Interlaboratory Comparison of Four *in Vitro* Assays for Assessing Androgenic and Antiandrogenic Activity of Environmental Chemicals

Wolfgang Körner,¹ Anne Marie Vinggaard,² Béatrice Térouanne,³ Risheng Ma,⁴ Carise Wieloch,⁵ Margret Schlumpf,⁴ Charles Sultan,³ and Ana M. Soto⁵

¹Bayerisches Landesamt für Umweltschutz, Augsburg, Germany; ²Danish Veterinary and Food Administration, Institute of Food Safety and Nutrition, Søborg, Denmark; ³INSERM U 439, Pathologie Moléculaire des Récepteurs Nucléaires, Montpellier, France; ⁴University of Zürich, Institute of Pharmacology and Toxicology, Zürich, Switzerland; ⁵Department of Anatomy and Cellular Biology, Tufts University School of Medicine, Boston, Massachusetts, USA

We evaluated and compared four in vitro assays to detect androgen agonists and antagonists in an international interlaboratory study. Laboratory 1 used a cell proliferation assay (assay 1) with human mammary carcinoma cells stably transfected with human androgen receptor. The other laboratories used reporter gene assays, two based on stably transfected human prostate carcinoma cells (assay 2) or human mammary carcinoma cells (assay 4), and the third based on transient transfection of Chinese hamster ovary cells (assay 3). Four laboratories received four coded compounds and two controls: two steroidal androgens, two antiandrogens, an androgenic control, 5a-dihydrotestosterone (DHT), and an antiandrogenic control, bicalutamide (ICI 176,334). All laboratories correctly detected the androgenic activity of 4-androsten-3,17-dione and 17α -methyltestosterone. For both compounds, the calculated androgenic potencies relative to the positive control (RAPs) remained within one order of magnitude. However, laboratory 3 calculated a 50-fold higher RAP for 4-androsten-3,17-dione. All assays detected and quantified the antiandrogenic effect of vinclozolin [median inhibitory concentration (IC₅₀) values ranging from 1.1×10^{-7} M to 4.7×10^{-7} M]. In assays 2 and 3, vinclozolin showed partial and rogenic activity at the highest concentrations tested. For vinclozolin, calculated antiandrogenic potencies relative to bicalutamide (RAAPs) differed no more than a factor of 10, and IC50 values matched those of bicalutamide. Similarly, we found antiandrogenic activity for tris-(4-chlorophenyl)methanol. RAAP values were between 0.086 and 0.37. Three assays showed cytotoxicity for this compound at or above 1×10^{-5} M. In summary, all assays proved sensitive screening tools to detect and quantify androgen receptor-mediated androgenic and antiandrogenic effects of these chemicals accurately, with coefficients of variation between 8 and 90%. Key words: androgenicity, 4-androsten-3,17-dione, antiandrogenicity, A-SCREEN, bicalutamide, tris-(4-chlorophenyl)methanol (TCPM), vinclozolin. Environ Health Perspect 112:695-702 (2004). doi:10.1289/ehp.6715 available via http://dx.doi.org/ [Online 22 January 2004]

In recent years, researchers have shown that a large and continuously growing number of man-made chemicals and naturally occurring compounds can mimic endogenous estrogens of vertebrates including humans (Colborn et al. 1993; Schlumpf et al. 2001; Soto et al. 1995). In addition to compounds naturally occurring in plants and some mycotoxins (Breithofer et al. 1998; Miksicek 1994), various pesticides and industrial chemicals have shown estrogenic effects in vitro (Jobling et al. 1995; Klotz et al. 1996; Körner et al. 1998; Miller et al. 2001; Soto et al. 1991; Soto et al. 1995) and in vivo in mammals (Dodds and Lawson 1936; Milligan et al. 1998; Nagel et al. 1997; Soto et al. 1991) as well as fish (Christiansen et al. 2000; Donohoe and Curtis 1996; Gimeno et al. 1996; Jobling et al. 1996). However, we have not identified a common chemical substructure responsible for the estrogenic activity of the so-called xenoestrogens and phytoestrogens. The elucidation of the crystal structure of the ligand-binding domain of the human estrogen receptor (ER)-a (Brzozowski et al. 1997) explained, in part, this surprising structural diversity. It revealed a ligand-binding domain gap almost twice as large as that required by 17β -estradiol (E₂), allowing space for a variety of other molecules to interact with the ER. Therefore, we must assume that more xenobiotics with unknown estrogenic activity exist, requiring specific, sensitive, and practical experimental screening systems for detection.

In contrast, we know comparatively little about the interference of chemicals with the human androgen receptor (hAR). A few derivatives of the known nonsteroidal antiandrogen bicalutamide (ICI 176,334) have shown androgenic effects in vitro (Dalton et al. 1998), and the bioaccumulating DDT metabolite *p*,*p*'-DDE (dichlorodiphenyldichloroethylene), the fungicides vinclozolin, procymidone, and prochloraz, and the herbicide linuron (Cook et al. 1993; Fail et al. 1995; Gray et al. 1994; Kelce et al. 1995; Vinggaard et al. 2002) have demonstrated androgen receptor (AR)mediated antiandrogenic activities in vitro and in vivo. The two vinclozolin metabolites, M1 and M2, cause antiandrogenic effects in male rats (Kelce et al. 1994). These metabolites (Figure 1) and linuron, but not p,p'-DDE, relate structurally to the therapeutic antiandrogen flutamide. The natural insecticide pyrethrin and some synthetic pyrethroids act as competitive AR antagonists in human fibroblasts (Eil and Nisula 1990). Tyler et al. identified 3-phenoxybenzylalcohol, a metabolite of the pyrethroid permethrin, as an antiandrogen in genetically modified yeast cells (Tyler et al. 2000). In addition, polycyclic aromatic hydrocarbons block AR activation *in vitro* (Vinggaard et al. 2000) and suppress androgen-dependent growth of accessory sex organs in juvenile male rats (Chang and Liao 1987). Recently, several phenolic chemicals demonstrated antiandrogenic activity in a reporter cell line (Paris et al. 2002).

These examples indicate that, analogous to xenoestrogens and phytoestrogens, a common chemical substructure may not exist for ARmediated antiandrogenic activity. This stresses the need for adequate screening systems. Although many in vitro and in vivo bioassays can detect ER-mediated activity, relatively few test systems measure androgen agonist and antagonist effects in vitro. To devise strategies for screening new and existing chemicals, we must test the accuracy and comparability of existing assays. While recent comparison studies of bioassays for detection of estrogenic activity of pure substances have revealed comparable results in most cases (Andersen et al. 1999; Fang et al. 2000), this work presents, for the first time, an interlaboratory comparison of in vitro assays for measuring AR-mediated androgenic and antiandrogenic activity.

Materials and Methods

Participants and test systems. All four laboratories participating in this study applied cellular

Address correspondence to W. Körner, Bayerisches Landesamt für Umweltschutz, Referat Z5, 86177 Augsburg, Germany. Telephone: 49 821 9071 5287. Fax: 49 821 9071 5559. E-mail: wolfgang.koerner@ lfu.bayern.de

We thank N. Servant (laboratory 2) and B. Møller Plesning (laboratory 3) for helpful technical assistance.

This work was supported by the Institute de la Santé et de la Recherche Medicale, contract QLK4CT-1999-0142 of the European Community (laboratory 2); the National Institutes of Health, grant ES08314C (laboratory 1); and the Danish Medical Research Council, grant 9801270 (laboratory 3).

The authors declare they have no competing financial interests.

Received 29 August 2003; accepted 21 January 2004.

test systems. Three of the four laboratories worked with reporter gene assays with different human and mammalian cell lines. These assays rely on the expression of the firefly luciferase gene under control of the hAR. Laboratory 1 used the A-SCREEN assay, which measures androgen-dependent inhibition of proliferation of an AR-positive human mammary carcinoma cell line (Szelei et al. 1997). Table 1 lists an overview of the four participating laboratories and their respective test systems. All participants have previously described their assays in the literature (Szelei et al. 1997; Térouanne et al. 2000; Vinggaard et al. 1999; Wilson et al. 2002) and performed them in this study with the following modifications:

Laboratory 1 used the A-SCREEN assay. This assay can be run in serumless or serumsupplemented medium. The serum-free medium does not require E_2 to achieve maximal cell yield, and the serum-supplemented medium does. Both methods give comparable results (Szelei et al. 1997), so in conjunction with the other assays, laboratory 1 used serum-supplemented medium. They seeded 25,000–35,000 AR-positive MCF-7-AR1



Figure 1. Structures of the test compounds and the positive controls. Abbreviations: DHT, 5α -dihydrotestosterone; TCPM, tris-(4-chlorophenyl)methanol.

cells in Dulbecco's modification of Eagle's medium (DMEM) with phenol red and 5% fetal calf serum (FCS) into each well of 24-well plates. Twenty-four hours later, they changed the medium to phenol red-free DMEM supplemented with 5% charcoal-dextran-treated FCS (CDFCS). Because these cells express ER as original MCF-7 cells do, E2 was added to each well, except for a control row of four wells, to yield a final concentration of 0.1 nM, allowing maximal proliferation of MCF-7-AR1 cells. Androgens inhibit proliferation in a concentration-dependent manner to about 30% of the maximal cell count (Figure 2A). Under the conditions of the A-SCREEN assay, the cells arrest in G₀/G₁ phase (Soto et al. 1999; Szelei et al. 1997). Coincubation with an antiandrogen blocks that inhibition of cell proliferation. Laboratory 1 ran an extended dose-response curve to 5α -dihydrotestosterone (DHT) with each experiment and added DHT (to yield a concentration of 0.1 nM) to the bottom two rows of each plate with test compounds and solvent controls to test for antiandrogenicity. They left the second column in each plate as an androgenless control. After 5 days of incubation, they fixed the cells and stained them with sulforhodamine B (SRB). They resuspended the dye in 10 mM Tris base (pH 10.5), transferred aliquots from each well into a 96-well plate, then calculated cell numbers from the optical densities measured at 515 nm.

Laboratory 2 seeded stably transfected human prostate adenocarcinoma (PC-3) cells (PALM cells) in white opaque tissue culture 96-well plates (Becton Dickinson, Le Pont de Claix, France). They used 20,000 cells in 150 uL HAM-F12 medium supplemented with 3% CDFCS in each well. After an 8-hr incubation, they added the compounds in 50 µL of the same medium without replacement of the seeding medium. After 30 hr, they removed the culture medium with the tested compounds and replaced it with luminescent buffer (50 µL/well phenol red-free DMEM, 3×10^{-4} M luciferin). They measured luciferase activity in intact cells with a luminometer (Trilux Wallac; PerkinElmer, Courtaboeuf Cedex, France).

Laboratory 3 tested the compounds in a reporter gene assay based on transient transfections as originally described (Vinggaard et al. 1999), but with major modifications. They maintained Chinese hamster ovary cells (CHO K1) in DMEM/F12 (Gibco, Paisley, U.K.) supplemented with 100 U/mL penicillin, 100 µg/mL streptomycin, and 0.25 µg/mL amphotericin B (Sigma-Aldrich, St. Louis, MO, USA) and 10% FCS (BioWhitaker, Walkersville, MD, USA). They seeded cells in white 96-well plates (PerkinElmer Life Sciences, Packard, Groningen, the Netherlands) at a density of 7,000 cells/well in DMEM/F12 medium containing 10% charcoal-treated FCS

(BioWhitaker) and incubated the plates at 37°C in a humidified atmosphere of 5% CO₂/air. After 24 hr, they transfected the cells for 5 hr with a total of 75 ng DNA/well consisting of the AR expression vector pSVAR0 and the mouse mammary tumor virus-luciferase (MMTV-LUC) reporter plasmid (both provided by A. Brinkmann, Erasmus University, Rotterdam, the Netherlands) in a ratio of 1:100 using 0.30 µL of the transfection reagent FuGene (Boehringer, Mannheim, Germany). They kept the ratio of DNA (micrograms) to FuGene (microliters) at 0.25. After removing the transfection medium, they added the test compounds. Laboratory 3 tested all concentrations in quadruplicate. After incubation for 20 hr, they aspirated the medium and lysed the cells by adding 20 µL/well of a lysis buffer containing 25 mM trisphosphate, pH 7.8, 15% glycerol, 1% Triton X-100, 1 mM dithiothreitol, and 8 mM MgCl₂, followed by shaking at room temperature for 10 min. They measured luciferase activity directly in the culture plates using a BioOrbit Galaxy luminometer (Anthos Labtec Instruments, Wals, Austria).

Laboratory 4 carried out a test according to the protocol previously described by Wilson et al. (2002) with some modifications and simplifications. They routinely maintained MDA-kb2 human breast carcinoma cells in Leibowitz's L-15 medium supplemented with 10% heat-inactivated (56°C, 30 min) FCS and 1% antibiotic-antimycotic (all from Gibco) at 37°C in a humidified incubator without CO2. They detached cells by trypsinization and seeded them into 96-well plates at a density of about 10,000 cells/well in 100 µL medium. After cells had attached for 24 hr, they replaced the medium with a total of 100 µL/well of medium containing various dilutions of test chemicals. In each plate they filled four wells in one column with medium only as the negative control, four wells in another column with 1×10^{-10} M DHT (final concentration) as the androgen control, and four wells in a third column with 1×10^{-6} M bicalutamide (final concentration) as an antiandrogen control. They assayed each

concentration of each test compound in at least four wells. After incubation for 24 hr, they removed the medium and washed the cells gently twice with phosphate-buffered saline. To measure luciferase activity, they added 25 μ L/well lysis buffer (Promega, Wallisellen, Switzerland), transferred these into a microtiter plate with a multichannel pipettor after 30 min, and read them on a luminometer (ML 1000; Dynex, Frankfurt/Main, Germany).

Test compounds, controls, and solvent. Each laboratory received four commercially available test compounds, all sent in a coded manner, plus two control compounds. Each laboratory used the following controls and test compounds: androgen control, DHT (Sigma-Aldrich,) [laboratory 2 used methyltrienolone (R1881; Perkin-Elmer)]; antiandrogen control, bicalutamide (Casodex, ICI 176,334), a gift from Zeneca Pharmaceuticals (Macclesfield, U.K.); compound 1, vinclozolin (Dr. Ehrenstorfer, Augsburg, Germany); compound 2, 4-androsten-3,17-dione (Sigma-Aldrich, Taufkirchen, Germany); compound 3, 17α-methyltestosterone (Fluka, Taufkirchen, Germany); and compound 4, tris-(4-chlorophenyl)methanol (TCPM) (Lancaster, Mühlheim, Germany).

Figure 1 illustrates the structures of the assayed substances. Each laboratory received the solid substances in amber glass vials with screw caps and Teflon and parafilm sealings. Each vial contained the amount required to obtain 4 mL of a 0.01 M stock solution (weighed on microscales with an uncertainty of \pm 0.05 mg). Participants stored all substances at 4°C. The vials of the androgen and antiandrogen controls contained a minimum of 30 mg, which was enough for 10 mL and 5 mL of 0.01 M stock solution, respectively. Laboratory 2 used the synthetic androgen R1881 instead of DHT as an androgen control; the PALM assay has a



Figure 2. Concentration–response curves of androgen control, 4-androsten-3,17-dione (compound 2), and 17 α -methyltestosterone (compound 3) for all four assays. (*A*) Assay 1. (*B*) Assay 2. (*C*) Assay 3. (*D*) Assay 4. Values represent means \pm SD of at least three independent experiments.

Tab	le	1.	Parti	cipating	labo	oratories	and	test	systems.	
-----	----	----	-------	----------	------	-----------	-----	------	----------	--

Laboratory	Cell type	End point	Exposure time (hr)	Statistical program ^a	Reference
1 (Tufts University School of Medicine)	MCF-7-AR1 human mammary adenocarcinoma cells expressing hAR	Cell number (determined with SRB assay)	120	Lotus 1-2-3 (logit)	Szelei et al. 1997
2 (INSERM U 439, Pathologie	Human prostate adenocarcinoma PC-3 cells	Luciferase activity	30	VBA program for	Térouanne et al. 2000
Moléculaire des Récepteurs	stably transfected with pSG ₅ -puro-hAR			EXCEL 5 ^b (log-probit)	
Nucléaires)	and pMMTV-neo-Luc				
3 (Danish Veterinary and Food Administration, Institute of Food Safety and Nutrition)	Chinese hamster ovary cells transiently transfected with pSVAR0 expression vector and MMTV-LUC reporter plasmid	Luciferase activity	20	SigmaPlot [Chapman (three parameter); four parameter logistic (for antagonists)]	Vinggaard et al. 1999
4 (University of Zürich, Institute of Pharmacology and Toxicology)	MDA-MB-453-KB2 human mammary carcinoma cells, endogenously expressing hAR and stably	Luciferase activity	24	GraphPad Prism (variable slope)	Wilson et al. 2002
67 67.	transfected with pMMTV-neo-Luc				

^aType of sigmoid regression. ^bProgram designed by J. Greve, Fraunhofer-Institute of Environmental Chemistry and Ecotoxicology, Schmallenberg, Germany.

lower sensitivity for DHT than for R1881 because of rapid metabolism of natural androgens in PC-3 cells, as described by Térouanne et al. (2000). These cells metabolize testosterone and DHT quickly, in less than 2 hr, whereas detection of luciferase activity requires at least 24 hr incubation with compounds. Castagnetta et al. (1994, 1997) have reported the ability of PC-3 cells to metabolize androgens in addition to a high expression of 17 β -hydroxysteroid dehydrogenase. All laboratories used high-purity DMSO for dissolving test and control compounds.

All test systems showed that bicalutamide and hydroxyflutamide act as potent androgen antagonists (Ma et al. 2003; Szelei et al. 1997; Térouanne et al. 2000; Vinggaard et al. 1999). Although hydroxyflutamide proved more potent than bicalutamide in assays 2 and 3 (Térouanne et al. 2000; Vinggaard et al. 1999), we selected bicalutamide as the antiandrogen control. We used the two steroidal androgens (compounds 2 and 3) because of their expected different potencies. Because Kelce et al. (1994) detected the two metabolites of vinclozolin, M1 and M2, as AR antagonists in vivo, we expected the results on vinclozolin would provide information on the metabolic capacities of the in vitro systems. The persistent, bioaccumulating environmental contaminant TCPM (DeBoer 1997) showed antiandrogenic activity in a cell proliferation assay (Körner et al. 1997, 2000). Therefore, because TCPM has high resistance to metabolic conversion, its use offered the opportunity to confirm these results.

Performance of experiments. The laboratories received instructions that each single experiment must contain a) a negative control; b) a solvent control; c) at least five appropriate concentrations of the androgen control DHT that encompass the whole range of the concentration-response curve; d) one concentration of DHT that gives nearly maximum response (0.1 nM) together with the following five concentrations of the antiandrogen control bicalutamide: 0.0001, 0.001, 0.01, 0.1, and 1 µM; e) 1 μ M bicalutamide; f) five (or more) concentrations of each test compound covering the whole range of the concentration-response curve (the highest concentration tested should equal 10 µM for compounds 1, 2 and 4, and 1 µM for compound 3); and g) five (or more) concentrations of each test compound together

with 0.1 nM DHT covering the whole range of the concentration-response curve.

Each laboratory tested the compounds in at least three independent experiments alone and together with 0.1 nM DHT.

Quantitative evaluation. Laboratory 1 used cell number as the end point. They used raw cell numbers for quantitative evaluation, but for graph representation, they normalized the experiments, set the maximum cell number to 1, and reported ratios in relationship to this.

The three reporter gene assays (assays 2–4) used luciferase activity relative to the hormonefree negative control as the basic end point. For quantitative evaluation of antiandrogenic activity, all laboratories set the luciferase activity of the androgen control (0.1 nM DHT or R1881) to 100%.

Each laboratory used different statistical software (Table 1) for performance of sigmoid regression of concentration–response curves and calculation of median effective concentration (EC_{50}) values for androgens and IC_{50} values for antiandrogens, respectively.

The androgenic potency of a test compound relative to the positive control (RAP) equals the quotient of the EC_{50} values of DHT (or R1881) and the compound:

> $RAP = EC_{50} (DHT \text{ or } R1881)$ + $EC_{50} (test compound).$

The antiandrogenic potency of a test compound relative to the positive control bicalutamide (RAAP) equals the quotient of the IC_{50} values of bicalutamide and the compound. The IC_{50} equals the concentration required for the compensation of half of the androgenic effect of 0.1 nM DHT (or R1881):

> RAAP = IC_{50} (bicalutamide) + IC_{50} (test compound).

RAP and RAAP have no dimension.

Results and Discussion

Overview. The solvent control (laboratories 1–3: 0.1% DMSO; laboratory 3 also tested 0.25 and 0.5%; laboratory 4: 0.2%) showed no significant effect in any assay; for example, in assay 2 the luciferase activity relative to the negative control was 0.97 \pm 0.37 (n = 3). All four laboratories correctly detected the androgenic activity of the testosterone metabolite

4-androsten-3,17-dione and of 17α -methyltestosterone. All laboratories calculated similar androgenic potencies relative to the positive control (RAP) for both test substances and, with the exception of 4-androsten-3,17-dione in assay 3, fell within one order of magnitude.

Each test system detected and quantified the antiandrogenic effect of the fungicide vinclozolin. Moreover, this compound showed slight androgenic activity at the highest tested concentrations with assays 2 and 3. Calculated antiandrogenic potencies relative to the positive control bicalutamide (RAAP) differed less than a factor of 10. Similarly, all assays found antiandrogenic activity for TCPM. This chemical showed cytotoxic effects at $\geq 1 \times 10^{-5}$ M in the assays of laboratories 1, 2, and 3. Nevertheless, all laboratories calculated RAAP values well within one order of magnitude.

Androgens. Table 2 summarizes data on the performance of the assays and the quantitative results for the androgen control DHT (or R1881). Assays 1 and 3 obtained EC_{50} values for DHT that were 3- and 10-fold lower, respectively, than those obtained with assay 4. In general, assay 3 resulted in the lowest EC_{50} values for all three androgens tested.

One might question the comparability of the results of assay 2 and those of the others because of the use of two different positive controls. However, the potency of the synthetic androgen R1881 in assay 2 fell within the range found for DHT in the other three test systems. In all other assays, R1881 gives comparable results, but we preferred DHT because it is a natural androgen.

Table 3 summarizes information on the performance of the tests and the quantitative results for 4-androsten-3,17-dione and 17 α -methyltestosterone. Figure 2 illustrates the concentration–response curves of both androgens and DHT (or R1881) for all four test systems. The coefficients of variation (CVs) of the EC₅₀ values of DHT (or R1881) fell within a reasonable range (12–57%) for all test systems (Table 2). All assays showed acceptable repeatability for the RAP values of the two androgenic test compounds, with CVs between 8 and 73% (Table 3).

The cell proliferation assay showed a 3-fold dynamic range. The three reporter gene assays revealed differences in the magnitude of the fold induction obtained after androgen exposure. In assays 2 and 3 the androgen con-

Table 2. Overview of the results for the androgen control DHT.

Laboratory	No. of assays	Replicates per assay	Concentrations tested	Lowest concentration (M)	Highest concentration (M)	EC ₅₀ mean (M)	SD (M)	CV (%)	Percent induction caused by 0.1 nM (mean ± SD)
1	5	4	15	1.0 × 10 ⁻¹²	1.0 × 10 ^{−8}	7.8 × 10 ⁻¹¹	9.3 × 10 ⁻¹²	12	69 ± 12
2	6	4	7	1.0 × 10 ⁻¹²	1.0 × 10 ⁻⁷	1.1 × 10 ^{-10a}	2.3 × 10 ⁻¹¹	22	53 ± 11
3	4	4	11	1.0 × 10 ⁻¹²	1.0 × 10 ⁻⁷	2.2 × 10 ⁻¹¹	6.7 × 10 ⁻¹²	30	85 ± 11
4	5	4	9	1.0×10^{-12}	1.0 × 10 ⁻⁸	2.1×10^{-10}	1.2 × 10 ⁻¹⁰	57	35 ± 8

^aLaboratory 2 used R1881.

trol induced a luciferase activity of > 20-fold of that of the negative control, whereas laboratory 4 obtained a 7-fold induction. However, none of the parameters measured covaried with the dynamic range.

With all assays 17α -methyltestosterone induced about the same maximal effect as the androgen control, thereby confirming this compound as a full AR agonist. Assays 1-4 produced EC₅₀ values of 4.3×10^{-10} M, 7.4 × 10^{-10} M, 3.3×10^{-11} M, and 5.3×10^{-10} M, respectively. 4-Androsten-3,17-dione also induced about the same maximal effect as the androgen control with assays 1, 2, and 3. This compound induced only 72% of the maximal effect of DHT with assay 4. Assays 1-4 produced EC₅₀ values of 5.9×10^{-8} M, $3.5 \times$ 10^{-8} M, 2.7×10^{-10} M, and 1.4×10^{-7} M, respectively. Compared with other test systems, assay 3 yielded lower EC₅₀ values of both compounds, as well as DHT. Because DHT and 17α -methyltestosterone varied in the same way, the resulting RAP value compared with those of the other assays. However, laboratory 3 found a 100-fold lower EC₅₀ for 4-androsten-3,17-dione than the other groups. As this compound was relatively more potent compared with DHT in this assay, a 30- to 60-fold higher RAP than that obtained in the other assays resulted. The possibility that the cells used in these assays may have different metabolic capacities may partly explain this discrepancy. Generally, synthetic androgens such as 17\alpha-methyltestosterone resist metabolism better than natural androgens. For example, Koh et al. (2001) recently demonstrated that PC-3 cells rapidly convert 4-androsten-3,17-dione predominantly to the inactive dehydroepiandrosterone and to a lesser extent to testosterone and subsequently to DHT.

Antiandrogens. Table 4 summarizes data on the performance of the assays and the quantitative results for the antiandrogen control bicalutamide. In all reporter gene assays, bicalutamide showed slight but measurable androgenic activity at the highest concentration of 1×10^{-6} M. However, compared with the activity induced by 0.1 nM DHT (or R1881), we found the androgenic effect of bicalutamide negligible (Figure 3). Laboratory 2 tested bicalutamide concentrations up to 1×10^{-5} M. They found the androgenic effect low compared with 0.1 nM R1881 but higher than that of 1×10^{-6} M. Laboratory 1 did not detect



Figure 3. Partial androgenic effects of antiandrogens in reporter gene assays. (*A*) Bicalutamide (mean \pm SD of three to five independent experiments); the effects of 0.1 nM DHT (or R1881) are also shown for comparison. (*B*) Vinclozolin (assay 2 and assay 3; mean \pm SD of three and four independent experiments, respectively). Figure 2B and 2C show the maximal effect of the androgen control.

Table 3. Overview of the results for the androgenic compounds 4-andros	en-3,17-dione (compound 2) and	17α-methyltestosterone	(compound 3)
--	--------------------------------	------------------------	--------------

Compound, laboratory	No. of assays	Concentrations tested	Lowest concentration (M)	Highest concentration (M)	EC ₅₀ mean (M)	SD (M)	RAP mean	SD	CV (%)
4-Androsten-3,17-dione									
1	4	5	1.0 × 10 ⁻⁹	1.0 × 10 ⁻⁵	5.9 × 10 ⁻⁸	2.1 × 10 ⁻⁸	1.4 × 10 ⁻³	5.4 × 10 ⁻⁴	38
2	3	6	1.0 × 10 ⁻¹⁰	1.0 × 10 ⁻⁵	3.5 × 10 ⁻⁸	7.9 × 10 ⁻⁹	3.1 × 10 ^{−3}	2.5 × 10 ⁻⁴	8
3	4	12	3.0×10^{-11}	6.0 × 10 ⁻⁸	2.7 × 10 ⁻¹⁰	6.4×10^{-11}	8.7 × 10 ⁻²	2.1 × 10 ⁻²	24
4	5	8	1.0 × 10 ⁻⁹	1.0 × 10 ⁻⁵	1.4 × 10 ⁻⁷	4.0 × 10 ⁻⁸	1.5 × 10 ^{−3}	7.7 × 10 ⁻⁴	50
17 α -Methyltestosterone									
1	4	5	1.0 × 10 ⁻¹⁰	1.0 × 10 ⁻⁶	4.3 × 10 ⁻¹⁰	1.8 × 10 ⁻¹⁰	0.23	0.16	70
2	3	6	1.0 × 10 ⁻¹¹	1.0 × 10 ⁻⁶	7.4 × 10 ⁻¹⁰	1.3 × 10 ⁻¹⁰	0.15	0.04	27
3	4	12	5.0 × 10 ⁻¹³	9.4 × 10 ⁻¹⁰	3.3 × 10 ⁻¹¹	2.2 × 10 ⁻¹¹	1.03	0.75	73
4	5	8	1.0×10^{-11}	1.0×10^{-6}	5.3 × 10 ⁻¹⁰	1.4 × 10 ⁻¹⁰	0.44	0.32	73

Table 4. Overview of the results for the antiandrogen control bicalutamide.

Laboratory	No. of assays	Replicates per assay	Concentrations together with 0.1 nM DHT ^a	Lowest concentration (M)	Highest concentration (M)	IC ₅₀ mean (M)	SD (M)	CV (%)
1	4	2	5	1.0 × 10 ⁻¹⁰	1.0 × 10 ⁻⁶	1.3 × 10 ⁻⁷	1.6 × 10 ⁻⁸	12
2	3	4	5	1.0 × 10 ⁻⁹	1.0 × 10 ⁻⁵	1.7 × 10 ^{−7}	2.7 × 10 ⁻⁸	16
3	3	4	10	1.0 × 10 ⁻¹⁰	5.0 × 10 ⁻⁶	2.4 × 10 ⁻⁷	1.0 × 10 ⁻⁷	42
4	5	4	8	1.0 × 10 ⁻¹⁰	1.0 × 10 ⁻⁶	3.8 × 10 ⁻⁷	3.2 × 10 ⁻⁷	85

^aLaboratory 2 used 0.1 nM R1881.

an androgen-like decrease of cell count for bicalutamide.

Figure 4 illustrates the concentration– response curves for all four assays of bicalutamide and the antiandrogenic chemicals vinclozolin and TCPM. Each laboratory derived the IC₅₀ values from the coincubation of cells with 0.1 nM DHT (or R1881) and different concentrations of antiandrogen, leading to a subsequent increasing inhibition of the androgenic effect of DHT (or R1881). A concentration of 1×10^{-6} M bicalutamide did not lead to complete inhibition of the androgenic effect in any of the assays. For calculation of IC₅₀ values, the laboratories extrapolated the concentration–response curves to maximal effect.



(compound 1), and TCPM (compound 4) for all four assays. (*A*) Assay 1: antiandrogens plus 0.1 nM DHT. (*B*) Assay 2: antiandrogens plus 0.1 nM R1881. (*C*) Assay 3: antiandrogens plus 0.1 nM DHT. (*D*) Assay 4: antiandrogens plus 0.1 nM DHT. Values represent mean ± SD of at least three independent experiments. Laboratories 2 and 3 disregarded the values measured for coincubation of androgen with 1×10^{-5} M bicalutamide and > 3.0×10^{-6} M vinclozolin, respectively, for IC₅₀ calculation because of the slight androgenic activity of these compounds at higher concentrations. All four laboratories calculated comparable IC₅₀ values, and the difference between the highest and the lowest value was a factor of 3.

Table 5 and Figure 4 summarize and illustrate the results on the activity of vinclozolin and TCPM. With the exception of assay 4, neither compound inhibited the androgenic effect of 0.1 nM DHT (or R1881) completely. Therefore, for calculation of IC₅₀ values, all laboratories extrapolated the concentration–response curves to maximal effect as for bicalutamide.

Table 2 shows the percentage of maximal effect induced by 0.1 nM of the androgen control. Because the response to antiandrogens depends on the magnitude of this response (the stronger the effect of the androgen control, the greater competition of the antiandrogen required), the different experimental conditions used in the assays may explain some of the observed differences in IC₅₀ values of the antiandrogens. Thus, laboratories 1-4 tested the antiandrogens at 69, 53, 85, and 35% induction of maximum androgen response, respectively, resulting in the most favorable experimental conditions in terms of sensitivity for detecting antiandrogens for assay 4 and the least favorable conditions for assay 3.

In assays 2 and 3, vinclozolin showed slight and moderate androgenic activity at the highest concentrations, respectively (Figure 3B). These results agree with results obtained by Wong et al. (1995), who found agonistic activity of the vinclozolin metabolite M2 at 10 μ M, and by Nellemann et al. (2003), who observed agonistic activity of vinclozolin itself at concentrations \geq 3 μ M. TCPM showed cytotoxic effects \geq 10 μ M in test systems 1, 2, and 3 (Figure 4).

For both chemicals, IC_{50} values as well as calculated RAAP values remained similar among all assays, although the percentage of the maximal effect induced by 0.1 nM DHT (or R1881) differed (Table 2). The highest and

Table 5. Overview of the results for antiandrogenic compounds vinclozolin (compound 1) and TCPM (compound 4)

Compound, laboratory	No. of assays	Concentrations tested	Lowest concentration (M)	Highest concentration (M)	IC ₅₀ mean (M)	SD (M)	RAAP mean	SD	CV (%)		
Vinclozolin											
1	4	5	1.0 × 10 ⁻⁹	1.0 × 10 ^{−5}	2.1 × 10 ⁻⁷	1.3 × 10 ⁻⁷	1.0	0.31	31		
2	3	8	1.0 × 10 ⁻⁹	1.0 × 10 ^{−5}	4.7 × 10 ⁻⁷	1.3 × 10 ⁻⁷	0.37	0.07	18		
3	4	12	2.5 × 10 ⁻⁸	5.0 × 10 ⁻⁵	3.7 × 10 ⁻⁷	1.2 × 10 ⁻⁷	1.6	0.5	32		
4	5	8	1.0 × 10 ⁻¹⁰	1.0 × 10 ⁻⁵	1.1 × 10 ⁻⁷	3.2 × 10 ⁻⁸	3.7	3.1	83		
TCPM											
1	3	5	1.0 × 10 ⁻⁹	1.0 × 10 ⁻⁵	5.7 × 10 ⁻⁷	1.4 × 10 ⁻⁷	0.22	0.04	16		
2	3	7	1.0 × 10 ⁻⁹	1.0 × 10 ⁻⁵	8.0 × 10 ⁻⁷	1.8 × 10 ⁻⁷	0.22	0.04	20		
3	4	12	2.5 × 10 ⁻⁸	5.0 × 10 ⁻⁵	3.1 × 10 ⁻⁶	1.4 × 10 ⁻⁶	0.086	0.03	36		
4	5	8	1.0 × 10 ⁻⁹	1.0 × 10 ⁻⁵	1.1 × 10 ⁻⁶	7.6 × 10 ⁻⁸	0.37	0.33	90		

the lowest IC_{50} differed by a factor of 3 for vinclozolin and a factor of 10 for TCPM. Vinclozolin and bicalutamide had similar antiandrogenic potencies, whereas TCPM had about one fifth the potency of these compounds. Two proliferation assays using AR-positive human mammary carcinoma cell lines have demonstrated the antiandrogenic effect of TCPM (Körner et al. 1997, unpublished data). The fact that all four assays unambiguously confirmed the antiandrogenic properties of TCPM implicates environmental relevance, because this compound is a ubiquitous and highly bioaccumulating chemical and we know little about its sources and toxicologic properties (DeBoer 1997).

The test systems are applicable to other environmental antiandrogens such as p,p '-DDE. In assay 4, DDE reduced DHTinduced luciferase activity with an IC₅₀ value of 2.8 ± 0.8 × 10⁻⁶ M (n = 5), whereas cytotoxicity was observed at 1 × 10⁻⁴ M. In assay 3, p,p'-DDE had an IC₅₀ of 1.1 × 10⁻⁶ M (Vinggaard et al., unpublished data).

When testing for agonists and antagonists of AR, it is important to consider any cytotoxicity of test compounds to avoid classification of false positives. In assay 1 cytotoxicity is assessed by inspecting the appearance of the cells using an inverted microscope before the cells are fixed. Cytotoxicity is recognized by the presence of floating (dead) cells and the presence of cytoplasmic vacuoles in those still attached to the substrate. In this assay, agonists inhibit cell proliferation and antagonists overcome this inhibition; hence, evaluation of cytotoxicity can also be made by testing whether the inhibitory response observed in the presence of a putative agonist is totally reversed by excess antiandrogen. If the antagonist does not reverse the low cell yield, the effect is considered cytotoxic. In the other three assays, agonists increase and antagonists decrease the expression of the reporter gene. In the experimental conditions of this study, in which antiandrogens were tested for both agonistic and antagonistic effects, any cytotoxicity in the reporter gene assays was indirectly revealed in the agonism test by an inhibition of the transcriptional/translational process (i.e., a decreased luciferase activity). When only antagonism is tested, a very specific cytotoxicity test has been developed for assays based on transient transfections. This method involves transfection of cells with a constitutive active AR expression vector that lacks the ligand-binding domain of the receptor (Kelce et al. 1995; Vinggaard et al. 2002). Cytotoxicity is measured directly at the transcriptional/translational level with this method.

All assays showed comparable repeatability for agonistic and antagonistic androgenic activity. The CVs of the IC_{50} values of bicalutamide ranged from 12 to 85% (Table 4), and those of the RAAP values ranged from 16 to 90% (Table 5). Comparing CVs between the four test systems, one should take into account that the different number of independent experiments (three to five) performed with the assays influence the CV values.

The various statistical programs and types of regression used for calculation of EC_{50} and IC_{50} values had little contribution to the differences of the quantitative results between the four assays.

Conclusions

We compared four different cellular in vitro assays for the detection of AR-mediated agonistic and antagonistic effects of chemicals in an interlaboratory study. All four test systems produced comparable quantitative results for two androgens and two antiandrogens. The EC50 values calculated for the androgens and the resulting androgenic potencies relative to the positive control differed by less than a factor of 10 between the assays, with one exception. We saw this exception in assay 3, which also shows a general greater sensitivity toward detection of androgens. For the antiandrogenic chemicals, differences of IC50 values and calculated relative antiandrogenic potencies between the assays remained well within one order of magnitude.

The CVs we obtained for the different EC50 and IC50 values stayed generally within the same range, and we found no obvious differences in performance between assays. Discrepancies among IC₅₀ values may stem from the fact that all laboratories tested the antiandrogens with a fixed concentration of 0.1 nM androgen, giving rise to diverging percentages of maximum induction. Different metabolic capacities of each cell line for each compound may also play a role. Differences in RAP may reflect differences in metabolism of the test compound, DHT, or both. This could explain why 4-androsten-3,17-dione showed considerably greater differences than the other three test compounds and the positive controls.

In summary, all four cellular in vitro assays proved sensitive screening tools to detect and quantify AR-mediated androgenic and antiandrogenic effects of these chemicals with reasonable accuracy. We did not design this experiment to test which assay was best for screening purposes, because a number of other factors must be taken into account. All laboratories have reported specificities of the assays in the original publications, which should be considered when deciding which assay to use for screening purposes. In choosing a test system to use, the equipment of the laboratory, specific background and experience of the staff, and cost-effectiveness must also be evaluated.

REFERENCES

- Andersen HR, Andersson AM, Arnold SF, Autrup H, Barfoed M, Beresford NA, et al. 1999. Comparison of short-term estrogenicity tests for identification of hormone-disrupting chemicals. Environ Health Perspect 107(suppl 1):89–108.
- Breithofer A, Graumann K, Scicchitano MS, Karathanasis SK, Butt TR, Jungbauer A. 1998. Regulation of human estrogen receptor by phytoestrogens in yeast and human cells. J Steroid Biochem Mol Biol 67:421–429.
- Brzozowski AM, Pike ACW, Dauter Z, Hubbard RE, Bonn T, Engström O, et al. 1997. Molecular basis of agonism and antagonism in the gestrogen receptor. Nature 389:753–758.
- Castagnetta LA, Carruba G, Traina A, Granata OM, Markus M, Pavone-Macaluso M, et al. 1997. Expression of different 17 beta-hydroxysteroid dehydrogenase types and their activities in human prostate cancer cells. Endocrinology 138:4876–4882.
- Castagnetta LA, Granata OM, Polito L, Blasi L, Cannella S, Carruba G. 1994. Different conversion metabolic rates of testosterone are associated to hormone-sensitive status and response of human prostate cancer cells. J Steroid Biochem Mol Biol 49:351–357.
- Chang C, Liao S 1987. Topographic recognition of cyclic hydrocarbons and related compounds by receptors for androgens, estrogens, and glucocorticoids. J Steroid Biochem Mol Biol 27:123–131.
- Christiansen LB, Pedersen KL, Pedersen SN, Korsgaard B, Bjerregaard P. 2000. *In vivo* comparison of xenoestrogens using rainbow trout vitellogenin induction as a screening system. Environ Toxicol Chem 19:1867–1874.
- Colborn T, vom Saal FS, Soto AM. 1993. Developmental effects of endocrine-disrupting chemicals in wildlife and humans. Environ Health Perspect 101:378–384.
- Cook JC, Mullin LS, Frame SR, Biegel LB. 1993. Investigation of a mechanism for Leydig cell tumorigenesis by linuron in rats. Toxicol Appl Pharmacol 119:195–204.
- Dalton JT, Mukherjee A, Zhu Z, Kirkovsky L, Miller DD. 1998. Discovery of nonsteroidal androgens. Biochem Biophys Res Commun 244:1–4.
- DeBoer J. 1997. Environmental distribution and toxicity of tris(4chlorophenyl)methanol and tris(4-chlorophenol)methane. Rev Environ Contamin Toxicol 150:95–106.
- Dodds EC, Lawson W. 1936. Synthetic estrogenic agents without the phenanthrene nucleus. Nature 137:996.
- Donohoe RM, Curtis LR. 1996. Estrogenic activity of chlordecone, o,p⁻DDT and o,p⁻DDE in juvenile rainbow trout: induction of vitellogenesis and interaction with hepatic estrogen binding sites. Aquat Toxicol 36:31–52.
- Eil C, Nisula BC. 1990. The binding properties of pyrethroids to human skin fibroblast androgen receptors and to sex hormone binding globulin. J Steroid Biochem Mol Biol 35:409–414.
- Fail PA, Pearce SW, Anderson SA, Tyl RW, Gray LE Jr. 1995. Endocrine and reproductive toxicity of vinclozolin (vin) in male Long-Evans Hooded rats. Fundam Appl Toxicol 15:293–298.
- Fang H, Tong W, Perkins R, Soto AM, Prechtl NV, Sheehan DM. 2000. Quantitative comparisons of *in vitro* assays for estrogenic activities. Environ Health Perspect 108:723–729.
- Gimeno S, Gerritsen A, Bowmer T, Komen H. 1996. Feminization of male carp. Nature 384:221–222.
- Gray LE Jr, Ostby JS, Kelce WR. 1994. Developmental effects of an environmental antiandrogen: the fungicide vinclozolin alters sex differentiation of the male rat. Toxicol Appl Pharmacol 129:46–52.
- Jobling S, Reynolds T, White R, Parker MG, Sumpter JP. 1995. A variety of environmentally persistent chemicals, including some phthlate plasticizers, are weakly estrogenic. Environ Health Perspect 103:582–587.
- Jobling S, Sheahan D, Osborne JA, Matthiessen P, Sumpter JP. 1996. Inhibition of testicular growth in rainbow trout (*Oncorhynchus mykiss*) exposed to estrogenic alkylphenolic chemicals. Environ Toxicol Chem 15:194–202.
- Kelce WR, Monosson E, Gamcsik MP, Laws SC, Gray LE Jr. 1994. Environmental hormone disruptors: evidence that vinclozolin developmental toxicity is mediated by antiandrogenic metabolites. Toxicol Appl Pharmacol 126:276–285.
- Kelce WR, Stone CR, Laws SC, Gray LE, Kemppainen JA, Wilson EM. 1995. Persistent DDT metabolite p,p⁻DDE is a potent androgen receptor antagonist. Nature 375:581–585.
- Klotz DM, Beckman BS, Hill SM, McLachlan J, Walters MR, Arnold SF. 1996. Identification of environmental chemicals

with estrogenic activity using a combination of *in vitro* assays. Environ Health Perspect 104:1084–1089.

- Koh E, Kanaya J, Namiki M. 2001. Adrenal steroids in human prostatic cancer cell lines. Arch Androl 46:117–125.
- Körner W, Hanf V, Oettling G, Schuller W, Hagenmaier H. 1997. Effects of tris(4-chlorophenol)methanol on proliferation of human breast cancer cells. Organohalogen Compounds 34:364–369.
- Körner W, Hanf V, Schuller W, Bartsch H, Zwirner M, Hagenmaier H. 1998. Validation and application of a rapid in vivo assay for assessing the estrogenic potency of halogenated phenolic chemicals. Chemosphere 37:2395–2407.
- Ma R, Cotton B, Lichtensteiger W, Schlumpf M. 2003. UV filters with antagonistic action at androgen receptors in the MDA-kb2 cell transcriptional-activational assay. Toxicol Sci 74:43–50.
- Miksicek R. 1994. Interaction of naturally occurring nonsteroidal estrogens with expressed recombinant human estrogen receptor. J Steroid Biochem Mol Biol 49:153–160.
- Miller D, Wheals BB, Beresford N, Sumpter J. 2001. Estrogenic activity of phenolic additives determined by an *in vitro* yeast bioassay. Environ Health Perspect 109:133–138.
- Milligan SR, Balasubramanian AV, Kalita JC. 1998. Relative potency of xenobiotic estrogens in an acute *in vivo* mammalian assay. Environ Health Perspect 106:23–26.

Nagel SC, vom Saal FS, Thayer KA, Dhar MG, Boechler M, Welshons WV. 1997. Relative binding affinity-serum modified access (RBA-SMA) assay predicts the relative *in vivo* bioactivity of the xenoestrogens bisphenol A and octylphenol. Environ Health Perspect 105:70–76.

- Nellemann C, Dalgaard M, Lam HR, Vinggaard AM. 2003. The combined effects of vinclozolin and procymidone do not deviate from expected additivity *in vitro* and *in vivo*. Toxicol Sci 71:251–262.
- Paris F, Balaguer P, Térouanne B, Servant N, Lacoste C, Cravedi J, et al. 2002. Phenylphenols, biphenols, bisphenol-A and 4-tert-octylphenol exhibit alpha and beta estrogen activities and antiandrogen activity in reporter cell lines. Mol Cell Endocrinol 193:43–49.
- Schlumpf M, Cotton B, Conscience M, Haller V, Steinmann B, Lichtensteiger W. 2001. *In vitro* and *in vivo* estrogenicity of UV screens. Environ Health Perspect 109:239–244.
- Soto AM, Justicia H, Wray JW, Sonnenschein C. 1991. p-Nonylphenol: an estrogenic xenobiotic released from "modified" polystyrene. Environ Health Perspect 92:167–173.
- Soto AM, Sonnenschein C, Chung KL, Fernandez MF, Olea N, Olea-Serrano MF. 1995. The E-SCREEN assay as a tool to identify estrogens: an update on estrogenic environmental pollutants. Environ Health Perspect 103:113–122.
- Soto AM, Sonnenschein C, Prechtl NV, Weill BC, Olea N. 1999. Methods to screen estrogen-agonists and antagonists. J Med Food 2:139–142.
- Szelei J, Jimenez J, Soto AM, Luizzi MF, Sonnenschein C. 1997. Androgen-induced inhibition of proliferation in human breast cancer MCF7 cells transfected with androgen receptor. Endocrinology 138:1406–1412.

- Térouanne B, Tahiri B, Georget V, Belon C, Poujol N, Avances C, et al. 2000. A stable prostatic bioluminescent cell line to investigate androgen and antiandrogen effects. Mol Cell Endocrinol 160:39–49.
- Tyler CR, Beresford N, Van der Woning M, Sumpter JP, Thorpe K. 2000. Metabolism and environmental degradation of pyrethroid insecticides produce compounds with endocrine activities. Environ Toxicol Chem 19:801–809.
- Vinggaard AM, Hnida C, Larsen JC. 2000. Environmental polycyclic aromatic hydrocarbons affect androgen receptor activation in vitro. Toxicology 145:159–169.
- Vinggaard AM, Joergensen ECB, Larsen JC. 1999. Rapid and sensitive reporter gene assays for detection of antiandrogenic and estrogenic effects of environmental chemicals. Toxicol Appl Pharmacol 155:150–160.
- Vinggaard AM, Nellemann C, Dalgaard M, Jørgensen EB, Andersen HR. 2002. Antiandrogenic effects in vitro and in vivo of the fungicide prochloraz. Toxicol Sci 69:344–353.
- Wilson VS, Bobseine K, Lambright CR, Gray LE Jr. 2002. A novel cell line, MDA-kb2, that stably expresses an androgenand glucocorticoid-responsive reporter for the detection of hormone receptor agonists and antagonists. Toxicol Sci 66:69–81.
- Wong C, Kelce WR, Sar M, Wilson EM. 1995. Androgen receptor antagonist versus agonist activities of the fungicide vinclozolin relative to hydroxyflutamide. J Biol Chem 270:19998–20003.