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Particulate Matter (PM) Research Centers (1999-2004) and the Role of Interdisciplinary Center-Based Research

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Elinor Fanning UCLA School of Public Health 650 Charles E. Young Drive, South Los Angeles, CA 90095-1772 (310) 206-1229 – phone (310) 206-9903 – fax efanning@ucla.edu Running Title: U.S. EPA PM Centers Accomplishments (1999-2004)

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List of Abbreviations

CAPs – concentrated ambient particles CHS – Children's Health Study CO – carbon monoxide COPD – chronic obstructive pulmonary disease U.S. EPA – Environmental Protection Agency HRV – heart rate variability NAAQS – National Ambient Air Quality Standards NRC – National Research Council NMMAPS – National Morbidity and Mortality Air Pollution Study OC – organic carbon PAH – polycyclic aromatic hydrocarbons PM – particulate matter ROS – reactive oxygen species UFP – ultrafine PM

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ABSTRACT

Objective: The U.S. Environmental Protection Agency funded five academic Centers in 1999 to address the uncertainties in exposure, toxicity and health effects of airborne particulate matter identified in the National Research Council's (NRC) "Research Priorities for Airborne Particulate Matter". The Centers were structured to promote interdisciplinary approaches to address the NRC's research priorities. This report presents selected accomplishments from the PM Centers' first six years, with a focus on the advantages afforded by the interdisciplinary, center-based research approach. The review highlights advances in the area of ultrafine particles and traffic related health effects as well as cardiovascular and respiratory effects, mechanisms, susceptibility and PM exposure and characterization issues. Data sources and Synthesis: The collective publications of the Centers served as the data source. To provide a concise synthesis of overall findings, authors representing each of the five Centers identified a limited number of topic areas that serve to illustrate the key accomplishments of the PM Centers program and a consensus statement was developed. Conclusions: The PM Centers program has effectively applied interdisciplinary research approaches to advance PM science.

Introduction

The U.S. Environmental Protection Agency (U.S. EPA) funded five academic Centers in 1999 to address the uncertainties in exposure, toxicity and health effects of airborne particulate matter (PM) identified in the National Research Council's (NRC) "Research Priorities for Airborne Particulate Matter" (1998). Centers were established at Harvard University, New York University, the University of Rochester, the University of Washington, and three universities in Southern California. All Centers were structured to promote interdisciplinary approaches to address the NRC's research priorities. A midterm report of PM Center findings was published previously (Lippmann et al. 2003). This report highlights selected accomplishments from the PM Centers' first six years, with a focus on the advantages of interdisciplinary, center-based research. A more detailed summary of research findings and bibliography may be found in the supplemental material available on the U.S. EPA PM Centers website at http://es.epa.gov/ncer/science/pm/centers.html.

PM Exposure Research Highlights

Characterization of Ambient PM

The PM Centers worked to characterize ambient PM and the substantial variation of concentration and composition with source, region, seasonal and diurnal patterns, and size fraction. Examples of findings follow. In the eastern United States, $PM_{2.5}$ (PM with aerodynamic diameter < 2.5 µm) composition varies seasonally, with relatively more sulfate from long range transport in the winter, and nitrate in the summer. Substantial spatial variability in PM components and co-pollutants was observed (Maciejczyk and Chen 2005). In the Pacific

Northwest, organic carbon (OC) derived from wood burning is a major contributor to fine particle mass (Larson et al. 2006). PM_{10} (PM < 10µm in aerodynamic diameter) collected in Southern California derive largely from road dust and soil and contain significant quantities of metals, whereas $PM_{2.5}$ from the same locations contain primarily nitrates, OC and elemental carbon (EC). Ultrafine PM (UFP, PM < 0.1 µm) is especially high in OC (Sardar et al. 2005). Semivolatile components of PM have received increased attention in recent investigations, especially with regard to combustion-derived UFP in which a significant fraction of emissions by mass can consist of semi-volatile material that has condensed onto a non-volatile, primarily carbon core (Kuhn et al. 2005a; Robinson et al. 2007). Atmospheric processes generate UFP in regions of the Los Angeles air basin that receive advected pollutant air masses (Fine et al. 2004; Singh et al. 2006). The role of atmospheric chemistry in formation of UFP is important: photooxidation of diesel emissions rapidly generates organic PM (Ntziachristos et al. 2007).

Source Apportionment

Research on sources emphasized mobile sources/traffic during the first six years of the PM Centers (see below) A workshop was held by the PM Centers to compare different methods for source apportionment of PM. The outcomes of different analytical methods found good agreement across different investigators and methods in apportioning sources of PM_{2.5} mass in two US cities, Phoenix, Arizona and Washington, D.C. (Hopke et al. 2006; Thurston et al. 2005). Center research also included identification of tracer compounds for use in identifying sources of ambient particles (Fine et al. 2004).

Personal Exposure

A significant body of data on personal exposure resulted from the PM Centers' field studies, including longitudinal studies conducted in different airsheds, populations and housing. Extensive intra- and inter- personal variability in the ratio of personal to ambient exposure measures was observed in some studies (Liu et al. 2003), but taken collectively the data establish that ambient air concentrations at central site monitors can yield valid estimates of average personal exposure for population-based epidemiological studies (Sarnat et al. 2000, 2002). The location of central site monitors, extent of PM penetration into indoor environments, personal activities, and the influence of indoor PM sources can affect personal/ambient exposure ratios (Larson et al. 2004; Sarnat et al. 2006). The effects of these factors differ with PM size and composition; for example, freeway derived UFP in the 70-100nm range penetrated indoors to a greater extent than 10-20nm PM (Zhu et al. 2005). The relationship of ambient criteria pollutant concentrations to ambient and personal PM_{2.5} was explored. Ambient criteria pollutant levels were better predictors of personal PM_{2.5} than they were of personal exposure to the gaseous species themselves, suggesting that the criteria pollutants may be useful as surrogates of PM_{2.5} exposure, but are unlikely to act as confounders in epidemiological studies (Sarnat et al. 2005). In a study of ambient UFP, hourly and 24-hr number concentrations were not significantly associated with concentrations of gaseous co-pollutants (Sardar et al. 2004).

PM Health Effects and Mechanisms of Injury Highlights

During U.S. EPA's effort to establish a national ambient air quality standard for fine particles, considerable questions about the biological plausibility of epidemiological findings on

hospitalization and mortality from cardiopulmonary effects arose. As a result the NRC committee recommended research into the mechanisms of injury that underlie PM health effects, especially daily mortality. Developments in defining toxicological mechanisms and intermediate clinical conditions that may explain the observed cardiovascular mortality are one of the highest impact area of the PM Centers' scientific contributions in particular by addressing PM size specific research, e.g., ultrafine, fine and coarse PM.

PM Effects on the Cardiovascular System

The PM Centers convened a workshop to discuss potential mechanisms of PM-associated cardiovascular effects, and to identify fruitful research approaches (Frampton et al. 2008; Utell et al. 2002) (Figure 1). During the first 6 years, Center investigators have contributed to several review papers on cardiovascular responses to inhaled UFP and PM_{2.5} (Brook et al. 2004; Delfino et al. 2005; Godleski 2006; Mar et al. 2006; Pope and Dockery 2006). New statistical methodology was developed and applied to strengthen the interpretation of acute mortality studies (Coull et al. 2001; Janes et al. 2005; Schwartz and Coull 2003; Zanobetti et al. 2000, 2001; Zeka and Schwartz 2004). Epidemiological studies that focused on specific cardiovascular outcomes, such as myocardial infarction (Peters et al. 2001, 2004; Zanobetti and Schwartz 2005) or cause-specific mortality (Franklin et al. 2007; Miller et al. 2007; Pope et al. 2002; Zeka et al. 2005) produced hypotheses for testing in laboratory animal research and human clinical studies. Toxicologists have contributed by identifying cellular and biomolecular mechanisms involved in the cardiovascular effects that result from acute and long-term exposures to ambient PM (Araujo et al. 2008; Corey et al. 2006; Lippmann et al. 2005a, 2006; Sun et al. 2005). Most recently, toxicological studies (Ghelfi et al. 2008) have shown that increases in reactive oxygen species in

the heart associated with inhalation of concentrated ambient particles may be abrogated by blocking neural receptors in the lung (Figure 2).

Investigations in the PM Centers and elsewhere supported the hypothesis that inflammatory responses contribute to cardiovascular toxicity. Possible mechanisms were proposed. Pulmonary inflammation could release reactive oxygen species (ROS), cytokines and chemokines from the lung to the systemic circulation (Frampton et al. 2006b). Vascular inflammatory markers were associated with PM2.5 exposure in a subchronic mouse study (Sun et al. 2005). Gong et al. (2007) demonstrated that both diesel extract and oxidized lipid components synergistically affect the expression profile of several gene modules that are related to vascular inflammatory processes. Evidence for an increase in C-reactive protein and a shift to a pro-coagulatory state of the blood was seen in coronary artery disease patients exposed to various size fractions of PM (Rückerl et al. 2006). Temporal and other parameters differed with the specific air pollution mixture in this study, limiting interpretation. Pope et al. (2004), concluded fine particulate air pollution is a risk for cause specific cardiovascular disease mortality via inflammation, accelerated atherosclerosis and altered autonomic function. Zeka et al. 2006 reached similar conclusions; their epidemiological study supports the hypothesis that particles can induce cardiovascular disease through inflammatory pathways and suggests greater toxicity of traffic-related particles.

Autonomic function effects manifested as alterations in heart rate and heart rate variability (HRV) have been associated with $PM_{2.5}$ exposure. Decreased HRV was associated with $PM_{2.5}$ exposure in panel studies of elderly subjects (Adar et al. 2007; Henneberger et al. 2005;

Schwartz et al. 2005a). No associations with altered heart rate or HRV were seen in Seattle during the winter woodburning season (Mar et al. 2005b; Sullivan et al. 2005). A population-based study that drew on an established cohort (the Normative Aging Study) confirmed the association between decreased HRV and $PM_{2.5}$ seen in other studies; history of ischemic heart disease, hypertension, and diabetes modified the effects of $PM_{2.5}$ (Park et al. 2005). Cardiac arrhythmias and vascular changes such as endothelial cell responses and alterations in blood pressure are other important clinical signs of cardiovascular toxicity that have been identified in both humans and animals exposed to PM (Frampton et al. 2006b; Gong et al. 2004; Nadziejko et al. 2002).

Atherosclerosis is emerging as an important toxic endpoint of $PM_{2.5}$ exposure. Atherosclerosis findings may be related to reports of myocardial infarction associated with $PM_{2.5}$ in epidemiologic studies (Peters et al. 2004; Zanobetti and Schwartz 2005). The Peters study relates traffic exposures and myocardial infarction. Atherosclerotic lesions in a susceptible mouse model were enhanced by $PM_{2.5}$ exposure in a number of reports (Araujo et al. 2008; Chen and Hwang 2005; Chen and Nadziejko 2005; Lippmann et al. 2005b; Sun et al. 2005). Araujo et al. (2008) compared the proatherogenic effects of ambient UFP with $PM_{2.5}$ in apolipoprotein Edeficient mice. UFP exposed mice exhibited significantly larger atherosclerotic lesions than mice exposed to $PM_{2.5}$ or filtered air (Figure 3).

Respiratory Effects of PM Exposure

PM Centers research has added to a wide body of literature investigating toxicological mechanisms and effects of PM in the respiratory system. Overall, the issue of respiratory effects and PM exposure has been reviewed recently with reference to work produced by the PM

Centers as well as others (Boothe and Shendell 2008; Salam et al. 2008). Salam focuses on asthma whereas the Boothe and Shendell paper also addresses some other endpoints besides respiratory. Results from clinical and panel studies in asthmatic and elderly subjects, as well as experimental studies in animals and in vitro cellular systems with relevance to respiratory tissues were performed. The discovery that UFP deposition is increased in asthmatic patients and during exercise has important implications for defining populations at greater risk of PM-related effects (Chalupa et al. 2004; Daigle et al. 2003). Adjuvant effects of ambient PM in promoting allergic airways responses occurred in a sensitized mouse model (Kleinman et al. 2005). Acute exposures to ambient PM in Seattle were associated with increased inflammation in asthmatic subjects as measured by exhaled nitric oxide (Jansen et al. 2005; Koenig et al. 2005; Mar et al. 2005a). Respiratory effects in children were a focus. Increased risk of infant hospitalization for bronchiolitis was significantly associated with subchronic and chronic exposures to PM in Los Angeles (Karr et al. 2007) where exposures in the month prior to hospitalization (subchronic) and mean lifetime exposure (chronic) referenced to the case diagnosis date were assessed on the basis of data derived from the California Air Resources Board. Epidemiological studies that linked the PM Centers and the Children's Health Study (CHS) contributed findings that identify infants and children as important populations of concern for respiratory effects of PM (Gauderman et al. 2004, 2005, 2007; Molitor et al. 2007; Trenga et al. 2006). These studies demonstrate that exposure to PM_{2.5} and other air pollutants were associated with reduced lung function growth in children and provided evidence for compromised lung function. The CHS/PM Center studies identified traffic as a factor (Gauderman et al. 2004, 2005, 2007; McConnell et al. 2006).

Identification of New Target Tissues

UFP of Carbon-13 were detected in the olfactory bulbs of rats after inhalation exposure (Oberdörster et al. 2004), suggesting that the central nervous system is a potentially important toxicological target of $PM_{2.5}$ (Figure 4). In support of this significant result, studies of mice chronically exposed to ambient $PM_{2.5}$ documented loss of brain neurons (Veronesi et al. 2005) and changes in gene expression in the brain consistent with inflammatory effects (Gunnison and Chen 2005). In another study, proinflammatory cytokines were increased in brains of mice exposed to concentrated $PM_{2.5}$ compared to that of control animals. (Campbell et al. 2005).

Chemical Mechanisms of PM Toxicity

To better identify the most toxic PM components and sources, the PM Centers have pursued experimental linkages between toxicological properties and specific physical/chemical characteristics of particles including size, surface area, and PM components such as transition metals, endotoxin, and organics including reactive organic compounds. Multiple chemical and biological mechanisms by which PM can induce toxic effects in a variety of target cell types have been proposed (Frampton 2006; Yang et al. 2008). Oxidative stress, a common effect of toxicant exposure, is a change in the redox environment of the cell (Schafer and Buettner 2001) through changes in the ratios of concentrations of oxidized to reduced cellular antioxidants. Oxidative stress occurs by increasing intracellular ROS or by depleting Glutathione (GSH). GSH is the predominant antioxidant in cells and plays important roles in protecting against oxidative and electrophile stress (Rahman and Macnee 2000). A number of PM Center studies during the first 6 years contributed to what is now a strong evidentiary basis for oxidative damage as a general toxicological mechanism of PM injury (Delfino et al. 2005; Ghelfi et al. 2008; Gonzalez-Flecha 2004; Gurgueria et al. 2002; Li et al. 2003b; Rhoden et al. 2004,

2005 ; Tao et al. 2003; Xia et al. 2006). There is widespread agreement throughout the PM Centers that oxidative stress may be a mechanism of major importance for cardiorespiratory effects.

Studies of reactive chemical components of ambient PM samples reported that particles possess intrinsic chemical reactivity that may play an important role in toxicity (Cho et al. 2005; Venkatachari et al. 2005). Covalent modification of biological molecules by reactive electrophilic compounds, particularly organics, and ROS production are two key chemical mechanisms by which PM can disrupt intracellular biochemistry, ultimately altering gene expression and subcellular organelle function in target cells. Center investigators demonstrated covalent binding of a cellular enzyme by electrophilic agents, including organic compounds, present in ambient PM (Rodriguez et al. 2005; Samet et al. 1999), and that PM can directly inhibit the activity of enzymes involved in oxidative stress response in a cell free assay (Hatzis et al. 2006). There is accumulating evidence that transition metals, such as copper, vanadium, chromium, nickel, cobalt and iron, as well as aromatic and polar organic substances play a role in ROS production. An important role of metals may be alteration of signal transduction pathways involving oxidative stress (Samet et al. 2003). Assays that can screen for both oxidative and covalent binding properties of PM are of interest for comparing the toxicological potential of PM from different sources, locations of interest, season, and other parameters of interest (Borm et al. 2007).

Life Shortening Associated with Exposure to Particulate Matter

In analyses at the Harvard Center in which daily deaths in 10 European cities were investigated by examining all cause, respiratory, and cardiovascular deaths, for all ages and stratifying by age groups, it was found that the effect of air pollution is not limited to advancing mortality by a few weeks, but that effects persist for over a month after exposure. The short-term mortality effect size estimate for PM_{10} doubles when longer-term effects for all mortality and cardiovascular mortality were considered and becomes five times higher for respiratory mortality (Zanobetti et al. 2003). Reduction of ambient air pollution levels was associated with reduced total, cardiovascular, and lung cancer mortality in the Harvard Six Cities Cohort (Laden et al. 2006). Long term exposure was associated with excess lung cancer in cohort studies of Pope et al. (2002), Laden et al. (2006) and Pope and Dockery (2006).

Susceptibility Factors and Populations of Concern for PM-Induced Health Effects

When the PM Centers research was initiated, epidemiological studies had indicated that the elderly and people with cardiovascular or chronic lung disease were at greater risk for morbidity and mortality associated with acute PM exposure. The PM Centers explored the basis for this susceptibility and also produced research findings that expand the spectrum of populations of concern. Support for the epidemiological observations that elderly and chronic obstructive pulmonary disease (COPD) patients have higher rates of hospitalization and mortality associated with acute PM exposure has come from human clinical studies showing that elderly people experience greater effects of PM on HRV and blood parameters (Park et al. 2005; Pope and Dockery 2006; Schwartz et al. 2005a, 2005b). Further support for the elderly as a population of concern comes from studies of geriatric laboratory animals (Elder et al. 2004a, 2004b).

A study of PM-related daily mortality found greater effects in diabetic subjects (Zeka et al. 2006). The increase in mortality in diabetics may related to increased susceptibility to the cardiovascular effects of PM exposure, as indicated by greater rate of hospitalization for heart

disease (Zanobetti and Schwartz 2002), sensitivity to changes in HRV (Park et al. 2005) and altered vasomotor function (O'Neill et al. 2005) in diabetic subjects. It is possible that these patients may be more susceptible to inflammatory effects of PM, which in turn affect vascular tissues (O'Neill et al. 2007). In contrast, recent results from the Women's Health Initiative suggest that diabetics in this cohort were not at increased risk (Miller et al. 2007). More work on this subject is needed, and controlled human exposures in diabetic studies have been initiated by the PM Centers (Frampton et al. 2006a). Schwartz et al. (2005b) reported an association between presence or absence of the allele for glutathione S transferase M1 (GSTM1) and the high frequency component of heart rate variability; genetic susceptibility is an area in which the PM Centers are currently increasing research focus.

Advances in Critical Interdisciplinary Research Areas

Interdisciplinary research has been a hallmark of the PM Centers since their inception. Two subject areas that were exemplary in terms of bringing together multiple investigative perspectives were investigations of UFP and mobile sources.

Ultrafine Particles: Unique in Composition and Toxicity

Center-based research allowed a major effort to characterize size distributions, chemical speciation and the effect of atmospheric processes of UFP to be integrated with toxicological research (Donaldson and Stone 2003). UFP in urban airsheds are largely derived from fresh combustion sources, although secondary formation of UFP from atmospheric photochemical processes is also an important source (Sioutas et al. 2005). UFP freshly generated by combustion are short lived and subsequently grow to form aggregates. UFP dominate particle number concentration in ambient PM samples while contributing little to PM mass concentrations. Due

in part to a complex fractal structure (Friedlander and Xiong 2000), UFP possess much greater surface area per unit mass than larger ambient particles. The large surface area in turn allows greater per mass concentrations of adsorbed or condensed toxic air pollutants (oxidant gases, organic compounds, transition metals) to collect on UFP (Sioutas et al. 2005). Studies on ambient and model particles have concluded that the large specific surface area of UFP may be a key component in their toxicology (Oberdörster 2001).

The PM Centers produced an integrated body of exposure and toxicological studies on ambient and model UFP as well as studies of controlled human exposures. Dosimetry work showed that UFP will have significant accumulation in the lung (Kreyling et al. 2006). In addition, UFP of varying composition can cross cellular membranes by diffusion (Geiser et al. 2005) and gain access to vulnerable targets within cells. The potential for translocation from the site of lung deposition into systemic circulation, while rates have been low with test particles (Kreyling et al. 2002), could have major mechanistic implications (Elder and Oberdörster 2006). Electron microscopy indicated subcellular penetration and mitochondrial damage by UFPs in *in vivo* studies and, to a lesser extent, fine particles (Li et al. 2003b). Disruption of mitochondrial functions may play an important role in PM-mediated health effects (Xia et al. 2007).

In a study of size segregated concentrated ambient PM samples the ability of PM to catalyze ROS generation, an initial step in the induction of oxidative stress, was greatest in the UFP fraction (Cho et al. 2005). Li et al. (2003a) summarized contrasting features of coarse, fine and ultrafine particles from Southern California, including relevant chemical and biological

parameters. The toxicological findings correlated with PM organic carbon and PAH composition, suggesting a role of organic agents in generating redox activity (Table 1).

The PM Centers conducted controlled human exposure studies with UFP. Results from these studies were limited, perhaps due to small group sizes and because these exposures are necessarily brief and conducted at low concentrations compared to the background PM exposures that may be experienced by urban study subjects. In the first set of studies, short term exposures were conducted with 10-50µg/m³ carbon UFP generated in the laboratory. Alterations in blood cell adhesion molecules and in a marker of vascular perfusion suggest that UFP exposure may produce subtle changes in pulmonary vasoconstriction (Frampton 2007; Pietropaoli et al. 2004). A small but statistically significant reduction in arterial oxygen saturation and some evidence for reduced HRV were found although the small study size limited interpretation (Gong et al. 2008). An expanded focus on UFP in epidemiological studies is needed, but has been limited to date by the challenges of assessing exposure to UFP.

Traffic: Mobile Sources are Highly Relevant to the Public Health Impacts of PM

The Center-based research context was particularly useful in advancing the science on mobile sources of PM, the focus of an extensive international research effort. Numerous investigations of the physical and chemical attributes of PM collected alongside freeways and in roadway tunnels were performed. The results have yielded data on size distribution, number and mass concentrations, chemical speciation, emissions factors, volatility, penetration indoors and the impact of atmospheric processes on roadway PM (Biswas et al. 2007; Fine et al. 2004; Geller et al. 2006; Kuhn et al. 2005b, 2005c; Phuleria et al. 2007; Sardar et al. 2005; Zhu et al. 2005). Detailed spatial profiles of UFP concentration at varying distances from freeways were generated

(Zhu et al. 2002a, 2002b). Concentrations of UFP drop exponentially with distance from the center of the freeway, reaching upwind levels at approximately 300 meters. The size distribution of UFP also changed markedly with distance reflective of coagulation and other atmospheric particle processes. Winter particle number concentrations are greater than summer indicating formation of UFP from vapor condensation. Exposure to motor-vehicle exhaust emissions during commuting may constitute a substantial fraction of daily personal PM exposure, especially to UFP (Sioutas et al. 2005; Zhu et al. 2007).

Toxicological studies of traffic-derived aerosols studied by PM Centers included in vitro findings that implicate PM collected in freeway microenvironments in the production of reactive chemical species, stimulation of pro-inflammatory effects and altered gene expression in cellular test systems. UFP fraction, carbonaceous content, and an organic tracer for vehicles were linked with toxicological activity of PM in a variety of assays (Cho et al. 2005; Li et al. 2003a, 2003b). Several studies of laboratory animals exposed to PM on or near busy roadways have identified cardiovascular and allergic airways effects (Elder et al. 2004b, 2007; Kleinman et al. 2005). Evidence that traffic-derived air pollution affects humans has expanded significantly during the first six years of PM Centers funding, to implicate mobile source in respiratory effects in children (Gauderman et al. 2004, 2005, 2007; McConnell et al. 2006), cardiovascular effects (Riediker et al. 2004) including myocardial infarction (Peters et al. 2004; Tonne et al. 2007), and low birth weight (Wilhem and Ritz 2003). Toxicological studies are needed to follow up the epidemiological findings of effects on the fetus. In a reanalysis of data from the Harvard Six Cities study of daily mortality and PM, source apportionment approaches identified the mobile source factor as most strongly associated with increased daily mortality (Laden et al. 2000).

Policy Implications of PM Centers Research

Research findings from the PM Centers have had a significant influence on science policy, most directly in terms of the science that underlies the National Ambient Air Quality Standards (NAAQS) for PM. The findings of morbidity and mortality that form the scientific basis for the short-term and annual PM NAAQS were strengthened through epidemiological and statistical research. Mechanistic investigations and studies of preclinical markers established biological plausibility for observed relationships between ambient air PM and observed acute mortality. Validation of the use of central site ambient concentrations, in personal exposure studies, provided crucial support to the interpretation of epidemiological results.

The PM NAAQS are based on mass concentration. The state of the science suggests that no single parameter, whether mass, size fraction, surface area, or a particular chemical component is responsible for all the diverse mechanisms and toxicological endpoints that have been associated with PM and a more sophisticated approach to standards will be needed. Based on findings from the PM Centers and others, the potential efficacy of number and component based standards should be assessed. As more data becomes available to link specific PM emissions sources, chemical composition, and physical characteristics with quantitative measures of toxicity, the question of source-specific control strategies to maximize public health protection also needs to be considered.

The increasing level of evidence that UFP are toxic, but may not be controlled well by existing regulatory approaches, raises other policy issues including mitigation of the risk of health effects

associated with housing, schools, parks and other heavily populated public facilities located near heavily-traveled roadways, busy seaports, and other combustion sources that are the major urban sources of exposure to UFP. There are potential environmental justice concerns associated with transportation derived combustion, since it is often areas of lower socioeconomic status that are most affected by proximity to these sources.

Looking Forward: Research Priorities and Current Directions

As the PM Centers program moved forward into the second phase, the original guiding research priorities were reevaluated, and new priorities have emerged. Several areas of investigation that were identified during the development of the 1997 PM NAAQS are still of critical relevance today, but the scientific questions being asked have been refined. Some research topics being pursued in the current round of PM Centers are described below.

Particle Source Characterization and PM Components as Factors in PM Toxicity

The PM Centers current research agenda includes detailed studies of the physical and chemical attributes of ambient PM associated with specific sources. The current science indicates multiple mechanisms of injury, in backgrounds modified by host susceptibility factors, can be activated by a variety of PM components and characteristics. To address the complexity associated with assessing the health effects associated with specific PM components, the current PM Centers research agenda compares toxicological properties of PM by source type in addition to compositional attributes. Mobile sources continue to be a priority focus and there is a need to better understand the fate of fossil fuel combustion emissions from a variety of mobile and stationary sources, including airports, seaports, and other sources as well as roadways. Building upon the productive body of work on mobile source PM in the first six years of PM Center work,

the current PM Centers include human panel and clinical studies and toxicological studies in laboratory animals and *in vitro* systems that test hypotheses about the effects of mobile source PM exposures. Source apportionment efforts are ongoing as well, to build on previous work that found mobile sources are dominant contributors to urban UFP loads. *In vitro* studies will pay particular attention to UFP, to organic compounds, and to transition metals. UFP formed from nucleation of ambient air vapors are a new focus, as they may be especially toxic.

Dosimetry and Toxicokinetics

Research at the PM Centers is addressing particle deposition, uptake, distribution, and fate, including the effects of developmental stage on disposition of PM. Cell culture systems with gene expression and proteomics methods are being used for studies of metabolic and genetic responses that will be useful for toxicokinetics. Studies of the dosimetry and toxicokinetics associated with UFP are especially important given the previous PM Centers' findings that these particles distribute into systemic circulation and secondary target organs such as the CNS and can enter cells and subcellular organelles.

Mechanisms

All the current PM Centers have a strong focus on continuing to develop understanding of the toxic mechanisms that underlie clinically and epidemiologically defined adverse health effects of PM. Mechanisms being pursued include reactive chemical species that cause cellular oxidative stress responses. In the first six years, studies of oxidative damage associated with PM were performed in studies using diverse chemical species, cell culture experiments and laboratory animal studies. Evolving out of that work, the current PM Centers include studies that are looking at markers of oxidative stress processes in humans, and a range of clinical and pre-

clinical biomarkers. The list of gene products that can be used as indicators of PM exposure or toxicity in various cell types has expanded. Mechanistic hypotheses are being tested in panel and other epidemiological studies.

Susceptibility

Susceptibility is a major theme, drawing on the work from the earlier center and non-center investigators showing that individuals with pulmonary and cardiac health conditions, the elderly, children, diabetics, and others may be more susceptible to the adverse effects of PM exposure than the general population. The PM Centers are looking at early life exposures to PM in animal models, performing panel studies of elderly subjects or subjects with compromised health status, using a large established cohort to identify how risk factors for PM related health outcomes may be modified by individual factors such as medication use, diet and genotype. Compromised animal models are a key theme of current research into susceptibility. PM exposure studies on ApoE-/- mice (an atherosclerosis prone model), hypertensive rats, and diabetic rats are all planned or underway.

Conclusions

In 1998, a committee of the NRC published the first of a 4 volume report entitled: "Research Priorities for Airborne Particulate Matter" identifying the 10 highest priority targets for PM research (NRC 1998). Within the PM Centers' research portfolio, the priority areas have been addressed. A subsequent NRC report (2001) emphasized that these research priorities require multidisciplinary approaches. Recognizing that progress in understanding the health effects consequent to air pollution exposure requires talents from highly divergent fields, we believe that

the PM Centers effectively promote interdisciplinary cross-fertilization. The next 5 years of this program will bring the experience and results of the first Centers to fruition in new, focused studies and will hope to be instrumental in addressing the difficult scientific and public health policy problems that arise from ubiquitous particulate air pollution.

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Parameters		Particle mode	
Size	<u>Coarse (PM₁₀₎</u> 2.5 - 10μm	<u>Fine (PM₁₀₎</u> 2.5 -0.15μm	<u>Ultrafine (PM₁₀₎</u> < 0.15µm
Organic carbon content	+	++	+++
Elemental carbon content	+	++	+++
Metals as % of total elements	+++	++	+
PAH content	+	+	+++
Redox activity (DTT assay)	+	++	+++
HO-1 induction	+	++	+++
GSH depletion	+	+++	+++
Mitochondrial damage	None	Some	Extensive

Table 1: Contrasting features of coarse, fine and ultrafine particles (Li et al. 2003a)

Figure and Table Legends

Figure 1. Figure modified from Frampton et al. 2008. Mechanistic pathways for PM cardiovascular effects. ET, endothelin. MI, myocardial infarction. NO, nitric oxide. PM, particulate matter. TF, tissue factor.

Figure 2. Figure taken from Ghelfi et al. 2008. Capsazepine (CPZ) aerosolization prevents oxidative stress and damage in the heart of rats exposed to Concentrated ambient particles (CAPs). Adult Sprague Dawley rats received aerosols containing either 500 μ M CPZ or saline for 20 min immediately prior to exposure to CAPs. Values represent the mean of 8 independent determinations ± SEM. * p <0.05. (Ghelfi et al. 2008)

Figure 3. Figure taken from Araujo et al. 2008. UFP is the most proatherogenic fraction. Atherosclerotic lesions were quantitatively analyzed in serial aortic root sections and stained with oil red O. Lesional area was scored as square micrometer per section, averaged ≥ 25 sections per animal. Group averages are indicated by straight horizontal bars. One FA mouse was an obvious outlier in its group and removed from the atherosclerotic lesion analysis. However, its inclusion did not modify the overall significance. FA mice are represented by dotted circles (n+14), FPs by stripped circles (n=16), and UFPs by filled circles (n=15).

Figure 4. Figure taken from Oberdörster et al. 2004. Time course of ¹³C tissue concentrations in lung, olfactory bulb, cerebrum, and cerebellum of rats following a 6-h inhalation exposure to ultrafine (36 nm CMD) elemental ¹³C particles (n = 3 rats per time point): red= lung; blue =

olfactory; green = cerebellum; purple = cerebrum. Asterisks indicate values significantly greater than controls, p < 0.05 (ANOVA).







