the groundfish trawl fisheries. Pot handling mortality rate is 0.2, and the first length group is 67.5 mm.
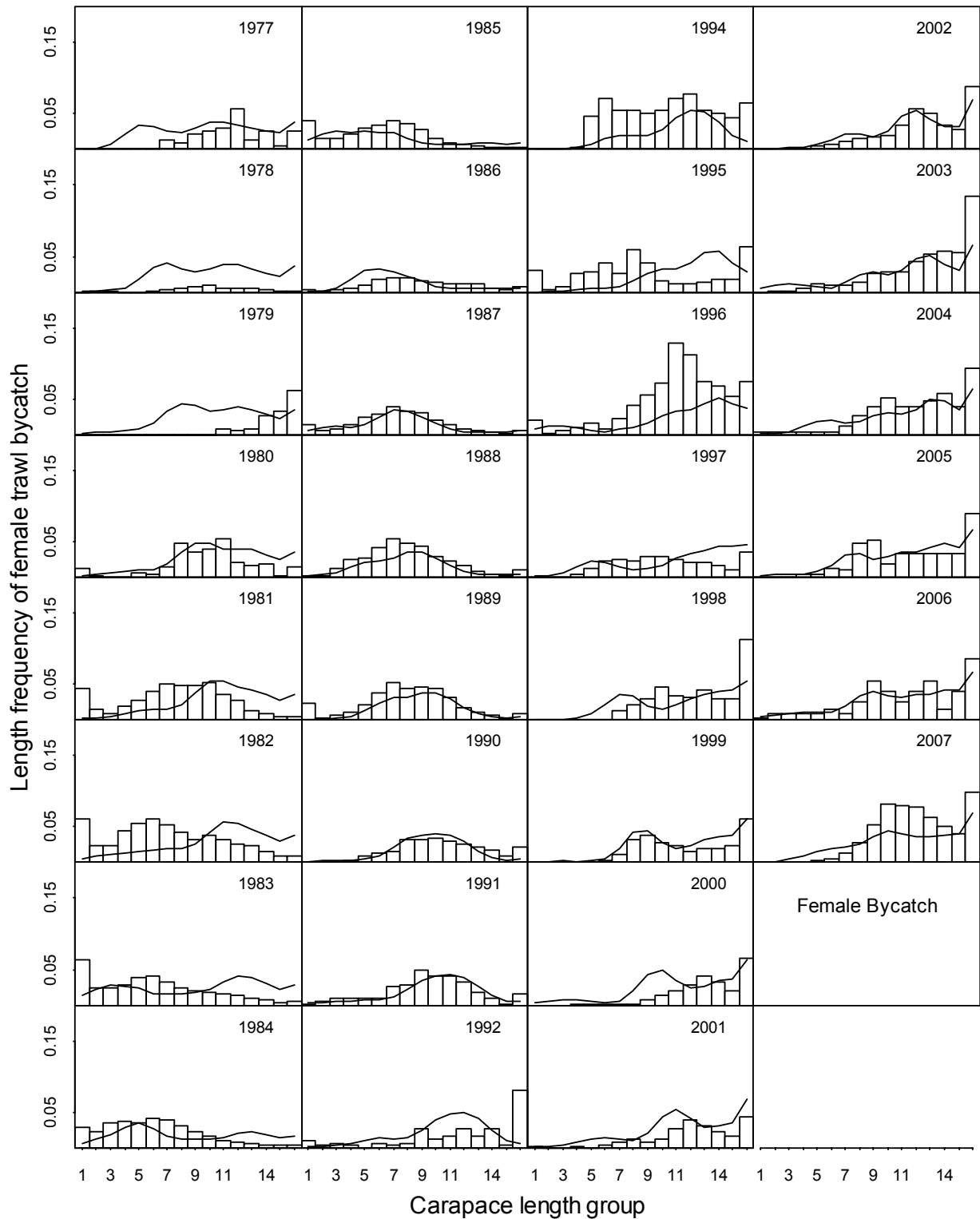
Figure 20. Comparison of observer and model estimated discarded length frequencies of Bristol Bay female red king crabs by year in the groundfish trawl fisheries. Pot handling mortality rate is 0.2, and the first length group is 67.5 mm.

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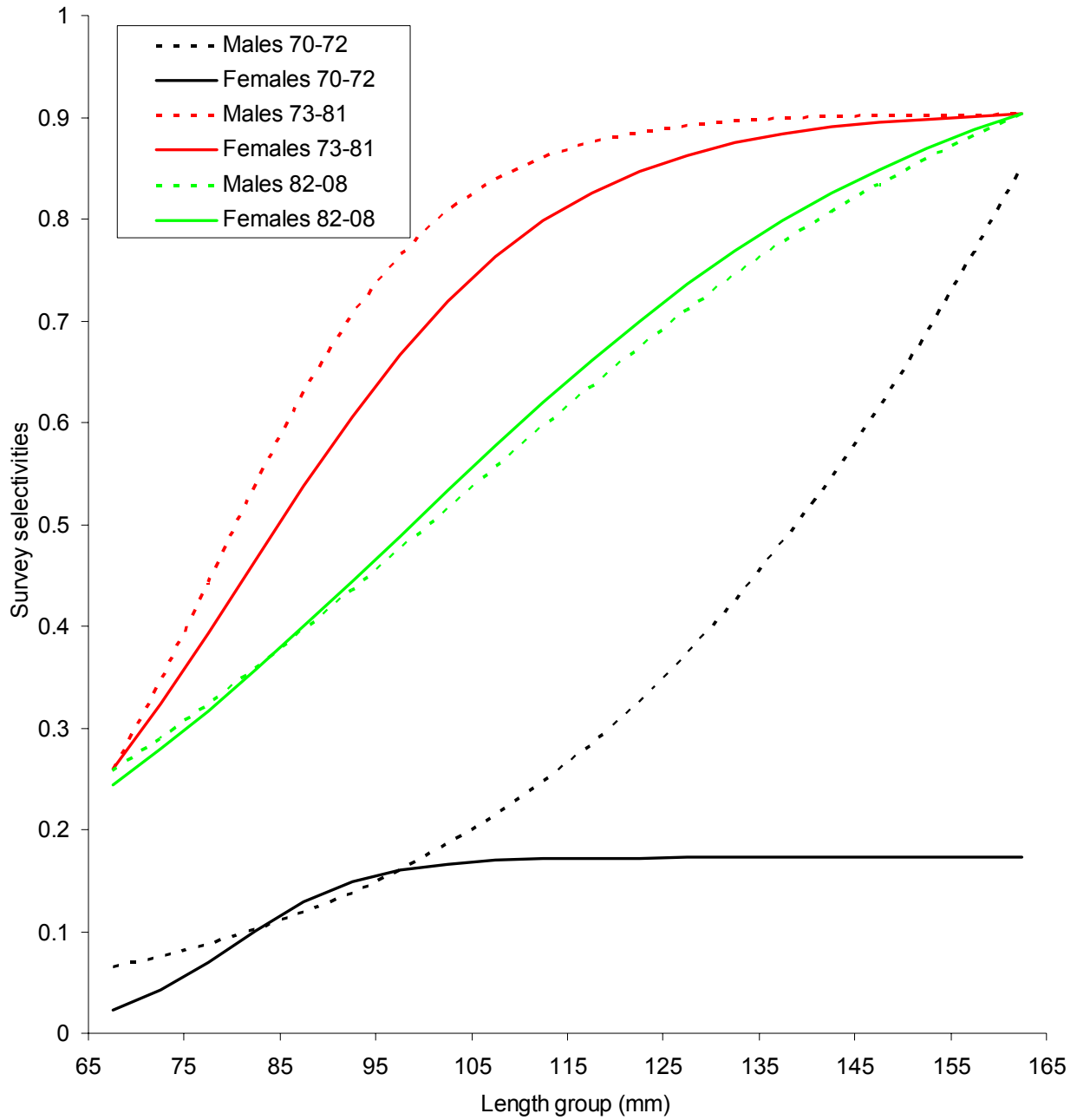
Figure 21a. Estimated trawl survey selectivities. Pot handling mortality rate is 0.2.
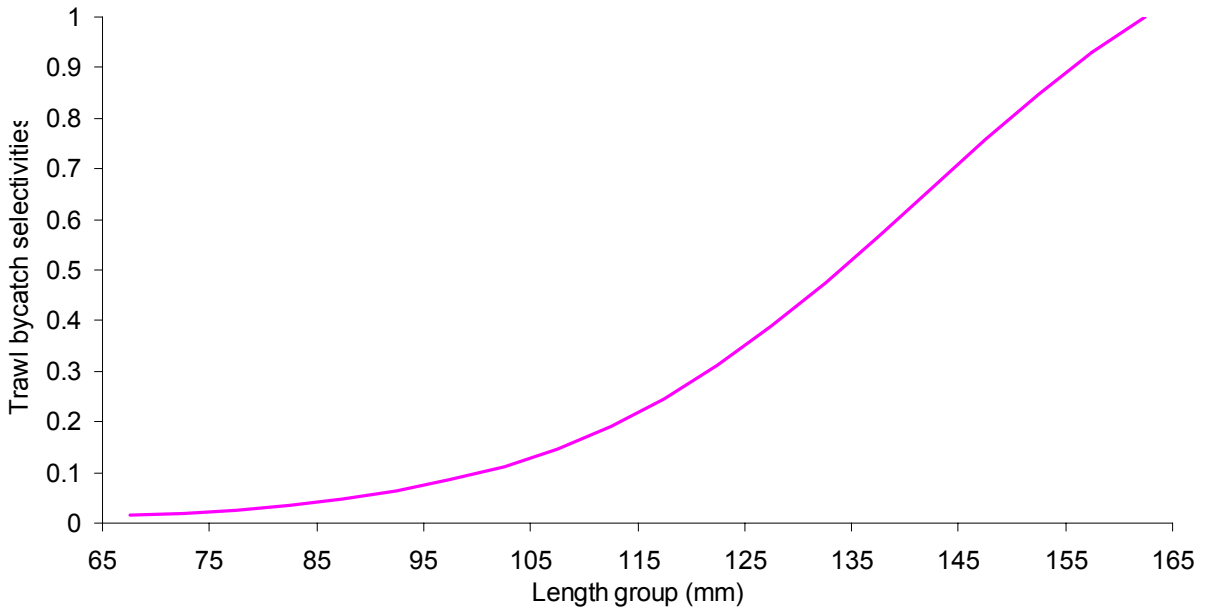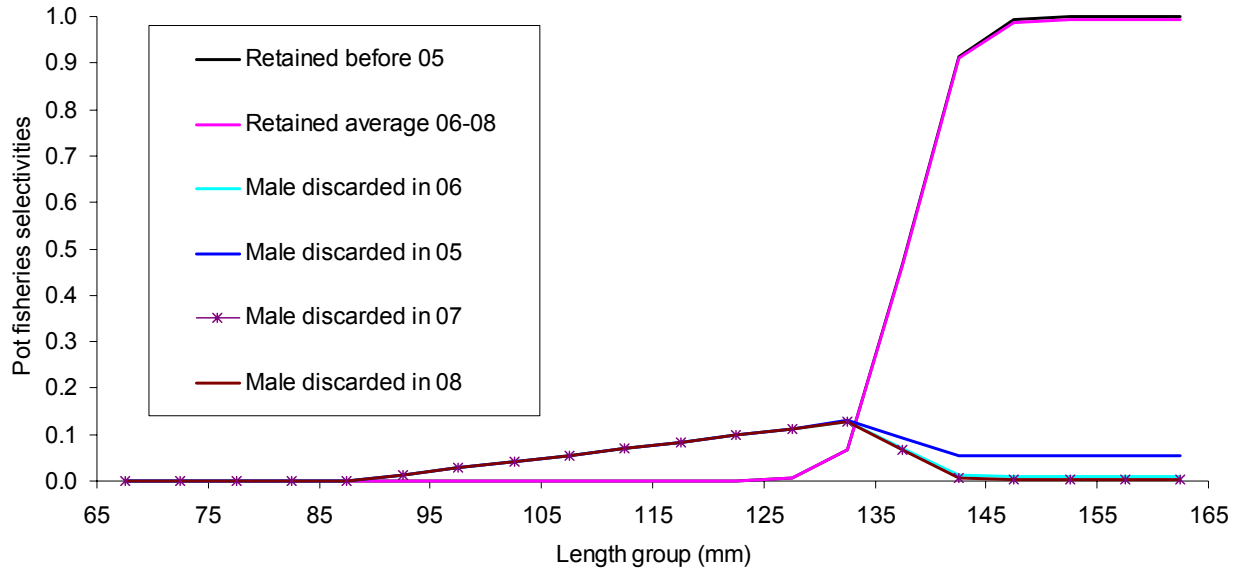
Figure 21b. Estimated pot fishery selectivities and groundfish trawl bycatch selectivities. Pot handling mortality rate is 0.2.
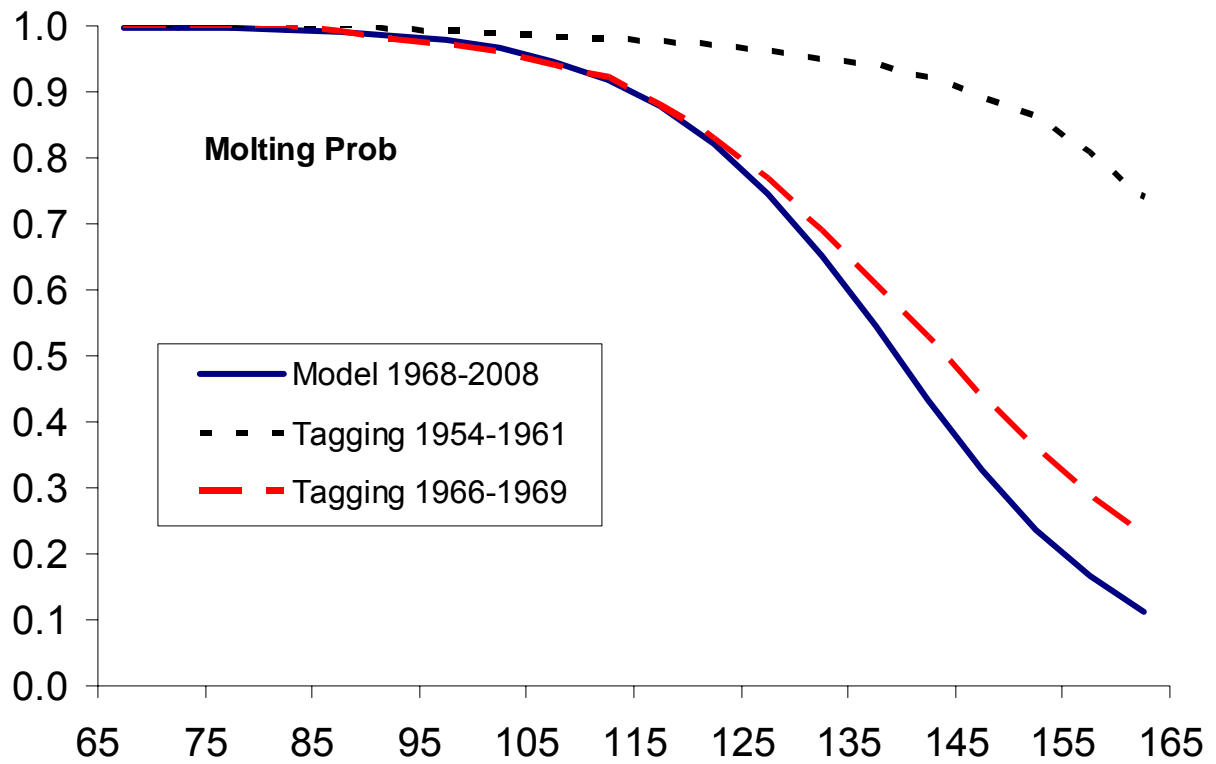
Figure 22. Comparison of estimated probabilities of molting of male red king crabs in Bristol Bay for different periods. Molting probabilities for periods 1954-1961 and 1966-1969 were estimated by Balsiger (1974) from tagging data. Molting probabilities for 1968-2008 were estimated with a length-based model with pot handling mortality rate to be 0.2.
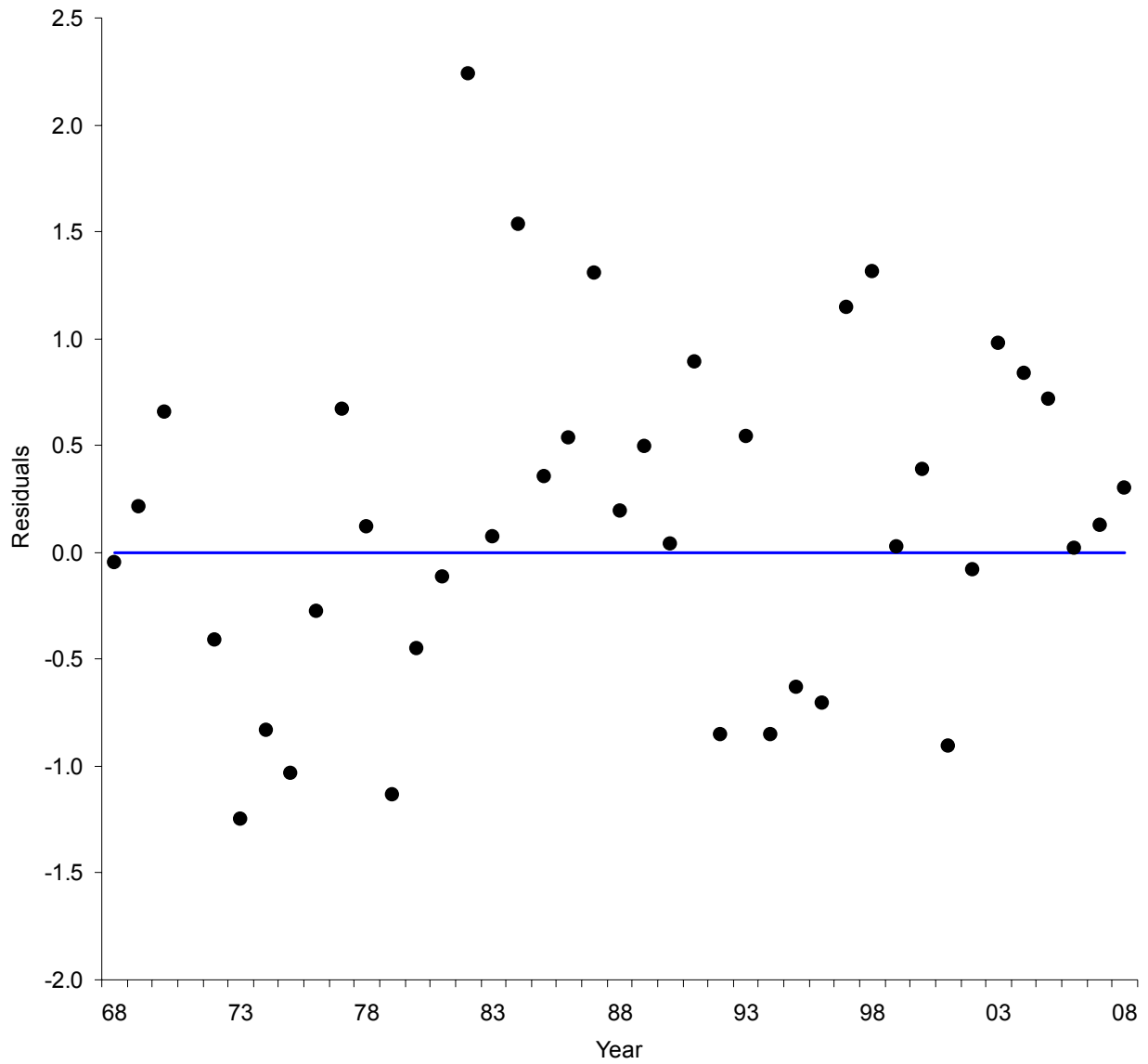
Figure 23. Standardized residuals of total survey biomass. Pot handling mortality rate is 0.2.
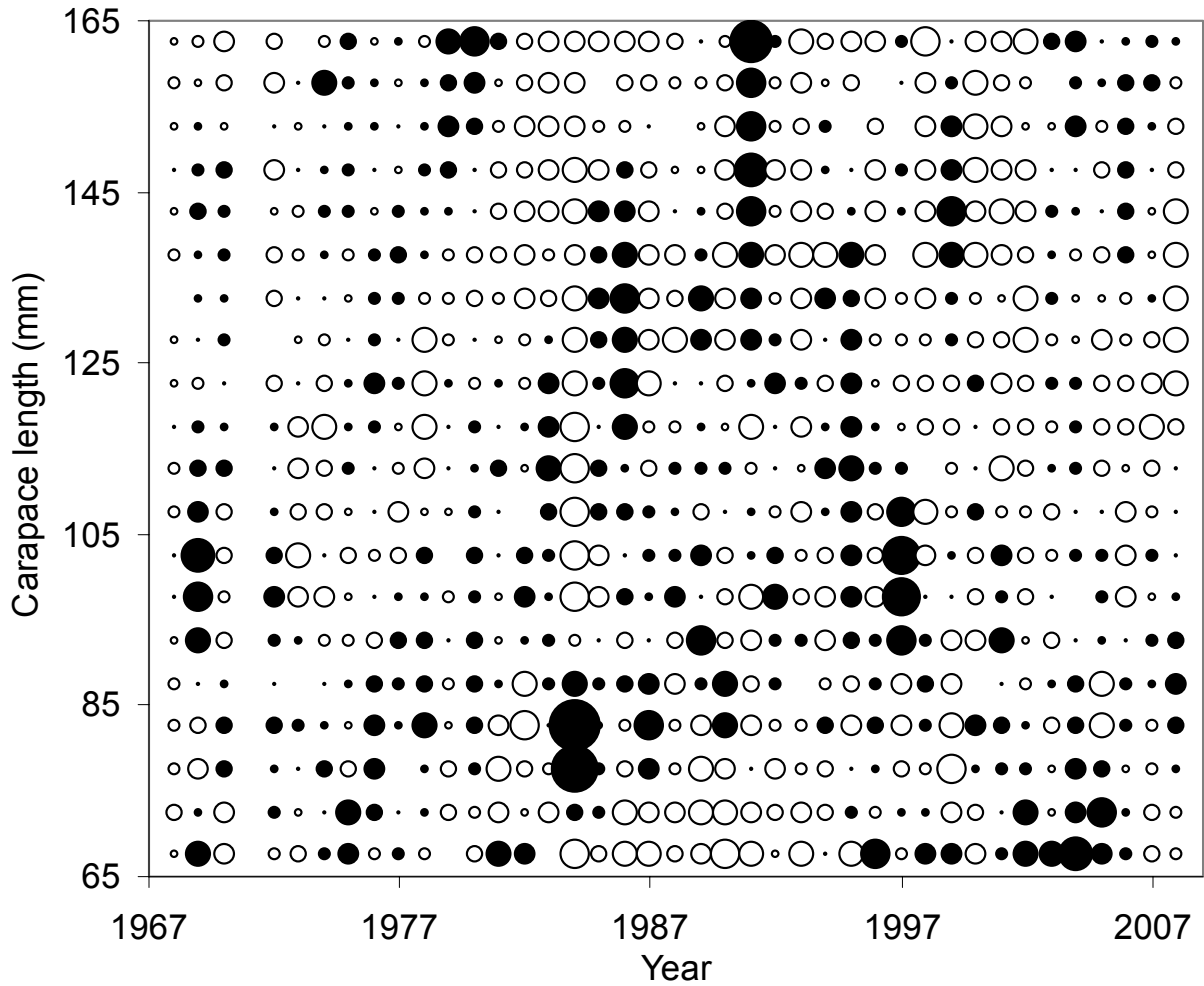
Figure 24. Standardized residuals of proportions of survey all-shell (1968-1985) and newshell (1986-2008) male red king crabs. Solid circles are positive residuals, and open circles are negative residuals. Pot handling mortality rate is 0.2.
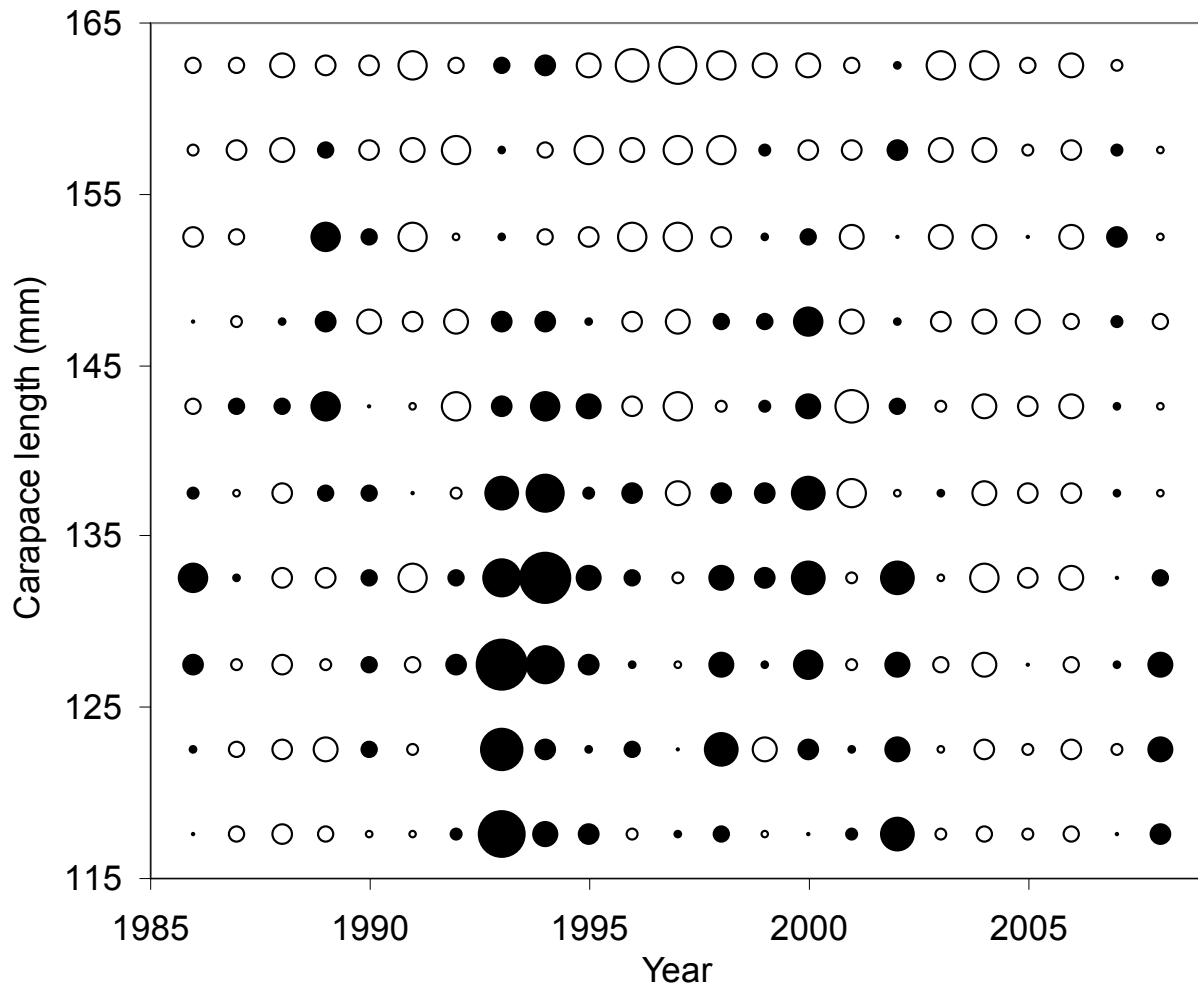
Figure 25. Standardized residuals of proportions of survey oldshell male red king crabs. Solid circles are positive residuals, and open circles are negative residuals. Pot handling mortality rate is 0.2.
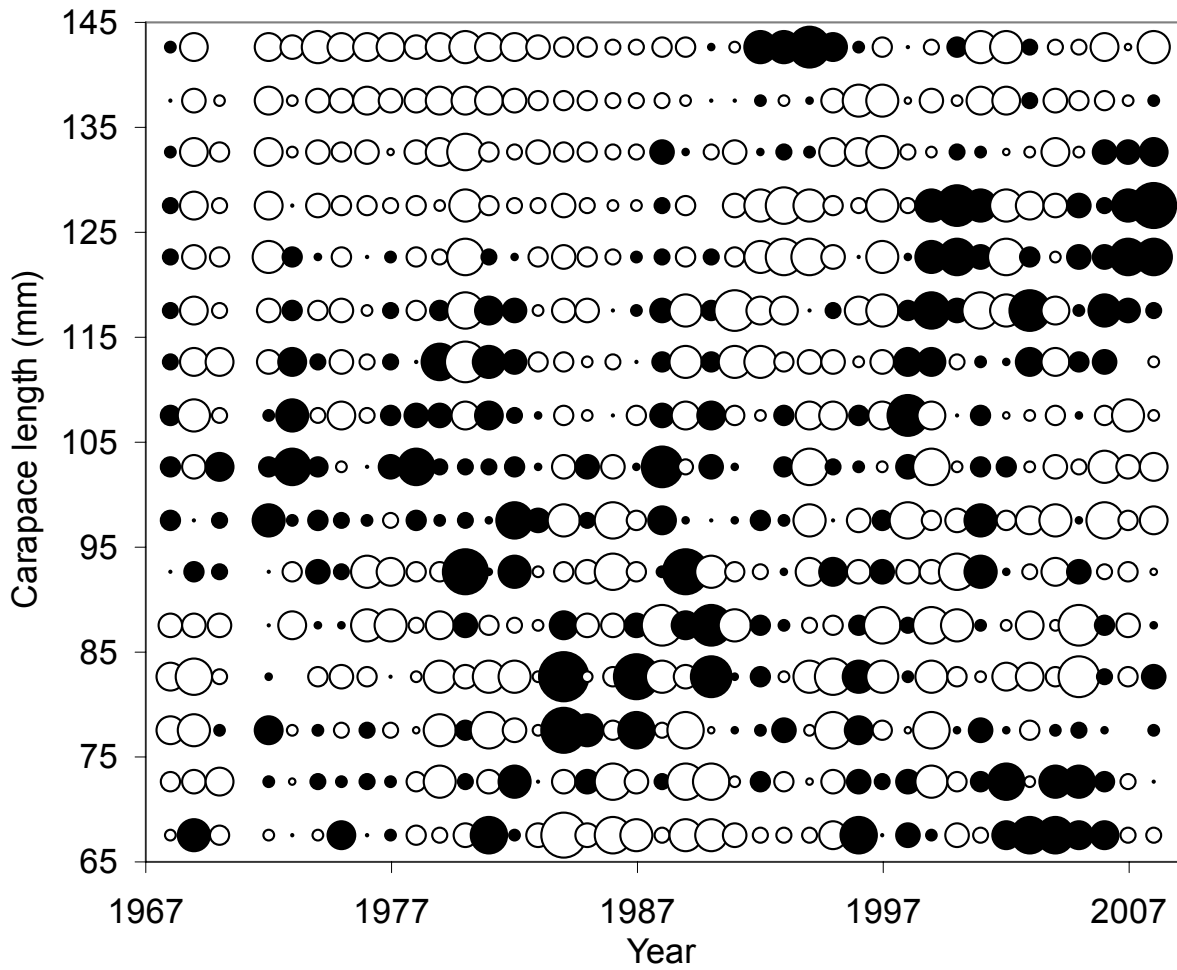
Figure 26. Standardized residuals of proportions of survey female red king crabs. Solid circles are positive residuals, and open circles are negative residuals. Pot handling mortality rate is 0.2.
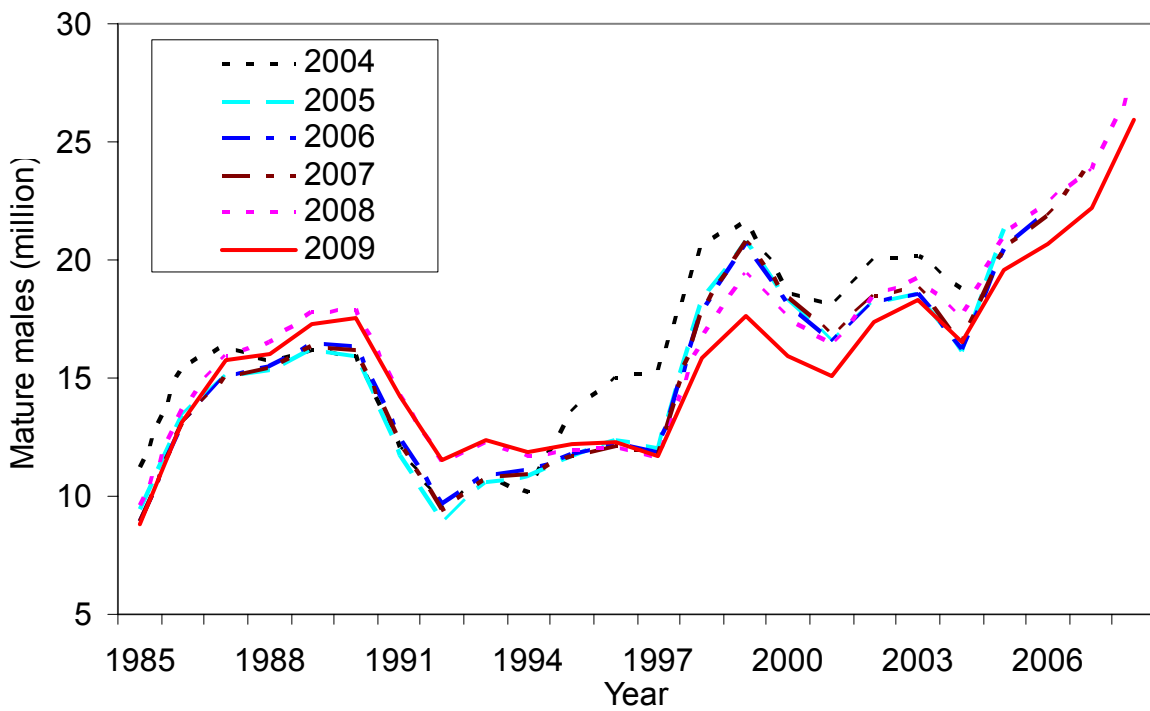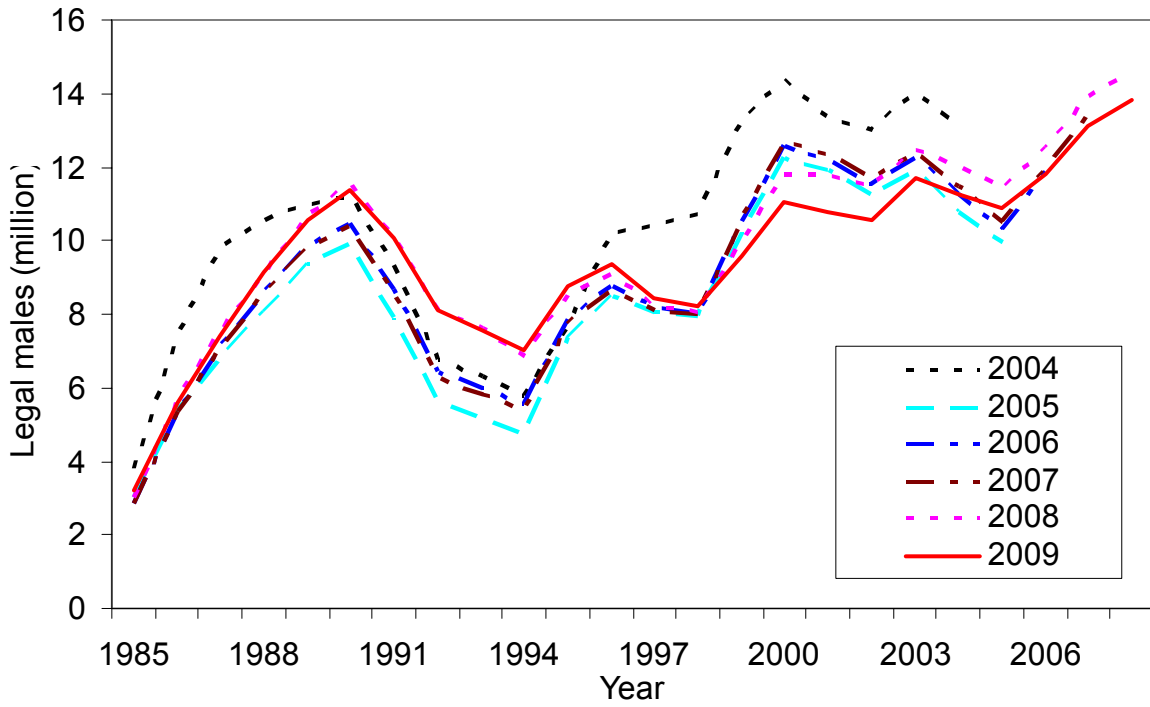
Figure 27. Comparison of estimates of legal male abundance (top) and mature males (bottom) of Bristol Bay red king crab from 1985 to 2008 made with terminal years 2004-2009. These are results of historical assessments. Legend shows the year in which the assessment was conducted. Pot handling mortality rate is 0.2.
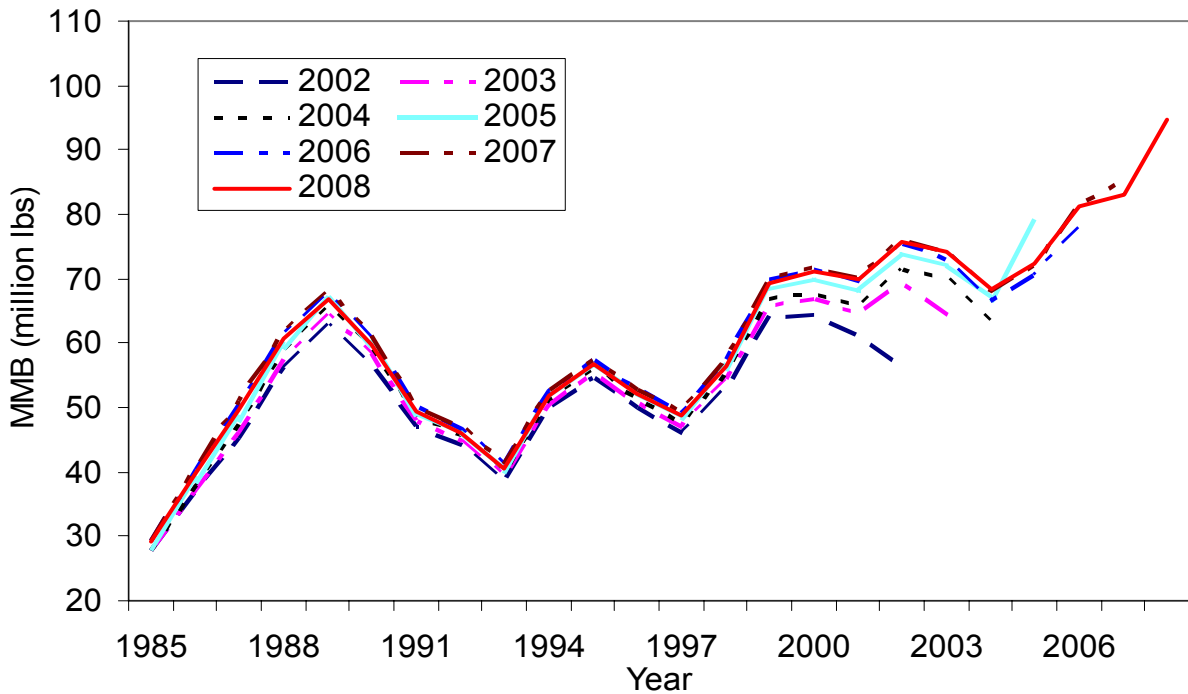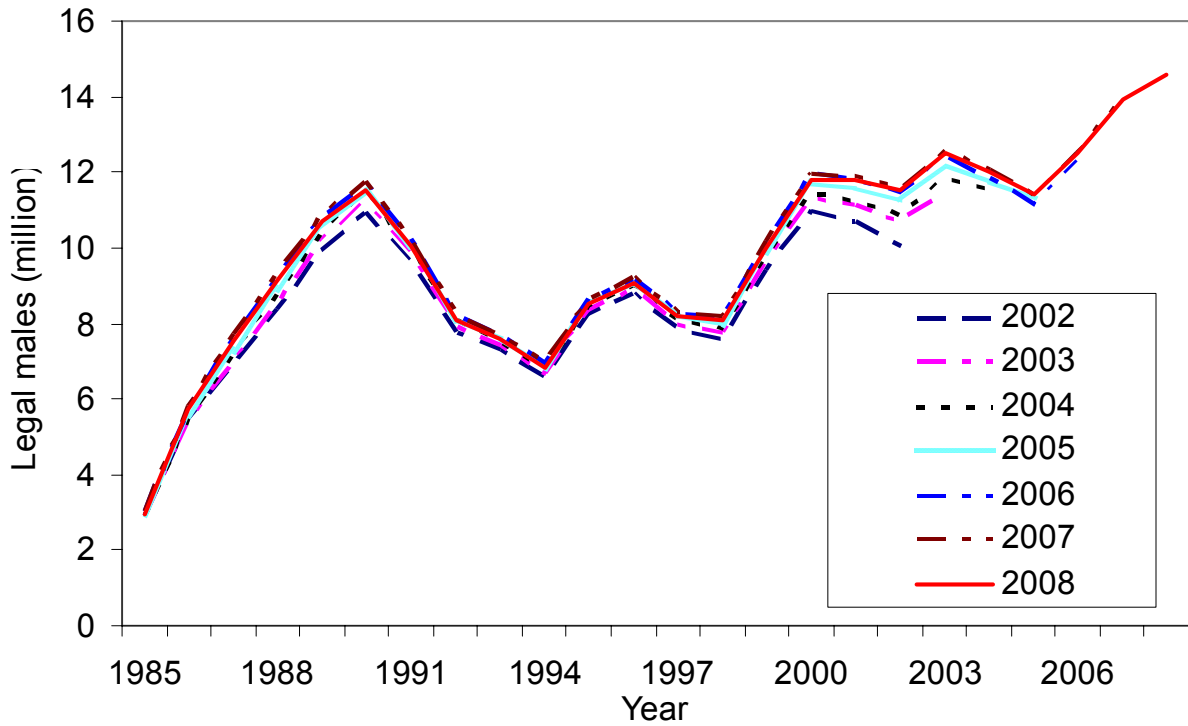
Figure 28. Comparison of estimates of legal male abundance (top) and mature male biomass (bottom) on Feb. 15 of Bristol Bay red king crab from 1985 to 2008 made with terminal years 2002-2008. These are results of the 2008 model. Legend shows the year in which the assessment was conducted. Pot handling mortality rate is 0.2. (These are results from the model in August 2008, which will be updated in August 2009)
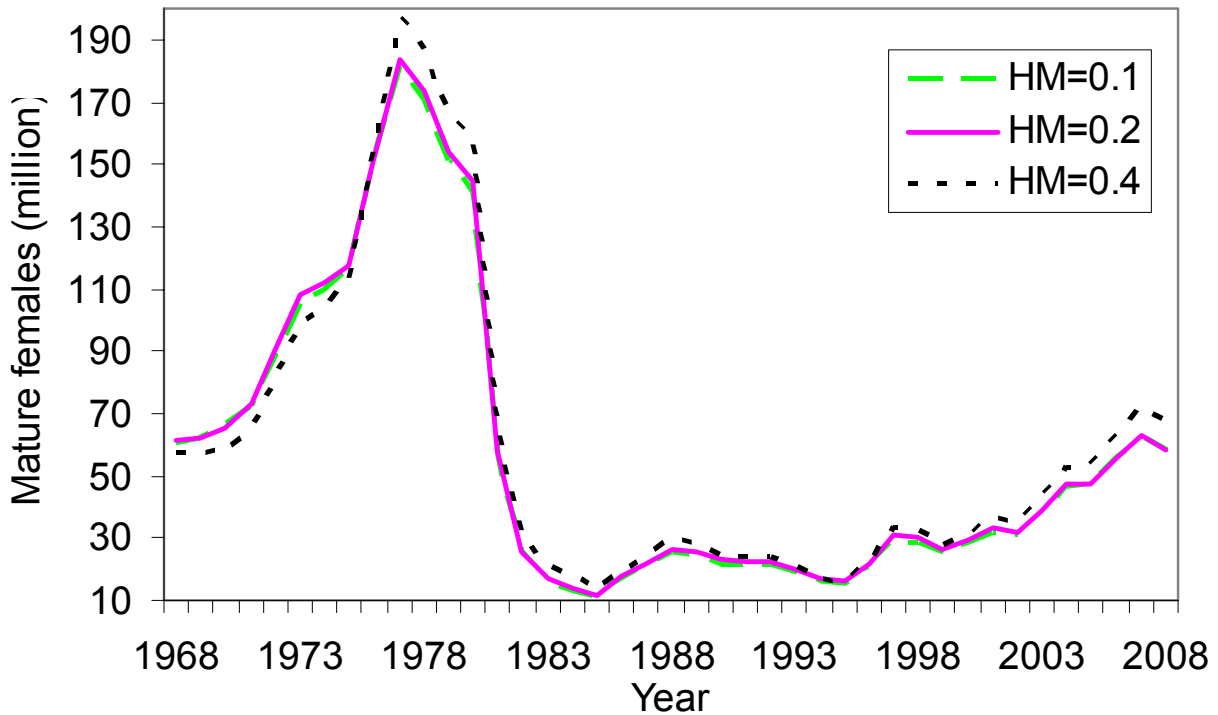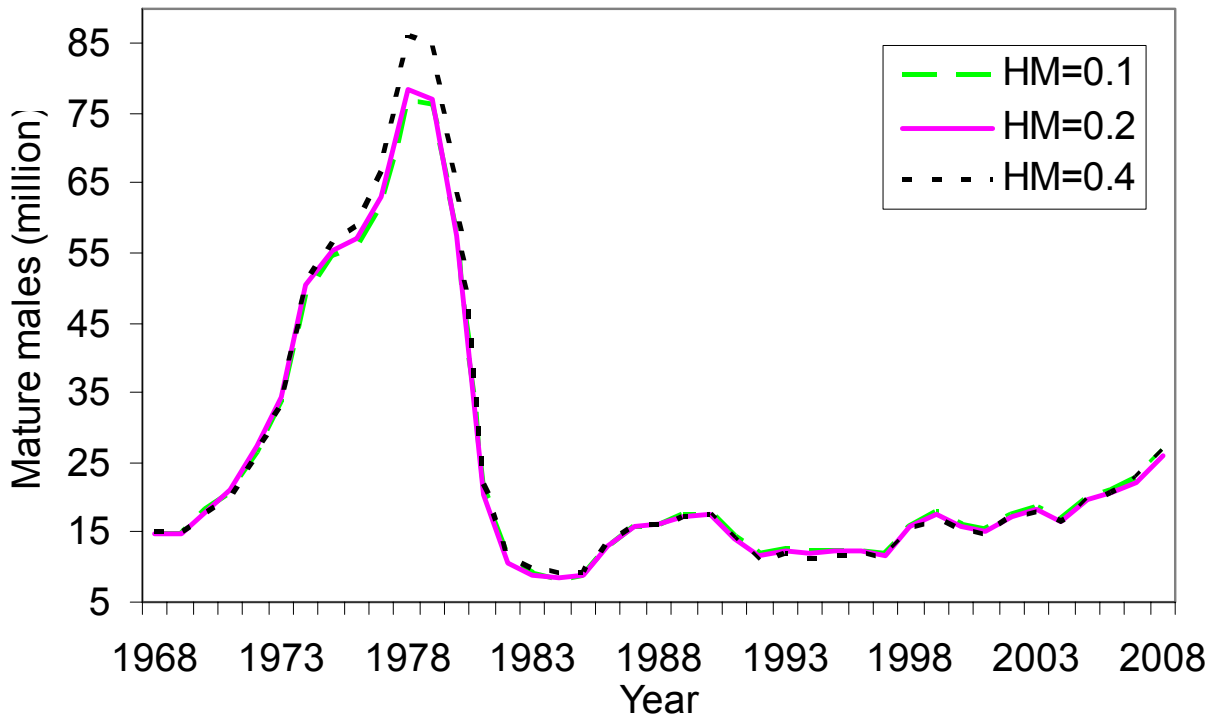
Figure 29. Comparison of mature abundance estimates for pot handling mortality rates of 0.1, 0.2 and 0.4. Mature females are for crabs >89 mm CL in this plot.
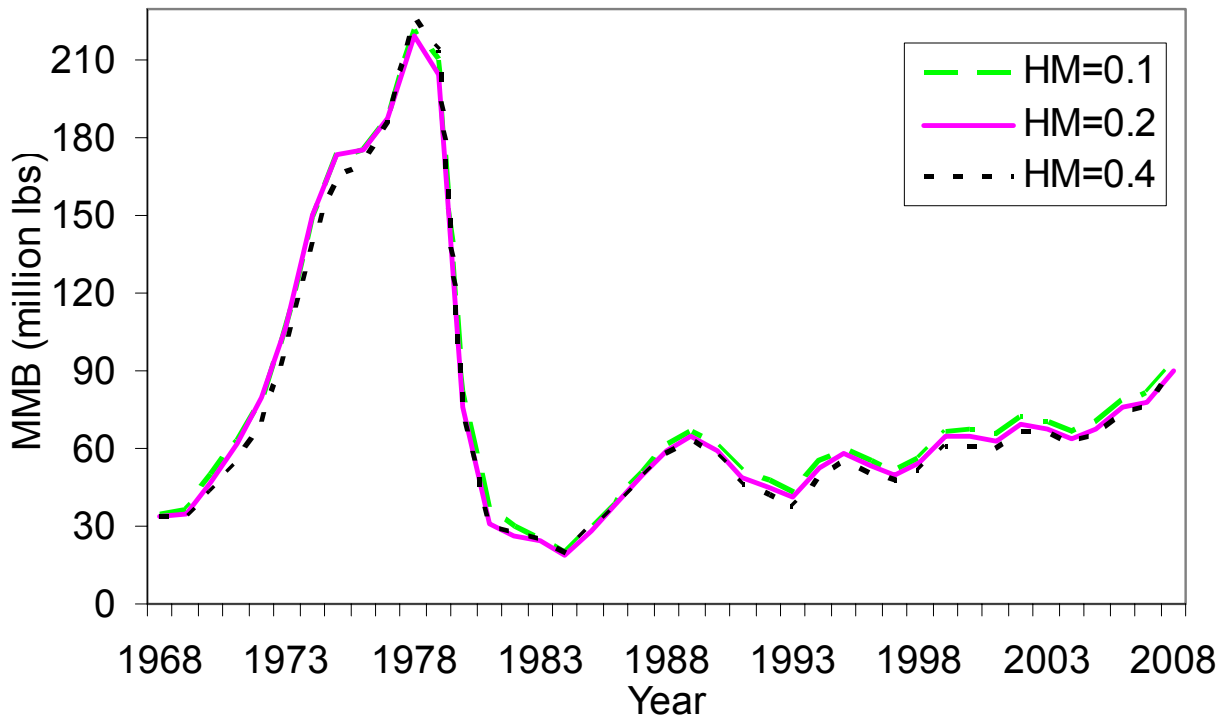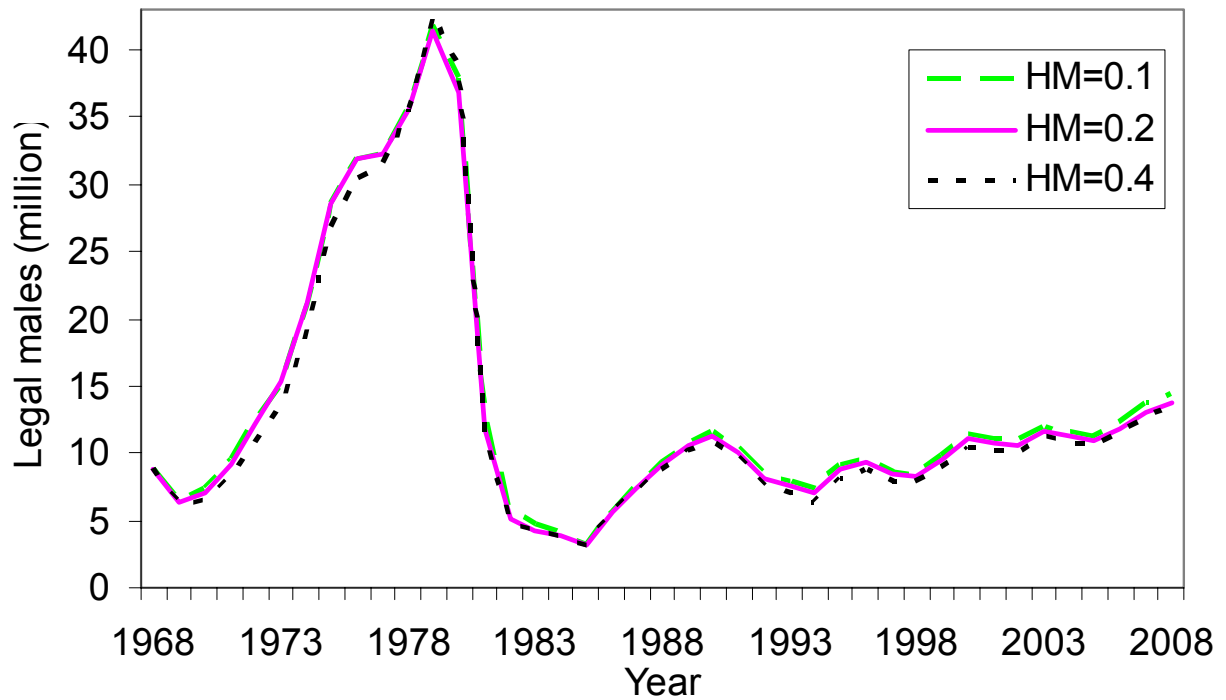
Figure 30. Comparison of legal male abundance estimates and mature male biomass on Feb. 15 for pot handling mortality rates of 0.1, 0.2 and 0.4.
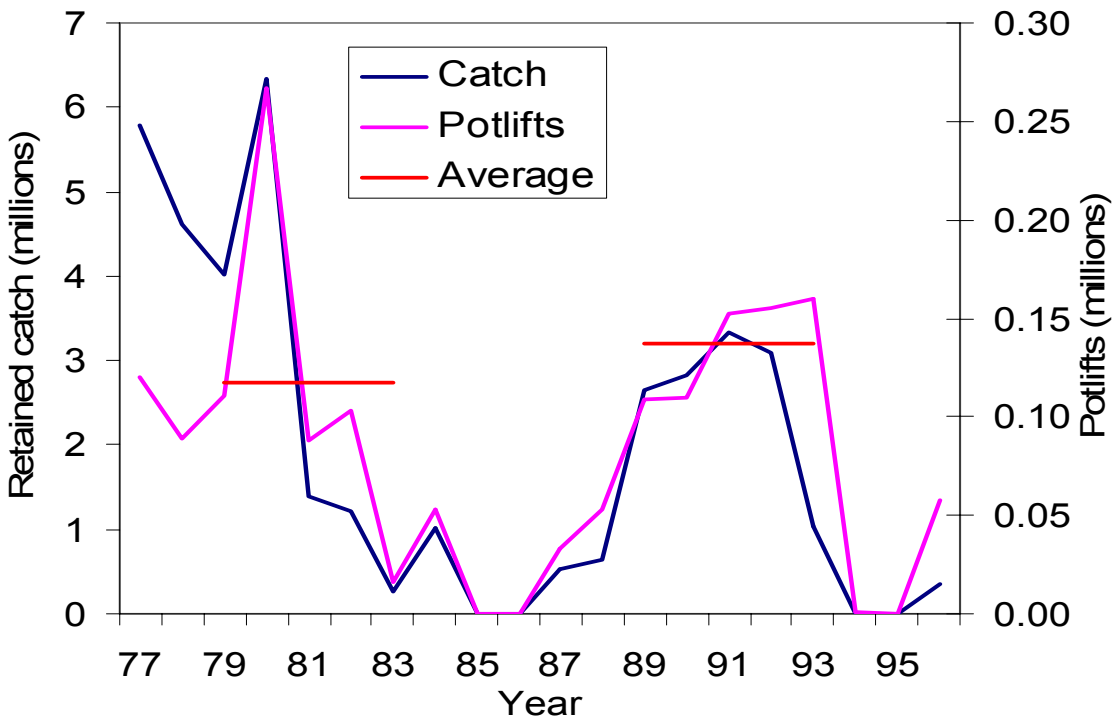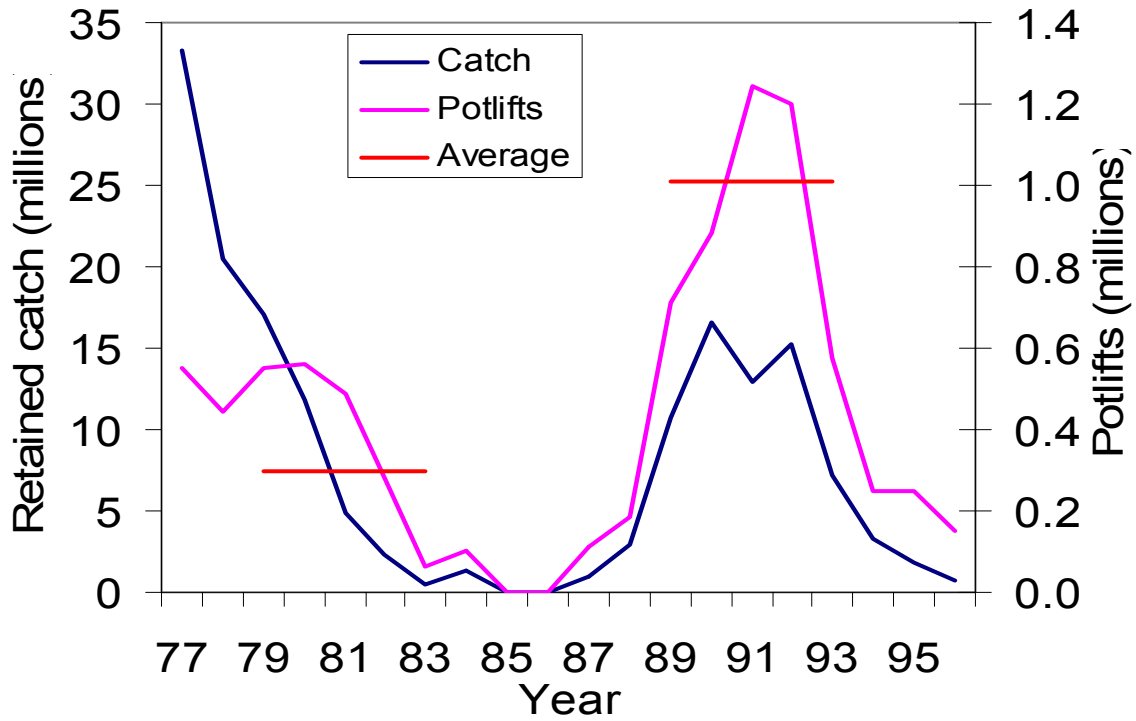
Figure 31. Retained catch and potlifts for total eastern Bering Sea Tanner crab fishery (upper plot) and the Tanner crab fishery east of 163° W (bottom).

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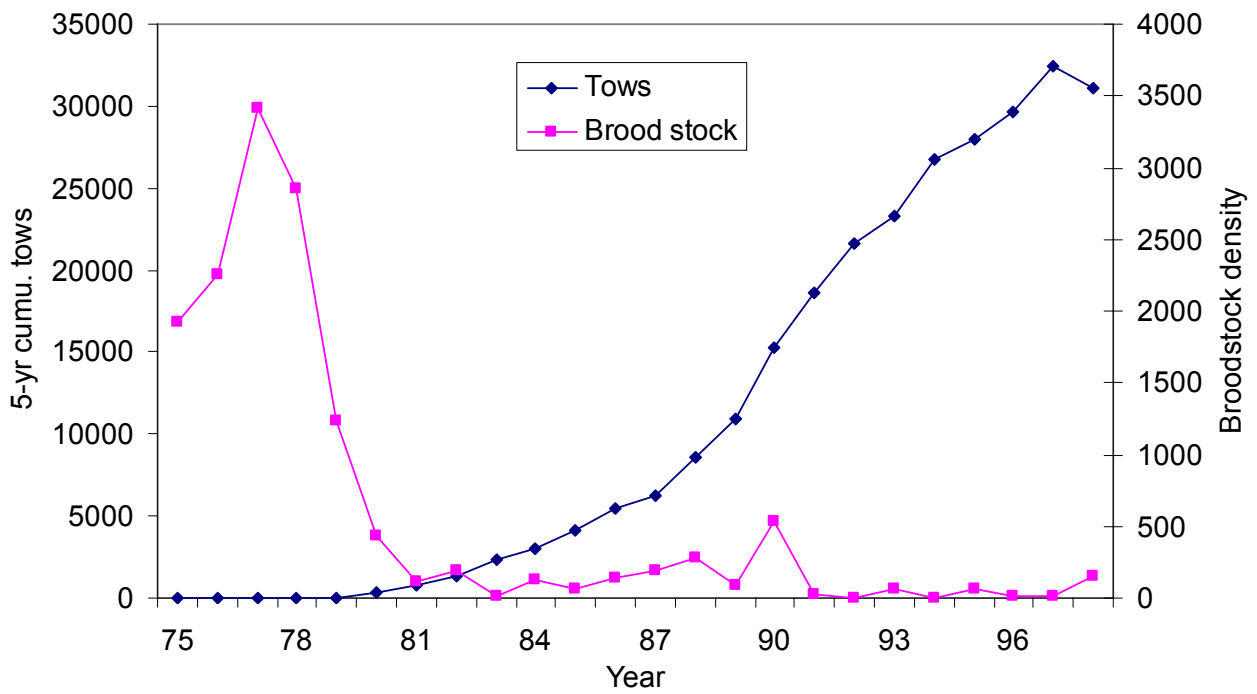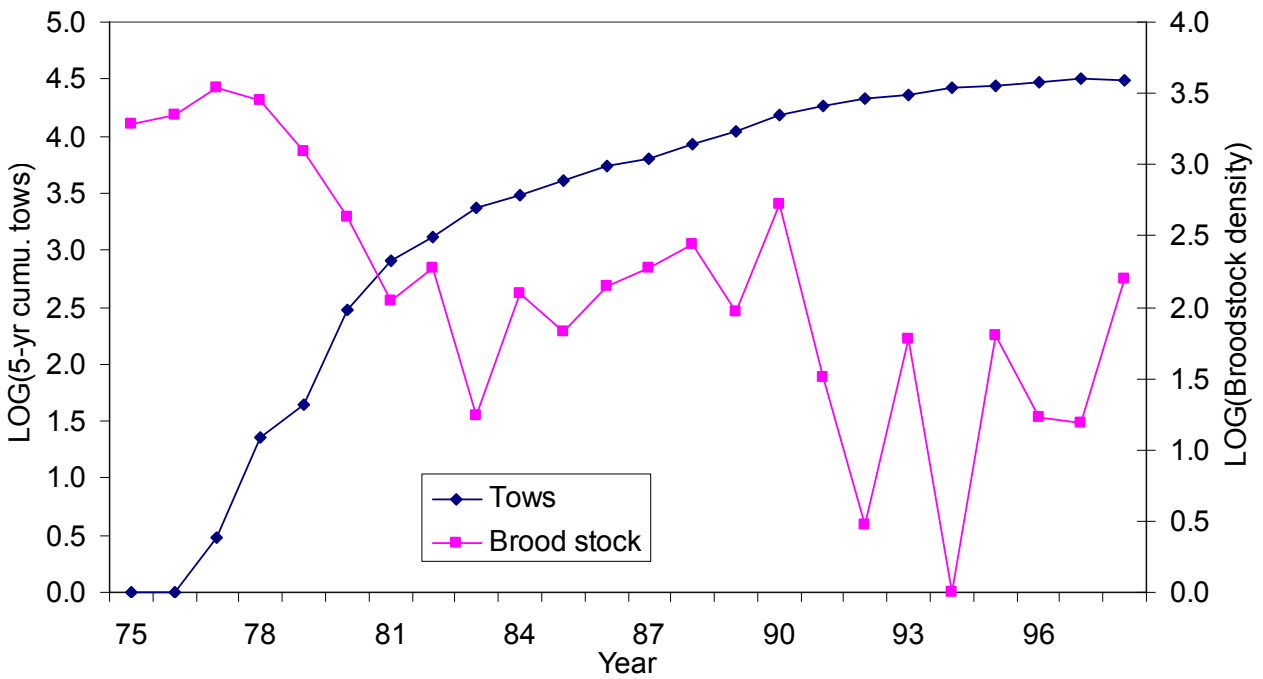
Figure 32a. Numbers of bottom trawl tows and red king crab brookstock density in heavily trawled Unimak and Amak of the southern Bristol Bay during 1975-1998. The upper plot is log-transformed data used by Dew and Mcconnaughey (2005), and the bottom plot is in a normal scale. All data are from Dew and Mcconnaughey (2005).

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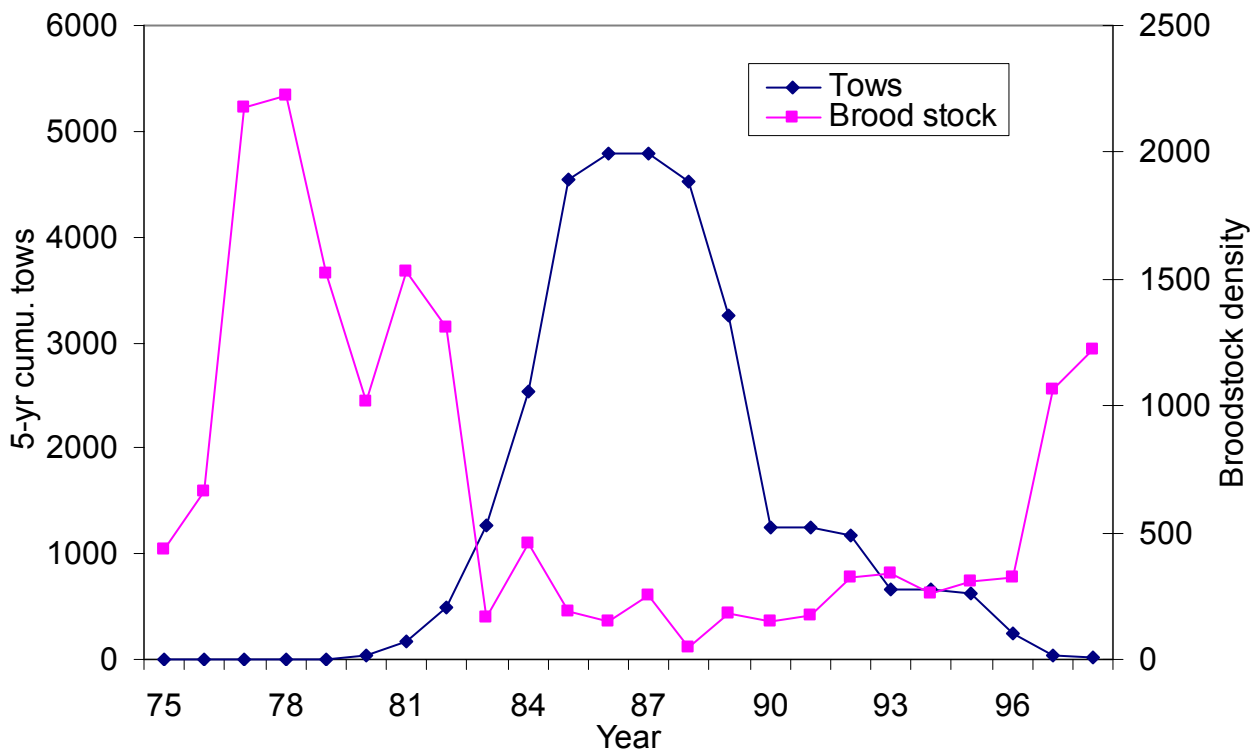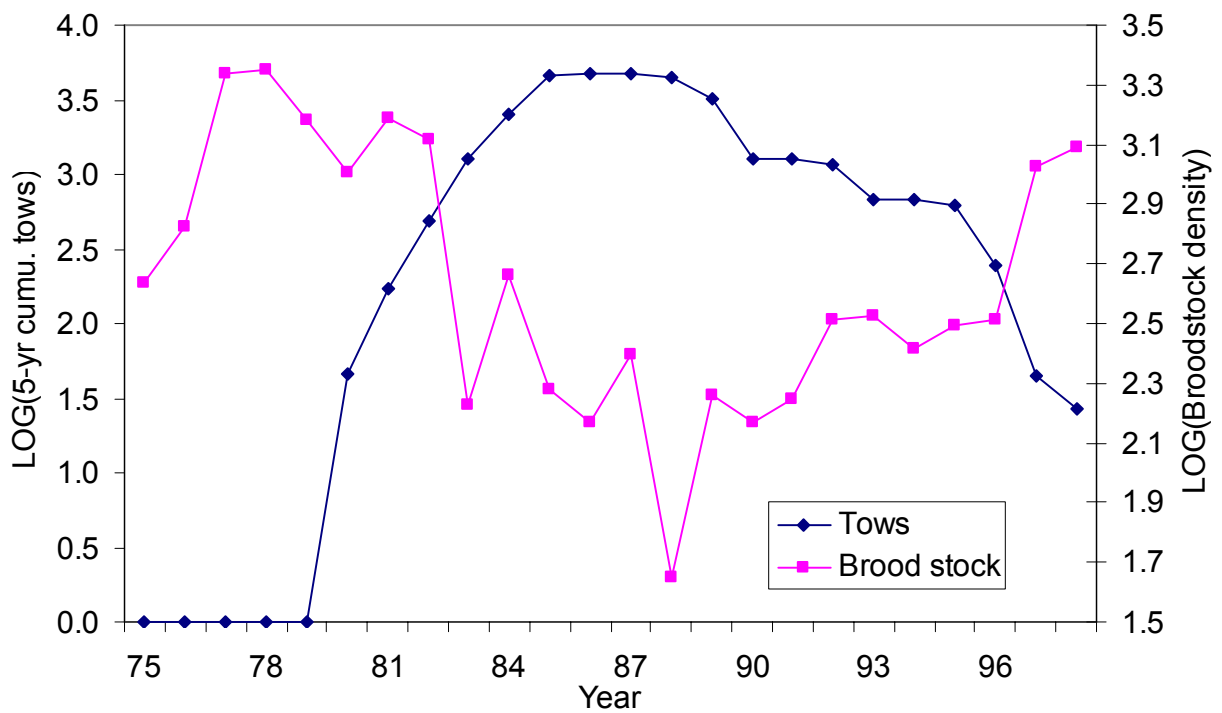
Figure 32b. Numbers of bottom trawl tows and red king crab brookstock density in lightly trawled north of Port Moller, Bristol Bay, during 1975-1998. The upper plot is log-transformed data used by Dew and Mcconnaughey (2005), and the bottom plot is in a normal scale. All data are from Dew and Mcconnaughey (2005).
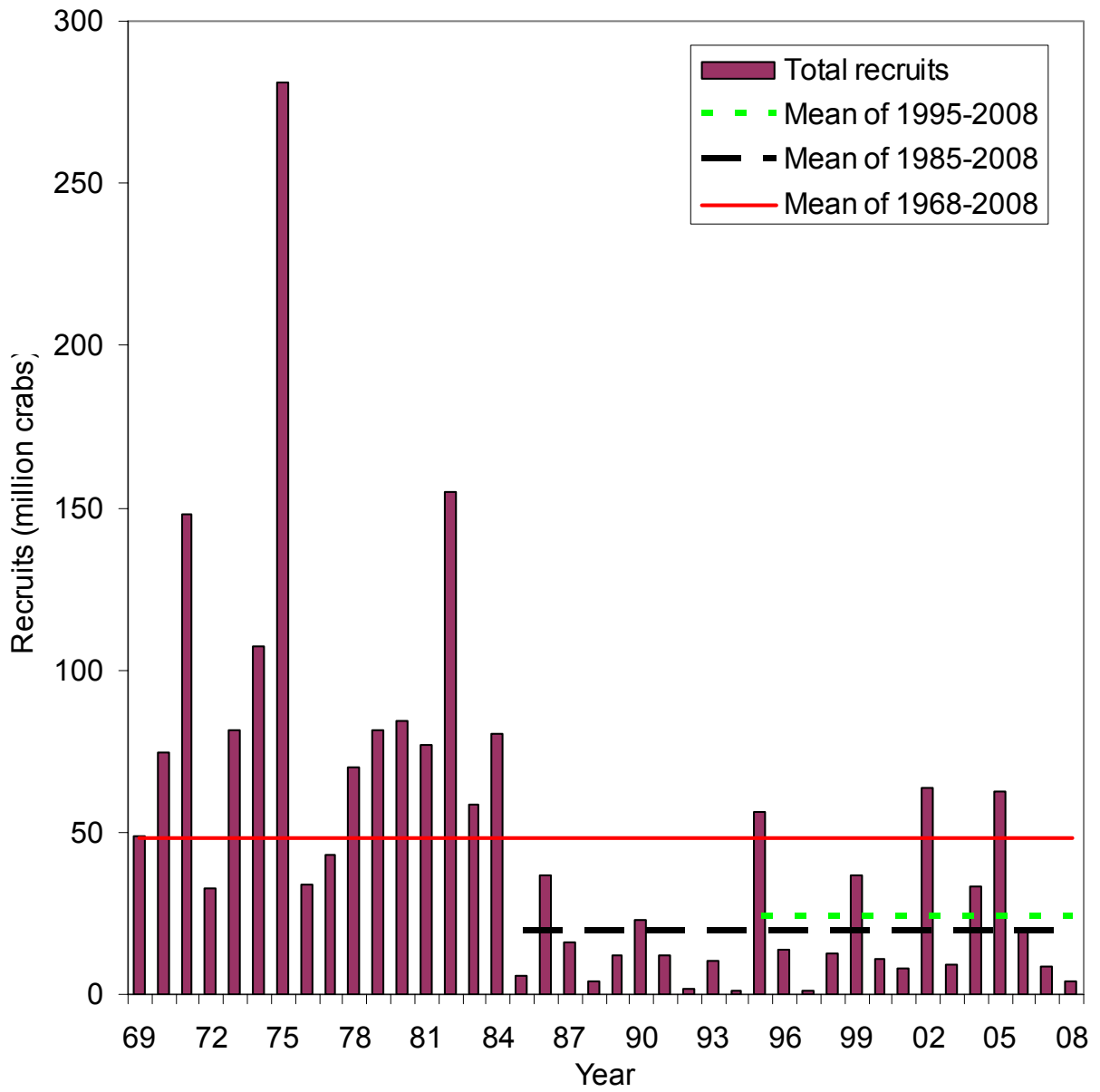
Figure 33. Estimated recruitment time series during 1969-2008. Mean male recruits during 1995-2008 was used to estimate $B_{35\%}$.
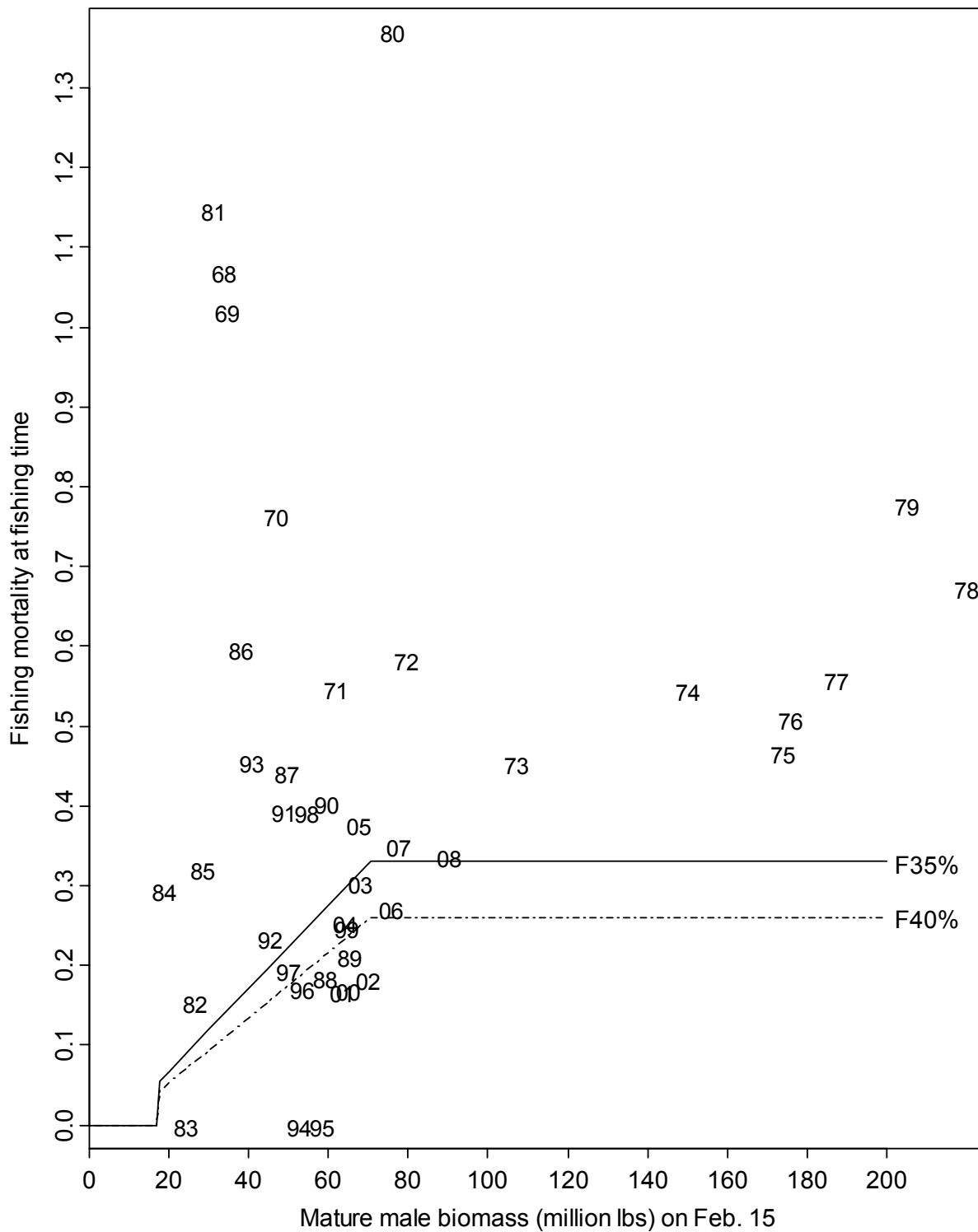
Figure 34a. Relationships between full fishing mortalities for the directed pot fishery and mature male biomass on Feb. 15 during 1968-2008. Average of recruitment from 1995 to 2008 was used to estimate B_{MSY}. Pot handling mortality rate is 0.2.
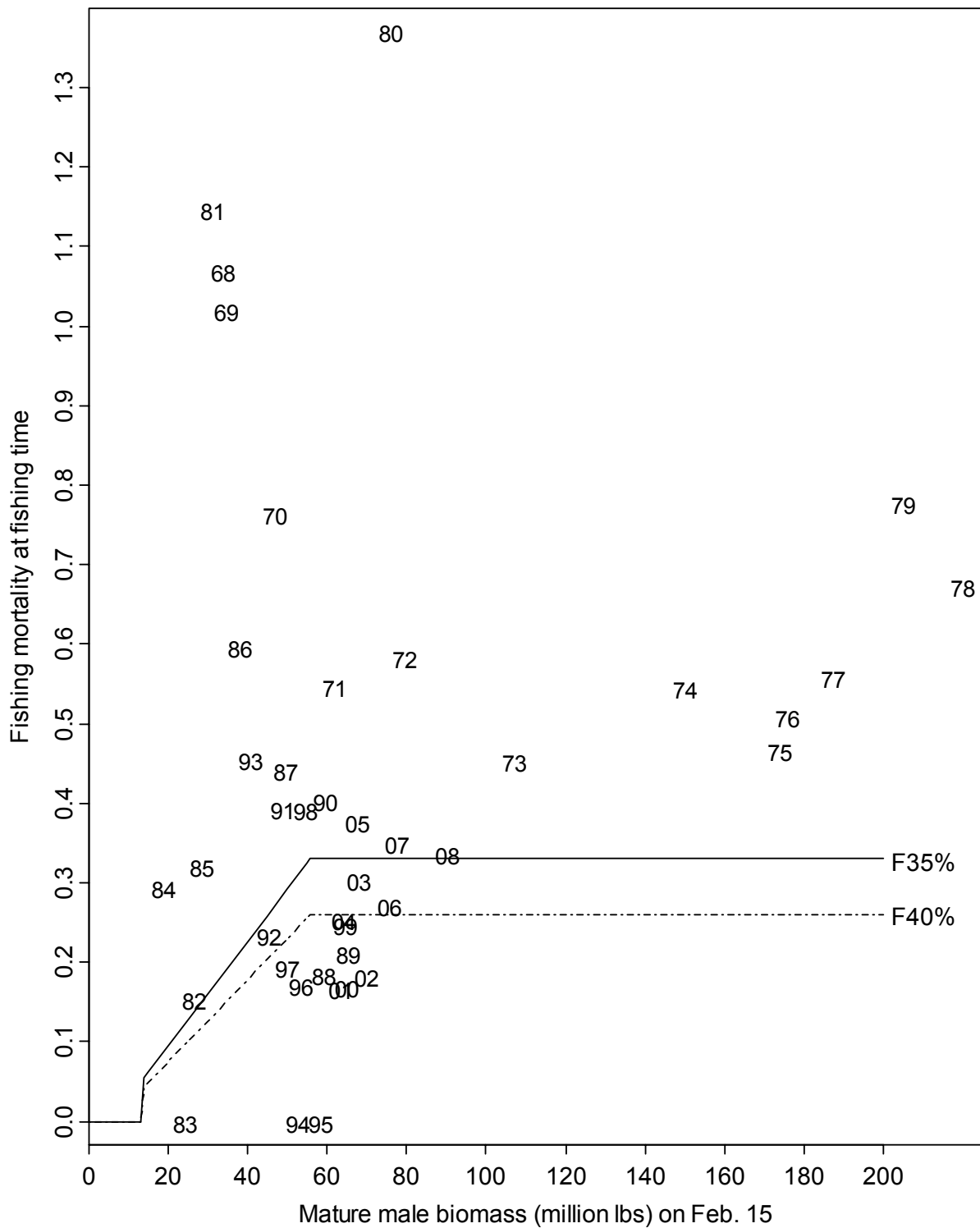
Figure 34b. Relationships between full fishing mortalities for the directed pot fishery and mature male biomass on Feb. 15 during 1968-2008. Average of recruitment from 1985 to 2008 was used to estimate B_{MSY}. Pot handling mortality rate is 0.2.
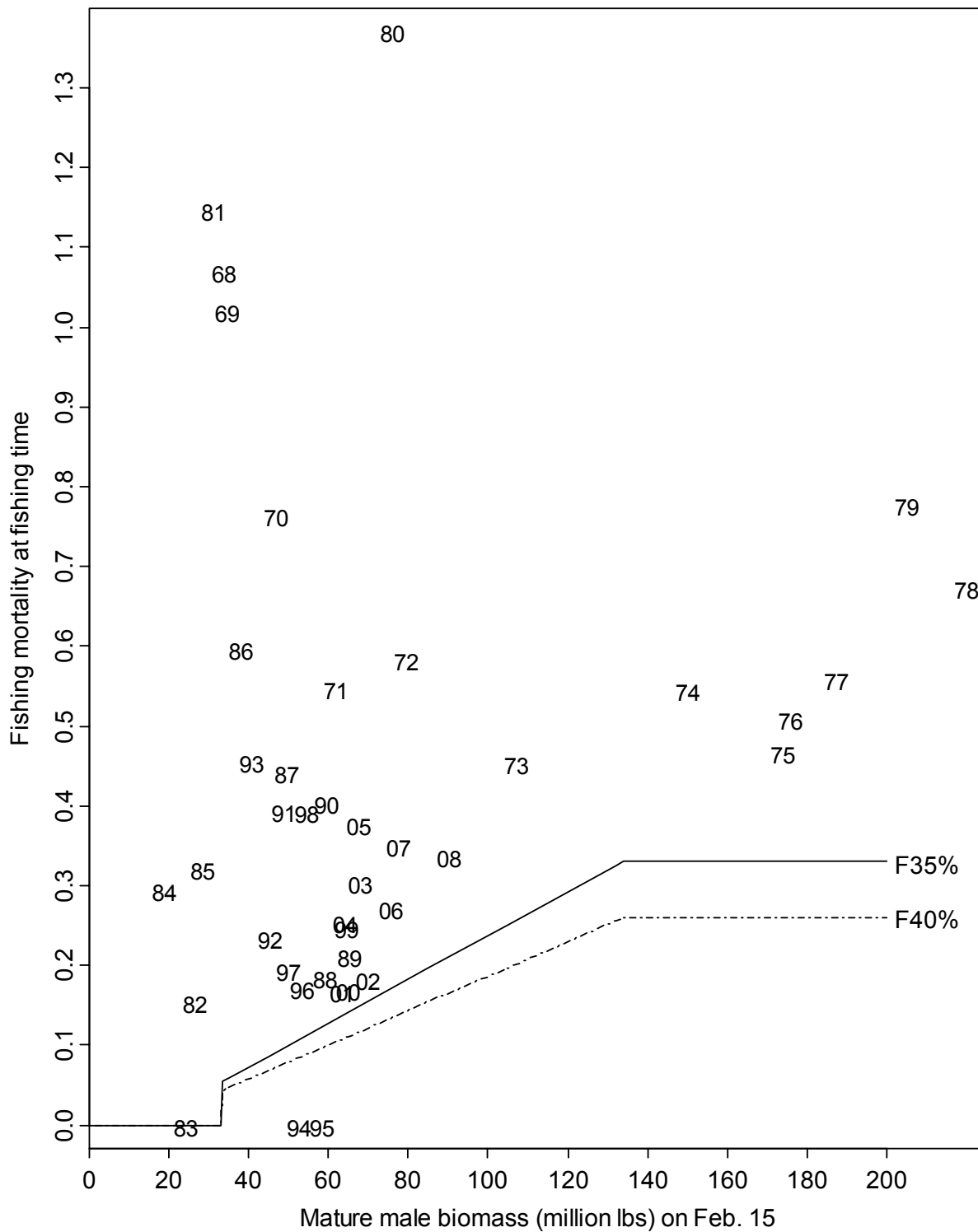
Figure 34c. Relationships between full fishing mortalities for the directed pot fishery and mature male biomass on Feb. 15 during 1968-2008. Average of recruitment from 1968 to 2008 was used to estimate B_{MSY}. Pot handling mortality rate is 0.2.
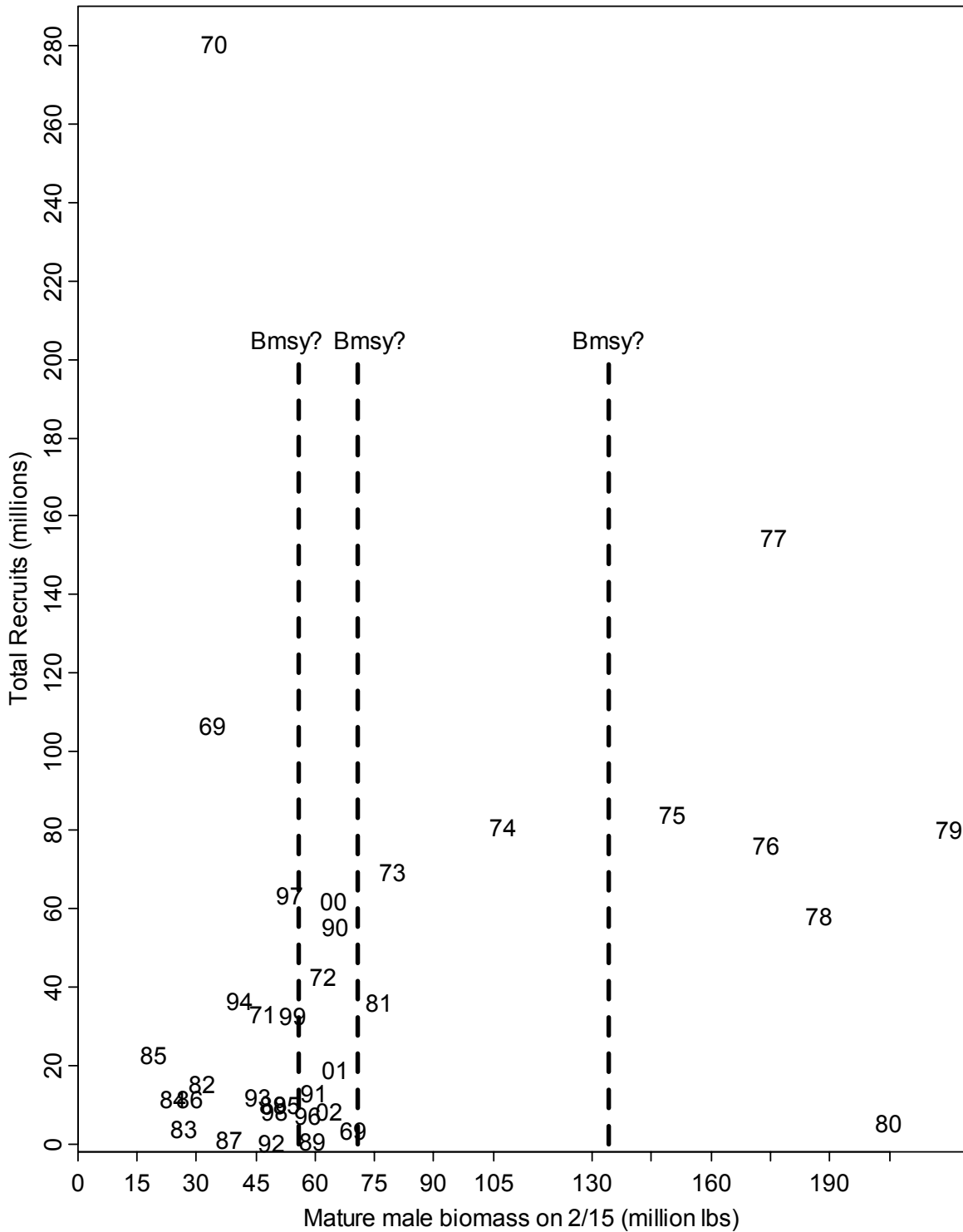
Figure 35. Relationships between mature male biomass on Feb. 15 and total recruits at age 5 (i.e., 6-year time lag) for Bristol Bay red king crab with pot handling mortality rate to be 0.2. Numerical labels are years of mating, and the vertical dotted lines are the estimated $B_{35\%}$ based on three different recruitment levels.
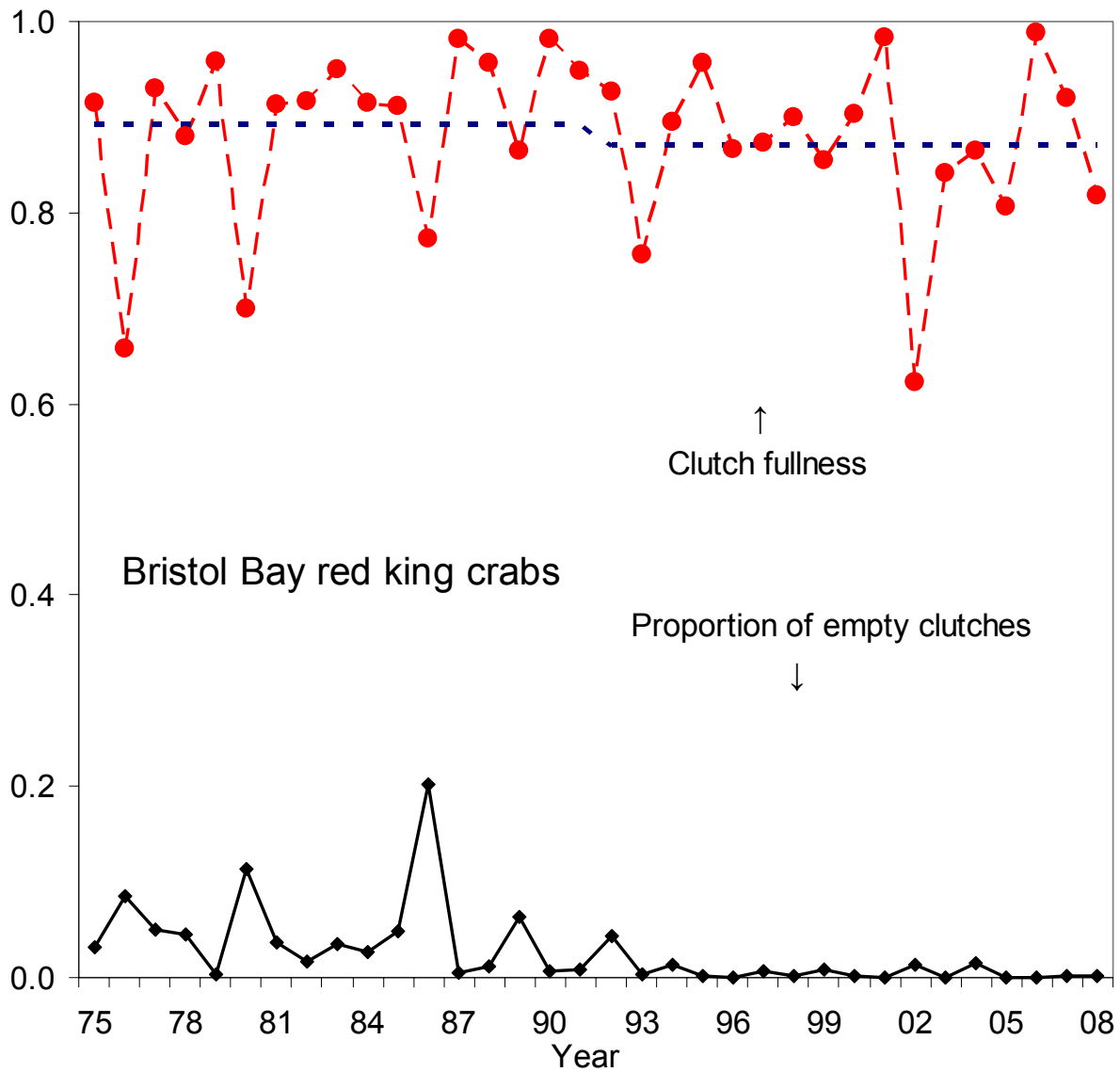
Figure 36. Average clutch fullness and proportion of empty clutches of newshell (shell conditions 1 and 2) mature female crabs >89 mm CL from 1975 to 2008 from survey data. Oldshell females were excluded.

This figure space is reserved for the SAFE report in August 2009.

Figure 37. Likelihood profiles for estimated mature male biomass on Feb. 15 and overfishing limits of retained catch and total catch for the 2009 season with $F_{35\%}$. Pot and trawl handling mortality rates were assumed to be 0.2 and 0.8, respectively.

This figure space is reserved for the SAFE report in August 2009.

Figure 38. Projected mature male biomass on Feb. 15 with $F_{40\%}$, $F_{35\%}$ and the ADF&G harvest strategy with $F_{35\%}$ constraint during 2009-2118. Pot handling mortality rate is 0.2 and the confidence limits are for the ADF&G harvest strategy.

This figure space is reserved for the SAFE report in August 2009.

Figure 39. Projected retained catch biomass on Feb. 15 with $F_{40\%}$, $F_{35\%}$ and the ADF&G harvest strategy with $F_{35\%}$ constraint during 2009-2118. Pot handling mortality rate is 0.2 and the confidence limits are for the ADF&G harvest strategy.
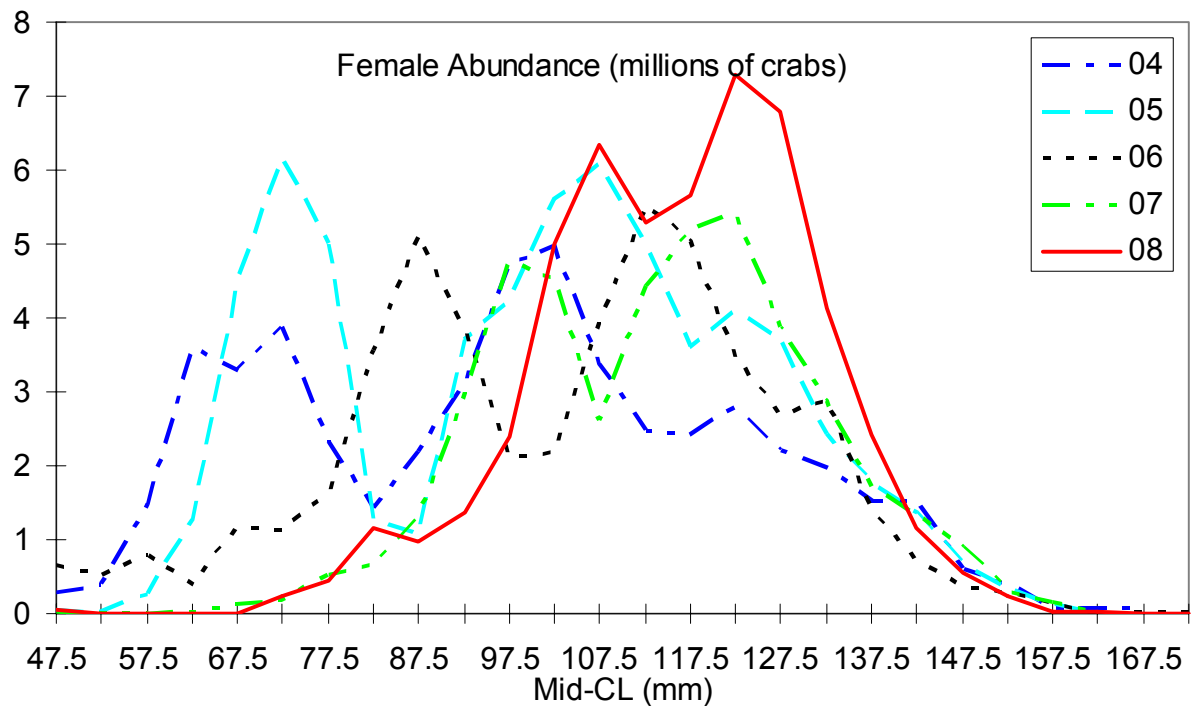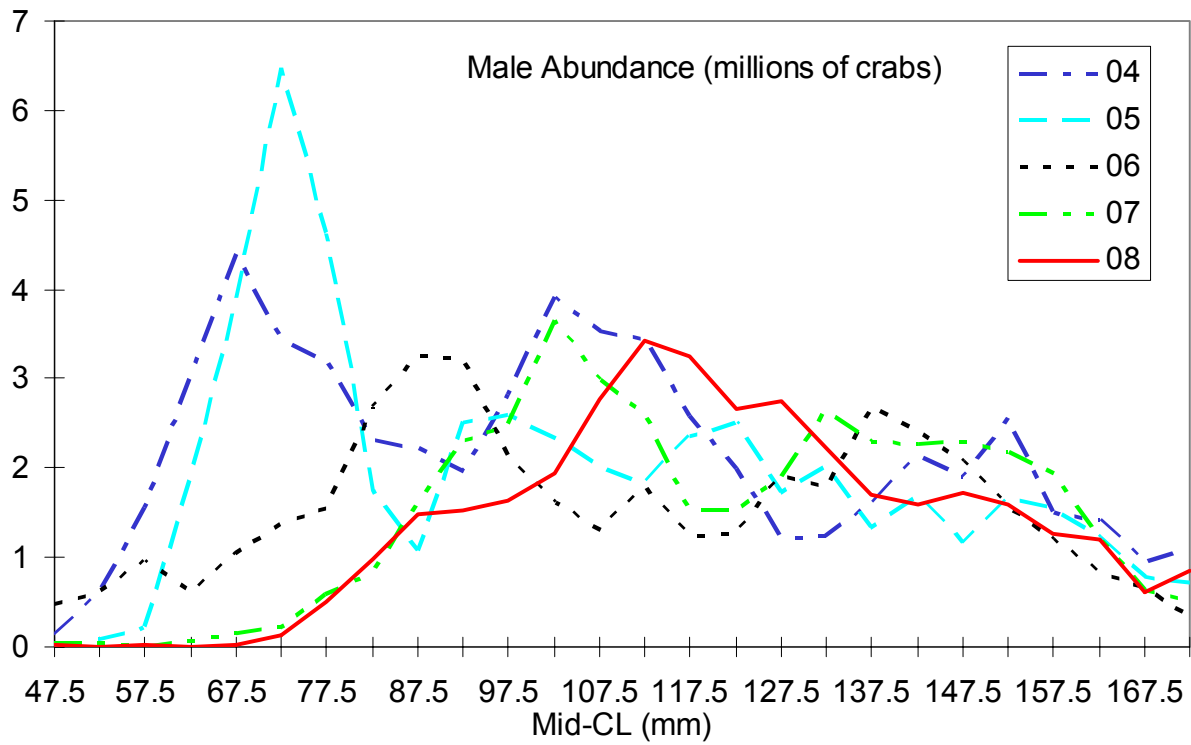
Figure 40. Length frequency distributions of male (top panel) and female (bottom panel) red king crabs in Bristol Bay from NMFS trawl surveys during 2004-2008. For purposes of these graphs, abundance estimates are based on area-swept methods.
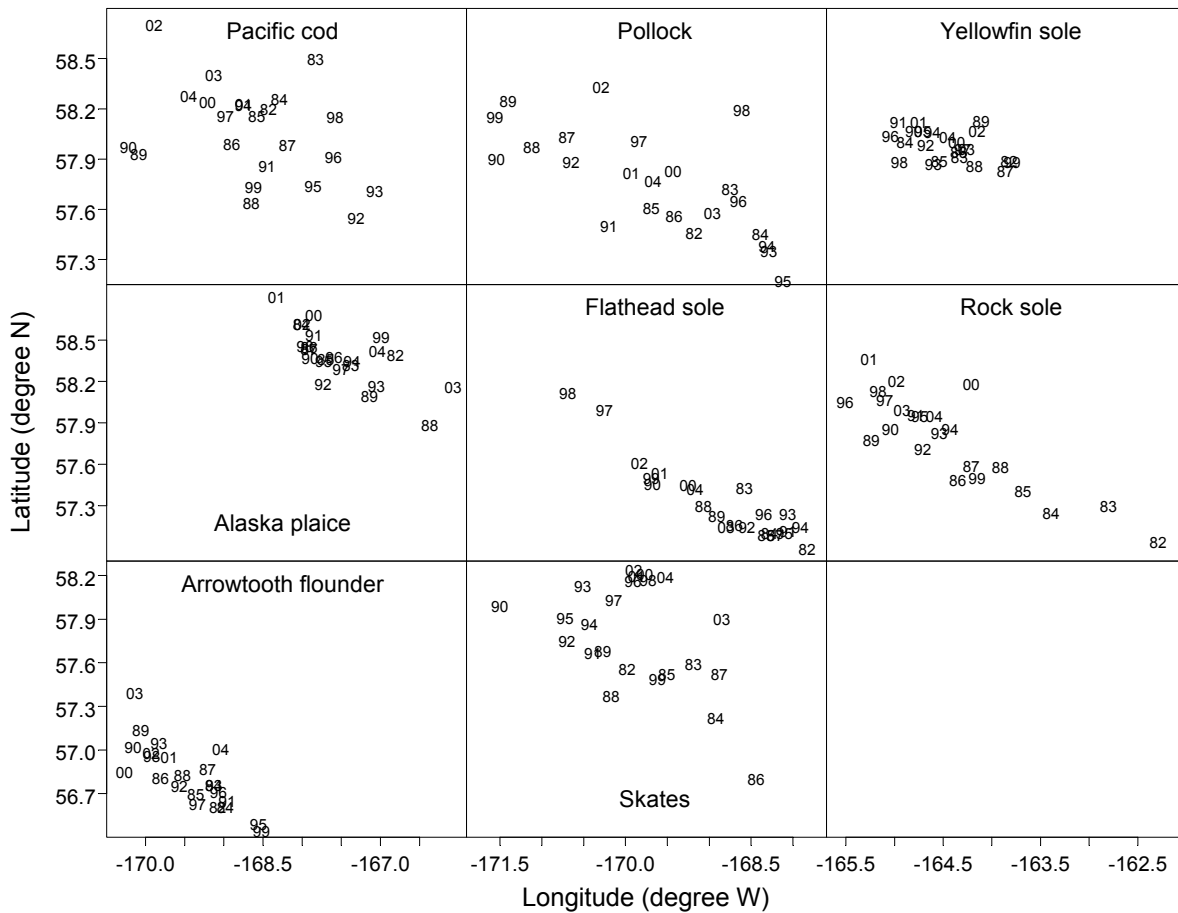
100

Figure 41. Biomass distribution centers of Pacific cod, walleye pollock, yellowfin sole, Alaska plaice, flathead sole, rock sole, arrowtooth flounder, and skates derived from NMFS summer trawl survey data in the eastern Bering Sea. (Source: Zheng and Kruse 2006).
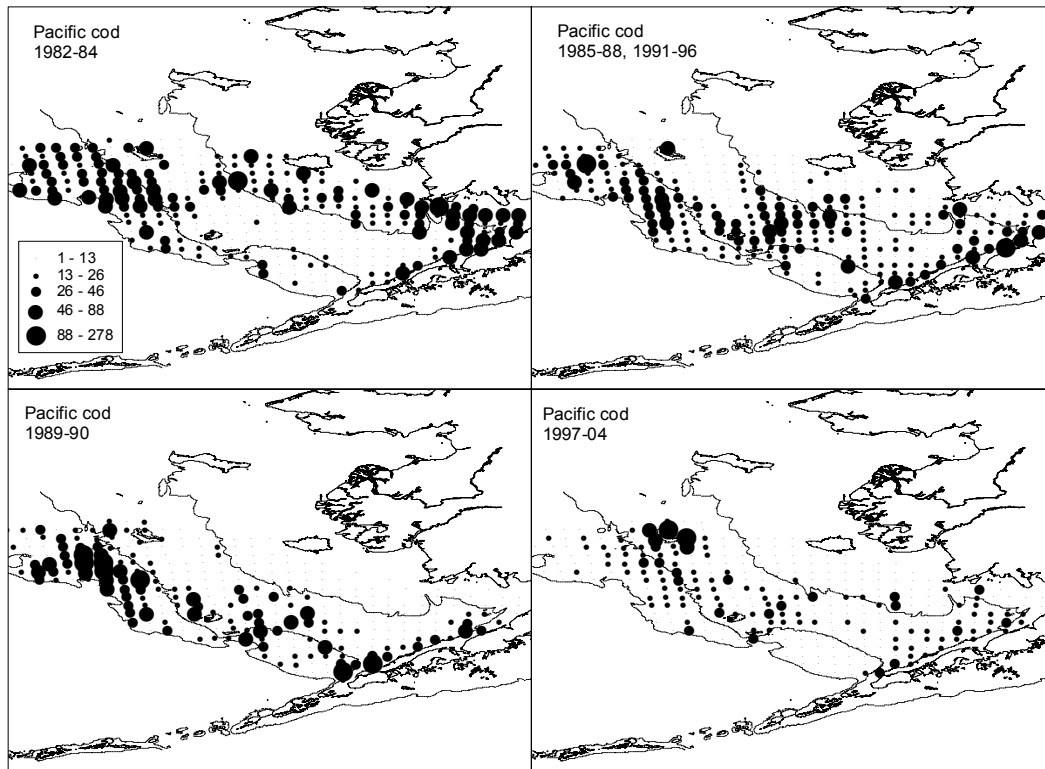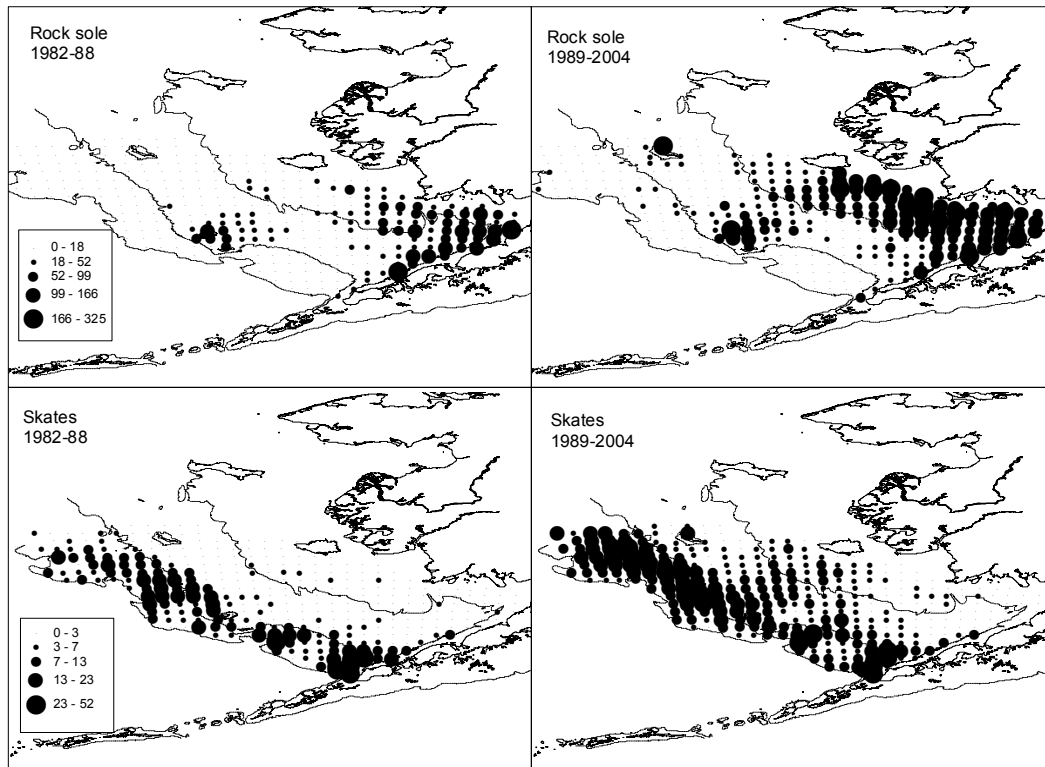
Figure 42. Distributions of relative biomass of Pacific cod in the eastern Bering Sea from 1982 to 2004 derived from NMFS summer trawl survey data. Relative biomass is expressed as kg/ha. Three depth contour lines are 50, 100, and 200 m. (Source: Zheng and Kruse 2006).

Figure 43. Distributions of relative biomass of rock sole and skates in the eastern Bering Sea from 1982 to 2004 derived from NMFS summer trawl survey data. Relative biomass is expressed as kg/ha. Three depth contour lines are 50, 100, and 200 m. (Source: Zheng and Kruse 2006).
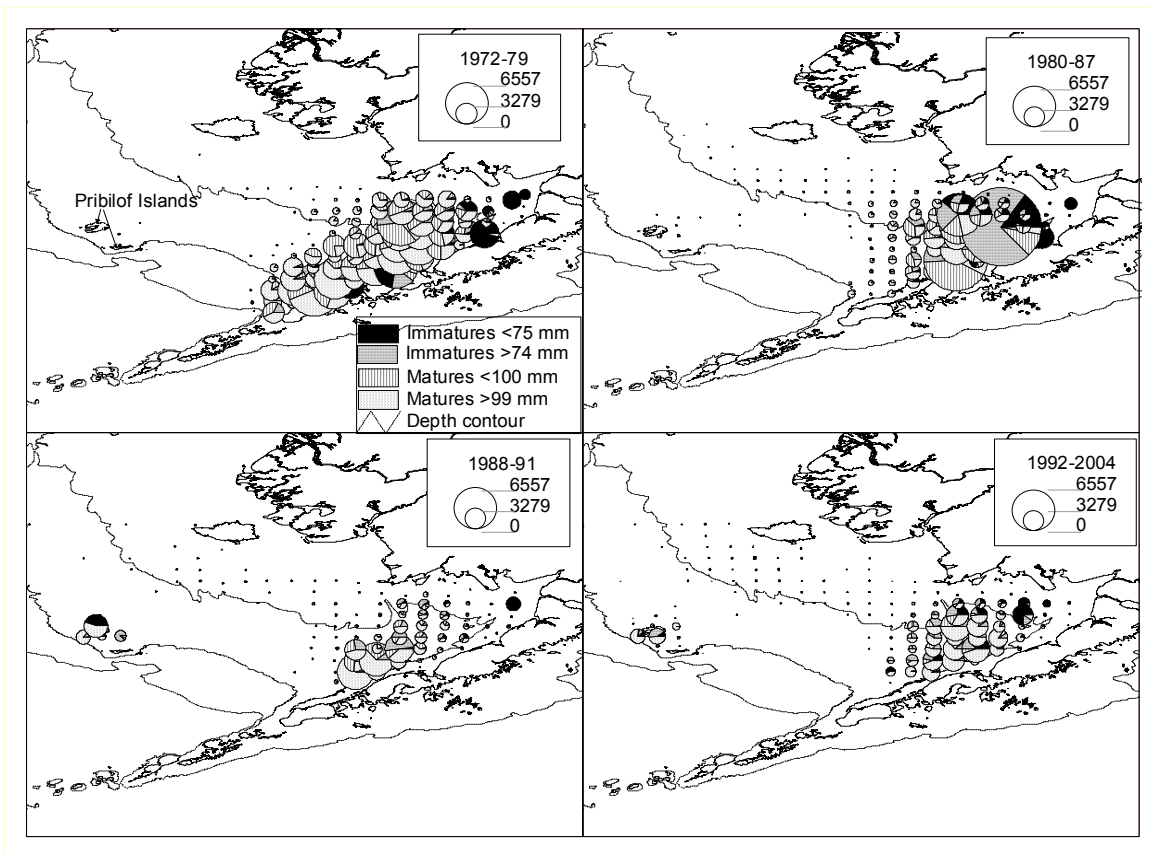
Figure 44. Geographic distributions of immature and mature female red king crabs from 1972 to 2004 in the eastern Bering Sea derived from NMFS summer trawl survey data. The diameter of each pie represents crab density expressed as the number of crabs per square nautical mile. Three depth contour lines are 50, 100, and 200 m. (Source: Zheng and Kruse 2006).
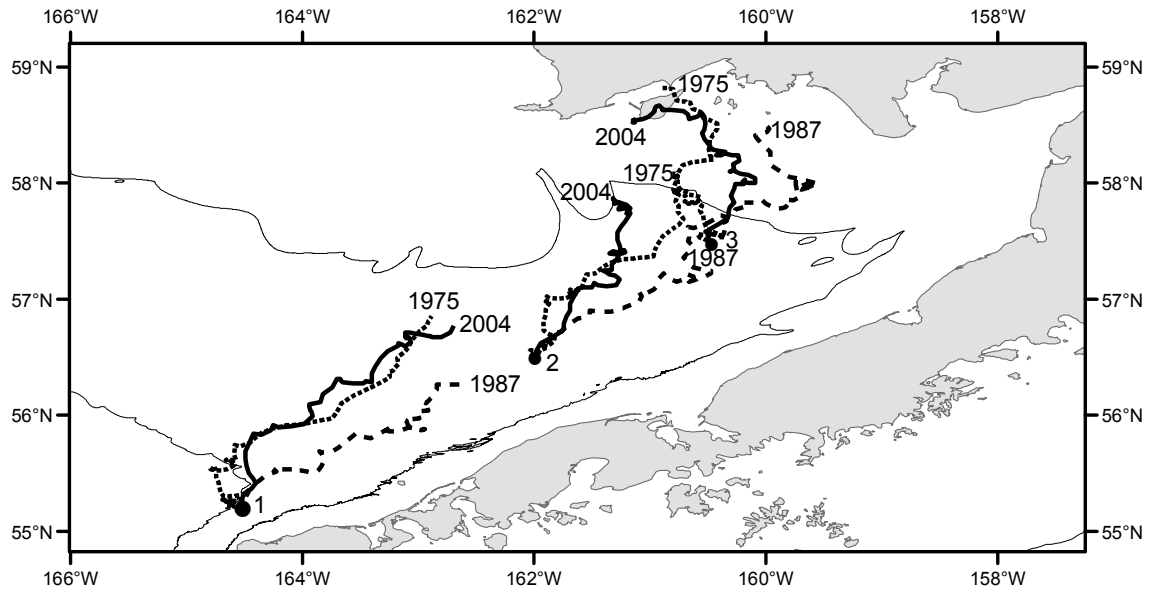
Figure 45. Larval movements after hatching on May 15, 1975, 1987, and 2004 from three different locations for Bristol Bay red king crab during two months. (Source: Zheng and Kruse 2008).
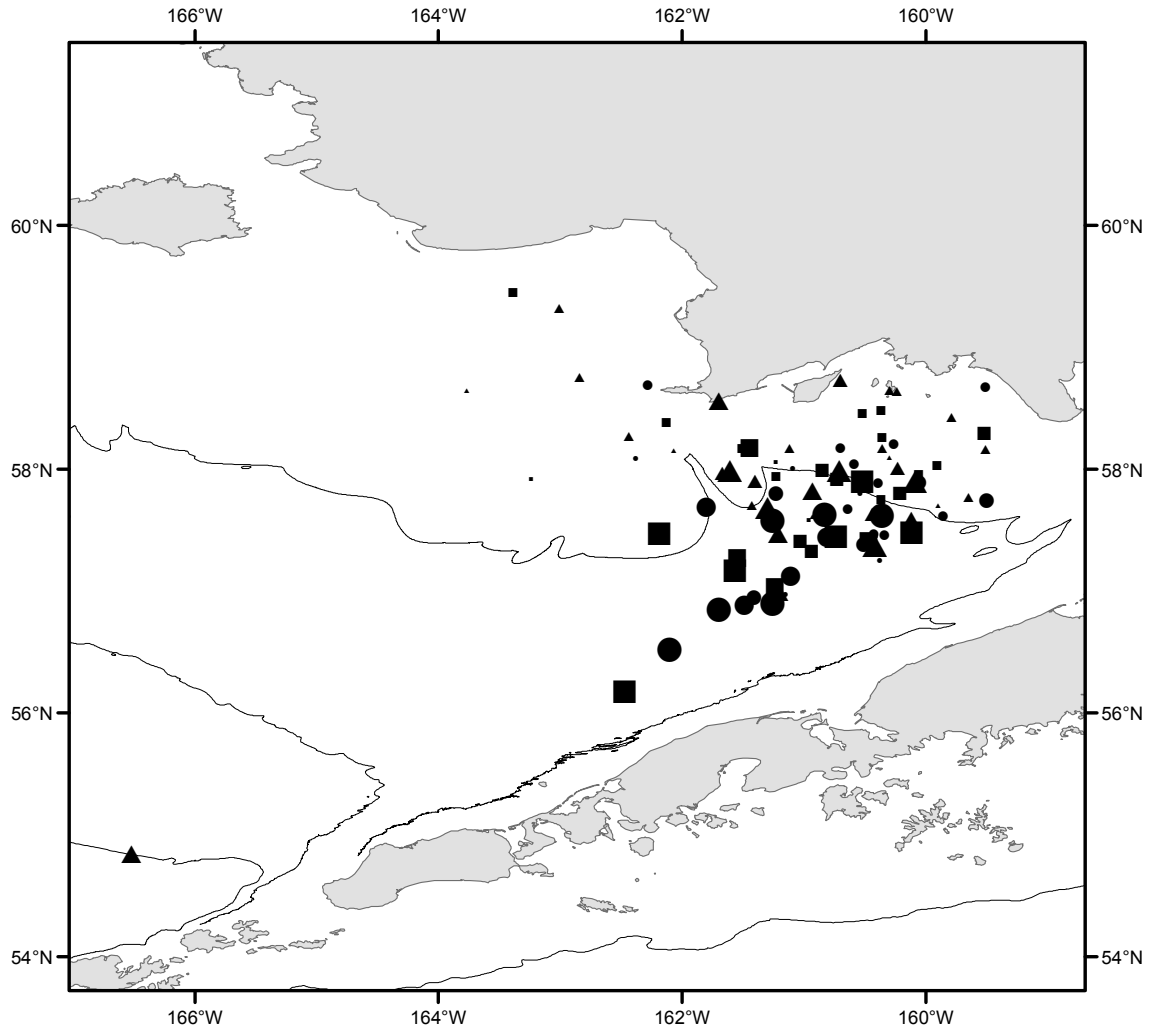
Figure 46. Estimated settling locations from the distribution centers of Bristol Bay mature female red king crabs >99 mm CL during 1967-1999. Hatching dates of April 15, May 15, and June 15 are triangles, squares, and circles, respectively. Symbol sizes are proportional to year-class strength.