

Comparison of Water-Quality Samples Collected by Siphon Samplers and Automatic Samplers in Wisconsin

Introduction

In small streams, flow and water-quality concentrations often change quickly in response to meteorological events. Hydrologists, field technicians, or locally hired stream observers involved in water-data collection are often unable to reach streams quickly enough to observe or measure these rapid changes. Therefore, in hydrologic studies designed to describe changes in water quality, a combination of manual and automated sampling methods have commonly been used—manual methods when flow is relatively stable and automated methods when flow is rapidly changing. Automated sampling, which makes use of equipment programmed to collect samples in response to changes in stage and flow of a stream, has been shown to be an effective method of sampling to describe the rapid changes in water quality (Graczyk and others, 1993). Because of the high cost of automated sampling, however, especially for studies examining a large number of sites, alternative methods have been considered for collecting samples during rapidly changing stream conditions. One such method employs the siphon sampler (fig. 1), also referred to as the “single-stage sampler”. Siphon samplers are inexpensive to build (about \$25–\$50 per sampler), operate, and maintain, so they are cost effective to use at a large number of sites. Their ability to collect samples representing the average quality of water passing through the entire cross section of a stream, however, has not been fully demonstrated for many types of stream sites.

The Inter-Agency Committee on Water Resources, Subcommittee on Sedimentation (ICWR-SS) developed and tested siphon samplers under laboratory conditions and concluded that siphon samplers are able to collect a sample representative of near-surface water quality during rising stages. ICWR-SS (1961) developed several models of samplers to collect representative samples for distinct ranges of stream velocity, water-surface surge, water temperature, and sediment size. The study concluded that siphon samplers are useful when sediment concentrations near the water surface are of value and sampling by other, possibly more accurate methods is not practical or feasible. Edwards and Glysson (1988) outlined some of the limitations of siphon samplers. The primary limitation was that, because samples are collected near the water surface at one point in the stream, adjustments may be needed to describe the vertical and horizontal distributions in water quality, especially if the stream transports large sand-size particles. This is also a limitation for automatic samplers because automatic sam-

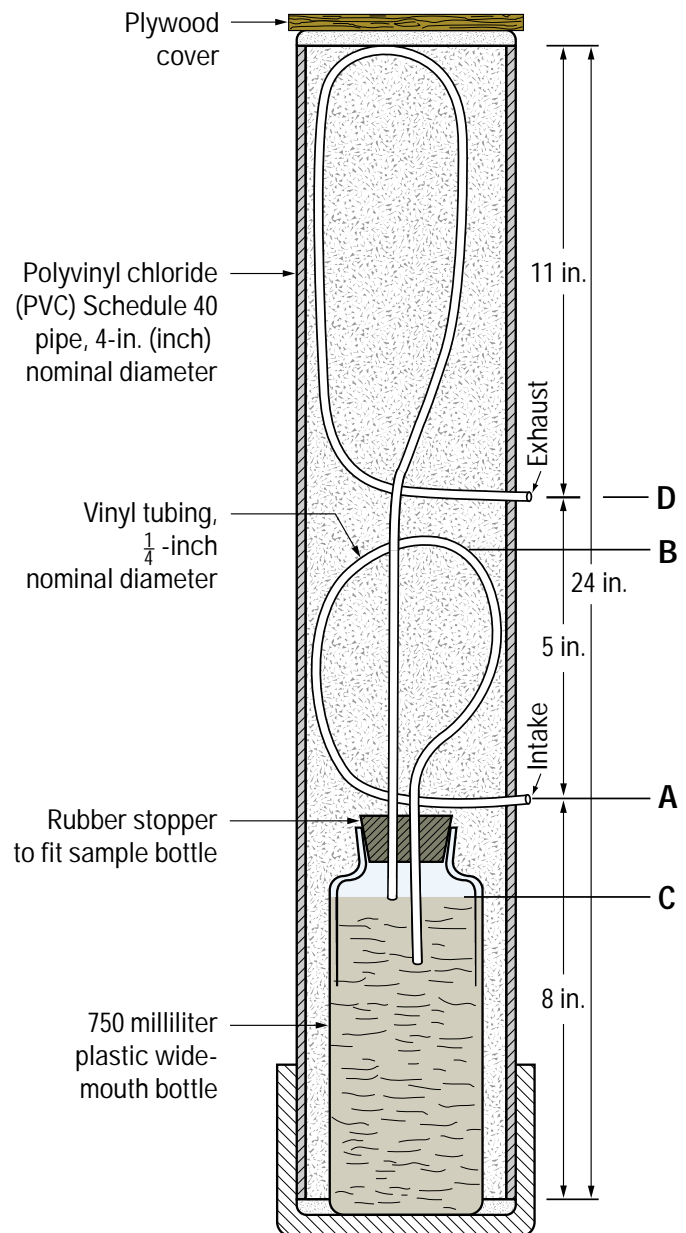


Figure 1. Typical siphon sampler.

plers collect a sample at a fixed horizontal and vertical location in the stream cross section.

The U.S. Geological Survey (USGS) and Wisconsin Department of Natural Resources (WDNR) are currently monitoring the water quality of several streams with a combination of manual and automated sampling methods. Future studies are aimed at describing changes in water quality at many sites; therefore, siphon samplers are being considered as a means to augment manual sampling and minimize sampling costs. Siphon samplers

have had limited use in Wisconsin but have not been thoroughly tested to determine their ability to collect representative samples in Wisconsin streams. This fact sheet describes how successfully siphon samplers can be used to collect representative samples at selected stream sites in Wisconsin. Concentrations of suspended sediment, total phosphorus, and ammonia nitrogen in samples collected by siphon samplers in three streams in southwestern Wisconsin are compared with those collected with the more thoroughly investigated stage-change-activated automated samplers (Krug and Goddard, 1986).

Siphon-Sampler Design and Operation

The design of the siphon sampler used in this study is similar to that described by the ICWR-SS (1961) and by Edwards and Glysson (1988) and shown in figure 1. The operation of a siphon sampler during an event with increased stage and flow is simple. As the stream stage rises to the elevation of the intake level A (fig. 1), water enters the 1/4-inch-diameter plastic tube. As the stream continues to rise, water continues to move up the intake tube until the stream

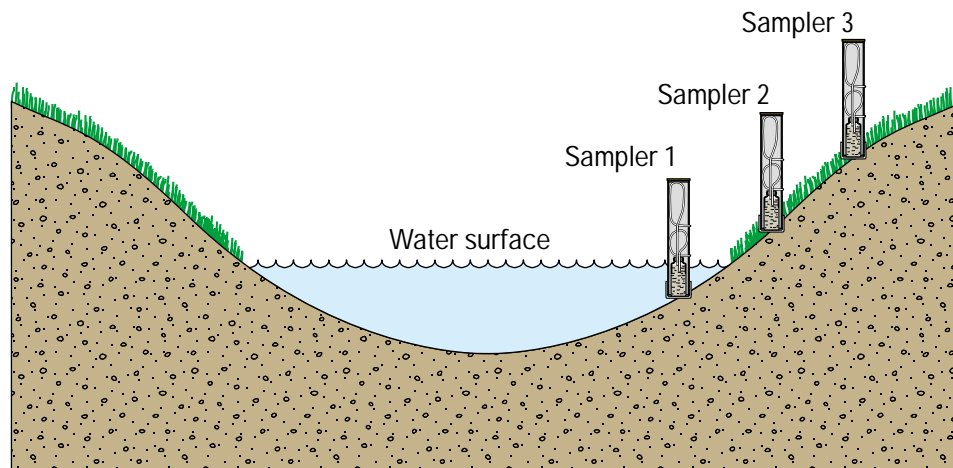


Figure 2. Typical stream-site installation of siphon samplers.

and the water in the tube reach level B; when the water levels rise past level B, a siphon is created and the sample bottle starts to fill. The sample bottle fills rapidly because the flow rate is driven by the hydraulic head, which is approximately the height difference between the stream stage (level B) and the discharge end of the intake tube (level C). As the water level in the sample bottle reaches the bottom end of the exhaust port (level C), filling is substantially completed; however, a small amount of additional water, equal to the water volume in the exhaust tube between levels C and D, enters the bottle after the water level rises past level C. After the stream stage reaches level D, an airlock is established in the loop of the exhaust tube, which precludes further filling of the bottle. Changes in the water level after this point do not significantly affect the contents in the bottle. After the event, the bottles are collected and the contents analyzed. Siphon samplers are unrefrigerated; therefore, analytical results may have to be qualified for certain constituents that are unstable at temperatures above about 4° Celsius. Several samplers can be installed at different levels at each site to collect samples throughout the anticipated range in water levels (fig 2).

Table 1. Comparison of water-quality data from siphon samplers with those from automated (ISCO) samplers. [Statistics are based on 47 paired samples for total phosphorus and ammonia nitrogen and 41 paired samples for suspended sediment]

Constituent and method	Concentration, in milligrams per liter					Difference in concentration (Siphon – ISCO), in milligrams per liter (and percent)				
	Minimum	Maximum	Standard deviation	Mean	Median	Maximum negative difference	Maximum positive difference	Standard deviation	Mean difference	Median difference
Total phosphorus										
Siphon sampler	0.10	4.16	0.73	0.60	0.41	-2.16	1.08	0.43	-0.05	-0.03
ISCO sampler	0.13	3.58	0.73	0.65	0.41				(-23%)	(7%)
Ammonia nitrogen										
Siphon sampler	0.02	4.15	0.67	0.39	0.22	-2.09	1.49	0.43	-0.07	-0.02
ISCO sampler	0.05	3.25	0.68	0.46	0.22				(-7%)	(-8%)
Suspended sediment										
Siphon sampler	12	526	193	155	75	-291	287	133	14	5
ISCO sampler	23	449	120	141	95				(-41%)	(4%)

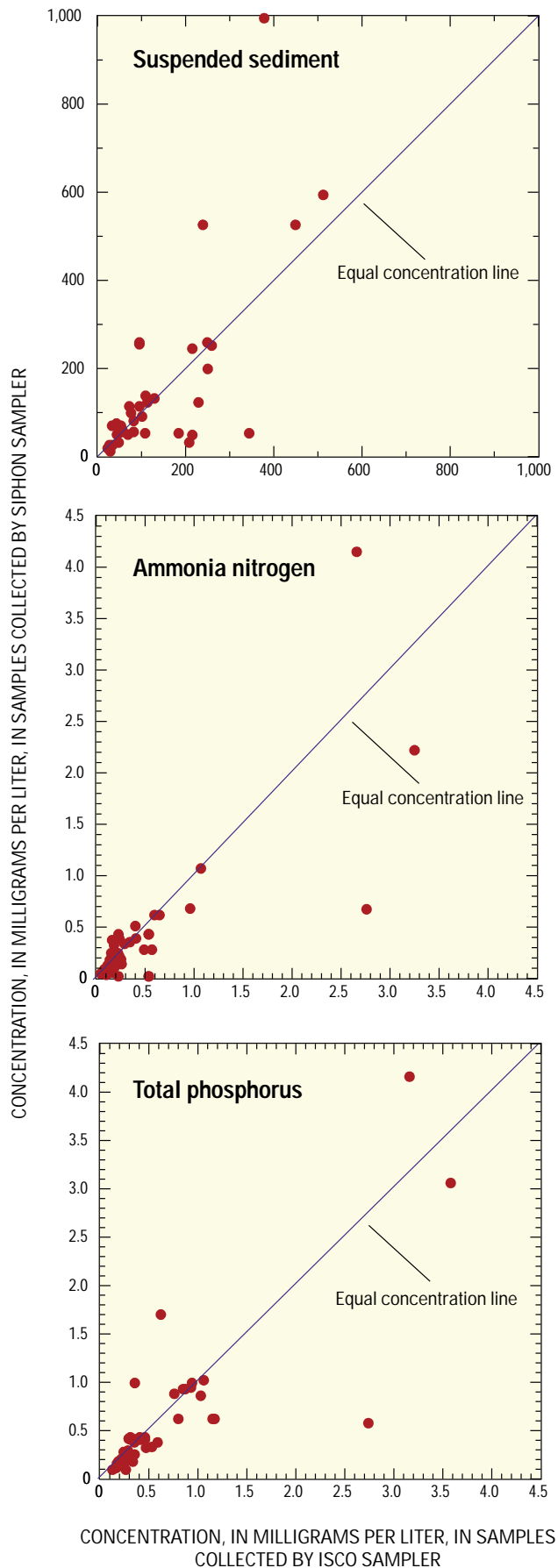


Figure 3. Concentrations of suspended sediment and chemical constituents in samples collected by the siphon sampler and ISCO sampler.

Sampling Sites

Siphon samplers were installed at three sites near USGS offices in Middleton, Wis.: North Fork of Pheasant Branch Creek (North Fork) and Pheasant Branch Creek at Highway 12 (Pheasant Branch), which are perennial streams; and South Fork of Pheasant Branch Creek (South Fork), an ephemeral stream. A USGS streamflow-gaging station was operational at each site, along with an automated water-quality sampler (ISCO) programmed to collect samples during runoff events. The drainage area above the North Fork site, 9.8 mi² (square miles), is primarily agricultural, whereas the drainage area above the South Fork site (5.7 mi²) is predominantly urban. The drainage area above Pheasant Branch (18.3 mi²), which is downstream from both the North and South Forks, encompasses both agriculture and urban development. Historical data from samples collected at Pheasant Branch indicate that most suspended-sediment particles were silt- and clay- sized (particles < 0.062 millimeters).

At each site, siphon samplers were installed at three distinct elevations to sample different stages and times during an episode of increasing streamflow (fig. 2). The samplers were placed so that the first sample would be collected when the water stage rose approximately 0.2 feet. The second sample would be collected approximately 0.3–0.5 feet above where the first sample was collected, and the third sample would be collected 0.4–0.6 feet above where the second was collected. A fence post was driven into the stream bottom, and each siphon sampler was attached to the post by a large hose clamp. The intake nozzles of the samplers were oriented perpendicular to the direction of streamflow to minimize the likelihood of the nozzles being clogged with sediment or debris. The automated ISCO samplers at each site were programmed to collect discrete samples at the stages and times similar to those for the siphon samplers.

Samples were removed from both the automated refrigerated samplers and siphon samplers as soon as possible after each runoff event and preserved by either chilling (for suspended-sediment analyses) or chilling and acidifying (for total phosphorus and ammonia nitrogen analyses). After sample removal, each sampler was cleaned by flushing the intakes with streamwater and distilled water. Suspended-sediment analyses were done by the USGS sediment laboratory in Iowa City, Iowa, and total phosphorus and ammonia nitrogen analyses were done by the Wisconsin State Laboratory of Hygiene in Madison, Wis. All samples were analyzed by use of standard methods (American Public Health Association, 1995; Guy, 1969).

Water-Quality Comparison

Pairs of samples (ISCO and siphon) were collected from the three sites over a range of flows and water-quality conditions and were aggregated into one data set for the statistical analyses. Forty-seven pairs of samples were analyzed for total phosphorus and ammonia nitrogen and 41 pairs for suspended sediment. As is evident from table 1, constituent concentrations in the paired samples were similar, but the ranges in values were slightly smaller in the

samples collected with the ISCO samplers than in those collected with the siphon samplers.

No systematic biases are evident in the distribution of data points about the 1:1 line (the line of equal concentrations) in the graphs shown in figure 3. The mean concentrations of the total phosphorus and ammonia nitrogen in the 47 sample pairs were within 0.07 mg/L (milligrams per liter) of each other, although the mean percentage difference for total phosphorus was almost 23 percent. The mean concentrations of suspended sediment were within 17 mg/L of each other, with a mean percentage difference of 41 percent. Differences between medians were even smaller. The median total phosphorus and ammonia nitrogen concentrations were identical, with the median percentage difference about 8 percent. The median suspended-sediment concentrations were within 5 mg/L of each other, and the median difference was 4 percent.

A nonparametric Wilcoxon signed-rank test (Conover, 1980) applied to the data indicated no statistically significant differences in the constituent concentrations between the samples collected by the two types of samplers. The null hypotheses of the tests performed were that there were no differences between the constituent concentrations using either sampler. At the 5-percent significance level ($P < 0.05$), there were no statistically significant differences found in concentrations between the sampling methods for any of the constituents. Therefore the null hypotheses were not rejected.

In general, the constituent concentrations of samples collected with automated samplers (ISCO) have been shown to be similar to those of manually collected, cross-sectionally integrated water-quality samples (Krug and Goddard, 1986). Therefore, the similarity found in the means and medians for each of the three water-quality constituents indicates that siphon samplers also collect representative water samples over the range of sampled flow conditions for the type of streams examined. It follows that samples collected with siphon samplers typically should have about the same accuracy (bias) as automated samplers; however, individual measurements may be less precise (as seen in the variance around the 1:1 lines in fig. 3). Part of the scatter around the 1:1 line of equal concentrations may have resulted from the pair of samples not being collected exactly at the same time and, therefore, may have been samples of water of different concentrations. Additional work is needed to determine if this variability between data sets is caused by sampler performance or by slight differences in sample-collection timing.

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Conclusions

Siphon samplers are low-cost alternatives to automatic samplers that have been traditionally used to collect representative water-quality samples. Siphon samplers can be used to augment manual sampling of “flashy” streams and remote streams by collecting samples during rapidly increasing stream stage—a generally impractical condition to be sampled adequately with a manual sampling program. Siphon samplers would also be a cost-effective alternative to automatic samplers if samples need to be collected at numerous sites. Siphon samplers do not collect water samples when the stream stage is decreasing; therefore, manual samples still need to be collected during this period. Decreases in stage, however, are generally more protracted than increases in stage and commonly can be manually sampled by a field person dispatched at the beginning of the event. Additional studies may help to determine the reason for the variability between individual constituent concentrations of samples collected with an automated sampler and the siphon sampler as demonstrated by the scatter around the 1:1 lines in figure 3.

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