

# Prepared in Cooperation with the City of Aurora, Colorado, Utilities Department

# Selenium and Other Elements in Water and Adjacent Rock and Sediment of Toll Gate Creek, Aurora, Arapahoe County, Colorado, December 2003 through March 2004



Scientific Investigations Report 2007–5018

U.S. Department of the Interior U.S. Geological Survey

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By J.R. Herring and Katherine Walton-Day

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# **Conversion Factors**

#### SI to Inch/Pound

Multiply	Ву	To obtain
	Length	
meter (m)	3.281	foot (ft)
centimeter (cm)	0.3937	inch
kilometer (km)	0.6214	mile (mi)
	Volume	
liter (L)	0.2642	gallon (gal)
cubic meter (m <sup>3</sup> )	264.2	gallon (gal)
cubic meter (m <sup>3</sup> )	35.31	cubic foot (ft <sup>3</sup> )
	Flow rate	
cubic meter per second (m <sup>3</sup> /s)	35.31	cubic feet per second (ft <sup>3</sup> /s)
liter per second (L/s)	15.85	gallon per minute (gal/min)
cubic meter per day (m <sup>3</sup> /d)	264.2	gallon per day (gal/d)
	Mass	
gram (g)	0.03527	ounce
kilogram (kg)	2.205	pound, avoirdupois (lb)

Temperature in degrees Celsius (°C) may be converted to degrees Fahrenheit (°F) as follows:

°F=(1.8×°C)+32

Temperature in degrees Fahrenheit (°F) may be converted to degrees Celsius (°C) as follows:

### °C=(°F-32)/1.8

Horizontal coordinate information is referenced to the North American Datum of 1983 (NAD 83)

Specific conductance is given in microsiemens per centimeter at 25 degrees Celsius ( $\mu$ S/cm at 25 °C).

Concentrations of chemical constituents in water are given either in milligrams per liter (mg/L) or micrograms per liter ( $\mu$ g/L).

# **Additional Abbreviations**

DIW	Deionized water
EDXRF	Energy dispersive X-ray fluorescence
EPA	U.S. Environmental Protection Agency
ETGC	East Toll Gate Creek
g/d	gram per day
HNO <sub>3</sub>	Ultrex nitric acid
ICP-MS	inductively coupled plasma mass spectrometry
kV	kilovolt
LLD	lower limit of detection
L/s	liter per second
mA	milliampere
MCL	maximum contaminant level
mg/Kg	milligram per kilogram
mm	millimeter
mL	milliliter
MΩ	meg-ohm
mS/cm	millisiemens per centimeter at 25 degrees Celsius
ΝΑΨΩΑ	National Water Quality Assessment
NGD	National Geochemical Database
NIST	National Institute of Standards and Technology
ORP	oxidation-reduction potential
ppm	part per million
Se	selenium
TDS	total dissolved solids
TGC	Toll Gate Creek
TMDL	total maximum daily load
μm	micrometer
USGS	U.S. Geological Survey
WTGC	West Toll Gate Creek
XRD	X-ray diffraction analyses

# Selenium and Other Elements in Water and Adjacent Rock and Sediment of Toll Gate Creek, Aurora, Arapahoe County, Colorado, December 2003 through March 2004

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# Abstract

Streamwater and solid samples (rock, unconsolidated sediment, stream sediment, and efflorescent material) in the Toll Gate Creek watershed, Colorado, were collected and analyzed for major and trace elements to determine trace-element concentrations and stream loads from December 2003 through March 2004, a period of seasonally low flow. Special emphasis was given to selenium (Se) concentrations because historic Se concentrations exceeded current (2004) stream standards. The goal of the project was to assess the distribution of Se concentration and loads in Toll Gate Creek and to determine the potential for rock and unconsolidated sediment in the basin to be sources of Se to the streamwater.

Streamwater samples and discharge measurements were collected during December 2003 and March 2004 along Toll Gate Creek and its two primary tributaries - West Toll Gate Creek and East Toll Gate Creek. During both sampling periods, discharge ranged from 2.5 liters per second to 138 liters per second in the watershed. Discharge was greater in March 2004 than December 2003, but both periods represent low flow in Toll Gate Creek, and results of this study should not be extended to periods of higher flow. Discharge decreased moving downstream in East Toll Gate Creek but increased moving downstream along West Toll Gate Creek and the main stem of Toll Gate Creek, indicating that these two streams gain flow from ground water. Se concentrations in streamwater samples ranged from 7 to 70 micrograms per liter, were elevated in the upstream-most samples, and were greater than the State stream standard of 4.6 micrograms per liter. Se loads ranged from 6 grams per day to 250 grams per day, decreased in a downstream direction along East Toll Gate Creek, and increased in a downstream direction along West Toll Gate Creek and Toll Gate Creek. The largest Se-load increases occurred between two sampling locations on West Toll Gate Creek during both sampling periods and between the two sampling locations on the main stem of Toll Gate Creek during the December 2003 sampling. These load increases may indicate that sources of Se exist between these two locations; however, Se loading along West Toll Gate Creek and Toll Gate Creek primarily was characterized by gradual downstream increases in load.

Linear regressions between Se load and discharge for both sampling periods had large, significant values of  $r^2$  ( $r^2 > 0.96$ , p < 0.0001) because increases in Se load (per unit of flow increase) were generally constant. This relation is evidence for a constant addition of water having a relatively constant Se concentration over much of the length of Toll Gate Creek, a result which is consistent with a ground-water source for the Se loads.

Rock outcroppings along the stream were highly weathered, and Se concentrations in rock and other solid samples ranged from below detection (1 part per million) to 25 parts per million. One sample of efflorescence (a surface encrustation produced by evaporation) had the greatest selenium concentration of all solid samples, was composed of thenardite (sodium sulfate), gypsum (calcium sulfate) and minor halite (sodium chloride), and released all of its Se during a 30minute water-leaching procedure. Calculations indicate there was an insufficient amount of this material present throughout the watershed to account for the observed Se load in the stream. However, this material likely indicates zones of ground-water discharge that contain Se.

This report did not identify an unequivocal source of Se in Toll Gate Creek. However, multiple lines of evidence indicate that ground-water discharge supplies Se to Toll Gate Creek: (1) the occurrence of elevated Se concentrations in the stream throughout the watershed and in the headwater regions, upstream from industrial sources; (2) the progressive increase in Se loads moving downstream, which indicates a continuous input of Se along the stream rather than input from point sources; (3) the occurrence of efflorescence deposits in several locations along the stream channel that contain elevated concentrations of readily soluble Se and probably represent ground-water discharge zones; and (4) the occurrence of rock types within the Toll Gate Creek watershed that may contain elevated concentrations of Se (ash layers and lignite deposits). Ground water likely acquires the Se through water-rock interaction with the aquifers underlying Toll Gate Creek. Shallow ground water containing soluble Se discharges to Toll Gate Creek.

# Introduction

In response to a request by the City of Aurora, Colorado, the U.S. Geological Survey (USGS) began a series of geologic and water-quality studies from December 2003 through March 2004 to study water quality and nearby rock and unconsolidated sediment of Toll Gate Creek (TGC) (fig. 1). The request arose because of recent concerns about elevated selenium (Se) concentrations in TGC. Historical data indicate that elevated Se concentrations occur in surface water in East Toll Gate Creek (ETGC), West Toll Gate Creek (WTGC), and TGC (Brown and Caldwell and Chadwick Ecological Consultants, 2003) which are tributary to Sand Creek (fig. 1). Se concentrations occasionally exceed the national primary drinking water standard Maximum Contaminant Level (MCL) of 50 micrograms per liter ( $\mu$ g/L), established by the U.S. Environmental Protection Agency (EPA) (U.S. Environmental Protection Agency, 2002) and greatly exceed the State stream standard 4.6 µg/L (Colorado Department of Public Health and Environment, 2005). Neither Sand Creek nor TGC are used for drinking-water supply. Sand Creek, TGC, WTGC, and ETGC are on the most recent 303(d) list for Se. Therefore, a Total Maximum Daily Load (TMDL) may need to be developed. The City of Aurora is currently (2007) working on development of an alternative standard based on the assumption that the source of Se is naturally occurring and that remediation of the Se will result in widespread social or environmental impacts. This study was conducted to assess whether elevated Se concentrations in TGC (tributary to Sand Creek) result from the natural interaction of rock solids and water within the drainage area. The information provided will help with formulation of a new standard for Sand Creek.

## **Purpose and scope**

The purpose of this report is to describe and interpret results from surveys of streamwater quality and discharge and geochemistry of solid samples collected from the TGC watershed from December 2003 to March 2004. Mass loads for Se were calculated from streamflow and water-quality data collected at as many as 13 stream sites. Twenty-one rock, unconsolidated sediment, and stream-sediment samples, and samples of efflorescent salts were collected from areas within and adjacent to TGC and were retrieved from archived drillcore samples from wells drilled within the watershed. Water and solid samples and some digestions of solid samples were analyzed for concentrations of Se and other trace and major elements. This report provides analysis and interpretation of major and trace-element concentrations and Se loads in the streamwater and of major and trace-element concentrations in rocks and unconsolidated sediment that might be contributing Se or other trace elements to the streamwater.

### Physiographic and Geologic Setting

Land use is varied within the drainage basin of TGC. The stream has its headwaters in rolling prairie, urbanized within the past 15 years with residential housing, and flows northwest toward the confluence with Sand Creek (fig. 1). ETGC flows through the southwest corner of Buckley Air National Guard Base but was dry upstream from ETGS-2 during the December 2003 and March 2004 sampling events. WTGC is diverted around the west side of Quincy Reservoir and has no direct interaction with the reservoir water. In addition, WTGC can accept water from the overflow spillway of Cherry Creek Reservoir, but this is not known to have happened within the last decade. In its lower reaches, TGC flows next to the Fitzsimmons Army Medical Center. The stream is perennial sustaining base flow throughout the year and has sharp peaks in flow in response to storm events (C.R. Bossong, U.S. Geological Survey, oral commun., 2007).

Geology of the TGC watershed is characterized by Quaternary unconsolidated surficial materials as much as 10 meters (m) thick (Robson, 1996) overlying the Cretaceous to early Tertiary-aged Denver Formation bedrock. Surficial materials include alluvial, colluvial, and eolian deposits. The alluvial deposits consist of gravel, sand, silt, and clay. Colluvial deposits consist of the same components and commonly are derived from the underlying bedrock. Eolian deposits consist of fine- to medium-grained windblown sand. The bedrock underlying the unconsolidated sediments is the Denver Formation, which consists of weakly consolidated mudstone and occasional beds of consolidated siltstone or sandstone, localized seams of low-rank coal, and volcanic ash layers. Volcanic ash can contain Se; however, there is no documentation of Se in the Denver Formation volcanic-ash layers.

### **Previous Work**

A report prepared for the annual Selenium Stakeholders Meeting, April 2003, summarized Se concentrations, loads, and other water-quality parameters for various locations on TGC, nearby sites, Sand Creek, and the South Platte River for various dates 1995 to 2002 (Brown and Caldwell and Chadwick Ecological Consultants, 2003). These data had precision of +/- 7 percent based on analysis of three replicate sample pairs, but contained no assessment of bias. The Se concentration and streamflow data from four locations sampled during 2002 were summarized (table 1) to depict Se loads that originated in TGC and contributed to Sand Creek and eventually the South Platte River. The monthly data collected from January to November 2002 enabled comparison of Se loads between WTGC and ETGC and between TGC and Sand Creek (SWA, fig.1). These data indicate the following general conclusions. ETGC Se loads ranged from 1.0 to 14 percent and averaged 8 percent of the Se load in WTGC (average of 6 values); the Se load in TGC continued to increase and approximately doubled by station TG-1, downstream from



Figure 1. Location of Toll Gate Creek drainage area, water and solid sampling locations, and locations of historic data collection.

#### Table 1. Historical Data on Dissolved Selenium Concentrations and Loads in Toll Gate Creek and comparison with Sand Creek.

[Data from Brown and Caldwell and Chadwick Ecological Consultants, 2003; date in month/day/year; µg/L, micrograms per liter, L/s, liters per second; g/d, grams per day; - indicates no data ; \* indicates value eliminated because of inconsistency in reported discharge; WTG-2, West Toll Gate Creek station 2; ETG-2, East Toll Gate Creek station 2; TG-1, Toll Gate Creek station 1; SWA, Sand Creek sampling location]

Component/Location	Average, all dates	1/9/2002	2/6/2002	3/6/2002	4/3/2002	5/1/2002	6/5/2002	7/3/2002	8/6/2002	9/4/2002	10/2/2002	11/6/2002
				Diss	olved Se, µg	/L						
WTG-2-at Chambers	13	22	22	9	14	11	5	12	10	19	2	15
ETG-2-at Chambers	7	13	13	10	12	4	5	2	0.5	_	1	-
TG-1-at 6th Ave.	14	25	26	12	18	15	5	11	5	16	2	16
SWA-Sand Ck. At Peoria	12	21	19	10	13	9	6	8	10	14	4	13
				Di	scharge, L/s							
WTG-2-at Chambers	113	33	45	253	65	-	465	9	20	3	220	18
ETG-2-at Chambers	12	5	7	33	1	-	17	0	0	0	55	0
TG-1-at 6th Ave.	186	65	103	301	102	-	594	66	91	239	458	40
SWA–Sand Ck. At Peoria	369	209	200	578	525	-	752	86	133	*	740	101
					Load, g/d							
WTG-2-at Chambers	72	62.7	85.5	197	78.6	-	201	9.3	17.3	4.92	38.0	23.3
ETG-2-at Chambers	9	5.6	7.8	28.5	1.0	-	7.3	_	_	-	4.8	-
TG-1-at 6th Ave.	167	140	231	312	159	-	257	62.7	39.3	330	79.1	55.2
SWA–Sand Ck. At Peoria	273	379	328	500	590	_	390	59.4	115	*	256	113
Average percent load ETGC compared to WTGC	8	9	9	14	1	-	4	-	-	-	13	-
Average percentage of selenium load of TGC compared to combined load with Sand Creek	53	37	70	62	27	_	66	105	34	_	31	49

the confluence of ETGC and WTGC and upstream from the confluence with Sand Creek (table 1). The main stem of TGC (as estimated at TG-1) contributed an average of 53 percent (9 values), albeit with high scatter, of the combined Se load of TGC and Sand Creek (as estimated at station SWA, down-stream from the confluence of Sand Creek and TGC (table 1). For the combined east and west reaches of TGC at 6<sup>th</sup> Avenue, there was no coherent relation of load versus discharge. These historical data indicate that: elevated Se concentrations have existed in TGC since at least 2002; WTGC carried a greater Se load than ETGC; and the Se load of TGC accounted for approximately one-half of the load in Sand Creek downstream from the confluence with TGC.

Brown and Chadwick and Caldwell Ecological Associates (2003) also provided some measurements of Se in the bottom sediments of TGC that were analyzed in 2001 and 2002. For the 2002 data, the analyte was listed as "total Se (3050)" which presumably refers to EPA analytical method 3050 B which entails acid digestion of sediments, sludges, and soils with analysis of Se by graphite furnace atomic adsorption spectrometry, or inductively coupled plasma-mass spectrometry (U.S. Environmental Protection Agency, 1996; U.S. Environmental Protection Agency, variously dated). However, in Brown and Chadwick and Caldwell Ecological Associates (2003), the analytical method is not specified, and no qualityassurance/quality-control data were provided. The sediment Se concentrations mostly were less than 0.5 milligram per kilogram (mg/Kg or parts per million (ppm)). However, Se concentrations of 1.8 and 6.2 mg/Kg were reported for WTGC at Chambers Road and ETGC at Chambers Road, respectively.

This study focuses on assessing Se concentrations and loads in TGC and on assessing Se concentration in geologic material near TGC. Analysis of solids and water focused primarily on Se, but the analytical techniques that were used also provide concentrations of several other trace elements and major elements. The goal of the study was to assess the distribution of Se concentration and loads in streamwater and to determine the potential for rock and unconsolidated sediment in TGC to contribute Se to streamwater.

### **Acknowledgments**

Rhonda Driscoll provided assistance in samples preparation and X-ray diffraction analysis. Dave Siems and Tammy Hannah are acknowledged for Energy Dispersive X-Ray Fluorescence (EDXRF) analysis of the solid samples. Janel Servis provided field assistance, and Paul Lamothe furnished timely analysis of water samples. Suzanne Paschke and Kip Bossong kindly provided reviews of the report. Carma San Juan and John Horton assisted with preparation of figures.

# Water Sampling and Analytical Methods

Water samples and discharge measurements were collected in the TGC watershed at 10 sites during December 2003 and 13 sites during March 2004. Samples were analyzed to determine concentrations of Se, major elements, and other trace elements. A few whole-water samples were collected and digested to help assess the contribution of suspended sediment to element loading in the TGC watershed. Se loads were constructed as the product of Se concentration and discharge.

## **Sampling Locations**

Water samples were collected and discharge measurements were made December 16 and 17, 2003 and March 2, 2004. Water-sampling sites on TGC, WTGC, and ETGC were designated by alphanumeric characters where the initial abbreviation represents each stream or location name (Sand Creek stream, SCS-; Toll Gate Creek stream, TGS-; West Toll Gate Creek stream, WTGS-; East Toll Gate Creek stream, ETGS-; outfall sample, OFS-) and the numeric designations increased moving upstream along each stream. The December 2003 sampling effort included sampling 10 sites on December 16, and resampling all 10 sites the following day to assess daily variability in Se concentration. In addition, sequential replicates were collected at three sites on December 16, 2003 (SCS-1, TGS-1, WTGS-2) and one site on December 17, 2003 (WTGS-3) to assess short-term environmental variability. All the same sites were sampled during the December 2003 and March 2004 events except that during March 2004 one new site was added on WTGC upstream from Quincy Reservoir to examine Se concentrations near the headwaters (fig. 1; table A1 in Appendix A at the back of the report). Two sites were added between WTGS-2 and WTGS-3 to bracket inflow from a pond located along that reach (fig. 1, table A1). The farthest downstream sample site on TGC (TGS-1, fig. 1) was a few hundred meters upstream from the confluence of TGC and Sand Creek. A site was sampled on Sand Creek (SCS-1), directly upstream from the TGC confluence, to assess the relative contribution of TGC flow and loads to Sand Creek. In TGC, the farthest upstream location in the December sampling was a site on WTGC approximately 0.5 kilometer (km) downstream from Quincy Reservoir (WTGS-3, fig. 1). In March 2004, an additional sample was collected upstream from Quincy Reservoir (WTGS-3.5, fig. 1). There was no flow upstream from ETGS-2 during both sampling events. These 13 sample sites in the TGC watershed encompassed about 80 percent of the perennial length of the channel of about 20 km.

In addition to the stream sites, discharge from a 3-m diameter outfall that drains the area around the Aurora Municipal Complex and Aurora Mall and flows into WTGC near Chambers Road about 0.8 km upstream from the confluence with ETGC (OFS-1, fig. 1) was sampled to evaluate how much Se load it contributed to TGC. A site in WTGC (WTGS-0.5) was included in the sample design to evaluate Se in WTGC directly upstream from OFS-1.

For each water-sampling event, all sampling sites of the stream drainage were sampled in a single day to minimize any day-to-day changes that might occur in the composition of the streamwater. Water sampling began at the most downstream locality then progressed upstream to avoid downstream disturbance of the stream.

#### **Discharge Measurement**

Discharge measurements were made by City of Aurora personnel at each site immediately following water sampling. The stream was divided into at least 3 and, if stream width permitted, as many as 10 measurement cells perpendicular to flow. For each cell, the width, average depth, and cross-sectional area were recorded. In each cell, stream velocity was measured at 0.6 times total depth using a Marsh McBirney Model 201 flowmeter. Discharge for each sampling site was calculated as the product of the cross-sectional area and the average flow velocity for the entire cross section. At some locations, the channel was too shallow to use the flowmeter. For example, at the outfall site (OFS-1) flow typically was about 30-centimeters (cm) wide and only 1- to 3-cm deep. In this case, flow velocity was measured by pouring colored dye into the water, timing its transit along a measured distance, and multiplying flow by the measured cross-sectional area of the channel. There was a problem with the discharge measurement made at ETGS-1 during the December 16, 2003 sampling event. Therefore, discharge measurements were made at all sites on January 15, 2004. Because there were only minor differences in discharge at the other sites between the two dates, we assumed that the January 15, 2004 discharge measurement at ETGS-1 was representative of conditions during the December 16, 2003 sampling event and assigned the January discharge measurement to the site for the December sampling event. This is a reasonable assumption given that December and January are considered periods of baseflow in TGC and that paired December-January discharge measurements at other sites in the watershed showed very little difference. In addition, flow at the nearest USGS streamgaging station, Sand Creek at the Mouth near Commerce City, Colorado, showed similar flow for both days 13 ft<sup>3</sup>/s (368 L/s) on December 16, 2003 and 15 ft<sup>3</sup>/s (425 L/s) on January 1, 2004 http://waterdata.usgs.gov/co/nwis/dv/?site\_ no394839104570300, accessed, July 2006) indicating that flow conditions were similar on the 2 days.

## **Sampling and Processing**

Streamwater was sampled directly into 60 milliliter (mL) plastic syringes. At each site, a new syringe and sample container were triple rinsed with streamwater to be sampled then shaken dry. The syringe was held facing upstream and moved across the width of the stream while sampling water. Using visual estimates of streamflow to guide sampling, sample volumes were obtained that were approximately proportional to flow.

Samples were filtered through new, 0.45 micrometer ( $\mu$ m) pore-size, cellulose nitrate, surfactant-free, filters attached directly to the sampling syringe. The filters were prerinsed with 10 mL of the sample solution. Filtered samples were immediately measured for specific conductance, total dissolved solids (TDS), oxidation-reduction potential (ORP), and pH using a Myron L Ultrameter with automatic temperature compensation. The meter was calibrated onsite several times each day using commercially prepared conductivity and pH standards. For each day's sampling, a field procedural blank was collected using the standard field procedures and 18 meg-ohm (M\Omega) deionized water (DIW).

Filtered samples for trace-element analysis were acidified immediately upon collection in the field using three drops Ultrex nitric acid ( $HNO_3$ ) per 20 mL of sample, which produced a pH value between 1 and 2. The samples were stored at room temperature until analysis within 1 to 2 days after collection.

Two samples of raw (unfiltered) streamwater collected in December 2003 were analyzed using a modified USGS digestion procedure (Garbarino and Hoffman, 1999) to determine trace-element concentrations associated with suspended solids. Whole-water samples for this digestion procedure were collected (without any filtration or acid preservation) and refrigerated in the field and then frozen upon return to the laboratory until digestion. The digestion involved adding 1 mL of concentrated HNO<sub>3</sub> to 30 mL of sample, heating to 60 degrees Celsius for 8 hours, cooling, and then filtering through a 0.45-µm filter. Filtered solutions were diluted tenfold with 1-percent HNO<sub>3</sub> solution before analysis.

### Analytical Methods for Water Samples

Acidified-water samples and whole-water digestions were analyzed for major- and trace-element concentrations using Inductively Coupled Plasma-Mass Spectrometry (ICP-MS) because the technique has a detection sensitivity for Se of about 1 µg/L and also because the technique measures 36 other trace elements (table A1), including iron (Fe) and phosphorous (P) which are sometimes associated with water-quality degradation. Other trace-elements measured include lithium (Li), beryllium (Be), aluminum (Al), silicon (Si), scandium (Sc), titanium (Ti), vanadium (V), chromium (Cr), manganese (Mn), cobalt (Co), nickel (Ni), copper (Cu), zinc (Zn), gallium (Ga), germanium (Ge), arsenic (As), rubidium (Rb), strontium (Sr), yttrium(Y), zirconium (Zr), niobium (Nb), molybdenum (Mo), cadmium (Cd), antimony (Sb), cesium (Cs), barium (Ba), lanthanum (La), cerium (Ce), tungsten (W), thallium (Tl), lead (Pb), bismuth (Bi), thorium (Th), and uranium (U). This technique also includes the major ions sodium (Na), magnesium (Mg), potassium (K), sulfur (S)

(which can be expressed as its common anion for the environmental conditions in TGC as sulfate  $(SO_4)$ ), and calcium (Ca). In addition alkalinity, expected to be present as bicarbonate in these solutions with their near-neutral pH, was measured on a few filtered samples using a sulfuric acid titration to an end-point pH of 4.5 (LaMotte Alkalinity Kit DR-A).

Analysis by ICP-MS was completed in the USGS Geology Discipline laboratories located in Denver, Colorado, using a A Perkin-Elmer Elan 6000 equipped with a conventional pneumatic nebulization sample-introduction system (Lamothe and others, 2002; Taggart 2002). Because concentrations of major ions in most samples were beyond the linear calibration range of the instrument, all samples were diluted tenfold with 1 percent HNO<sub>3</sub> to obtain solutions within the calibrated concentration range. Even with this dilution, a few samples had Na concentrations that exceeded the calibrated concentration range, and these samples were reanalyzed after an additional tenfold dilution. All element concentrations for the water samples presented in table A1 represent the undiluted concentrations of the elements in the environmental samples.

The ICP-MS instrument was calibrated using three commercial multielement aqueous standards, plus two aqueous standards prepared inhouse, and one standard reference sample (T-143) available through the USGS Water Resources Discipline. Quality-assurance/quality-control procedures included assessment of analytical bias and precision and contamination. Standard reference materials and certified standards were analyzed to assess analytical bias (the difference between measured values and certified or "true" values). A standard reference solution from National Institute of Standards and Technology (NIST) (NIST-1643d) was analyzed at the beginning and end of each analytical run to assess bias for the analysis of samples collected in December 2003 and in March 2004. In addition, two standard reference water samples, T-161 and T-147, were analyzed to help assess bias during analysis of the March 2004 environmental samples. To assess precision (the spread in reported data values), three environmental sample replicates were analyzed during the analysis of the December 2003 samples. Replicate analysis of NIST-1634d, T-147, and T-161 provided estimates of precision for the samples collected and analyzed in March 2004. NIST-1634d was analyzed to assess bias and precision of analysis of the whole-water digestions of the two raw-water samples. To assess and prevent contamination, all sample analytical runs included a DIW wash or blank, and the ICP-MS procedure included a 90-second wash using a 2-percent HNO<sub>2</sub> solution between each sample analysis. In addition, one field blank was collected during each sampling day using DIW and the standard field sampling procedure to help assess contamination introduced during sample collection and processing. Approximately 20 percent of the solutions analyzed with December 2003 environmental samples and approximately 30 percent of the solutions analyzed with the March 2004 environmental samples were quality-assurance/ quality-control samples. Approximately 40 percent of the solutions analyzed with the whole-water digestion samples were

quality-assurance/quality-control samples. Analytical data for quality assurance and quality control for water samples collected during December 2003 and March 2004 are listed in table A2, in Appendix A at the back of the report, and the results are described in Appendix B.

#### **Calculation of Load**

Se loads are the product of discharge and selenium concentration. The term instantaneous load is used because the loads were calculated using discharge measurements that represented one point in time. Instantaneous loads were reported as grams per day (g/d) and pounds per year (lbs/yr; table A1) because those terms are easier to conceptualize than micrograms per second. These calculations required converting the units of load (micrograms per second) by appropriate conversion factors to obtain grams per day and pounds per year. In some parts of the following discussion, the term "daily load" is used. This term refers the amount of selenium transported during 1 day as represented by the instantaneous load. For example, a Se load of 250 g/d represents a daily Se load of 250 g.

# Solid Sampling and Analytical Methods

Twenty-one solid samples were collected and analyzed to determine the total amount of Se and other trace and major elements present in the bulk sample. A subset of these samples was treated using a deionized water-leaching procedure to assess the water-soluble concentrations of Se and other trace and major elements. X-ray diffraction analysis (XRD) was used to determine mineralogy of one sample.

### Solid Sampling Strategy and Locations

Twenty-one solid samples comprised of rock (2 samples), a mixture of rock and unconsolidated sediment (13 samples), stream-sediment (3 samples), and efflorescence (surface encrustations produced by evaporation, 3 samples) were collected during January and February 2004. The results of the water analysis were used to develop a strategy to sample rocks and unconsolidated sediment that might contribute Se and other trace elements to the TGC. The December 2003 watersampling results indicated elevated Se in the waters throughout most of TGC and that Se loads increase in a downstream direction. Therefore, rocks and surficial material in proximity to TGC were sampled at several sites over most of the length of TGC (fig. 1). Sampling sites were numbered to correspond to water-sampling sites or were interpolated for locations between water-sampling sites, and A, B, C suffixes indicate

#### 8 Selenium and Other Elements in Water and Adjacent Rock and Sediment of Toll Gate Creek, Colorado

different types of samples collected at the same locations (table 2, samples: TGS/R-; ETGS/R-; and WTGS/R-). No solid samples were obtained from the drainage of Sand Creek. Sample locations and descriptions are provided in table 2. Samples of the cuttings obtained during previous drilling and installation of two monitoring wells are listed as Ur-LUS samples in table 2.

Outcrops of natural rock and unconsolidated sediments are sparse along TGC. Much of the stream channel consists of added fill and slopes that have been graded for flood control (fig. 2). Most bedrock exposures in the drainage are several



Figure 2. View of typical graded slope along bank of Toll Gate Creek.

of TGC has distinct rocks of the Denver Formation exposed in the streambank and bottom (samples TGS/R-1.3 and -2.1A, B).

Some sample locations were selected based on field observations. One sample was collected from anoxic mud in the stream bottom of WTGC (WTGS/R-2.2B) to represent sediments from the wetlands that occur in some locations along WTGC and ETGC. A sample of silty-stream sediment was taken at Station 3.5 of WTGC (WTGS/R-3.5B). Another stream sediment consisting of iron-rich sand was collected about midway down the course of WTGC (WTGS/R-1.4C).

> The portable EDXRF analyzer (see Analytical Methods for Solid Samples section) indicated this iron-rich sample had an approximate composition of 89 percent Fe and 3 percent Mn. This sample was included because Fe and Se occasionally are associated in some sedimentary rock systems (Grauch et al., 2004). Last, three samples were taken of a white efflorescence that occurred on rock surfaces above water level along the stream channel (WTGS/R-1.4B, -2.2A, and -3.5A). This encrustation apparently results from evaporative concentration of dissolved solids in ground water and(or) the capillary fringe of the stream.

> Based on work in a seleniferous shale sequence, (Herring, 2004; Herring and Grauch, 2004), the Se content in exposed, weathered surficial rocks is typically much less than in deeper underlying equivalents of those rocks, because the Se in the exposed rock weathers into highly soluble oxidized species and is removed by ground-water leaching. Consequently, part of the sample design included obtaining some

deeper, unexposed samples of the same rock units as the

hundred meters away from the stream. Because the interest was in sampling only those rocks in direct or most proximal contact with the streamwater, bedrock sections away from the stream were not sampled. Where exposed, the rock and unconsolidated sediments generally were light tan and appeared well weathered. Rock and unconsolidated sediment samples were taken at locations where there was a relatively uniform presence in color and apparent composition through the streambank deposits (fig. 3). These uniform rock sections were sampled along the vertical extent of exposure without regard to the unconsolidated sediment or rock type to obtain a single, composite sample that was representative of the rock sequence at that site. The sample was collected by compositing a continuous set of rock chips along a straight line, perpendicular to bedding planes in the outcrop. This type of composite sample is termed a channel sample. The contact between the unconsolidated sediments and underlying bedrock can be transitional and often was difficult to identify. Only the lowermost reach



**Figure 3.** View of streambank channel sample WTGS/R-1.4A directly north of the Mississippi Street overpass. Note efflorescence zone, sampled as WTGS/R-1.4B.

#### Table 2. Location and description of solid sampling locations, Toll Gate Creek.

[m, meter; km, kilometer; USGS, U.S. Geological Survey; NAWQA, National Water-Quality Assessment; UrLUS, urban land use study; Rk. Btm, bedrock from stream bottom; Ch. Sed., channel sample of streambank sediments; Latitude and longitude referenced to North American Datum 1983; Fe, iron]

USGS lab no.	Sample	Latitude (decimal degrees)	Longitude (decimal degrees)	Depth (m)	Comments
				Toll Gate	Creek
C-230083	TGS/R-1.3	39.74517	-104.82692	Surface	Composite along 30 m of streambank of stream-channel material consisting of in-place bedrock and channel deposits
C-230081	TGS/R-2	39.73008	-104.82043	Surface	Channel sample of 3-m high streambank above water of running stream, 30 m upstream from footbridge and approximately 0.4 km downstream from water-sample site TGS-2
C-230087	TGS/R-2.1 A	39.72463	-104.81670	Surface	Sandstone bedrock bottom in stream
C-230088	TGS/R-2.1 B	39.72463	-104.81670	Surface	Composite channel sample of streambank deposits through 2 m of vertical height on east side of streambank approximately 100 m upstream from Urban Drainage and Flood Control District stream gazing station
				East Toll	Gate Creek
C-230077	ETGS/R-1	39.71751	-104.80638	Surface	Channel sample of 3-m high streambank on east side of stream above running water at location of ETGS-1 water-sampling site
C-230075	ETGS/R-2	39.70082	-104.77437	Surface	Equal mix of: channel sample of bank above water of running stream; and channel sample of bank in dry reach of stream
				West Toll	Gate Creek
C-230086	WTGS/R-1.1	39.71202	-104.80660	Surface	Composite channel sample of streambank deposits through 4 m of vertical height on north side of streambank just after sharp bend in stream to west
C-230090	WTGS/R-1.4A	39.69693	-104.80483	Surface	Streambank sediment channel sampled along 2.5-m vertical section above flowing water; alternations of silt, gravel, and sand layers
C-230084	WTGS/R-1.4B	39.69693	-104.80483	Surface	White efflorescence in streambank deposits directly above flowing water
C-230085	WTGS/R-1.4C	39.69693	-104.80483	Surface	Fe-rich sand in stream channel
C-230089	WTGS/R-1.7	39.68753	-104.80460	Surface	Composite channel sample through 8 m of vertical height along bulldozer scrape on west side of streambank
C-230079	WTGS/R-2.2A	39.67660	-104.79087	Surface	Efflorescence on streambank
C-230082	WTGS/R-2.2B	39.67660	-104.79087	Surface	Anoxic mud in flowing stream bottom
C-230080	WTGS/R-3.5A	39.62608	-104.75158	Surface	Efflorescence on streambank
C-230076	WTGS/R-3.5B	39.62608	-104.75158	Surface	Stream sediment
<u>C-230078</u>	WTGS/R-3.5C	39.62608	-104.75158	Surface	Mud/rock bank along stream
C 220070	L. UC10 1	20 (4(2)	104 79229	USGS NA	WQA well auger cuttings
C-230070	ULUSIS-1	39.04031	-104.78228	8 60-8 00	Contains gypsum crystals
C-230073	Url IIS19-1	39 64006	-104 76611	3 35-3 66	Contains By Pount Crystans
C-230072	UrLUS19-2	39 64006	-104.76611	6.40-6.71	
C-230074	UrLUS19-3	39.64006	-104.76611	12.80-13.1	1

streambank deposits. Five rock and unconsolidated sediment samples were obtained from the base of two shallow USGS National Water-Quality Assessment Program (NAWQA) monitoring wells that were installed (using a hollow stem auger) in the watershed of TGC (see: http://co.water.usgs. gov/nawqa/splt/index.html). The borehole samples ranged in depth from 3 to 13 m below the ground surface.

### Analytical Methods for Solid Samples

A mix of analytical techniques was used for solid samples. All solid samples collected were analyzed using EDXRF. Selected solid samples were treated using a deionized water or nitric-acid leaching procedure and then analyzed using ICP-MS. X-ray diffraction was used to identify minerals in selected samples.

Solids were analyzed using nondestructive, direct measurement of bulk chemical content using EDXRF. The technique, and its accuracy and precision, were described by Siems (2002). The laboratory technique provides detection sensitivity for Se at about 1 part per million (ppm) and also measures 29 other elements. The lower limit of detection (LLD) for each element is listed in table 3. Elements that were detected but for which all samples had concentrations either below the LLD or only a few values with a maximum concentration equal to twice the LLD were, with their LLD in parts per million in parentheses: Ge (2), Ag (1), Cd (1), Sb (2), W (5), and Bi (5). These elements were eliminated from the data set and subsequent evaluation.

Initially, a field-portable EDXRF unit was used to help select rock and unconsolidated sediment sampling for more detailed laboratory analysis. Unfortunately, its lower detection limit of about 30 ppm Se was too high to detect Se in most of the rocks along the TGC drainage. However, the unit was used to help locate samples for other analytes of interest.

A deionized water-leaching procedure was used for the three efflorescence samples and the anoxic mud (samples WTGS/R-1.4B, WTGS/R-2.2A, WTGS/R-3.5A, and WTGS/ R-2.2B) and is similar to the method of Herring (2004). Briefly, solid samples ground to <100 mesh (< 0.15 millimeter (mm)) were reacted with DIW in a mass ratio of 20:1 water to solids without agitation for 30 minutes, centrifuged, and the decanted solution then was filtered at 0.45 µm, acidified, and injected directly into the ICP-MS. Two samples of stream sediment (WTGS/R-2.2B, anoxic mud; TGS/R-2.1A bedrock from the stream bottom) were digested using nitric acid and heat to compare the amount of Se that could be liberated using this procedure with the amount detected by EDXRF in the bulk-sediment sample. This digestion technique is similar to that used for the raw water, described above (see Analytical Methods for Water Samples section), except that 1 gram (g) of ground sediment was added to 30 mL of deionized water, acidified, heated, and cooled. The resulting solution was filtered, diluted tenfold with 1-percent HNO<sub>3</sub>, and injected directly into the ICP-MS.

XRD patterns were collected (Cu K-alpha radiation, Peltier counter) on a Scintag X-1 theta-theta diffractometer with 2-mm divergence slit and 4-mm scatter slit for the tube and 0.5-mm scatter slit and 0.2-mm receiving slit for the detector. Patterns were scanned from 4 to 70 degrees 2-theta at a power setting of 45 kilovolts (kV) and 35 milliamperes (mA) with a step size of 0.02 degree 2-theta and a counting time of 2 seconds per step with a sample spinner to reduce preferred orientation (Steve Sutley, U.S. Geological Survey, written commun., July 2006).

# Selenium and Other Trace Elements in Water

Results from discharge measurements and chemical analysis of samples were used to assess Se occurrence in the water of the TGC watershed. Load profiles of Se within the TGC watershed were used to evaluate the location and distribution of Se input to TGC.

#### Discharge

Discharge ranged from 2.5 L/s at ETGS-2 in December 2003 to 138 L/s at TGS-1 in March 2004 (fig. 4). Discharge increased in a downstream direction in WTGC, ETGC, and downstream from the confluence of ETGC and WTGC to TGS-1 during both sampling events. Discharge decreased in a downstream direction along ETGC during both sampling events. Upstream from ETGS-2, the stream channel was dry. These data indicate that TGC is a gaining stream along



**Figure 4.** Discharge profile along Toll Gate Creek for December 2003 and March 2004 sampling events. Results from Sand Creek and outfall sampling locations not included on this graph.

# Table 3. Energy Dispersive X-Ray Fluorescence (EDXRF) analysis of rock and unconsolidated sediment samples adjacent to Toll Gate Creek.

[ppm, parts per million; USGS, U.S. Geological Survey; NAWQA, National Water-Quality Assessment; Se, selenium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Ba, barium; Ni, nickel; Cu, copper; Zn, zinc; V, vanadium; Cr, chromium; Ga, gallium; As, arsenic; Br, bromine; Mo, molybdenum; Sn, tin; Cs, cesium; La, lanthanum; Ce, cerium; Nd, neodymium; Pb, lead; Th, thorium; U, uranium; TGS/R, Toll Gate Creek solid sample; ETGS/R, East Toll Gate Creek solid sample; WTGS/R, West Toll Gate Creek solid sample; UrLUS, urban land use study; <, less than; A, B, or C suffix on sample indicates different sample types collected at same location, see table 2]

USGS lab number	Sample	Se (ppm)	Sr (ppm)	Y (ppm)	Zr (ppm)	Nb (ppm)	Ba (ppm)	Ni (ppm)
Lower Limit of	Detection	1	2	2	2	2	5	2
Toll Gate Creel	ζ.							
C-230083	TGS/R-1.3	2	888	30	227	15	1,080	12
C-230081	TGS/R-2	<1	513	37	463	22	1,020	19
C-230087	TGS/R-2.1A	3	974	35	235	15	671	10
C-230088	TGS/R-2.1B	<1	631	34	347	16	867	12
East Toll Gate	e Creek							
C-230077	ETGS/R-1	<1	393	43	526	20	1,010	18
C-230075	ETGS/R-2	1	382	42	400	18	961	22
West Toll Gate	Creek							
C-230086	WTGS/R-1.1	<1	404	55	670	24	1,030	24
C-230090	WTGS/R-1.4A	<1	500	63	951	34	1,100	23
C-230084	WTGS/R-1.4B	25	582	42	360	14	654	12
C-230085	WTGS/R-1.4C	<1	202	168	1,470	57	324	25
C-230089	WTGS/R-1.7	<1	495	64	942	32	1,110	26
C-230079	WTGS/R-2.2A	3	830	30	200	11	696	13
C-230082	WTGS/R-2.2B	5	1,220	17	208	13	555	17
C-230080	WTGS/R-3.5A	4	464	12	104	8	391	9
C-230076	WTGS/R-3.5B	4	557	34	349	19	938	20
C-230078	WTGS/R-3.5C	2	786	31	289	20	1,290	24
USGS NAWQA	well auger cuttings							
C-230070	UrLUS18-1	<1	265	17	143	9	1,070	8
C-230073	UrLUS18-2	<1	469	33	332	18	1,130	18
C-230072	UrLUS19-1	<1	316	30	275	23	504	25
C-230071	UrLUS19-2	<1	327	13	277	23	463	11
C-230074	UrLUS19-3	<1	402	35	262	24	654	18

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# Table 3. Energy Dispersive X-Ray Fluorescence (EDXRF) analysis of rock and unconsolidated sediment samples adjacent to Toll Gate Creek.—Continued

[ppm, parts per million; USGS, U.S. Geological Survey; NAWQA, National Water-Quality Assessment; Se, selenium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Ba, barium; Ni, nickel; Cu, copper; Zn, zinc; V, vanadium; Cr, chromium; Ga, gallium; As, arsenic; Br, bromine; Mo, molybdenum; Sn, tin; Cs, cesium; La, lanthanum; Ce, cerium; Nd, neodymium; Pb, lead; Th, thorium; U, uranium; TGS/R, Toll Gate Creek solid sample; ETGS/R, East Toll Gate Creek solid sample; WTGS/R, West Toll Gate Creek solid sample; UrLUS, urban land use study; <, less than; A, B, or C suffix on sample indicates different sample types collected at same location, see table 2]

USGS lab number	Sample	Cu (ppm)	Zn (ppm)	V (ppm)	Cr (ppm)	Ga (ppm)	As (ppm)	Br (ppm)	Mo (ppm)	Sn (ppm)
Lower Limit of	Detection	2	2	5	5	2	2	1	2	2
Toll Gate Creek	ζ.									
C-230083	TGS/R-1.3	37	97	108	28	20	10	4	10	2
C-230081	TGS/R-2	24	97	123	35	16	3	8	<2	2
C-230087	TGS/R-2.1A	23	124	96	16	23	2	6	2	3
C-230088	TGS/R-2.1B	27	95	110	35	17	3	13	<2	<2
East Toll Gate	e Creek									
C-230077	ETGS/R-1	32	87	117	42	17	6	5	<2	3
C-230075	ETGS/R-2	27	88	90	41	17	4	12	3	3
West Toll Gate	Creek									
C-230086	WTGS/R-1.1	28	112	229	46	18	5	2	<2	4
C-230090	WTGS/R-1.4A	36	141	168	50	19	5	6	<2	4
C-230084	WTGS/R-1.4B	19	79	90	28	10	3	91	<2	<2
C-230085	WTGS/R-1.4C	79	469	669	545	35	8	2	<2	8
C-230089	WTGS/R-1.7	32	135	167	45	19	4	5	<2	3
C-230079	WTGS/R-2.2A	23	95	58	25	11	3	38	<2	3
C-230082	WTGS/R-2.2B	84	104	88	32	12	2	14	2	<2
C-230080	WTGS/R-3.5A	15	46	32	18	8	2	80	<2	2
C-230076	WTGS/R-3.5B	30	118	103	50	19	7	17	2	4
C-230078	WTGS/R-3.5C	29	93	103	39	17	4	19	2	3
USGS NAWQA	well auger cuttings									
C-230070	UrLUS18-1	11	35	47	26	12	<2	4	<2	<2
C-230073	UrLUS18-2	24	75	104	35	18	4	3	<2	2
C-230072	UrLUS19-1	53	118	117	43	26	9	10	4	3
C-230071	UrLUS19-2	22	87	118	26	27	<2	1	<2	<2
C-230074	UrLUS19-3	54	117	148	36	26	<2	1	2	<2

# Table 3. Energy Dispersive X-Ray Fluorescence (EDXRF) analysis of rock and unconsolidated sediment samples adjacent to Toll Gate Creek.—Continued

[ppm, parts per million; USGS, U.S. Geological Survey; NAWQA, National Water-Quality Assessment; Se, selenium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Ba, barium; Ni, nickel; Cu, copper; Zn, zinc; V, vanadium; Cr, chromium; Ga, gallium; As, arsenic; Br, bromine; Mo, molybdenum; Sn, tin; Cs, cesium; La, lanthanum; Ce, cerium; Nd, neodymium; Pb, lead; Th, thorium; U, uranium; TGS/R, Toll Gate Creek solid sample; ETGS/R, East Toll Gate Creek solid sample; WTGS/R, West Toll Gate Creek solid sample; UrLUS, urban land use study; <, less than; A, B, or C suffix on sample indicates different sample types collected at same location, see table 2]

USGS lab number	Sample	Cs (ppm)	La (ppm)	Ce (ppm)	Nd (ppm)	Pb (ppm)	Th (ppm)	U (ppm)
Lower Limit of	Detection	5	5	5	10	3	4	4
Toll Gate Creek	ζ.							
C-230083	TGS/R-1.3	<5	52	94	44	23	11	<4
C-230081	TGS/R-2	6	52	103	51	26	9	5
C-230087	TGS/R-2.1A	<5	41	88	44	21	4	<4
C-230088	TGS/R-2.1B	<5	49	100	47	26	9	4
East Toll Gate	e Creek							
C-230077	ETGS/R-1	6	82	161	77	30	21	5
C-230075	ETGS/R-2	10	63	118	58	28	18	4
West Toll Gate	Creek							
C-230086	WTGS/R-1.1	<5	93	189	77	24	21	7
C-230090	WTGS/R-1.4A	6	77	160	82	33	22	7
C-230084	WTGS/R-1.4B	<5	49	101	44	20	12	8
C-230085	WTGS/R-1.4C	<5	219	473	215	83	107	16
C-230089	WTGS/R-1.7	<5	73	152	70	35	15	6
C-230079	WTGS/R-2.2A	6	35	67	31	20	8	5
C-230082	WTGS/R-2.2B	<5	25	53	21	18	8	7
C-230080	WTGS/R-3.5A	<5	23	40	17	11	5	21
C-230076	WTGS/R-3.5B	8	49	94	46	27	14	5
C-230078	WTGS/R-3.5C	11	56	101	46	23	9	5
USGS NAWQA	well auger cuttings							
C-230070	UrLUS18-1	11	33	51	36	24	<4	<4
C-230073	UrLUS18-2	11	58	111	54	23	10	<4
C-230072	UrLUS19-1	7	56	110	55	26	15	4
C-230071	UrLUS19-2	<5	20	33	16	20	11	<4
C-230074	UrLUS19-3	6	54	101	53	22	10	<4

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the reaches sampled for this study. The increase in streamflow along most of the stream length likely results from ground-water discharge to the stream. The greatest increases in discharge occurred between TGS-2 and TGS-1 during the December sampling event, and between WTGS-2 and WTGS-1 during the March sampling event. The discharge at TGS-1 was greater than the flow in Sand Creek upstream from the confluence with TGC by at least a factor of 4 during both sampling events (table A1). Discharge was greater during March 2004 than in December 2003. The greater flow during March 2004 may be due to snowmelt or spring rain or increased ground-water discharge to the stream resulting from infiltration of recent snowmelt and spring rain. Streamflow in December represents base-flow conditions. The slightly higher flow in March was likely caused by infiltration and release of snowmelt or spring rain. At the USGS streamgage at Sand Creek at the mouth, daily mean discharge was 11 and 12 ft<sup>3</sup>/s (311 and 340 L/s) December 16 and 17, 2003 and approximately 20 ft<sup>3</sup>/s (566 L/s) on March 2, 2004, although during March this flow may have been elevated from discharge from a ditch or Aurora Sand Creek Waste Water Reclamation plant that discharge to Sand Creek between the confluence with TGC and the mouth of Sand Creek

(http://waterdata.usgs.gov/co/nwis/dv/?site\_

*no394839104570300*, accessed, July 2006). Thus, the December 2003 and March 2004 sampling events were conducted during low-flow conditions on TGC, and conclusions presented herein are valid only for those flow conditions.

### **Element Concentrations in Water**

Concentrations of Se, other analytes, and Se loads from all water samples are listed in table A1, and Se concentrations are shown for most TGC sites in figure 5. Table A2 lists results of analysis of quality-assurance samples. Details of qualityassurance and quality-control procedures and calculations are reported in Appendix B. The results indicated that precision and bias of all analyses were generally within acceptable limits (about 10 percent for bias and about 5 percent for precision). There was a possible high bias for Fe analyses. There was no contamination in the procedural blank samples except for Zn; blank samples collected during December 2003 had Zn concentrations similar to those measured in environmental samples. It is possible that Zn was present in the DIW used to prepare the blank samples. However, analysis of the Zn data is not critical to conclusions presented herein, so the contamination is inconsequential with regard to this study.

Se concentrations ranged from 7 to 70  $\mu$ g/L and had an average value of about 23  $\mu$ g/L for all TGC watershed streamwater samples (excluding the water samples from the outfall, OFS-1). These values are greater than the State stream standard of 4.6  $\mu$ g/L. The average for the December 2003 stream samples, 28  $\mu$ g/L, exceeded that for the March 2004 stream samples, 14  $\mu$ g/L. In addition, the Se concentrations during December 2003 had a greater range in concentrations, whereas the March 2004 samples had a lower range in concentrations and a flatter concentration profile, particularly downstream from and including WTGS-2 (fig. 5).

The two samples with the greatest Se concentration, again excluding the outfall, came from the most upstream



**Figure 5.** Selenium concentration profile along Toll Gate Creek for December 2003 and March 2004 sampling events. Results from Sand Creek and outfall sampling locations not included on this graph.

location on ETGC. In addition, in March, WTGS-3.5 had elevated Se concentrations. These data from the upstreammost sites in the basin combined with elevated Se concentrations all the way downstream to the mouth of TGC indicate that elevated Se concentrations are widespread throughout the TGC watershed and that elevated Se concentrations exist well upstream (WTGS-3.5 and ETGS-2) from any industrial sources.

The greatest increase in Se concentration occurred between WTGS-3 and WTGS-2 during the December sampling (fig. 5). This large increase may indicate a source between the two locations. However, more detailed sampling between WTGS-3 and WTGS-2 that occurred in March 2004, included sampling upstream and downstream from a pond (WTGS-2.4 and WTGS-2.3, table A1, fig. 5). These samples did not identify the pond or the tributary that flows into it from the east as the source of elevated Se concentration, but they indicate a possible source located between WTGS-2.3 and WTGS-2 (fig. 5).

Arsenic (As) in TGC often was below the detection limit. However, As concentrations ranged from 2 to 6  $\mu$ g/L in stream samples taken in TGC on December 17, 2003.

During December 2003 the outfall (OFS-1) had the greatest Se concentrations measured during the study (133 and 131  $\mu$ g/L). In addition, the outfall (OFS-1) had concentrations of several elements that were greater than those in any

of the streamwater samples (table A1). Se and sulfate had their highest measured concentrations of all water samples in the December 16, 2003 outfall sample. For Sb, the December sampling shows an average of 2.2  $\mu$ g/L and the March sample was about 6.9  $\mu$ g/L at the outfall. The lesser Sb concentrations downstream from the outfall on TGC appear to be dilutions of this larger concentration. Most streamwater samples upstream from the outfall were near to or below the detection limit of 0.3  $\mu$ g/L for Sb. For As, the December water samples at the outfall averaged about 5  $\mu$ g/L.

For the TGC water samples, anion composition can be approximated using the alkalinity and ICP-MS sulfate data. Alkalinity ranged from 230 to 340 mg/L as CaCO<sub>2</sub> (table A1). The presence of substantive amounts of other anions can be tested by estimating a charge balance for the samples by summing the concentration (in equivalents per liter) of the measured anions (sulfate and alkalinity) and comparing this sum to the sum of the concentration (in equivalents per liter) of the major cations (Ca, Mg, K, and Na) that were determined using the ICP-MS. Estimating an average alkalinity of 300 mg/L CaCO<sub>2</sub> for all samples, the calculated bicarbonate concentration, using a conversion factor of 1.22, is 366 mg/L. A calculated charge balance for all water samples (using Na, K, Mg, and Ca for the major cations and bicarbonate and sulfate as the only major anions) produces an average positive excess charge of 11 milliequivalents per liter. This calculation indicates that there is at least one other anion in the streamwater, possibly chloride. Chloride occurs in soluble halite in efflorescence deposits that occasionally occur on the streambank (see "Efflorescence" section) and, consequently, also would be in the streamwater. If the only other anion were chloride, it would require an average concentration of approximately 390 mg/L to equal the excess positive charge from the major cations. Note that this anion evaluation also indicates that sulfate is the dominant anion in these water samples, with an average concentration of 829 mg/L, and that sulfate and bicarbonate together account for approximately two-thirds of the anions on an equivalence basis.

The solutions from the samples of raw water that were digested using heat and acidification were compared with filtered samples of the same water to determine if suspended solids contributed to constituent concentrations and thus loads in TGC under base-flow conditions sampled during December 2003 (table A1). For Se and U, there was less than a 6-percent difference between the two digested raw-water samples and their filtered counterparts, which indicates that suspended solids in the raw-water sample do not contribute to the loading of these trace elements in the streamwater during base-flow conditions. For Cu, Zn, and Mo, the digested raw concentrations were 1.2 to 3 times that of the filtered-water sample, which indicates there was a contribution to the concentrations of these elements in the streamwater by the suspended solids. Concentrations of P in the digested raw samples were elevated about fivefold and those for Fe and Al as much as fiftyfold over the filtered samples, which indicates that suspended

solids in the stream contain major acid-soluble amounts of these elements during base-flow conditions.

### Selenium Load

Selenium-load results are listed in table A1 and shown in figure 6. Selenium loads ranged from less than 6 g/d at ETGS-1 to 250 g/d at TGS-1. Selenium load generally increased downstream along WTGS and TGC, and load decreased downstream along ETGC. Although the greatest Se concentration occurred at ETGS-1, the load at that site was small because the discharge was small. In addition, despite the elevated Se concentrations in the outfall (OFS-1) the Se load was only about 18 percent of that in WTGS-0.5 because of the small discharge of the outfall (table A1). Loads were greater in WTGC than in ETGC. The greatest increases in Se load occurred between WTGS-3 and WTGS-2 during both sampling events, and between TGS-2 and TGS-1 during the December 2003 sampling event, and to a lesser extent during the March 2004 sampling event. The load at TGS-2, downstream from the confluence of WTGC and ETGC was slightly greater than the sum of the loads at ETGC-1 and WTGC-0.5.

For the most part, the load profile shows that load consistently increased along WTGC and TGC during both sampling periods, supporting the hypothesis that ground-water discharge accounts for the elevated Se concentrations and increasing Se loads in TGC. Larger load increases between WTGS-3 and WTGS-2 during both sampling events and between TGS-2 and TGS-1 during the December 2003 sampling event may indicate possible point sources in these locations, or alternately, may indicate that ground water discharging to the stream in



**Figure 6.** Selenium load profile along Toll Gate Creek for December 2003 and March 2004 sampling events. Results from Sand Creek and outfall sampling locations not included on this graph.

these areas had greater Se concentrations than along the rest of the stream.

The progressive downstream increase in Se loads during both sampling events, particularly downstream from WTGS-2, indicates progressive addition of an element provided by and proportional to flow into the stream (fig. 7). The plot of load versus discharge yields a typical linear addition curve for an element with load proportional to inflow. Note that the December 2003 and March 2004 sampling events have different slopes, and the March 2004 measurements indicated slightly higher discharge and lower Se loads. The high values of  $r^2$ , low p-values, and the lack of scatter about the regression lines, particularly at flow greater than 50 L/s (all sites downstream



**Figure 7.** Selenium load and discharge and linear regressions for December 2003 and March 2004 sampling events. Results from Sand Creek and outfall sampling locations not included on this graph.

from and including WTGS-2), indicates that the amount of load being added per unit of flow increase is constant. If this amount were not constant, there would be greater scatter about the regression line and a lower value of r<sup>2</sup> and a greater pvalue. During March 2004 sampling, snow on the ground had recently melted and likely added to shallow ground water, possibly producing the slightly higher flows and lower Se loads compared to the December 2003 sampling. Note that these regression relations cannot be used to predict concentrations in the stream at any location or any flow regime because the regressions were developed for a limited set of locations during generally low-flow conditions and do not encompass the complete range of conditions and locations on TGC. The high values of r<sup>2</sup> of the regressions illustrate that the load increases (per unit volume of flow increase) to WTGC and TGC particularly downstream from WTGS-2 were constant, which is

evidence for discharge of ground water along that stream reach that had a relatively constant concentration of Se.

#### **Pearson Correlation Analysis**

Concentrations of elements in the water samples were tested for correlations using standard Pearson correlation methods. Correlation analysis requires numeric data sets. Hence, when some, but no more than one-third, of the reported values were non-numeric (that is censored, by being at or below the LLD), that variable was eliminated from the correlation subset. Duplicate samples were averaged to eliminate weighting bias, and samples taken at the same site but on different days were treated as separate samples. Elements with concentrations less than the LLD had those values replaced with 0.7 of the LLD (Cohen, 1959). Most elements had complete data sets with no replacements, and Al, Co, and Mo had four or fewer, less than 12 percent, replaced values. However, 32 percent of the values for Cd and Pb and 50 percent of the values for As were replaced. For these elements, especially As, the associations should be considered tenuous. Observations in another seleniferous sedimentary rock system noted that a logarithmic transformation of the concentration data better normalized the data set (Herring, 2004). Here, however, because of the limited number of observations, concentration data were not transformed. Correlations greater than 0.7 are in boldface type (table 4). This degree of correlation is significant at p < 0.005.

In the TGC water samples, Se significantly associates with the major phases SO<sub>4</sub>, K, Ca, Ti, and with the trace elements Cu, Sr, and U (table 4). The Se associations with Na and conductivity have strong but slightly less significant correlation coefficients. Selenium correlations with Ca and Na indicate a soluble host mineral phase with these elements as dominant cations. The association with sulfate indicates that this possibly is a dominant anion in those minerals. The apparent significant association in TGC waters between Se and K is anomalous. In water leachates of seleniferous marine shale of the Phosphoria Formation, Se does not strongly correlate with K (Herring, 2004). In the water of TGC, the seeming significant correlation between these two elements is an artifact of two anomalous water samples. These samples, both from the outfall (OFS-1), have the highest concentrations of Se and K for all samples. The reason for the seeming correlation is that the calculation of the correlation coefficient overly weights the values of the outfall samples by the square of their distance from the center of mass of the data points. Because these two outfall samples are most distant from the center of mass, a seemingly significant linear correlation coefficient is generated. If the two outfall samples are removed from the data set, the correlation coefficient between Se and K reduces to 0.46, which is not significant at p < 0.005.

#### Table 4. Correlation coefficients among elements in Toll Gate Creek water samples.

[Sp. Cond., specific conductance; Na, sodium; Mg, magnesium; Al, aluminum; Si, silicon; P, phosphorus; SO<sub>4</sub>, sulfate; K, potassium; Ca, calcium, Ti, titanium; V, vanadium; Cr, chromium; Mn, manganese; Co, cobalt; Cu, copper; Se, selenium; Sr, strontium; Mo, molybdenum; Cd, cadmium; Ba, barium; Pb, lead; U, uranium; Bold indicates correlation coefficient is significant at p less than 0.005]

Element	Sp. Cond.	Na	Mg	AI	Si	Р	S0₄	К	Ca	Ti	v	Cr	Mn	Co	Cu	Se	Sr	Мо	Cd	Ba	Pb
Na	0.98																				
Mø	0.80	0.70																			
Al	0.14	0.18	-0.07																		
Si	0.37	0.31	0.47	-0.61																	
Р	0.08	0.11	0.01	0.43	-0.26																
SO4	0.84	0.82	0.63	0.18	0.45	0.03															
Κ	0.51	0.49	0.47	0.03	0.60	0.28	0.63														
Ca	0.81	0.73	0.84	-0.12	0.73	-0.06	0.86	0.70													
Ti	0.78	0.72	0.79	-0.16	0.73	-0.10	0.87	0.68	0.97												
V	-0.41	-0.37	-0.52	0.46	-0.61	0.46	-0.37	-0.29	-0.61	-0.65											
Cr	0.58	0.53	0.60	-0.53	0.78	-0.30	0.48	0.43	0.68	0.67	-0.50										
Mn	0.64	0.57	0.77	0.16	-0.08	-0.02	0.41	-0.01	0.44	0.37	-0.24	0.22									
Co	0.68	0.64	0.71	0.20	-0.16	0.11	0.46	0.00	0.41	0.36	-0.17	0.21	0.96								
Cu	0.76	0.82	0.51	0.19	0.35	0.25	0.68	0.59	0.61	0.65	-0.25	0.39	0.26	0.31							
Se	0.66	0.66	0.43	0.09	0.57	0.02	0.89	0.74	0.78	0.83	-0.31	0.45	0.06	0.10	0.74						
Sr	0.90	0.84	0.86	-0.10	0.63	-0.07	0.88	0.62	0.96	0.94	-0.62	0.69	0.57	0.56	0.65	0.73					
Mo	0.07	0.13	-0.10	0.33	-0.33	-0.01	0.12	-0.41	-0.10	-0.04	0.17	-0.21	0.10	0.14	0.20	0.04	-0.05				
Cd	0.73	0.82	0.46	0.25	0.12	0.24	0.49	0.37	0.43	0.43	-0.31	0.28	0.40	0.44	0.80	0.38	0.56	0.20			
Ва	0.38	0.37	0.44	-0.57	0.56	-0.42	0.18	0.06	0.45	0.44	-0.76	0.60	0.27	0.20	0.15	0.04	0.50	-0.06	0.37		
Pb	0.23	0.35	0.00	0.45	-0.27	0.35	0.05	0.12	-0.11	-0.09	0.08	-0.22	0.17	0.19	0.49	0.04	0.03	0.24	0.67	-0.08	
U	0.68	0.66	0.53	0.08	0.54	0.04	0.82	0.79	0.77	0.75	-0.30	0.53	0.24	0.25	0.68	0.80	0.76	-0.10	0.47	0.11	0.17

# Selenium and Other Trace Elements in Solid Samples

Results from sampling and analysis of solid samples were used to help determine a mineralogic or bedrock source for the Se that occurs throughout TGC.

#### Selenium in Solid Samples

Element LLD and element concentration data for the sampled rocks and unconsolidated sediment are listed in table 3. Selenium in the solid samples generally had concentrations at or only moderately elevated above the EDXRF LLD of 1 ppm (table 3). Only 9 of 21 solid samples had a detectable Se concentration that exceeded the LLD. Most channel samples through the streambank were at or below the LLD. The Se concentrations of two sandstone-bedrock samples from the bottom of the stream channel of TGC (2 and 3 ppm) were greater than the mean of about 0.1 ppm that is reported for various sandstone samples in the conterminous U.S. (Connor and Shacklette, 1975). Samples of stream sediment and of anoxic mud from within the stream channel contained 4 and 5 ppm Se, respectively. The largest concentrations of Se in the solid samples were in a white efflorescent material that occurred along some of the streambank above water level. Three samples of the efflorescence had bulk Se concentrations that ranged from 3 to 25 ppm.

In the solid samples, Se was less than the detection limit for 12 of 21 samples. This poor detection ratio precludes a Pearson correlation analysis.

Nearby sedimentary rock and unconsolidated sediment samples that are listed in the USGS National Geochemical Database (NGD) (http://minerals.cr.usgs.gov/projects/ geochem database/index.html, accessed, July 2006) show some slightly elevated Se concentrations relative to average samples for the conterminous United States. The average Se concentration in local soil is 0.3 ppm, although a few samples had concentrations of 2 ppm as compared to surface horizon cultivated and uncultivated soil samples taken throughout Colorado that averaged 0.23 ppm Se (Connor and Shacklette, 1975). Also, a sample of a continental low-grade coal seam taken from near Quincy Reservoir contained 2.4 ppm Se, and a mudstone sample from about 50 m depth in a drill hole about 10 km to the northeast of the TGC drainage basin contained 9 ppm Se. These concentrations are similar to the concentrations reported for samples collected during this study.

#### **Concentration of Elements in Leachate**

Element concentration data from 20:1 (mass water: mass solid) leachates are listed in table 5. Leachate salinity (conductivity) ranged from 1 to 26 millisiemens per centimeter at 25 degrees Celsius (mS/cm) and was several times that of the streamwater samples. Specifically, leachates from the

efflorescence samples had specific-conductance values greater than 15 mS/cm, roughly one-third that of seawater, with Na concentrations near to or exceeding 5,000 mg/L and SO<sub>4</sub> concentrations ranging from 7,870 to 17,200 mg/L. Note that the concentrations of some major ions (Na, Mg, and SO<sub>4</sub>) in leachate solutions from WTGS/R2.2A and WTGS/R3.5A greatly exceeded typical water concentrations of these elements in TGC. In addition, because the concentrations of these major elements were much greater in the leachate solutions than in TGC water samples, and other major elements (Ca and K) had similar concentrations in the leachate solutions and TGC water samples, the efflorescence probably is not evaporated from the capillary fringe of the stream, but rather from water having a different composition than the streamwater-ground water entering the stream. If the efflorescence formed by evaporation of streamwater, element ratios, or relative concentrations of major elements in the streamwater and the efflorescence would be similar. They are not.

In water leachates of the efflorescence, several elements  $(SO_4, Se, Na, V, As, Cu, Mo, and Pb)$  and conductivity were enriched at least sixfold over streamwater samples. Selenium in leachate from efflorescence samples ranged from 146 to 1,460 µg/L. Comparing the amount of Se in the leachate with that known to be in the solid sample (table 3), all Se in the efflorescence was water extractable within the 30-minute extraction time used for these experiments. However, even these elevated Se concentrations were a tiny fraction of the SO<sub>4</sub> concentration in the leachate solutions and in the streamwater (table A1). Molar SO<sub>4</sub>/Se ratios range from about 3,000 to 64,000 in the leachates and from about 10,000 to 69,000 in the streamwater.

#### Efflorescence

X-ray diffraction analysis of one efflorescence sample indicated that it was composed of thenardite (NaSO<sub>4</sub>), gyp-sum, and minor halite, in addition to minor silicate minerals from the rock substrate underlying the efflorescence that were included during sampling.

The solubility of the Se in this material indicates that the efflorescence would release considerable Se into streamwater during snowmelt or rain along the streambanks or during higher water levels associated with seasonal changes in flow.

An approximate evaluation of the possible contribution of the efflorescence material to the Se load in TGC basin indicates that the efflorescence, though an important contributor of Se to the streamwater, is insufficient to supply a Se load of 250 g/d if that load is maintained throughout the year. A daily Se load of 250 g in TGC requires the dissolution of 10 tons of the efflorescence material with a concentration of 25 ppm Se. If an average mass per unit area of this material as exposed on the streambank face is assumed to be about 1 g per 100 cm<sup>2</sup>, dissolution of an amount of efflorescence to produce the daily load of 250 g would require an exposure of about  $10^5$  m<sup>2</sup> of efflorescence along the streambanks. Continuing the

Sample	Procedure	Specific conductance (mS/cm)	TDS (ppt)	рН	ORP (mV)	Alkalinity (mg/L CaCO <sub>3</sub> )	EDXRF Se (ppm)	ICP-MS Se Leach (µg/L)	Amount Se extracted (percent)
Lower Limit of Detection								1	_
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	-	_	_	_	_	25	1,460	117
WTGS/R-2.2B anoxic mud	HNO <sub>3</sub> Digest	_	_	_	_	_	5	111	44
TGS/R-2.1A rock bottom in ck.	HNO <sub>3</sub> Digest	_	_	_	-	-	3	36	24
WTGS/R-2.2A efflorescence	20:1 H <sub>2</sub> O Leachate	17.94	10.61	6.4	_	_	3	146	97
WTGS/R-3.5A efflorescence	20:1 H <sub>2</sub> O Leachate	25.55	15.48	6.85	_	_	4	241	121
WTGS/R-1.4B efflorescence	20:1 $H_2O$ Leachate	18.56	11.02	7.19	-	-	25	1,360	109
WTGS/R-2.2A efflorescence	20:1 H <sub>2</sub> O Leachate	16.7	9.793	6.71	111	28	3	164	109
WTGS/R-3.5A efflorescence	20:1 H <sub>2</sub> O Leachate	26.57	16.18	7.68	85	45	4	255	128
WTGS/R-2.2B anoxic mud	20:1 H <sub>2</sub> O Leachate	1.182	0.6229	7.37	65	32	5	68	27
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	17.26	10.2	8.15	54	30	25	1,120	90

Sample	Procedure	ICP SO <sub>4</sub> (mg/L)	Min. pct. SO₄ in solid sample	Molar S0₄ se	Pct. solid soluble	Li (µg/L)	Be (µւg/L)	Na (mg/L)	Mg (mg/L)	<b>ΑΙ</b> (μ <b>g/L</b> )
Lower Limit of Detection		2	_	-	_	0.9	0.05	0.5	0.01	2
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	9,970	20	5,617	_	28.3	0.07	5,020	69.7	26.4
WTGS/R-2.2B anoxic mud	HNO <sub>3</sub> Digest	893	2	6,617	_	114	17.7	53.6	120	99,000
TGS/R-2.1A rock bottom in ck.	HNO <sub>3</sub> Digest	124	0.2	2,825	-	134	40.8	63.0	108	153,000
WTGS/R-2.2A efflorescence	20:1 H <sub>2</sub> O Leachate	11,300	23	63,659	21	34.7	0.09	_	174	203
WTGS/R-3.5A efflorescence	20:1 H <sub>2</sub> O Leachate	17,200	34	58,701	31	22.4	0.08	_	262	15.9
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	11,500	23	6,955	22	32.5	< 0.05	_	82.0	21.4
WTGS/R-2.2A efflorescence	20:1 H <sub>2</sub> O Leachate	8,320	17	41,727	20	33.2	< 0.05	4,380	149	58.0
WTGS/R-3.5A efflorescence	20:1 H <sub>2</sub> O Leachate	14,800	30	47,737	32	< 0.9	0.2	7,420	246	34.9
WTGS/R-2.2B anoxic mud	20:1 H <sub>2</sub> O Leachate	451	1	5,447	1	< 0.9	< 0.05	43.7	18.0	48.1
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	7,870	16	5,780	20	5.1	0.8	4,390	66.3	43.1

Sample	Procedure	Si (mg/L)	P (mg/L)	K (mg/L)	Ca (mg/L)	<b>Տշ</b> (μ <b>g/L</b> )	<b>Τi</b> (μ <b>g/L</b> )	<b>V</b> (μ <b>g/L</b> )	<b>Сr</b> (µ <b>g/L)</b>	<b>Mn</b> (μ <b>g/L)</b>	<b>Fe</b> (μ <b>g/L</b> )
Lower Limit of Detection		0.2	0.01	0.03	0.2	0.6	0.5	0.5	1	0.2	50
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	< 0.2	0.1	6.21	336	< 0.6	197	8.6	< 1	17.3	164
WTGS/R-2.2B anoxic mud	HNO <sub>3</sub> Digest	95.4	42.2	20.8	7,920	51.5	501	555	148	38,600	443,000
TGS/R-2.1A rock bottom in ck.	HNO <sub>3</sub> Digest	124	66.5	22.4	1,700	55.6	781	275	24.8	78,200	195,000
WTGS/R-2.2A efflorescence	20:1 H <sub>2</sub> O Leachate	3.8	1.4	11.1	418	< 0.6	161	13.2	1.2	139	601
WTGS/R-3.5A efflorescence	20:1 H <sub>2</sub> O Leachate	3.8	0.7	9.97	430	1.3	277	16.0	1.5	31.5	331
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	3.2	0.4	6.41	360	1.0	170	8.8	< 1	14.7	230
WTGS/R-2.2A efflorescence	20:1 H <sub>2</sub> O Leachate	< 0.2	< 0.01	9.76	373	< 0.6	< 0.5	44.9	< 1	64.6	398
WTGS/R-3.5A efflorescence	20:1 H <sub>2</sub> O Leachate	< 0.2	< 0.01	9.45	389	< 0.6	60.5	54.5	< 1	33.4	< 50
WTGS/R-2.2B anoxic mud	20:1 H <sub>2</sub> O Leachate	< 0.2	< 0.01	8.01	124	< 0.6	< 0.5	47.5	< 1	57.4	< 50
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	< 0.2	< 0.01	6.02	307	< 0.6	< 0.5	47.7	< 1	14.4	< 50

Sample	Procedure	Со (µg/L)	<b>Ni</b> (μ <b>g/L</b> )	<b>Си</b> (µ <b>g/L</b> )	<b>Ζn</b> (μ <b>g/L</b> )	<b>Ga</b> (μ <b>g/L</b> )	<b>Ge</b> (µ <b>g/L</b> )	<b>As</b> (μ <b>g/L</b> )	<b>Rb</b> (μ <b>g/L</b> )	<b>Sr</b> (μ <b>g/L</b> )	Υ (μ <b>g/L</b> )
Lower Limit of Detection		0.02	0.4	0.5	0.5	0.05	0.05	1	0.01	0.5	0.01
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	1.06	8.0	24.8	10.8	< 0.05	< 0.05	62.2	3.36	11,300	0.70
WTGS/R-2.2B anoxic mud	HNO <sub>3</sub> Digest	203	152	2,580	2,430	36.6	2.2	95.5	59.9	55,600	308
TGS/R-2.1A rock bottom in ck.	HNO <sub>3</sub> Digest	286	159	575	2,190	49.4	3.4	34.4	237	15,300	720
WTGS/R-2.2A efflorescence	20:1 H <sub>2</sub> O Leachate	3.53	6.0	48.5	15.8	0.08	0.75	11.4	4.19	8,960	1.22
WTGS/R-3.5A efflorescence	20:1 H <sub>2</sub> O Leachate	1.67	9.0	43.7	26.2	< 0.05	0.90	18.7	4.38	8,860	0.58
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	0.92	1.7	30.2	19.9	< 0.05	0.25	23.8	2.88	10,200	0.38
WTGS/R-2.2A efflorescence	20:1 H <sub>2</sub> O Leachate	2.44	< 0.4	44.2	189	< 0.05	1.4	20.3	3.56	7,970	0.14
WTGS/R-3.5A efflorescence	20:1 H <sub>2</sub> O Leachate	1.58	0.9	50.6	12.9	< 0.05	1.7	27.4	4.61	8,600	0.30
WTGS/R-2.2B anoxic mud	20:1 H <sub>2</sub> O Leachate	1.35	0.9	50.5	< 0.5	< 0.05	1.0	2	3.49	1,570	0.06
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	0.56	< 0.4	26.5	< 0.5	< 0.05	1.8	7.5	2.41	9,470	0.10

Sample	Procedure	Zr (µg/L)	<b>Nb</b> (μ <b>g/L</b> )	<b>Μο</b> (μ <b>g/L)</b>	Cd (µg/L)	<b>Sb</b> (μ <b>g/L</b> )	<b>Сs</b> (µ <b>g/L</b> )	<b>Βa</b> (μ <b>g/L</b> )	<b>La</b> (μ <b>g/L</b> )	<b>Се</b> (µ <b>g/L</b> )	<b>W</b> (μ <b>g/L</b> )
Lower Limit of Detection		0.2	0.2	2	0.02	0.3	0.02	0.2	0.01	0.01	0.5
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	0.2	1.64	20.8	0.09	0.70	0.03	74.0	0.02	0.05	1.20
WTGS/R-2.2B anoxic mud	HNO <sub>3</sub> Digest	44.2	< 0.2	13.3	11.0	1.19	2.24	5,250	375	892	1.87
TGS/R-2.1A rock bottom in ck.	HNO <sub>3</sub> Digest	19.9	< 0.2	4.3	14.9	0.68	11.1	6,120	993	2200	1.37
WTGS/R-2.2A efflorescence	20:1 H <sub>2</sub> O Leachate	3.0	< 0.2	11.6	0.23	0.88	0.02	43.6	1.10	2.48	1.45
WTGS/R-3.5A efflorescence	20:1 H <sub>2</sub> O Leachate	2.8	< 0.2	73.6	0.20	0.52	0.05	57.7	0.06	0.24	0.93
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	0.97	< 0.2	21.4	0.10	0.42	< 0.02	57.7	0.02	0.04	0.71
WTGS/R-2.2A efflorescence	20:1 H <sub>2</sub> O Leachate	1.8	16.8	39.0	< 0.02	6.82	0.15	33.2	0.08	< 0.01	6.06
WTGS/R-3.5A efflorescence	20:1 H <sub>2</sub> O Leachate	2.4	15.3	80.9	0.10	3.92	0.16	59.3	0.07	0.06	5.54
WTGS/R-2.2B anoxic mud	20:1 H <sub>2</sub> O Leachate	0.94	12.1	17.9	< 0.02	2.54	0.07	47.3	< 0.01	< 0.01	4.06
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	1.2	12.2	25.9	< 0.02	2.00	0.08	61.0	< 0.01	< 0.01	3.47

#### 24 Selenium and Other Elements in Water and Adjacent Rock and Sediment of Toll Gate Creek, Colorado

# **Table 5.** Composition of water leachate and digestions of various rock and unconsolidated rock samples, Toll Gate Creek.—Continued Continued

[TDS, total dissolved solids; ORP, oxidation-reduction potential; EDXRF, energy dispersive X-ray fluorescence; ICP-MS, inductively coupled plasma-mass spectrometry, ICP, inductively coupled plasma; mS/cm, millisiemens per centimeter; ppt, parts per thousand; mV, millivolts; mg/L, milligram per liter; CaCO<sub>3</sub>, calcium carbonate;  $\mu$ g/L, micrograms per liter; ppm, parts per million; Min., minimum; Pct., percent; ck., creek; Se, selenium; Li, lithium; Be, beryllium; Na, sodium; Mg, magnesium; Al, aluminum; Si, silicon; P, phosphorus; SO<sub>4</sub>, sulfate; K, potassium, Ca, calcium; Sc, scandium, Ti, titanium; V, vandium; Cr, chromium; Mn, manganese; Fe, iron; Co, cobalt; Ni, nickel; Cu, copper; Zn, zinc; Ga, gallium; Ge, germanium; As, arsenic; Rb, rubidium; Sr, strontium; Y, yttrium;, Zr, zirconium; Nb, niobium; Mo, molybdenum; Cd, cadmium; Sb, antimony; Cs, cesium; Ba, barium; La, lanthanum; Ce, cerium; W, tungsten; Tl, thallium; Pb, lead; Bi, bismuth; Th, thorium; U, uranium; H<sub>2</sub>O, water; HNO<sub>3</sub>, nitric acid, <, less than; -, no value]

Sample	Procedure	<b>ΤΙ</b> (μ <b>g/L)</b>	<b>Pb</b> (μ <b>g/L)</b>	Bi (μg/L)	<b>Th</b> (μ <b>g/L</b> )	Ս (µg/L)
Lower Limit of Detection		0.1	0.05	0.2	0.2	0.1
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	< 0.1	0.2	< 0.2	0.40	15.6
WTGS/R-2.2B anoxic mud	HNO <sub>3</sub> Digest	0.88	390	1.23	47.6	175
TGS/R-2.1A rock bottom in ck.	HNO <sub>3</sub> Digest	0.4	473	1.16	21.0	53.3
WTGS/R-2.2A efflorescence	20:1 H <sub>2</sub> O Leachate	< 0.1	0.68	< 0.2	0.74	3.18
WTGS/R-3.5A efflorescence	20:1 H <sub>2</sub> O Leachate	< 0.1	0.06	< 0.2	< 0.2	82.4
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	< 0.1	< 0.05	< 0.2	< 0.2	10.3
WTGS/R-2.2A efflorescence	20:1 H <sub>2</sub> O Leachate	3.2	1.2	4.11	1.15	3.31
WTGS/R-3.5A efflorescence	20:1 H <sub>2</sub> O Leachate	1.6	0.4	1.93	1.14	91.2
WTGS/R-2.2B anoxic mud	20:1 H <sub>2</sub> O Leachate	0.91	0.3	1.20	0.47	5.45
WTGS/R-1.4B efflorescence	20:1 H <sub>2</sub> O Leachate	0.77	0.3	0.78	0.56	10.2

approximation, if the typical efflorescence zone along the streambank is one-half meter in height (figs. 3 and 8), a 20-km length of streambank efflorescence would have to dissolve each day to supply the observed daily load. This would constitute a significant fraction of the total length of all reaches of TGC. Based on field observation, the amount of streambank with efflorescence is estimated to be 5 percent or less, therefore it is not possible for dissolution of the efflorescence to supply the observed Se loads to the stream.

The source of the Se in this efflorescence material is unknown, but it likely results from movement of ground water through rock with relatively low but leachable Se content, discharge of this ground water along the streambank, evaporation of some of the water on the face of the streambank, and concentration of the Se into the resulting evaporite minerals. Alternatively, the efflorescence could originate from percolation and evaporation of Se-laden streamwater that saturates the bank during high flow in the stream then seeps back toward the stream during lower flow. However, if this were the case, the streambanks would have relatively uniform coverage of the efflorescence, and this is not the case. In addition, if the efflorescence originated from streamwater, it should have similar element ratios to streamwater, and as described in a previous section ("Concentration of Elements in Leachate"), it does not.



**Figure 8.** View of efflorescence zones along West Toll Gate Creek directly north of Mississippi Street overpass.

Efflorescence on the bank has an irregular occurrence, which is more consistent with the seepage of Se-laden ground water through rock and discharge along permeable zones of the bank.

# Implications Regarding Selenium Sources, Remediation, and Future Work

The results of this study can be used to discuss possible sources of Se to TGC, possible remediation strategies for the elevated Se concentrations, and directions for future work. The discussion presented in this section arose from interpretation of the scientific data presented herein and is not intended to be an official recommendation by the U.S. Geological Survey.

## **Possible Sources of Selenium to Toll Gate Creek**

Dissolution of efflorescence along the streambanks is insufficient to supply a sustained daily Se load of 250 g on the basis of the mass-balance calculations presented herein. However, the efflorescent material may indicate that ground water having small Se concentrations is discharging to the stream. The ground water eventually transports the Se to the water of TGC, likely contributing Se along the entire length of the stream, which is consistent with the observed surfacewater-quality data.

The possible supply of Se from shallow ground water in the drainage basin of TGC to the streamwater can be calcu-

lated. If a daily load of 250 g Se in the stream is maintained throughout the year, the annual load would be about  $9 \times 10^4$  g. The drainage basin of the stream is about 116 km<sup>2</sup> in extent. A rock and unconsolidated sediment layer with an average density of about 2 g/cm and volume of 10 m thickness over an areal extent of 116 km<sup>2</sup> with an average Se content of 2 ppm contains  $4 \times 10^9$  g Se. Only a small fraction of this Se would have to dissolve into the water of TGC to supply an annual load of about 9 x 10<sup>4</sup> g.

Geologic materials that outcrop and subcrop in the TGC watershed consist of Quaternary colluvial and alluvial deposits overlying the Tertiary to Cretaceous-age Denver Formation. The surficial colluvial and alluvial deposits are composed of unconsolidated gravel, sand, and clay with a dominant mineralogy of quartz and feldspar. The Denver Formation in the study area consists of interbedded sandstones, claystones, and lignite beds of fluvial and volcanic origin and is the likely source of Se to ground water and surface water in the TGC watershed. The Pierre Shale has been identified as a source for Se in water in the Arkansas River basin south of the study area (Zielinski and others, 1995). However, in the study area, the top of the Pierre Shale occurs at least 360 m below the bottom of the Denver Formation (Suzanne Paschke, U.S. Geological Survey, oral commun., 2006). It is not a likely source for Se occurring in streamwater in TGC.

# Remediation of Selenium and Subsequent Monitoring of Trace Elements

Concentrations of dissolved, oxidized Se in flowing surface water can markedly decrease as the water flows through wetlands (Stillings and Amacher, 2004; Mackowiak and others, 2004). In wetlands, the oxidized Se in the water is chemically reduced and the Se becomes incorporated into the sediment, either in a mineralized form or in organic matter. In this study, there is a possible example of this wetland removal indicated by a downstream decrease in Se load in ETGC (between ETGC-2 and ETGC-1, fig. 6) after the streamwater passes through a wetland and golf course (fig. 9). Furthermore, the anoxic stream mud of sample WTGS/R-2.2B had a Se concentration of 5 ppm, which is the second highest concentration of the solid samples and greater than two of the efflorescence samples. The value in this mud is consistent with removal of Se into reducing sediment. Note that these considerations also indicate a possible remedial technique for the removal of Se from the streamwater by incorporating flow through natural or constructed wetlands (Lin and Terry, 2003)



Figure 9. View of wetland and golf course located on East Toll Gate Creek.

### **Future Work**

An inventory of the extent of the efflorescent material on the streambanks and its Se content may help understand source areas for the efflorescence and Se where the Se content is elevated. The amount of this material on the streambanks likely changes throughout the year because of dissolution by snowmelt or seasonal storm runoff. Therefore, it would be appropriate to inventory and sample this material several times, for example quarterly, throughout the year.

The origin of the Se in the white efflorescence deposits on the TGC banks has not been identified. Although it seems likely that the efflorescence results from shallow ground-water seepage toward the stream and consequent discharge onto the surface with evaporation and concentration of Se, there is no direct evidence in this study to support this hypothesis. Therefore, assessment of the Se concentration and the amount of soluble Se in shallow ground water within the drainage area of TGC that might contribute Se to the stream via ground-water discharge would help test the hypothesis that ground water is the source for Se. This assessment could be accomplished by installing a series of monitoring wells. Drilling these wells would allow recovery of rock that can be analyzed for Se content and water-extractable Se content that might help identify which rock layers contribute Se to ground water.

This study concentrated on Se concentrations and load in TGC during low flow. Seasonal sampling would assess how Se concentrations and loads vary over the full range of hydrologic conditions expected in TGC. In addition, more detailed spatial sampling of TGC, its tributaries, and all flowing outfalls may help identify zones where Se load is increasing.

# Conclusions

Streamwater and solid samples were collected in the Toll Gate Creek (TGC) watershed to assess the distribution of selenium (Se) in these materials and evaluate the potential for rocks and solids in the watershed to be sources of Se to the stream. Water samples were collected and discharge measurements were made at as many as 13 sites along East Toll Gate Creek (ETGC), West Toll Gate Creek (WTGC), and TGC during December 2003 and March 2004. Samples were analyzed by inductively coupled plasma mass spectrometry (ICP-MS) and profiles of discharge, Se concentration and Se load were constructed. Twenty-one samples of rock, rock and unconsolidated sediment, stream sediment, and efflorescence were collected along the stream channel of TGC during January and February 2004. Samples were analyzed by energy dispersive X-ray fluorescence (EDXRF), a deionized water (DIW) leaching procedure, a weak nitric-acid leaching procedure, and by X-ray diffraction (XRD). Solutions from the leaching procedures were analyzed by ICP-MS.

Discharge ranged from 2.5 liters per second (L/s) at ETGS-2 in December 2003 to 138 L/s at TGS-1 in March

2004. Except in ETGC, discharge increased moving downstream, and the discharge in WTGC was greater than in ETGC. Downstream increases in discharge indicate that Toll Gate Creek and West Toll Gate Creek gain flow from ground water. Discharge was greater in March 2004 than in December 2003, but both periods were low-flow periods for the TGC watershed. Therefore, the results of this study should not be extrapolated to higher flow periods.

Selenium concentrations in stream samples ranged from 7 to 70 micrograms per liter ( $\mu$ g/L), were elevated in the upstream-most samples collected in WTGC and ETGC, and were greater than the State stream standard of 4.6  $\mu$ g/L in all stream samples collected along TGC. Average Se concentrations were greater in December (28  $\mu$ g/L) than in March 2004 (14  $\mu$ g/L). The greatest Se concentration (133  $\mu$ g/L) occurred at the outfall site (OFS-1).

Se loads ranged from 6 grams per day (g/d) at ETGS-1 to 250 g/d at TGS-1. Se loads increased along WTGC and TGC, but decreased along ETGC. Se loads generally were greater in December 2003 than in March 2004. Despite elevated concentrations, the outfall sample contained only18 percent of the load measured at the site directly upstream from where it flowed into TGC (WTGS-0.5). The largest Se-load increases occurred between two sampling locations on West Toll Gate Creek during both sampling periods and between the two sampling locations in Toll Gate Creek during the December 2003 sampling. This result may indicate sources for Se in these locations. However, Se loading along West Toll Gate Creek and Toll Gate Creek generally indicated gradual load increases moving downstream, which is characteristic of a ground-water source for Se. Linear regressions between discharge and Se load indicated significant regressions (p < 0.0001) having large r<sup>2</sup> values because the load increases (per unit volume of flow increase) into WTGC and TGC (particularly downstream from WTGS-2) generally were constant moving downstream. This pattern is evidence for discharge of ground water along that stream reach that had a relatively constant concentration of Se.

Pearson correlation analysis of water-sample results indicated strong correlations (p < 0.005) between Se and calcium (Ca), and sulfate (SO<sub>4</sub>). A strong, but slightly less significant correlation occurred between Se and sodium (Na). These correlations suggest that the source of Se to the water is a mineral containing Na, Ca, and SO<sub>4</sub> as some of its major constituents.

Rock outcroppings along the stream channel are highly weathered and do not contain elevated concentrations of Se. Selenium was below the detection limit of 1 ppm in 12 of 21 solid samples analyzed. One sample of the efflorescent material had the greatest concentration of Se (25 ppm), and the other efflorescent samples had Se concentrations of 3 and 4 ppm.

Results of deionized water leaching of the efflorescent material indicated that all of the Se dissolved during the leaching procedure, and that some elements in the leachate were concentrated more than sixfold over concentrations in streamwater samples. X-ray diffraction analysis indicated that one efflorescent sample was composed of thenardite, gypsum, some halite, and minor silicate minerals likely incorporated into the sample from the underlying outcrop. Calculations indicate that the efflorescent material is not present in sufficient quantity along the banks of TGC to account for the Se load in TGC. However, its presence, and different element ratios in its leachate relative to streamwater indicate that ground-water discharge containing elevated Se concentrations probably produces the efflorescent material.

This report did not identify an unequivocal source for Se in TGC. However, multiple lines of evidence indicate that ground-water discharge supplies Se to TGC: (1) the occurrence of elevated Se concentrations in the stream throughout the watershed and in the headwater regions, upstream from industrial sources; (2) the progressive increase in Se loads moving downstream (particularly the constant increase in load per unit flow downstream from WTGS-2) which indicates a continuous input of Se along the stream reach, rather than input from point sources; (3) the occurrence of efflorescence deposits in several locations along the stream channel that contain elevated concentrations of readily soluble Se, and probably represent ground-water discharge zones; and (4) the occurrence of rock types within the TGC watershed that may contain elevated concentrations of Se (ash layers and lignite deposits). Ground water acquires the Se through water-rock interaction within the aquifers underlying TGC, and shallow ground water containing soluble Se discharges to TGC. This study did not identify which rock formation or mineral was the source of Se to ground water. However, the results of this study indicate that ground water that acquires Se through water-rock interaction within the TGC watershed is the likely source for Se in streamwater in TGC.

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Site	Date	Latitude	Longitude	Location notes
Lower Limit of Detection				
Sand Creek				
SCS-1	12/16/2003	N 39.75658	W 104.80752	Sand Ck.,150 m upstream from confluence with Toll Gate Ck.
	12/17/2003 3/2/2004			
Toll Gate Creek, after confluen	nce of East and West Reaches			
TGS-1	12/16/2003	N 39.75395	W 104.83175	Toll Gate Ck., across from Fitzsimmons
	12/16/2003			
	12/17/2003			
	3/2/2004			
TGS-2	12/16/2003	N 39.72604	W 104.81747	Toll Gate Ck., at 6th Ave.
	12/17/2003			
	3/2/2004			
TGS-2 Raw, digested	12/16/2003			
West Toll Gate Creek				
OFS-1	12/16/2003	N 39.71933	W 104.81210	City outfall at West Toll Gate Ck.
	12/17/2003			
	3/2/2004			
WTGS-0.5	12/16/2003	N 39.70287	W 104.81211	West Toll Gate Ck., 30 m upstream from outfall discharge
	3/2/2004			
WTGS-1	12/16/2003	N 39.71525	W 104.80959	West Toll Gate Ck., 100 m upstream from Chambers Rd.; bedrock shale exposed

Site	Date	Latitude	Longitude	Location notes
Lower Limit of Detection				
West Toll Gate Creek				
WTGS-2	12/17/2003 3/2/2004 12/16/2003	N 39.68107	W 104.79910	West Toll Gate Ck., Horse Shoe Park
	12/17/2003 3/2/2004			
WTGS-2.3	3/2/2004	N 39.67045	W 104.77913	West Toll Gate Ck., directly downstream from pond
WTGS-2.4	3/2/2004	N 39.66844	W 104.77809	West Toll Gate Ck., directly upstream from pond
WTGS-3	12/16/2003	N 39.63887	W 104.78188	West Toll Gate Ck., diversion flow around Quincy Reservoir
	12/17/2003			
	3/2/2004			
WTGS-3.5	3/2/2004	N 39.62608	W 104.75158	West Toll Gate Ck., 1 km upstream from Quincy Reservoir; Himalaya Ct. and Chenango St.
East Toll Gate Creek				
ETGS-1	12/16/2003 12/17/2003 3/2/2004	N 39.71751	W 104.80638	East Toll Gate Ck., 1st and Chambers
ETGS-2	12/16/2003	N 39.70514	W 104.77606	East Toll Gate Ck., 50 m north of Alameda Ave.
	12/17/2003 3/2/2004			
ETGS-2 Raw, digested	12/16/2003			

Site	Date	Comment	Temp., (°C)	рН	Specific conductance (µS/cm)	Total dissolved solids (ppm)	ORP (mV)	Alkalinity (mg/L CaCO <sub>3</sub> )
Lower Limit of Detection								
Sand Creek								
SCS-1	12/16/2003	Average of duplicate pair	0	6.64	1,623	862	163	340
	12/17/2003 3/2/2004		0.3	7.39 6.02	1,709 1,676	910 889	136	-
Toll Gate Creek, after confluen	ce of East and West Reaches							
TGS-1	12/16/2003	Average of duplicate pair	0	7.08	4,045	2,230	175	308
	12/16/2003	Ice on stream	_	_	_	_	_	_
	12/17/2003 3/2/2004		0 4.5	7.57 6.43	3,250 2,484	1,774 1,337	155 -	
TGS-2	12/16/2003 12/17/2003 3/2/2004		3 1 2	7.29 7.45 6.66	3,205 2,935 2,402	1,751 1,593 1,291	171 175 -	- - -
TGS-2 Raw, digested	12/16/2003					,		
West Toll Gate Creek								
OFS-1	12/16/2003 12/17/2003 3/2/2004		4 7 5	7.60 7.82 6.73	4,525 4,254 2,813	2,504 2,351 1,524	188 170	
WTGS-0.5	12/16/2003		3	7.70	2,993	1,626	181	-
	3/2/2004		4.5	6.87	2,372	1,275		_
WTGS-1	12/16/2003		3	7.55	2,925	1,588	207	_

Site	Date	Comment	Temp., (°C)	рН	Specific conductance (µS/cm)	Total dissolved solids (ppm)	ORP (mV)	Alkalinity (mg/L CaCO <sub>3</sub> )
Lower Limit of Detection								
West Toll Gate Creek								
	12/17/2003 3/2/2004		2 5	7.39 6.98	2,995 2,368	1,625 1,273	184	-
WTGS-2	12/16/2003	Average of duplicate pair	3	6.93	3,268	1,457	134	322
	12/17/2003 3/2/2004		2 5.5	7.04 6.3	3,496 2,292	1,312 1,230	171	
WTGS-2.3	3/2/2004		6	6.73	2,209	1,185	-	-
WTGS-2.4	3/2/2004		11.5	7.12	2,041	1,081	-	-
WTGS-3	12/16/2003		2	7.43	2,693	1,785	136	_
	12/17/2003	Average of duplicate pair	2	7.54	2,772	1,501	175	_
	3/2/2004	L	4	7.24	1,579	838	-	-
WTGS-3.5	3/2/2004		6.5	7.13	2,027	1,057	_	-
East Toll Gate Creek								
ETGS-1	12/16/2003 12/17/2003		3 4	7.01 7.38	2,365 2,455	1,271 1,317	112 144	230
	3/2/2004		7	6.96	3,584	1,958	107	-
EIGS-2	12/16/2003		3	6.82	6,130	3,430	125	-
	12/17/2003		2	7.08	5,400	3,005	175	_
ETGS-2 Raw, digested	3/2/2004 12/16/2003		6.5	6.97	4,463	2,466	_	-

Site	Date	Discharge (L/s)	Discharge, (cubic feet per second)	<b>Se</b> (μ <b>g/L</b> )	Instantaneous Se load (g/d)	Instantaneous Se Ioad (pounds per year)	Li (μ <b>g/L)</b>	Be (μg/L)
Lower Limit of Detection				< 1			< 0.9	< 0.05
Sand Creek								
SCS-1	12/16/2003	19	0.68	6.8	18	15	62.7	< 0.06
	12/17/2003	_	_	7.5	_	_	69.4	< 0.05
	3/2/2004	17	0.60	4.1	6.0	4.8	103	0.2
Toll Gate Creek, after confluen	ce of East and West Reaches							
TGS-1	12/16/2003	79	2.8	37	250	201	198	< 0.05
	12/16/2003	_	_	7.8	_	_	35.3	< 0.05
	12/17/2003	_	_	37	_	_	149	< 0.05
	3/2/2004	138	4.9	15	184	148	118	0.06
TGS-2	12/16/2003	55	1.9	32	150	120	144	< 0.05
	12/17/2003	_	_	32	_	_	112	< 0.05
	3/2/2004	122	4.3	13	138	111	111	< 0.05
TGS-2 Raw, digested	12/16/2003			31.6			148	< 0.05
West Toll Gate Creek								
OFS-1	12/16/2003	1.22	0.04	133	14	11	111	< 0.05
	12/17/2003	_	_	131	_	_	98.1	< 0.05
	3/2/2004	6.4	0.2	51	28	23	69.6	< 0.05
WTGS-0.5	12/16/2003	44	1.6	30.3	116	93	116	< 0.05
	3/2/2004	105	3.7	14	128	103	112	< 0.05
WTGS-1	12/16/2003	44	1.6	32	124	99	107	< 0.05

Site	Date	Discharge (L/s)	Discharge, (cubic feet per second)	<b>Se</b> (μ <b>g/L</b> )	Instantaneous Se Ioad (g/d)	Instantaneous Se load (pounds per year)	Li (µg/L)	<b>Ве</b> (µ <b>g/L)</b>
Lower Limit of Detection				< 1			< 0.9	< 0.05
West Toll Gate Creek								
	12/17/2003	-	_	30	-	_	116	< 0.05
	3/2/2004	104	3.7	13	114	92	112	0.06
WTGS-2	12/16/2003	30	1.1	36	91	73	109	< 0.05
	12/17/2003	_	_	38	_	_	132	< 0.05
	3/2/2004	56	2.0	17	81	65	72.6	0.06
WTGS-2.3	3/2/2004	-	_	7.4	-	-	47.9	0.06
WTGS-2.4	3/2/2004	-	_	10.2	-	_	56.0	< 0.05
WTGS-3	12/16/2003	11	0.4	17.9	17	13	75.2	< 0.05
	12/17/2003	_	_	18.6	_	_	79.3	< 0.05
	3/2/2004	18	0.7	7.2	11	9	47.1	0.09
WTGS-3.5	3/2/2004	6.7	0.2	12.3	7	6	57.2	< 0.05
East Tall Cate Create								
East Toll Gate Creek								
ETGS-1	12/16/2003	4.7	0.17	14.3	5.8	4.7	45.1	< 0.05
	12/17/2003	-	-	14.3	_	-	53.3	< 0.05
	3/2/2004	7.2	0.3	13.0	8	7	154	0.06
ETGS-2	12/16/2003	2.5	0.09	48.8	10	8.4	257	< 0.05
	12/17/2003	_	_	69.7	_	_	274	< 0.05
	3/2/2004	4.1	0.1	37.2	13	11	127	0.06
ETGS-2 Raw, digested	12/16/2003			49.4			266	0.05

Site	Date	Na (mg/L)	Mg (mg/L)	Al (ug/L)	Si (mg/L)	P (ma/L)	SO <sub>4</sub> (mg/L)	K (mg/L)	Ca (mg/L)
Lower Limit of Detection		< 0.5	< 0.01	< 2	< 0.2	< 0.01	< 2	< 0.03	< 0.2
		<b>C</b> 0.5	<b>CO.01</b>		0.2	0.01		< 0.05	0.2
Sand Creek									
SCS-1	12/16/2003	156	28.7	< 2	12.6	0.02	302	4.9	180.0
	12/17/2003	146	29.2	< 2	11.8	0.01	290	4.46	184
	3/2/2004	156	27.6	8.0	8.3	0.01	323	4.55	156
Toll Gate Creek, after confluen	ce of East and West Reaches								
TGS-1	12/16/2003	596	65.4	2.5	9.8	0.02	832	5.5	288
	10/17/2002	07.5	10.0	52.6	1.0	0.04	170	1.05	(0.5
	12/16/2003	97.5	12.2	53.6	1.8	0.04	179	1.05	60.5
	12/17/2003	399	66.7	3.4	8.9	0.02	777	5.37	288
	3/2/2004	313	38.2	8.6	4.7	0.07	636	4.92	178
TGS-2	12/16/2003	414	58.7	2.9	9.0	0.03	812	5.31	277
	12/17/2003	340	65.6	3.2	8.5	0.03	750	5.21	291
	3/2/2004	286	36.2	10	4.7	0.09	572	4.84	166
TGS-2 Raw, digested	12/16/2003	439	60	158	6.9	0.2	806	5.57	273
West Toll Gate Creek									
OFS-1	12/16/2003	700	59.4	8.5	11.4	0.04	1850	9.97	440
	12/17/2003	597	72.5	10.1	11.7	0.08	1680	8.78	477
	3/2/2004	370	29.9	14.7	5.5	0.07	1020	3.74	221
WTGS-0.5	12/16/2003	372	58.6	2.9	9.9	0.02	804	4.76	275
	3/2/2004	282	36.0	0.2	16	0.10	550	4.80	161
	31212004	203	50.0	7.5	4.0	0.10	550	4.07	101
WTGS-1	12/16/2003	370	57.5	3.4	9.6	0.03	804	4.63	272

Site	Date	Na (mg/L)	Mg (mg/L)	Al (μg/L)	Si (mg/L)	P (mg/L)	SO <sub>4</sub> (mg/L)	K (mg/L)	Ca (mg/L)
Lower Limit of Detection		< 0.5	< 0.01	< 2	< 0.2	< 0.01	< 2	< 0.03	< 0.2
West Toll Gate Creek									
WTGS-2	12/17/2003 3/2/2004 12/16/2003	361 287 391	78.2 35.8 66.5	5.0 9.9 4.9	9.3 4.5 9.6	0.05 0.10 0.07	789 550 954	6.19 5.02 5.9	315 158 320
	12/17/2003 3/2/2004	395 244	94.7 40.2	3.9 10.5	11.9 5.2	0.06 0.08	904 670	6.73 4.86	381 205
WTGS-2.3	3/2/2004	199	39.7	16.7	4.4	0.03	632	5.09	185
WTGS-2.4	3/2/2004	225	39.6	10.1	5.2	0.05	641	5.18	198
WTGS-3	12/16/2003	307	53.1	< 2	9.4	0.08	758	6.72	274
	12/17/2003	284	73.5	7.8	10.0	0.10	728	7.9	323
	3/2/2004	157	26.3	8.2	4.5	0.04	435	5.17	142
WTGS-3.5	3/2/2004	169	37.1	9.2	7.2	0.05	662	5.42	226
East Toll Gate Creek									
ETGS-1	12/16/2003	272	42.1	< 2	7.9	0.01	812	3.81	262
	12/17/2003	259	59.6	2.3	8.7	0.03	767	4.51	305
	3/2/2004	446	48.0	12.9	5.8	0.03	833	4.04	260
ETGS-2	12/16/2003	1,040	107	14.8	6.4	0.10	1,380	7.08	385
	12/17/2003	723	160	10.1	8.2	0.05	1,580	6.98	515
	3/2/2004	592	79.0	9.5	5.4	0.06	1,440	4.50	325
ETGS-2 Raw, digested	12/16/2003	1,040	99.1	681	5.9	0.3	1,300	7.35	363

Site	Date	Տc (μg/L)	<b>Ті</b> (µ <b>g/L</b> )	۷ (µg/L)	Cr (µg/L)	<b>Mn</b> (μ <b>g/L</b> )	<b>Fe</b> (μ <b>g/L</b> )	<b>Со</b> (µ <b>g/L)</b>	<b>Ni</b> (μ <b>g/L</b> )
Lower Limit of Detection		< 0.6	< 0.5	< 0.5	< 1	< 0.2	< 50	< 0.02	< 0.4
Sand Creek									
SCS-1	12/16/2003	3.3	4.2	3.6	8.6	81.1	< 50	0.1	< 0.4
	12/17/2003 3/2/2004	4.2 2.4	5.4 < 0.5	2.6 5.3	6.3 3.9	70.8 217	< 50 < 50	0.11	< 0.4
Toll Gate Creek, after confluen	ce of East and West Reaches		4 010	0.0	017	211		0117	
TGS-1	12/16/2003	2.8	12.4	3.8	10.2	735	< 50	1.0	< 0.4
	12/16/2003	< 0.6	3.4	0.9	1.3	190	83	0.34	< 0.4
	12/17/2003	3.4	15.1	2.7	5.7	707	< 50	0.94	< 0.4
	3/2/2004	1.3	2.0	5.2	3.8	491	49	1.01	< 0.4
TGS-2	12/16/2003	2.7	12.6	2.5	5.2	741	< 50	0.91	< 0.4
	12/17/2003	3.1	14.5	2.7	5.7	739	< 50	0.90	< 0.4
	3/2/2004	1.4	3.8	5.6	3.5	604	76	1.06	< 0.4
TGS-2 Raw, digested	12/16/2003	1.5	21.5	0.9	< 1	743	517	1.25	4.9
West Toll Gate Creek									
OFS-1	12/16/2003	3.2	26.0	3.0	7.3	75.3	< 50	0.10	< 0.4
	12/17/2003	4.0	30.2	3.0	6.8	57.6	< 50	0.04	< 0.4
	3/2/2004	1.6	9.1	5.4	3.4	58.5	78	0.04	< 0.4
WTGS-0.5	12/16/2003	2.8	11.5	3.6	8.6	622	< 50	0.73	< 0.4
	3/2/2004	1.4	1.4	4.8	3.3	544	< 50	1.10	< 0.4
WTGS-1	12/16/2003	2.7	11.8	3.3	7.2	649	< 50	0.77	< 0.4

Site	Date	Տc (µg/L)	Ті (µg/L)	۷ (µg/L)	Cr (μg/L)	<b>Mn</b> (μ <b>g/L</b> )	<b>Fe</b> (μ <b>g/L</b> )	<b>Co</b> (μ <b>g/L)</b>	<b>Ni</b> (μ <b>g/L</b> )
Lower Limit of Detection		< 0.6	< 0.5	< 0.5	< 1	< 0.2	< 50	< 0.02	< 0.4
West Toll Gate Creek									
WTGS-2	12/17/2003 3/2/2004 12/16/2003	3.1 1.2 2.9	14.6 0.8 14.4	3.0 5.2 2.7	6.1 3.2 4.9	791 522 785	51 200 50	0.88 0.93 0.75	< 0.4 < 0.4 < 0.4
	12/17/2003 3/2/2004	4.0 1.7	17.4 5.1	3.3 4.8	7.1 4.3	982 447	71 < 50	0.92 0.47	< 0.4 < 0.4
WTGS-2.3	3/2/2004	1.6	5.2	3.4	2.9	667	97	0.47	< 0.4
WTGS-2.4	3/2/2004	1.6	3.5	5.5	3.6	568	58	0.50	< 0.4
WTGS-3	12/16/2003	2.6	10.4	3.7	7.5	41.1	< 50	< 0.02	< 0.4
	12/17/2003	3.3	14.1	3.6	6.2	60.3	48	0.0	< 0.4
	3/2/2004	1.5	0.5	3.6	3.3	55.4	< 50	0.07	< 0.4
WTGS-3.5	3/2/2004	2.2	2.7	5.1	4.0	939	61	0.49	< 0.4
East Toll Gate Creek									
ETGS-1	12/16/2003 12/17/2003 3/2/2004	2.1 3.0 1.5	11.1 14.8 5.3	2.2 1.9 3.9	6.3 4.8 4.6	604 793 654	63 64 55	0.75 0.89 0.78	< 0.4 < 0.4 1.1
ETGS-2	12/16/2003	1.9	19.3	2.5	6.2	2,210	77	4.01	< 0.4
	12/17/2003 3/2/2004	3.1 1.9	29.4 15.0	2.9 4.0	8.0 5.7	3,130 2,200	72 113	4.98 4.64	< 0.4 < 0.4
ETGS-2 Raw, digested	12/16/2003	1.2	47.3	2.8	< 1	2,120	1,140	4.88	7.8

Site	Date	<b>Cu</b> (μ <b>g/L</b> )	Zn (µg/L)	<b>Ga</b> (μ <b>g/L</b> )	<b>Ge</b> (µ <b>g/L</b> )	<b>As</b> (μ <b>g/L</b> )	<b>Rb</b> (μ <b>g/L</b> )	<b>Sr</b> (μ <b>g/L</b> )	Υ (μ <b>g/L)</b>
Lower Limit of Detection		< 0.5	< 0.5	< 0.05	< 0.05	< 1	< 0.01	< 0.5	< 0.01
Sand Creek									
SCS-1	12/16/2003	1.4	51.8	< 0.05	< 0.05	< 1	1.8	2,365	0.1
	12/17/2003	1.5	45.9	< 0.05	< 0.05	2	1.64	2,500	0.07
	3/2/2004	2.2	79.8	< 0.05	0.27	< 1	1.58	2,040	0.06
Toll Gate Creek, after confluen	ce of East and West Reaches								
TGS-1	12/16/2003	4.0	65.7	< 0.05	< 0.05	2.0	2.8	4,745	0.1
	12/16/2003	1.4	321	< 0.05	< 0.05	< 1	0.59	982	0.29
	12/17/2003	3.8	60.4	< 0.05	< 0.05	4.3	2.65	4,560	0.11
	3/2/2004	3.0	86.2	< 0.05	0.22	< 1	2.43	2,810	0.05
TGS-2	12/16/2003	3.6	72.8	< 0.05	< 0.05	1	2.70	4,680	0.11
	12/17/2003	3.3	77.4	< 0.05	< 0.05	3.5	2.41	4,460	0.11
	3/2/2004	2.9	91.7	< 0.05	0.23	< 1	2.56	2,650	0.05
TGS-2 Raw, digested	12/16/2003	4.8	145	0.08	0.2	3.1	2.89	4,420	0.23
West Toll Gate Creek									
OFS-1	12/16/2003	5.0	128	< 0.05	0.09	4.4	5.50	7.130	0.19
	12/17/2003	5.1	114	< 0.05	0.06	6.4	5.19	6,630	0.17
	3/2/2004	4.4	148	< 0.05	0.2	< 1	2.90	3.320	0.11
WTGS-0.5	12/16/2003	3.2	64.8	< 0.05	< 0.05	1	2.23	4,680	0.11
	3/2/2004	3.1	86.5	< 0.05	0.27	< 1	2.39	2,540	0.05
WTGS-1	12/16/2003	3.0	55.7	< 0.05	< 0.05	1	2.20	4,690	0.10

Site	Date	<b>Cu</b> (μ <b>g/L</b> )	Zn (µg/L)	<b>Ga</b> (μ <b>g/L</b> )	<b>Ge</b> (µ <b>g/L</b> )	<b>A</b> s (μ <b>g/L</b> )	<b>Rb</b> (μ <b>g/L</b> )	<b>Sr</b> (μ <b>g/L</b> )	Υ (μ <b>g/L</b> )
Lower Limit of Detection		< 0.5	< 0.5	< 0.05	< 0.05	< 1	< 0.01	< 0.5	< 0.01
West Toll Gate Creek									
WTGS-2	12/17/2003 3/2/2004 12/16/2003	3.6 3.3 3.1	67.2 131 62.2	< 0.05 < 0.05 < 0.05	< 0.05 0.32 < 0.05	3.3 < 1 1.0	2.49 2.54 2.7	4,540 2,510 5,600	0.10 0.05 0.10
	12/17/2003 3/2/2004	3.2 2.3	47.9 6.6	< 0.05 < 0.05	< 0.05 0.2	3.9 < 1	2.64 1.92	5,460 3,140	0.10 0.05
WTGS-2.3	3/2/2004	2.2	5.2	< 0.05	0.32	< 1	1.35	2,890	0.05
WTGS-2.4	3/2/2004	1.7	4.6	< 0.05	0.24	< 1	1.61	3,000	0.05
WTGS-3	12/16/2003	2.7	57.1	< 0.05	< 0.05	< 1	2.11	4,420	0.10
	12/17/2003	2.8	67.1	< 0.05	< 0.05	2.5	2.3	4,410	0.1
	3/2/2004	2.4	5.4	< 0.05	0.24	< 1	1.57	2,110	0.03
WTGS-3.5	3/2/2004	2.7	5.8	< 0.05	0.29	< 1	1.56	3,160	0.06
Fast Toll Cate Creek									
ETGS-1	12/16/2003 12/17/2003 3/2/2004	1.9 2.2 2.1	42.3 55.3 3.3	< 0.05 < 0.05 < 0.05	< 0.05 < 0.05 0.21	< 1 2 < 1	1.34 1.40 1.39	4,350 4,320 3,930	0.11 0.11 0.10
ETGS-2	12/16/2003	6.3	86.8	< 0.05	< 0.05	3	3.63	7,340	0.18
	12/17/2003 3/2/2004	4.4 2.6	46.4 5.6	0.06 < 0.05	< 0.05 0.23	6.3 < 1	2.58 1.84	8,460 5,890	0.16 0.39
ETGS-2 Raw, digested	12/16/2003	8.2	299	0.25	0.27	11.6	4.74	6,570	0.68

Site	Date	<mark>Zr</mark> (µg/L)	<b>Nb</b> (μ <b>g/L</b> )	<b>Μο</b> (μ <b>g/L</b> )	Cd (µg/L)	<b>Տ</b> ե (μg/L)	Сs (µg/L)	Ba (μg/L)	<b>La</b> (μ <b>g/L</b> )
Lower Limit of Detection		< 0.2	< 0.2	< 2	< 0.02	< 0.3	< 0.02	< 0.2	< 0.01
Sand Creek									
SCS-1	12/16/2003	< 0.2	0.2	3.2	< 0.02	< 0.3	< 0.02	89.7	0.0
	12/17/2003	0.2	0.33	2.8	< 0.02	< 0.3	< 0.02	92.1	0.01
	3/2/2004	< 0.2	0.48	3.0	< 0.02	< 0.3	< 0.02	72.9	< 0.01
Toll Gate Creek, after confluen	ce of East and West Reaches								
TGS-1	12/16/2003	0.2	0.2	3.3	0.1	0.5	< 0.02	65.5	0.0
	12/16/2003	0.2	< 0.2	< 2	0.16	< 0.3	< 0.02	15.6	0.34
	12/17/2003	0.2	0.28	2.8	0.05	0.30	< 0.02	56.8	0.02
	3/2/2004	< 0.2	0.66	2.7	< 0.02	0.48	< 0.02	37.1	0.01
TGS-2	12/16/2003	0.2	< 0.2	2.9	0.06	0.32	< 0.02	62.4	0.02
	12/17/2003	< 0.2	0.25	2.5	0.04	0.50	< 0.02	56.3	0.02
	3/2/2004	< 0.2	0.65	3.2	0.04	0.34	< 0.02	37.5	0.02
TGS-2 Raw, digested	12/16/2003	< 0.2	1.16	4.1	0.11	0.62	< 0.02	63.4	0.28
West Toll Gate Creek									
OFS-1	12/16/2003	0.2	< 0.2	2.7	0.09	1.60	0.02	51.1	0.03
	12/17/2003	0.2	0.25	2.1	0.08	2.86	0.03	39.6	0.02
	3/2/2004	< 0.2	0.52	7.4	0.06	6.94	< 0.02	31.1	0.03
WTGS-0.5	12/16/2003	< 0.2	< 0.2	2.5	0.04	< 0.3	< 0.02	61.8	0.02
	3/2/2004	< 0.2	0.66	2.7	< 0.02	0.36	< 0.02	35.6	0.02
WTGS-1	12/16/2003	< 0.2	< 0.2	2.5	0.04	< 0.3	< 0.02	62.6	0.02

Site	Date	<b>Zr</b> (μ <b>g/L</b> )	<b>Nb</b> (μg/L)	<b>Μο</b> (μ <b>g/L</b> )	Сd (µg/L)	Sb (µg/L)	Сs (µg/L)	Ba (μg/L)	La (µg/L)
Lower Limit of Detection		< 0.2	< 0.2	< 2	< 0.02	< 0.3	< 0.02	< 0.2	< 0.01
West Toll Gate Creek									
WTGS-2	12/17/2003 3/2/2004 12/16/2003	0.2 < 0.2 0.2	0.20 0.45 < 0.2	2.3 2.7 2.3	0.05 0.03 0.05	< 0.3 0.38 < 0.3	< 0.02 < 0.02 < 0.02	56.9 35.0 55.8	0.02 < 0.01 0.02
	12/17/2003 3/2/2004	0.2 0.2	< 0.2 1.20	2.1 4.2	0.03 0.02	< 0.3 0.59	< 0.02 < 0.02	56.0 36.7	0.01 < 0.01
WTGS-2.3	3/2/2004	< 0.2	0.81	2.4	< 0.02	< 0.3	< 0.02	36.2	0.02
WTGS-2.4	3/2/2004	< 0.2	0.71	2.3	< 0.02	0.38	< 0.02	37.9	0.01
WTGS-3	12/16/2003	< 0.2	< 0.2	< 2	0.03	< 0.3	< 0.02	56.0	0.01
	12/17/2003	0.2	< 0.2	2.0	0.02	< 0.3	< 0.02	55.1	0.0
	3/2/2004	0.2	0.85	2.8	< 0.02	0.35	< 0.02	28.2	< 0.01
WTGS-3.5	3/2/2004	0.2	1.00	2.6	< 0.02	0.47	< 0.02	42.1	< 0.01
Fast Toll Gate Creek									
ETGS-1	12/16/2003 12/17/2003 3/2/2004	< 0.2 < 0.2 < 0.2	< 0.2 < 0.2 0.49	3.9 3.9 2.8	0.02 0.02 < 0.02	< 0.3 < 0.3 < 0.3	< 0.02 < 0.02 < 0.02	78.4 81.5 57.8	0.01 0.01 0.02
ETGS-2	12/16/2003	0.3	< 0.2	3.8	0.35	0.40	< 0.02	76.2	0.08
	12/17/2003 3/2/2004	0.3 0.2	< 0.2 0.84	3.5 3.2	0.07 0.02	< 0.3 < 0.3	< 0.02 < 0.02	55.0 40.7	0.03 0.01
ETGS-2 Raw, digested	12/16/2003	< 0.2	1.61	7.3	0.48	1.23	0.13	83.3	0.96

Site	Date	Се (µg/L)	<b>W</b> (μ <b>g/L)</b>	<b>ΤΙ</b> (μ <b>g/L)</b>	<b>Рb</b> (µ <b>g/L</b> )	Bi (μg/L)	Th (µg/L)	Ս (µg/L)
Lower Limit of Detection		< 0.01	< 0.5	< 0.1	< 0.05	< 0.2	< 0.2	< 0.1
Sand Creek								
SCS-1	12/16/2003	0.0	< 0.5	< 0.1	< 0.05	< 0.2	< 0.2	29.3
	12/17/2003 3/2/2004	0.02 0.02	< 0.5	< 0.1 < 0.1	< 0.05 < 0.05	< 0.2 < 0.2	< 0.2 < 0.2	33.4 27.2
Toll Gate Creek, after confluen	ce of East and West Reache	s						
TGS-1	12/16/2003	0.0	< 0.5	< 0.1	0.1	< 0.2	< 0.2	36.4
	12/16/2003	0.74	< 0.5	< 0.1	0.53	< 0.2	< 0.2	6.80
	12/17/2003 3/2/2004	0.04 0.02	< 0.5	< 0.1 < 0.1	0.08 0.06	< 0.2 < 0.2	< 0.2 < 0.2	33.9 20.0
TGS-2	12/16/2003 12/17/2003	0.04 0.04	< 0.5 < 0.5	< 0.1 < 0.1	0.08 0.06	< 0.2 < 0.2	< 0.2 < 0.2	31.7 28.4
	3/2/2004	0.02		< 0.1	0.1	< 0.2	< 0.2	17.1
TGS-2 Raw, digested	12/16/2003	0.6	< 0.5	< 0.1	0.56	< 0.2	0.24	30.2
West Toll Gate Creek								
OFS-1	12/16/2003 12/17/2003	0.03 0.02	< 0.5 < 0.5	< 0.1 < 0.1	0.1 < 0.05	< 0.2 < 0.2	< 0.2 0.21	67.6 57.0
	3/2/2004	0.06		< 0.1	0.09	< 0.2	< 0.2	30.5
WTGS-0.5	12/16/2003	0.04	< 0.5	< 0.1	< 0.05	< 0.2	< 0.2	28.8
	3/2/2004	0.02		< 0.1	0.08	< 0.2	< 0.2	15.8
WTGS-1	12/16/2003	0.05	< 0.5	< 0.1	0.1	< 0.2	< 0.2	28.4

Site	Date	Ce (μg/L)	<b>W</b> (μg/L)	<b>ΤΙ</b> (μ <b>g/L</b> )	<b>Pb</b> (μg/L)	Bi (μg/L)	Th (µg/L)	Ս (μ <b>g/L)</b>
Lower Limit of Detection		< 0.01	< 0.5	< 0.1	< 0.05	< 0.2	< 0.2	< 0.1
West Toll Gate Creek								
WTGS-2	12/17/2003 3/2/2004 12/16/2003	0.05 0.02 0.05	< 0.5 < 0.5	< 0.1 < 0.1 < 0.1	0.06 0.1 0.1	< 0.2 < 0.2 < 0.2	< 0.2 < 0.2 < 0.2	26.0 15.2 28.9
	12/17/2003 3/2/2004	0.03 0.02	< 0.5	< 0.1 0.60	0.06 0.1	< 0.2 0.33	< 0.2 < 0.2	29.1 21.6
WTGS-2.3	3/2/2004	0.06		< 0.1	0.1	< 0.2	< 0.2	24.3
WTGS-2.4	3/2/2004	0.04		< 0.1	< 0.05	< 0.2	< 0.2	21.9
WTGS-3	12/16/2003	0.02	< 0.5	< 0.1	< 0.05	< 0.2	< 0.2	40.9
	12/17/2003	0.1	< 0.5	< 0.1	0.1	< 0.2	< 0.2	38.7
	3/2/2004	0.03		< 0.1	0.09	< 0.2	< 0.2	25.6
WTGS-3.5	3/2/2004	0.02		0.1	0.1	< 0.2	< 0.2	47.2
East Toll Gate Creek								
ETGS-1	12/16/2003 12/17/2003 3/2/2004	0.03 0.04 0.06	< 0.5 < 0.5	< 0.1 < 0.1 < 0.1	< 0.05 < 0.05 < 0.05	< 0.2 < 0.2 < 0.2	< 0.2 < 0.2 < 0.2	19.6 17.9 18.1
ETGS-2	12/16/2003	0.18	< 0.5	< 0.1	0.2	< 0.2	< 0.2	49.2
	12/17/2003 3/2/2004	0.07 0.04	< 0.5	< 0.1 < 0.1	0.05 0.05	< 0.2 < 0.2	< 0.2 < 0.2	52.3 39.8
ETGS-2 Raw, digested	12/16/2003	2.21	1.08	0.3	1.6	0.68	0.33	45.2

Appendix A2. Analytical Quality-Assurance and Quality-Control Data for Toll Gate Creek Water Samples

 $[\mu g/L, microgram per liter; mg/L, microgram per liter; Dec., December; Mar., March; QC, Quality control; NIST, National Institute of Standards$ and Technology; %, percent; <, less than; -, no value; A and B suffix on sample names indicate different replicate samples; Li, lithium; Be, beryllium, Na, sodium; Mg, magnesium; Al, aluminum; Si, silicon; P, phosphorus; SO<sub>4</sub>, sulfate; K, potassium; Ca, calcium; Sc, scandium; Ti, titanium;V, vanadium; Cr, chromium; Mn, manganese; Fe, iron; Co, cobalt; Ni, nickel; Cu, copper; Zn, zinc; Ga, gallium; Ge, germanium; As, arsenic; Se,selenium; Rb, rubidium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Mo, molybdenum; Cd, cadmium; Sb, antimony; Cs, cesium; Ba,barium; La, lanthanum; Ce, cerium; Tl, thallium; Pb, lead; Bi, bismuth; Th, thorium; U, uranium; % Difference, the difference between the certified (true) and reported value (or average of reported values) divided by the certified value and expressed as percent; Avg. Abs. Diff., the averageof the absolute values of % Difference for all analytes for one standard reference sample; Grand Avg. Abs. Diff., the average of the absolute valuesof % Difference for all analytes and all standard reference samples; Relative % Diff., the difference between replicate analyses divided by the average of the analyses and expressed as percent; Grand Avg. Precision, the average of Relative % Diff. for all analytes and all replicate samples]

OC sample, statistic, and(or) date	Li (µg/L)	<b>Ве</b> (µ <b>g/L)</b>	Na (mg/L)	Mg (mg/L)	<b>ΑΙ</b> (μg/L)	Si (µg/L)
BIAS						
Lower Limit of Detection	< 0.9	< 0.05	< 0.5	< 0.01	< 2	< 0.2
18 meg-ohm deionized water wash, all						
sample runs	< 0.1	< 0.05	< 0.5	< 0.01	< 2	< 0.2
Dec. 16/17. 2003 OC Samples						
NIST-1643d OC standard	16.1	11.8	18.6	6.89	116	2.8
Procedural Field Blank 12/16	< 0.9	< 0.05	< 0.5	< 0.01	< 2	< 0.2
Procedural Field Blank 12/17	< 0.9	< 0.05	< 0.5	< 0.01	< 2	< 0.2
NIST-1643d QC standard	20	13.6	24	9.32	152	3.1
QC Summary:						
NIST-1643d true	16.5	12.5	22.1	8	128	2.7
% Difference	9	2	-4	1	5	9
Avg. Abs. Diff.	5					
Mar. 2, 2004 QC Samples						
NIST-1643d	16.9	12.8	20.7	7.58	118	2.7
T-161	18.2	14.7	42.5	1.42	62.6	6.8
T-147	18.4	18.1	51.3	7.71	12.9	11.2
Procedural Field Blank	3.8	0.1	< 0.5	0.02	7.6	< 0.2
NIST-1643d	16.8	12.9	21.5	7.84	123	2.8
T-161	18.2	14.3	43.6	1.45	63.9	7
T-147	18.5	17.7	51.4	7.74	13	11.3

[ $\mu g/L$ , microgram per liter; mg/L, microgram per liter; Dec., December; Mar., March; QC, Quality control; NIST, National Institute of Standards and Technology; %, percent; <, less than; -, no value; A and B suffix on sample names indicate different replicate samples; Li, lithium; Be, beryllium, Na, sodium; Mg, magnesium; Al, aluminum; Si, silicon; P, phosphorus; SO<sub>4</sub>, sulfate; K, potassium; Ca, calcium; Sc, scandium; Ti, titanium; V, vanadium; Cr, chromium; Mn, manganese; Fe, iron; Co, cobalt; Ni, nickel; Cu, copper; Zn, zinc; Ga, gallium; Ge, germanium; As, arsenic; Se, selenium; Rb, rubidium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Mo, molybdenum; Cd, cadmium; Sb, antimony; Cs, cesium; Ba, barium; La, lanthanum; Ce, cerium; Tl, thallium; Pb, lead; Bi, bismuth; Th, thorium; U, uranium; % Difference, the difference between the certified (true) and reported value (or average of reported values) divided by the certified value and expressed as percent; Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes for one standard reference sample; Grand Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes and all standard reference samples; Relative % Diff., the difference between replicate analyses divided by the average of the analyses and expressed as percent; Grand Avg. Precision, the average of Relative % Diff. for all analytes and all replicate samples]

ΩC sample, statistic, and(or) date	Li (µg/L)	<b>Ве</b> (µ <b>g/L)</b>	Na (mg/L)	Mg (mg/L)	Al (μg/L)	<b>Si</b> (μ <b>g/L</b> )
BIAS						
QC Summary:						
NIST-1643d true	16.5	12.5	22.1	8.0	128	2.7
% Difference	2	2	-5	-4	-6	4
Avg. Abs. Diff.	6					
T-161 true	17.8	12.9	43.0	1.51	32.4	6.92
% Difference	2	12	0	-4	95	0
Avg. Abs. Diff.	8					
T-147 true	18.0	16.0	52.6	8.20	14.0	11.2
% Difference	2	12	-2	-6	-7	0
Avg. Abs. Diff.	6					
Grand Avg. Abs. Diff.	6					
PRECISION, FIELD REPLICATES						
SCS-1A, 12/6/03	62.5	< 0.05	155	28.7	< 2	12.6
SCS-1B, 12/6/03	62.8	< 0.05	156	28.7	< 2	12.5
Relative % Diff.	0.5	-	0.6	0	_	1.0
TGS-1A, 12/6/03	200	< 0.05	596	66.5	2.5	9.8
TGS-1B, 12/6/03	195	< 0.05	596	64.2	2.5	9.8
Relative % Diff.	3	-	0	4	0	0
WTGS-2A, 12/6/03	109	< 0.05	396	66.9	4.2	9.6
WTGS-2B, 12/6/03	108	< 0.05	385	66.1	5.6	9.6
Relative % Diff.	1	-	3	1	29	0

5

[ $\mu$ g/L, microgram per liter; mg/L, microgram per liter; Dec., December; Mar., March; QC, Quality control; NIST, National Institute of Standards and Technology; %, percent; <, less than; -, no value; A and B suffix on sample names indicate different replicate samples; Li, lithium; Be, beryllium, Na, sodium; Mg, magnesium; Al, aluminum; Si, silicon; P, phosphorus; SO<sub>4</sub>, sulfate; K, potassium; Ca, calcium; Sc, scandium; Ti, titanium; V, vanadium; Cr, chromium; Mn, manganese; Fe, iron; Co, cobalt; Ni, nickel; Cu, copper; Zn, zinc; Ga, gallium; Ge, germanium; As, arsenic; Se, selenium; Rb, rubidium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Mo, molybdenum; Cd, cadmium; Sb, antimony; Cs, cesium; Ba, barium; La, lanthanum; Ce, cerium; Tl, thallium; Pb, lead; Bi, bismuth; Th, thorium; U, uranium; % Difference, the difference between the certified (true) and reported value (or average of reported values) divided by the certified value and expressed as percent; Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes for one standard reference sample; Grand Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes and all standard reference samples; Relative % Diff., the difference between replicate analyses divided by the average of the analyses and expressed as percent; Grand Avg. Precision, the average of Relative % Diff. for all analytes and all replicate samples]

QC sample, statistic, and(or) date	P (mg/L)	SO <sub>4</sub> (mg/L)	K (mg/L)	Ca (mg/L)	<b>Տշ</b> (μ <b>g/L</b> )	Ті (µg/L)	۷ (µg/L)	Сr (µg/L)	<b>Mn</b> (μ <b>g/L</b> )
BIAS									
Lower Limit of Detection 18 meg-ohm deionized water wash, all	< 0.01	< 2	< 0.03	< 0.2	< 0.6	< 0.5	< 0.5	< 1	< 0.2
sample runs	< 0.01	< 2	< 0.03	< 0.2	< 0.6	< 0.5	< 0.5	< 1	< 0.2
Dec. 16/17, 2003 OC Samples									
NIST-1643d QC standard	< 0.01	< 2	2.19	29.6	1	< 0.5	36.5	18.9	37.5
Procedural Field Blank 12/16	< 0.01	< 2	< 0.03	< 0.2	< 0.6	< 0.5	< 0.5	< 1	< 0.2
Procedural Field Blank 12/17	< 0.01	< 2	< 0.03	< 0.2	< 0.6	< 0.5	< 0.5	< 1	< 0.2
NIST-1643d QC standard	0.01	< 2	2.44	32.3	1.2	< 0.5	40.4	20.8	43.2
QC Summary:									
NIST-1643d true	_	_	2.36	31	_	_	35.1	18.5	37.7
% Difference	_	_	-2	0	_	_	10	7	7
Avg. Abs. Diff.									
Mar. 2, 2004 QC Samples									
NIST-1643d	0.01	< 2	2.14	28.4	0.9	< 0.5	34.6	17.8	37.2
T-161	< 0.01	5	1.18	6.63	1.8	1.5	18.1	34.2	36
T-147	< 0.01	47	3.33	38.5	3	< 0.5	15.3	11.9	16
Procedural Field Blank	< 0.01	13	0.03	< 0.2	< 0.6	< 0.5	< 0.5	< 1	< 0.2
NIST-1643d	0.01	3	2.26	29.9	0.8	< 0.5	35.5	18.3	37.3
T-161	< 0.01	6	1.22	6.83	1.9	1.3	19.1	35.3	36.7
T-147	< 0.01	48	3.37	38.8	2.9	< 0.5	15.6	12.2	15.9

 $[\mu g/L, microgram per liter; mg/L, microgram per liter; Dec., December; Mar., March; QC, Quality control; NIST, National Institute of Standards and Technology; %, percent; <, less than; -, no value; A and B suffix on sample names indicate different replicate samples; Li, lithium; Be, beryllium, Na, sodium; Mg, magnesium; Al, aluminum; Si, silicon; P, phosphorus; SO<sub>4</sub>, sulfate; K, potassium; Ca, calcium; Sc, scandium; Ti, titanium; V, vanadium; Cr, chromium; Mn, manganese; Fe, iron; Co, cobalt; Ni, nickel; Cu, copper; Zn, zinc; Ga, gallium; Ge, germanium; As, arsenic; Se, selenium; Rb, rubidium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Mo, molybdenum; Cd, cadmium; Sb, antimony; Cs, cesium; Ba, barium; La, lanthanum; Ce, cerium; Tl, thallium; Pb, lead; Bi, bismuth; Th, thorium; U, uranium; % Difference, the difference between the certified (true) and reported value (or average of reported values) divided by the certified value and expressed as percent; Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes for one standard reference sample; Grand Avg. Abs. Diff., the average of the absolute values of % Difference between replicate analyses divided by the average of the analyses and expressed as percent; Grand Avg. Precision, the average of Relative % Diff. for all analytes and all replicate samples]$ 

OC sample, statistic, and(or) date	P (mg/L)	SO <sub>4</sub> (mg/L)	K (mg/L)	Ca (mg/L)	<b>Տշ</b> (μ <b>g/L</b> )	Ті (µ <b>g/L</b> )	۷ (µ <b>g/L</b> )	Сr (µg/L)	<b>Mn</b> (μ <b>g/L)</b>
BIAS									
QC Summary:									
NIST-1643d true	-	_	2.36	3.1	_	_	35.1	18.5	37.7
% Difference	-	_	-7	-6	_	_	0	-2	-1
Avg. Abs. Diff.									
T-161 true	_	_	1.26	7.17	_	_	18.4	34.6	37.4
% Difference	-	_	-5	-6	_	_	1	0	-3
Avg. Abs. Diff.									
T-147 true	_	_	3.52	41.1	_	_	15.2	12.8	17.2
% Difference	_	_	-5	-6	_	_		-6	-7
Avg. Abs. Diff.									
Grand Avg. Abs. Diff.									
PRECISION, FIELD REPLICATES									
SCS-1A, 12/6/03	0.02	301	4.9	180	3.3	4.1	3.7	9.2	80.8
SCS-1B, 12/6/03	0.01	302	4.81	180	3.3	4.2	3.4	8	81.4
Relative % Diff.	67.0	0.3	2	0	0	2	8	14	0.7
TGS-1A, 12/6/03	0.02	822	5.47	288	2.8	12	3.6	9.7	738
TGS-1B, 12/6/03	0.02	842	5.47	287	2.8	12.7	3.9	10.7	732
Relative % Diff.	0	2	0	0.3	0	6	8	10	0.8
WTGS-2A, 12/6/03	0.07	963	5.94	321	2.8	13.8	2.5	4.1	783
WTGS-2B, 12/6/03	0.07	944	5.87	318	2.9	15	2.9	5.7	786
Relative % Diff.	0	2	1	0.9	4	8	15	33	0.4

Grand Avg. Abs. Diff.

[ $\mu$ g/L, microgram per liter; mg/L, microgram per liter; Dec., December; Mar., March; QC, Quality control; NIST, National Institute of Standards and Technology; %, percent; <, less than; -, no value; A and B suffix on sample names indicate different replicate samples; Li, lithium; Be, beryllium, Na, sodium; Mg, magnesium; Al, aluminum; Si, silicon; P, phosphorus; SO<sub>4</sub>, sulfate; K, potassium; Ca, calcium; Sc, scandium; Ti, titanium; V, vanadium; Cr, chromium; Mn, manganese; Fe, iron; Co, cobalt; Ni, nickel; Cu, copper; Zn, zinc; Ga, gallium; Ge, germanium; As, arsenic; Se, selenium; Rb, rubidium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Mo, molybdenum; Cd, cadmium; Sb, antimony; Cs, cesium; Ba, barium; La, lanthanum; Ce, cerium; Tl, thallium; Pb, lead; Bi, bismuth; Th, thorium; U, uranium; % Difference, the difference between the certified (true) and reported value (or average of reported values) divided by the certified value and expressed as percent; Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes for one standard reference sample; Grand Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes and all standard reference samples; Relative % Diff., the difference between replicate analyses divided by the average of the analyses and expressed as percent; Grand Avg. Precision, the average of Relative % Diff. for all analytes and all replicate samples]

QC sample, statistic, and(or) date	<b>Fe</b> (μ <b>g/L</b> )	<b>Со</b> (µ <b>g/L)</b>	<b>Νi</b> (μ <b>g/L)</b>	<b>Cu</b> (μ <b>g/L</b> )	<b>Ζn</b> (μ <b>g/L</b> )	<b>Ga</b> (μ <b>g/L)</b>	<b>Ge</b> (µ <b>g/L)</b>	<b>A</b> s (μ <b>g/L</b> )	Se (μ <b>g/L</b> )
BIAS									
Lower Limit of Detection 18 meg-ohm deionized water wash, all	< 50	< 0.02	< 0.4	< 0.5	< 0.5	< 0.05	< 0.05	< 1	< 1
sample runs	< 50	< 0.02	< 0.4	< 0.5	< 0.5	< 0.05	< 0.05	< 1	< 1
Dec. 16/17, 2003 QC Samples									
NIST-1643d QC standard	109	25.1	57	20.1	68.9	< 0.05	0.1	52.3	11
Procedural Field Blank 12/16	< 50	< 0.02	< 0.4	< 0.5	49.8	< 0.05	< 0.05	< 1	< 1
Procedural Field Blank 12/17	< 50	< 0.02	< 0.4	< 0.5	62.1	< 0.05	< 0.05	< 1	< 1
NIST-1643d QC standard	122	27.3	61.1	21.3	76	< 0.05	0.2	57.9	12
QC Summary:									
NIST-1643d true	91.2	25	58.1	20.5	72.5	_	_	56	11.4
% Difference	27	5	2	1	0	_	_	-2	1
Avg. Abs. Diff.									
Mar. 2, 2004 QC Samples									
NIST-1643d	107	24.6	55	19.7	66.8	< 0.05	0.1	48.7	10.3
T-161	57	12.3	29.2	21.9	44.3	< 0.05	< 0.05	24.5	10.2
T-147	< 50	0.09	12.4	11.4	14.2	< 0.05	0.09	2	11.1
Procedural Field Blank	< 50	< 0.02	< 0.4	0.57	3.8	< 0.05	0.2	< 1	< 1
NIST-1643d	111	24.7	56.5	20.3	70.2	< 0.05	0.1	48.3	10.7
T-161	62	12.3	28.9	22.2	45.8	< 0.05	< 0.05	24.9	10.9
T-147	< 50	0.08	12.4	11.7	14.7	< 0.05	0.1	2	11.4

 $[\mu g/L, microgram per liter; mg/L, microgram per liter; Dec., December; Mar., March; QC, Quality control; NIST, National Institute of Standards and Technology; %, percent; <, less than; -, no value; A and B suffix on sample names indicate different replicate samples; Li, lithium; Be, beryllium, Na, sodium; Mg, magnesium; Al, aluminum; Si, silicon; P, phosphorus; SO<sub>4</sub>, sulfate; K, potassium; Ca, calcium; Sc, scandium; Ti, titanium; V, vanadium; Cr, chromium; Mn, manganese; Fe, iron; Co, cobalt; Ni, nickel; Cu, copper; Zn, zinc; Ga, gallium; Ge, germanium; As, arsenic; Se, selenium; Rb, rubidium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Mo, molybdenum; Cd, cadmium; Sb, antimony; Cs, cesium; Ba, barium; La, lanthanum; Ce, cerium; Tl, thallium; Pb, lead; Bi, bismuth; Th, thorium; U, uranium; % Difference, the difference between the certified (true) and reported value (or average of reported values) divided by the certified value and expressed as percent; Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes for one standard reference sample; Grand Avg. Abs. Diff., the average of the absolute values of % Difference between replicate analyses divided by the average of the analyses and expressed as percent; Grand Avg. Precision, the average of Relative % Diff. for all analytes and all replicate samples]$ 

QC sample, statistic, and(or) date	<b>Fe</b> (µ <b>g/L</b> )	<b>Со</b> (µ <b>g/L</b> )	<b>Νi</b> (μ <b>g/L)</b>	Си (µ <b>g/L)</b>	<b>Ζn</b> (μ <b>g/L</b> )	<b>Ga</b> (μ <b>g/L)</b>	Ge (µg/L)	<b>As</b> (μ <b>g/L</b> )	<b>Se</b> (μ <b>g/L</b> )
BIAS									
QC Summary:									
NIST-1643d true	91.2	25	58.1	20.5	72.5	_	_	56	11.4
% Difference	20	-1	-4	-2	-6	_	_	-13	-8
Avg. Abs. Diff.									
T-161 true	61.7	12.5	29.0	22.0	40.6	_	_	26.1	9.58
% Difference	-4	-2	0	0	11	_	_	-5	10
Avg. Abs. Diff.									
T-147 true	8.4	_	13.6	11.4	14.0	_	_	2.39	10.1
% Difference	_	_	-9	1	3	_	_	-16	11
Avg. Abs. Diff.									
Grand Avg. Abs. Diff.									
PRECISION, FIELD REPLICATES									
SCS-1A, 12/6/03	< 50	0.07	< 0.4	1.4	_	_	< 0.05	< 1	6.8
SCS-1B, 12/6/03	< 50	0.07	< 0.4	1.4	_	_	< 0.05	< 1	6.8
Relative % Diff.	_	0	-	0	-	-	-	-	0
TGS-1A, 12/6/03	< 50	0.99	< 0.4	3.9	_	_	< 0.05	2	36.8
TGS-1B, 12/6/03	< 50	0.98	< 0.4	4	_	_	< 0.05	2	38.1
Relative % Diff.	-	1	-	3	-	_	_	0	3
WTGS-2A, 12/6/03	< 50	0.73	< 0.4	3	_	_	< 0.05	1	34.6
WTGS-2B, 12/6/03	50	0.77	< 0.4	3.2	_	_	< 0.05	1	36.6
Relative % Diff.	_	5	_	6	_	_	_	0	6

Grand Avg. Abs. Diff.

[ $\mu$ g/L, microgram per liter; mg/L, microgram per liter; Dec., December; Mar., March; QC, Quality control; NIST, National Institute of Standards and Technology; %, percent; <, less than; -, no value; A and B suffix on sample names indicate different replicate samples; Li, lithium; Be, beryllium, Na, sodium; Mg, magnesium; Al, aluminum; Si, silicon; P, phosphorus; SO<sub>4</sub>, sulfate; K, potassium; Ca, calcium; Sc, scandium; Ti, titanium; V, vanadium; Cr, chromium; Mn, manganese; Fe, iron; Co, cobalt; Ni, nickel; Cu, copper; Zn, zinc; Ga, gallium; Ge, germanium; As, arsenic; Se, selenium; Rb, rubidium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Mo, molybdenum; Cd, cadmium; Sb, antimony; Cs, cesium; Ba, barium; La, lanthanum; Ce, cerium; Tl, thallium; Pb, lead; Bi, bismuth; Th, thorium; U, uranium; % Difference, the difference between the certified (true) and reported value (or average of reported values) divided by the certified value and expressed as percent; Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes for one standard reference sample; Grand Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes and all standard reference samples; Relative % Diff., the average of the analyses and expressed as percent; Grand Avg. Precision, the average of Relative % Diff. for all analytes and all replicate samples]

QC sample, statistic, and(or) date	<b>Rb</b> (μ <b>g/L</b> )	<b>Sr</b> (μ <b>g/L</b> )	Υ (μ <b>g/L</b> )	Zr (µg/L)	<b>Nb</b> (μ <b>g/L)</b>	<b>Мо</b> (µ <b>g/L</b> )	Сd (µg/L)	<b>Տ</b> ե (μg/L)
BIAS								
Lower Limit of Detection	< 0.01	< 0.5	< 0.01	< 0.2	< 0.2	< 2	< 0.02	< 0.3
18 meg-ohm deionized water wash, all								
sample runs	< 0.01	< 0.5	< 0.01	< 0.2	< 0.2	< 2	< 0.02	< 0.3
Dec. 16/17, 2003 QC Samples								
NIST-1643d QC standard	11.9	302	< 0.01	< 0.2	0.26	122	6.23	56.2
Procedural Field Blank 12/16	< 0.01	< 0.5	< 0.01	< 0.2	< 0.2	< 2	< 0.02	< 0.3
Procedural Field Blank 12/17	< 0.01	< 0.5	< 0.01	< 0.2	< 0.2	< 2	< 0.02	< 0.3
NIST-1643d QC standard	12.5	308	< 0.01	< 0.2	< 0.2	122	6.33	56.7
QC Summary:								
NIST-1643d true	13	295	_	_	_	113	6.47	54.1
% Difference	-6	3	_	_	_	8	-3	4
Avg. Abs. Diff.								
Mar. 2, 2004 QC Samples								
NIST-1643d	11.4	290	< 0.01	< 0.2	0.22	116	5.87	50.8
T-161	0.37	50.5	0.15	< 0.2	< 0.2	18.4	17.5	13
T-147	2.09	298	0.06	< 0.2	< 0.2	12.3	15.4	9.61
Procedural Field Blank	< 0.01	1.39	< 0.01	< 0.2	0.99	< 2	< 0.02	< 0.3
NIST-1643d	11.4	290	< 0.01	< 0.2	< 0.2	114	5.89	50.2
T-161	0.37	51.2	0.14	< 0.2	< 0.2	18.5	17.7	13.2
T-147	2.08	299	0.06	< 0.2	< 0.2	12.1	15.4	9.57

[µg/L, microgram per liter; mg/L, microgram per liter; Dec., December; Mar., March; QC, Quality control; NIST, National Institute of Standards and Technology; %, percent; <, less than; -, no value; A and B suffix on sample names indicate different replicate samples; Li, lithium; Be, beryllium, Na, sodium; Mg, magnesium; Al, aluminum; Si, silicon; P. phosphorus; SO., sulfate; K, potassium; Ca, calcium; Sc, scandium; Ti, titanium; V, vanadium; Cr, chromium; Mn, manganese; Fe, iron; Co, cobalt; Ni, nickel; Cu, copper; Zn, zinc; Ga, gallium; Ge, germanium; As, arsenic; Se, selenium; Rb, rubidium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Mo, molybdenum; Cd, cadmium; Sb, antimony; Cs, cesium; Ba, barium; La, lanthanum; Ce, cerium; Tl, thallium; Pb, lead; Bi, bismuth; Th, thorium; U, uranium; % Difference, the difference between the certified (true) and reported value (or average of reported values) divided by the certified value and expressed as percent; Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes for one standard reference sample; Grand Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes and all standard reference samples; Relative % Diff., the difference between replicate analyses divided by the average of the analyses and expressed as percent; Grand Avg. Precision, the average of Relative % Diff. for all analytes and all replicate samples]

QC sample, statistic, and(or) date	<b>Rb</b> (μ <b>g/L</b> )	Sr (μg/L)	Υ (μg/L)	<b>Ζr</b> (μ <b>g/L</b> )	<b>Nb</b> (μ <b>g/L)</b>	<b>Mo</b> (μ <b>g/L</b> )	Cd (µg/L)	<b>Տ</b> ե (μg/L)
BIAS								
QC Summary:								
NIST-1643d true	13	295	_	_	_	113	6.47	54.1
% Difference	-12	-2	_	_	_	2	-9	-7
Avg. Abs. Diff.								
T-161 true	_	54.2	_	_	_	18.9	17.5	14.0
% Difference	_	-6	_	_	_	-2	1	-6
Avg. Abs. Diff.								
T-147 true	_	313	_	_	_	11.8	15.9	10.5
% Difference	_	-5	_	_	_	3	-3	-9
Avg. Abs. Diff.								
Grand Avg. Abs. Diff.								
PRECISION, FIELD REPLICATES								
SCS-1A, 12/6/03	1.84	2,360	0.07	< 0.2	0.2	3.2	< 0.02	< 0.3
SCS-1B, 12/6/03	1.69	2,370	0.06	< 0.2	< 0.2	3.1	< 0.02	< 0.3
Relative % Diff.	8	0.4	15	_	_	3	_	-
TGS-1A, 12/6/03	2.81	4,720	0.1	0.2	0.2	3.3	0.08	0.46
TGS-1B, 12/6/03	2.81	4,770	0.11	0.2	< 0.2	3.2	0.08	0.48
Relative % Diff.	0	1	10	0	-	3	0	4
WTGS-2A, 12/6/03	2.69	5,590	0.1	0.2	< 0.2	2.3	0.04	< 0.3
WTGS-2B, 12/6/03	2.69	5,610	0.1	0.2	< 0.2	2.2	0.05	< 0.3
Relative % Diff.	0	0.4	0	0	_	4	22	_

Grand Avg. Abs. Diff.

[µg/L, microgram per liter; mg/L, microgram per liter; Dec., December; Mar., March; QC, Quality control; NIST, National Institute of Standards and Technology; %, percent; <, less than; -, no value; A and B suffix on sample names indicate different replicate samples; Li, lithium; Be, beryllium, Na, sodium; Mg, magnesium; Al, aluminum; Si, silicon; P, phosphorus; SO<sub>4</sub>, sulfate; K, potassium; Ca, calcium; Sc, scandium; Ti, titanium; V, vanadium; Cr, chromium; Mn, manganese; Fe, iron; Co, cobalt; Ni, nickel; Cu, copper; Zn, zinc; Ga, gallium; Ge, germanium; As, arsenic; Se, selenium; Rb, rubidium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Mo, molybdenum; Cd, cadmium; Sb, antimony; Cs, cesium; Ba, barium; La, lanthanum; Ce, cerium; Tl, thallium; Pb, lead; Bi, bismuth; Th, thorium; U, uranium; % Difference, the difference between the certified (true) and reported value (or average of reported values) divided by the certified value and expressed as percent; Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes for one standard reference sample; Grand Avg. Abs. Diff., the average of the absolute values of % Diff., the difference between replicate analyses divided by the average of the analyses and expressed as percent; Grand Avg. Precision, the average of Relative % Diff. for all analytes and all replicate samples]

QC sample, statistic, and(or) date	Cs (μg/L)	Ba (μg/L)	<b>La</b> (μ <b>g/L)</b>	Се (µ <b>g/L)</b>	<b>ΤΙ</b> (μ <b>g/L)</b>	<b>Ρb</b> (μ <b>g/L)</b>	Bi (μg/L)	<b>Th</b> (μ <b>g/L)</b>	<b>Ս</b> (μ <b>g/L</b> )
BIAS									
Lower Limit of Detection 18 meg-ohm deionized water wash, all	< 0.02	< 0.2	< 0.01	< 0.01	< 0.1	< 0.05	< 0.2	< 0.2	<0.1
sample runs	< 0.02	< 0.2	< 0.01	< 0.01	< 0.1	< 0.05	< 0.2	< 0.2	<0.1
Dec. 16/17, 2003 QC Samples									
NIST-1643d QC standard	4.68	532	0.02	0.01	6.7	17	11.7	< 0.2	< 0.1
Procedural Field Blank 12/16	< 0.02	< 0.2	< 0.01	< 0.01	< 0.1	< 0.05	< 0.2	< 0.2	0.1
Procedural Field Blank 12/17	< 0.02	< 0.2	< 0.01	< 0.01	< 0.1	< 0.05	< 0.2	< 0.2	0.12
NIST-1643d QC standard	4.63	529	0.02	0.01	6.9	16.9	11.3	< 0.2	<0.1
QC Summary:									
NIST-1643d true	_	507	_	_	7.3	18.2	_	_	_
% Difference	_	5	_	_	-7	-7	_	_	-
Avg. Abs. Diff.									
Mar. 2, 2004 QC Samples									
NIST-1643d	4.32	504	0.02	0.01	6.3	16.2	11	< 0.2	< 0.1
T-161	< 0.02	65.4	0.14	0.14	46.6	15.2	< 0.2	< 0.2	7.33
T-147	< 0.02	67.3	0.04	0.05	18.1	12.5	< 0.2	< 0.2	2.9
Procedural Field Blank	< 0.02	< 0.2	< 0.01	< 0.01	0.1	< 0.05	< 0.2	< 0.2	< 0.1
NIST-1643d	4.19	487	0.02	0.01	6.3	16.9	10.9	< 0.2	< 0.1
T-161	< 0.02	65.1	0.15	0.14	46.8	15.3	< 0.2	< 0.2	7.21
T-147	< 0.02	66.7	0.04	0.05	18.1	12.5	< 0.2	< 0.2	2.87

[µg/L, microgram per liter; mg/L, microgram per liter; Dec., December; Mar., March; QC, Quality control; NIST, National Institute of Standards and Technology; %, percent; <, less than; -, no value; A and B suffix on sample names indicate different replicate samples; Li, lithium; Be, beryllium, Na, sodium; Mg, magnesium; Al, aluminum; Si, silicon; P, phosphorus; SO<sub>4</sub>, sulfate; K, potassium; Ca, calcium; Sc, scandium; Ti, titanium; V, vanadium; Cr, chromium; Mn, manganese; Fe, iron; Co, cobalt; Ni, nickel; Cu, copper; Zn, zinc; Ga, gallium; Ge, germanium; As, arsenic; Se, selenium; Rb, rubidium; Sr, strontium; Y, yttrium; Zr, zirconium; Nb, niobium; Mo, molybdenum; Cd, cadmium; Sb, antimony; Cs, cesium; Ba, barium; La, lanthanum; Ce, cerium; Tl, thallium; Pb, lead; Bi, bismuth; Th, thorium; U, uranium; % Difference, the difference between the certified (true) and reported value (or average of reported values) divided by the certified value and expressed as percent; Avg. Abs. Diff., the average of the absolute values of % Difference for all analytes for one standard reference sample; Grand Avg. Abs. Diff., the average of the absolute values of % Difference between replicate analyses divided by the average of the analyses and expressed as percent; Grand Avg. Precision, the average of Relative % Diff. for all analytes and all replicate samples]

QC sample, statistic, and(or) date	Сs (µg/L)	<b>Ва</b> (µ <b>g/L)</b>	La (μg/L)	Се (µg/L)	TI (μg/L)	<b>Pb</b> (μ <b>g/L</b> )	<b>Ві</b> (µ <b>g/L</b> )	<b>Th</b> (μ <b>g/L)</b>	Ս (µ <b>g/L)</b>
BIAS									
QC Summary:									
NIST-1643d true	_	507	_	_	7.3	18.2	_	_	_
% Difference	-	-2	_	-	-14	-9	-	—	-
Avg. Abs. Diff.									
T-161 true	_	70.4	_	_	50.3	16.5	_	_	7.97
% Difference	_	-7	_	_	-7	-8	_	_	-9
Avg. Abs. Diff.									
T-147 true	_	73.0	_	_	20.0	13.8	_	_	3.21
% Difference	_	-8	_	_	-9	-9	_	_	-4
Avg. Abs. Diff.									
Grand Avg. Abs. Diff.									
PRECISION, FIELD REPLICATES									
SCS-1A, 12/6/03	_	89.2	0.01	0.03	_	< 0.05	_	_	29.3
SCS-1B, 12/6/03	_	90.1	< 0.01	0.03	_	< 0.05	_	_	29.2
Relative % Diff.	-	1	-	0	_	_	-	_	0.3
TGS-1A, 12/6/03	_	65	0.01	0.03	_	0.07	_	_	35.7
TGS-1B, 12/6/03	-	66	0.02	0.04	_	0.07	_	_	37
Relative % Diff.	-	2	67	29	-	0	-	-	4
WTGS-2A, 12/6/03	_	55.7	0.02	0.04	_	0.08	_	_	28.9
WTGS-2B, 12/6/03	_	55.9	0.02	0.05	_	0.08	_	_	28.8
Relative % Diff.	_	0.4	0	22	_	0	-	_	0.3

Grand Avg. Abs. Diff.

**APPENDIX B.** Quality Assurance and Quality Control Procedural Details

#### 58 Selenium and Other Elements in Water and Adjacent Rock and Sediment of Toll Gate Creek, Colorado

Bias was evaluated using percent difference (the difference between the certified and reported value (or average of reported values) divided by the certified value and expressed as percent for each certified element for a standard reference sample). This value generally was 10 percent or less for NIST-1634d for the December 2003 and March 2004 analyses except for Fe which was 20 percent or greater for each event. Percent-difference values for standard reference samples T-147 and T-161 were generally less than 15 percent and did not validate the high Fe bias shown by NIST-1643d. The average of the absolute values of percent difference (Average Absolute Difference in table A2) for all elements for one standard reference sample estimates the overall bias. This value was between 5 and 8 percent for standard reference samples that were associated with environmental samples collected during December 2003 and March 2004 (table A2). Another estimate of overall bias is the grand average absolute difference (which is the average of the absolute values of percent difference for all elements for all standard reference samples) which was 6 percent for both sampling events (table A2). These results indicate that there is good control on bias, except possibly for Fe analyses.

For precision, the relative percent difference (the difference between replicate analyses divided by the average of the analyses and expressed as percent) for analysis of three laboratory replicates of samples from the December 2003 sampling event was  $\pm$  5 percent, which indicates good precision for the samples collected and analyzed in December 2003 (table A2). Replicate analyses of NIST-1634d, T-147, and T-171 provided information to assess precision for the March 2004 samples. Relative percent difference was generally less than 5 percent for all elements (except in a few instances where the value was near the detection limit) and the average absolute difference for each reference material was less than 2 percent (raw data, but not relative percent difference calculations, are listed in table A2).

Element concentrations for the deionized water (DIW) blanks (wash) that accompanied ICP-MS analysis in the

laboratory were all less than the detection limits. For the procedural field blank that used DIW from the same laboratory system, the concentrations of most elements were at or near the detection limits (table A2). However, zinc (Zn) was detected in the procedural blanks from December 16 and December 17, 2003, samples at concentrations of 50 and 68 micrograms per liter ( $\mu$ g/L). For several of the water samples collected on this date, the Zn concentration in the procedural blank was near to or exceeded Zn concentrations for the samples indicating possible Zn contamination in the environmental samples introduced by sample collection and processing. However, there were also several environmental samples that had much lower Zn concentrations than the procedural blank samples, which indicates that the contamination probably was not in the acid used to preserve the samples. Contamination may have been in the laboratory DIW used for the blank, or it may be random. The procedural blank for the March 2004 sampling event had a much lower Zn concentration (3.8  $\mu$ g/L) indicating that Zn concentrations in the environmental samples probably were reliable. Because interpretation of Zn data was not a major part of this study, the possible contamination of environmental samples collected during December 2003 with Zn does not interfere with the interpretations presented herein.

The NIST 1643d standard was analyzed for qualityassurance/quality-control purposes with the two whole-water digestions. The percent difference between known and measured values ranged from 0 to 9, and the average absolute percent difference for all elements was 1 percent indicating good control for theses analyses (these values are not listed in table A2).

The water-concentration data for Se in the duplicate site samples indicate that there was little to no variability at these sites during the sampling events. In addition, selenium results for the December sampling events 1 day apart show that there was little compositional variation at each site from day-to-day during base flow.