

Levels of Methyleugenol in a Subset of Adults in the General U.S. Population as Determined by High Resolution Mass Spectrometry

Dana B. Barr,¹ John R. Barr,¹ Sandra L. Bailey,¹ Chester R. Lapeza, Jr.,¹ Michelle D. Beeson,¹ Samuel P. Caudill,¹ Vincent L. Maggio,¹ Arnold Schechter,² Scott A. Masten,² George W. Lucier,² Larry L. Needham,¹ and Eric J. Sampson¹

¹Division of Laboratory Sciences, National Center for Environmental Health, Centers for Disease Control and Prevention, Atlanta, Georgia, USA; ²Environmental Toxicology Program, National Institute of Environmental Health Sciences, National Institutes of Health, Research Triangle Park, North Carolina, USA

We developed a sensitive and accurate analytical method for quantifying methyleugenol (ME) in human serum. Our method uses a simple solid-phase extraction followed by a highly specific analysis using isotope dilution gas chromatography–high resolution mass spectrometry. Our method is very accurate; its limit of detection is 3.1 pg/g and its average coefficient of variation is 14% over a 200-pg/g range. We applied this method to measure serum ME concentrations in adults in the general U.S. population. ME was detected in 98% of our samples, with a mean ME concentration of 24 pg/g (range < 3.1–390 pg/g). Lipid adjustment of the data did not alter the distribution. Bivariate and multivariate analyses using selected demographic variables showed only marginal relationships between race/ethnicity and sex/fasting status with serum ME concentrations. Although no demographic variable was a good predictor of ME exposure or dose, our data indicate prevalent exposure of U.S. adults to ME. Detailed pharmacokinetic studies are required to determine the relationship between ME intake and human serum ME concentrations. *Key words:* mass spectrometry, methyleugenol, reference range, serum. *Environ Health Perspect* 108:323–328 (2000). [Online 22 February 2000]

<http://ehpnet1.niehs.nih.gov/docs/2000/108p323-328barr/abstract.html>

Methyleugenol (ME), or 4-allyl-1,2-dimethoxybenzene (Figure 1A), is a compound that occurs naturally in clove oil, nutmeg, allspice, walnuts, and a variety of other spices and herbs (1). Currently, ME is approved by the U.S. Food and Drug Administration for use in foods either as a component of a natural product additive or as a food additive itself. ME is commonly used in its natural and synthetic forms as a flavoring agent in dessert foods, an attractant in insecticides, and a fragrance in perfumes and soaps (2). More than 30,000 kg ME is used per year by the food, perfume, and pesticide industries in the United States (3). Some commercial products that may contain ME include ice cream, cookies, pies, candy, soft drinks, chewing gum, gingerbread, eggnog, pâtés, ketchup, chutney, apple butter, cigarettes, potpourri, perfumes, and insecticides. As a flavorant and fragrance, ME is used in commercial products at concentrations ranging from 5 to 52 ppm and 0.002 to 0.3%, respectively (2). It has been estimated that the average human consumes approximately 6 µg ME/day (1).

Because of the structural similarity of ME to other carcinogenic allylbenzene flavorants such as safrole and estragole (Figure 1B and C, respectively), attention has been focused on the carcinogenic potential of ME. Miller et al. (1) dosed mice with 4.75 µmol (846 µg) ME administered by intraperitoneal injections from 1 to 22 days after birth (1). Hepatic tumors were found

in 70 and 96% of the mice sacrificed after 13 months and between 13 and 18 months, respectively. More recently, data obtained by the National Toxicology Program at the National Institute of Environmental Health Sciences (NIEHS) clearly implicated ME as a rodent carcinogen (4,5). In a 2-year study with doses given 5 days/week, liver neoplasms and other tumors were observed in rats administered ME by oral gavage (5).

Although ME toxicity has been studied in laboratory animals (1,3,4–6), little or no information is available on human exposure and possible adverse health outcomes. Both toxicologic and human exposure data are needed to make accurate risk evaluations. The National Center for Environmental Health (NCEH) at the Centers for Disease Control and Prevention (CDC) in Atlanta, Georgia, and the NIEHS are collaborating to acquire these data. NCEH investigators are characterizing human exposure to a variety of environmental chemicals by measuring internal doses (the concentration of that chemical, its primary metabolite, or reaction product in a human specimen) in the general population to determine the so-called reference range for a chemical in a population (7). Such ranges supply information about the prevalence of exposure to selected chemicals and the background concentration range found in humans. These ranges also serve as a basis for trend studies, which are designed to determine whether human exposure to a given chemical is increasing or decreasing over a given time

period. For toxicologic purposes, these reference ranges can help prioritize chemicals for testing. For example, chemicals found in a high proportion of the population or in high concentrations in certain segments of the population could be given a high priority for these toxicologic studies (8,9). In the case of ME, however, the toxicologic studies were performed first, and they indicated significant carcinogenic activity (5). Efforts then focused on human exposure to ME and the comparative pharmacokinetics of ME in humans and rodents. To our knowledge, these are the first data reported on the levels of ME or its metabolites in the blood or urine of humans.

The literature detailing the measurement of ME or its metabolites in biologic matrices is scant. In support of NIEHS toxicokinetic studies that identified ME as a multisite rodent carcinogen, Graves and Runyon (3) developed a method for measuring ME in denatured rat plasma using high-performance liquid chromatography (HPLC) with ultraviolet (UV) detection. The range of quantification of the method was 50 µg/L to 10 mg/L, with coefficients of variation (CVs) ranging from 0.5% at the high calibration end to 12.5% at the low calibration end. Fischer and Dengler (10) reported a more sensitive HPLC–UV method for the analysis of a similar compound, eugenol, in bile, urine, and serum after a hexane or C₁₈ solid phase extraction (SPE). This method had limits of detection (LODs) of 2, 10, and 10 µg/L in urine, serum, and bile, respectively, and CVs

Address correspondence to D.B. Barr, Centers for Disease Control and Prevention (CDC), 4770 Buford Highway NE, Mailstop F17, Atlanta, GA 30341 USA. Telephone: (770) 488-7886. Fax: (770) 488-4609. E-mail: dlb1@cdc.gov

We thank those individuals from the National Center for Health Statistics at the CDC who participated in the planning and implementation of the Third National Health and Nutrition Examination Survey. We thank A. Woolfitt, S. Stanfill, D. Ashley, E. Smith, D. LaVoie, and E. Gunter for technical assistance. We also thank C. Smith and M. Cunningham for reviewing this manuscript.

The use of trade names is for identification purposes only and does not constitute endorsement by the Public Health Service, the Department of Health and Human Services, or the CDC.

Received 11 February 1999; accepted 29 October 1999.

of < 4%. Although both of these methods may have adequate sensitivity for measuring ME concentrations in biologic media from dosed animals, they lack both the sensitivity and selectivity for measuring trace levels of ME in biologic samples that result from everyday human exposures.

We developed a sensitive and accurate method for quantifying ME in human serum. Our method uses a simple solid-phase extraction followed by a highly specific analysis using isotope dilution gas chromatography–high resolution mass spectrometry (GC–HRMS). As a part of a CDC/NIEHS collaborative effort (including detailed pharmacokinetic studies), we applied this method to measure serum concentrations of ME in the general U.S. population.

Materials and Methods

Materials. We obtained ME and 3',4'-(methylenedioxy)-acetophenone (MDA) (Figure 1D), the recovery standard, from Aldrich Chemical Co. (Milwaukee, WI). ME and MDA had purities of 99 and 98%, respectively. Allyl- $^{13}\text{C}_3$ methyleugenol ($^{13}\text{C}_3$ -ME) of 99% isotopic purity was synthesized by Cambridge Isotope Laboratories (Andover, MA). We purchased formic acid (98%) and anhydrous sodium sulfate from EM Industries (Gibbstown, NJ) and Mallinkrodt Chemical Co. (Paris, KY), respectively. All solvents were analytical grade and were purchased from Burdick and Jackson (Muskegon, MI). We used all of the chemicals and solvents without further purification. All reagents were made daily with bioanalytical grade I water, which we prepared in-house using a 0.2- μm water filtration system (Millipore, Bedford, MA) and an Organic-Pure ultraviolet light treatment reservoir (Barnstead, Newton, MA).

Analytical standards. Individual stock solutions were prepared by dissolving 5-mg amounts of ME, $^{13}\text{C}_3$ -ME, and MDA (Figure 1D) in 100 mL toluene. We created

seven standard sets (0.2, 0.4, 2, 4, 10, 40, and 100 $\text{pg}/\mu\text{L}$) to encompass the entire linear range of the method. The ME concentration in each standard set was varied, but the $^{13}\text{C}_3$ -ME and MDA concentrations were kept constant at 40 $\text{pg}/\mu\text{L}$. The standard sets were divided into aliquots and stored at -20°C until used.

Internal standard. We prepared an internal standard spiking solution by diluting the stock $^{13}\text{C}_3$ -ME solution with acetonitrile to a concentration of 20 $\text{pg}/\mu\text{L}$. The concentration was set to obtain the most accurate analysis possible while maintaining a repeatable analytical signal.

Recovery standard and diluent. We made a recovery standard/diluent solution by diluting the MDA stock standard with toluene to a concentration of 40 $\text{pg}/\mu\text{L}$. This standard, which we added as the final step during sample preparation, had a dual purpose. The toluene served as a keeper to prevent the extract from completely evaporating. The MDA was the standard against which $^{13}\text{C}_3$ -ME recovery in individual samples was determined.

Quality control materials. We prepared quality control (QC) materials from residual sera from multiple donors. We purchased the sera from the local Red Cross. Sera were combined and well mixed. Particles > 0.2 μm were filtered from the pooled serum using a sterile filtration apparatus. We split the filtered serum into three equal volume pools. One pool was not enriched, and therefore reflected the native or endogenous concentration of ME in the serum. The other two pools were enriched with different levels of ME. Thus, we obtained QC pools with native low (≈ 100 pg/g) and high (≈ 250 pg/g) ME concentrations. After enrichment, all pools were mixed for 24 hr under refrigeration. We dispensed serum from each pool into vials in 4-mL aliquots. The vials were capped, labeled, and stored at -20°C until their use. We determined the mean concentration and the analytic variance by the repeat measurement of at least 20 samples in different analytical runs for each QC pool. A QC run was unacceptable if either the QC sample result for the current run was outside the upper or the lower 99% control limit, or the QC sample results for the current and most recent previous run were both outside the same upper or lower 95% control limit.

Reagent blanks. Because virtually all serum samples that we tested had detectable levels of ME, reagent blanks consisted of 4 mL freshly prepared water. The blank contained the same water used in the daily preparation of reagents. We prepared the reagent blanks in the same manner as the unknown samples.

Sample preparation. We prepared unknown serum samples, QC materials, and reagent blanks identically. All sera, reagents, and standards were brought to room temperature. We weighed a 4-g aliquot of serum into a test tube. The serum was spiked with 400 pg $^{13}\text{C}_3$ -ME as an internal standard, mixed, and allowed to equilibrate for approximately 5 min. The serum proteins were denatured with 4 mL 50% formic acid. We passed the denatured serum through a preconditioned Empore C_{18} SPE column (3M, Harbor City, CA) and then discarded it. The SPE column was washed with 2 mL purified water and eluted with 5 mL methylene chloride. The eluate was passed through a 500-mg silica gel SPE column topped with approximately 1 g anhydrous sodium sulfate and then collected. We rinsed the column with 2 mL methylene chloride and collected and combined the rinse with the sample. A TurboVap evaporator (Zymark Corporation, Hopkinton, MA) set at 37°C and 15 psi head pressure of nitrogen concentrated the extract to approximately 300 μL . We transferred the concentrate to a 1-mL conical vial. We added a 10- μL aliquot of the recovery standard/diluent to the vial and then allowed the sample to evaporate to approximately 10 μL at ambient temperature. We capped the vial and stored it under refrigeration until analysis.

Instrumental analysis. We analyzed 2 μL of the concentrated extract using splitless injection GC–HRMS. We performed the analyses using an HP 5890 or HP 6890 gas chromatograph (GC; Hewlett Packard Co., Wilmington, DE) interfaced with a VG250/70S or a VG70SE mass spectrometer (MS; Micromass, Manchester, UK) with Opus operating software (version 3.5, Micromass) and equipped with a low-energy (30 eV) electron impact ionization source. We achieved separation on a 30-m J & W DB-5MS [(5% phenyl)-methyl polysiloxane, 0.25- μm film thickness, 0.25-mm id] capillary column (J&W Scientific, Folsom, CA). We used helium with a linear velocity of 35 cm/sec as the carrier gas. The injector and transfer line temperatures were 260°C . The initial column temperature, 80°C , was held for 1 min, increased to 122°C at $3^\circ\text{C}/\text{min}$, then increased to 272°C at $30^\circ\text{C}/\text{min}$, and held for 1 min. We operated the MS in single ion monitoring (SIM) mode with an initial accelerating voltage of 7,000 and a 10,000 resolution, as defined at 10% valley. We used perfluorokerosene (PFK) ions as lock masses.

We monitored two ions for ME: one for quantifying and one for confirming the presence of ME. One ion each was monitored for $^{13}\text{C}_3$ -ME and MDA. Table 1 shows the ions monitored in each channel, the channel times, and the interchannel delay times. We recorded the appropriate

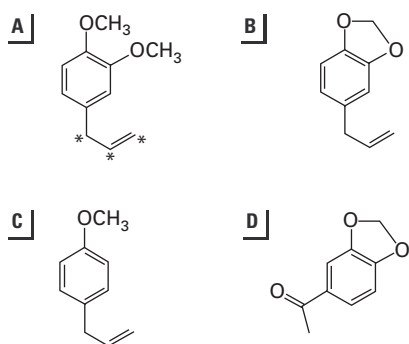


Figure 1. The chemical structures of (A) ME, (B) safrole, (C) estragole, and (D) methylenedioxyacetophenone. *Carbon atoms that were labeled with ^{13}C for the ME internal standard.

analysis specifications in an acquisition program initiated immediately after the injection of the sample into the GC. The total analysis time per sample was approximately 20 min.

Data processing and analysis. Data were automatically processed using OpusQuan software (version 6.1; Micromass), which was supplied with the mass spectrometers. The detection threshold and baseline were both set at 0% in OpusQuan; the peak differential was 3, and the minimum peak width was 1. In addition, the background signal was subtracted and all data were smoothed. The retention times and areas were automatically entered into an R:BASE database (Microrim, Redmond, WA) and the ratios of the quantification and confirmation ions were calculated. Because of the specificity of HRMS, interferences were rare; however, any interferences that occurred were easily recognized because of a dramatic shift in the ratio of the areas of the quantification and confirmation ions. In these instances, the data were deemed unacceptable and the analysis was repeated.

Quantification. We constructed calibration curves with seven ME concentrations plotted against the response factors. We calculated the response factors as the area of the ME quantification ion divided by the area of the $^{13}\text{C}_3$ -ME ion. At least five repeat determinations were performed for each concentration on the calibration curve.

Calibration standard concentrations encompassed the entire linear range of the analysis. The lowest standard concentrations were at or below the LOD to ensure linearity and accuracy at the low concentration end. A linear regression analysis of the calibration

plot provided a slope and intercept from which unknown sample concentrations could be determined. The intercept was not statistically different from zero.

Method validation. We calculated the analytical LOD for the method as $3s_0$, where s_0 was the average signal in the blanks. For the instrument LOD, we estimated s_0 as the y -intercept of a linear regression analysis of a plot of the absolute standard deviation versus the concentration (11). To evaluate ME recovery, we used 20 serum samples whose endogenous ME concentrations were well characterized. Before extraction, we spiked four samples with ME to a final concentration of 20 pg/g and four to 100 pg/g; four were not spiked. The samples were extracted as previously described. Control samples were extracts of the unspiked serum spiked after extraction with ME to final concentrations of 20 and 100 pg/g. The extracts of all samples were spiked with the internal standard to correct for instrumental variation during analysis. We determined ME recovery at each concentration by comparing the spiked samples to the control samples. Additionally, we determined the recovery of $^{13}\text{C}_3$ -ME of each individual sample by referencing the area of the $^{13}\text{C}_3$ -ME ion to the area of the MDA ion.

We determined the method accuracy by enriching serum samples with a known amount of ME, preparing and analyzing the samples, and then comparing the calculated and the expected ME concentrations. We performed a linear regression analysis on a plot of the calculated concentration versus the expected concentration. With this analysis, a slope of 1.0 would be indicative of 100% accuracy.

Reference range determination. Using this method, we determined the range of ME in a subset of serum samples collected from human adults who participated in the Third National Health and Nutrition Examination Survey (NHANES III). NHANES III was conducted between 1988 and 1994 by the National Center for Health Statistics (NCHS/CDC). All protocols were reviewed and approved by a human subjects review committee and complied with all national and institutional guidelines for the protection of human subjects. NHANES III was designed to accurately represent the U.S. civilian noninstitutionalized population; however, the serum specimens used in our study were a convenience sample of the residual NHANES III specimens and were not necessarily representative of the U.S. population. However, the samples analyzed in our study were from adults who represented a diverse spectrum of sex, age, race and ethnicity, urban and rural residences, and geographic location variables. We obtained questionnaire data from each participant. These data encompassed a variety of topics ranging from dietary intake to health status. The questionnaire data that were considered potentially important factors affecting serum ME concentrations were used in the statistical analysis and interpretation of the serum ME data. All ME data were log-transformed before analysis using univariate, bivariate, and multivariate procedures. We considered data statistically significant when $p < 0.05$. All data analyses were performed using SAS statistical software (SAS Institute, Cary, NC).

Results and Discussion

The lack of ionizable functional groups on the allylbenzene carbon skeleton of ME (Figure 1A) facilitated its simple and efficient extraction from the serum matrix using a C_{18} SPE sorbent. We added a second SPE column in the extraction procedure to further clean the sample. This silica gel cleanup of the serum extract removed coextracted compounds with polar functional groups and also removed residual water. In repeated recovery experiments at three ME concentrations, the total recovery of ME from serum was essentially quantitative. In addition, the recovery of $^{13}\text{C}_3$ -ME, which was determined independently for each sample, was consistently $> 90\%$.

The addition of an isotopically labeled standard ($^{13}\text{C}_3$ -ME) before sample manipulation, a technique known as isotope dilution (12), afforded us many advantages. Chemically, $^{13}\text{C}_3$ -ME behaves almost identically to ME, but they are distinguishable based on the 3 atomic-mass-unit (amu) difference in their masses and respective fragment ions (Figure 2). For this reason, the ratio between their ions can internally correct

Table 1. Characteristics of analysis by GC–HRMS.

Analyte	Ion	Mass assignment	Channel time (msec)	Interchannel delay time (msec)	Ion use
ME	178.0994	$[\text{M}]^+$	160	20	Quantification
ME	163.0759	$[\text{M}-\text{CH}_3]^+$	160	20	Confirmation
PFK	168.9888	–	50	20	Lock mass
ME- $^{13}\text{C}_3$	181.1094	$[\text{M}]^+$	160	20	Quantification
MDA	164.0473	$[\text{M}]^+$	160	20	Recovery

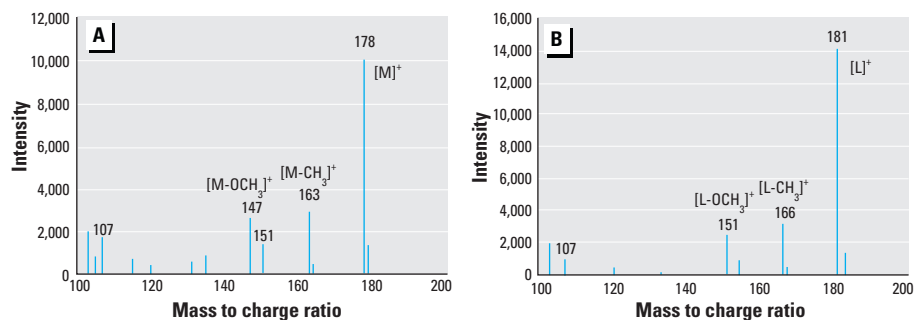


Figure 2. Electron impact mass spectra of (A) ME and (B) $^{13}\text{C}_3$ -ME with important mass assignments. M and L refer to the molecular ion of the native and labeled ME, respectively. The fragmentation of the two compounds is nearly identical with only a 3-amu difference in the predominant ions. Only fragments with masses > 100 amu were monitored. Smaller fragments (< 100 amu) were in a region with a higher background signal, therefore increasing the potential for interferences.

for recovery of ME in each individual sample, which eliminates the need for recovery surrogates, although a surrogate can still be used for validation purposes. The automatic recovery correction reduces the error associated with the measurement and ultimately increases the method sensitivity.

During GC–HRMS analysis of ME with low energy (30 eV) electron impact ionization (EI), the predominant ions were $[M]^+$, $[M-CH_3]^+$, and $[M-OCH_3]^+$ at mass to charge ratio (m/z) of 178, 163, and 147, respectively (Figure 2). As expected, higher energy (70 eV) EI produced more fragmentation, especially lower molecular weight fragment ions, and chemical ionization using methane as a reagent gas produced only one ion at m/z 179, $[M+H]^+$. Because the controlled fragmentation at 30 eV resulted in higher molecular weight ions in a region with fewer background ions and ions of higher intensity, we opted to use low energy EI for the analysis. Under these conditions, the $^{13}C_3$ -ME formed similar ions as ME in the same relative abundances except that the m/z of the ions were 3 amu greater (Figure 2). This confirmed that the fragment losses were not from the allyl group that was labeled with ^{13}C atoms. We used the MDA as a recovery standard because of its structural similarity to ME and because its predominant ions were within a few atomic mass units of ME ions, as is necessary for high-resolution SIM analysis. MDA eluted from the gas chromatography column approximately 2 min after ME, thus allowing their ions to be monitored in separate windows of time (Figure 3).

The instrument LOD was 282 fg on-column. With the method recovery, this translates to approximately 350 fg/g serum (parts per quadrillion); however, ME was endogenous in the air and water. Using a solvent trap on the vacuum system or pulling excess air through the SPE columns caused a low-level ME contamination of approximately 5 pg/g (parts per trillion). By removing the solvent trap and carefully monitoring the volume of air pulled through the columns, we were able to reduce the contamination to approximately 1 pg/g or lower. Additionally, we used in-house purified water in reagent preparation because bottled water and distilled water contained higher endogenous ME levels. As a result, the method LOD was 3.1 pg/g. Thereafter, an occasional contaminant appeared, but it was readily apparent in the quality control samples.

Figure 4 shows a calibration curve. The ME analysis was linear over 3 orders of magnitude; $r^2 = 0.997$. We obtained similar calibration curves on multiple high-resolution instruments.

The method's accuracy was essentially 100%. A linear regression analysis of a plot

of the calculated concentrations of spiked samples versus the expected concentrations of the same samples (Figure 5) yielded a slope of 0.997, which is indicative of a high degree of accuracy. We used a similar plot to compare data from multiple instruments; the plot yielded a slope of 0.996, signifying good agreement among instruments.

A typical quality control Shewart plot is shown in Figure 6. This plot includes samples analyzed on multiple instruments and reflects both intra- and interday variation. The overall CV and the intra- and interday variations at three concentrations over the linear range of the method are shown in Table 2. As expected, the variation among days was a greater contributor to the overall CV than the variation within days. Additionally, the variation was greater as the

concentration approached the method LOD; although this is a normal occurrence, the low-level contamination of ME probably resulted in increased variation at the low concentration end.

Overall, the data from the QC materials proved that ME was stable in serum over the testing period of approximately 2 months. We did not conduct stability studies over longer periods of time. However, our data indicate that NHANES III samples had ME concentrations comparable to freshly collected serum from volunteers. These data suggest that ME is stable in frozen serum stored for up to 5 years.

Our method is more sensitive and more selective than the only published method (3); for example, the LOD of our method is > 4 orders of magnitude lower than the

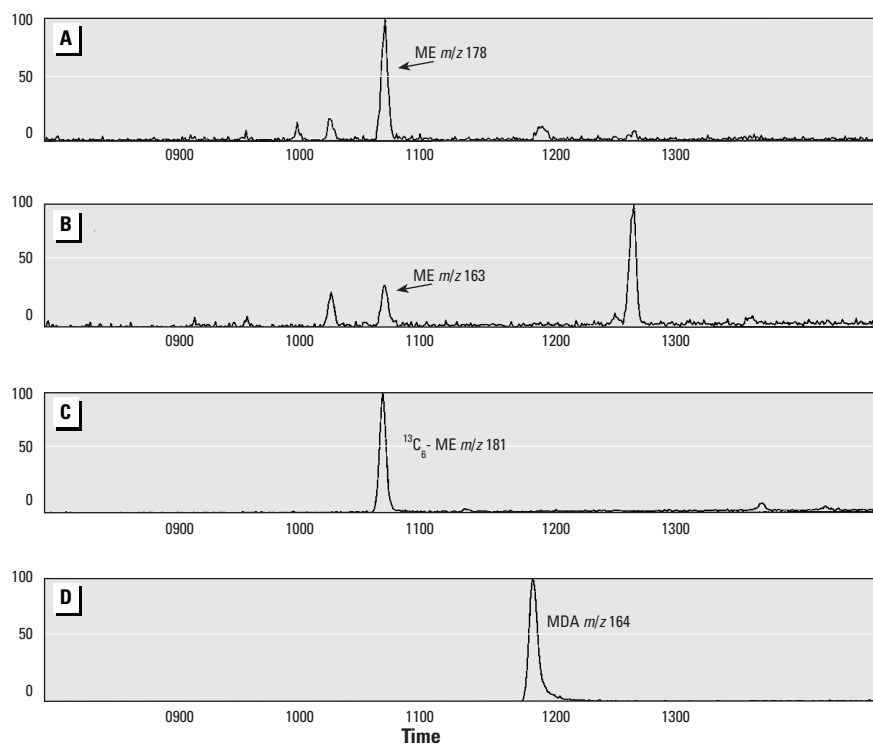


Figure 3. Ion chromatograms of a 0.2-pg/ μ L standard show the clean separation of the components. The (A) quantification, (B) confirmation, and (C) internal standard ions of ME. (D) The MDA recovery standard.

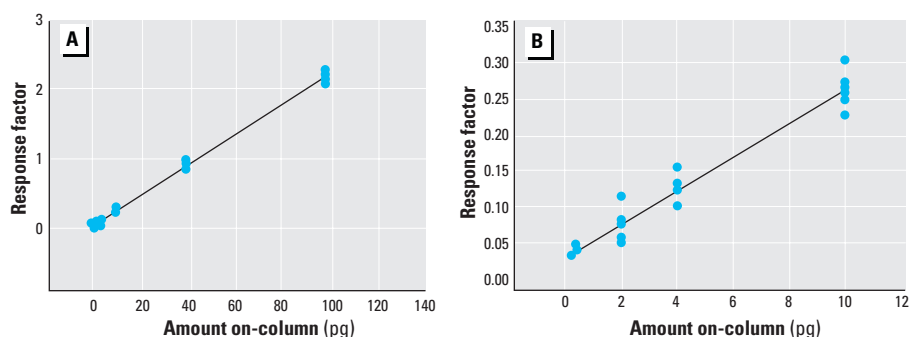


Figure 4. (A) A calibration curve for ME over a 100-pg range. $r^2 = 0.997$. (B) The low concentration end demonstrates the linearity even near the LOD.

LOD of the HPLC–UV method. Some of the sensitivity differences between the two methods can be attributed to the decreased sample size for the HPLC–UV method. The Graves and Runyon (3) method uses only 200 μ L plasma, whereas our method requires 4 g serum (the approximate amount of serum from one 10-mL blood draw). Because our method uses HRMS, it provides a greater selectivity than the HPLC–UV method; this selectivity probably accounts for a large portion of the increase in method sensitivity. In addition, our sample preparation provides a much cleaner extract, which can help minimize interfering serum components. The CVs of our method at very low serum concentrations (Table 2) are comparable to the CVs (12–12.5%) of the HPLC–UV method at 50 mg/L plasma (3). At higher plasma concentrations, the HPLC–UV method reports CVs ranging from 1–5%.

The specificity of HRMS at 10,000 resolution was required to eliminate interfering components in the human serum extracts that in turn provided the low detection limits

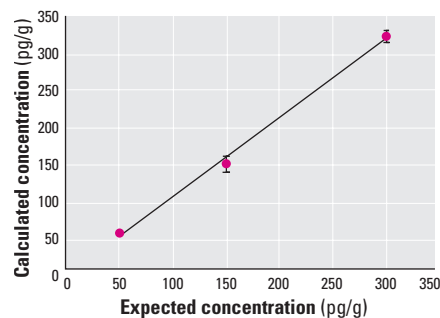


Figure 5. Plots of the calculated ME concentration from spiked serum sample extracts against the expected ME concentration. The slope of the resultant linear regression analysis was 0.997, indicating a high degree of accuracy in the calculation.

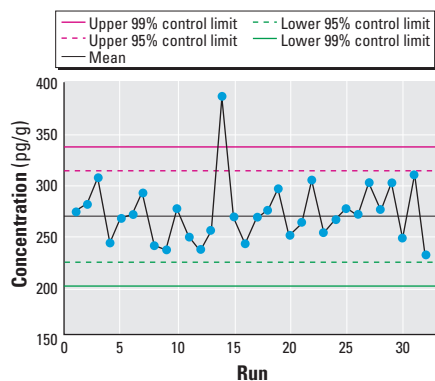


Figure 6. A Shewart plot of quality control materials demonstrates the precision of our method. The mean is 255 pg/g, with a total CV of 8.9%. The outlier on this plot indicates an occurrence of contamination (i.e., blank samples that had high ME concentrations).

of the method. Analysis at lower resolutions resulted in recurring interferences, as evidenced by significant changes in the ratios of the quantification and confirmation ions. These specificity requirements precluded the use of single quadrupole or other low-resolution mass spectrometers. However, we did evaluate the effectiveness of GC–MS (SIM mode) and GC–tandem mass spectrometry (MS/MS) for ME analysis using a quadrupole ion trap with external ionization (GCQ; Finnigan MAT, San Jose, CA). The instrument LODs in the SIM and MS/MS modes were 50 and 3.4 pg on-column, respectively. Although the SIM mode did not provide the sensitivity required for ME analysis in serum, the MS/MS LOD was adequate. The calculated concentrations of QC materials analyzed using both the GC–HRMS and GC–MS/MS analyses are shown in Table 3. The GC–HRMS clearly provided more accurate and precise data.

Table 4 shows a summary of the serum ME concentrations (whole-weight basis) in 206 adult participants of NHANES III. The distribution of the data (Figure 7) was not altered when the individual ME concentrations were adjusted for total lipid content in each serum sample. The frequency of ME detection was 98%, which verified that our analytical method had adequate sensitivity to detect incidental exposure to ME and that

Table 2. Variation in ME analyses.

Parameter	Concentration		
	Low (12 pg/g)	Mid (130 pg/g)	High (225 pg/g)
Total CV (%)	21	13	8.9
Interday CV (%)	22	12.5	9.3
Intraday CV (%)	9.6	7.5	7.2
Replicates (n)	69	60	48

Table 3. Calculated concentrations in spiked quality control pools using GC–HRMS and GC–MS/MS.

Pool	Concentration	
	GC–HRMS (pg/g)	GC–MS/MS (pg/g)
Low concentration	12 (21%)	Not quantifiable
Mid concentration	130 (13%)	122 (63%)
High concentration	225 (9%)	360 (19%)

CVs are shown in parentheses.

Table 4. ME distribution in serum of adult participants in NHANES III.

Statistic	Value
Adults (n)	206
Detection frequency	98%
Mean	24 pg/g
Median	16 pg/g
Minimum	ND
Maximum	390 pg/g
5th percentile	5 pg/g
95th percentile	78 pg/g

ND, not detected or < 3.1 pg/g.

the population as a whole is exposed to ME to some extent. Only four individuals had ME concentrations below the LOD (< 3.1 pg/g). Five individuals had serum ME levels \geq 100 pg/g, which is approximately 2 times the average peak serum ME concentrations observed in fasting adults fed a meal containing approximately 60 μ g ME (4).

Bivariate and multivariate analyses of the data using the demographic information proved only marginally significant at best. Males who fasted for > 9 hr before providing a serum sample had, on average, lower ME levels than those who fasted for \leq 9 hr (Table 5). Females displayed the opposite trend. The interaction between sex and fasting status was marginally significant ($p = 0.0449$). Even though the interaction was marginally significant, further testing did not indicate any significant differences among the groups. Non-Hispanic blacks had lower average ME levels than non-Hispanic whites, Mexican Americans, and other ethnic groups (Table 5); however, the differences were only nominally significant ($p = 0.0847$).

Differences among the other demographic variables were nonsignificant; however, some general trends were interesting and worth noting. Smokers had slightly higher ME levels than nonsmokers. In general, people living with nonsmokers had lower ME levels than those living with one smoker; both of these groups had lower ME levels than people living with two or more smokers. Other data demonstrating nonsignificant trends are shown in Table 5.

Although the demographic information did not provide data that would help predict which factors were strong determinants of ME exposure in the general U.S. population, the data provided interesting information. Most of the trends that we observed were not surprising. For instance, because ME is a flavoring agent in many commercial products, it seems plausible that smoke resulting from commercial cigarettes could contribute, in part, to serum ME levels. Although the differences were not significant, the fact that these trends existed warrants further investigation.

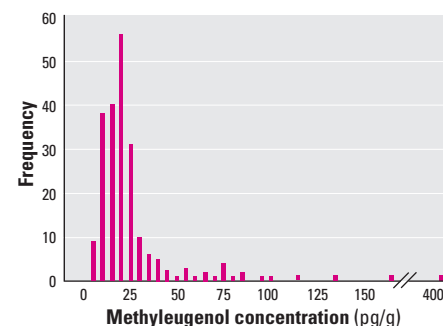


Figure 7. Distribution of ME in the serum of NHANES III participants.

Table 5. Geometric mean (GM) concentration and distribution characteristics of serum ME in selected demographic categories.

Demographic group	No. ^a	ME concentrations (pg/g)					Max
		GM	Min	Percentile			
				25th	50th	75th	
Male							
All	64	17	3.9	12	16	27	110
Fasting ≤ 9 hr	23	19	4.3	13	17	31	110
Fasting > 9 hr	41	16	3.7	11	15	24	84
Female							
All	142	16	2.2	10 ^b	17	22	390
Fasting ≤ 9 hr	56	14	2.2	10 ^b	16	20 ^b	54
Fasting > 9 hr	86	18	2.2	10 ^b	18	24	390
Race/ethnicity							
Non-Hispanic white	53	19	4.9	12	18	29	84
Non-Hispanic black	96	14	2.2	8.8	15	22	390
Mexican American	41	19	5.4	14	18	23	110
Other	16	16	5.2	13	17	19	78
Age group							
18–26 years	48	16	2.2	13	17	22	390
27–34 years	54	16	4.1	10 ^b	15	22	95
35–42 years	43	16	2.2	11	17	22	110
43+ years	61	17	2.2	11	18	24	160
Smoking exposure status							
Nonsmoker and no smokers in home	107	5.6	2.2	9.5	16	22	390
Smoker or living with 1 smoker	69	17	2.2	11	15	23	160
Smoker or living with ≥ 2 smokers	28	19	4.9	12	16	29	110
Drinks bottled water							
Yes	29	19	5.8	14	20 ^b	24	95
No	175	16	2.2	10 ^b	15	22	390
Diet							
≤ 1 cake/cola per day	114	16	2.2	11	17	23	160
> 1 cake/cola per day	92	17	2.2	11	16	22	390

Abbreviations: max, maximum; min, minimum.

^aTotal number in demographic group. ^bZero is a significant digit.

The ME exposure data in adults of the general U.S. population provide invaluable information. Considering that in male and female rats, oral ME doses of 75–150 mg/kg resulted in peak average plasma ME concentrations ranging from 1.5 to 8.2 mg/L (1.5–8.2 ppm) and that these same dose levels clearly induced malignant lesions in rats and mice on a chronic daily basis (5), the human serum ME concentrations observed in our study population suggest that more extensive risk evaluation is needed. Assuming that humans respond similarly to rodents, these human ME data coupled with hazard identification data (i.e., Does a chemical represent a potential health hazard?), comparative human and rodent pharmacokinetic data, and rodent dose–response data should provide adequate information to characterize human risk of cancer as well as noncancer health effects resulting from ME exposure.

It is important to emphasize that these human data are from adults only. Considering the potential sources of ME exposure and children's small size, it is likely that children would have higher concentrations of ME. Exposure and risk assessment in children is

an extremely important area in which to focus future ME studies.

Conclusions

We developed a highly specific, accurate, and sensitive method for the measurement of ME in human serum using isotope dilution GC–HRMS. We applied this method to analyze ME concentrations in the serum of adult participants in NHANES III. The high frequency of ME detection in this population verified that our method possessed adequate sensitivity for ME analysis in the general population. The data also indicated that exposure to ME in the U.S. population is prevalent. Because no potential determinants of exposure were significantly correlated with ME concentrations, we surmise that the total ME exposure is from a variety of sources. Substantial serum concentrations of ME in fasting adults, coupled with ubiquitous low levels of ME in air and water, suggest that ME exposure results from foods where ME is not intentionally added as well as from nonfood sources. These human data coupled with hazard identification and rodent dose–response data provide the

necessary information for proper human risk assessment for ME.

REFERENCES AND NOTES

- Miller EC, Swanson AB, Phillips DH, Fletcher TL, Liem A, Miller JA. Structure-activity studies of the carcinogenesis in the mouse and rat of some naturally occurring and synthetic alkenylbenzene derivatives related to safrole and estragole. *Cancer Res* 43:1124–1134 (1983).
- Burdock GA, ed. *Fenaroli's Handbook of Flavor Ingredients*. Boca Raton, FL: CRC Press, 1995.
- Graves SW, Runyon S. Determination of methyleugenol in rodent plasma by high-performance liquid chromatography. *J Chromatogr B* 663:255–262 (1995).
- Centers for Disease Control and Prevention and the National Institute of Environmental Health Sciences. Unpublished results.
- NTP. *Toxicology and Carcinogenesis Studies of Methyleugenol (CAS No. 93-15-12) in F344/N Rats and B6C3F₁ Mice (Gavage Studies)*. Peer Review Draft, TR-491. Research Triangle Park, NC: National Toxicology Program, 1998.
- Gardner I, Bergin P, Stening P, Kenna JG, Caldwell J. Immunochemical detection of covalently modified protein adducts in livers of rats treated with methyleugenol. *Chem Res Toxicol* 9(4):713–721 (1996).
- Needham LL, Patterson DG Jr, Burse VW, Paschal DC, Turner WE, Hill RH Jr. Reference range data for assessing exposure to selected environmental toxicants. *Toxicol Ind Health* 12:507–513 (1996).
- Lucier GW, Schechter A. Human exposure assessment and the National Toxicology Program. *Environ Health Perspect* 106:623–627 (1998).
- Lucier GW, Needham LL. NIEHS, CDC collaborate to improve exposure assessment. *Environ Health Lett* 37:127–128 (1998).
- Fischer IU, Dengler HJ. Sensitive high-performance liquid chromatographic assay for the determination of eugenol in body fluids. *J Chromatogr B* 525:369–377 (1990).
- Taylor JK. *Quality Assurance of Chemical Measurements*. Boca Raton, FL: CRC Press, 1987.
- Colby BN, McCaman MW. A comparison of calculation procedures for isotope dilution determinations using gas chromatography mass spectrometry. *Biomed Mass Spectrom* 6(6):225–230 (1979).