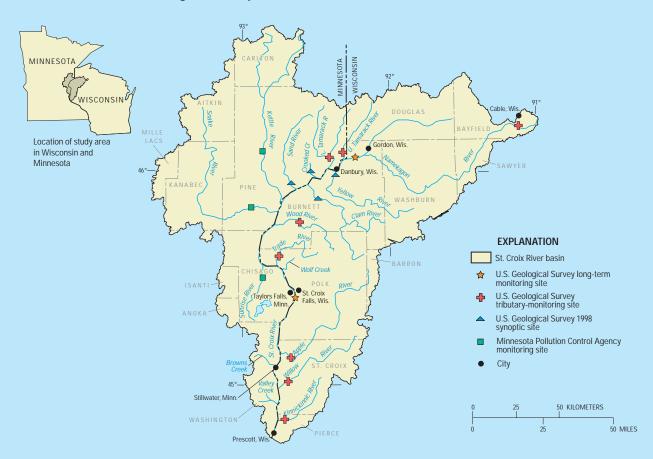
Nutrient and Suspended-Sediment Concentrations and Loads, and Benthic-Invertebrate Data for Tributaries to the St. Croix River, Wisconsin and Minnesota, 1997–99

Water-Resources Investigations Report 01-4162



Prepared in cooperation with the National Park Service, Minnesota-Wisconsin Boundary Area Commission, St. Croix County, and Wisconsin Department of Natural Resources



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By Bernard N. Lenz, Dale M. Robertson, James D. Fallon, and Randy Ferrin

U.S. GEOLOGICAL SURVEY
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Middleton, Wisconsin 2003



U.S. DEPARTMENT OF THE INTERIOR Gale A. Norton, Secretary

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CONVERSION FACTORS AND ABBREVIATED WATER-QUALITY UNITS

Multiply	Ву	To Obtain
millimeter (mm)	0.03937	inch
centimeter (cm)	.3937	inch
meter (m)	3.281	foot
kilometer (km)	.6214	mile
square kilometer (km ²)	.3861	square mile
cubic meter per second (m ³ /s)	35.3107	cubic foot per second
kilogram (kg)	2.2045	pound
liter (L)	.262	gallon
milligram (mg)	.000002205	pound
liters per second (L/s)	.0353	cubic foot per second
cubic meter per day (m ³ /d)	.1834	gallon per minute
centimeter per year (cm/yr)	.3937	inch per year

Temperature in degrees Celsius (°C) can be converted to degrees Fahrenheit (°F) by use of the following equation: $^{\circ}F = 1.8$ (°C) + 32.

Water year: water year is defined as the period beginning October 1 and ending September 30, designated by the calendar year in which it ends.

Abbreviated water-quality units used in this report: Chemical concentrations and water temperature are given in metric units. Chemical concentration is given in milligrams per liter (mg/L). Milligrams per liter is a unit expressing the concentration of chemical constituents in solution as weight (milligrams) of solute per unit volume (liter) of water. One thousand micrograms per liter is equivalent to one milligram per liter. For concentrations less than 7,000 mg/L, the numerical value is the same as for concentrations in parts per million. Specific conductance of water is expressed in mircosiemens per centimeter at 25°Celsius (μ S/cm). This unit is equivalent to micromhos per centimeter at 25°Celsius (μ mho/cm), formerly used by the U.S. Geological Survey.

Other Abbreviations Used in this Report:

GIS	Geographic Information System
MDNR	Minnesota Department of Natural Resources
MPCA	Minnesota Pollution Control Agency
NAWQA	National Water-Quality Assessment Program
NPS	National Park Service
NWQL	National Water-Quality Laboratory, U.S. Geological Survey
UMIS	Upper Mississippi River Basin
USGS	U.S. Geological Survey
WDNR	Wisconsin Department of Natural Resources

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Nutrient and Suspended-Sediment Concentrations and Loads, and Benthic-Invertebrate Data for Tributaries to the St. Croix River, Wisconsin and Minnesota, 1997–99

By Bernard N. Lenz, Dale M. Robertson, James D. Fallon¹, and Randy Ferrin²

Abstract

Nutrient and suspended-sediment data were collected on major tributaries to the St. Croix River during 1997–99 as part of three studies. The first study, done in 1997 as part of the U.S. Geological Survey's National Water-Quality Assessment Program Upper Mississippi Study Unit, was a widespread synoptic survey of nutrient and suspendedsediment concentrations, loads, and yields during snowmelt. Runoff from snowmelt in agricultural areas and other areas with low permeability soils had significantly greater nutrient concentrations than forested areas, whereas differences in suspended-sediment loading were not detected. In 1998, synoptic samplings of 11 tributaries were done during snowmelt, base-flow, and storm-runoff periods. These studies showed that the Apple, Willow, and Kinnickinnic Rivers were major contributors of suspended sediments and nutrients to the St. Croix River during base flow and storm-runoff events. Nitrate concentrations were high during base flow in the agricultural tributaries—specifically, the Kinnickinnic (4.83 mg/L), Willow (1.53 mg/L), and Apple (0.79 mg/L) Rivers—possibly from ground-water recharge or point-source contributions. Extensive water-quality sampling was done monthly and during high-flow events in water year 1999 (October 1, 1998 to September 30, 1999) in coordination with continuous streamflow monitoring at 12 sites in the St. Croix River Basin. These data were used to compute annual nutrient and suspended-sediment loads and yields at the monitored sites for water year 1999. Relations among environmental characteristics and calculated annual nutrient and suspended-sediment yields were used to estimate loading from unmonitored parts of the basin. The environmental characteristics found to best estimate annual yields were soil characteristics (clay, permeability of soil, and erodibility), basin slope and area, and the percentages of wetland and urban areas in the basins. Variability in 1999 rainfall intensity resulted in annual yields from several northern, forested basins being higher than those from the southern, agricultural basins. The Sunrise River had the highest annual suspended-sediment and nutrient yields in the basin in 1999. Concentrations and instantaneous loading rates varied as much among various flow conditions at individual sites as among sites during the three years of study.

Benthic invertebrates were sampled and indices of water quality were calculated at 16 tributaries in fall 1999. Benthic invertebrate indices indicated excellent to good water quality at all tributaries except Valley Creek, Willow River, and Kettle River. No relations were found between benthic invertebrate indices and the calculated and estimated 1999 annual tributary loads and yields.

INTRODUCTION

The St. Croix National Scenic Riverway was established in 1968 as one of the original eight components of the National Wild and Scenic Rivers Act (National Park Service, 1995a). The Riverway includes the St. Croix River from Gordon Dam near Gordon, Wis. to the confluence with the Mississippi River at Prescott, Wis., a distance of 247 km. The upper 40 km of the Riverway is solely within Wisconsin, whereas the remaining reaches of the river form the boundary between Wisconsin and Minnesota. The Riverway also includes the entire 158-km-long Namekagon River from Namekagon Dam near Cable, Wis., to the confluence with the St. Croix River. The part of the St. Croix River south of Taylors Falls, Minn., was added to the system in 1972. The St. Croix River Basin drains 20.010 km² of Minnesota and Wisconsin. The basin includes over 15 major tributaries and hundreds of minor tributaries to the St. Croix River (fig. 1).

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²National Park Service

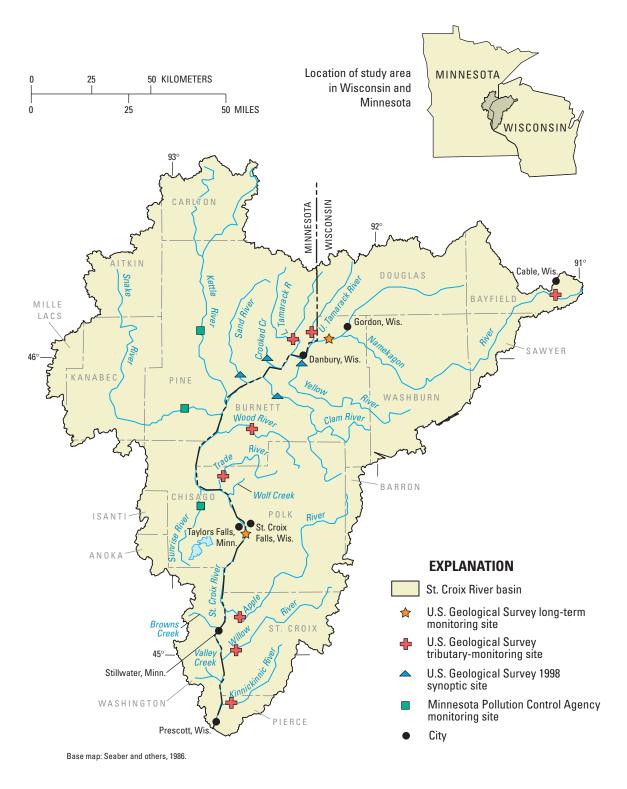


Figure 1. St. Croix River Basin sampling locations, Wisconsin and Minnesota.

The Riverway, from its headwaters to Stillwater, Minn., is administered by the National Park Service (NPS). The NPS owns a narrow corridor, approximately 0.4 to 0.8 km wide, along the parts of the Riverway it administers. The NPS also has made extensive use of scenic easements for properties near the rivers that remain in private ownership. Less than 1 percent of the river basin is in NPS jurisdiction. The lower reach of the river from Stillwater to the confluence with the Mississippi River is administered cooperatively by a commission with representatives from the Minnesota Department of Natural Resources (MDNR), the Wisconsin Department of Natural Resources (WDNR), and the NPS. This lower reach is more lacustrine than riverine and is called Lake St. Croix.

The Riverway contains more than 60 state and federally listed endangered and threatened species, indicating that it provides one of the few remaining relatively well preserved and biologically diverse aquatic environments in the Upper Midwest. Among the 40 species of freshwater mussels found in the Riverway, 15 are on state endangered species lists and two are on the federal list. Seven of the nearly 90 species of fish found within the Riverway are on state threatened or endangered species lists. More than 200 species of aquatic macroinvertebrates have been identified in the Riverway.

Because the Riverway is within a one hour drive of the Minneapolis/St. Paul metropolitan area, which has a rapidly expanding population that already is over 2 million, the Riverway continues to experience increased use and developmental pressure. Recreational use on the river has doubled since 1973 and has now reached nearly 1 million visitors per year (National Park Service, 1995a,b). Most recreational use in the park involves contact with the water, such as boating and canoeing, accentuating the need for good water quality.

The Riverway is protected effectively by its status not only as a National Wild and Scenic River but also as a unit of the National Park System; however, its tributaries do not have that same status. To protect the Riverway, managing agencies understand the need to protect these tributaries. In 1994, the NPS, WDNR, MDNR, Minnesota Pollution Control Agency (MPCA), and Minnesota-Wisconsin Boundary Area Commission signed a memorandum of understanding that formed the interagency St. Croix Basin Water Resources Planning Team (Basin Team) and agreed to investigate waterresources issues throughout the basin. Issues identified initially in 1994 included effects on the Riverway from

point and nonpoint sources of nutrients and suspended sediment, ground-water contamination, storm-water discharge, toxic contaminants, exotic species, impoundments, recreation and biomonitoring. Since 1994, the Basin Team has grown with the addition of other agencies including the U.S. Geological Survey (USGS), Metropolitan Council Environmental Services, University of Minnesota, Minnesota Department of Agriculture, St. Croix Band of Chippewa Indians, and U.S. Environmental Protection Agency. Additional issues including water-quality standards, ammonia toxicity, fish advisories, monitoring and agency coordination have been added to the list of issues to be addressed by the Basin Team.

One of the first accomplishments of the Basin Team was working with the Upper Mississippi River Basin Study Unit of the USGS National Water-Quality Assessment (NAWQA) program (Stark and others, 1996) to include the St. Croix River Basin as part of its study area. Three sites: Namekagon River near Leonards, Wis., St. Croix River near Danbury, Wis., and St. Croix River at St. Croix Falls were intensively monitored as NAWQA "fixed sites." Numerous other sites in the St. Croix River Basin were included as parts of various studies by the NAWQA program (Stark and others, 2000). Another accomplishment was the development of a Water Resources Management Plan for the St. Croix National Scenic Riverway (Holmberg and others, 1997). This plan, required by the NPS, has been instrumental in securing funding for water-quality studies. From this funding, the Basin Team implemented an interagency nutrient and suspended-sediment monitoring project with more than 25 sites monitored by six agencies. These studies included two tributary monitoring projects by the USGS in 1998 and 1999 and continued monitoring at the three NAWQA "fixed sites."

This report summarizes the results of the USGS nutrient and suspended-sediment studies done since 1997. The Basin Team will incorporate results from this report into future plans and studies. The purpose of this report is to summarize information on streamflow, concentrations, and loading rates of phosphorus, nitrogen, and suspended sediment in the major tributaries to the St. Croix River. Thirteen tributaries to the St. Croix River were sampled during 1997–99. Annual phosphorus and suspended-sediment loads and yields (load per unit area) were calculated and used to determine which tributaries were the largest contributors. Relations between environmental characteristics (land use, land cover, slope, permeability, soils) and annual nutrient

and suspended-sediment yields at monitored sites were determined and used to estimate loading from unmonitored parts of the basin. Relations between benthic invertebrate populations and nutrients also are discussed.

APPROACH AND METHODS

The USGS Upper Mississippi River Basin (UMIS) study unit of the NAWQA Program sampled tributaries to the Upper Mississippi River, including many in the St. Croix River Basin, as part of a synoptic study to characterize nutrient and suspended-sediment concentrations, loads, and yields during snowmelt (Fallon and McNellis, 2000). Streams were sampled during March and April 1997, when a combination of melting snowpack and rainfall produced near-record streamflow. Snowmelt-runoff samples were collected from 42 stream sites. The sampled streams drained areas with various combinations of land use and surficial geology common in the UMIS NAWQA study unit. Streams were classified by the predominant land use (forested, agricultural, urban, or mixed) and the predominant type of soil and surficial geology deposit. Drainage areas of sampled streams ranged from 26 to 121,000 km². Samples were collected during increasing streamflows, when concentrations were expected to be highest. Ancillary data from 12 sites provided comparative concentrations from before and after streamflow increases, as well as data needed to estimate constituent loads delivered during the 1997 snowmelt period (Fallon and McNellis, 2000).

On the basis of results from the snowmelt synoptic study, a study was designed for 1998 to identify spatial and temporal variations in nutrients and suspended-sediment loading from tributaries identified as major contributors and from other tributaries with environmental characteristics thought to be related to high nutrient and suspended-sediment loading. Each site was sampled four times during the year: snowmelt, spring rainfall runoff, summer rainfall runoff, and summer base flow. Instantaneous yields (based on constituent concentration and corresponding one-time streamflow measurement) were compared to environmental characteristics of the sampled tributary basins to identify which characteristics were most strongly related to nutrient and suspended-sediment loading in the tributaries.

A follow-up study was then conducted from October 1998 to September 1999 on a subset of sites from the 1998 study. For each monitored site, continuous

streamflow was measured and water samples were collected monthly and during four runoff-producing events. Annual loads and yields of nutrients and suspended sediments were calculated for the monitored tributaries and compared to environmental characteristics in the basins. These relations were used to estimate loads and yields at unmonitored tributaries.

Benthic invertebrates were collected at 15 tributary sites in fall 1999 and used to calculate indices of water quality. The indices of water quality were compared to annual loads and yields calculated during this study to determine whether relations were present between nutrients and sediments transported by the tributaries and the benthic invertebrates at these sites.

Field and Laboratory Methods

Streamflow

Continuous streamflow was calculated at 15-minute intervals at the flow-monitoring stations. Continuous stage was recorded by use of pressure transducers and stilling wells. Flow measurements were made at varying stream stages throughout the year. Relations between stage and discharge were determined from these data and used to convert stage data into flow data. Flows were calculated for each 15-minute interval and were used to determine the daily mean streamflow at each station (Kennedy, 1983).

Nutrients and Suspended Sediment

Water samples were collected by USGS and MPCA personnel at 15 major tributaries to the St. Croix River in 1997–99. Data from two long-term USGS-NPS monitoring sites on the St. Croix River (near Danbury, Wis. and at St. Croix Falls, Wis.) also were included in the loading study. All sampling locations are shown in figure 1 and listed in table 1.

Snowmelt-runoff samples were collected in spring 1997 by USGS-UMIS NAWQA personnel. The snowmelt synoptic sampling sites (1997) are described by Fallon and McNellis (2000). In 1998, USGS and MPCA personnel collected four samples representing snowmelt runoff, spring-rainstorm runoff, summer-rainstorm runoff, and summer base flow. The tributaries in the northern part of the basin, however, did not receive a substantial summer storm runoff in 1998 because of the unusually dry summer. Instead, these tributaries were

Table 1. Selected sampled tributary streams to the St. Croix River, Wisconsin and Minnesota [km², square kilometers; --, no U.S. Geological Survey (USGS) station number]

Site name	USGS station number	Drainage area (km²)	Latitude	Longitude
Kinnickinnic River near River Falls, Wis.	05342000	449	44° 49' 51"	92° 43′ 59″
Valley Creek at Afton, Minn.		49	44° 54' 41"	92° 46′ 48″
Willow River at Burkhardt, Wis.	05341752	721	45° 01' 01"	92° 42' 23"
Browns Creek at Stillwater, Minn.	05341540	88	45° 04' 35"	92° 48′ 21″
Silver Creek near Stillwater, Minn.		20	45° 04' 33"	92° 48′ 30″
Apple River near Somerset, Wis.	05341500	1,420	45° 09' 27"	92° 42' 59"
Little Carnelian Lake outlet near Stillwater, Minn.		11	45° 06' 56"	92° 47′ 38″
Trade River near Trade River, Wis.	05340390	345	45° 35' 54"	92° 46′ 02″
Sunrise River at Sunrise, Minn.	05340050	439	45° 32' 49"	92° 51' 23"
Wood River near Grantsburg, Wis.	05338955	217	45° 47' 07"	92° 37' 52"
Snake River near Pine City, Minn.	05338500	2,525	45° 50' 30"	92° 56' 00"
Kettle River below Sandstone, Minn.	05336700	2,252	46° 06' 20"	92° 51' 50"
Lower Tamarack River near Markville, Minn.	05335151	470	46° 04' 49"	92° 23' 37"
Sand Creek near Hinckley, Minn.	05335900	284	45° 57' 08"	92° 40′ 04″
Crooked Creek near Hinckley, Minn.	05335170	244	46° 00' 42"	92° 31' 45"
Clam River near Webster, Wis.	05335500	926	45° 52' 52"	92° 29' 16"
Upper Tamarack River near Markville, Minn.	05333579	245	46° 05' 30"	92° 18' 32"
Yellow River near Danbury, Wis.	05335031	940	46° 00' 44"	92° 21' 25"
Namekagon River near Leonards, Wis.	05331833	333	46° 10' 17"	91° 19' 45"
St. Croix River near Danbury, Wis.	05333500	4,092	46° 04' 28"	92° 14' 50"
St. Croix River at St. Croix Falls, Wis.	05340500	16,160	45° 24' 25"	92° 38' 49"

sampled during base flow and again at extremely low flows late in the summer. All samples from 1997 and 1998 were analyzed for nutrients and suspended sediment. Streamflow was measured at each site concurrently with sample collection, and loads and yields were computed on the basis of these measurements. (These are referred to hereafter as "instantaneous" loads and yields.)

In 1999, USGS and MPCA monitored streamflow continuously and did routine sampling at selected sites in the St. Croix River Basin. Water-quality samples were collected at the monitoring sites monthly and during runoff-producing events (snowmelt, spring storm, and two summer storms) to determine nutrient and suspended-sediment concentrations at differing flow conditions and seasons. The selected tributary monitoring locations were at established stream gaging locations or as close as possible to the St. Croix River while yet enabling sampling and measurements during high water. This location generally was at the first road crossing upstream from the tributary mouth. Care was taken to stay below impoundments so as to measure flow and

constituent concentration actually reaching the St. Croix River. This was possible on all but the Willow River, which was sampled above Lake Mallalieu.

Water-quality sampling by USGS personnel followed USGS protocols (Shelton, 1994), and analyses were done at the USGS National Water-Quality Laboratory (NWQL) in Denver, Colo. MPCA personnel followed standard agency protocols (Minnesota Pollution Control Agency, 1993), and analyses were done at the Minnesota Department of Health Chemical Laboratory in St. Paul, Minn. The greatest difference in the sampling methods was that USGS collected samples by use of the equal-width-increment method, whereas MPCA collected grab samples.

Quality-assurance/quality-control testing was done in 1998 and 1999. Each year both agencies collected similar samples at three tributaries that varied in size and environmental characteristics. Each agency used their respective protocols and sent samples to their respective laboratories for analysis. No consistent differences in nutrient concentrations were found among agencies; however, suspended-sediment concentrations

measured by the USGS by use of the equal-width-increment method consistently were higher than those measured by other agencies by use of grab samples (appendix 1). No attempts were made to adjust the suspended-sediment concentrations measured at the Kettle, Snake, and Sunrise Rivers to account for these differences; therefore, the concentrations, loads, and yields of suspended sediment may be slightly negatively biased. Similar results have been observed in interagency monitor studies conducted by the Interagency Task Force on Monitoring (Kammerer and others, 1998).

Benthic Invertebrates

Benthic invertebrates were collected during base flow in fall 1999. The benthic invertebrates collected were in the stream for all or part of their lives, depending on their individual life cycles, during the period that concentrations were measured and loads were estimated. These relatively immobile organisms are subjected continually to changes in water quality. The number, type, and diversity of the organisms found in each tributary are affected by these water-quality changes. Because of this effect, water-quality indices calculated from these invertebrate populations provide an indication of the overall water quality in a stream and are a valuable supplement to discrete chemical analysis of water samples.

Benthic-invertebrate samples were collected from the habitat most likely to have the greatest variety and quantity of organisms, as inferred from substrate, water depth, flow, and canopy cover. Typically, suitable habitats were found at the two largest riffles closest to the chemical water-quality sampling location. Riffles are the part of the stream where the water flows swiftly over completely or partially submerged coarse substrate that produces surface agitation. Sample locations generally had an open canopy, gravel or cobble substrate, high flow velocities, and water depth sufficient to keep most of the benthic invertebrates submerged during low flows.

Invertebrate samples were collected according to NAWQA protocols (Cuffney and others, 1993a,b). Riffle samples were collected using a modified Slack kick sampler with a 425-µm mesh at 13 of the 15 tributaries sampled. A stream-reach sample consisted of a composite of six kick samples—three collected at the upstream riffle and three collected at the downstream riffle in the same reach—that was processed and identified as one sample. Distances between kick samples in the same rif-

fle rarely were greater than 5 m. The sample area for each kick sample was 0.5 m by 0.5 m. Stream reaches near the chemical water-quality sampling sites at the Wood and Kettle River lacked wadeable coarse-grained riffles; therefore, samples were collected at these sites using a D-frame net in vegetation and woody debris and by scraping and picking debris clumps and logs within the main flow.

Samples were processed in the field in accordance with NAWQA protocols, which include swirling and sieving (425-µm mesh) until sample volumes were less than 750 mL. NAWQA protocols do not include the collection of mussels and these were returned to the stream during sieving. Samples, preserved in the field in 70 percent ethanol, were drained within 3 days and refilled with 70 percent ethanol, and were stored until they could be shipped for identification. Enumeration and taxonomic identifications of benthic invertebrates were done at the University of Wisconsin-Stevens Point.

Methods for Determining Environmental Characteristics

The environmental characteristics examined for the study area include land-use/land-cover types, surficial-deposit types and thickness, slope, and soil characteristics. Methods for determining the characteristics of each basin are described below.

Land Use

Land-use/land-cover information (hereafter referred to as land use) for the study area was obtained from high-altitude aerial photographs collected by the USGS from the early 1970s to the early 1980s (Feagus and others, 1983) and manually interpreted on the basis of the land-use classification of Anderson and others (1976). In addition, land-use data were updated with urban population information obtained from the 1990 census (Hitt, 1994). Land-use maps were produced from the interpreted data and digitized into a geographic information system (GIS). Land-use types summarized for each basin represent Anderson's Level I categories (Anderson and others, 1976) and include urban, agriculture, forest, water, wetland (further subdivided into forested and nonforested categories), and barren land (fig. 2).

Surficial Deposits

Surficial-deposit classification and thickness information for the study area was compiled from a digital coverage of Quaternary sediments for the glaciated United States east of the Rocky Mountains (Soller and Packard, 1998). Quaternary-sediment classifications included six categories: coarse-grained, fine-grained, till, patchy, exposed bedrock, and organic-rich. The digital coverage also included an attribute describing ranges of depth to bedrock. Those ranges were generalized by assigning a single depth to bedrock equal to the average of the range values. The digital coverage was intersected with the basin outlines, and the percentage of each surficial sediment type and an area-weighted average depth to bedrock were calculated for each basin.

Soil Characteristics

Soil-characteristic information for the study area was compiled from USSOILS, a national digital coverage of the State Soil Geographic (STATSGO) (Schwarz and Alexander, 1995). Soil characteristics examined in this study included clay content (percentage of soil less than 2 µm in size), organic-matter content (percentage by weight), soil erodibility (K factor used in the universal soil loss equation), permeability rates (millimeters per hour) (fig. 3), and slope (percent). The USSOILS coverage was intersected with the drainage basin outlines, and each of the soil characteristics of interest was calculated as an area-weighted average.

Load Estimation

Monitored Basins

Annual loads (calculated by summing daily loads) and annual standard errors of the predictions of those loads (calculated using daily standard errors of the predictions) were estimated by a regression approach by use of the USGS computer program Estimator (Cohn and others, 1989). In this study, estimated daily loads (*L*) were calculated based on relations between constituent load (in kilograms) and two variables: streamflow (*Q*, in cubic meters per day) and time of the year (*T*, in radians). The general form of the model was

$$\ln(L) = a + b [\ln(Q) - c] + d [\ln(Q) - c]^{2} + e [\sin(T)] + f [\cos(T)],$$
 (1)

where a, b, c, d, e, and f are regression coefficients.

Values for the regression coefficients in equation 1 were calculated for each site and for each time period by the use of multiple regression analyses between daily loads (daily average streamflows multiplied by instantaneous measured concentrations, in milligrams per liter) and daily streamflows (Q) and time of the year (T). For each sampling strategy and for each time interval, only terms that were significant at P < 0.05 were included in the regression. Because a logarithmic transformation was used in equation 1, computed daily loads were adjusted to account for a retransformation bias by use of the minimum variance unbiased estimate procedure. (See Cohn and others (1989) for a complete discussion.)

Equation 1 was calibrated for each site using all available data from 1997 to 1999 in the regression analysis to determine a relation between flow and loads. This relation and the flow data from water year 1999 were used to calculate annual loads and yields for the monitored tributaries. The standard error of prediction was used to calculate the range of loads within a 95-percent confidence interval.

Unmonitored Basins

Loads and yields from unmonitored tributaries were estimated from the environmental characteristics of their respective basins and the relations found between annual yields and the environmental characteristics of the monitored basins. These relations were found by use of forward stepwise linear regressions. The regression equation for each water-quality constituent was limited to the first three environmental characteristic variables in the stepwise regression that were statistically significant at P < 0.05. The 95-percent confidence intervals for the predicted annual yields were computed based on the prediction of a single future observation (Draper and Smith, 1966). When the prediction error was greater than the best approximation of the annual yield, the minimum estimated annual yield was set to zero rather than a negative value.

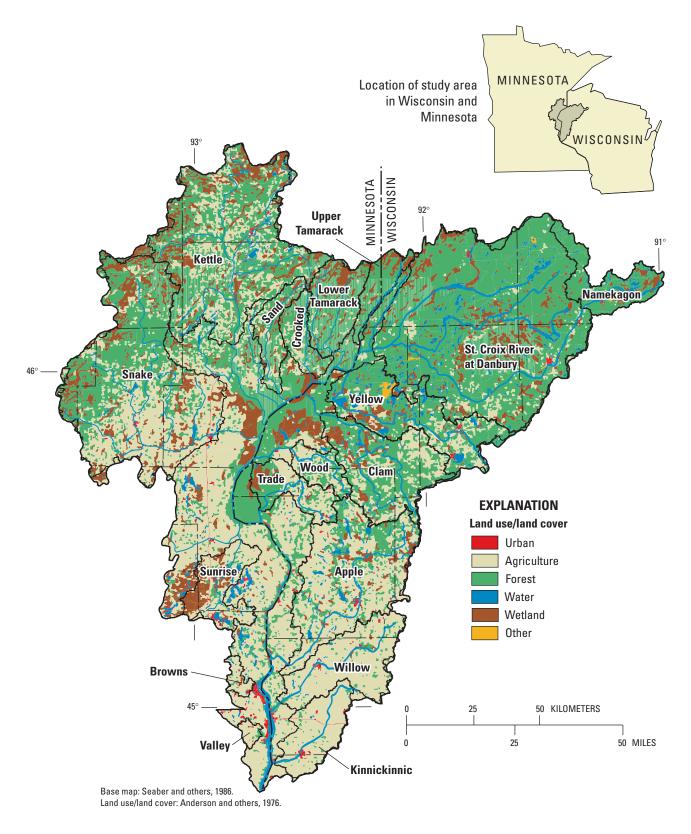


Figure 2. Land use/land cover of the St. Croix River Basin, Wisconsin and Minnesota.

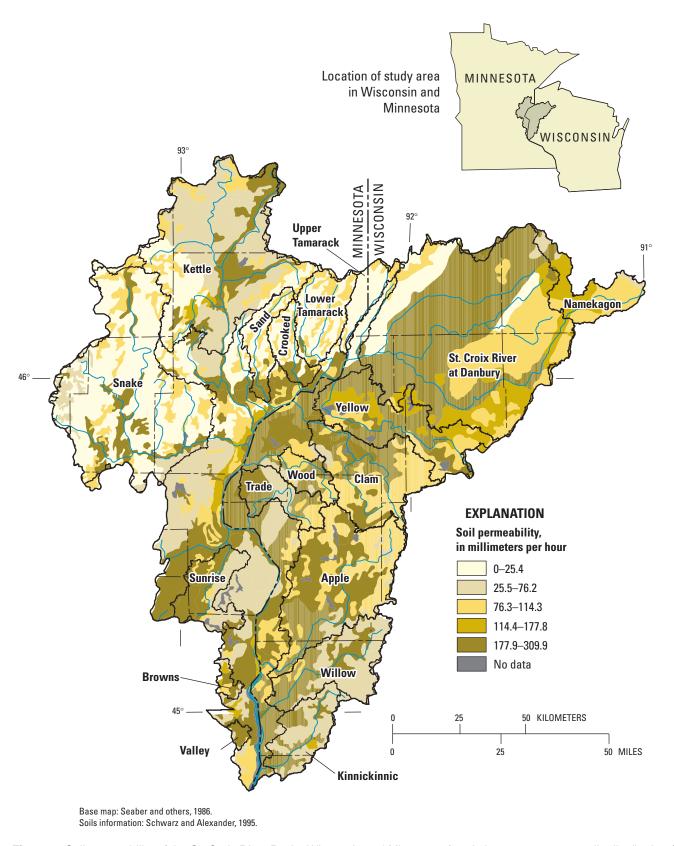


Figure 3. Soil permeability of the St. Croix River Basin, Wisconsin and Minnesota (gradations represent quantile distributions).

RESULTS OF NUTRIENT, SEDIMENT, AND BENTHIC-INVERTEBRATE SAMPLING

1997 Spring Snowmelt Synoptic Sampling

The USGS sampled snowmelt runoff from 42 stream sites in the Upper Mississippi River Basin, Minnesota and Wisconsin, during March and April 1997 as part of the UMIS-NAWQA synoptic study to characterize nutrient and suspended-sediment concentrations, loads, and yields. A detailed account of this study can be found in Fallon and McNellis (2000). A brief summary is included here.

Total phosphorus concentrations ranged from 0.02 to 0.57 mg/L, total nitrogen concentrations ranged from 0.52 to 7.7 mg/L, and suspended-sediment concentrations ranged from 6.2 to 249 mg/L. Concentrations at sites within the St. Croix River Basin were below median values for the entire Upper Mississippi Basin and generally in the lowest quartile, with the exception of nitrogen at Crooked Creek and suspended sediment at various Upper St. Croix River Basin tributary sites.

Streams in agricultural areas had significantly greater instantaneous yields (P < 0.05) of dissolved nitrite, nitrate, total nitrogen, orthophosphate, and total phosphorus than those in forested areas, and significantly greater instantaneous yields (P < 0.05) of dissolved nitrate and orthophosphate than all other land uses. In forested areas, instantaneous yields of suspended sediment and all nutrient forms were significantly greater (P < 0.05) for streams draining impermeable deposits than for those draining permeable deposits.

Additional water-quality data available for 12 of the sites provided data to estimate constituent loads delivered during the entire period of snowmelt and all of 1997. The snowmelt period contributed from 1 to 50 percent of 1997 annual loads of total nitrogen, total phosphorus, and suspended sediment at small stream sites, and 17 to 70 percent of annual loads at main-stem river sites. Agricultural streams transported the greatest proportions of their annual loads during snowmelt. Small urban streams transported the least proportions of their annual loads during snowmelt. Data collected at 11 sites in the St. Croix River Basin in spring 1997 as part of the UMIS-NAWQA study are listed in appendix 2. Additional data and a detailed account of this study can be found in Fallon and McNellis (2000).

1998 Synoptic Sampling

Eleven tributaries were sampled in 1998 to determine how concentrations vary during differing flow regimes: snowmelt, spring storm, summer storm, and typical summer base flow. A summary of the data collected by the USGS as part of this synoptic is presented in appendix 2. A summary of the data collected by the MPCA as part of this study is presented in appendix 3.

Climatological Factors

The winter of 1997–98 was nearly 6°C above the 30-year normal because of warm temperatures associated with a very strong El Niño event in the Pacific Ocean. Accumulated snow depth was below normal in the northern part of the St. Croix River Basin and moderately below normal in the southern part. A period of warm weather caused what snow had accumulated to melt in mid-February, and the resulting minimal snowmelt runoff was sampled. Snow did not accumulate again. A warm spring began with a large storm on March 30 and April 1, 1998, and spring storm samples were collected at all tributaries during this event. A summer storm occurred in the southern part of the basin on June 25, and samples were collected at the Apple, Willow, and Kinnickinnic Rivers during this event. No large summer runoff events occurred in the northern part of the basin in 1998. Consequently, the study was modified and extreme low-flow samples were collected from northern tributaries on September 22, 24, and October 1. Typical base-flow samples were collected on July 14, 22, and 28 throughout the basin following periods of a week to 10 days without rain.

Suspended Sediment

Suspended-sediment concentrations ranged from 2 to 424 mg/L and were related positively to flow for each stream. The forested northern tributaries generally had less suspended sediment than the southern agricultural tributaries, but a correlation with land use was not evident. The Apple, Yellow, Snake, and Clam Rivers (fig. 1) routinely had low suspended-sediment concentrations because they were sampled below impoundments. For the stream basins with large amounts of accumulated snow, highest suspended-sediment concentrations were collected during snowmelt runoff events, whereas for those that had only a small amount of snowmelt runoff, the highest suspended-sediment

concentrations were collected during the first spring rain event.

The highest suspended-sediment concentration was measured at the Kinnickinnic River during the spring storm (424 mg/L). The Willow River had a very high suspended-sediment concentration (414 mg/L) during the spring storm, probably because the spring runoff caused the highest discharge in the river since a dam upstream was removed in 1997. The exposed accumulated sediments from the former impoundment were transported easily by the river, as evidenced by the bank erosion in the former impoundment basin. Spring-storm runoff samples contained the highest suspended-sediment concentrations of the year at the Apple (13 mg/L), Clam (100 mg/L), and Upper Tamarack (15 mg/L) Rivers. The highest concentrations at the Sand (50 mg/L), Wood (35 mg/L), Trade (136 mg/L), Yellow (9 mg/L) and Lower Tamarack (26 mg/L) Rivers were found in snowmelt samples. Concentrations in base-flow and low-flow samples were less than 10 mg/L everywhere except for the Kinnickinnic (16 mg/L) and Clam Rivers (14 mg/L).

The highest instantaneous suspended-sediment daily loads and yields calculated from 1997–98 sampling data came from the agricultural Kinnickinnic (660,000 kg, 1,530 kg/km²) and Willow River Basins (780,000 kg, 1,100 kg/km²) as well as the heavily forested Crooked Creek Basin (170,000 kg, 700 kg/km²), and the mixed forest and agricultural Trade River Basin (177,000 kg, 510 kg/km²).

Phosphorus

Total phosphorus concentrations ranged from 1.32 mg/L during snowmelt at the Kinnickinnic River to less than the detection level (< 0.01 mg/L) during base flow at Upper Tamarack, Yellow, and Clam Rivers. Phosphorus concentrations were low in all streams. Only the Kinnickinnic, Willow, Trade, Sunrise, Snake, and Kettle Rivers had one or more samples with concentrations greater than 0.1 mg/L. Base-flow concentrations of phosphorus were all below 0.08 mg/L. Total phosphorus concentrations were highest during events in the tributaries from the southern part of the basin with predominantly agricultural land use.

The Kinnickinnic River had the highest total phosphorus concentrations of 1998 during snowmelt (1.32 mg/L) and a spring storm (0.38 mg/L). The next highest phosphorus concentrations (0.18 mg/L) were from snowmelt at the Trade River and a spring storm at

the Sunrise River. Large streams had the highest instantaneous daily phosphorus loads and included the Kinnickinnic (1,003 kg), Sunrise (444 kg), Kettle (398 kg), Apple (195 kg), Snake (188 kg), and Willow (184 kg) Rivers. Several small streams had high instantaneous daily yields, including the Wood (0.34 kg/km²), Crooked (0.25 kg/km²), and Upper Tamarack (0.21 kg/km²) Rivers.

Nitrogen

Total nitrogen concentrations ranged from 0.30 to 5.53 mg/L. Concentrations of total nitrogen always were less than 2 mg/L except during 1997 snowmelt runoff at the Kinnickinnic River (5.53 mg/L). Base-flow and event samples had relatively similar concentrations in all tributaries; thus, instantaneous daily loads and yields were directly related to flow volume—higher during periods of increased flow. The Kettle River, because of its large flow volume (1,800 cfs), had the highest instantaneous daily load of total nitrogen (4,670 kg). The Kinnickinnic River (9.7 kg/km²) and two smaller basins with high runoff (Upper Tamarack (4.1 kg/km²) and Wood (4.0 kg/km²) Rivers) had the highest instantaneous daily yields.

Dissolved nitrate concentrations were highest in agricultural tributaries in the southern part of the basin: the Kinnickinnic (4.8 mg/L), Willow (4.4 mg/L), and Apple (1.3 mg/L) Rivers. Concentrations were highest at low flows and decreased because of dilution during rainfall-runoff events, an indication that a point source or ground-water recharge or both were possible contributing factors to the amount nitrate in the streams; however, instantaneous loads and yields still were controlled by flow, and the highest instantaneous loads and yields were recorded during events. Nitrogen was not analyzed in samples collected at the Sunrise, Snake, and Kettle Rivers.

Major Contributing Tributaries

The agricultural land-use basins of the Kinnickinnic, Apple, and Willow Rivers, as well as the clay-rich Kettle River Basin, were the major contributors of suspended sediments and nutrients to the St. Croix River in 1998, not only during events, but also during base flow. The Trade, Wood, Snake, Sand, Crooked, and Lower Tamarack Rivers had greater event and seasonal variations than the other tributaries, but they all appear to be substantial contributors of sediments and nutrients to the St. Croix River during events. The Snake and Sunrise Rivers, though they did not have the highest instantaneous loading rates in 1998, nevertheless were important contributors of nutrients and sediment because of their large flow volumes and the increase in loading rates in these streams during event related increases in flow.

The only statistically significant relation between land use and instantaneous nutrient and sediment concentrations, loads, and yields found in the 1998 data was that the percentage of agricultural land use was related to nitrite plus nitrate concentrations during base flow and to dissolved phosphorus concentrations during the spring storm sampling. Other data showed weak relations to land use but the small data set size resulted in the relations being statistically insignificant.

1998-99 Intensive Tributary Monitoring

Ten tributaries (Wood, Trade, Apple, Willow, Kinnickinnic, Upper Tamarack, Lower Tamarack, Kettle, Snake, and Sunrise Rivers) were selected for continuous streamflow monitoring and intensive water-quality sampling. Water-quality data from this monitoring effort are listed in appendixes 3 and 4. Two main-stem sites on the St. Croix River (at St. Croix Falls and near Danbury, Wis.) also were sampled in 1998–99, and data for these sites are listed in appendixes 5 and 6. Summaries of the data used in loading computations are given in tables 2, 3, and 4.

Climatological Factors

Total precipitation for the period from October 1, 1998, to September 30, 1999, was above normal throughout the basin. During the fall and winter of 1998–99, precipitation and snowfall were normal to less than normal (National Oceanographic and Atmospheric Administration, 1998, 1999). Snowmelt in the spring of 1999 was gradual throughout the basin, resulting in minimal runoff events at most tributaries. Spring and summer rainfall in 1999 was above normal, and May and July were both much wetter than normal. In the northern forested part of the basin, several intense shortduration storms created large runoff events; however, in the more agricultural and mixed agricultural portion of the southern part of the basin, high-intensity storms were rare. Rainfall events in the southern part of the basin generally lasted many hours, sometimes days.

Runoff events in the agriculture-dominated southern part of the basin correspondingly were small. Samples were collected near the maximum, the minimum, and throughout the entire range of flows at all sites.

Suspended Sediment

Suspended-sediment concentrations measured during events were much lower in 1999 than in 1998. Suspended-sediment concentrations, some as low as 1 mg/L, were determined for samples collected under the ice at the Upper Tamarack River in December 1998. Concentrations less than 10 mg/L were typical during base flow across the basin, and even various samples collected during small events had concentrations below 10 mg/L. The highest suspended-sediment concentration measured in water year 1999 was 64 mg/L at the forested Upper Tamarack River (compared to the 424 mg/L measured at the Kinnickinnic River in 1998). The majority of the suspended sediment in base-flow samples was fine grained (less than 0.062 mm), whereas suspended sediment in runoff samples primarily was composed of particles coarser than 0.062 mm. The volume-weighted concentrations (total load divided by total flow) of suspended sediment were highest in the Sunrise River (24.8 mg/L) and Trade River (24.6 mg/L).

Phosphorus

Concentrations of total phosphorus were highest during the small events at tributaries in the southern, agricultural part of the basin and all year at the Sunrise River, which drains a mixed agricultural, urban, and forested basin. Concentrations in 9 of the 12 samples from the Sunrise River were greater than 0.1 mg/L, even during base flow. The average phosphorus concentration at the Sunrise was 0.12 mg/L. Only 16 samples from all 9 other tributaries had concentrations greater than 0.1 mg/L in water year 1999. Of those 16 samples, 5 were collected at the Kinnickinnic River, 4 at the Willow River, and 3 each from the Snake and Upper Tamarack Rivers.

The highest total phosphorus concentration (0.492 mg/L) collected in the basin in water year 1999 came from the Apple River on October 23 and was almost entirely in the particulate form. The next highest concentration (0.264 mg/L) came from the Willow River during a spring event, and nearly half was in the dissolved form. Generally, more particulate phosphorus

Table 2. Summary of selected water-quality characteristics analyzed by the U.S. Geological Survey, St. Croix River tributary synoptic sampling sites, 1997–99 [°C, degrees Celsius; µS/cm, microsiemens per centimeter; milligrams per liter; mm, millimeter; <, less than; %, percent]

	Upper	Upper Tamarack River near Markville. Minn.	River inn.	Lower	Lower Tamarack River near Markville. Minn.	liver inn.	JW G	Wood River near Grantsburg, Wis.	ar is.	Trac	Trade River near Trade River, Wis.	ar S
			Nimber			Nimber		ò	Nimber			Nimber
Constituent	Minimum- maximum	Median	of samples	Minimum- maximum	Median	of samples	Minimum- maximum	Median	of samples	Minimum- maximum	Median	of samples
Temperature, water (°C)	-0.1–23.1	10	. 61	-0.1–25.5	11.0	18	0–27.4	10.8	20	0.8 - 21.8	9.3	21
Specific conductance (µS/cm)	40–358	109	19	40–257	112	18	129–267	202	20	140-242	189	21
Oxygen, dissolved (mg/L)	5.0-18.2	10.5	18	5.6-15	6.7	17	5.4-13.8	8.6	19	5.1–14	10.8	20
pH, water, field (standard units)	6.4–7.9	7.4	19	6.3–7.7	7.25	18	7.3–7.9	7.6	19	7.1–8	7.7	21
Nitrogen, ammonia, dissolved (mg/L as N)	<.02192	.05	19	<.02-134	.054	19	<.02148	.072	20	<.0233	60:	21
Nitrogen, nitrite, dissolved (mg/L as N)	<.01014	<.01	19	<:01-013	<.01	19	<.01021	<.01	20	<.01022	<.01	21
Nitrogen, ammonia + organic, dissolved (mg/L as N)	.43–1.6	76.	19	.35–1.3	68.	19	.29–.78	.58	20	.29–.8	.47	21
Nitrogen, ammonia + organic, total (mg/L as N)	.38-3.1	1.26	19	.36–2.3	1.15	19	.28–1.6	.81	20	.05-1.5	89:	21
Nitrogen, nitrite plus nitrate, dissolved (mg/L as N)	<.05-2.7	.285	19	<.05518	.141	19	<.05317	.174	20	.10349	.297	21
Phosphorus, total (mg/L as P)	<.01124	.049	20	.019-093	.052	19	.018098	.061	20	01722	.061	21
Phosphorus, dissolved (mg/L as P)	<.01044	.023	19	.014051	.026	19	<.01046	.00	20	<.0103	.015	21
Phosphorus ortho, dissolved (mg/L as P)	<.01032	.013	19	<.0103	910.	19	<.01035	.015	20	<.01031	.015	21
Chloride, dissolved (mg/L as Cl)	.56–12	3.34	15	.05–6.6	1.95	15	3.3–6.6	5.2	14	3.6-13	5.1	14
Sediment, suspended (mg/L)	1–64	13.7	18	2–26	9.3	18	4–53	16.4	19	3-204	33	20
Sediment, suspended, % finer than 0.062 mm	22–96	69	18	67-100	85	18	21–93	29	19	31–96	74	19
	•	i -			ï		2	i -				
	or S	Apple River near Somerset, Wis.	, g	ā	Willow River near Burkhart, Wis.	₩ .		Amnickinnic River near River Falls, Wis.	r near S.			
		ŕ	Number		,	Number		•	Number			
Constituent	Minimum- maximum	Median	of samples	Minimum- maximum	Median	of	Minimum- maximum	Median	of			
Temperature, water (°C)	0-23	10.7	20	1.1–23.1	11.7	19	0.05-21.3	10.6	20			
Specific conductance (µS/cm)	173-430	270	19	171–427	353	19	228–496	431	20			
Oxygen, dissolved (mg/L)	6.8–16.7	10.8	19	8–15.2	10.8	18	7.5–8.8	8.1	20			
pH, water field (standard units)	7.6–8.5	8.0	20	7.6–9	8.2	19	7.5-8.8	8.1	20			
Nitrogen, ammonia, dissolved (mg/L as N)	<:02126	.048	20	<.02298	.087	19	<.02-2.27	.138	20			
Nitrogen, nitrite, dissolved (mg/L as N)	<.01052	.0121	20	.01053	.026	19	<.01051	.020	20			
Nitrogen, ammonia + organic, dissolved (mg/L as N)	.24–.69	.36	20	.18–1.2	.422	19	.05-5.3	.483	20			
Nitrogen, ammonia + organic, total (mg/L as N)	.21–.74	.46	20	.28–1.6	.70	19	.05-5.5	.662	20			
Nitrogen, nitrite plus nitrate, dissolved (mg/L as N)	.439–1.43	.835	20	.731–4.36	2.29	19	1.70-5.74	4.51	20			
Phosphorus, total (mg/L as P)	.024492	.071	25	.013–.264	680.	24	.046 - 1.32	.154	23			
Phosphorus, dissolved (mg/L as P)	<.01057	.025	20	<.0112	.038	19	.049-1.03	.120	20			
Phosphorus ortho, dissolved (mg/L as P)	<.01038	.0176	20	<.01073	.03	19	.03928	.103	20			
Chloride, dissolved (mg/L as Cl)	6.8 - 13	7.8	14	7.9–91	17.8	14	13–17	16.3	14			
Sediment, suspended (mg/L)	2–13	4.6	20	3-414	32	18	3-424	49	19			
Sediment, suspended, % finer than .062 mm	44–100	81	20	25–100	84	18	32–97	89	19			

Table 3. Summary of selected water-quality constituents at Minnesota Pollution Control Agency, St. Croix River tributary synoptic sampling sites, Minnesota,1998–99

[mg/L, milligrams per liter]

	Kettle	River, M	linn.	Snake	River, N	linn.	Sunris	e River, I	Minn.
Constituent	Minimum- maximum	Median	Number of samples	Minimum- maximum	Median	Number of samples	Minimum- maximum	Median	Number of samples
Nitogen, ammonia + organic, dissolved (mg/L)	0.36-1.29	0.85	15	0.62-1.47	0.94	15	0.63-1.45	1.03	15
Phosphorus, total (mg/L)	.0312	.06	15	.0415	.07	18	.0618	.12	15
Chloride, dissolved (mg/L)	2.8-8.9	3.9	10	3.5-10.0	5.4	10	11.0-17.0	13.5	10
Sediment, suspended (mg/L)	1.5-15.0	5.2	12	1.6-10.0	4.7	14	1.0-46.0	15.0	12

Table 4. Summary of selected water-quality constituents at long-term monitoring sites on the St. Croix River, Wisconsin and Minnesota, October 1998 to December 1999

[°C, degrees Celsius; μ S/cm, microsiemens per centimeter; μ g/L, micrograms per liter; mg/L, milligram per liter; mm, millimeter; <, less than; %, percent]

		oix River n nbury, Wis.			roix River oix Falls, W	
Constituent	Minimum– Maximum	Median	Number of samples	Minimum- Maximum	Median	Number of samples
Temperature, water (°C)	-0.1–25.3	10.3	19	-0.12-4.9	9.6	16
Specific conductance (µS/cm)	48-156	110	19	114-217	168	15
Oxygen, dissolved (mg/L)	4.8-14.2	9.4	16	6.3-26	10.5	16
pH, water, field (standard units)	6.6-7.9	7.4	19	7.1-8.4	7.6	16
Alkalinity, dissolved field (mg/L as CaCO ₃)	20-74	49.5	16	49–97	74	15
Bicarbonate, dissolved field (mg/L as HCO ₃)	25-90	61.2	17	60-119	91	15
Nitrogen, ammonia, dissolved (mg/L as N)	<.02065	.20	18	<.02094	.032	16
Nitrogen, nitrite, dissolved (mg/L as N)	<.01	<.01	18	<.01018	<.01	16
Nitrogen, ammonia + organic, dissolved (mg/L as N)	.15-1.5	.40	18	.2276	.43	16
Nitrogen, ammonia + organic, total (mg/L as N)	.2195	.45	19	.2995	.59	16
Nitrogen, nitrite plus nitrate, dissolved (mg/L as N)	<.05214	.090	19	.087445	.203	16
Phosphorus, total (mg/L as P)	<.008081	.031	19	.012081	.042	16
Phosphorus, dissolved (mg/L as P)	.004021	.011	18	<.004037	.020	16
Phosphorus, ortho, dissolved (mg/L as P)	<.01025	<.01	18	<.01257	.026	16
Carbon, organic, dissolved (mg/L as C)	2.9-20	9.1	19	3.2-20	9.5	16
Carbon, organic suspended, total (mg/L as C)	.2-2.2	.65	19	<.20-2.6	.7	15
Magnesium, dissolved (mg/L as Mg)	2.4-5.8	4.5	18	4.8-9.7	7.0	16
Sodium, dissolved (mg/L as Na)	1.2-2.8	2.4	18	2.5-4.2	3.5	16
Chloride, dissolved (mg/L as Cl)	<.1-3.9	2.40	19	2.4-5.5	4.3	16
Sulfate, dissolved (mg/L as SO ₄)	1.5-5.7	3.7	19	.95-6.3	3.9	16
Fluoride, dissolved (mg/L as F)	<.1	<.1	18	<.1	<.1	16
Silica, dissolved (mg/L as SiO ₂)	8–16	11.4	18	8.1-17	12.1	16
Iron, dissolved (μ g/L as Fe)	88-790	254	18	88-940	306	16
Manganese, dissolved (µg/L as Mn)	7.9–85	21.8	18	15–57	29	16
Solids, residue at 180°C, dissolved (mg/L)	71–104	90.7	19	96-140	121	16
Sediment, suspended (mg/L)	1–26	7.6	19	2-18	7.2	12
Sediment, suspended, % finer than 0.062 mm	40-100	76	19	50-100	89	12

¹⁴ Nutrient and Suspended-Sediment Concentrations and Loads and Benthic-Invertebrate Data for Tributaries to the St. Croix River, Wisconsin and Minnesota, 1997–99

was collected during events, whereas 50 to 100 percent of the phosphorus was in the dissolved form in baseflow samples. Volume-weighted concentrations of total phosphorus were greatest at the Sunrise River (0.118 mg/L) followed by the Kinnickinnic (0.096 mg/L), Willow (0.084 mg/L), Snake (0.083 mg/L), and Kettle (0.070 mg/L) Rivers.

Nitrogen

The highest total nitrogen concentration measured in 1999 was 3.10 mg/L at the Upper Tamarack River on June 10. Concentrations between 2 and 3 mg/L were collected at the Upper Tamarack River in April and May, and at the Lower Tamarack River in April, May, and June. Both basins have numerous bogs. All other samples had concentrations less than 2 mg/L. Most of the nitrogen was in the dissolved form, except during the events mentioned previously, when only one-half to one-third of the nitrogen was in the dissolved form. Total nitrogen concentrations demonstrated seasonality, the highest concentrations were collected in the spring and summer.

Although the highest total nitrogen concentrations were found at the northern forested sites in water year 1999, concentrations of the nitrite form of nitrogen were highest in tributaries in the agricultural southern part of the basin, including the Apple (0.052 mg/L), Willow (0.037 mg/L), and Kinnickinnic (0.027 mg/L) Rivers. Nitrogen concentrations varied little with flow in all tributaries so the volume-weighted concentrations show similar trends as the measured concentrations.

Comparison to 1998 Concentrations

Concentrations of nutrients and suspended sediments in 1999 generally were higher in the northern, forested tributaries and lower in the southern, agricultural and mixed agricultural tributaries when compared to those measured in 1998. The highest suspended-sediment concentration in water year 1999 (64 mg/L) was collected at the Upper Tamarack River in the northern-forested part of the basin during a June runoff event. The highest suspended-sediment concentration measured during any 1998 event at that site was 15 mg/L. Similar patterns in suspended-sediment and nutrient concentration were found for all northern sites; much higher in 1999 than in 1998.

Concentrations in the tributaries in the southern part of the basin demonstrated just the opposite pattern between years. There were no major events in the southern part of the basin in 1999, and concentrations were lower than those in measured 1998. The Kinnickinnic River in the southern part of the basin had the highest suspended-sediment concentration sampled in 1999, but was only 56 mg/L (compared to 424 mg/L in a sample collected during a large summer storm event in 1998).

The Upper Tamarack, Lower Tamarack, Kettle, and Snake Rivers in the northern, forested areas had the most variability in flow and concentration in 1999, whereas the Sunrise River had consistently high nutrient concentrations. Tributaries in the southern part of the basin (Kinnickinnic and Willow Rivers) in the agricultural areas showed highly variable flows and concentrations in 1998; however, because of minimal runoff in 1999, flows were much less variable and concentrations consistently were lower than in 1998. These data are good examples of the disparity among concentrations in the basin from the northern, forested areas to the southern, agricultural areas and from 1998 to 1999.

1999 Benthic-Invertebrate Sampling

Diverse populations of invertebrates were found throughout the basin. The five most common families of benthic invertebrates identified at each stream are given in appendix 7. The most commonly found families of benthic invertebrates in the basin were Hydropsychidae, Elmidae, Baetidae, and Chironomidae. Benthic-invertebrate taxonomic data were used to calculate the following water-quality indices: Biotic Index (Hilsenhoff, 1982, 1987); Family-level Biotic Index (Hilsenhoff, 1988); Mean Tolerance Value (Lillie and Schlesser, 1994); Shannon-Weaver Diversity Index (Shannon and Weaver, 1949); number of Ephemeroptera, Plecoptera, Trichoptera individuals and genera; number of species; and number of genera (table 5).

Benthic-invertebrate indices indicated no water-quality effects in all but three St. Croix tributaries. Valley Creek and Willow River have diminished invertebrate population health based on species abundance (16 and 12), genera abundance (13 and 12), and diversity measures (3.02 and 2.43). Biotic indices (7.43) results indicated that the Kettle River was affected by excessive nutrients. Although the number of individuals (31) was similar to many other tributaries, the diversity (2.18) at

Table 5. Benthic invertebrate water-quality measures calculated for selected tributaries to the St. Croix River, Wisconsin and Minnesota, September 1999 [EPT, Ephemeroptera, Plecoptera, and Trichoptera.]

Stream names	Hilsenhoff Biotic Index	Hilsenhoff Biotic Index water-quality class	Hilsenhoff Family- Level Biotic Index	Mean tolerance value	Shannon- Weaver Diversity Index	Number of EPT genera	Number of EPT individuals	Number of species	Number of genera
Apple River near Somerset, Wis.	3.90	Very Good	4.03	4.24	4.02	20	435	50	41
Browns Creek at Stillwater, Minn.	3.74	Very Good	3.99	4.00	3.28	8	118	22	19
Clam River near Webster, Wis.	5.54	Good	4.11	4.73	3.21	13	159	29	28
Crooked Creek near Hinckley, Minn.	3.19	Excellent	3.58	3.10	4.13	19	95	35	33
Kettle River below Sandstone, Minn.	7.43	Fairly Poor	6.64	5.96	2.18	10	38	31	28
Kinnickinnic River near River Falls, Wis.	3.42	Excellent	4.04	3.52	3.79	11	101	31	22
Lower Tamarack River near Markville, Minn.	3.33	Excellent	3.41	3.55	4.19	21	86	36	33
Sand Creek near Hinckley, Minn.	2.99	Excellent	3.48	2.88	4.36	21	119	41	36
Snake River near Pine City, Minn.	4.91	Good	4.60	4.59	3.34	10	178	23	23
Sunrise River at Sunrise, Minn.	3.00	Excellent	3.62	3.60	4.53	19	155	39	32
Trade River near Trade River, Wis.	2.98	Excellent	3.23	3.23	4.08	15	77	33	29
Upper Tamarack River near Markville, Minn.	2.56	Excellent	3.08	3.26	4.48	26	121	46	42
Valley Creek at Afton, Minn.	3.53	Very Good	3.84	4.18	3.02	9	142	16	13
Willow River at Burkhardt, Wis.	5.31	Good	4.17	5.46	2.43	2	78	12	12
Wood River near Grantsburg, Wis.	5.56	Good	4.71	5.35	3.78	12	159	34	29
Yellow River near Danbury, Wis.	3.74	Very Good	3.79	4.44	4.50	19	196	42	39

Kettle River also was low. The Kettle and Wood Rivers were sampled from woody debris rather than gravel riffles because gravel or cobble substrates were not available in the reach near the sample site. Therefore, habitat effects could be a factor in the benthic invertebrate measures at these streams (Shepard, 2001; Hooper, 1993).

NUTRIENT AND SUSPENDED-SEDIMENT LOADING IN THE ST. CROIX NATIONAL SCENIC RIVERWAY

Loads at the monitored tributaries were calculated by use of water-quality data from 1997 to 1999, flow data from 1999, and the USGS computer program Estimator. Annual loads and yields for water year 1999 from these monitored basins are listed in table 6 and shown in figure 4. In general, the tributaries with the largest drainage areas had the largest calculated loads. Loads transported during events were a large proportion of the annual load for many tributaries.

Annual yields from unmonitored tributaries were estimated from relations found between annual yields and the environmental characteristics of the monitored basins and the environmental characteristics of the individual unmonitored basins. Stepwise regressions indicated that annual nutrient and suspended-sediment yields were most strongly related to runoff, percentage of clay, permeability of the soil, soil erodibility, and basin slope. Also important were drainage area and the percentage of urban land use and wetlands in a basin. Fallon and McNellis (2000) found similar relations between nutrient and suspended-sediment yields during snowmelt and percentage clay and permeability of soil in their 1997 synoptic study.

Annual loads and yields estimated from regression equations using the environmental characteristics of the unmonitored tributaries are listed in table 6 and shown in figure 4, along with those calculated by use of the USGS computer program Estimator at monitored basins. Data are subdivided into quartiles in the figures. The range of the 95-percent confidence limits of loads and yields often were quite large because of the weakness of some of these relations (table 6). Errors of greater than 100 percent were not uncommon for the estimates at unmonitored tributaries. Volume-weighted concentrations are given in table 7.

Annual suspended-sediment loads in the tributaries were highest at the St. Croix River at Danbury (7,100,000 kg), Kettle (5,970,000 kg), Sunrise

(3,690,000 kg), and Snake (3,050,000 kg) Rivers, whereas annual suspended-sediment yields were highest at the Sunrise (8,400 kg/km²), Sand (7,860 kg/km²), and Upper Tamarack (6,350 kg/km²) Rivers. Annual total phosphorus loads were highest at Kettle (43,400 kg), Snake (37,400 kg), Apple (25,800 kg), and Sunrise (17,500 kg) Rivers, whereas Sunrise (39.9 kg/km²), Upper Tamarack (22.4 kg/km²), and Kinnickinnic (21.2 kg/km²) Rivers had the highest annual total phosphorus yields (table 6).

Nitrogen data were not collected at the Sunrise, Snake, and Kettle Rivers. Of the other tributaries monitored, the Apple and Lower Tamarack Rivers were major contributors of ammonia and organic nitrogen, whereas the Upper Tamarack, Lower Tamarack, and Wood Rivers were the tributaries with the highest annual total nitrogen yields. The tributaries in the southern, agricultural part of the basin (Kinnickinnic, Willow, and Apple Rivers) were those with the highest annual nitrate and nitrite loads and yields. The Sunrise River had the highest annual suspended-sediment and total phosphorus yields of the monitored tributaries in 1999.

Relation Between Streamflow and Tributary Loading

Examination of the hydrographs and concentrations of samples collected through the 1999 water year (fig. 5) helps to better explain why 1999 annual loads from the northern, forested basins were higher than loads from the southern, agricultural basins. A large portion of the annual loading to the St. Croix River tributaries comes from storm runoff events. Most of the 1999 loading from the northern, forested sites occurred during runoff-producing storm events when flows were orders of magnitude above base flow and concentrations were elevated as much as 10 to 50 times above baseflow levels. Runoff-producing events were fewer and smaller in the southern part of the basin in 1999 resulting in increases in flows that were less than double base flow and concentrations only slightly above base- flow levels. The southern tributaries had small annual loads in 1999 because of the lack of runoff events at those tributaries.

The Upper Tamarack River had the highest suspended-sediment and nutrient concentrations and yields of the northern tributaries. Loading at the Upper Tamarack River, similar to most of the loading from the small northern tributaries, was driven by runoff during events.

Table 6a. Calculated and estimated 1999 annual sediment and nutrient loads of St. Croix River Tributaries, Wisconsin and Minnesota [The best value is from Estimator or from a regression equation. The error represents the 95th-percent confidence range from Estimator or the regression equation. Max and min refer to the upper and lower 95th- percent confidence limits (best value plus or minus the error). All data are annual loads for water year 1999; USGS, U.S. Geological Survey, MPCA, Minnesota Pollution Control Agency; kg/yr, kilograms per year; best, best statistical estimate; error, standard error of prediction; NA, not available]

			Sec	liment			Phospho	orus, total	
			Annual I	oad (kg/yr)			Annual lo	oad (kg/yr)	
Sample locations	Agency	best	error	max	min	best	error	max	min
Kinnickinnic River near River Falls, Wis.	USGS	1,420,000	469,000	1,890,000	949,000	9,520	1,940	11,500	7,580
Valley Creek at Afton, Minn.	Estimated	1,840,000	162,000	346,000	21,300	748	714	1,460	34
Willow River at Burkhardt, Wis.	USGS	1,300,000	401,000	1,700,000	897,000	10,300	2,540	12,800	7,720
Browns Creek at Stillwater, Minn.	Estimated	280,000	292,000	572,000	0	2,950	1,640	4,580	1,310
Silver Creek near Stillwater, Minn.	Estimated	45,400	69,900	115,000	0	306	274	581	32
Apple River near Somerset, Wis.	USGS	1,810,000	346,000	2,150,000	1,460,000	25,800	5,600	31,400	20,200
Little Carnelian Lake outlet near Stillwater, Minn.	Estimated	24,500	37,900	62,300	0	65	136	201	0
Trade River near Trade River, Wis.	USGS	1,690,000	1,470,000	3,160,000	228,000	3,700	1,400	5,100	2,300
Sunrise River at Sunrise, Minn.	MPCA	3,690,000	2,710,000	6,400,000	982,000	17,500	3,360	20,900	14,200
Wood River near Grantsburg, Wis.	USGS	945,000	227,000	1,170,000	717,000	3,530	658	4,180	2,870
Snake River near Pine City, Minn.	MPCA	3,050,000	1,270,000	4,320,000	1,780,000	37,400	7,970	45,400	29,500
Kettle River below Sandstone, Minn.	MPCA	5,970,000	2,320,000	8,290,000	3,660,000	43,400	8,430	51,800	34,940
Lower Tamarack River near Markville, Minn.	USGS	1,600,000	519,000	2,120,000	1,080,000	8,100	1,680	9,780	6,420
Sand Creek near Hinckley, Minn.	Estimated	2,230,000	1,180,000	3,410,000	1,050,000	4,490	3,540	8,030	948
Crooked Creek near Hinckley, Minn.	Estimated	1,190,000	774,000	1,960,000	413,000	4,140	2,990	7,130	1,140
Clam River near Webster, Wis.	Estimated	1,750,000	3,110,000	4,850,000	0	7,740	11,500	19,300	0
Upper Tamarack River near Markville, Minn.	USGS	1,560,000	974,000	2,530,000	581,000	5,490	1,500	6,990	4,000
Yellow River near Danbury, Wis.	Estimated	1,400,000	3,330,000	4,730,000	0	12,600	14,100	26,800	0
Namekagon River near Leonards, Wis. 1	USGS	520,000	104,000	624,00	415,000	1,570	651	2,220	918
St. Croix River near Danbury, Wis. ²	USGS	7,100,000	2,120,000	9,210,000	4,970,000	16,900	6,370	23,300	10,500
St. Croix River at St. Croix Falls, Wis. ²	USGS	35,600,000	8,880,000	44,500,000	26,700,000	118,000	40,100	159,000	78,400

¹Loads calculated with 1996–98 data, calculated with 1999 flows.

²Loads calculated with 1997–99 data, calculated with 1999 flows.

³Because many concentration values were less than the detection limit, the estimated loads were negative and, therefore, set to 0.

Table 6a. Calculated and estimated 1999 annual sediment and nutrient loads of St. Croix Tributaries, Wisconsin and Minnesota—Continued

		Phosphoru	ıs, dissolved	t	P	hosphorus	ortho, disso	lved
		Annual le	oad (kg/yr)		· -	Annual	load (kg/yr)	
Sample locations	best	error	max	min	best	error	max	min
Kinnickinnic River near River Falls, Wis.	7,340	1,740	9,100	5,600	6,510	1,810	8,320	4,700
Valley Creek at Afton, Minn.	501	492	994	9	267	362	629	0
Willow River at Burkhardt, Wis.	4,170	1,900	6,060	2,270	3,490	1,150	4,650	2,340
Browns Creek at Stillwater, Minn.	912	896	1,810	16	539	654	1,190	0
Silver Creek near Stillwater, Minn.	204	200	404	4	27	163	190	0
Apple River near Somerset, Wis.	10,300	2,000	12,300	8,260	7,550	1,870	9,420	5,670
Little Carnelian Lake outlet near Stillwater, Minn.	110	108	218	2	13	89	101	0
Trade River near Trade River, Wis.	1,090	286	1,380	805	916	193	1,110	723
Sunrise River at Sunrise, Minn.	NA	NA	NA	NA	NA	NA	NA	NA
Wood River near Grantsburg, Wis.	1,140	193	1,330	946	905	200	1,110	705
Snake River near Pine City, Minn.	NA	NA	NA	NA	NA	NA	NA	NA
Kettle River below Sandstone, Minn.	2,480	587	3,060	1,890	NA	NA	NA	NA
Lower Tamarack River near Markville, Minn.	3,850	701	4,550	3,150	2,180	386	2,570	1,800
Sand Creek near Hinckley, Minn.	0^3	3,430	2,480	0	555	2,150	2,710	0
Crooked Creek near Hinckley, Minn.	1,260	2,290	3,560	0	71	1,910	1,980	0
Clam River near Webster, Wis.	6,450	8,740	15,200	0	1,770	6,950	8,720	0
Upper Tamarack River near Markville, Minn.	2,480	587	3,060	1,890	1,140	272	1,410	863
Yellow River near Danbury, Wis.	6,550	8,870	15,400	0	0^3	8,480	5,530	0
Namekagon River near Leonards, Wis. 1	872	515	1,390	357	905	422	1,330	483
St. Croix River near Danbury, Wis. ²	11,500	1,750	13,200	9,780	12,000	2,380	14,400	9,610
St. Croix River at St. Croix Falls, Wis. ²	57,000	19,900	76,900	37,100	51,700	8,350	60,100	43,400

¹Loads calculated with 1996–98 data, calculated with 1999 flows.

²Loads calculated with 1997–99 data, calculated with 1999 flows.

³Because many concentration values were less than the detection limit, the estimated loads were negative and, therefore, set to 0.

Table 6a. Calculated and estimated 1999 annual sediment and nutrient loads of St. Croix Tributaries, Wisconsin and Minnesota—Continued

	N	itrogen, amm	onia, dissolv	ed		Nitrogen, n	itrite, dissolv	ed	
		Annual lo	oad (kg/yr)			Annual	load (kg/yr)		
Sample locations	best	error	max	min	best	error	max	min	
Kinnickinnic River near River Falls, Wis.	3,140	2,220	5,360	925	1,940	436	2,370	1,500	
Valley Creek at Afton, Minn.	2,270	10,500	12,800	0	2,150	7,350	9,510	0	
Willow River at Burkhardt, Wis.	8,240	2,830	11,100	5,410	3,060	644	3,700	2,420	
Browns Creek at Stillwater, Minn.	0^3	27,400	20,400	0	0^3	19,200	12,000	0	
Silver Creek near Stillwater, Minn.	451	5,340	5,790	0	1,220	3,750	4,960	0	
Apple River near Somerset, Wis.	19,800	5,650	25,400	14,100	5,130	1,450	6,580	3,690	
Little Carnelian Lake outlet near Stillwater, Minn.	503	3,060	3,560	0	908	2,140	3,050	0	
Trade River near Trade River, Wis.	4,500	2,080	6,580	2,430	723	157	880	565	
Sunrise River at Sunrise, Minn.	NA	NA	NA	NA	NA	NA	NA	NA	
Wood River near Grantsburg, Wis.	4,110	1,290	5,400	2,830	515	129	643	386	
Snake River near Pine City, Minn.	NA	NA	NA	NA	NA	NA	NA	NA	
Kettle River below Sandstone, Minn.	NA	NA	NA	NA	NA	NA	NA	NA	
Lower Tamarack River near Markville, Minn.	7,360	2,830	10,200	4,530	759	1,380	2,140	0	
Sand Creek near Hinckley, Minn.	93,600	59,300	153,000	34,200	71,700	41,600	113,000	30,100	
Crooked Creek near Hinckley, Minn.	73,500	50,900	124,000	22,500	57,600	35,700	93,300	21,900	
Clam River near Webster, Wis.	78,900	194,000	273,000	0	77,300	136,000	213,000	0	
Upper Tamarack River near Markville, Minn.	4,100	2,490	6,590	1,610	310	630	940	0	
Yellow River near Danbury, Wis.	0^3	224,000	152,000	0	0^3	157,000	132,000	0	
Namekagon River near Leonards, Wis. 1	2,740	1,070	3,800	1,670	1,050	315	1,370	736	
St. Croix River near Danbury, Wis. ²	26,400	7,080	33,500	19,400	7,080	3,260	10,300	3,820	
St. Croix River at St. Croix Falls, Wis. ²	174,000	59,400	233,000	114,000	29,200	10,000	39,200	19,200	

¹Loads calculated with 1996–98 data, calculated with 1999 flows.

²Loads calculated with 1997–99 data, calculated with 1999 flows.

³Because many concentration values were less than the detection limit, the estimated loads were negative and, therefore, set to 0.

Table 6a. Calculated and estimated 1999 annual sediment and nutrient loads of St. Croix Tributaries, Wisconsin and Minnesota—Continued

	Nitro	Nitrogen, ammonia + organic, dissolved				Nitrogen, ammonia + organic, dissolved, total				
		Annual le	oad (kg/yr)		_	Annual lo	oad (kg/yr)			
Sample locations	best	error	max	min	best	error	max	min		
Kinnickinnic River near River Falls, Wis.	22,800	9,250	32,000	13,500	34,700	11,400	46,100	23,400		
Valley Creek at Afton, Minn.	360	311	671	49	37,700	14,400	52,100	23,300		
Willow River at Burkhardt, Wis.	44,100	6,830	50,900	37,300	79,300	18,100	97,400	61,200		
Browns Creek at Stillwater, Minn.	0^{3}	875	0	0	83,700	37,900	122,000	45,800		
Silver Creek near Stillwater, Minn.	456	156	612	300	8,390	6,020	14,400	2,360		
Apple River near Somerset, Wis.	146,000	15,300	161,000	130,000	192,000	23,800	216,000	168,000		
Little Carnelian Lake outlet near Stillwater, Minn.	276	86	362	190	2,510	3,430	5,940	0		
Trade River near Trade River, Wis.	29,500	3,280	32,800	26,300	3,700	1,400	5,100	2,300		
Sunrise River at Sunrise, Minn.	NA	NA	NA	NA	NA	NA	NA	NA		
Wood River near Grantsburg, Wis.	33,700	3,210	36,900	30,500	46,700	7,500	54,200	39,200		
Snake River near Pine City, Minn.	NA	NA	NA	NA	NA	NA	NA	NA		
Kettle River below Sandstone, Minn.	NA	NA	NA	NA	NA	NA	NA	NA		
Lower Tamarack River near Markville, Minn.	128,000	17,500	146,000	111,000	178,000	43,000	221,000	135,000		
Sand Creek near Hinckley, Minn.	5,070	1,770	6,830	3,300	44,000	76,900	121,000	0		
Crooked Creek near Hinckley, Minn.	4,210	1,530	5,740	2,680	0^3	65,100	54,700	0		
Clam River near Webster, Wis.	10,600	5,450	16,000	5,120	0^3	261,000	46,800	0		
Upper Tamarack River near Markville, Minn.	95,100	13,900	109,000	81,300	125,000	36,900	162,000	88,200		
Yellow River near Danbury, Wis.	5,660	5,700	11,400	0	0^3	273,000	10,200	0		
Namekagon River near Leonards, Wis. 1	28,200	4,540	32,700	23,600	40,300	3,660	43,900	36,600		
St. Croix River near Danbury, Wis. ²	349,000	67,900	417,000	281,000	455,000	51,800	506,000	403,000		
St. Croix River at St. Croix Falls, Wis. ²	1,630,000	234,000	1,860,000	1,390,000	2,210,000	245,000	2,460,000	1,970,000		

¹Loads calculated with 1996–98 data, calculated with 1999 flows.

²Loads calculated with 1997–99 data, calculated with 1999 flows.

³Because many concentration values were less than the detection limit, the estimated loads were negative and, therefore, set to 0.

Table 6a. Calculated and estimated 1999 annual sediment and nutrient loads of St. Croix Tributaries, Wisconsin and Minnesota—Continued

	Nitro	ogen, nitrite p	lus nitrate, dis	solved	Chloride, dissolved			
		Annual	load (kg/yr)			Annual	load (kg/yr)	
Sample locations	best	error	max	min	best	error	max	min
Kinnickinnic River near River Falls, Wis.	483,000	23,400	507,000	459,000	1,620,000	53,600	1,670,000	1,560,000
Valley Creek at Afton, Minn.	174	21	195	153	108,000	53,200	161,000	54,500
Willow River at Burkhardt, Wis.	286,000	69,100	355,000	216,000	1,930,000	578,000	2,510,000	1,350,000
Browns Creek at Stillwater, Minn.	337	38	375	299	205,000	97,600	303,000	108,000
Silver Creek near Stillwater, Minn.	54	9	63	45	26,900	20,900	47,800	6,080
Apple River near Somerset, Wis.	340,000	48,700	389,000	392,000	3,110,000	270,000	3,380,000	2,840,000
Little Carnelian Lake outlet near Stillwater, Minn.	29	5	34	24	14,200	11,300	25,500	2,920
Trade River near Trade River, Wis.	20,600	5,520	26,100	15,000	347,000	54,000	401,000	293,000
Sunrise River at Sunrise, Minn.	NA	NA	NA	NA	NA	NA	NA	NA
Wood River near Grantsburg, Wis.	9,330	1,450	10,800	7,880	287,000	22,600	310,000	265,000
Snake River near Pine City, Minn.	NA	NA	NA	NA	NA	NA	NA	NA
Kettle River below Sandstone, Minn.	NA	NA	NA	NA	NA	NA	NA	NA
Lower Tamarack River near Markville, Minn.	19,300	8,990	28,200	10,300	165,000	67,900	232,000	96,700
Sand Creek near Hinckley, Minn.	29	157	185	0	0^3	316,000	306,000	0
Crooked Creek near Hinckley, Minn.	0^3	136	0	0	0^3	284,000	158,000	0
Clam River near Webster, Wis.	2,540	402	2,940	2,140	1,040,000	978,000	2,020,000	64,300
Upper Tamarack River near Markville, Minn.	19,100	12,300	31,500	6,760	177,000	29,900	207,000	147,000
Yellow River near Danbury, Wis.	1,860	527	2,380	1,330	0^3	1,060,000	895,000	0
Namekagon River near Leonards, Wis. 1	15,600	5,040	20,600	10,500	394,000	89,400	483,000	304,000
St. Croix River near Danbury, Wis. ²	126,000	32,600	158,000	93,100	2,900,000	478,000	3,380,000	2,430,000
St. Croix River at St. Croix Falls, Wis. ²	822,000	185,000	1,010,000	637,000	18,300,000	2,310,000	20,600,000	16,000,000

¹Loads calculated with 1996–98 data, calculated with 1999 flows.

²Loads calculated with 1997–99 data, calculated with 1999 flows.

³Because many concentration values were less than the detection limit, the estimated loads were negative and, therefore, set to 0.

Table 6b. Calculated and estimated 1999 annual sediment and nutrient yields of St. Croix River Tributaries, Wisconsin and Minnesota

			Sec	liment			Phosph	orus, total	
			Annual yie	ld (kg/km²/y	r)		Annual yie	ld (kg/km²/y	/r)
Sample locations	Agency	best	error	max	min	best	error	max	min
Kinnickinnic River near River Falls, Wis.	USGS	3,160	1,050	4,210	2,120	21.2	4.3	25.5	16.9
Valley Creek at Afton, Minn.	Estimated	3,780	3,340	7,130	439	15.4	14.7	30.1	.7
Willow River at Burkhardt, Wis.	USGS	1,800	557	2,360	1,240	14.2	3.5	17.8	10.7
Browns Creek at Stillwater, Minn.	Estimated	3,170	3,300	6,470	0	33.3	18.5	51.8	14.8
Silver Creek near Stillwater, Minn.	Estimated	2,300	3,540	5,840	0	15.5	13.9	29.4	1.6
Apple River near Somerset, Wis.	USGS	1,270	243	1,510	1,030	18.2	3.9	22.1	14.2
Little Carnelian Lake outlet near Stillwater, Minn.	Estimated	2,290	3,550	5,850	0	6.1	12.8	18.8	.0
Trade River near Trade River, Wis.	USGS	4,910	4,250	9,150	660	10.7	4.1	14.8	6.7
Sunrise River at Sunrise, Minn.	MPCA	8,400	6,160	14,600	2,240	39.9	7.6	47.5	32.2
Wood River near Grantsburg, Wis.	USGS	4,350	1,050	5,390	3,300	16.2	3.0	19.3	13.2
Snake River near Pine City, Minn.	MPCA	1,210	502	1,710	706	14.8	3.2	18.0	11.7
Kettle River below Sandstone, Minn.	MPCA	2,650	1,030	3,680	1,620	19.3	3.7	23.0	15.5
Lower Tamarack River near Markville, Minn.	USGS	3,400	1,100	4,510	2,300	17.2	3.6	20.8	13.7
Sand Creek near Hinckley, Minn.	Estimated	7,860	4,150	12,000	3,710	15.8	12.5	28.2	3.3
Crooked Creek near Hinckley, Minn.	Estimated	4,860	3,170	8,030	1,690	16.9	12.3	29.2	4.7
Clam River near Webster, Wis.	Estimated	1,890	3,350	5,240	0	8.4	12.4	20.8	.0
Upper Tamarack River near Markville, Minn.	USGS	6,350	3,980	10,300	2,370	22.4	6.1	28.5	16.3
Yellow River near Danbury, Wis.	Estimated	1,490	3,550	5,030	0	13.4	15.0	28.5	.0
Namekagon River near Leonards, Wis. 1	USGS	1,560	314	1,880	1,250	4.7	2.0	6.7	2.8
St. Croix River near Danbury, Wis. ²	USGS	1,730	518	2,250	1,220	4.1	1.6	5.7	2.6
St. Croix River at St. Croix Falls, Wis. ²	USGS	2,200	550	2,750	1,650	7.3	2.5	9.8	4.9

¹Loads calculated with 1996–98 data, calculated with 1999 flows.

²Loads calculated with 1997–99 data, calculated with 1999 flows.

³Because many concentration values were less than the detection limit, the estimated loads were negative and, therefore, set to 0.

Table 6b. Calculated and estimated 1999 annual sediment and nutrient yields of St. Croix River Tributaries, Wisconsin and Minnesota—Continued

	PI	nosphorus,	dissolve	d	Phosphorus ortho, dissolved					
	Ar	nual yield	(kg/km²/y	r)	Α	nnual yiel	d (kg/km²	/yr)		
Sample locations	best	error	max	min	best	error	max	min		
Kinnickinnic River near River Falls, Wis.	16.3	3.9	20.2	12.5	14.5	4.0	18.5	10.5		
Valley Creek at Afton, Minn.	10.3	10.1	20.5	.2	5.5	7.4	12.9	.0		
Willow River at Burkhardt, Wis.	5.8	2.6	8.4	3.1	4.8	1.6	6.4	3.2		
Browns Creek at Stillwater, Minn.	10.3	10.1	20.5	.2	6.1	7.4	13.5	.0		
Silver Creek near Stillwater, Minn.	10.3	10.1	20.5	.2	1.3	8.3	9.6	.0		
Apple River near Somerset, Wis.	7.2	1.4	8.6	5.8	5.3	1.3	6.6	4.0		
Little Carnelian Lake outlet near Stillwater, Minn.	10.3	10.1	20.5	.2	1.2	8.3	9.5	.0		
Trade River near Trade River, Wis.	3.2	.8	4.0	2.3	2.7	.6	3.2	2.1		
Sunrise River at Sunrise, Minn.	NA	NA	NA	NA	NA	NA	NA	NA		
Wood River near Grantsburg, Wis.	5.2	.9	6.1	4.4	4.2	.9	5.1	3.2		
Snake River near Pine City, Minn.	NA	NA	NA	NA	NA	NA	NA	NA		
Kettle River below Sandstone, Minn.	1.1	.3	1.4	.8	NA	NA	NA	NA		
Lower Tamarack River near Markville, Minn.	8.2	1.5	9.7	6.7	4.6	.8	5.5	3.8		
Sand Creek near Hinckley, Minn.	$.0^{3}$	12.0	8.7	.0	2.0	7.6	9.5	.0		
Crooked Creek near Hinckley, Minn.	5.2	9.4	14.6	.0	.3	7.8	8.1	.0		
Clam River near Webster, Wis.	7.0	9.4	16.4	.0	1.9	7.5	9.4	.0		
Upper Tamarack River near Markville, Minn.	10.1	2.4	12.5	7.7	4.6	1.1	5.7	3.5		
Yellow River near Danbury, Wis.	7.0	9.4	16.4	.0	$.0^{3}$	9.0	5.9	.0		
Namekagon River near Leonards, Wis. ¹	2.6	1.5	4.2	1.1	2.7	1.3	4.0	1.5		
St. Croix River near Danbury, Wis. ²	2.8	.4	3.2	2.4	2.9	.6	3.5	2.3		
St. Croix River at St. Croix Falls, Wis. ²	3.5	1.2	4.8	2.3	3.2	.5	3.7	2.7		

¹Loads calculated with 1996–98 data, calculated with 1999 flows.

²Loads calculated with 1997–99 data, calculated with 1999 flows.

³Because many concentration values were less than the detection limit, the estimated loads were negative and, therefore, set to 0.

Table 6b. Calculated and estimated 1999 annual sediment and nutrient yields of St. Croix River Tributaries, Wisconsin and Minnesota—Continued

	Nitro	gen, amm	onia, diss	olved	Nitrogen, nitrite, dissolved				
	Aı	nnual yield	d (kg/km²/	yr)	А	nnual yie	ld (kg/km²	/yr)	
Sample locations	best	error	max	min	best	error	max	min	
Kinnickinnic River near River Falls, Wis.	7.0	6.1	13.1	0.9	4.3	1.0	5.3	3.3	
Valley Creek at Afton, Minn.	46.6	216	263	.0	44.3	151	196	.0	
Willow River at Burkhardt, Wis.	11.4	3.9	15.4	7.5	4.2	.9	5.1	3.4	
Browns Creek at Stillwater, Minn.	$.0^{3}$	310	230	.0	$.0^{3}$	217	135	.0	
Silver Creek near Stillwater, Minn.	22.9	271	293	.0	61.7	190	251	.0	
Apple River near Somerset, Wis.	13.9	4.0	17.9	9.9	3.6	1.0	4.6	2.6	
Little Carnelian Lake outlet near Stillwater, Minn.	47.2	287	334	.0	85.2	201	286	.0	
Trade River near Trade River, Wis.	13.1	6.0	19.1	7.0	2.1	.5	2.6	1.6	
Sunrise River at Sunrise, Minn.	NA	NA	NA	NA	NA	NA	NA	NA	
Wood River near Grantsburg, Wis.	18.9	5.9	24.9	13.0	2.4	.6	3.0	1.8	
Snake River near Pine City, Minn.	NA	NA	NA	NA	NA	NA	NA	NA	
Kettle River below Sandstone, Minn.	NA	NA	NA	NA	NA	NA	NA	NA	
Lower Tamarack River near Markville, Minn.	15.7	6.0	21.7	9.7	1.6	2.9	4.6	.0	
Sand Creek near Hinckley, Minn.	329	209	538	120	252	146	398	106	
Crooked Creek near Hinckley, Minn.	301	209	510	92.3	236	146	382	89.5	
Clam River near Webster, Wis.	85.2	209	295	0.0	83.4	147	230	.0	
Upper Tamarack River near Markville, Minn.	16.7	10.2	26.9	6.6	1.3	2.6	3.8	.0	
Yellow River near Danbury, Wis.	$.0^{3}$	238	162	.0	$.0^{3}$	167	141	.0	
Namekagon River near Leonards, Wis. 1	8.2	3.2	11.4	5.0	3.2	.9	4.1	2.2	
St. Croix River near Danbury, Wis. ²	6.5	1.7	8.2	4.7	1.7	.8	2.5	.9	
St. Croix River at St. Croix Falls, Wis. ²	10.7	3.7	14.4	7.1	1.8	.6	2.4	1.2	

¹Loads calculated with 1996–98 data, calculated with 1999 flows.

²Loads calculated with 1997–99 data, calculated with 1999 flows.

³Because many concentration values were less than the detection limit, the estimated loads were negative and, therefore, set to 0.

Table 6b. Calculated and estimated 1999 annual sediment and nutrient yields of St. Croix River Tributaries, Wisconsin and Minnesota—Continued

	Nitr	•	nonia + org	janic,	Nitrog	en, ammo	nia + organ	ic, total
		Annual yiel	d (kg/km²/	yr)	Α	nnual yie	ld (kg/km²/	yr)
Sample locations	best	error	max	min	best	error	max	min
Kinnickinnic River near River Falls, Wis.	50.7	20.6	71.3	30.1	77.4	25.3	103	52.1
Valley Creek at Afton, Minn.	7.4	6.4	13.8	1.0	776	297	1,070	479
Willow River at Burkhardt, Wis.	61.2	9.5	70.7	51.7	110	25.1	135	84.9
Browns Creek at Stillwater, Minn.	$.0^{3}$	9.9	.0	.0	947	429	1,380	518
Silver Creek near Stillwater, Minn.	23.1	7.9	31.0	15.2	425	305	730	120
Apple River near Somerset, Wis.	103	10.8	113	91.8	135	16.7	152	118
Little Carnelian Lake outlet near Stillwater, Minn.	25.9	8.1	33.9	17.8	235	321	557	.0
Trade River near Trade River, Wis.	85.6	9.5	95.1	76.1	10.7	4.1	14.8	6.7
Sunrise River at Sunrise, Minn.	NA	NA	NA	NA	NA	NA	NA	NA
Wood River near Grantsburg, Wis.	155	14.8	170	140	215	34.5	250	181
Snake River near Pine City, Minn.	NA	NA	NA	NA	NA	NA	NA	NA
Kettle River below Sandstone, Minn.	NA	NA	NA	NA	NA	NA	NA	NA
Lower Tamarack River near Markville, Minn.	274	37.3	311	236	380	91.7	471	288
Sand Creek near Hinckley, Minn.	17.8	6.2	24.0	11.6	155	270	425	.0
Crooked Creek near Hinckley, Minn.	17.2	6.3	23.5	11.0	.03	267	224	.0
Clam River near Webster, Wis.	11.4	5.9	17.3	5.5	.03	282	50.5	.0
Upper Tamarack River near Markville, Minn.	388	56.5	445	332	511	151	661	360
Yellow River near Danbury, Wis.	6.0	6.1	12.1	.0	$.0^{3}$	290	10.9	.0
Namekagon River near Leonards, Wis. 1	84.7	13.6	98.3	71.0	121	11.0	132	110
St. Croix River near Danbury, Wis. ²	85.3	16.6	102	68.7	111	12.6	124	98.4
St. Croix River at St. Croix Falls, Wis. ²	101	14.5	115	86.1	137	15.2	152	122

¹Loads calculated with 1996–98 data, calculated with 1999 flows.

²Loads calculated with 1997–99 data, calculated with 1999 flows.

³Because many concentration values were less than the detection limit, the estimated loads were negative and, therefore, set to 0.

Table 6b. Calculated and estimated 1999 annual sediment and nutrient yields of St. Croix River Tributaries, Wisconsin and Minnesota—Continued

	Nitroge	n, nitrite p	olus nitrate,	dissolved		Chlorid	e, dissolved	
		Annual yie	eld (kg/km²/	yr)		Annual yi	eld (kg/km²/y	ır)
Sample locations	best	error	max	min	best	error	max	min
Kinnickinnic River near River Falls, Wis.	1,080	52.1	1,130	1,020	3,600	119	3,720	3,480
Valley Creek at Afton, Minn.	3.6	.4	4.0	3.2	2,220	1,100	3,310	1,120
Willow River at Burkhardt, Wis.	396	95.9	492	300	2,680	802	3,480	1,880
Browns Creek at Stillwater, Minn.	3.8	.4	4.2	3.4	2,320	1,100	3,430	1,220
Silver Creek near Stillwater, Minn.	2.7	.5	3.2	2.3	1,360	1,060	2,420	308
Apple River near Somerset, Wis.	240	34.3	274	205	2,190	190	2,380	2,000
Little Carnelian Lake outlet near Stillwater, Minn.	2.7	.5	3.2	2.3	1,330	1,060	2,390	274
Trade River near Trade River, Wis.	59.6	16.0	75.6	43.6	1,010	156	1,160	850
Sunrise River at Sunrise, Minn.	NA	NA	NA	NA	NA	NA	NA	NA
Wood River near Grantsburg, Wis.	42.9	6.7	49.6	36.2	1,320	104	1,430	1,220
Snake River near Pine City, Minn.	NA	NA	NA	NA	NA	NA	NA	NA
Kettle River below Sandstone, Minn.	NA	NA	NA	NA	NA	NA	NA	NA
Lower Tamarack River near Markville, Minn.	41.0	19.1	60.1	21.8	350	145	495	206
Sand Creek near Hinckley, Minn.	.1	.6	.7	.0	$.0^{3}$	1,110	1,080	.0
Crooked Creek near Hinckley, Minn.	03	.6	.0	.0	$.0^{3}$	1,160	647	.0
Clam River near Webster, Wis.	2.7	.4	3.2	2.3	1,130	1,060	2,180	69.4
Upper Tamarack River near Markville, Minn.	78.0	50.4	128	27.6	724	122	846	602
Yellow River near Danbury, Wis.	2.0	.6	2.5	1.4	$.0^{3}$	1,120	952	.0
Namekagon River near Leonards, Wis. 1	46.8	15.2	62.0	31.7	1,180	260	1,450	915
St. Croix River near Danbury, Wis. ²	30.7	8.0	38.7	22.8	710	117	827	593
St. Croix River at St. Croix Falls, Wis. ²	50.9	11.5	62.4	39.4	1,130	143	1,270	988

¹Loads calculated with 1996–98 data, calculated with 1999 flows.

²Loads calculated with 1997–99 data, calculated with 1999 flows.

³Because many concentration values were less than the detection limit, the estimated loads were negative and, therefore, set to 0.

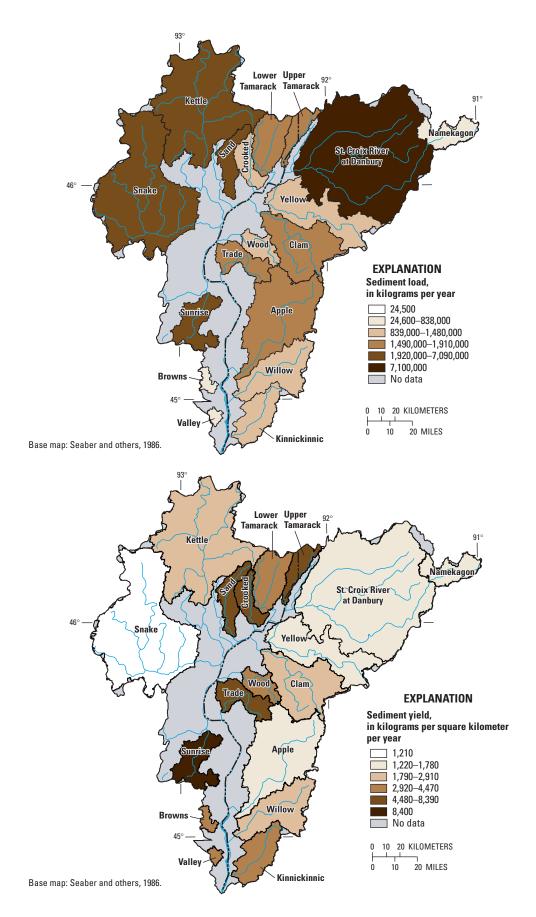


Figure 4a. Suspended-sediment loads and yields of the St. Croix tributaries, Wisconsin and Minnesota (categories are minimum, first quartile, second quartile, third quartile, fourth quartile, and maximum).

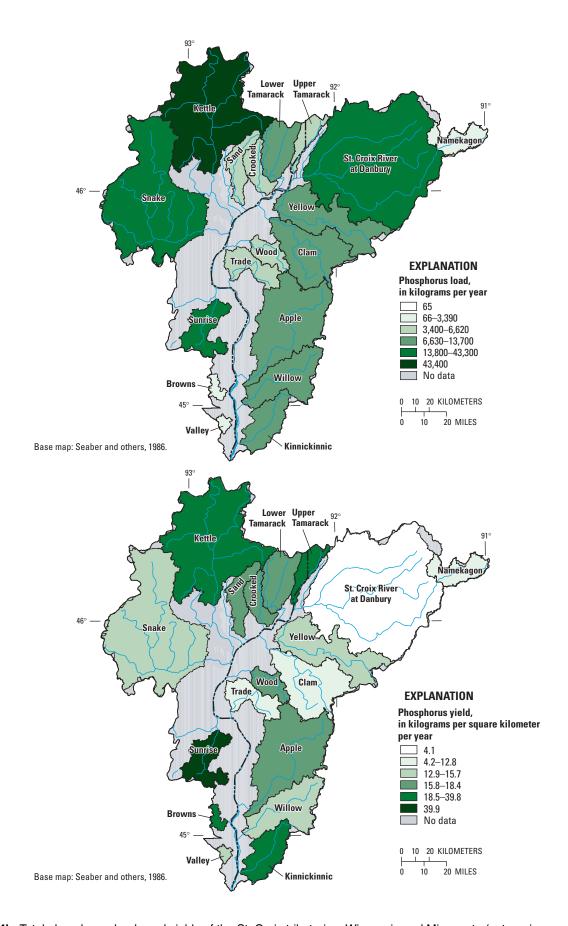


Figure 4b. Total phosphorus loads and yields of the St. Croix tributaries, Wisconsin and Minnesota (categories are minimum, first quartile, second quartile, third quartile, fourth quartile, and maximum).

Table 7. Volume-weighted concentrations of suspended sediment and nutrients at selected tributaries to the St. Croix River, Wisconsin and Minnesota, October 1, 1998—September 30, 1999 [mg/L, milligrams per liter; NA, not available; USGS, U.S. Geological Survey; MPCA, Minnesota Pollution Control Agency]

Sample locations	Agency	Drainage area (square kilometers)	Annual mean daily flow (cubic meters per second)	Runoff (centi- meters)	Suspended Phosphorus, sediment total (mg/L) (mg/L)	Phosphorus, total (mg/L)	Phosphorus, dissolved (mg/L)	Phosphorus, ortho, dissolved (mg/L)	Nitrogen, ammonia, dissolved (mg/L)	Nitrogen, nitrite, dissolved (mg/L)	Nitrogen, ammonia + organic, dissolved (mg/L)	Nitrogen, ammonia + organic, total (mg/L)	Nitrogen, nitrite plus nitrate, dissolved (mg/L)
Kinnickinnic River near River Falls, Wis.	USGS	449	3.14	22.1	14.3	960:0	0.074	0.066	NA	0.020	0.230	0.351	4.88
Willow River at Burkhardt, Wis.	USGS	721	3.86	16.9	10.6	.084	.034	.029	.068	.025	.362	.651	2.34
Apple River near Somerset, Wis.	NSGS	1,420	13.0	28.9	4. 4.	.063	.025	.018	.048	.013	.355	.467	.829
Trade River near. Trade River, Wis.	NSGS	345	2.18	19.9	24.6	.054	.016	.013	990.	.011	.430	.054	.299
Sunrise River at Sunrise, Minn.	MPCA	439	4.70	33.7	24.9	.118	NA	NA	NA	NA	NA	NA	NA
Wood River near Grantsburg, Wis.	NSGS	217	1.79	26.0	16.7	.062	.020	.016	.073	600°	.595	.826	.165
Snake River near Pine City, Minn.	MPCA	2,530	14.3	17.8	8.9	.083	000.	NA	NA	NA	NA	NA	NA
Kettle River below Sandstone, Minn.	MPCA	2,250	19.7	27.5	9.6	.070	.004	NA	NA	NA	NA	NA	NA
Lower Tamarack River near Markville, Minn.	NSGS	470	4.24	28.5	12.0	.061	.029	.016	.055	900.	.961	1.33	144
Upper Tamarack River near Markville, Minn.	NSGS	245	2.84	36.6	17.4	.061	.028	.013	.046	.003	1.06	1.40	.213
Namekagon River near Leonards, Wis.	NSGS	333	4.04	38.3	4.1	.012	.007	.007	.021	800.	.221	.316	.122
St. Croix River near Danbury, Wis.	USGS	4,090	38.3	29.6	5.9	.014	.010	.010	.022	900.	.289	.376	.104
St. Croix River at St. Croix Falls, Wis.	USGS	16,200	127	24.7	8.8	.038	.014	.013	.044	.007	.407	.554	.206

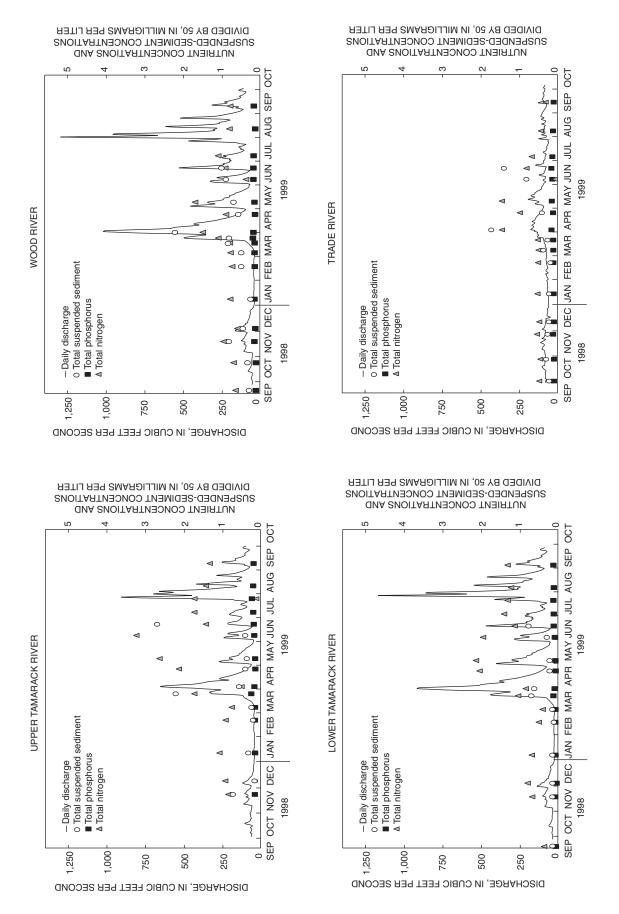


Figure 5a. Discharge hydrograph and sample concentrations from the St. Croix River tributary monitoring, Wisconsin and Minnesota, 1998-99.

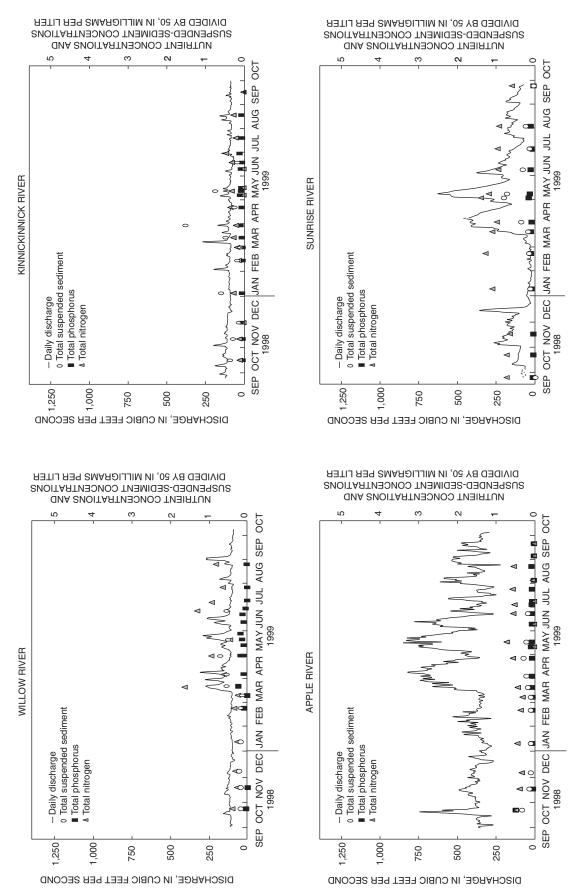


Figure 5b. Discharge hydrograph and sample concentrations from the St. Croix River tributary monitoring, Wisconsin and Minnesota, 1998–99

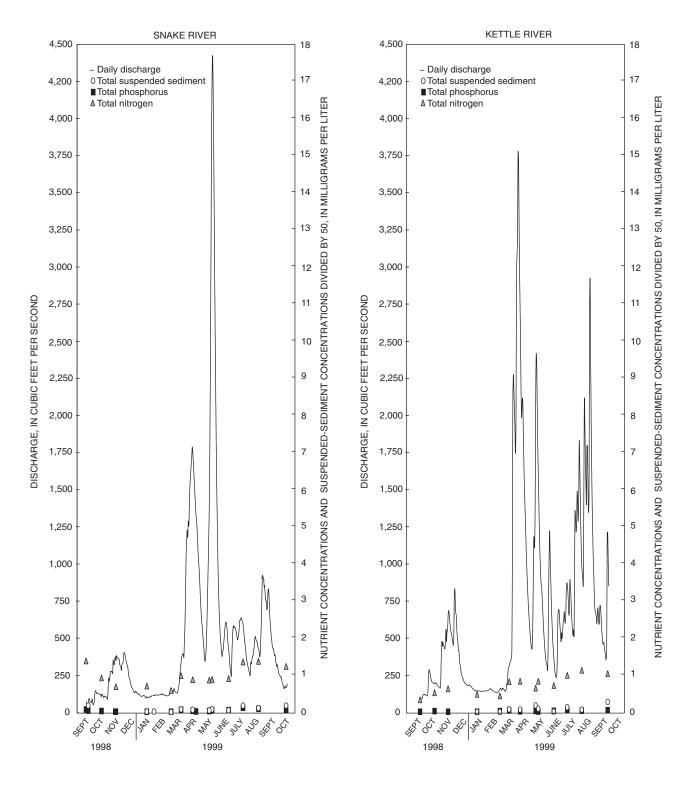


Figure 5c. Discharge hydrograph and sample concentrations from the St. Croix River tributary monitoring, Wisconsin and Minnesota, 1998–99.

At the Wood and Lower Tamarack Rivers, also, large events were associated with high concentrations of suspended sediment and nutrients. Data from 1997 and 1998 reinforce the finding that these northern tributaries can contribute substantial amounts of suspended sediment and nutrients to the St. Croix River during events but contribute little during base flow. In fact, these northern tributaries become insignificant to the overall nutrient and suspended-sediment loading to the St. Croix River during periods without event runoff.

Because of lack of runoff, event-related streamflow peaks were small and concentrations were consistently low for most of the southern tributaries, leading to the low overall basin loading from these tributaries in water year 1999, especially the Trade, Willow, and Kinnickinnic Rivers. In contrast, instantaneous loads and yields calculated for events sampled in 1998 show that these tributaries are not always characterized by a few small events and low rates of loading as seen in 1999. Data from the Kinnickinnic River can be used to illustrate the substantial between-year differences. The peak discharge at the Kinnickinnic River in 1999 was only 460 cfs, whereas a discharge of 640 cfs was measured on April 1, 1998. Historical records show an estimated peak discharge of 4,800 cfs on March 15, 1920, which shows the intensity of runoff that can occur at the site. The calculated instantaneous daily load for the April 1998 event was about 50 percent (664,000 kg) of the entire 1999 annual suspended-sediment load. Extrapolating loading rates to the record flow in March 1920 produces an estimated suspended-sediment load several orders of magnitude higher than the 1999 annual load, in just that one day. One event (or absence thereof) can greatly affect the year-to-year annual loads at these tributaries.

Base flows during winter at the Sunrise River were low and similar to those of the northern tributaries, although nutrient and suspended-sediment concentrations were higher. Events at the Sunrise River consisted of long periods of elevated flow rather than rapid, short-duration increases in flow. Also, summer base flows of the Sunrise River were high, compared to the northern tributaries. These elevated flows and consistently higher concentrations of nutrients led to the high loads and yields in the Sunrise River in 1999.

In general, loads from the Kettle, Snake, and Apple Rivers were driven by basin size and flow volume. The sheer volume of water entering the St. Croix River from these tributaries makes them substantial contributors to overall basin loading. Samples collected from the Apple River, which were collected below the Apple Falls Flowage, had consistently low concentrations, but the high annual flow volume of the river led to high annual loads. The Snake and the Kettle Rivers also had high flow volumes but were more variable in flow and concentration than the Apple River.

Relation Between Land Use and Tributary Loading

The 1997-snowmelt synoptic and previous studies examined the relation between land use and loading. Robertson and Saad (1996), for example, have shown that percentage of agriculture in a basin has a major effect on the magnitude of the nutrient yields. During the 1999 study, however, a relation between annual nutrient and sediment loading and the percentage of agriculture was not found. The tributaries with the highest percentage of agriculture are in the southern part of the basin (fig. 2), where only low-intensity, long-duration storms occurred in 1999, causing little runoff and loading. Tributaries with the lowest percentage of agriculture were in the northern part of the basin, where four or more high-intensity, short-duration storms caused large runoff events and high loads. This variability in runoff was caused by climatological factors rather than land use and appears to be more important than the differences in land use. This variability also led to the high errors when estimating loads in the unmonitored parts of the basins using environmental characteristics.

Base-flow, water-quality samples from agricultural areas had higher nitrogen concentrations than other areas of the St. Croix River Basin. Although nitrogen and phosphorus commonly are associated with each other, especially in agricultural areas, nitrogen is more likely than phosphorus to be transported through subsurface flow and released during base flow. Phosphorus is more adhesive to soil particles and is likely to be stored in the soils and released with eroded sediment during high-flow events (Wetzel, 1983). The elevated nitrogen concentrations (relative to other tributaries in the St. Croix River Basin) in base flow, the minimal runoff and suspended-sediment loading in 1999, and the results of the 1997 and 1998 event sampling showing higher phosphorus concentrations during periods of runoff all indicate that high phosphorus loading occurs during runoff events in agricultural areas of the southern St. Croix River Basin. Therefore, more phosphorus loading to the St. Croix River would be expected from

agricultural areas in normal to wet years than was calculated in 1999.

An inverse relation between annual yields and the percentage of urban area in the tributary basins was seen and is likely an effect of the less intense runoff and the resultant lower loads in the southern tributary basins, where most of the urban development is located rather than the urban development itself. The highest amount of urban area in any of the monitored tributaries was only 2.6 percent.

Point-Source Loading

In general, most nonpoint-source nutrient loading occurs during runoff events, which are prominent in the spring and early summer, whereas relative contributions from point sources and ground-water recharge are most consequential during low flows that are typical of fall and winter. Seasonal loads were estimated from the 1999 data in an attempt to document the relative importance of point and nonpoint sources of nutrients in the St. Croix River tributaries (fig. 6a). Most of the loading occurs in spring and summer, the seasons with the most runoff events. Loading in fall and winter varied among basins. Coinciding annual total phosphorus and suspended-sediment yields are shown in figure 6b.

The estimated point-source loadings from permitted discharges on each tributary are shown in figure 6a. These are most, but not all of the point-source discharges on these tributaries. Point-source data were provided by P. Prusak (Wisconsin Department of Natural Resources, written commun., 2000) and R. Wedlund, (Minnesota Pollution Control Agency, written commun., 2000). A large percentage of the annual load can be attributed to point-source discharges on tributaries where runoff was minimal in 1999, such as the Kinnickinnic and Willow Rivers. Whereas, point sources contributed only a small proportion of the annual load in the northern forested tributaries such as the Upper and Lower Tamarack Rivers, where nonpoint-source loads resulting from intense runoff events in the spring and summer dominated the annual loads.

Historical Loading

Calculated annual loads for 1999 were compared to estimated historical loading rates at four sites where long-term flow records were available, including the

Apple and Kettle Rivers and the two St. Croix River sites, at St. Croix Falls and near Danbury, Wisconsin. Annual loads for the previous 10 years were calculated at these long-term gaging stations using 1997–99 load/streamflow relations and historical flow. Annual loading rates varied greatly from year to year, but the 1999 annual loads were near the 10-year average: 75 to 114 percent of the average annual phosphorus load and 70 to 109 percent of the average annual suspended-sediment load, although the 1999 annual loads were lower than those computed for the early to mid-1990's (fig. 7). No long-term flow records were available for stations in the southern part of the basin to use as a comparison to long-term loading.

A study by Graczyk (1986) estimated average annual suspended-sediment and total phosphorus yields for 1974-81 at five tributaries and two main stem St. Croix River sites (fig. 8). Annual suspended-sediment yields in 1999 were greater at Yellow River (152 percent), Snake River (181 percent), and at St. Croix River at St. Croix Falls (157 percent) but less at the Clam River (69 percent), Kettle River (57 percent), and at the St. Croix River near Danbury (72 percent) in comparison. Annual phosphorus yields from 1999 were higher than those reported by Graczyk (1986) at the Kettle (125 percent), Snake (122 percent), and Apple (126 percent) Rivers and lower at the Yellow River (76 percent), Clam River (69 percent), and the two St. Croix River sites (near Danbury (38 percent) and at St. Croix Falls (29 percent)).

Difficulties with Load Analyses Based on One Year of Data

Data from the 1997 and 1998 synoptic studies demonstrated an increase in nutrient concentrations in agricultural areas and an increase in nutrient and sediment yields from basins with more impervious soils. Although the 1997 snowmelt synoptic (Fallon and McNellis, 2000) and previous studies (Robertson and Saad, 1996) have shown increased amounts of agricultural land in a basin to be related to increased loading, loading data from tributaries to the St. Croix River from 1999 did not show this relation. In water year 1999, nutrient and suspended-sediment yields throughout the St. Croix River Basin appear to be driven more by precipitation intensity than environmental characteristics of the tributary basins.

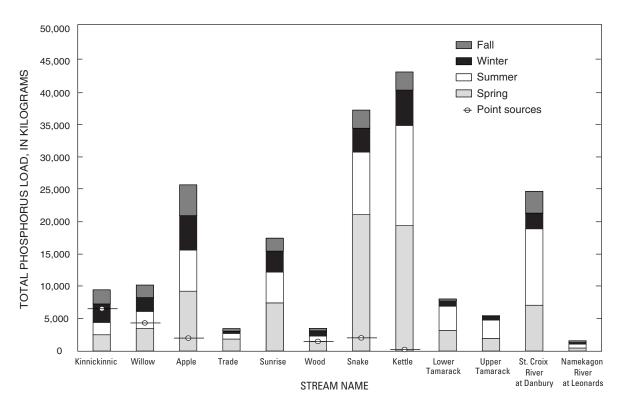


Figure 6a. Calculated phosphorus loads by season and from monitored point sources, at tributaries of the St. Croix River, Wisconsin and Minnesota, based on samples collected between October 1, 1998, and September 30, 1999.

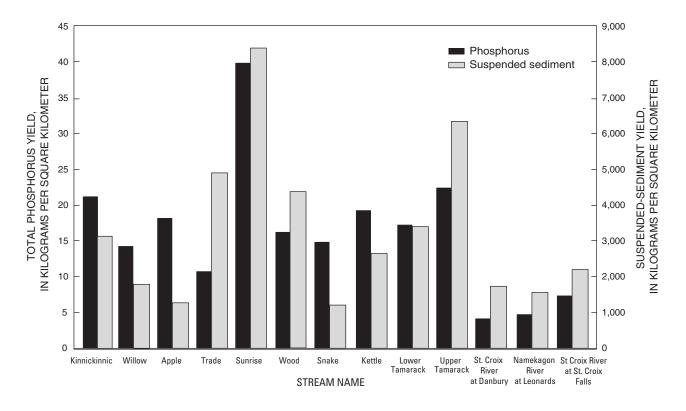
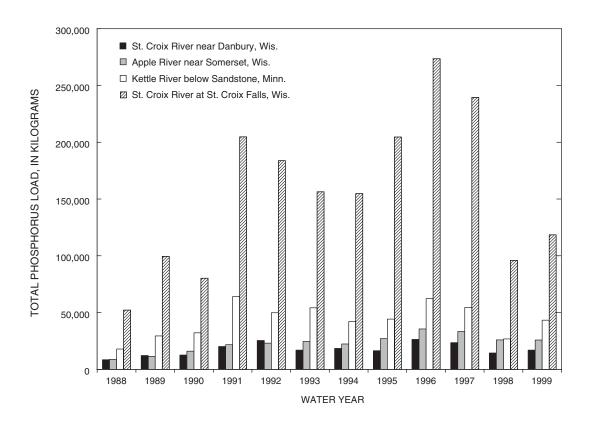


Figure 6b. Calculated phosphorus and suspended-sediment yields at tributaries of the St. Croix River, Wisconsin and Minnesota, based on samples collected between October 1, 1998, and September 30, 1999.



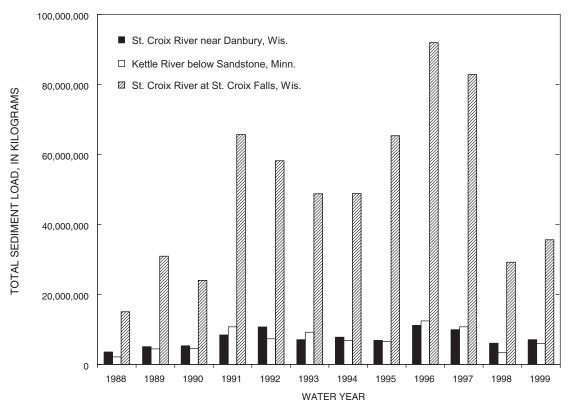


Figure 7. Estimated historical annual total phosphorus and suspended-sediment loads, Wisconsin and Minnesota, 1988–99.

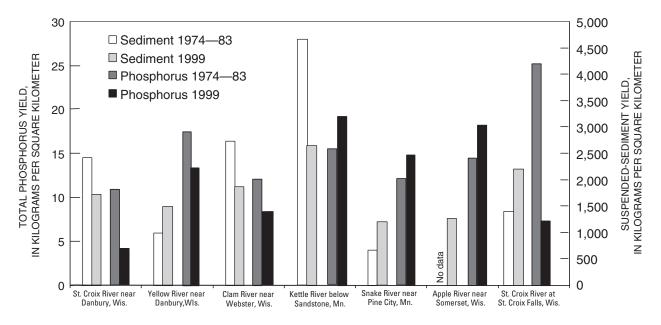


Figure 8. Average annual yield for 1974–83 (after Gracyzk, 1986) compared to annual yield for 1999 at tributaries to the St. Croix River, Wisconsin and Minnesota.

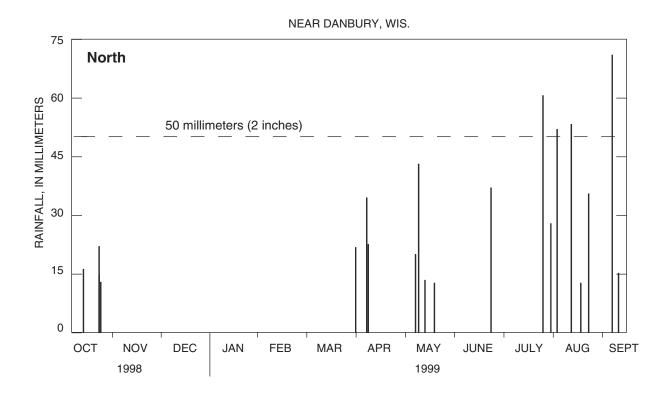
Precipitation amounts, intensity, and timing of events all affect the amount of runoff and the resulting magnitude of nutrient and suspended-sediment loading. Variations in weather in 1999 made geographical and temporal comparisons difficult. The lack of spring snowmelt runoff led to lower than average springtime flows and lower than normal springtime loads in the sampled tributaries. Although the entire basin had nearly average rainfall, the rain events in the northern part of the study area were larger, of shorter duration, and of higher erosivity potential. The southern part of the basin received longer duration storms that generated less runoff. Because a large part of the sediment and nutrient loading occurs during events in the St. Croix River Basin, this absence of snowmelt runoff and the disparity in rainfall intensity across the basin was a major explanatory factor for the unusual pattern of loading to the St. Croix River in 1999.

Loading during snowmelt was nearly indistinguishable from base-flow loading because the minimal snow accumulation in winter 1998–99 resulted in minimal runoff. A large portion of the annual load from the tributaries to the St. Croix River usually occurs in the spring, as documented in instantaneous snowmelt-runoff samples collected in 1997 (Fallon and McNellis, 2000) and 1998. The instantaneous daily loads, calculated during runoff events at tributaries to the St. Croix River in 1997 and 1998 often were greater than half the

entire 1999 annual loads calculated at those same sites. The 1999 annual loads would have been higher at all tributaries had there been more snowpack and a larger snowmelt event in 1999.

Disparity in rainfall intensity between the northern and southern parts of the St. Croix River Basin in 1999 is shown in figure 9, where size, intensity, and timing of rains in excess of 13 mm recorded by tipping bucket rain gages at two USGS gages are shown. (Rainfall was considered to be from separate events if a period of 2 hours resulted without rainfall.) The northern part of the basin is represented by the station near Danbury Wis., and the southern part of the basin is represented by the station in Wheeler, Wis.

Four storms, each producing more than 50 mm of rainfall occurred at the station near Danbury, Wis., whereas only one storm occurred at the station at Wheeler, Wis. The Wischmeier and Smith's erosivity index (1958), a measure of the kinetic energy and intensity of rainfall, was calculated for each event based on the total accumulation and duration of the precipitation (fig. 10). Five northern storms had erosivities greater than 25, whereas only one storm at Wheeler exceeded this value. The one storm at Wheeler was very localized and developed outside the St. Croix River Basin; daily precipitation that day measured 27 mm from River Falls, Wis. (56 km southwest) and less than 2.5 mm was measured at Amery, Wis. (48.3 km northwest), and



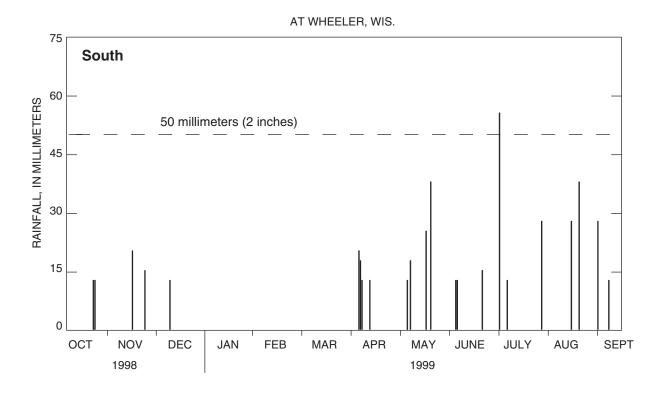
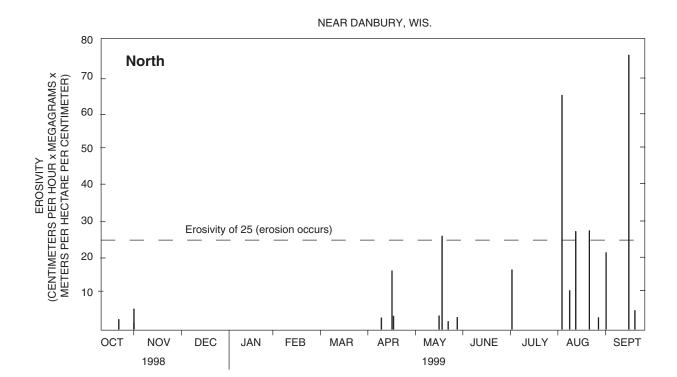


Figure 9. Rain events greater than 13 millimeters (0.5 inch) at two USGS stream-gaging stations, Wisconsin, 1998–99.



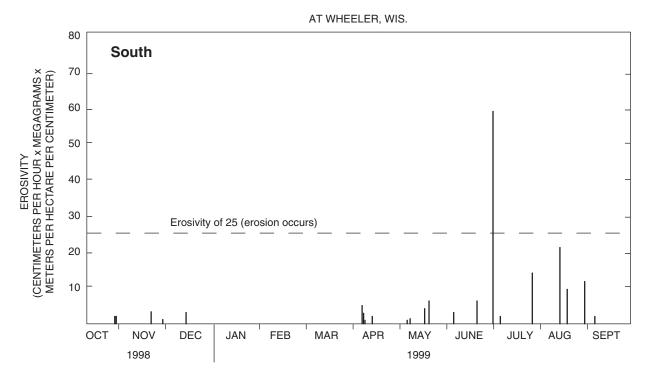


Figure 10. Rainstorm erosivity for events greater than 13 millimeters (0.5 inch) at two USGS stream-gaging stations, Wisconsin, 1998–99.

40

streamflow in the southern tributaries only slightly increased. Because of this disparity in runoff event size and intensity, the loading to the St. Croix River during events from tributaries in the northern part of the basin was far greater than that from the southern tributaries in 1999. Instantaneous loading rates from event samples in 1998 indicate that the southern tributaries can be substantial contributors of loads to the St. Croix River and confirm that event loading is negligible in the 1999 annual loads from these tributaries.

The variability in climatic conditions made it difficult to develop relations between environmental characteristics and annual yields, and weakened those relations that were found. With only 1 year of data, it could not be determined how the absence of snowmelt and the intensity of rain events affected various annual loads or the relation between annual loads and environmental characteristics; however, the variability in the runoff among tributaries was definitely a major factor in 1999 annual loading. Interestingly, the environmental characteristics found to be most related to annual loads in 1999 (increased clay soils in the basin, reduced permeability, and increased slope) are also those that have been shown to be related to runoff.

RELATIONS BETWEEN WATER QUALITY AND BENTHIC INVERTEBRATES

Benthic invertebrate measures were calculated (table 5) using data collected in the fall of 1999 and compared to 1999 annual loads. There were no significant relations evident between benthic invertebrate indices and 1999 annual loads. Benthic invertebrate indices indicated good to excellent water quality at all St. Croix River tributaries except for diversity and abundance measures at Valley Creek and Willow River and diversity and biotic index at Kettle River. No significant difference in the 1999 loads were evident at these three streams relative to the other tributaries.

The Kinnickinnic River, with the highest suspended-sediment (424 mg/L) and phosphorus (5.74 mg/L) concentrations sampled in 1998–99, and the Sunrise River, with the highest annual suspended-sediment (8,400 kg/km²) and phosporus (39.9 kg/km²) yields in 1999, both had excellent water quality based on benthic invertebrate biotic indexes used to indicate nutrient enrichment (and the associated drop in dissolved oxygen concentration). Diversity and abundance measures at these tributaries also indicated healthy invertebrate populations.

Benthic invertebrate health depends on habitat, chemical water quality, and the effect of the water quality—primarily sedimentation—on the habitat. Many of the tributaries had excellent water quality according to invertebrate indices and appeared not to be affected by nutrient and sediment loading. Those invertebrate indices that did show slight effects do not match the sites with high 1999 annual yields, an indication that invertebrate populations were affected by habitat rather than chemical water quality, that the 1999 annual loading rates were not an accurate representation of the long-term average chemical water quality of the tributaries, or more likely a combination of both of these effects.

The invertebrate indices for the Kettle and Wood Rivers likely were affected by the type of habitat sampled as much as the chemical water quality of the stream. These streams did not have hard substrate of gravel and/or cobble at the sampling sites, so benthic invertebrates were collected from woody debris and detritus. Habitat often has a greater effect on abundance and diversity than on Biotic Index (Rheaume and others, 1995; Lenz and Miller, 1996), although all measures at these streams could be affected. Willow River and Vallev Creek had good habitat for benthic invertebrates, but the benthic invertebrate indices indicated that the invertebrate communities were slightly impacted. These streams had minimal runoff and low annual loads in 1999 and may typically have higher average loads and yields than were measured in 1999. The concentrations and instantaneous loads measured at the Willow River in 1997 and 1998 support the possibility that nutrient or sediment loading may be affecting the invertebrates at these sites. Benthic-invertebrate measures can also be affected by other variables not measured in this study.

SUMMARY AND CONCLUSIONS

The St. Croix River Basin drains 20,010 km² of Minnesota and Wisconsin. The basin includes over 15 major tributaries and hundreds of minor tributaries to the St. Croix River. The Riverway is protected effectively by its status not only as a National Wild and Scenic River but also as a unit of the National Park System; however, its tributaries do not have that same status. Most recreational use in the park and tributaries involves contact with the water, such as boating and canoeing, accentuating the need for good water quality. In 1994, the NPS, WDNR, MDNR, MPCA, and Minnesota-Wisconsin Boundary Area Commission signed a memorandum of understanding that formed the inter-

agency St. Croix Basin Water Resources Planning Team (Basin Team) and agreed to investigate water-resources issues throughout the basin. Since 1994, the Basin Team has grown with the addition of other agencies including the USGS, Metropolitan Council Environmental Services, University of Minnesota, Minnesota Department of Agriculture, St. Croix Band of Chippewa Indians, and U.S. Environmental Protection Agency.

The USGS Upper Mississippi River Basin (UMIS) study unit of the NAWQA Program sampled tributaries to the Upper Mississippi River, including many in the St. Croix River Basin, as part of a synoptic study to characterize nutrient and suspended-sediment concentrations, loads, and yields during snowmelt in 1997. On the basis of results from the 1997 snowmelt synoptic study, a study was designed for 1998 to identify spatial and temporal variations in nutrients and suspended-sediment loading from tributaries identified as major contributors and other tributaries with environmental characteristics thought to be related to high nutrient and suspended-sediment loading. A follow-up study then was done from October 1998 to September 1999 on a subset of 11 sites from the 1998 study. Extensive waterquality data collected in coordination with continuous streamflow-monitoring data were used to compute annual nutrient and suspended-sediment loads and yields at the 11 tributary sites for water year 1999. Relations between environmental characteristics and calculated annual nutrient and suspended-sediment yields were used to predict loading from unmonitored tributaries of the St. Croix River Basin.

The Kettle River had the highest annual loads of total phosphorus and suspended sediment, whereas the Sunrise River had the highest annual yields of total phosphorus and suspended sediment among the tributaries to the St. Croix River in 1999. Sand Creek, and the Snake and Upper Tamarack Rivers were also substantial suspended-sediment contributors. Browns Creek, and the Snake, Apple, Upper Tamarack, and Kinnickinnic Rivers were also substantial contributors of total phosphorus in water year 1999. Instantaneous loading rates calculated for events in 1997 and 1998 demonstrated that the Willow, Trade, Wood, Crooked, and Lower Tamarack Rivers can also be substantial contributors to the total loading to the St. Croix River Basin.

The most important factor affecting annual yields in 1999 was rainfall intensity and runoff. Loading during storm events was a large portion of the annual loading of the St. Croix River tributaries. Storm events affecting the southern tributary basins were few and

small, resulting in loads and yields that were correspondingly small. In contrast, storm events affecting northern tributaries were more frequent and more intense, leading to higher loads and yields. Annual loads for 1999 in the southern St. Croix River Basin tributaries probably were lower than the long-term average loading rates. Trend analyses of loading rates in the St. Croix River Basin are not possible without long-term data.

Loading during events appears to be a large portion of the nutrient and suspended-sediment loading from tributaries to the St. Croix River. In some cases, a single event contributed most of the annual load of a given tributary. Because of the importance of event loads in the St. Croix River Basin, a single year of data collection is inadequate to fully evaluate the variability in annual tributary loading. Results from limited-length sampling depend highly on weather patterns and the amount and duration of runoff throughout the basin during the study period. The disparities in runoff caused by climatological factors throughout the basin may be more important to 1999 annual tributary loading than variability in land use and environmental characteristics.

The 1999 annual loads and yields in monitored tributaries to the St. Croix River calculated using Estimator had lower standard errors of prediction than the loads and yields calculated for unmonitored tributaries using relations between annual yields and environmental characteristics. Because of the climatological variability in the basin, relations between annual yields and environmental characteristics of the tributary basins were poorly defined resulting in large errors when predicting annual loads and yields in the unmonitored basins.

Benthic invertebrate indices indicated excellent to good water quality at all tributaries except Valley Creek, Willow River, and Kettle River. No relations were found between benthic invertebrate indices and the calculated and predicted annual tributary loads. Nutrient and sediment concentrations, loads, and yields from the tributaries to the St. Croix River were low and have had little effect on the benthic invertebrate populations at these streams.

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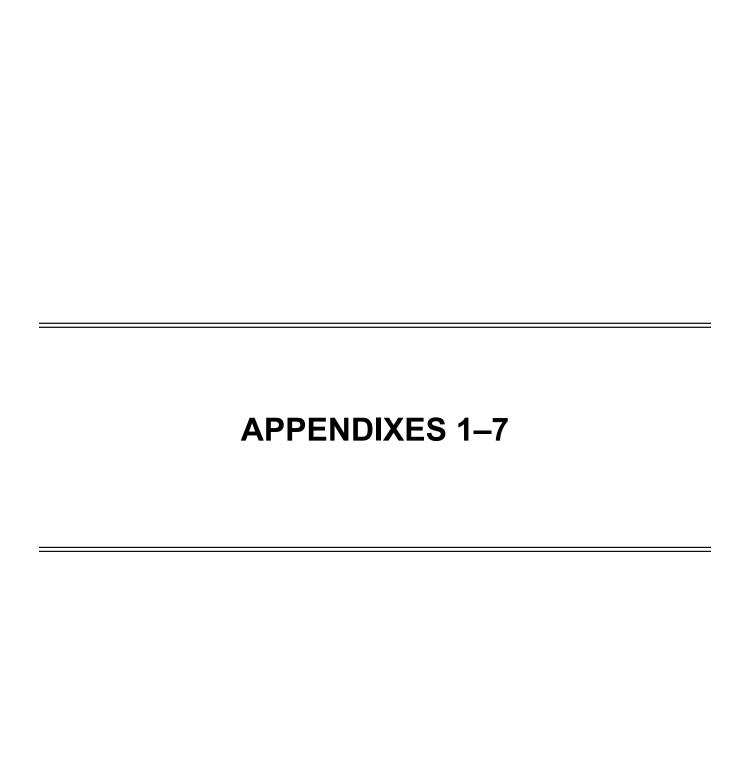
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Appendix 1. Results of St. Croix River tributaries quality-assurance/quality-control sampling, Wisconsin and Minnesota, 1998–99

[mg/L, milligrams per liter; %, percent; USGS, U.S. Geological Survey; MPCA, Minnesota Pollution Control Agency; Met. Council, Metropolitan Council Environmental Services; NA, not available; <, less than]

	Apple	River - June	22, 1998	Willow	River - June	22, 1998	Kinnickinr	nic River - J	lune 22, 1998
	USGS	MPCA	Met. Council	USGS	MPCA	Met. Council	USGS	MPCA	Met. Council
Total kjeldahl nitrogen, mg/L	0.57	0.64	0.76	1.29	1.23	1.2	0.306	0.37	NA
Dissolved kjeldahl nitrogen, mg/L	.385	NA	.53	.534	NA	.65	.206	NA	.29
Dissolved ammonia nitrogen, mg/L	.035	.03	.1	<.02	.05	.05	.020	.11	.05
Dissolved nitrite N, mg/L	.017	NA	<.03	.037	NA	.04	.023	NA	<.03
Dissolved nitrite plus nitrate, mg/L	.818	.77	NA	1.18	1.10	NA	5.05	4.6	NA
Total phosphorus, mg/L	.052	.057	.02	.088	.114	.09	.077	.079	NA
Dissolved phosphorus, mg/L	.023	NA	<.01	.011	NA	<.01	.063	NA	<.01
Dissolved orthophosphate as P, mg/L	.021	.026	.020	.02	.014	.019	.060	.062	.063
Chloride, dissolved, mg/L	7.3	NA	6	10.6	NA	9	17.0	NA	16
Suspended sediment, mg/L	4	3.6	3	15	19	17	5	.3	4
Sediment, % finer than 0.062 mm	95.24	NA	NA	97.96	NA	NA	76.67	NA	NA

	Clam I	River - Sept	. 24, 1999	Wood River -	Sept. 24, 1999	Trade Rive	r - Sept. 25, 99
	USGS	MPCA	Met. Council	USGS	MPCA	USGS	MPCA
Total kjeldahl nitrogen, mg/L	0.63	0.88	0.67	NA	0.56	0.44	0.32
Dissolved kjeldahl nitrogen, mg/L	.370	NA	NA	NA	NA	.500	NA
Dissolved ammonia nitrogen, mg/L	.035	.02	<.02	.020	<.02	.033	<.02
Dissolved nitrite N, mg/L	<.01	NA	<.03	<.01	NA	<.01	NA
Dissolved nitrite plus nitrate, mg/L	.124	.08	.13	.252	.25	.103	.06
Total phosphorus, mg/L	.034	.04	NA	.025	.022	.019	.019
Dissolved phosphorus, mg/L	.010	NA	NA	<.01	NA	<.01	NA
Dissolved orthophosphate as P, mg/L	<.01	NA	.027	.01	NA	.010	NA
Suspended sediment, mg/L	14	4.8	7	5	<1	3	1.6
Sediment, % finer than 0.062 mm	33.65	NA	NA	61.11	NA	83.33	NA

Appendix 2. Summary of analytical results from the U.S. Geological Survey 1997–98 synoptic sampling of tributaries to the St. Croix River, Wisconsin and Minnesota

		less than detection]
		no data; $0 =$
		millimeter;,
		liter: mm,
		nilligrams per
		iter; mg/L, r
		rograms per l
		: ug/L, micr
เรอบเล		square miles
	c	ſmi²,

Suspended sediment siner (percent finer (mm 200.0 nsnt)		78.48	50.00	80.00	76.32		14.90	68.18	63.16	80.95	80.00		72.33	83.87	91.67	77.78		64.00	76.40	68.6	82.35	94.74		19.05	43.88	22.13	12.69	33.65		84.37	65.60
Suspended toment, totot (∆\gm)		111	15	2	6		6	2	ю	5	4		26	14	4	2		160	12	201	9	4		78	13	100	7	14		20	32
Phosphorus, ortho, dissolved (mg/L)		0.032	.013	.015	0		0	.022	.014	910.	.026		.031	.015	.018	.010		.010	.032	.023	.023	.017		.010	.028	910.	.027	0		.032	.022
Phosphorus, dissolved (mg/L)		0.013	.022	.013	0		0	0	.013	710	.024		.014	.027	.025	910		.030	.018	.034	0	610:		.00	0	.011	010	010		.023	.034
Phosphorus, total (mg/L)		0.031	.049	0	.014		.020	0	.036	0	.045		\$	9	.024	610		.110	650	.072	610.	.023		.07	.037	.038	0	.034		.067	.064
Mitrogen, nitrite plus nitrate, desolved (mg/L)		0.383	.176	.053	820.		.460	.363	.241	.124	.175		.518	.133	.054	.064		.200	.490	.139	.196	.197		.280	.394	.071	.138	.124		.410	.161
Nitrogen, ammonia + organic, total (mg/L)		0.980	.937	1.145	ı		.300	.304	.342	.855	.630		1.03	.772	1.10	.360		006.	1.09	916.	.599	.730		009	765.	.370	830	.630		1.17	.935
Nitrogen, ammonia + organic, dissolved (mg/L)	, Minn.	1.027	.788	1.015	;	Vis.	0	.201	.192	.618	.530	rille, Minn.	.810	.832	.923	.350	Minn.	c .	878	.774	.466	.630	inn.	.500	.465	.211	.712	.370	Ë.	.835	.800
,ejitrio, nitrite, bevloselb (mg/L)	Upper Tamarack near Marksville, Minn.	0	0	0	0	ear Danbury, Wis.	0	0	0	.028	0	Lower Tamarack River near Marksville, Minn	0	0	0	0	Crooked Creek near Hinkley, Minn	0	0	0	0	0	ar Webster, Minn	0	.012	0	.021	0	near Hinkley, Minn	0	0
Nitrogen, ammonia, dissolved (mg/L)	er Tamarack r	0.192	.075	.064	0	Yellow River near	.050	0	.034	.112	.057	Tamarack Rive	.134	.081	.035	0	rooked Creek	.07	.178	.101	.034	.078	Clam River near	.17	.155	.033	.082	.035	Sand River ne	.195	.109
Hq (stinu brabnata)	ğ	7.7	7.8	7.5	7.7			7.8	7.9	8.6	7.8	Lower	7.6	7.4	7.4	7.5	o	;	7.8	7.7	7.6	7.7		7.3	7.8	8.0	8.3	8.0		7.7	9.7
Oxygen, dissolved (mg/L)		18.2	14.1	8.3	10.8		1	12.5	13.9	8.9	8.2		15.0	13.6	7.3	9.5		1	15.8	13.5	8.8	11.4		11.4	13.2	13.3	82.1	9.1		15.2	12.8
Specific conductance (mg/L)		26	51	138	220		1	190	155	24	172		98	53	159	257		ŀ	103	72	202	256		112	217	152	214	214		88	70
Temperature (degrees Celsius)		0.3	3.1	19.8	9.6		:	3.0	4.5	29.2	18.9		0.5	3.9	23.4	13.0		ŀ	1.4	4.1	19.9	11.3		4.3	3.4	5.0	28.8	15.4		4.	5.0
Discharge (cubic meters per minute)		167	780	15	S		758	646	870	498	306		498	1,208	49	10		1,465	263	583	30	13		1,305	472	901	321	287		318	532
Drainage area (square kilometers)		257	257	257	257		808^{1}	896	896	896	896		487	487	487	487		241	241	241	241	241		935	935	935	935	935		285	285
Date		02/25/98	03/30/98	07/22/98	09/22/98		04/03/97	02/24/98	03/31/98	07/14/98	09/22/98		02/24/98	03/30/98	07/22/98	09/22/98		04/03/97	02/24/98	03/30/98	07/22/98	09/22/98		04/03/97	02/23/98	03/31/98	07/14/98	09/24/98		02/24/98	03/30/98
Event Type		Snowmelt	Spring storm	Base flow	Low flow		Snowmelt	Snowmelt	Spring storm	Base flow	Low flow		Snowmelt	Spring storm	Base flow	Low flow		Snowmelt	Snowmelt	Spring storm	Base flow	Low flow		Snowmelt	Snowmelt	Spring storm	Base flow	Low flow		Snowmelt	Spring storm

Appendix 2. Summary of analytical results from the U.S. Geological Survey 1997–98 synoptic sampling of tributaries to the St. Croix River, Wisconsin and Minnesota—Continued

[mi², square miles; µg/L, micrograms per liter; mg/L, milligrams per liter; mm, millimeter; --, no data; 0 = less than detection]

Event Type	Date	Drainage area (square kilometers)	Discharge (cubic meters per minute)	Temperature (degrees Celsius)	Specific (mg/L)	bəvlossib ,nəgyxO (mg/L)	Hq (stinu bາsbnsts)	Nitrogen, ammonia, dissolved (mg/L)	Nitrogen, nitrite, dissolved (mg/L)	Vitrogen, ammonia + organic, dissolved (mg/L)	Nitrogen, ammonia + organic, total (mg/L)	Mitrogen, nitrite plus nitrate, bevlossib (J\pm)	Phosphorus, total (mg/L)	Phosphorus, dissolved (mg/L)	Phosphorus, ortho, dissolved (mg/L)	bendeau2 total (mg/L)	Suspended sediment (percent finer than 0.062 mm)
Base flow	07/22/98	285	46	23.7	142	7.6	7.7	.031	010.	929.	.793	960°	.042	.024	.019	5	89.29
Low flow	09/22/98	285	15	14.0	182	9.01	7.8	.021	0	390	.460	.067	.024	010	0	3	46.15
							North Fo	North Fork of Wood River	liver near Grant	near Grantsburg, Wis.							
Snowmelt	04/03/97	412^{1}	666	ı	;	;	;	080	0	.500	.700	.110	.070	.020	.020	92	80.60
Snow melt	02/23/98	4	155	3	210	12.9	7.7	.146	.013	.782	944	.317	.071	.025	.035	35	55.93
Spring storm	03/31/98	4	493	3.2	129	11.8	7.4	.052	0	989.	.807	.081	690:	.046	.028	25	21.34
Base flow	07/14/98	4	92	27.4	240	7.7	7.8	.043	.021	.648	.849	.170	.073	010	.025	17	80.00
Low flow	09/24/98	4	99	15.1	250	9.5	7.6	.020	0	;	;	.252	.025	0	.010	5	61.11
							F	Trade River nea	ar Trade River, Wis.	Wis.							
Snowmelt	04/01/97	347	859	3.2	160	11.55	7.2	.330	.010	.800	1.10	.390	220	.030	.020	204	31.32
Snowmelt	04/02/97	347	906	8.8	165	12.6	7.1	.300	.010	.700	1.10	.420	.180	0	.010	136	
Snowmelt	02/23/98	347	199	5.0	212	12.6	7.9	.139	.015	.583	.758	.424	.049	0	.030	18	69.17
Spring storm	03/31/98	347	372	4.0	145	13.3	7.8	.160	0	.571	.844	.360	.093	.029	.031	40	66.27
Base flow	07/22/98	347	124	21.8	202	8.4	8.0	.036	0	.439	.555	.105	.025	.012	.014	S	91.67
Low flow	09/24/98	347	82	12.2	216	10.2	8.0	.033	0	.500	.440	.103	010	0	.010	3	83.33
							₹	Apple River near	ar Sommerset, Wis.	Wis.							
Snowmelt	04/03/97	1,199	1,332	ı	ı	:	1	.030	0	300	.400	.950	060	.020	0	∞	21.90
Snowmelt	02/19/98	1,199	770	3.6	291	14.2	8.3	.122	0	.326	.379	1.29	.038	.029	.032	2	91.67
Spring storm	04/01/98	1,199	2,423	4.9	173	13.1	7.6	950.	0	.476	.463	.483	.056	.057	.038	13	79.05
Summer storm	06/26/98	1,199	1,021	22.7	239	9.7	7.8	.022	.018	.415	.576	.439	.050	.031	0	5	80.95
Base flow	07/28/98	1,199	440	22.8	280	7.8	8.0	.042	.013	300	.512	.804	.030	0	.017	2	90.00
							-	Willow River	near Hudson, Wis	Vis.							
Snowmelt	03/28/97	658^{1}	1,431	ı	ŀ	;	1	.520	.030	1.00	1.20	2.00	330	200	.170	19	7.10
Snowmelt	02/18/98	717	200	3.6	409	14.4	8.1	.231	.015	356	.611	4.36	.134	.034	.036	26	100
Spring storm	04/01/98	717	1,320	0.9	171	12.8	7.6	.298	.028	1.20	1.08	1.21	760	.106	.072	414	25.03
Summer storm	06/25/98	717	452	23.1	343	8.9	9.8	.043	.053	.516	.934	1.58	.024	0	0	11	83.50
Base flow	07/28/98	717	221	21.6	360	8.5	8.1	.078	.046	305	.700	1.58	090	0	.011	3	76.19
							Kinn	Kinnickinnic River	r near River Falls, Wis.	lls, Wis.							
Snowmelt	03/21/97	432	214	ı	ŀ	;	1	800°	.020	0	£.	5.00	80.	.05	.03	6	29.20
Snowmelt	02/18/98	432	528	3.1	335	14.3	7.9	2.26	.051	5.28	5.53	3.31	1.318	1.029	.928	252	44.12
Spring storm	04/01/98	432	1,087	5.8	228	12.8	7.8	.114	910.	.741	1.40	1.70	381	202	.158	424	31.70
Summer storm	06/25/98	432	511	20.6	316	7.9	7.8	.028	.043	.713	1.74	2.61	41.	.125	.064	318	57.16
Base flow	07/28/98	432	194	20.6	471	11.4	8.1	.027	.021	199	.353	4.85	.046	.054	.041	16	60.27
-		00:1															

¹Sampling location in 1997 differed from that in 1998.

Appendix 3. Summary of analytical results from Minnesota Pollution Control Agency (MPCA) sampling of the St. Croix River tributaries, Wisconsin and Minnesota, 1998–99 [--, no data; mg/L, milligrams per liter]

Date	Total sediment (mg/L)	Total nitrogen (mg/L)	Total phosphorus (mg/L)	Chloride (mg/L)
		ttle River below Sand	*	
02/25/98	15	1.29	0.121	
03/31/98	14	1.06	.090	
10/01/98		0.36	.035	
10/29/98		.55	.053	
11/24/98		.66	.033	
01/19/99	1.5	.51	.037	6.6
03/04/99	1.6	.47	.056	8.9
03/22/99	4.8	.85	.070	8.5
04/12/99	4.4	.86	.047	3.8
05/12/99	10	.68	.058	3.3
05/17/99	5.6	.86	.043	2.8
06/16/99	2.0	.75	.057	4.6
07/12/99	7.6	1.02	.081	4.0
08/09/99	4.4	1.16	.066	3.5
09/28/99	15	1.07	.079	3.5
	5	Snake River near Pine	City, Minn.	
02/25/98	3.0	1.47	.115	
04/01/98	5.0	.78	.059	
10/01/98		1.42	.102	
10/05/98	10		.073	
10/29/98		.96	.059	
11/23/98			.044	
11/24/98		.72	.044	
02/01/99	2.0		.050	
03/22/99	5.2	1.02	.096	8.5
04/12/99	4.4	.91	.065	4.4
05/17/99	5.2	.92	.078	3.5
06/16/99	4.0	.94	.084	4.1
07/12/99	10	1.39	.148	10
01/19/99	2.0	.74	.043	6.2
03/04/99	1.6	.62	.044	8.4
05/12/99	3.6	.90	.059	4.5
08/09/99	7.0	1.41	.121	8.5
09/28/99	10	1.27	.086	4.3
33120133	10	Sunrise River at Sun		1.5
02/25/98	27	1.32	.131	
04/01/98	46	1.03	.182	
10/01/98		.77	.072	
10/29/98		.77	.062	
11/24/98		.68	.060	
01/19/99	7.0	1.15	.108	15
03/04/99	8.6	1.33	.124	17
03/31/99	9.8	1.15	.127	12
04/12/99	19	1.03	.106	14.0
05/12/99	42	1.45	.182	12.0
05/12/99	38	1.24	.153	11
06/16/99	38 17	.99	.130	12
07/12/99	8.0	1.01	.107	14
08/09/99	13	.97	.130	14.0

Appendix 4. Summary of analytical results from the U.S. Geological Survey sampling of St. Croix River tributaries, Wisconsin and Minnesota, used in 1999 annual-load calculations

than 0.062 mm)

Suspended sediment (percent finer

(µ6w)

Suspended funent

(mg/L as Cl)																				
Chloride, dissolved		1	1	1	1	3.5	2.9	12.0	8.4	8.1	2.2	1.3	2.4	1.4	1.4	1.2	1.4	9.	1	1.4
Phosphorus, ortho, dissolved (mg/L as P)		0.032	.013	.015	<.010	<.010	.012	.017	.015	.018	.010	<.010	.017	<.010	.018	.020	<.010	<.010	1	.011
Phosphorus, dissolved (mg/L as P)		0.013	.022	.013	<.010	.015	.010	610.	.021	.020	.023	.021	.014	.014	.026	.036	440.	.042	ŀ	.037
Phosphorus, total (mg/L as P)		0.031	.049	<.010	.014	.032	<.050	.025	.029	.027	.124	.043	.020	.025	.047	090.	.074	.100	.108	.075
Nitrogen, nitrite plus nitrate, dissolved (mg/L as N)	le, Minn.	0.383	.176	.053	820.	.189	.148	2.70	.209	.241	.383	.224	<.050	<.050	.071	.067	.104	.107	ŀ	.113
Nitrogen, ammonia + organic, total (mg/L as N)	ar Marksvill	86.0	96.	1.10	.50	.72	.80	.95	62.	.63	1.60	.38	2.00	2.50	3.10	1.30	1.60	1.60	1	1.30
Nitrogen, ammonia + organic, dissolved (mg/L as N)	Upper Tamarack River near Marksville, Minn	1.0	0.79	1.0	.43	.71	89.	.91	.81	.64	.83	.87	.82	1.10	1.20	1.20	1.60	1.40	1	1.30
Nitrogen, nitrite, bevloseib (M ss J\gm)	Upper Tama	<0.010	<.010	<.010	<.010	<.010	<.010	.014	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010	ŀ	<.010
Vitrogen, ammonia, dissolved (mg/L as V)		0.192	.075	.064	<.020	.029	.025	690.	.129	960:	.108	.030	.031	<.020	<.020	<.020	.021	<.020	ŀ	.038
Hq (stinu brabnate)		7.7	7.8	7.5	7.7	7.9	7.5	7.7	7.3	8.9	7.0	6.4	7.9	7.4	7.8	7.1	7.4	7.0	1	7.4
Oxygen, dissolved (mg/L)		18.2	14.1	8.3	10.8	7.6	5.0	12.6	;	12.7	13.5	12.9	10.9	10.9	7.7	8.7	8.8	8.1	ŀ	0.6
Specific conductance (µS/cm)		76	51	138	220	107	76	358	206	192	77	40	62	52	62	99	62	52	ŀ	69
Water temperature (degrees Celsius)		0.3	3.1	19.8	9.6	2.6	1.	9.	0	0	1.6	2.6	14.7	11.9	22.4	20.0	19.1	23.1	ŀ	19.3
Discharge (cubic meters per minute)		167	780	15	S	102	102	19	17	20	510	856	100	296	177	302	1,716	675	972	204
Date		02/25/98	03/30/98	07/22/98	09/22/98	11/23/98	12/10/98	01/14/99	02/24/99	03/12/99	03/29/99	04/07/99	04/29/99	05/12/99	06/110/99	06/24/99	66/60/20	07/26/99	07/26/99	08/11/99

76 62 67 84 70 70

50

79

61

447591

22

64

63

83

1.9

.012

.038

.064

.128

1.20

1.10

<.010

<.020

7.5

9.2

78

19.8

172

66/80/60

Appendix 4. Summary of analytical results from the U.S. Geological Survey sampling of St. Croix River tributaries, Wisconsin and Minnesota, used in 1999 annual-load calculations—Continued
[µS/cm; microsiemens per centimeter; mg/L, milligrams per liter; mm, millimeter; <, less than; --, no data]

Suspended sediment (percent fine) (mm S90.0 nsht		72	84	92	78	100	29	82	75	95	92	80	68	68	78	87	;	86	88	68
babnaqsu2 finamibas (J\gm)		26	14	4	2	ю	2	3	2	ю	18	16	S	ď	7	20	1	10	11	16
Chloride, dissolved (mg/L as Cl)		ŀ	;	;	;	2.7	1.7	8.4	5.9	9.9	1.2	49.	1.3	.52	.51	.92	.63	<.10	89:	1.2
Phosphorus, ortho, dissolved (mg/L as P)		0.031	.015	.018	.010	.014	.014	.021	.015	.020	.017	.010	.017	<.010	.018	.026	.015	.010	.014	.018
,eunondeson beyloseib (9 as J\gm)		0.014	.027	.025	.016	.016	.019	.023	.021	.018	.029	.026	.016	.018	.029	.038	.051	.034	.037	.036
Phosphorus, total (mg/L as P)		0.044	440.	.024	610.	.042	<.050	.032	.037	.034	060.	.055	.028	.034	.051	.082	.081	.092	620.	.093
Nitrogen, nitrite plus nitrate, dissolved (M Sa J\pm)	le, Minn.	0.518	.133	.054	.064	.146	.094	.101	.168	.193	.485	.224	<.050	<.050	.051	.055	.082	.119	.054	980.
Hitrogen, ammonia + organic, total (M sa J\pm)	ear Marksvil	1.0	77.	1.1	.36	.70	.83	.71	.51	.40	1.1	88.	2.2	2.3	2.1	1.2	1.5	1.4	1.3	1.4
Nitrogen, ammonia + organic, dissolved (mg/L as N)	Lower Tamarack River near Marksville, Minn	0.81	.83	.92	.35	.63	.71	.67	.48	.41	.81	8.	.78	1.0	1.2	1.0	1.3	1.1	1.2	1.1
Nitrogen, nitrite, dissolved (mg/L as N)	Lower Tama	<0.010	<.010	<.010	<.010	<.010	<.010	.013	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010
Nitrogen, ammonia, dissolved (mg/L as N)		0.134	.081	.035	<.020	.028	.028	.112	.113	.067	.106	.029	.032	.028	440.	.038	.027	.049	.053	<.020
Hq (etinu brabnate)		7.6	7.4	7.4	7.5	7.7	;	7.6	7.2	7.4	7.0	6.3	7.3	7.3	7.2	6.9	7.2	6.9	7.3	7.3
bevlossib ,negyxO (L/gm)		15.0	13.6	7.3	9.5	5.6	;	12.8	;	12.9	12.9	11.8	10.1	8.6	7.0	7.1	7.4	6.5	7.4	8.3
Specific conductance (m3/Sµ)		98	53	159	257	121	ŀ	192	231	232	99	40	69	55	29	09	77	78	72	86
Water temperature (degrees Celsius)		0.5	3.9	23.4	13	2.3	ŀ	1.	0	6:	1.8	3.3	11.9	11.5	21.8	20	19.7	25.5	18.8	19.3
Discharge (cubic meters per minute)		498	1,208	49	10	136	194	27	31	25	646	1,689	194	523	333	895	376	438	460	381
əfsQ		02/24/98	03/30/98	07/22/98	09/22/98	11/23/98	12/10/98	01/14/99	02/24/99	03/12/99	03/29/99	04/07/99	04/29/99	05/12/99	06/11/90	06/24/99	66/60/L0	04/126/99	08/11/99	66/80/60

Appendix 4. Summary of analytical results from the U.S. Geological Survey sampling of St. Croix River tributaries, Wisconsin and Minnesota, used in 1999 annual-load calculations—Continued

	Suspended sediment (percent finer mm 290.0 mm)		99	21	80	19	78	53	74	93	68	80	09	58	22	55	72	87	57	ŀ	87	87
	babneqsu& sediment (J\gm)		35	25	17	5	9	18	6	4	10	10	18	18	53	12	15	20	22	1	S	10
	Chloride, dissolved (mg/L as Cl)		ŀ	ŀ	ŀ	1	;	4.9	4.3	6.4	5.7	9.9	4.6	ŀ	5.8	4.8	5.2	4.3	3.3	5.8	5.6	5.7
	Phosphorus, ortho, dissolved (mg/L as P)		0.035	.028	.025	.010	<.010	.013	.017	.016	<.010	<.010	<.010	.018	.016	.020	.010	.017	.022	<.010	<.010	.020
	Phosphorus, dissolved (mg/L as P)		0.025	.046	.010	<.010	<.050	.014	<.050	.013	600.	.007	.017	.037	.030	.013	.016	.012	.028	.026	.011	.030
	Phosphorus, fotal (P ss کا(m)		0.071	690:	.073	.025	.018	.059	.035	.058	.050	.055	.053	860.	.085	.042	.057	.085	980.	620.	.039	.074
	Nitrogen, nitrite plus nitrate, dissolved (M as J\pm)	burg, Wis.	0.317	.081	.170	.252	.141	.148	.160	.245	.311	.291	.167	.123	911.	.092	<.050	.134	060.	.179	.274	.154
	Vitrogen, ammonia + organic, total (mg/L as V)	near Grants	0.94	.81	.85	.55	09:	98.	.50	69:	.64	89.	.70	1.0	1.4	∞;	1.6	.28	.82	1.0	.71	69.
, no data]	Hitrogen, ammonia + organic, dissolved (M ss J\pm)	Wood River	0.78	69:	.65	.41	.52	.51	.58	.50	54	.49	.45	.78	.73	.45	99.	.65	89.	77.	.29	.57
millimeter; <, less than;, no data]	Nitrogen, nitrite, dissolved (mg/L as V)	North Fork of Wood River near Grantsburg, Wis.	0.013	<.010	.021	<.010	<.010	<.010	<.010	.016	<.010	.010	.010	<.010	<.010	<.010	<.010	<.010	.011	.015	<.010	<.010
m, millimeter	Nitrogen, ammonia, dissolved (M as J\gm)		0.146	.052	.043	.020	090.	.071	.058	.124	.148	.124	.070	.122	.074	090.	.048	.070	.034	920.	.021	<.020
er liter; m	Hq (stinu bາsbnsts)		7.7	7.4	7.8	9.7	7.5	7.7	7.8	1	7.6	7.3	7.3	7.3	7.3	7.7	7.7	7.6	7.3	7.7	7.9	9.7
illigrams _l	bevlossib ,negyxO (mg/L)		12.9	11.8	7.7	9.5	10.5	11.0	5.4	13.8	ŀ	13.0	12.3	7.3	11.3	10.2	9.1	9.7	7.1	8.0	9.4	8.4
mg/L, m	Specific endictance (mɔ/Sɹi)		210	129	240	250	211	192	200	230	267	258	207	164	152	203	197	192	131	192	234	184
r centimeter;	Water femperature (degrees Celsius)		3	3.2	27.4	15.1	10.4	3.8	2.1	0	2.1	1.6	3.3	6.7	9	13.4	11.7	21.8	22	21.2	21.8	19.2
osiemens per	Discharge (cubic meters per minute)		155	493	92	99	93	153	129	31	37	22	102	197	280	173	180	126	258	151	63	133
[µS/cm; microsiemens per centimeter; mg/L, milligrams per liter; mm,	Date		02/23/98	03/31/98	07/14/98	09/24/98	10/28/98	11/23/98	12/09/98	01/14/99	02/23/99	03/12/99	03/24/99	03/30/66	04/06/99	04/28/99	05/13/99	06/10/99	06/24/99	66/60/L0	08/11/99	66/80/60

Appendix 4. Summary of analytical results from the U.S. Geological Survey sampling of St. Croix River tributaries, Wisconsin and Minnesota, used in 1999 annual-load calculations—Continued

[µS/cm; microsiemens per centimeter; mg/L, milligrams per liter; mm, millimeter; <, less than;, no data]		
ter; mg/L, milligrams per liter; mm, millimeter; <, less than;		no n
ter; mg/L, milligrams per liter; mm, mi		than;
ter; mg/L, milligrams per liter;		millimeter; <,
ter; mg/L, m	;	iter;
[µS/cm; microsiemens per centimeter; mg/L		, milligrams
[μS/cm; microsiemens per centimeter; mg	1	딋
[µS/cm; microsiemens p		per centimeter; mg
		[µS/cm; microsiemens p

Suspended sediment (percent finer (mm S90.0 nsrlt		31	ŀ	69	99	92	83	29	78	95	91	75	96	87	71	92	53	81	47	;	73	83
babnaqau& tnamibaz (A\gm)		204	136	18	40	S	С	9	S	4	4	3	∞	S	45	53	53	20	36	ŀ	∞	8
Chloride, dissolved (mg/L as Cl)		ł	ŀ	ŀ	ŀ	ŀ	;	ŀ	4.1	4.2	4.5	5.0	5.3	5.3	4.9	8.4	4.1	4.7	3.6	3.9	4.5	13.0
Phosphorus, ortho, dissolved (mg/L as P)		0.020	.010	.030	.031	.014	.010	.013	.010	.011	.020	.010	.013	<.010	.012	610.	<.010	.026	.017	.013	.011	<.010
Phosphorus, dissolved (mg/L as P)		0.030	<.010	<.010	.029	.012	<.010	.016	.016	.010	.012	.012	.012	600.	.016	.013	.010	.020	.021	.025	610.	.012
enrondeod9. Istot (9 es J\gm)		0.220	.180	.049	.093	.025	.019	.017	.039	.019	.024	.034	.036	.038	.087	.042	620.	.065	.092	.064	.044	.025
Mitrogen, nitrite plus nitrate, dissolved (M as M)	Vis.	0.390	.420	.424	.360	.105	.103	.300	.318	.363	.423	.490	.488	.426	.275	.198	.140	.258	.121	.263	.185	.179
+ sinomms ,negorilN organic, total (M 2s A)(mg/L as M)	ade River, V	1.1	1.1	.76	.84	.56	44.	.41	.50	44.	49	74.	.46	.48	1.5	66.	1.5	<.10	.80	.65	.39	.34
Nitrogen, ammonia + organic, dissolved (mg/L as N)	Trade River near Trade River, Wis.	08.0	.70	.58	.57	44.	.50	.35	.40	.37	.41	.46	.41	.36	.55	.38	.52	.52	.54	.51	.29	.31
Uitrogen, nitrite, bevloesib (M 3s J\gm)	Trade	0.010	.010	.015	<.010	<.010	<.010	.022	.010	.018	.014	<.010	<.010	.013	<.010	<.010	.012	.011	<.010	<.010	<.010	<.010
Nitrogen, ammonia, dissolved (mg/L as N)		0.330	.300	.139	.160	.036	.033	.024	.075	.077	441.	144	.111	.045	.102	.037	.036	920.	.022	.020	.020	<.020
Hq (standard units)		7.2	7.1	7.9	7.8	7.9	8.0	7.8	7.8	7.3	7.6	7.8	7.9	7.8	7.6	8.0	7.6	7.6	7.6	7.9	8.0	8.0
bəvlossib ,nəgүxO (J\gm)		11.6	12.6	12.6	13.3	8.4	10.2	11.9	14.0	5.1	13.3	ı	12.3	12.6	11.7	10.6	6.6	6.6	8.3	8.5	8.7	10.4
Specific conductance (µS/cm)		160	165	212	145	202	216	188	213	208	203	242	228	219	148	200	144	144	140	180	202	212
Water temperature (degrees Celsius)		3.2	4.8	5	4	21.8	12.2	8.5	4.6	1.2	∞.	1.7	3.8	3.1	9.9	11.8	10.6	18.9	21.4	18.3	20.1	11.9
Discharge (cubic meters per minute)		859	906	204	372	124	82	122	150	134	73	100	100	129	292	165	325	112	219	139	109	127
Date		04/01/97	04/02/97	02/23/98	03/31/98	07/22/98	09/24/98	10/22/98	11/23/98	12/09/98	01/14/99	02/23/99	03/11/99	03/24/99	04/06/99	04/28/99	05/13/99	06/10/99	06/24/99	66/60/20	08/11/99	09/17/99

Appendix 4. Summary of analytical results from the U.S. Geological Survey sampling of St. Croix River tributaries, Wisconsin and Minnesota, used in 1999 annual-load calculations—Continued

Particular Particula	ciii, iiiiciot	ieilielis per	posenti menostemens per centinieter, ing/L, minigrams per mer, min, min	, mg/L, m.	Ingrama I		, , , , , , , , , , , , , , , , , , ,										
1,463 5.2 1.16 7.8 0.130 4.000 0.140 0.500 0.000 0.020 0	Date	(cubic meters	temperature	conductance	\ /		bevlossib	bevlossib	organic, dissolved	organic, total	plus nitrate, bevlossib	lstot	bevlossib	bevlossib	bevlossib	tnəmibəs	(percent finer
1445 52 237 116 7.8 6090 6010 630 6090 6090 6000<								Apple F	liver near Sc	ommerset,	Wis.						
770 36 29 142 83 129 483 129 689 693 694 694 694 694 693 694 694 694 695 694 695 694 695 694 695 694 695 694 695 694 695 694 695 694 695 694 695 694 695	03/28/97	1,463	5.2	237	11.6	7.8	0.030	<0.010	0.30	0.40	0.950	0.090	0.020	<0.010	1	8	78
1,43 1,45 1,45 1,45 4,46 4,43 6,46 4,43 6,46 4,43 6,46 4,43 6,46 6,43 6,49 <th< td=""><td>02/19/98</td><td>770</td><td>3.6</td><td>291</td><td>14.2</td><td>8.3</td><td>.122</td><td><.010</td><td>.33</td><td>.38</td><td>1.29</td><td>.038</td><td>.029</td><td>.032</td><td>1</td><td>2</td><td>92</td></th<>	02/19/98	770	3.6	291	14.2	8.3	.122	<.010	.33	.38	1.29	.038	.029	.032	1	2	92
1,021 2,24 2,95 7,6 7,8 4,1 3,6 4,9 6,90 6,91 6,9	04/01/98	2,430	4.9	173	13.1	9.7	950.	<.010	84.	.46	.483	950.	.057	.038	1	13	79
440 228 380 613 36 45 56 613 613 36 464 639 610 610 617 2 918 90 112 77 605 615 32 49 584 492 610 61 7 8 44 265 167 78 609 610 38 33 888 601 61 73 8 578 62 128 126 610 38 48 104 607	86/97/90	1,021	22.7	239	9.7	7.8	.022	.018	.41	.58	.439	.050	.031	<.010	ŀ	5	81
918 90 - 112 77 626 615 32 49 584 492 617 617 618 8 44 656 112 77 626 616 78 626 626 617 78 626 617 78 626 617 78 626 618 78 620 624 617 624 617 624 617 624 617 624 617 627 617 627 617 627 617 627 617 627 617 627 617 627 617 627 627 627 627 627 627 627 627 627 627 627 627 627 627 627 627 627 627 628 627 628 628 628 628 628 628 629 629 629 629 629 629 629 629 629 629 629 6	28/98	440	22.8	280	7.8	8.0	.042	.013	.30	.51	.804	.030	<.010	.017	1	2	06
- 44 265 167 7.8 629 610 24 39 760 024 017 017 017 017 017 017 017 017 017 017 017 017 017 017 018 019 019 017 017 018 019 019 019 011 018 019 019 011 018 019 011 018 019 011 018 019 019 019 019 019 019 019 019 019 019 019 019 011 018 019 011 018 019 011 018 019 011 018 019 011 018 019 011 018 019	23/98	918	0.6	ŀ	11.2	7.7	.026	.015	.32	.49	.584	.492	.017	.014	1	∞	99
95 21 26 6.8 7.8 6.01 38 33 858 077 047 0.02 71 2 447 0 29 127 8.1 1.26 6.01 34 1.16 0.43 0.03 0.09	18/98	1	4.4	265	16.7	7.8	.028	<.010	24	.39	.760	.024	.017	.01	7.3	3	94
447 0 292 127 811 126 017 35 45 116 043 023 019 80 2 538 0.5 3.2 1.2 0.83 6.01 34 36 131 0.99 0.01 0.99 0.91 0.99 0.91 0.99 0.91 0.99 0.91 0.99 0.91 0.99 0.91 0.99 0.91 0.91 0.99 0.91 0.99 0.91 0.91 0.99 0.91 0.91 0.92 0.91 0.92 0.91 0.92 0.91 0.92 0.91 0.92 0.91 0.92 0.91 0.92 0.91 0.92 0.91 0.92 0.91 0.92 0.91 0.92 0.91 0.92 0.92 0.91 0.92 0.91 0.92 0.92 0.92 0.92 0.92 0.92 0.92 0.92 0.92 0.92 0.92 0.92 0.92 0.92 0.92 0.92 0.92	86/80	595	2.1	265	8.9	7.8	.049	<.010	.38	.33	.858	.077	.047	.02	7.1	2	91
538 1.5 3.6 1.3 3.6 1.31 3.6 1.31 3.6 1.31 3.6 1.31 3.6 1.31 3.6 1.31 3.6 1.31 3.6 1.31 3.6 3.11 3.6 3.11 3.6 3.11 3.6 3.11 3.6 3.11 3.6 3.11 3.6 3.11 3.6 3.11 3.6 3.11 3.12 <t< td=""><td>13/99</td><td>447</td><td>0</td><td>292</td><td>12.7</td><td>8.1</td><td>.126</td><td>.017</td><td>.35</td><td>.45</td><td>1.16</td><td>.043</td><td>.023</td><td>.019</td><td>8.0</td><td>2</td><td>98</td></t<>	13/99	447	0	292	12.7	8.1	.126	.017	.35	.45	1.16	.043	.023	.019	8.0	2	98
330 1.5 288 13.2 7.7 045 011 26 31 1.25 035 018 019 8.2 2 693 54 270 13.5 8.1 0.32 0.15 28 44 1.02 0.53 0.19 0.19 7.3 4 1154 7.2 1.16 8.5 0.31 <0.10	23/99	578	0.5	302	1	7.8	.083	<.010	.34	.36	1.31	.039	.021	.021	7.8	2	44
693 54 270 135 81 032 015 64 102 053 019 012 73 4 1134 72 116 85 031 601 33 21 619 053 015 015 74 57 647 138 245 116 8.2 031 601 32 57 586 050 016 79 <t< td=""><td>11/99</td><td>530</td><td>1.5</td><td>288</td><td>13.2</td><td>7.7</td><td>.045</td><td>.011</td><td>.26</td><td>.31</td><td>1.25</td><td>.035</td><td>.018</td><td>.019</td><td>8.2</td><td>2</td><td>75</td></t<>	11/99	530	1.5	288	13.2	7.7	.045	.011	.26	.31	1.25	.035	.018	.019	8.2	2	75
1154 72 213 116 85 031 <010 33 21 619 053 015 011 74 5 647 138 245 164 82 035 <010	23/99	693	5.4	270	13.5	8.1	.032	.015	.28	4.	1.02	.053	.019	.012	7.3	4	87
647 13.8 245 10.4 8.2 6.01 32 57 586 050 016 0.02 70 7 1,327 - <	66/90	1154	7.2	213	11.6	8.5	.031	<.010	.33	.21	619.	.053	.015	.011	7.4	5	88
1,327 - <td>66/87</td> <td>647</td> <td>13.8</td> <td>245</td> <td>10.4</td> <td>8.2</td> <td>.035</td> <td><.010</td> <td>.32</td> <td>.57</td> <td>.586</td> <td>.050</td> <td>.016</td> <td>.022</td> <td>7.0</td> <td>7</td> <td>77</td>	66/87	647	13.8	245	10.4	8.2	.035	<.010	.32	.57	.586	.050	.016	.022	7.0	7	77
1,074 15.2 237 9.6 8.1 0.72 0.14 6.9 7.4 5.25 0.57 0.53 6.0 6.9 7.4 5.25 0.57 0.53 6.0 7.2 6.0 7.2 6.0 7.2	66/71	1,327	1	ŀ	ŀ	ŀ	ŀ	ı	ŀ	1	ŀ	620.	I	ı	ŀ	ŀ	ŀ
1,169	66/81	1,074	15.2	237	9.6	8.1	.072	.014	69.	.74	.525	.057	.023	<.010	6.9	5	100
472 20.1 28.1 8.5 8.0 0.35 0.17 38 56 818 0.52 0.58 0.58 0.58 0.59 0.79 0.79 7.3 4 673 22.3 26.3 28.3 5.8 5.8 0.58 0.58 0.59 0.79 0.79 7.7 8 559 -	66/60	1,169	ŀ	1	1	ŀ	ŀ	1	1	;	ŀ	.053	1	1	ŀ	ŀ	1
673 22.3 266 8.1 8.3 .031 .32 .58 .58 .58 .058 .058 .058 .058 .058 .058 .058 .058 .058 .058 .059 .07	22/99	472	20.1	281	8.5	8.0	.035	710.	.38	.56	.818	.052	.023	.021	7.3	4	95
595	03/99	673	22.3	266	8.1	8.3	.032	.011	.32	.53	.588	.058	.034	.034	7.2	∞	89
559 23.0 27.3 7.1 8.1 <.020 .052 .42 .58 1.43 .054 .016 <.010 13 2 907	66/L0	595	1	ŀ	ŀ	ŀ	ŀ	ı	ŀ	1	ŀ	.047	I	ı	ŀ	ŀ	ŀ
907	22/99	559	23.0	273	7.1	8.1	<.020	.052	.42	.58	1.43	.054	.016	<.010	13	2	83
607 16.4 430 11.9 8.3 .068 .014 .41 .55 .643 .067 .046 .027 6.8 4	02/99	200	1	ŀ	ŀ	ŀ	ŀ	ı	ŀ	:	ŀ	.065	ı	ŀ	ŀ	ŀ	ŀ
797	66/61	209	16.4	430	11.9	8.3	890.	.014	.41	.55	.643	.067	.046	.027	8.9	4	29
127 14.1 279 10.0 8.1 <.020 <.010 .25 .45 .579 .052 .031 .016 7.4 4	01/99	797	1	1	1	;	;	;	;	;	;	.048	;	;	;	;	;
	66/91	127	14.1	279	10.0	8.1	<.020	<.010	.25	.45	.579	.052	.031	910.	7.4	4	68

Appendix 4. Summary of analytical results from the U.S. Geological Survey sampling of St. Croix River tributaries, Wisconsin and Minnesota, used in 1999 annual-load calculations—Continued
[µS/cm; microsiemens per centimeter; mg/L, milligrams per liter; mm, millimeter; <, less than; --, no data]

orspended sediment (jedica) (j		100	25	84	92	96	87	29	87	77	92	75	93	94	:	1	83	1	1	ŀ	86	88	;	86	93
babnaqsu& fnamibas (J\gm)		26	414	11	3	S	5	9	5	4	4	14	17	20	ŀ	ŀ	10	ŀ	ŀ	ł	15	6	1	6	3
Chloride, dissolved (mg/L as Cl)		1	1	;	ŀ	1	13	13	14	15	13	12	12	12	ŀ	ŀ	12	ŀ	;	ŀ	11	12	7.9	12	91
Phosphorus, ortho, dissolved (mg/L as P)		0.036	.072	<.010	.011	.042	.036	.042	.037	.032	.023	.073	.043	.020	ŀ	ŀ	.032	ŀ	ŀ	ł	.020	.018	<.010	<.010	.018
,eunokpond bevloesib (P as J\gm)		0.034	.106	<.010	<.010	.038	.036	.031	.041	.036	.026	.120	.057	.013	ŀ	;	.041	ŀ	;	ł	.011	.016	.049	.013	.040
ensohqeohq. Sofal (A ss J\gm)		0.134	760.	.024	090.	.073	.057	.013	.064	.075	.054	.264	.121	.072	.111	.085	.091	.100	960.	.158	.088	.071	9200	680.	.061
Nitrogen, nitrite plus nitrate, dissolved (M ss J\pm)	ý	4.36	1.21	1.59	1.58	2.56	3.00	3.17	4.20	3.83	3.69	2.09	2.16	1.56	ŀ	ŀ	1.76	ŀ	1	ŀ	1.18	1.26	.731	1.45	2.17
Hitrogen, ammonia + organic, total (M ss J\pm)	River near Hudson, Wi	0.61	1.1	.93	.70	.43	.30	.35	.30	.38	.28	1.6	.74	1.0	ŀ	1	.46	ŀ	;	ł	1.3	86.	.65	.85	.40
Nitrogen, ammonia + organic, dissolved (mg/L as N)	v River near	0.36	1.2	.52	.31	.18	.20	.32	.24	.33	.18	76.	.48	.45	ŀ	ŀ	.52	ŀ	1	ł	.53	.41	.27	.30	.25
Nitrogen, nitrite, dissolved (mg/L as N)	Willow	0.015	.028	.053	.046	.023	.010	.015	.027	.015	.014	.023	.021	.017	ŀ	ŀ	.037	ŀ	ŀ	ŀ	.037	.03	.017	.035	.025
Nitrogen, ammonia, dissolved (mg/L as N)		0.231	.298	.043	.078	.041	.042	.064	.100	.075	.028	.206	.051	.042	ı	1	.220	ı	1	ŀ	<.020	.051	<.020	950.	<.020
Hq (stinu brabnata)		8.1	7.6	8.6	8.1	7.8	7.8	7.8	8.0	7.8	7.8	7.8	8.0	0.6	ŀ	1	8.1	ŀ	1	ŀ	8.2	8.9	8.6	8.5	8.5
bevlossib ,degyxO (mg/L)		14.4	12.8	8.9	8.5	10.1	13.5	8.2	15.2	;	14.0	12.8	10.9	11.5	1	1	10.3	1	;	1	8.2	8.4	8.0	8.9	10.2
Specific conductance (mS/Sµ)		409	171	343	360	357	391	385	427	420	400	285	302	331	ŀ	ŀ	363	ŀ	ŀ	ł	327	335	364	342	396
Water temperature (degrees Celsius)		3.6	0.9	23.1	21.6	10.3	4.4	5.1	1.1	2.1	2.9	4.1	9.1	14.0	ŀ	ŀ	14.4	ŀ	ŀ	ł	19.2	22.5	22.2	20.2	15.6
Discharge (cubic meters per minute)		200	1,320	452	221	206	228	228	180	195	187	491	343	195	199	314	319	411	352	442	207	206	195	246	184
Date		02/18/98	04/01/98	06/26/98	07/28/98	10/23/98	11/18/98	12/08/98	01/13/99	02/23/99	03/11/99	03/22/99	04/06/99	04/28/99	04/28/99	05/13/99	05/18/99	05/22/99	66/L0/90	06/14/99	06/22/99	01/03/99	07/22/99	08/19/99	09/16/99

Appendix 4. Summary of analytical results from the U.S. Geological Survey sampling of St. Croix River tributaries, Wisconsin and Minnesota, used in 1999 annual-load calculations—Continued

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Minimite Minet Falls Miles 0.020 6.020 0.030 - 9 7.1 0.021 6.02 0.030 0.030 - 9 7.1 0.021 6.02 0.03 0.030 - 9.2 7.1 0.02 6.02 0.03 1.03 9.28 - 9.2 7.1 0.01 7.4 1.7 2.61 1.4 1.25 0.64 - 9.2 4.4 3.2 0.03 7.1 1.7 2.61 1.4 1.25 0.64 0.64 0.64 0.64 0.64 0.67 0.64 0.67	(degrees Celsius) Specific conductance (µS/cm) Oxygen, dissolved (mg/L) PH (standard units)
0.3 5.00 0.080 0.030 9 5.5 3.31 1.32 1.03 928 922 1.4 1.70 3.81 2.02 1.58 424 1.7 2.61 1.44 1.25 0.64 424 1.3 4.85 0.46 0.64 1.6 424 1.1 5.14 0.60 0.62 0.64 1.6 4 5.43 0.60 0.62 0.67 1.7 8 1 5.43 0.62 0.67 0.74 16 4 4 1.1 0.62 0.67 0.74 16 4	Kinnick
5.5 3.31 1.32 1.03 928 252 1.4 1.70 381 202 1.58 424 1 1.4 1.20 1.58 424 1 1.7 2.61 1.44 1.15 0.64 318 1 1.4 5.14 0.06 0.05 0.04 116 1 5.14 0.06 0.06 0.06 0.07 0.07 17 8 2 5.10 0.74 0.07 0.07 0.07 17 8 3 5.14 0.07 0.05 0.07 17 8 16 4 4 1.1 5.53 0.62 0.05 0.04 17 4 16 4 16 4 16 4 16 4 16 4 16 16 4 16 16 17 16 17 17 18 16	458 13.9 8.3 <0.015 0.020
1.4 1.70 .381 .202 .188 - 424 1.7 2.61 .144 .125 .064 - .318 1.7 2.61 .144 .125 .064 - .318 1.9 .35 .485 .046 .054 .041 - .16 1.1 5.14 .060 .062 .057 .17 .8 .10 1.1 5.13 .072 .067 .074 .16 .4 .1 .1 .1 .1 .1 .1 .1 .1 .1 .1 .1 .1 .1 .0 .0 .0 .0 .1	335 14.3 7.9 2.27 .051
1.7 2.61 .144 .125 .064 318 3.5 4.85 .046 .054 .041 .16 1.4 5.14 .060 .062 .064 .0 .0 .0 1.1 5.43 .060 .062 .057 .17 .8 2.10 5.43 .072 .067 .074 .16 .4 3.1 5.74 .078 .062 .047 .17 .4 1.1 .053 .062 .047 .17 .4 2.1 .049 .049 .049 .049 .049 .049 .049 .072 .049 .16 .0 .0 2.4 .04 .087 .072 .072 .0<	228 12.8 7.8 .114 .016
35 4.85 .046 .054 .041 16 .14 5.14 .060 .062 .064 10 .15 5.43 .060 .062 .057 .17 8 <.10	316 7.9 7.8 .028 .043
.14 5.14 .060 .062 .064 10 .15 5.43 .060 .062 .057 17 8 <.10	471 11.4 8.1 .027 .021
.15 5.43 .060 .062 .057 17 8 <.10	433 13.8 8.0 .023 .018
<.10 5.43 .072 .067 .074 16 4 .31 5.74 .078 .065 .051 17 16 .17 5.53 .062 .052 .047 17 4 .21 5.41 .061 .049 .048 17 4 .24 3.89 .115 .063 .051 16 40 .24 3.89 .115 .063 .072 <td>474 17.8 7.8 <.020 <.010</td>	474 17.8 7.8 <.020 <.010
31 5.74 .078 .065 .051 17 16 1.7 5.53 .062 .052 .047 17 4 2.1 5.41 .061 .049 .048 17 3 3.1 4.62 .083 .054 .039 16 13 2.4 3.89 .115 .063 .072 17 40 3.9 4.94 .087 .072 .072 17 7 .130 <	471 7.4 7.5 .026 <.010
117 5.53 .062 .052 .047 17 4 21 5.41 .061 .049 .048 17 3 31 4.62 .083 .054 .039 16 13 24 3.89 .115 .063 .051 16 40 39 4.94 .087 .072 .072 17 7 .130 .095 <td>474 14.6 7.9 .025 .025</td>	474 14.6 7.9 .025 .025
21 5.41 .061 .049 .048 17 3 31 4.62 .083 .054 .039 16 13 24 3.89 .115 .063 .051 16 40 39 4.94 .087 .072 .072 17 7 .130 .095 .096 .31 5.05 .077 .063 .061 17 5 .25 4.94 .077 .064 .055 17 .31 4.08 .079 .078 .059 15 6 .28 4.78 .076 .066 .059 17 .31 4.08 .076 .056 17	485 8.2 <.020 <.010
.31 4.62 .083 .054 .039 16 13 .24 3.89 .115 .063 .051 16 40 .39 4.94 .087 .072 .072 .17 7 .1 .130 .2 .2 .2 .2 .2 .2 .2 .2 .2 .2 .2 .2 .1 .2 .2 .2 .2 .2 .2 .2 .2	467 15.8 8.1 <.020 .017
.24 3.89 .115 .063 .051 16 40 .39 4.94 .087 .072 .072 .17 7 .130	441 13.4 8.4 <.020 .021
.39 4.94 .087 .072 .17 7	400 11.8 7.9 .045 .016
.36 4.25 .102 .058 .048 16 20 .10 .095 .10 .21 .31 3.48 <	470 11.4 8.2 .049 .018
36 4.25 .102 .058 .048 16 20 .095 <td>: : : : : : : : : : : : : : : : : : : :</td>	: : : : : : : : : : : : : : : : : : : :
<td< td=""><td>461 11.6 8.2 .020 .026</td></td<>	461 11.6 8.2 .020 .026
1.06 5.05	1 1
31 5.05 .077 .063 .060 17 5 .51 3.48 .143 .066 .061 13 56 .25 4.94 .077 .064 .055 17 .31 4.08 .094 .078 .059 15 6 .28 4.78 .076 .066 .056 17 7	1 1
.51 3.48 .143 .066 .061 13 56 .25 4.94 .077 .064 .055 17 .31 4.08 .094 .078 .059 15 6 .28 4.78 .076 .066 .056 17 7	496 11.6 8.4 .020 .023
.25 4.94 .077 .064 .055 17 .31 4.08 .094 .078 .059 15 6 .28 4.78 .076 .066 .056 17 7	367 9.3 8.2 .030 .020
.31 4.08 .094 .078 .059 15 6 .28 4.78 .076 .066 .056 17 7	467 13.1 8.8 <.020 .019
.28 4.78 .076 .066 .056 17 7	430 11.9 8.3 <.020 .027
	485 12.3 8.8 <.020 .012

Appendix 5. Selected water-quality constituents analyzed at St. Croix River at St. Croix Falls, Wis., October 1, 1998 to September 30, 1999 [μS/cm, microsiemens per centimeter; mg/L, milligrams per liter; <, less than; --, no data]

Phosphorus, ortho, dissolved (mg/L as P)	<.010	.010	.017	.014	<.010	<.010	<.010	.012	.013	.019	.014	.014	.013	<.010	.257	<.010
Phosphorus, dissolved (mg/L as P)	<.050	<.050	<.050	600.	<.004	.008	.022	.021	.022	.022	.028	.037	.032	.005	.013	.007
Phosphorus, total (9 ss J\gm)	0.012	.023	<.050	.023	.029	.025	.081	.054	.062	.054	.065	920.	.053	.018	.035	610.
Mitrogen, nitriite plus nitrate, dissolved	960:0	.212	.177	.357	.428	.445	.150	860:	960:	.209	.103	.140	.193	.087	.123	.333
Hitrogen, ammonia+ organic, total (M SK J\gm)	0.40	.45	.36	.34	.29	.29	88.	68.	.83	.63	.95	.92	77.	.45	.58	.37
Mitrogen, ammonia + organic, dissolved (mg/L as M)	0.29	.29	.32	.29	.26	.22	.52	.62	.62	.50	.65	.78	.58	.30	4	.27
Nitrogen, nitrite, plus nitrate, dissolved (M as J\gm)	<0.010	<.010	<.010	.010	<.010	.018	<.010	<.010	.013	<.010	<.010	<.010	<.010	<.010	<.010	<.010
Vitrogen, ammonia, dissolved (mg/L as V)	0.043	.028	.029	.082	.094	.055	.031	<.020	.028	<.020	.022	.025	<.020	<.020	.029	<.020
Bicarbonate (mg/L as HCO ₃)	107	ŀ	78	110	117	103	09	71	73	83	83	89	94	117	08	119
Alkalinity (mg/L as CaCO ₃)	87	ł	49	06	96	98	49	58	09	89	89	99	77	96	65	26
Specific conductance (mɔ/Sɹɹ)	207	197	167	224	221	217	125	143	151	163	148	118	169	182	159	220
Hq (stinU brabnata)	7.8	8.4	7.5	7.2	7.1	7.7	7.5	7.6	7.5	7.5	7.6	7.6	7.6	8.2	7.8	7.5
bevlossib ,negyxO (J\gm)	9.5	26.0	6.3	8.5	10.3	11.8	14.1	6.6	8.6	6.5	6.9	7.9	8.9	13.7	10.1	12.2
Specific Specific Charles Char	188	174	166	208	199	195	114	134	143	159	142	ŀ	162	169	147	217
Water temperature (degrees Celsius)	12.0	2.8	1.2	1.	0	4.	5.0	13.0	17.4	24.0	24.9	19.5	20.9	6.7	6.3	0
Discharge (cubic meters per second)	4,571	8,258	8,207	3,025	ł	2,736	21,240	16,159	10,399	8,955	10,569	14,919	8,156	ł	ł	1
Date	10/28/98	11/17/98	12/09/98	01/14/99	02/12/99	03/02/99	04/06/99	05/13/99	05/27/99	06/24/99	07/14/99	08/16/99	66/L0/60	10/26/99	11/03/99	12/23/99

Appendix 5. Selected water-quality constituents analyzed at St. Croix River at St. Croix Falls, Wis., October 1, 1998 to September 30.

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1999—Continued [mg/L, micrograms per liter; °C, degrees Celsius; mm, millimeter; <, less than;, no data]	inued ams per liter	r; μg/L, mic	rograms F	er liter; °(C, degrees	Celsius; n	g/L, micrograms per liter; °C, degrees Celsius; mm, millimeter;	ter; <, less th	c, less than;, no data]	a]	- - - -		
Date	Carbon, organic, dissolved (mg/L as C)	Carbon, organic, suspended, total (D as L\@)	,muisəngsM bəvlossib (BM ss J\Qm)	,muibo2 beyloseib (BN 2s J\gm)	Chloride, dissolved (mg/L as Cl)	eyllate, beyloseib (mg/L as SO ₄)	Fluoride, dissolved (mg/L as F)	Sillica, dissolved (μg/L as SiO2)	lron, dissolved (µg/L as Fe)	Manganese, dissolved (mg/L as Mn)	Solids, residue at 180°C, dissolved (mg/L)	fnemibes bebneqsu?	Suspended sediment, (percent finer (mm S60.0 nsnt)
10/28/98	4.6	0.4	7.9	4.1	5.5	4.6	<0.10	11	88	24	119	5	92
11/17/98	0.9	٠ċ	7.4	3.7	5.4	4.9	<.10	12	140	17	121	3	100
12/09/98	7.7	4.	6.5	3.6	4.3	6.3	<.10	12	190	18	118	3	73
01/14/99	4.4	<.20	8.7	3.9	4.6	5.0	<.10	17	160	28	140	2	50
02/12/99	3.2	<.20	8.8	4.2	8.8	5.5	<.10	17	110	30	136	4	100
03/02/99	3.4	<.20	7.9	4.0	5.2	4.6	<.10	15	140	24	135	4	100
04/06/99	10	2.1	8.4	2.9	3.7	8.8	<.10	9.2	280	4	96	18	91
05/13/99	41	∞.	5.6	3.3	3.8	4.1	<.10	8.4	220	31	105	12	81
05/27/99	13	2.6	5.9	3.3	3.6	3.0	<.10	8.1	290	35	109	7	95
06/24/99	8.6	6:	6.5	3.2	4.1	3.5	<.10	9.1	250	27	120	6	66
07/14/99	16	∞.	6.3	2.8	3.9	2.5	<.10	11	290	30	119	∞	95
08/16/99	20	1.3	5.2	2.5	2.4	.95	<.10	13	940	36	115	12	88
66/L0/60	16	<.20	9.9	3.2	3.5	2.6	<.10	14	740	57	126	ı	ŀ
10/26/99	7.4	ŀ	7.8	3.9	4.2	3.6	<.10	11	240	21	116	70	76
11/03/99	6.6	L.	6.4	3.2	4.5	3.0	<.10	7.6	320	15	124	70	82
12/23/99	6.4	2.	6.7	3.9	5.3	3.5	<.10	15	200	33	140	10	100

Appendix 6. Selected water-quality constituents analyzed at St. Croix River near Danbury, Wis., October 1, 1998 to September 30, 1999 [μS/cm, microsiemens per centimeter; mg/L, milligrams per liter; <, less than; --, no data]

Phosphorus, ortho, dissolved (mg/L as P)	<0.010	<.010	.013	.025	<.010	<.010	ŀ	<.010	.014	.010	<.010	.013	<.010	<.010	<.010	<.010	<.010	<.010	<.010
Phosphorus, dissolved (mg/L as P)	<0.050	<.050	<.050	900.	.007	900.	1	.012	.011	.012	.020	.013	.018	.018	.021	.007	.004	.004	.007
Phosphorus, total (mg/L as P)	<0.050	.015	<.050	.018	.018	.021	.039	.040	.034	.037	.045	.037	.081	.062	.031	<:008	.014	.010	.017
Nitrogen, nitrite plus nitrate, dissolved (m as J\gm)	<0.050	.101	.110	.162	.214	.165	;	.127	.054	620.	.158	<.050	960.	<.050	.128	<.050	<.050	690.	.127
Hitrogen, ammonia+ organic, total (M SK L)(m)	0.21	.24	.24	.27	.25	.22	.55	.47	.53	09.	62:	5.	.95	.90	.62	.30	.30	.22	.35
+ Mitrogen, ammonia + croganic, dissolved (Mgk L Mgk)	0.19	17	18	19	.19	.15	;	.47	.46	.43	.59	.49	.62	1.5	.59	.26	.24	.21	.20
Nitrogen, nitrite, dissolved (mg/L as N)	<0.010	<.010	<.010	<.010	<.010	<.010	1	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010	<.010
Vitrogen, ammonia, dissolved (mg/L as V)	0.037	.024	<.020	.065	.030	<.020	;	.028	.037	<.020	.021	<.020	<.020	<.020	<.020	<.020	<.020	<.020	<.020
Bicarbonate (mg/L as HCO ₃)	77	ŀ	06	81	73	89	51	29	ŀ	37	44	55	42	25	62	71	72	62	84
Alkalinity (mg/L as CaCO ₃)	63	;	74	99	09	99	42	24	;	31	36	45	34	20	50	58	ŀ	64	69
Specific conductance (mɔ/Sɹɹ)	146	147	144	161	160	150	124	75	117	113	96	112	46	62	121	130	135	145	160
Hq (standard Units)	7.9	7.9	7.4	9.9	7.3	7.3	7.2	6.9	7.6	7.4	7.1	7.6	7.0	9.9	7.6	7.6	7.9	7.9	6.9
bevlossib, degyxO (Algm)	10.7	1	4.8	1	ŀ	12.2	12.7	12.1	95	8.1	73	73	65	4.9	8.4	95	10.4	14.2	12.2
Specific conductance (us)(cm)	131	130	135	156	147	130	92	99	101	104	72	103	71	48	114	115	119	121	142
Water temperature (degrees Celsius)	11.0	2.4	4.	0	0	4	4.1	5.5	13.9	19.2	21.1	23.0	25.3	22.1	19.0	14.4	10.4	3.4	0
Discharge (cubic meters per second)	1,511	1,716	2,175	1	ł	1	2,889	5,998	3,279	2,430	3,925	2,396	4,520	11,657	2,294	2,073	1	1,463	1
Date	10/02/98	11/17/98	12/10/98	01/26/99	02/24/99	03/17/99	03/30/99	04/07/99	05/11/99	06/11/90	06/24/99	07/20/99	07/26/99	07/27/99	08/31/99	09/27/99	10/12/99	11/18/99	12/22/99

Appendix 6. Selected water-quality constituents analyzed at St. Croix River near Danbury, Wis., October 1, 1998 to September 30, 1999—Continued [mg/L, micrograms per liter; ag/L, micrograms per lite

	Carbon, organ dissolved (mg/L as C)	Carbon, organi suspended, tot (D ss J\pm)	muisəngsM, bəvlossib (pM ss J\pm)	,muibo2 bevlossib (BN ss J\gm)	Chloride, dissolved (mg/L as Cl)	Sulfate, dissolved (mg/L as SO ₄)	Fluoride, dissolved (mg/L as F)	Silica, dissolved (µg/L as SiO ₂)	lron, dissolved (µg/L as Fe)	Asnganese, bevioesib (mM ss J\gm)	Solids, residue at 180°C, dissolved (mg/L)	nəmibəs bəbnəqsu2 (J\gm)	nemibes bebneqsu? (percent finer (mm 200.0 nsdt
10/02/98	3.2	0.3	5.0	2.7	2.6	4.2	<0.10	11	88	10	06	1	40
11/17/98	3.7	εċ	5.1	2.6	2.9	4.2	<.10	12	96	9.1	92	ю	77
12/10/98	5.9	ιċ	4.7	2.4	3.9	5.6	<.10	12	110	7.9	100	1	100
01/26/99	3.0	4.	5.5	2.8	3.4	8.4	<.10	16	120	14	100	ю	62
02/24/99	3.2	κi	5.4	2.8	2.9	4.4	<.10	15	220	9.5	96	ю	71
03/17/99	2.9	εċ	5.0	2.3	3.0	4.1	<.10	13	220	14	86	7	52
03/30/99	8.8	∞.	4. 4.	2.4	2.3	5.6	<.10	11	270	24	98	14	64
04/07/99	12	9:	2.8	1.8	1.8	5.7	<.10	8.6	290	37	75	13	74
05/11/99	111	ις	4.2	2.5	2.3	3.3	<.10	8.5	150	12	88	6	83
06/11/90	10	κi	4.2	2.2	2.1	2.8	<.10	8.0	210	18	87	13	98
06/24/99	20	2.2	3.7	2.0	2.2	2.2	<.10	9.0	400	26	94	14	88
07/20/99	16	9:	1	1	2.0	1.9	<.10	1	1	ł	06	7	77
04/26/99	14	2.2	3.3	1.7	.53	2.1	<.10	6.6	390	52	81	26	68
07/27/99	19	1.6	2.4	1.2	<.10	1.5	<.10	8.5	580	85	71	13	63
08/31/99	15	εċ	8.8	2.4	2.3	2.2	<.10	14	790	21	104	ю	94
09/27/99	9.3	5	5.0	2.6	3.0	3.2	<.10	12	220	12	94	ю	98
10/12/99	6.7	5.	8.8	2.6	2.6	3.4	<.10	10	160	8.6	83	8	85
11/18/99	4.7	ϵ i	5.1	2.6	2.8	3.4	<.10	12	120	12	92	4	99
12/22/99	4.1	"	8	с «	3.0	3.3	10	7	140	ć	103	u	Ċ

Appendix 7. Top five dominant benthic-invertebrate families and genera found in selected St. Croix River tributaries, Wisconsin and Minnesota [(xx%), percent of the total sample that were of a specific family, genera, or species]

Rank	Family	Genus	Species
	Sunri	ise River at Sunrise, Minn.	
1	Hydroptilidae (12%)	Leucotrichia (9%)	pictipes (9%)
2	Chironomidae (12%)	Optioservus (7%)	flavida (5%)
3	Hydropsychidae (11%)	Ceratopsyche (7%)	unidentified (5%)
4	Baetidae (10%)	Isoperla (6%)	flavistriga (4%)
5	Elmidae (10%)	Psychomyia (5%)	unidentified (4%)
	Snake	River near Pine City, Minn.	
1	Hydroptilidae (31%)	Stenelmis (19%)	unidentified (19%)
2	Chironomidae (19%)	Ceratopsyche (17%)	nr. convictum (14%)
3	Elmidae (19%)	Polypedilum (14%)	morosa bifida (14%)
4	Philopotamidae (14%)	Chimarra (13%)	obscura (13%)
5	Baetidae (9%)	Baetis (8%)	flavistriga (8%)
	Kettle R	tiver below Sandstone, Minn.	
1	Talitridae (59%)	Hyallela (59%)	azteca (59%)
2	Leptophlebiidae (13%)	Leptophlebia (4%)	unidentified (4%)
3	Chironomidae (7%)	unidentified (3%)	unidentified (3%)
4	Elmidae (3%)	Dubiraphia (2%)	minima (1%)
5	Tubificidae (3%)	Polypedilum (2%)	r. racovitzai (1%)
	Sand	Creek near Hinckley, Minn.	
1	Hydropsychidae (21%)	Ceratopsyche (16%)	unidentified (8%)
2	Tipulidae (13%)	Optioservus (9%)	angulata (7%)
3	Baetidae (10%)	Pseudolimnophila (8%)	flavida (6%)
4	Elmidae (10%)	Paracapnia (7%)	bronta (5%)
5	Capniidae (7%)	Baetis (7%)	unidentified (5%)
	Crooke	d Creek near Hinckley, Minn.	
1	Hydropsychidae (26%)	Cheumatopsyche (16%)	unidentified (16%)
2	Elmidae (16%)	Optioservus (15%)	angulata (11%)
3	Baetidae (12%)	Paracapnia (11%)	sparna (7%)
4	Chironomidae (11%)	Ceratopsyche (9%)	trivittatus (6%)
5	Capniidae (11%)	Baetis (6%)	flavistriga (5%)
	Lower Tama	rack River near Marksville, Minn.	
1	Elmidae (20%)	Paracapnia (16%)	angulata (16%)
2	Capniidae (16%)	Ceratopsyche (11%)	unidentified (8%)
3	Hydropsychidae (15%)	Optioservus (10%)	turbida (6%)
4	Baetidae (14%)	Stenelmis (8%)	morosa bifida (5%)
5	Philopotamidae (5%)	Baetis (6%)	sparna (5%)
	Upper Tama	rack River near Marksville, Minn.	
1	Capniidae (21%)	Paracapnia (21%)	angulata (21%)
2	Chironomidae (14%)	Ceratopsyche (8%)	flavida (4%)
3	Elmidae (10%)	Optioservus (8%)	vicarium (3%)
4	Hydropsychidae (9%)	Stenonema (5%)	mollis (3%)
5	Baetidae (8%)	Psychomyia (4%)	trivittatus (3%)

Appendix 7. Top five dominant benthic-invertebrate families and genera found in selected St. Croix River tributaries, Wisconsin and Minnesota—Continued [(xx%), percent of the total sample that were of a specific family, genera, or species]

Rank	Family	Genus	Species
	Browns	Creek near Stillwater, Minn.	
1	Baetidae (25%)	Baetis (24%)	brunneicolor (17%)
2	Elmidae (19%)	Optioservus (15%)	tuberosum (14%)
3	Hydropsychidae (17%)	Simulium (15%)	occidentalis (10%)
4	Simuliidae (15%)	Brachycentrus (10%)	betteni (9%)
5	Brachycentridae (10%)	Hydropsyche (10%)	tricaudatus (6%)
	Val	ley Creek at Afton, Minn.	
1	Baetidae (26%)	Baetis (26%)	tricaudatus (15%)
2	Hydropsychidae (24%)	Optioservus (14%)	pseudolimnaeus (14%)
3	Elmidae (14%)	Gammarus (14%)	slossonae (12%)
4	Gammaridae (14%)	Ceratopsyche (12%)	occidentalis (10%)
5	Brachycentridae (10%)	Brachycentrus (10%)	unidentified (9%)
	Willo	w River at Burkhardt, Wis.	
1	Hydropsychidae (60%)	Ceratopsyche (36%)	morosa bifida (33%)
2	Elmidae (31%)	Stenelmis (19%)	unidentified (19%)
3	Chironomidae (5%)	Cheumatopsyche (13%)	unidentified (13%)
4	Athericidae (2%)	Optioservus (12%)	fastiditus (7%)
5	Simuliidae (2%)	Atherix (2%)	variegata (2%)
	Yellov	w River near Danbury, Wis.	
1	Hydropsychidae (30%)	Serratella (14%)	deficiens (14%)
2	Chironomidae (18%)	Ceratopsyche (10%)	unidentified (8%)
3	Ephemerellidae (17%)	Cheumatopsyche (8%)	nr. convictum (5%)
4	Hydroptilidae (5%)	Polypedilum (5%)	pictipes (4%)
5	Psychomyiidae (4%)	Leucotrichia (4%)	flavida (4%)
	Clan	River near Webster, Wis.	
1	Heptageniidae (46%)	Stenonema (37%)	unidentified (10%)
2	Hydropsychidae (13%)	Cheumatopsyche (10%)	intercalaris (7%)
3	Baetidae (10%)	Baetis (7%)	azteca (4%)
4	Elmidae (8%)	Stenelmis (4%)	unidentified (2%)
5	Leptophlebiidae (6%)	Hyallela (4%)	unidentified (2%)

Appendix 7. Top five dominant benthic-invertebrate families and genera found in selected St. Croix River tributaries, Wisconsin and Minnesota—Continued [(xx%), percent of the total sample that were of a specific family, genera, or species]

Rank	Family	Genus	Species
		Wood River	
1	Baetidae (32%)	Plauditus (17%)	punctiventris (16%)
2	Simuliidae (17%)	Simulium (16%)	unidentified (11%)
3	Hydropsychidae (13%)	Cheumatopsyche (11%)	vittatum (11%)
4	Chironomidae (13%)	Hyallela (8%)	azteca (8%)
5	Talitridae (8%)	Brachycentrus (7%)	numerosus (7%)
	Kinnickini	nic River near River Falls, Wis.	
1	Baetidae (26%)	Optioservus (16%)	punctiventris (9%)
2	Elmidae (16%)	Baetis (14%)	tricaudatus (8%)
3	Chironomidae (12%)	Simulium (10%)	tuberosum (7%)
4	Simuliidae (11%)	Plauditus (9%)	flavistriga (6%)
5	Heptageniidae (9%)	Stenonema (8%)	pseudolimnaeus (6%)
	Trade F	River near Trade River, Wis.	
1	Chironomidae (21%)	Ephemerella (16%)	subvaria (11%)
2	Ephemerellidae (18%)	Optioservus (14%)	unidentified (8%)
3	Elmidae (15%)	Protoptila (8%)	sparna (7%)
4	Glossosomatidae (9%)	C. (cricotopus) (8%)	flavistriga (5%)
5	Hydropsychidae (8%)	Ceratopsyche (7%)	trifascia group (5%)
	Apple	River near Somerset, Wis.	
1	Hydropsychidae (29%)	Chimarra (20%)	obscura (20%)
2	Philopotamidae (20%)	Ceratopsyche (11%)	unidentified (10%)
3	Chironomidae (15%)	Cheumatopsyche (10%)	pictipes (6%)
4	Heptageniidae (9%)	Stenonema (8%)	trivittatus (4%)
5	Hydroptilidae (7%)	Leucotrichia (6%)	unidentified (4%)