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Survival Estimates of Migrant Juvenile Salmonids through Bonneville Dam Using Radio Telemetry, 2004

Final Report of Research

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## Table of Contents

List of Figures ..... iv
List of Appendix Figures ..... vi
List of Tables ..... vii
List of Appendix Tables ..... xii
Executive Summary ..... xviii
Yearling Chinook salmon ..... xix
Paired Release-recapture Model ..... xix
Minimum Gap Runner Turbine Unit ..... xix
Ice and Trash Sluiceway ..... xix
Spillway Flow Deflectors ..... xix
Route-specific Survival Model ..... XX
Steelhead Trout ..... xxi
Paired Release-recapture Model ..... xxi
Minimum Gap Runner Turbine Unit ..... xxi
Ice and Trash Sluiceway ..... xxi
Spillway Flow Deflectors ..... xxi
Route-specific Survival Model ..... xxii
Subyearling Chinook salmon ..... xxiii
Paired Release-recapture Model ..... xxiii
Ice and trash sluiceway ..... xxiii
Spillway Flow Deflectors ..... xxiii
Route-specific Survival Model ..... xxiv
Introduction .....  1
Methods ..... 5
Study Area ..... 5
Bonneville Dam ..... 6
System antenna configuration ..... 6
Radio Transmitters ..... 10
Fish tagging and releasing ..... 10
Converting radio signals into detection histories ..... 11
Statistical methods ..... 12
Paired release-recapture model ..... 12
Model Assumptions ..... 13
Estimable Parameters ..... 15
Route-Specific Survival Model ..... 17
Model Assumptions ..... 17
Parameter Estimation ..... 18
Results ..... 21
Releases of dead radio-tagged fish ..... 21
Run timing and radio telemetry tagging dates ..... 22
Radio-tagged fish size relative to run-of-river fish ..... 25
Tag-life performance for determining potential bias of survival estimates ..... 29
River Discharge and Project Operations ..... 30
Yearling Chinook salmon ..... 33
Paired Release-recapture Model ..... 33
Minimum Gap Runner Turbine Unit Survival Estimation ..... 33
Powerhouse 1 Ice and Trash Sluiceway Survival Estimation ..... 35
Spillway Flow Deflectors ..... 36
Route-specific Survival Model ..... 41
Survival estimation ..... 41
56 kcfs day spill operations ..... 44
Total dissolved gas cap night spill operations ..... 44
Steelhead trout ..... 48
Paired Release-recapture Model ..... 48
Minimum Gap Runner Turbine Unit Survival Estimation ..... 49
Powerhouse 1 Ice and Trash Sluiceway Survival Estimation ..... 51
Spillbay Flow Deflectors ..... 52
Route-specific Survival Model ..... 57
Survival estimation ..... 57
56 kcfs day spill operations. ..... 57
Total dissolved gas cap night spill operations ..... 57
Subyearling Chinook salmon ..... 63
Paired Release-recapture Model ..... 63
Powerhouse 1 Ice and Trash Sluiceway ..... 63
Spillway Flow Deflectors ..... 66
Route-specific Survival Model ..... 73
56 kcfs day/TDG night spill operations ..... 73
23 kcfs 24 h spill operations ..... 74
Discussion ..... 79
Acknowledgements ..... 82
References ..... 83
Appendix 1: Release Dates, Times, Fork Lengths and Weights. ..... 87
Appendix 2: Dead fish analysis ..... 115
Appendix 3: Tag-Life Performance for Determining Potential Bias of Survival Estimates ..... 126
Appendix 4: Burnham Tests 2 and 3. ..... 132
Appendix 5: Homogeneity of Arrival Times ..... 152
Appendix 6: Bonneville Lock and Dam, Revision to Fish Passage Plan Spill Patterns ..... 163

## List of Figures

Figure 1. Release and detection locations for Bonneville Dam survival evaluation, during 2004. $\mathrm{R}=$ release locations, yellow ovals are locations of radio telemetry antenna arrays.

Figure 2. Plan view of aerial antenna coverage during spring 2004 at Bonneville Dam's: (a) second powerhouse (B2) and spillway (SPI); and (b) first powerhouse (B1).

Figure 3. Plan view of underwater antenna coverage during spring 2004 at Bonneville Dam's: (a) second powerhouse (B2) and spillway (SPI); and (b) first powerhouse (B1).

Figure 4. Schematic of estimable capture and survival probabilities ( $S=$ survival estimate, $p=$ capture probability, and $\lambda=\mathrm{S} \cdot \mathrm{p}$ ) from site-specific releases ( $R_{\text {ROUTE }}$ ) at Bonneville Dam and in the Bonneville Dam tailrace.

Figure 5. The estimable parameters (see p .49 for definitions) for the route-specific survival model using the proposed release and detection schemes for 2004. Included in the detection scheme is a double radio-telemetry array at Bonneville Dam.

Figure 6. Yearling Chinook salmon daily passage index at Bonneville Dam, powerhouse 2. The vertical bars represent the passage index (see: www.fpc.org) for a given day. Vertical lines represent the start and end dates for radio telemetry tagging.

Figure 7. Steelhead trout daily passage index at Bonneville Dam, powerhouse 2 The vertical bars represent the passage index (see: www.fpc.org) for a given day. Vertical lines represent the start and end dates for radio telemetry tagging.

Figure 8. Subyearling Chinook salmon daily passage index at Bonneville Dam, powerhouse 2. The vertical bars represent the passage index (see: www.fpc.org) for a given day. Vertical lines represent the start and end dates for radio telemetry tagging.

Figure 9. The distribution of fork lengths of run-of-river yearling Chinook salmon that were sampled at the John Day Dam and Bonneville Dam smolt monitoring facilities and fish tagged with MCFT-3KM radio transmitters (Lotek Engineering, Newmarket, Ontario) and released at The Dalles Dam and Bonneville Dam during 2004. Based on length to weight regression equations and tag weight to fish weight criterion, fish to the left of the dashed lines were too small to be tagged with the transmitters.

Figure 10. The distribution of fork lengths of run-of-river steelhead trout that were sampled at the John Day Dam and Bonneville Dam smolt monitoring facilities and fish tagged with MCFT3KM radio transmitters (Lotek Engineering, Newmarket, Ontario) and released at The Dalles Dam and Bonneville Dam during 2004. Based on length to weight regression equations and tag weight to fish weight criterion, fish to the left of the dashed lines were too small to be tagged with the transmitters.

Figure 11. The distribution of fork lengths of run-of-river subyearling Chinook salmon that were sampled at the John Day Dam and Bonneville Dam smolt monitoring facilities and fish tagged with MCFT-3KM radio transmitters (Lotek Engineering, Newmarket, Ontario) and released at The Dalles Dam and Bonneville Dam during 2004. Based on length to weight regression equations and tag weight to fish weight criterion, fish to the left of the dashed lines were too small to be tagged with the transmitters.28

Figure 12. The estimated survival probabilities ( $95 \%$ confidence intervals error bars) from paired releases of radio-tagged yearling Chinook salmon passing through spillbays with deflectors located at either $7-\mathrm{ft}$ or 14 - ft above mean sea level at Bonneville Dam under two spill operations ( 56 kcfs day or total dissolved gas cap(TDG) night) by release. The paired release groupings were formed post-hoc from fish released into The Dalles Dam tailrace and in the Bonneville Dam tailrace during 2004.

Figure 13. Schematic of estimated route-specific passage and survival parameters for yearling Chinook salmon through Bonneville Dam during 56 kcfs day/total dissolved gas cap night spill operations. Estimated standard errors are in parentheses.

Figure 14. Schematic of estimated route-specific passage and survival parameters for yearling Chinook salmon through Bonneville Dam during 56 kcfs day spill operations. Estimated standard errors are in parentheses.

Figure 15. Schematic of estimated route-specific passage and survival parameters for yearling Chinook salmon through Bonneville Dam during total dissolved gas cap night spill operations. Estimated standard errors are in parentheses.

Figure 16. The estimated survival probabilities ( $95 \%$ confidence interval error bars) from paired releases of radio-tagged steelhead trout passing through spillbays with either deflectors located 7 feet or 14 feet above mean sea level at Bonneville Dam under two spill operations (TDG is total dissolved gas cap), by release. The paired release groupings were formed post-hoc from fish released at The Dalles Dam and in the Bonneville Dam tailrace during 2004

Figure 17. Schematic of estimated route-specific passage and survival parameters for steelhead trout through Bonneville Dam during 2004. Estimated standard errors are in parentheses.

Figure 18. Schematic of estimated route-specific passage and survival parameters for steelhead trout passing during daytime 56 kcfs spill operations through Bonneville Dam 2004. Estimated standard errors are in parentheses.

Figure 19. Schematic of estimated route-specific passage and survival parameters for steelhead trout passing during total dissolved gas cap nighttime spill operations through Bonneville Dam 2004. Estimated standard errors are in parentheses.

Figure 20. The estimated survival probabilities ( $95 \%$ CI error bars) from paired releases of radio-tagged subyearling Chinook salmon passing through spillbays with either deflectors located at 7 feet or 14 feet above mean sea level at Bonneville Dam under the 56 kcfs day/total dissolved gas night spill operation, by release. The paired release groupings were formed posthoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004

Figure 21. The estimated survival probabilities ( $95 \%$ CI error bars) from paired releases of radio-tagged subyearling Chinook salmon passing through spillbays with either deflectors located at 7 feet or 14 feet above mean sea level at Bonneville Dam under $24 \mathrm{~h}, 23 \mathrm{kcfs}$ spill, by release. The paired release groupings were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

Figure 22. Schematic of estimated route-specific passage and survival parameters for subyearling Chinook salmon through Bonneville dam during 20 June through 22 July, during 56 kcfs day and total dissolved gas cap at night spill operations in 2004. Estimated standard errors are in parentheses.77

Figure 23. Schematic of estimated route-specific passage and survival parameters for subyearling Chinook salmon through Bonneville dam during 20 June through 22 July, during 23 kcfs 24 h spill operations in 2004. Estimated standard errors are in parentheses

## List of Appendix Figures

Figure A2.1. Schematic of estimated route-specific passage and survival parameters for yearling Chinook salmon through Bonneville Dam with the removal of fish with travel times $>$ 99.7. Estimated standard errors are in parentheses.

Figure A2.2. Schematic of estimated route-specific passage and survival parameters for yearling Chinook salmon through Bonneville Dam with the removal of fish within 24 h of detected dead fish. Estimated standard errors are in parentheses 120

Figure A2.3. Schematic of estimated route-specific passage and survival parameters using an analysis with the removal of fish contacted within 24 h of detected dead fish, for subyearling Chinook salmon through Bonneville dam during 20 June through 22 July, at 56 kcfs day and total dissolved gas cap at night spill operations in 2004. Estimated standard errors are in parentheses.

Figure A2.4. Schematic of estimated route-specific passage and survival parameters using an analysis with the removal of fish contacted within 24 h of detected dead fish, for subyearling Chinook salmon through Bonneville dam during 20 June through 22 July, at 32 kcfs spill operations in 2004. Estimated standard errors are in parentheses.

Figure A3.1. Arrival distributions of treatment fish for the, a) Bonneville Dam and b) The Dalles Dam survival assessment during summer of 2004 for subyearling Chinook salmon................. 130

Figure A3.2. Fitted Gompertz model with tag-life data for a) spring and b) summer studies... 130

Figure A3.3. Probability distributions (a-c) for radio-tags being operational over time with cumulative arrival distributions at downstream survival gates for the Bonneville Dam survival assessment during 2004

## List of Tables

Executive Summary Table 1. The estimated survival probabilities and $95 \%$ confidence intervals of radio-tagged, yearling Chinook and subyearling Chinook salmon, and steelhead trout released directly into the Minimum Gap Runner turbine unit (MGR) and ice and trash sluiceway at Bonneville Dam's powerhouse 1 and in the Bonneville Dam tailrace. Also presented are the estimated survival probabilities of fish passing via spillbays with either $7-\mathrm{ft}$ or $14-\mathrm{ft}$ spillbay deflectors at Bonneville Dam spillway estimated from the post-hoc pairing of release groups formulated from fish released at The Dalles Dam (treatment) and Bonneville Dam tailrace (control). Spring spill operations were 56 kcfs during the day and total dissolved gas cap at night ( $56 \mathrm{kcfs} / \mathrm{TDG}$ ). Two spill operations were examined in the summer, $56 \mathrm{kcfs} / \mathrm{TDG}$ and 23 kcfs for 24 hour. xxvii

Executive Summary Table 2. The route-specific survival probabilities and associated profile likelihood $95 \%$ confidence intervals for radio-tagged, yearling Chinook salmon, steelhead trout, and subyearling Chinook salmon passing through Bonneville Dam. Fish were released in The Dalles Dam and Bonneville Dam tailraces. Spring spill operations were 56 kcfs during the day and night spill until the total dissolved gas cap (TDG) of $125 \%$ was reached in the tailrace ( 56 $\mathrm{kcfs} / \mathrm{TDG}$ ). Two spill operations were examined in the summer, $56 \mathrm{kcfs} / \mathrm{TDG}$ and 23 kcfs for 24 hour from 20 June to 22 July. xxviii

Table 1. Summary of the release dates, times and travel times to the radio telemetry detection arrays (Gates) for radio-tagged dead fish detections. The percentile ranking of the dead fish, with respect to the travel time distribution for all live radio-tagged fish released into the Bonneville Dam tailrace below the powerhouse 2 juvenile bypass outfall 2004, is presented for each detection array.

Table 2. Estimated probabilities (mean, SE in parentheses) a radio-tag was operational at Bonneville Dam and the downstream detection arrays for yearling Chinook salmon, steelhead trout, and subyearling Chinook salmon, during 2004

Table 3. Descriptive statistics for discharge (kcfs) at Bonneville Dam during spring (29 April to 7 June) 2004. Values have been rounded to the nearest tenth and are based on hourly averages.

Table 4. Descriptive statistics for discharge (kcfs) during day (0500 to 1959) and night (2000 to 0459 hours) by dam area at Bonneville Dam during spring (29 April to 7 June) 2004. Dam operations were 56 kcfs spill during the day and night spill up to $125 \%$ of the total dissolved gas cap (TDG) at night.

Table 5. Descriptive statistics for discharge (kcfs) during day (0400 to 2159) and night (2200 to 0359 hours) by dam area at Bonneville Dam during summer (21 June to 22 July) 2004 during. Dam operations were either 56 kcfs day spill with night spill up to $125 \%$ of the total dissolved gas cap (TDG) at night or 32 kcfs spill (day and night).

Table 6. The estimated survival and standard error (SE) of yearling Chinook salmon released into the minimum gap runner (MGR) turbine unit at Bonneville Dam's powerhouse 1 during 2004. Treatment releases were made directly into the MGR turbine unit and two control releases were evaluated; one directly below the front roll of the turbine unit and one below the powerhouse 2 juvenile bypass system (JBS) outfall at Bonneville Dam. Survival estimates are presented for both paired-release groupings (e.g., the MGR and the control group directly below the turbine unit and MGR and the control group below the powerhouse 2 juvenile bypass outfall). The survival estimates presented are the estimated survival of the release group into the MGR to the release location of the tailrace release group. The specific dates and times of the releases can be referenced in Appendix 1.

Table 7. The estimated survival and standard error (SE) of yearling Chinook salmon released into the ice and trash sluiceway at Bonneville Dam's powerhouse 1 during 2004. Treatment releases were made into the ice and trash sluiceway and the control release was released below the powerhouse 2, juvenile bypass (JBS) outfall at Bonneville Dam. The survival estimates presented are the estimated survival of the release group into the ice and trash sluiceway to the release location of the tailrace release group. The specific dates and times of the releases can be referenced in Appendix 1.

Table 8. The estimated survival probabilities (S), standard errors (SE), $95 \%$ confidence intervals (CI), number released ( N ; number of fish), and dates associated with paired releases of radiotagged yearling Chinook salmon passing via spillbays with deflectors located 7 feet above mean sea level during 56 kcfs day (05:00-19:59 hrs) spill operations and spill at night (20:00-04:59 hrs) up to the total dissolved gas cap of $125 \%$ in the tailrace at the Bonneville Dam spillway. The paired release groups were formed post-hoc from fish released into The Dalles Dam tailrace (treatment) and in the Bonneville Dam tailrace (control) during 2004. 37

Table 9. The estimated survival probabilities (S), standard errors (SE), $95 \%$ confidence intervals $(\mathrm{CI})$, number released ( N ; number of fish), and dates associated with paired releases of radiotagged yearling Chinook salmon passing via spillbays with deflectors located 14 feet above mean sea level during 56 kcfs day (05:00-19:59 hrs) spill operations and spill at night (20:00-04:59 hrs) up to the total dissolved gas cap of $125 \%$ in the tailrace at Bonneville Dam spillway. The paired release groups were formed post-hoc from fish released into The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004. 39

Table 10. Counts of radio-tagged yearling Chinook salmon for the releases from The Dalles Dam $\left(R_{l}\right)$ and in the tailrace of Bonneville Dam $\left(R_{2}\right)$ used in the route-specific survival model during 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $R_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B1 is powerhouse 1 and B 2 is powerhouse 2 at Bonneville Dam.

Table 11. Counts of radio-tagged yearling Chinook salmon for releases from The Dalles Dam $\left(R_{l}\right)$ and in the tailrace of Bonneville Dam $\left(R_{2}\right)$ used in the route-specific survival model during two spill operations: 56 kcfs day ( $0500-2000$ ) and total dissolved gas cap (TDG) at night in 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B1 is powerhouse 1 and B 2 is powerhouse 2 at Bonneville Dam. 45

Table 12. Summary table of estimated route-specific survival probabilities (S) and their associated standard errors (SE) of yearling Chinook salmon survival through Bonneville Dam (Dam survival) generated from the route-specific survival model. The results of $Z$-tests (i.e., $Z$ statistic) structured to assess whether the estimated survival probabilities during the 56 kcfs day spill operations were different than the estimated survival probabilities during the total dissolved gas cap (TDG) night spill operations. Significant results are indicated where $\mathrm{Z} \geq 1.645$ given a two-tailed test and $\alpha=0.10$. The JBS refers to the juvenile bypass system at powerhouse 2 . .. 48

Table 13. The estimated survival (S) and standard error (SE) of steelhead trout released into the Minimum Gap Runner (MGR) Turbine Unit at Bonneville Dam's powerhouse 1 during 2004. Treatment releases were made directly into the MGR turbine unit and two control releases were evaluated; one directly below the front roll of the turbine unit and one below the powerhouse 2 juvenile bypass system (JBS) outfall at Bonneville Dam. Survival estimates are presented for both paired-release groupings (e.g., the MGR and the control group directly below the turbine unit and MGR and the control group below the powerhouse 2 juvenile bypass outfall). The survival estimates presented are the estimated survival of the release group into the MGR to the release location of the tailrace release group. The specific dates and times of the releases can be referenced in Appendix 1. 50

Table 14. The estimated survival and standard error (SE) of steelhead trout released into the ice and trash sluiceway at Bonneville Dam's powerhouse 1 during 2004. Treatment releases were made into the ice and trash sluiceway and the control release was released below the powerhouse 2, juvenile bypass (JBS) outfall at Bonneville Dam. The survival estimates presented are the estimated survival of the release group into the ice and trash sluiceway to the release location of the tailrace release group.

Table 15. The estimated survival probabilities (S), standard errors (SE), 95\% confidence intervals, number released ( N ; number of fish), and dates associated with paired releases of radio-tagged steelhead trout passing via spillbays with deflectors located 7 feet above mean sea level during 56 kcfs day ( $05: 00-19: 59$ hours) spill operations and spill at night (20:00-04:59 hrs) up to the total dissolved gas cap of $125 \%$ in the tailrace at the Bonneville Dam spillway. The paired release groups were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

Table 16. The estimated survival probabilities (S), standard errors (SE), $95 \%$ confidence intervals, number released ( N ; number of fish), and dates associated with paired releases of radio-tagged steelhead trout passing via spillbays with deflectors located 14 feet above mean sea level during 56 kcfs day ( $05: 00-19: 59 \mathrm{hrs}$ ) spill operations and spill at night (20:00-04:59 hrs) up to the total dissolved gas cap of $125 \%$ in the tailrace at the Bonneville Dam spillway. The paired release groups were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

Table 17. Counts of radio-tagged steelhead trout for the releases from The Dalles Dam $\left(R_{I}\right)$ and in the tailrace of Bonneville $\operatorname{Dam}\left(R_{2}\right)$ used in the route-specific survival model during 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B1 is powerhouse 1 and B2 is powerhouse 2 at Bonneville Dam. 58

Table 18. Counts of radio-tagged steelhead trout for releases from The Dalles Dam $\left(R_{l}\right)$ and in the tailrace of Bonneville Dam $\left(R_{2}\right)$ used in the route-specific survival model during two spill operations: 56 kcfs day (0500-1959) and total dissolved gas cap (TDG) at night in 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B1 is powerhouse 1 and B2 is powerhouse 2 at Bonneville Dam. 59

Table 19. Summary table of estimated route-specific survival probabilities (S) and their associated standard errors (SE) of steelhead trout survival through Bonneville Dam (Dam survival) generated from the route-specific survival model. The results of $Z$-tests (i.e., $Z$ statistic) structured to assess whether the estimated survival probabilities during the 56 kcfs day spill operations were different than the estimated survival probabilities during the total dissolved gas cap (TDG) night spill operations. Significant results are indicated where $Z \geq 1.645$ given a two-tailed test and $\alpha=0.10$. The JBS refers to the juvenile bypass system at powerhouse 2 . ... 63

Table 20. The estimated survival (S) and standard error (SE) of subyearling Chinook salmon released into the ice and trash sluiceway at Bonneville Dam's powerhouse 1 during two dam operations, summer 2004. Dam operations were 56 kcfs during the day with total dissolved gas cap at night ( $56 \mathrm{kcfs} / \mathrm{TDG}$ ) or 23 kcfs for 24 h . Releases were made directly into the ice and trash sluiceway with the control release below the powerhouse 2 juvenile bypass (JBS) outfall at Bonneville Dam. The survival estimates are for the fish released directly into the ice and trash sluiceway to the release location of the tailrace release group. The specific dates and times of the releases can be referenced in Appendix 1.

Table 21. The estimated survival probabilities (S), standard errors (SE), 95\% confidence intervals, the number released ( N ), and dates and times associated with paired releases of radiotagged subyearling Chinook salmon passing via spillbays with deflectors located 7 feet above mean sea level during 24 h spill with 56 kcfs during the day and night spill until the total dissolved gas cap of $125 \%$ is reached in the tailrace. The paired release groups were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

Table 22. The estimated survival probabilities (S), standard errors (SE), $95 \%$ confidence intervals, number released ( N ), and dates and times associated with paired releases of radiotagged subyearling Chinook salmon passing via spillbays with deflectors located 7 feet above mean sea level during 23 kcfs 24 h spill operations at Bonneville Dam. The paired release groups were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

Table 23. The estimated survival probabilities (S), standard errors (SE), $95 \%$ confidence intervals, number released (N), and dates and times associated with paired releases of radiotagged subyearling Chinook salmon passing via spillbays with deflectors located 14 feet above mean sea level during 24 h spill with 56 kcfs during the day and night spill until the total dissolved gas cap of $125 \%$ is reached in the tailrace. The paired release groups were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

Table 24. The estimated survival probabilities (S), standard errors (SE), $95 \%$ confidence intervals, number released ( N ; number of fish), and dates and times associated with paired releases of radio-tagged subyearling Chinook salmon passing via spillbays with deflectors located 14 feet above mean sea level during $23 \mathrm{kcfs}, 24 \mathrm{~h}$ spill operations at Bonneville Dam. The paired release groups were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

Table 25. Counts of radio-tagged subyearling Chinook salmon for the releases from The Dalles Dam $\left(R_{l}\right)$ and in the tailrace of Bonneville Dam $\left(R_{2}\right)$ during two spill operations: 56 kcfs day/total dissolve gas cap at night (TDG) and 23 kcfs for 24 h used in the route-specific survival model during 2004. Detection history recorded as: 1, detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B1 is powerhouse 1 and B2 is powerhouse 2 at Bonneville Dam.

Table 26. Summary table of estimated route-specific survival probabilities (S) and their associated standard errors (SE) of subyearling Chinook salmon survival through Bonneville Dam (Dam survival) generated from the route-specific survival model at Bonneville Dam during 20 June to 22 July 2004. The results of $Z$-tests (i.e., $Z$-statistic and $P$-value) structured to assess whether the estimated survival probabilities during the 56 kcfs day/TDG night spill operations were different than the estimated survival probabilities during the 23 kcfs spill operations. Significant results are indicated where $Z \geq 1.645$ given a two-tailed test and $\alpha=0.10$. The JBS refers to the juvenile bypass system at powerhouse 2 .

## List of Appendix Tables

Table A1.1. Summary of yearling Chinook salmon releases at Bonneville Dam ice and trash sluiceway during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h posttagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later. 87

Table A1.2. Summary of yearling Chinook salmon releases at the Bonneville Dam powerhouse 1 minimum gap runner (MGR) turbine unit during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later. 88

Table A1.3. Summary of yearling Chinook salmon releases at Bonneville Dam in the tailrace directly below the front roll of the powerhouse 1 MGR turbine unit during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later. 89

Table A1.4. Summary of yearling Chinook salmon releases at The Dalles Dam ice and trash sluiceway (Sluiceway), and the tailrace 550 m downstream of the spillway beneath the I-197 bridge (Tailrace) during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

Table A1.5. Summary of yearling Chinook salmon releases at Bonneville Dam tailrace during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later. .. 94

Table A1.6. Summary of steelhead trout releases at Bonneville Dam ice and trash sluiceway during spring 2004. Dates, times, numbers of tagged fish released ( N ), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later. .. 97

Table A1.7. Summary of steelhead trout releases at the Bonneville Dam powerhouse 1 minimum gap runner (MGR) turbine unit during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

Table A1.8. Summary of steelhead trout releases at Bonneville Dam in the tailrace directly below the front roll of the powerhouse 1 MGR turbine unit during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

Table A1.9. Summary of steelhead trout releases at The Dalles Dam tailrace during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later. 100

Table A1.10. Summary of steelhead trout releases at Bonneville Dam tailrace during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later. 103

Table A1.11. Summary of subyearling Chinook salmon releases at Bonneville Dam ice and trash sluiceway during summer 2004. Dates, times, numbers of tagged fish released (N), 24 h posttagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later. 106

Table A1.12. Summary of subyearling Chinook salmon releases at The Dalles Dam ice and trash sluiceway (Sluiceway), and the tailrace 550 m downstream of the spillway beneath the I-197 bridge (Tailrace) during summer 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

Table A1.13. Summary of subyearling Chinook salmon releases at Bonneville Dam tailrace during summer 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

Table A2.1. Counts of radio-tagged yearling Chinook salmon with the removal of fish within 24 h of detected dead fish for the releases from The Dalles Dam $\left(R_{I}\right)$ and in the tailrace of Bonneville Dam $\left(R_{2}\right)$ used in the route-specific survival model during 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $R_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B 1 is powerhouse 1 and B 2 is powerhouse 2 at Bonneville Dam. .... 117

Table A2.2. Counts of radio-tagged yearling Chinook salmon, with the removal of fish with travel times $>99.7^{\text {th }}$ percentile, from The Dalles Dam $\left(R_{l}\right)$ and in the tailrace of Bonneville Dam $\left(R_{2}\right)$ used in the route-specific survival model during 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $R_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B 1 is powerhouse 1 and B 2 is powerhouse 2 at Bonneville Dam.

Table A2.3. Counts of radio-tagged subyearling Chinook salmon released from The Dalles Dam $\left(R_{l}\right)$ and the tailrace of Bonneville Dam $\left(R_{2}\right)$ during 56 kcfs day/TDG night spill operations with removal of fish within 24 h of deadfish detections used in the route-specific survival model during 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B1 is powerhouse 1 and B2 is powerhouse 2 at Bonneville Dam.

Table A2.4. Counts of radio-tagged subyearling Chinook salmon released from The Dalles Dam $\left(R_{1}\right)$ and the tailrace of Bonneville Dam $\left(R_{2}\right)$ during 32 kcfs spill operations with removal of fish within 24 h of deadfish detections used in the route-specific survival model during 2004.
Detection history recorded as: 1 , detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B1 is powerhouse 1 and B2 is powerhouse 2 at Bonneville Dam.

Table A3.1. Parameter estimates for tag-life using the Gompertz model during spring and summer during 2004, model estimate and (SE).

Table A3.2. Estimated probabilities (mean, SE in parentheses) a radio-tag was operational at Bonneville Dam and other downstream detection arrays for yearling Chinook salmon, hatchery steelhead trout, and subyearling Chinook salmon, during 2004.

Table A4.1. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for paired releases of yearling Chinook salmon, spring 2004. Treatment fish were released into the minimum gap runner (MGR) turbine 4A at Bonneville Dam powerhouse 1 and control fish were released directly downstream of the front roll below the (MGR) turbine unit at Bonneville Dam powerhouse 1.

Table A4.2. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for paired releases of yearling Chinook salmon, spring 2004. Treatment fish were released into the minimum gap runner turbine 4A at Bonneville Dam powerhouse 1 and control fish were released below the Bonneville Dam powerhouse 2 juvenile bypass outfall. 133

Table A4.3. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for paired releases of yearling Chinook salmon, spring 2004. Treatment fish were released at the top of the ice and trash sluiceway at Bonneville Dam powerhouse 1 and control fish were released below the Bonneville Dam powerhouse 2 juvenile bypass outfall. 134

Table A4.4. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for paired releases of hatchery steelhead trout, spring 2004. Treatment fish were released into the minimum gap runner (MGR) turbine 4A at Bonneville Dam powerhouse 1 and control fish were released directly downstream of the front roll below the (MGR) turbine unit at Bonneville Dam powerhouse 1. 135

Table A4.5. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for paired releases of hatchery steelhead trout, spring 2004. Treatment fish were released into the minimum gap runner turbine 4A at Bonneville Dam powerhouse 1 and control fish were released below the Bonneville Dam powerhouse 2 juvenile bypass outfall. 136

Table A4.6. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for paired releases of hatchery steelhead trout, spring 2004. Treatment fish were released at the top of the sluiceway at Bonneville Dam powerhouse 1 and control fish were released below the Bonneville Dam powerhouse 2 juvenile bypass outfall. 137

Table A4.7. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for each of 32 paired releases of subyearling Chinook salmon, summer 2004. Treatment fish were released at the ice and trash sluiceway of Bonneville Dam powerhouse 1 and control fish were released below the powerhouse 2 juvenile bypass outfall at Bonneville Dam. 138

Table A4.8. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of yearling Chinook salmon through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, spring 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam.

Table A4.9. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of hatchery steelhead trout through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, spring 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam. 143

Table A4.10. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of subyearling Chinook through spillbays with deflectors at 7 ft and $14-\mathrm{ft}$, summer 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam. BIOP spill $=56 \mathrm{kcfs}$ day/total dissolved gas cap at night.

Table A5.1. Summary of chi-square tests for homogeneity of arrival times of yearling Chinook salmon released into the minimum gap runner (MGR) turbine unit at Bonneville Dam powerhouse 1 and directly downstream of the front roll below the MGR turbine unit and detected at river kilometers 200, 194, and 181, spring 2004. 152

Table A5.2. Summary of chi-square tests for homogeneity of arrival times of yearling Chinook salmon released from a minimum gap runner (MGR) turbine at Bonneville Dam powerhouse 1 and in the Bonneville Dam tailrace below the juvenile bypass outfall at powerhouse 2 and detected at river kilometers 200, 194, and 181, spring 2004.

Table A5.3. Summary of chi-square tests for homogeneity of arrival times of yearling Chinook salmon released into the ice and trash sluiceway of powerhouse 1 at Bonneville Dam and below Bonneville Dam juvenile bypass outfall at powerhouse 2 and detected at river kilometers 200, 194, and 181, spring 2004 153

Table A5.4. Summary of chi-square tests for homogeneity of arrival times of hatchery steelhead trout released into a minimum gap runner (MGR) turbine unit at Bonneville Dam powerhouse 1 and directly downstream of the front roll below the MGR turbine unit and detected at river kilometers 200, 194, and 181, spring 2004.

Table A5.5. Summary of chi-square tests for homogeneity of arrival times of hatchery steelhead trout released from a minimum gap runner (MGR) turbine at Bonneville Dam powerhouse 1 and the Bonneville Dam tailrace below the juvenile bypass outfall at powerhouse 2 and detected at river kilometers 200, 194, and 181, spring 2004. 154

Table A5.6. Summary of chi-square tests for homogeneity of arrival times of hatchery steelhead trout released into the ice and trash sluiceway of powerhouse 1 at Bonneville Dam and below Bonneville Dam juvenile bypass outfall at powerhouse 2 and detected at river kilometers 200, 194 , and 181 , spring 2004 155

Table A5.7. Summary of chi-square tests for homogeneity of arrival times of subyearling Chinook salmon released into the ice and trash sluiceway of powerhouse 1 at Bonneville Dam and below Bonneville Dam juvenile bypass outfall at powerhouse 2 and detected at river kilometers 200, 194, and 181, summer 2004.

Table A5.8. Summary of chi-square tests for homogeneity of arrival times of yearling Chinook salmon released and used to estimate survival through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, spring 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam and detected at river kilometers 200, 194, and 181.

Table A5.9. Summary of chi-square tests for homogeneity of arrival times of hatchery steelhead trout used to estimate survival through spillbays with deflectors at $7-\mathrm{ft}$ and 14-ft, spring 2004. Releases were in The Dalles Dam tailrace and in the tailrace of Bonneville Dam and detected at river kilometers 200, 194, and 181, spring 2004.

Table A5.10. Summary of chi-square tests for homogeneity of arrival times of subyearling Chinook salmon used to estimate survival through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, summer 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were in the tailrace of Bonneville Dam and detected at river kilometers 200, 194, and 181. BIOP spill $=56 \mathrm{kcfs}$ day $/$ total dissolved gas cap at night.

## Executive Summary

During 2004, the USGS evaluated the survival of radio-tagged yearling and subyearling Chinook salmon and steelhead trout through the ice and trash sluiceway and the minimum gap runner (MGR) turbine unit at Bonneville Dam's powerhouse 1. Survival was estimated using paired release-recapture models with paired releases made directly into these passage routes and in the tailrace of Bonneville Dam. For the evaluations of survival through the MGR two separate control release locations were used; one location was directly downstream of the front roll below the turbine unit and the other release location was further downstream of the powerhouse 2 juvenile bypass outfall. During spring and summer releases of radio-tagged fish into the MGR and the ice and trash sluiceway, powerhouse 1 was not continuously operated due to a policy that prioritized the passage of water through powerhouse 2 . Because of this policy, powerhouse 1 was only operated sporadically for short time intervals before and after the releases of radiotagged fish associated with this study.

Using releases of radio-tagged yearling and subyearling Chinook salmon and steelhead trout released at The Dalles Dam, and releases made into the tailrace of Bonneville Dam, we also evaluated survival through Bonneville Dam spillway, powerhouse 1, powerhouse 2 turbines, and the corner collector and juvenile bypass system (JBS) at powerhouse 2. We also estimated dam survival (all routes combined) and project survival (product of pool and dam survival). Further we estimated the survival of fish passing via spillbays at Bonneville Dam with either $7-\mathrm{ft}$ or $14-\mathrm{ft}$ deflectors located above mean sea level using the paired release-recapture model. However, dissimilar to the releases into the ice and trash sluiceway and MGR, the paired release groups were formed post-hoc from releases of radio-tagged fish at The Dalles Dam and Bonneville Dam tailraces.

After the completion of the field component of this study, the Army Corps of Engineers identified a discrepancy between the reported inflow (The Dalles outflow + tributary inflow) and outflow from Bonneville Dam during times of spill. The reported spillway discharge was greater than the actual discharge (as measured downstream of the dam). The magnitude of this discrepancy varied but was on the order of 20 kcfs less than what had been reported (see: Bonneville Lock and Dam, Revision to Fish Passage Plan Spill Patterns FEB2005, Memorandum; Appendix 6). The USGS was to estimate the survival for spring migrants during the Biological Opinion (BIOP) spill operations of 75 kcfs during the day and spill up to $125 \%$ of the total dissolved gas cap at night ( $75 \mathrm{kcfs} / \mathrm{TDG}$ ). However, due to the discrepancy in reported and actual spillway discharge, the spring spill operation evaluated was on the order of 56 kcfs day/TDG at night. During the summer evaluation of radio-tagged subyearling Chinook salmon, two spill operations were to be examined, $75 \mathrm{kcfs} / \mathrm{TDG}$ and 50 kcfs for 24 hour. Again, the actual spill operations evaluated during the summer were approximately 56 kcfs day/TDG night and 23 kcfs for 24 h . This report contains the updated spill values given by the Army Corps of Engineers using methods explained in Bonneville Lock and Dam, Revision to Fish Passage Plan Spill Patterns FEB2005, Memorandum (Appendix 6).

# Yearling Chinook salmon 

## Paired Release-recapture Model

## Minimum Gap Runner Turbine Unit

## Control group released directly downstream of front roll below turbine unit

The estimated survival of yearling Chinook salmon released into the MGR turbine unit at Bonneville Dam's powerhouse 1 with the control group released in the tailrace below the MGR unit front roll ranged from 0.830 to 1.042 during 2004. The average survival was estimated to be $0.956(\mathrm{SE}=0.016,95 \%$ confidence interval [0.924, 0.988]).

## Control group released below the outfall of the powerhouse 2 juvenile bypass

The estimated survival of yearling Chinook salmon released into the MGR turbine unit at Bonneville Dam's powerhouse 1 with the control group released in the tailrace below the powerhouse 2 juvenile bypass outfall ranged from 0.83 to 1.02 during 2004. The average survival was estimated to be 0.944 ( $\mathrm{SE}=0.015,95 \%$ confidence interval [0.913, 0.976]).

## Ice and Trash Sluiceway

The estimated survival of yearling Chinook salmon released into the ice and trash sluiceway at Bonneville Dam's powerhouse 1 (control release in the tailrace below the outfall of powerhouse 2 JBS) during 2004 ranged from 0.84 tol.15. The average survival was estimated to be 0.947 ( $\mathrm{SE}=0.018,95 \%$ confidence interval [ $0.908,0.986]$ ).

## Spillway Flow Deflectors

## 7-Ft Spillbay Flow Deflectors

The point estimates of the survival for yearling Chinook salmon passing via spillbays with deflectors located 7 feet above mean sea level at Bonneville Dam spillway during daytime 56 kcfs spill operations ranged from 0.780 to 1.010
( $\bar{X}=0.937 ; S E=0.018 ; 95 \% C I=[0.898,0.976]$ ). Survival estimates for yearling Chinook salmon passing via spillbays with 7 -ft deflectors during total dissolved gas cap spill operations ranged from 0.780 to 1.077 ( $\bar{X}=0.943 ; S E=0.026 ; 95 \% C I=[0.886,1.000]$ ).

## 14-Ft Spillbay Flow Deflectors

The point estimates of the survival for yearling Chinook salmon passing via spillbays with deflectors located 14 feet above mean sea level at Bonneville Dam spillway during 56 kcfs spill operations ranged from 0.643 to $1.020(\bar{X}=0.773 ; S E=0.045 ; 95 \% C I=[0.667,0.879]$ ). Survival estimates for yearling Chinook salmon passing via spillbays with $14-\mathrm{ft}$ deflector during total dissolved gas cap spill operations ranged from 0.845 to 1.087
( $\bar{X}=0.946 ; S E=0.018 ; 95 \% C I=[0.907,0.985]$ ).

## Route-specific Survival Model

## 56 kcfs day/TDG night spill operations

During the 56 kcfs day/TDG night spill operations, the survival of yearling Chinook salmon through Bonneville Dam spillway was estimated to be 0.910 ( $\mathrm{SE}=0.011$, profile likelihood $95 \%$ confidence interval [ $0.888,0.931]$ ). For yearling Chinook salmon passing via powerhouse 1, the estimated survival was 0.913 ( $\mathrm{SE}=0.019$, profile likelihood $95 \%$ confidence interval [ $0.872,0.949]$ ) and for yearling Chinook salmon passing via powerhouse 2 turbines the estimated survival was 0.951 ( $\mathrm{SE}=0.011$, profile likelihood $95 \%$ confidence interval [0.929, 0.972 ]). For yearling Chinook salmon passing via the JBS the estimated survival was 0.970 (SE $=0.013$, profile likelihood $95 \%$ confidence interval [ $0.943,0.994]$ ) and passing via the corner collector at powerhouse 2 the estimated survival was 1.016 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval [0.999, 1.032]). Yearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.951 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval [ 0.937 , $0.966]$ ) and project survival was estimated to be 0.883 ( $\mathrm{SE}=0.008$, profile likelihood confidence interval [0.868, 0.898]).

## 56 kcfs day spill operations

The survival of yearling Chinook salmon through Bonneville Dam spillway during 56 kcfs day spill operations was estimated to be 0.861 ( $\mathrm{SE}=0.015$, profile likelihood $95 \%$ confidence interval [0.831, 0.889]). For yearling Chinook salmon passing via powerhouse 1 , the estimated survival was 0.894 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval [0.853, $0.930]$ ) and for yearling Chinook salmon passing via powerhouse 2 turbines, the estimated survival was 0.925 ( $\mathrm{SE}=0.012$, profile likelihood $95 \%$ confidence interval [0.902, 0.948]). For yearling Chinook salmon passing via the powerhouse 2 JBS the estimated survival was 0.938 ( $\mathrm{SE}=0.015$, profile likelihood $95 \%$ confidence interval [ $0.906,0.966]$ ). The survival of yearling Chinook salmon passing via the corner collector at powerhouse 2 was estimated to be 0.993 (SE $=0.008$, profile likelihood $95 \%$ confidence interval [ $0.977,1.008]$ ). Yearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.925 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval $[0.912,0.941])$ and project survival was estimated to be $0.860(\mathrm{SE}=$ 0.008 , profile likelihood $95 \%$ confidence interval [0.844, 0.876]).

## Total dissolved gas cap night spill operations

The survival of yearling Chinook salmon through Bonneville Dam spillway during total dissolved gas cap night spill operations was estimated to be 0.964 ( $\mathrm{SE}=0.016$, profile likelihood $95 \%$ confidence interval [ $0.932,0.996]$ ). For yearling Chinook salmon passing via powerhouse 1, the estimated survival was 0.903 ( $\mathrm{SE}=0.060$, profile likelihood $95 \%$ confidence interval [ $0.768,0.999]$ ) and for yearling Chinook salmon passing via powerhouse 2 turbines, the estimated survival was 0.981 ( $\mathrm{SE}=0.021$, profile likelihood $95 \%$ confidence interval [0.937, 1.020]). For yearling Chinook salmon passing via the powerhouse 2 JBS the estimated survival was 1.013 ( $\mathrm{SE}=0.022$, profile likelihood $95 \%$ confidence interval [ $0.964,1.053]$ ). The estimated survival of yearling Chinook salmon passing via the corner collector at powerhouse 2 was 1.028 ( $\mathrm{SE}=0.021$, profile likelihood $95 \%$ confidence interval [0.980, 1.065]). Yearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.979 ( $\mathrm{SE}=0.015$, profile likelihood $95 \%$ confidence interval [0.953, 1.007]) and project survival was estimated to be 0.908 ( $\mathrm{SE}=0.015$, profile likelihood $95 \%$ confidence interval [0.881, 0.937]).

## Steelhead Trout

## Paired Release-recapture Model

## Minimum Gap Runner Turbine Unit

## Control group released directly downstream of front roll below turbine unit

The estimated survival of steelhead trout released into the MGR turbine unit at Bonneville Dam's powerhouse 1 with the control group released in the tailrace below the MGR unit front roll ranged from 0.830 to 1.136 during 2004. The average survival was estimated to be 0.952 ( $\mathrm{SE}=0.024,95 \%$ confidence interval [0.900, 1.003]).

## Control group released below the outfall of the powerhouse 2 juvenile bypass

The estimated survival of steelhead trout released into the MGR turbine unit at Bonneville Dam's powerhouse 1 with the control group released in the tailrace below the outfall of the powerhouse 2 juvenile bypass outfall ranged from 0.74 to 1.09 during 2004. The average survival was estimated to be 0.926 , $(\mathrm{SE}=0.030,95 \%$ confidence interval [0.861, 0.992]).

## Ice and Trash Sluiceway

The estimated survival of steelhead trout released into the ice and trash sluiceway at Bonneville Dam's powerhouse 1 (control release in the tailrace below the outfall of powerhouse 2 juvenile bypass outfall) during 2004 ranged from 0.76 to1.07. The average survival was estimated to be 0.935 ( $\mathrm{SE}=0.024,95 \%$ confidence interval [0.884, 0.985]).

## Spillway Flow Deflectors

## 7-Ft Spillbay Flow Deflectors

The point estimates of survival for steelhead trout passing via spillbays with deflectors located 7 feet above mean sea level at Bonneville Dam spillway during 56 kcfs spill operations ranged from 0.667 to 1.064 ( $\bar{X}=0.927 ; S E=0.046 ; 95 \% C I=[0.818,1.036]$ ). Survival estimates for steelhead trout passing via spillbays with 7 -ft deflectors during total dissolved gas cap spill operations ranged from 0.926 to 1.143
( $\bar{X}=1.013 ; S E=0.016 ; 95 \% C I=[0.979,1.047]$ ).

## 14-Ft Spillbay Flow Deflectors

The point estimates of survival of steelhead trout passing via spillbays with deflectors located 14 feet above sea level at Bonneville Dam spillway during 56 kcfs spill operations ranged from 0.684 to $0.959(\bar{X}=0.850 ; S E=0.063 ; 95 \% C I=[0.650,1.050])$. Survival estimates for steelhead trout passing via spillbays with 14 -ft deflectors during total dissolved gas cap spill operations ranged from 0.878 to 1.143
( $\bar{X}=1.012 ; S E=0.015 ; 95 \% C I=[0.980,1.044]$ ).

## Route-specific Survival Model

## 56 kcfs day/TDG night spill operations

The survival of steelhead trout through Bonneville Dam spillway was estimated to be 0.979 ( $\mathrm{SE}=0.011$, profile likelihood $95 \%$ confidence interval [0.956, 1.002]). For steelhead trout passing via powerhouse 1 the estimated survival was 0.965 ( $\mathrm{SE}=0.019$, profile likelihood $95 \%$ confidence interval $[0.926,0.999]$ ) and for steelhead trout passing via powerhouse 2 turbines, the estimated survival was 0.889 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval $[0.848,0.927])$. For steelhead trout passing via the powerhouse 2 JBS the estimated survival was 0.951 ( $\mathrm{SE}=0.021$, profile likelihood $95 \%$ confidence interval [0.907, 0.989]). The estimated survival of steelhead trout passing via the powerhouse 2 corner collector was 1.030 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval [1.014, 1.047]). Steelhead trout dam survival through Bonneville Dam was estimated to be 0.991 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval $[0.975,1.008]$ ) and project survival was estimated to be 0.897 ( $\mathrm{SE}=0.009$, profile likelihood $95 \%$ confidence interval [ $0.881,0.915]$ ).

## 56 kcfs day spill operations

The survival of steelhead trout through Bonneville Dam spillway during 56 kcfs day spill operations was estimated to be 0.891 ( $\mathrm{SE}=0.024$, profile likelihood $95 \%$ confidence interval [ $0.840,0.936]$ ). For steelhead trout passing via powerhouse 1 , the estimated survival was 0.966 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval [ $0.922,1.003]$ ) and for steelhead trout passing via powerhouse 2 turbines the estimated survival was 0.863 ( $\mathrm{SE}=0.028$, profile likelihood $95 \%$ confidence interval $[0.804,0.915]$ ). For steelhead trout passing via the powerhouse 2 JBS , the estimated survival was 0.904 ( $\mathrm{SE}=0.031$, profile likelihood $95 \%$ confidence interval [0.837, 0.960]). For steelhead trout passing via the powerhouse 2, corner collector, the estimated survival was 1.018 ( $\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval $[0.9998,1.039])$. Steelhead trout dam survival through Bonneville Dam was estimated to be $0.980(\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval $[0.962,1.001])$ and project survival was estimated to be 0.888 ( $\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval [0.868, 0.909].

## Total dissolved gas cap night spill operations

The survival of steelhead trout through Bonneville Dam spillway during total dissolved gas cap spill operations was estimated to be $1.020(\mathrm{SE}=0.015$, profile likelihood $95 \%$ confidence interval [0.992, 1.050]). For steelhead trout passing via powerhouse 1, the estimated survival was $0.940(\mathrm{SE}=0.041$, profile likelihood $95 \%$ confidence interval [0.850, 1.009]) and for steelhead trout passing via powerhouse 2 turbines the estimated survival was 0.917 ( $\mathrm{SE}=$ 0.029 , profile likelihood $95 \%$ confidence interval [ $0.857,0.970]$ ). For steelhead trout passing via the powerhouse 2 JBS, the estimated survival was 1.003 ( $\mathrm{SE}=0.027$, profile likelihood $95 \%$ confidence interval [0.944, 1.050]). For steelhead trout passing via the powerhouse 2, corner collector the estimated survival was 1.028 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval [0.985, 1.066]). Steelhead trout dam survival through Bonneville Dam was estimated to be 0.998 ( $\mathrm{SE}=0.014,95 \%$ profile likelihood confidence interval [0.973, 1.027]) and project survival was estimated to be 0.904 ( $\mathrm{SE}=0.015$, profile likelihood confidence interval [0.876, 0.933]).

## Subyearling Chinook salmon

## Paired Release-recapture Model

## Ice and trash sluiceway

We evaluated the survival of subyearling Chinook salmon released into the ice and trash sluiceway from 21 June to 22 July at Bonneville Dam's powerhouse 1. The control group for this evaluation was released in the tailrace below the outfall of the powerhouse 2 juvenile bypass outfall. The estimated survival of subyearling Chinook salmon released into the ice and trash sluiceway ranged from 0.67 to 1.11 during 2004. The average survival was estimated to be 0.925 ( $\mathrm{SE}=0.019,95 \%$ confidence interval [0.887, 0.962]).

Since there were two dam operations implemented during the summer migration, we also evaluated the survival of subyearling Chinook salmon released into the ice and trash sluiceway during both spill operations. The two spill operations were, 24 h spill with 56 kcfs during the day ( 0400 to 2200 hrs ) and night spill until the total dissolved gas cap of $125 \%$ is reached in the tailrace of the dam ( 56 kcfs day/TDG night) and spill operations of 23 kcfs for 24 h ( 23 kcfs ). The average survival of subyearling Chinook salmon through the ice and trash sluiceway during the 56 kcfs day/TDG night spill operations was estimated to be 0.916 ( $\mathrm{SE}=0.025,95 \%$ confidence interval [0.862, 0.969]). For subyearling Chinook salmon passing via the ice and trash sluiceway during the 23 kcfs spill operations, the average survival was estimated to be 0.934 ( $\mathrm{SE}=0.028,95 \%$ confidence interval $[0.875,0.994]$ ). The average survival of subyearling Chinook salmon during 56 kcfs day/TDG night spill operations was not significantly different than survival during 23 kcfs spill operations (two-tailed t-test, $P=0.6234, \beta=0.077$ ).

## Spillway Flow Deflectors

## 7-ft spillbay flow deflectors during two spill operations

The point estimates of survival of subyearling Chinook salmon passing via spillbays with deflectors located 7 feet above mean sea level at Bonneville Dam spillway during 56 kcfs day and total dissolved gas cap night spill operations ranged from 0.832 to 1.055
$(\bar{X}=0.920 ; S E=0.010 ; 95 \% C I=[0.899,0.941])$. The estimated survival of subyearling
Chinook salmon passing via spillbays with 7 - ft deflectors during 23 kcfs spill operations ranged from 0.583 to $1.000(\bar{X}=0.822 ; S E=0.033 ; 95 \% C I=[0.758,0.886]$ ).

## 14-ft spillbay deflectors during two spill operations

The point estimates of survival of subyearling Chinook salmon passing via spillbays with deflectors located 14 feet above mean sea level at Bonneville Dam spillway during 56 kcfs day and total dissolved gas cap night spill operations ranged from 0.552 to 1.066
( $\bar{X}=0.803 ; S E=0.026 ; 95 \% C I=[0.749,0.857]$ ). The estimated survival of subyearling
Chinook salmon passing via spillbays with 14 - ft deflectors during 23 kcfs spill operations ranged from 0.553 to 0.913 ( $\bar{X}=0.741 ; S E=0.027 ; 95 \% C I=[0.683,0.799]$ ).

## Route-specific Survival Model

Using capture histories generated from the detections of radio-tagged subyearling Chinook salmon released at The Dalles Dam and in the tailrace of Bonneville Dam, we generated maximum likelihood estimates of the route-specific passage and survival probabilities for subyearling Chinook salmon through Bonneville Dam during two spill operations from 20 June to 22 July. The two spill operations were, 56 kcfs during the day ( 0400 to 2200 hrs ) and night spill until the total dissolved gas cap of $125 \%$ is reached in the tailrace ( 56 kcfs day/TDG night) and a spill volume of 23 kcfs for 24 h ( 23 kcfs ).

## 56 kcfs day/TDG night spill operations

During the 56 kcfs day/TDG night spill operations, the survival of subyearling Chinook salmon through the Bonneville Dam spillway was estimated to be 0.877 ( $\mathrm{SE}=0.013$ profile likelihood $95 \%$ confidence interval [0.848, 0.902]). For subyearling Chinook salmon passing via powerhouse 1, the estimated survival was 0.827 ( $\mathrm{SE}=0.061$, profile likelihood $95 \%$ confidence interval [0.694, 0.937]) and for subyearling Chinook salmon passing via powerhouse 2 turbines, the estimated survival was 0.824 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval [0.782, $0.864]$ ). For subyearling Chinook salmon passing via the powerhouse 2 JBS the estimated survival was 0.927 ( $\mathrm{SE}=0.027$, profile likelihood $95 \%$ confidence interval [0.863, 0.976]). For subyearling Chinook salmon passing via the powerhouse 2 corner collector the estimated survival was 0.981 ( $\mathrm{SE}=0.013$, profile likelihood $95 \%$ confidence interval [0.951, 1.005]). Subyearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.891 $(\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval $[0.871,0.910)$ and project survival was estimated to be 0.768 ( $\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval [0.747, 0.788].

## 56 kcfs day spill operations

During the 56 kcfs day spill operations, the survival of subyearling Chinook salmon through the Bonneville Dam spillway was estimated to be 0.851 ( $\mathrm{SE}=0.016$ profile likelihood $95 \%$ confidence interval [ $0.819,0.883]$ ). For subyearling Chinook salmon passing via powerhouse 1, the estimated survival was 0.845 ( $\mathrm{SE}=0.059$, profile likelihood $95 \%$ confidence interval $[0.715,0.943])$ and for subyearling Chinook salmon passing via powerhouse 2 turbines, the estimated survival was 0.834 ( $\mathrm{SE}=0.021$, profile likelihood $95 \%$ confidence interval [0.792, $0.873]$ ). For subyearling Chinook salmon passing via the powerhouse 2 JBS the estimated survival was 0.900 ( $\mathrm{SE}=0.031$, profile likelihood $95 \%$ confidence interval [0.833, 0.955]). For subyearling Chinook salmon passing via the powerhouse 2 corner collector the estimated survival was 0.966 ( $\mathrm{SE}=0.014$, profile likelihood $95 \%$ confidence interval [0.937, 0.991]). Subyearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.882 $(\mathrm{SE}=0.011,95 \%$ confidence interval $[0.861,0.903)$ and project survival was estimated to be 0.763 ( $\mathrm{SE}=0.012$, profile likelihood $95 \%$ confidence interval [0.741, 0.785]).

## Total dissolved gas night spill operations

During the TDG night spill operations, the survival of subyearling Chinook salmon through the Bonneville Dam spillway was estimated to be 0.913 ( $\mathrm{SE}=0.021$ profile likelihood $95 \%$ confidence interval [0.869, 0.953]). For subyearling Chinook salmon passing via powerhouse 1, the estimated survival was 0.519 ( $\mathrm{SE}=0.266$, profile likelihood $95 \%$ confidence interval $[0.109,0.947])$ and for subyearling Chinook salmon passing via powerhouse 2 turbines,
the estimated survival was 0.696 ( $\mathrm{SE}=0.062$, profile likelihood $95 \%$ confidence interval [0.570, $0.810]$ ). For subyearling Chinook salmon passing via the powerhouse 2 JBS the estimated survival was 0.998 ( $\mathrm{SE}=0.048$, profile likelihood $95 \%$ confidence interval [0.868, 1.060]). For subyearling Chinook salmon passing via the powerhouse 2 corner collector the estimated survival was $1.009(\mathrm{SE}=0.041$, profile likelihood $95 \%$ confidence interval [0.898, 1.063]). Subyearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.887 $(S E=0.022$, profile likelihood $95 \%$ confidence interval $[0.847,0.925)$ and project survival was estimated to be 0.757 ( $\mathrm{SE}=0.023$, profile likelihood $95 \%$ confidence interval [ $0.715,0.797]$ ).

## 23 kcfs 24 h spill operations

During the 23 kcfs spill operations, the survival of subyearling Chinook salmon through the Bonneville Dam spillway was estimated to be 0.744 ( $\mathrm{SE}=0.022$, profile likelihood $95 \%$ confidence interval [0.700, 0.786]). For subyearling Chinook salmon passing via powerhouse 1, the estimated survival was 0.829 ( $\mathrm{SE}=0.030$, profile likelihood $95 \%$ confidence interval [0.767, $0.884]$ ) and for subyearling Chinook salmon passing via powerhouse 2 turbines, the estimated survival was 0.833 ( $\mathrm{SE}=0.014$, profile likelihood $95 \%$ confidence interval [0.805, 0.860]). For subyearling Chinook salmon passing via the powerhouse 2 JBS , the estimated survival was 0.958 ( $\mathrm{SE}=0.019$, profile likelihood $95 \%$ confidence interval [0.918, 0.991]). For subyearling Chinook salmon passing via the powerhouse 2 corner collector, the estimated survival was 0.954 ( $\mathrm{SE}=0.013$, profile likelihood $95 \%$ confidence interval [ $0.926,0.978]$ ). Subyearling Chinook salmon survival through Bonneville Dam was estimated to be 0.858 ( $\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval [ $0.840,0.876]$ ) and project survival was estimated to be $0.736(\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval [ $0.717,0.754]$ ).

## 23 kcfs day spill operations

For subyearling Chinook salmon passing during day ( 0400 to 2200 hrs ) 23 kcfs spill operations, survival through the Bonneville Dam spillway was estimated to be 0.725 ( $\mathrm{SE}=$ 0.025 , profile likelihood $95 \%$ confidence interval [ $0.675,0.773]$ ). For subyearling Chinook salmon passing via powerhouse 1 , the estimated survival was 0.826 ( $\mathrm{SE}=0.033$, profile likelihood $95 \%$ confidence interval [ $0.758,0.886]$ ) and for subyearling Chinook salmon passing via powerhouse 2 turbines, the estimated survival was 0.820 ( $\mathrm{SE}=0.017$, profile likelihood $95 \%$ confidence interval [0.785, 0.853]). For subyearling Chinook salmon passing via the powerhouse 2 JBS , the estimated survival was 0.957 ( $\mathrm{SE}=0.024$, profile likelihood $95 \%$ confidence interval $[0.906,0.999]$ ) and for passing via the corner collector the estimated survival was 0.955 ( $\mathrm{SE}=0.015$, profile likelihood $95 \%$ confidence interval [0.926, 0.983]). Subyearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.852 ( $\mathrm{SE}=0.012$, profile likelihood $95 \%$ confidence interval [0.831, 0.875]) and project survival was estimated to be 0.731 ( $\mathrm{SE}=0.012$, profile likelihood $95 \%$ confidence interval [0.710, 0.753$]$ ).

## 23 kcfs night spill operations

For subyearling Chinook salmon passing during the night ( 2200 to 0400 hrs ) 23 kcfs spill operations, survival through the Bonneville Dam spillway was estimated to be 0.830 ( $\mathrm{SE}=$ 0.045 , profile likelihood $95 \%$ confidence interval [0.735, 0.909]). For subyearling Chinook salmon passing via powerhouse 1 , the estimated survival was 0.863 ( $\mathrm{SE}=0.077$, profile likelihood $95 \%$ confidence interval [ $0.687,0.982$ ]) and for subyearling Chinook salmon passing via powerhouse 2 turbines, the estimated survival was 0.864 ( $\mathrm{SE}=0.023$, profile likelihood $95 \%$
confidence interval $[0.818,0.907])$. For subyearling Chinook salmon passing via the powerhouse 2 JBS , the estimated survival was 0.960 ( $\mathrm{SE}=0.030$, profile likelihood $95 \%$ confidence interval $[0.891,1.010]$ ) and for passing via the corner collector the estimated survival was $0.959(\mathrm{SE}=0.045$, profile likelihood $95 \%$ confidence interval [ $0.847,1.025]$ ). Subyearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.881 ( $\mathrm{SE}=0.019$, profile likelihood $95 \%$ confidence interval [ $0.847,0.914]$ ) and project survival was estimated to be 0.755 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval [ $0.719,0.791$ ]).

## Comparison of estimators generated during 56 kcfs day/TDG night and 23 kcfs spill operations

The estimated survival probabilities for subyearling Chinook salmon passing via the powerhouse 1, powerhouse 2 turbines, and the powerhouse 2 corner collector and JBS were not found to be significantly different between the two spill operations. However, the estimated survival of subyearling Chinook salmon passing via the spillway was significantly different for fish passing during the 56 kcfs day/TDG night spill operations and the 23 kcfs spill operations. The estimated dam survival for fish passing during the two spill operations was also found to be significantly different.

Executive Summary Table 1. The estimated survival probabilities and $95 \%$ confidence intervals of radio-tagged, yearling Chinook and subyearling Chinook salmon, and steelhead trout released directly into the Minimum Gap Runner turbine unit (MGR) and ice and trash sluiceway at Bonneville Dam's powerhouse 1 and in the Bonneville Dam tailrace. Also presented are the estimated survival probabilities of fish passing via spillbays with either $7-\mathrm{ft}$ or $14-\mathrm{ft}$ spillbay deflectors at Bonneville Dam spillway estimated from the post-hoc pairing of release groups formulated from fish released at The Dalles Dam (treatment) and Bonneville Dam tailrace (control). Spring spill operations were 56 kcfs during the day and total dissolved gas cap at night ( $56 \mathrm{kcfs} / \mathrm{TDG}$ ). Two spill operations were examined in the summer, $56 \mathrm{kcfs} / \mathrm{TDG}$ and 23 kcfs for 24 hour.

| Paired release-recapture model |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spill Operations | MGR, downstream control |  | MGR, front roll control |  |  | Ice and trash sluiceway |  |
|  | S | 95\% CI | S | 95\% CI |  | S | 95\% CI |
| Yearling Chinook Salmon |  |  |  |  |  |  |  |
| $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.944 | 0.913, 0.976 | 0.956 | 0.924, 0.988 |  | 0.947 | 0.908, 0.986 |
| Steelhead trout |  |  |  |  |  |  |  |
| $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.926 | 0.861, 0.992 | 0.952 | 0.900, 1.003 |  | 0.935 | 0.884, 0.985 |
| Subyearling Chinook salmon |  |  |  |  |  |  |  |
| $56 \mathrm{kcfs} / \mathrm{TDG}^{\text {A }}$ |  |  |  |  |  | 0.916 | 0.862, 0.969 |
| $23 \mathrm{kcfs} 24 \mathrm{~h}^{\text {A }}$ |  |  |  |  |  | 0.934 | 0.875, 0.994 |
| Paired release-recapture model |  |  |  |  |  |  |  |
| Spill Operations | 7-ft Spillbay |  |  |  | 14-ft Spillbay |  |  |
|  | S | 95\% CI |  |  | S |  | 95\% CI |
| Yearling Chinook Salmon |  |  |  |  |  |  |  |
| 56 kcfs Day | 0.937 | 7 0.898, 0.976 |  |  | 0.773 |  | 0.667, 0.879 |
| TDG Night | 0.943 | - 0.886, 1.000 |  |  | 0.946 |  | 0.907, 0.985 |
| Steelhead trout |  |  |  |  |  |  |  |
| 56 kcfs Day | 0.927 | 7 0.818, 1.036 |  |  | 0.850 |  | 0.650, 1.050 |
| TDG Night | 1.013 | 3 0.979, 1.047 |  |  | 1.012 |  | 0.980, 1.044 |
| Subyearling Chinook salmon |  |  |  |  |  |  |  |
| $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.920 | 0 0.899, 0.941 |  |  | 0.803 |  | $0.749,0.857$ |
| 23 kcfs 24 h | 0.822 | 0.751, 0.893 |  |  | 0.741 |  | 0.683, 0.799 |

[^0]Executive Summary Table 2. The route-specific survival probabilities and associated profile likelihood 95\% confidence intervals for radio-tagged, yearling Chinook salmon, steelhead trout, and subyearling Chinook salmon passing through Bonneville Dam. Fish were released in The Dalles Dam and Bonneville Dam tailraces. Spring spill operations were 56 kcfs during the day and night spill until the total dissolved gas cap (TDG) of $125 \%$ was reached in the tailrace ( 56 $\mathrm{kcfs} / \mathrm{TDG}$ ). Two spill operations were examined in the summer, $56 \mathrm{kcfs} / \mathrm{TDG}$ and 23 kcfs for 24 hour from 20 June to 22 July.

| Route-Specific Survival Model |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spill Operations | Yearling Chinook |  | Steelhead Trout |  | Subyearling Chinook |  |
|  |  |  |  |  |  |  |
|  | S | 95\% CI | S | 95\% CI | S | 95\% CI |
|  | Spillway |  |  |  |  |  |
| 56 kcfs Day | 0.861 | 0.831, 0.889 | 0.891 | 0.840, 0.936 | 0.851 | 0.819, 0.883 |
| TDG Night | 0.964 | 0.932, 0.996 | 1.020 | 0.992, 1.050 | 0.913 | 0.869, 0.953 |
| $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.910 | 0.888, 0.931 | 0.979 | 0.956, 1.002 | 0.877 | 0.848, 0.902 |
| 23 kcfs Day | NA | NA | NA | NA | 0.725 | $0.675,0.773$ |
| 23 kcfs Night | NA | NA | NA | NA | 0.830 | 0.735, 0.909 |
| 23 kcfs 24 h | NA | NA | NA | NA | 0.744 | 0.700, 0.786 |
|  | Powerhouse 1 |  |  |  |  |  |
| 56 kcfs Day | 0.894 | 0.853, 0.930 | 0.966 | 0.922, 1.003 | 0.845 | 0.715, 0.943 |
| TDG Night | 0.903 | 0.768, 0.999 | 0.940 | 0.850, 1.009 | 0.519 | 0.109, 0.947 |
| $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.913 | 0.872, 0.949 | 0.965 | 0.926, 0.999 | 0.827 | 0.694, 0.937 |
| 23 kcfs Day | NA | NA | NA | NA | 0.826 | 0.758, 0.886 |
| 23 kcfs Night | NA | NA | NA | NA | 0.863 | 0.687, 0.982 |
| 23 kcfs 24 h | NA | NA | NA | NA | 0.829 | 0.767, 0.884 |
|  | Powerhouse 2 |  |  |  |  |  |
| 56 kcfs Day | 0.925 | 0.902, 0.948 | 0.863 | 0.804, 0.915 | 0.834 | 0.792, 0.873 |
| TDG Night | 0.981 | 0.937, 1.020 | 0.917 | 0.857, 0.970 | 0.696 | 0.570, 0.810 |
| $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.951 | 0.929, 0.972 | 0.889 | 0.848, 0.927 | 0.824 | 0.782, 0.864 |
| 23 kcfs Day | NA | NA | NA | NA | 0.820 | 0.785, 0.853 |
| 23 kcfs Night | NA | NA | NA | NA | 0.864 | 0.818, 0.907 |
| 23 kcfs 24 h | NA | NA | NA | NA | 0.833 | 0.805, 0.860 |

Executive Summary Table 2 (continued). The route-specific survival probabilities and associated profile likelihood $95 \%$ confidence intervals for radio-tagged, yearling Chinook salmon, steelhead trout, and subyearling Chinook salmon passing through Bonneville Dam. Fish were released in The Dalles Dam and Bonneville Dam tailraces. Spring spill operations were 56 kcfs during the day and night spill until the total dissolved gas cap (TDG) of $125 \%$ was reached in the tailrace ( $56 \mathrm{kcfs} / \mathrm{TDG}$ ). Two spill operations were examined in the summer, $56 \mathrm{kcfs} / \mathrm{TDG}$ and 23 kcfs for 24 hour from 20 June to 22 July.

| Route-specific Survival Model |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spill Operations | Yearling Chinook |  | Steelhead Trout |  | Subyearling Chinook |  |
|  | S | 95\% CI | S | 95\% CI | S | 95\% CI |
|  | Juvenile Bypass System |  |  |  |  |  |
| 56 kcfs Day | 0.938 | 0.906, 0.966 | 0.904 | 0.837, 0.960 | 0.900 | 0.833, 0.955 |
| TDG Night | 1.013 | 0.964, 1.053 | 1.003 | 0.944, 1.050 | 0.998 | 0.868, 1.060 |
| $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.970 | 0.943, 0.994 | 0.951 | 0.907, 0.989 | 0.927 | 0.863, 0.976 |
| 23 kcfs Day | NA | NA | NA | NA | 0.957 | 0.906, 0.999 |
| 23 kcfs Night | NA | NA | NA | NA | 0.960 | 0.891, 1.010 |
| 23 kcfs 24 h | NA | NA | NA | NA | 0.958 | 0.918, 0.991 |
|  | Corner Collector |  |  |  |  |  |
| 56 kcfs Day | 0.993 | 0.978, 1.008 | 1.018 | 0.9998, 1.039 | 0.966 | 0.937, 0.991 |
| TDG Night | 1.028 | 0.980, 1.065 | 1.028 | 0.985, 1.066 | 1.009 | 0.898, 1.063 |
| $56 \mathrm{kcfs} / \mathrm{TDG}$ | 1.016 | 0.999, 1.032 | 1.030 | 1.014, 1.047 | 0.981 | 0.951, 1.005 |
| 23 kcfs Day | NA |  | NA | NA | 0.955 | 0.926, 0.983 |
| 23 kcfs Night | NA | NA | NA | NA | 0.959 | 0.847, 1.025 |
| 23 kcfs 24 h | NA | NA | NA | NA | 0.954 | 0.926, 0.978 |
|  | Dam |  |  |  |  |  |
| 56 kcfs Day | 0.925 | 0.912, 0.941 | 0.980 | 0.962, 1.001 | 0.882 | 0.861, 0.903 |
| TDG Night | 0.979 | 0.953, 1.007 | 0.998 | 0.973, 1.027 | 0.887 | 0.847, 0.925 |
| $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.951 | 0.937, 0.966 | 0.991 | 0.975, 1.008 | 0.891 | 0.871, 0.910 |
| 23 kcfs Day | NA | NA | NA | NA | 0.852 | 0.831, 0.875 |
| 23 kcfs Night | NA | NA | NA | NA | 0.881 | 0.847, 0.914 |
| 23 kcfs 24 h | NA | NA | NA | NA | 0.858 | 0.840, 0.876 |
|  | Project |  |  |  |  |  |
| 56 kcfs Day | 0.860 | 0.844, 0.876 | 0.888 | 0.868, 0.909 | 0.763 | 0.741, 0.785 |
| TDG Night | 0.908 | 0.881, 0.937 | 0.904 | 0.876, 0.933 | 0.757 | 0.715, 0.797 |
| $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.883 | 0.868, 0.898 | 0.897 | 0.881, 0.915 | 0.768 | 0.747, 0.788 |
| 23 kcfs Day | NA | NA | NA | NA | 0.731 | 0.710, 0.753 |
| 23 kcfs Night | NA | NA | NA | NA | 0.755 | 0.719, 0.791 |
| 23 kcfs 24 h | NA | NA | NA | NA | 0.736 | $0.717,0.754$ |

## Introduction

As anadromous juvenile salmonids migrate from freshwater rearing habitats to the ocean, they are vulnerable to a host of factors that affect their survival. Direct effects associated with dam passage (e.g., instantaneous mortality, injury, loss of equilibrium, etc.) and indirect effects (e.g., predation, disease, and physiological stress) contribute to the total mortality of seaward migrating salmonids. Many studies have been conducted to determine the effects of hydroelectric dams on the survival of salmonid migrants (Raymond 1979; Stier and Kynard 1986; Iwamato et al. 1994; Muir et al. 1995; Smith et al. 1998; Bickford and Skalski 2000). Giorgi et al. (2002) noted that survival of salmonid migrants is variable among projects and across species. Thus, studies designed to estimate dam, project, and route-specific survival of juvenile salmon have been conducted to identify sources of mortality and potential mitigation opportunities. Based on these research studies that examine migrant salmonid behavior and survival at dams in the Columbia River Basin, management actions are being implemented to improve the survival of juvenile salmonid migrants.

Improved fish marking techniques and development and acceptance of statistical methodologies (see Lebreton et al. 1992) have led scientists to reevaluate past techniques used to assess survival of migrant salmonids in the Columbia River Basin. The development of the passive integrated transponder (PIT) tag, allowed for the unique identification of fish (Prentice et al. 1990), and recent technological advancements in radio-telemetry equipment have decreased the size and increased the life of transmitters allowing for use with juvenile fish passage behavior and survival studies (Skalski et al. 2001, 2002; Counihan et al. 2001, 2002). Consequently, PITtag recoveries, radio telemetry capture histories, and release-recapture models (Burnham et al. 1987; Smith et al. 1996) have been used to assess the survival of migrant salmonid smolts through various reaches of the Columbia and Snake Rivers (Muir et al. 1995; Skalski et al. 1998; Smith et al. 1998; Dawley et al. 1998; Skalski et al. 2001, 2002). Results from studies examining simultaneous releases of PIT-tagged and radio-tagged fish in the Snake River and mid Columbia River suggest similar trends in survival between the two groups (Hockersmith et al. 2003). Further, concurrent releases of radio- and PIT-tagged yearling Chinook salmon at The Dalles Dam also indicate that estimates from the two tagging techniques provide comparable estimates (Counihan et al. 2001). Estimates of survival generated from radio-tagged subyearling Chinook salmon were less comparable. However, the large confidence intervals associated with both PIT- and radio-tagged fish were not conducive to a meaningful evaluation of the comparability of the estimates.

Although the two techniques have been shown to produce similar results, there are important considerations with each method. The use of the PIT-tag technique relies on the availability of PIT-tag detectors at hydroelectric dams, which are not present at all locations in the Columbia River Basin (e.g. The Dalles Dam). The absence of PIT-tag detectors at certain dams and areas below Bonneville Dam has precluded or confounded survival estimation in some specific reaches of the Columbia River and limited the spatial scale over which survival estimates can be made. Further, the low detection probabilities associated with this technique requires that large numbers of fish be handled (although minimally) to obtain desired levels of precision in survival estimates (Skalski 1999b). Detection rates of marked fish affect the sample size required for a given level of precision and thus, the reliability of survival estimates (Skalski 1992). The radio-telemetry technique offers high detection rates, observed in migrant salmonid
studies at specific project sites and in-river sites in the lower Columbia River, suggesting that the numbers of fish necessary to generate survival estimates with similar or greater precision could be reduced using radio-tagged fish. Further, the flexibility of a radio-telemetry system deployment at hydroelectric projects and in-river locations can increase the geographic area over which estimates are generated (e.g. areas below Bonneville Dam).

Mitigation efforts in the Columbia River Basin have sought to increase survival of juvenile salmonid migrants through the federal hydrosystem (National Marine Fisheries Service 2000). To facilitate this objective, migrant salmonids are diverted from turbine passage by the development of turbine bypass systems and spill scenarios used to increase spillway passage. While there is a consensus that survival is greater for fish diverted from turbines, questions regarding the effectiveness of different spill patterns and other passage scenarios remain (Dawley et al. 1998). During 1999, tests of the efficacy of different spill scenarios were conducted at both John Day and The Dalles dams. The motivation for these evaluations was to identify which spill scenario would increase fish passage efficiency and reduce predation of migrant juvenile salmonids by altering the hydraulic conditions in the forebay environment, shortening travel times through tailrace areas, and manipulating passage routes through tailrace areas to divert fish from areas with high predator densities. Ultimately, these actions are designed to increase the survival of migrant salmonids as they migrate through hydroelectric projects in the lower Columbia River. Thus, there continues to be a need to estimate the dam survival and routespecific survival of migrant juvenile salmonids in the lower Columbia River to evaluate the utility of these management actions. Further, given the completion of the new corner collector at powerhouse 2 as a bypass system, a post construction survival program to evaluate dam and route-specific survival at Bonneville Dam will help fish managers understand the effect of this new passage route on migrating juvenile salmonids.

## Previous USGS survival studies at Bonneville Dam

## Pilot studies

Evaluations conducted during 1999 and 2000 demonstrated the feasibility of using radio telemetry to estimate the survival of juvenile salmonids passing through the John Day, The Dalles, and Bonneville dams (Counihan et al. 2001, 2002a). During 2000, radio-tagged yearling and subyearling Chinook salmon and steelhead trout were released in the lower Columbia River to evaluate fish passage efficiency and estimate survival (Beeman et al. 2001a and 2001b). During 2000, the evaluation of two spill conditions ( 12 vs. 24 h spill) at John Day Dam, indicated differences in survival for groups passing the dam during each operating scenario. However, further analyses suggest that other environmental conditions were variable within and between the two treatments and that the variability in conditions (including spill percent within treatments) may have affected the survival of both yearling Chinook salmon and steelhead trout and confounded the original intent of the experiment. Releases of yearling Chinook salmon during 2000 were made above and below Bonneville Dam to assess the feasibility of estimating survival at this dam. The results of the pilot study at Bonneville Dam suggested that the high capture probabilities observed in impounded reaches of the Columbia River were also possible in the un-impounded reach below Bonneville Dam.

## 2001

During 2001, we estimated the survival of yearling and subyearling Chinook salmon at Bonneville Dam (Counihan et al. 2002b). The survival of paired releases of radio-tagged fish was evaluated using paired release-recapture models of Burnham et al. (1987). The original objectives for the 2001 survival evaluation at Bonneville Dam were altered because of the low water conditions present during 2001. The objectives were to provide estimates of survival for fish passing via all routes at Bonneville Dam and to provide estimates of survival of fish passing through the JBS (JBS) at powerhouse 2.

The survival of yearling Chinook salmon passing via all routes at Bonneville Dam (based on detections at Bonneville Dam of fish released near Hood River, OR and in the tailrace of Bonneville Dam) ranged from 0.85 to 1.05 . The average dam survival at Bonneville Dam for yearling Chinook salmon was estimated to be 0.937 ( $\mathrm{SE}=0.014$ ). Dam survival during the day was estimated to be $0.923(\mathrm{SE}=0.024)$ and night survival was estimated to be $0.949(\mathrm{SE}=$ 0.016 ). No significant differences were detected between day and night dam survival (one-tailed $t$-test, $P=0.19$ ) but the power associated with this unplanned test was low ( $1-\beta=0.22$ ). No significant relations were detected (linear regression, $P>0.10$ ) between the dam survival of yearling Chinook salmon and total river discharge, total turbine discharge, or total powerhouse 2 discharge.

Because of the low water year during 2001, appreciable spill at Bonneville Dam occurred during only the last 7 releases of radio-tagged yearling Chinook salmon. The nature of the 2001 operations allowed us to conduct a post-hoc comparison of the survival of yearling Chinook passing Bonneville Dam during periods of spill and no spill. Prior to the initiation of spill at Bonneville Dam, the survival of yearling Chinook passing through all routes at Bonneville Dam was estimated to be $0.928(\mathrm{n}=8, \mathrm{SE}=0.023)$ and after spill was initiated, was $0.946(\mathrm{n}=7, \mathrm{SE}$ $=0.015$ ). The survival of yearling Chinook salmon passing Bonneville Dam before and after spill was initiated was not statistically different (one tailed t-test, $P=0.27$ ). However, the power associated with this unplanned test was again low $(1-\beta=0.14)$.

The estimated survival of yearling Chinook salmon released through the powerhouse 2 JBS ranged from 0.78 to 1.1. The average estimated survival through the JBS was estimated to be 0.962 ( $\mathrm{SE}=0.023$ ). Survival through the JBS during the day was estimated to be 0.953 (SE = 0.039 ) and night survival was estimated to be 0.971 ( $\mathrm{SE}=0.027$ ). No significant differences were detected between day and night survival through the JBS (one tailed $t$-test, $P=0.35$ ) with power $(1-\beta=0.10)$. Similar to the results for dam survival, no significant relations were detected (linear regression, $P>0.10$ ) between the estimated juvenile bypass survival of yearling Chinook salmon and total river discharge, total turbine discharge, or total powerhouse 2 discharge.

We also estimated the survival of guided and unguided yearling Chinook salmon through Bonneville Dam's second powerhouse. The estimated average survival of turbine passed yearling Chinook was $0.929(\mathrm{SE}=0.02)$ and for non-turbine passed fish was $0.937(\mathrm{SE}=0.02)$. For turbine passed yearling Chinook, the average survival of fish passing during periods of spill was $0.900(\mathrm{SE}=0.032)$ and during periods of no spill was $0.954(\mathrm{SE}=0.024)$. The survival of turbine passed yearling Chinook passing during periods of spill and no spill were significantly different (one-tailed t-test, $P=0.098$ ). The average survival of non-turbine passed fish during periods of spill was $0.96(\mathrm{SE}=0.018)$ and for periods of no spill was $0.91(\mathrm{SE}=0.029)$. The difference between the average estimated survival during periods of spill and no spill for nonturbine passed fish was found to be significantly different (one-tailed $t$-test, $P=0.086$ ).

The dam survival of subyearling Chinook salmon passing via all routes at Bonneville Dam was based on the same release locations as those used for yearling Chinook salmon. The dam survival of subyearling Chinook salmon ranged from 0.73 to 1.08 . The estimated average dam survival was $0.902(\mathrm{SE}=0.036)$. The average dam survival during day releases was estimated to be $0.895(\mathrm{SE}=0.044)$ and during night releases was $0.910(\mathrm{SE}=0.066)$. No significant differences between day and night dam survival were detected (one-tailed $t$-test, $P=$ 0.42 ). No significant relations (linear regression, $P>0.10$ ) between total river discharge, total turbine discharge, and total powerhouse 2 discharge were detected. Subyearling Chinook salmon were also released through the powerhouse 2 JBS during 2001. Subyearling Chinook salmon JBS survival ranged from 0.62 to 1.28 . The average JBS survival was estimated to be 0.90 ( $\mathrm{SE}=$ 0.053 ). The average JBS survival for the day releases was estimated to be $0.870(\mathrm{SE}=0.089)$ and for night releases was $0.946(\mathrm{SE}=0.0374)$. The average survival estimates were not found to be significantly different between day and night releases (variance weighted one-tailed $t$-test, $P=0.23$ ). Significant relations (linear regression, $P<0.1$ ) between total river discharge, total turbine discharge, and total powerhouse 2 discharge were detected.

## 2002

Evaluations of radio-tagged yearling Chinook salmon survival through a Minimum Gap Runner (MGR) Turbine Unit and the downstream migration channel (DSM) at Bonneville Dam's powerhouse 1 were conducted during 2002 (Counihan et al. 2003). Using releases of radiotagged yearling Chinook salmon released as part of the survival evaluation at The Dalles Dam, and releases made below the outfall of the powerhouse 2 JBS, we also evaluated survival through the spillway and the first and second powerhouses.

The average survival of yearling Chinook salmon released into the MGR turbine unit at powerhouse 1, given the control group was released directly below the front roll of the turbine unit was estimated to be 1.06 ( $[1.00,1.12] 95 \%$ confidence interval) and through the MGR turbine unit at Bonneville Dam's powerhouse 1 given the control release below the powerhouse 2 JBS outfall was 1.01 ([0.98, 1.04] $95 \%$ confidence interval) during the 2002 migration season. We estimated that the survival of yearling Chinook salmon released into the downstream migration channel at Bonneville Dam's powerhouse 1 (control release below the powerhouse 2 JBS outfall) during 2002 ranged from 0.60 to 1.05 . The average survival was estimated to be 0.91 ([0.83, 0.99] 95\% confidence interval).

Using capture histories generated from the detections of radio-tagged yearling Chinook salmon released at The Dalles Dam and below the powerhouse 2 JBS outfall, we generated maximum likelihood estimates of the route-specific passage and survival probabilities for yearling Chinook salmon passing Bonneville Dam. The estimated dam and route-specific survival probabilities generated during 2002 differ from those generated during 2001 in that a different survival estimation model (i.e., the Route-specific Survival Model) was used to generate the estimates. The survival of yearling Chinook salmon through the Bonneville Dam spillway was estimated to be 0.977 ( $\mathrm{SE}=0.0135$; profile likelihood $95 \%$ confidence interval [ $0.951,1.000]$ ). For yearling Chinook salmon passing via powerhouse 1 , the estimated survival was $0.902(\mathrm{SE}=0.036$, profile likelihood $95 \%$ confidence interval $[0.824,0.965])$ and for yearling Chinook passing via powerhouse 2 the estimated survival was 0.993 ( $\mathrm{SE}=0.036$, profile likelihood $95 \%$ confidence interval [0.964, 1.021]). Yearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.977 ( $\mathrm{SE}=0.019$ ).

During 2004, the USGS evaluated the survival of radio-tagged yearling Chinook salmon and steelhead trout using paired releases through the ice and trash sluiceway and the MGR turbine unit at Bonneville Dam's powerhouse 1. Site-specific releases were made directly into the ice and trash sluiceway, the MGR turbine unit and in the tailrace at the front roll of the MGR turbine unit at powerhouse 1. Radio-tagged subyearling Chinook salmon survival was also estimated through the ice and trash sluiceway at powerhouse 1. Using releases of radio-tagged yearling and subyearling Chinook salmon and steelhead trout released as part of the survival evaluation at The Dalles Dam, and releases made into the tailrace of Bonneville Dam (below the powerhouse 2 JBS outfall), we were able to evaluate survival through the spillway, powerhouses 1 and 2, the corner collector and the JBS at powerhouse 2. These fish were also used in a posthoc paired release analysis to estimate survival of fish passing the spillway via spillbays with $7-\mathrm{ft}$ or 14 -ft flow deflectors.

## Methods

## Study Area

The study area (zone of inference, see: Peven et al. 2005) extended from The Dalles Dam at river kilometer (RK) 308 downriver to the I-205 Glenn Jackson Bridge (RK 181, Figure 1). Antenna arrays within the study area were located at Bonneville Dam (RK 235), Reed Island, (RK 200), Lady Island near the mouth of the Washougal River (RK 194), and the I-205 Glenn Jackson Bridge. All detection arrays spanned the breadth of the river channel. The telemetry system at Bonneville Dam was set up so that passage route could be determined (Evans et al. 2003).


Figure 1. Release and detection locations for Bonneville Dam survival evaluation during 2004. $\mathrm{R}=$ release locations, yellow ovals are locations of radio telemetry antenna arrays.

## Bonneville Dam

Bonneville Dam is located on the Columbia River at RK 235. The dam consists of two powerhouses and a single spillway, each separated by an island. Powerhouse 2 consists of eight vertical-axis turbine units, each with three intakes, and is located on the north side of the river, spanning from Cascade Island to the Washington shore. Powerhouse 1 consists of 10 verticalaxis turbines, each with three intakes, and is located on the south side of the river, spanning between Bradford Island and the Oregon shore. At both powerhouses, juvenile fish are guided away from turbines by submersible traveling screens into a fish collection channel. At powerhouse 2, fish enter the JBS, a 1.22 m diameter high-density polyethylene plastic pipe, and are transported downriver $3,530 \mathrm{~m}$ to an outfall where fish and water plunge approximately 4 m into the main river channel. At powerhouse 1, the collection channel flows through a monitoring facility within the powerhouse, and then fish and water plunge into the tailrace. Also at powerhouse 2, fish can enter the corner collector channel, located on the southeastern corner of powerhouse 2, where a moveable gate can be raised to allow about $5,000 \mathrm{cfs}$ of water to spill through it. The corner collector channel is a 4.57 m wide by 6.1 m deep concrete trough, and stretches 914.4 m to beyond the western tip of Cascade Island, where fish and water are released into a 15.2 m plunge pool.

## System antenna configuration

We used four types of data acquisition equipment to monitor underwater and aerial antennas at Bonneville Dam in 2004. Ninety-seven aerial antennas, 35 stripped coax antennas, and 124 underwater dipole antennas were linked to 34 Lotek SRX-400 receivers (SRX; Lotek

Engineering, Newmarket, Ontario), five Lotek DSP-500 digital spectrum processors (DSP; Lotek Engineering, Newmarket, Ontario), three Orion DSP receivers (Grant Systems Engineering, King City, Ontario, Canada), and three Multiprotocol Integrated Telemetry Acquisition Systems (MITAS; Grant Systems Engineering, King City, Ontario, Canada). Each SRX monitored a maximum of six aerial antennas. Orions, DSPs, and MITASs were used to monitor underwater antennas. Orions and DSPs were also used to monitor aerial antennas in some areas. The combination of these technologies allowed us to monitor passage and survival through all routes at Bonneville Dam.

Aerial antennas were positioned in three locations: 1) along the periphery of the forebay, 2) along the tailrace shoreline, and 3) along the corner collector flume (Figure 2). Aerial antennas were located in the forebay to detect fish within 100 m of the dam, in the tailrace to confirm fish passage for fish passage efficiency studies, and in the corner collector flume to detect fish passing through the corner collector. Underwater dipole and stripped coax antennas had limited ranges (about 6 m ) compared to aerial antennas ( 100 to 300 m depending on transmitter depth, receiver gain, and number of antenna elements). Underwater antennas allowed us to obtain fine scale fish behavior information by limiting the range of signal detection.

Three MITAS systems were incorporated at B1, B2, and the spillway (Figure 3). The MITAS at B1 was composed of 22 underwater stripped coax antennas and one aerial antenna. Twenty stripped coax antennas were positioned mid-channel in the sluiceway, two at each unit, to monitor unit-specific sluiceway entrance and passage through the sluiceway. In addition, two stripped coax antennas and one aerial antenna were placed at the outfall of the sluiceway to confirm sluiceway passage.

The MITAS at B2 was composed of 61 underwater antennas. Forty-eight dipole underwater antennas attached to the submersible traveling screens monitored unguided turbine passage: Two dipole antennas were mounted to the bottom of each of three submersible traveling screens in front of each of eight turbine units. Antennas from each of three gatewell slots per unit were combined to provide turbine unit specific passage information. Nine stripped coax antennas placed within the downstream salmonids migrant channel (DSM) monitored guided fish passage. One antenna was located just downstream of each "C-slot" gatewell orifice and one additional antenna was located at the terminus of the DSM. Four dipole underwater antennas monitored approach and entrance of fish to the corner collector.

The spillway MITAS consisted of 72 underwater antennas attached to the forebay pier noses. Each spillbay had four antennas; two antennas on each piernose at about 4.5 m below mean pool level and 2 antennas at about 10.5 m below mean pool level. All four antennas in each spillbay were combined to one input to provide spillbay-specific passage.


Figure 2. Plan view of aerial antenna coverage during spring 2004 at Bonneville Dam's: (a) second powerhouse (B2) and spillway (SPI); and (b) first powerhouse (B1).


Figure 3. Plan view of underwater antenna coverage during spring 2004 at Bonneville Dam's: (a) second powerhouse (B2) and spillway (SPI); and (b) first powerhouse (B1).

## Radio Transmitters

The radio telemetry tags used in this study were pulse-coded transmitters (tags) manufactured by Lotek Engineering, Inc, (Newmarket, Ont.). Transmitters operated at frequencies between 150.280 and 150.800 MHz and used a pulse-coding scheme with 212 unique codes per frequency that allow each individual fish to be recognized. A radio signal was emitted every 2 seconds. Two sizes of these transmitters were used to accommodate the different sizes of the spring and summer migrants. Transmitters implanted in yearling Chinook salmon and steelhead trout were 7.3 mm in diameter $\times 18.0 \mathrm{~mm}$ in length and weighed 1.4 g in air (Lotek Wireless model 3 KM ). Expected battery life was 8 days for the 3 KM tags. Transmitters implanted in subyearling Chinook salmon were 6.3 mm wide x 14.5 mm length x 4.5 mm high and weigh 0.85 g in air (Lotek Model NTC 3-1, expected battery life was 8 days).

## Fish tagging and releasing

Juvenile salmonids to be implanted with radio transmitters and released at The Dalles Dam were collected from the juvenile collection and bypass facility at John Day Dam at night and in the morning. After collection, fish were transported to The Dalles Dam and were tagged 12 to 36 h later. Juvenile salmonids to be released at Bonneville Dam were collected at Bonneville Dam's powerhouse 2 juvenile bypass monitoring facility and were held approximately 12 to 36 hours prior to tagging. Fish were considered suitable for tagging if they were free of injuries, severe descaling, external signs of gas bubble trauma, or other abnormalities. Fish size criteria were also established such that the radio tag weight in air would not exceed $6.5 \%$ of a fish weight in air. For yearling Chinook salmon and steelhead trout implanted with a Lotek Wireless model 3 KM tag (weight $=1.4 \mathrm{~g}$ in air) the minimum weight for tagging was 21.5 g , and for subyearling Chinook salmon implanted with a Lotek Model NTC 31 , "nanotag" (weight $=0.85 \mathrm{~g}$ in air) the minimum weight was 13 g . Transmitters were gastrically implanted using the methods of Martinelli et al. (1998). Fish were held 18 to 28 h after tagging to check for spit tags and mortalities before being released. Fork lengths and weights of the various release groups and number of spit tags and mortalities by release are presented in Appendix 1.

To evaluate survival through the MGR and the ice and trash sluiceway, approximately 25 radio-tagged yearling Chinook salmon and steelhead trout (per release) were released at each location (MGR, MGR front roll tailrace, and ice and trash sluiceway) during 16 releases and 19 subyearling Chinook salmon (per release) into the ice and trash sluiceway during 32 releases (see Appendix 1 for the exact times, dates, and numbers of fish released during the study). Control fish for the MGR evaluation were released from the tailrace deck from a tank connected to a 10 cm diameter flexible tube that extended from the release tank to a barge anchored at the downstream end of the front roll created by the outflow from the turbine. The flexible tube was connected to an aluminum pipe on the barge that terminated 10 ft below the water surface. Water pumped into the tube created a siphon that actively transported fish to the barge where they were discharged into the river channel at the same depth that fish exited the turbine.

To evaluate the route-specific survival and dam survival at Bonneville Dam, radio-tagged fish were released from The Dalles Dam. Radio-tagged fish were released in the The Dalles Dam tailrace from a boat approximately 550 m downriver of the dam beneath the I-197 bridge as well as through and below the ice and trash sluiceway. Releases into The Dalles Dam tailrace
occurred twice a day at approximately 1300 and 0100 hrs and releases at The Dalles Dam sluiceway occurred once a day at $0700,1300,1900$, and 0100 hrs . The sluiceway releases were randomized and equally allocated among the four release times for the study period. Release times were the midpoints of 6-hour blocks of divergent discharge conditions seen in diel discharge patterns at The Dalles Dam (see Appendix 1 for exact dates, times and number of fish per release). A control group was released by boat, mid-channel in the Bonneville Dam tailrace 2 km downstream and below the powerhouse 2 JBS outfall. This control group was also the control group for the paired release-recapture evaluations through the MGR and ice and trash sluiceway. Bonneville Dam tailrace releases occurred twice a day at approximately 1300 and 0100 hrs. Releases into the tailrace at Bonneville Dam were timed to approximately coincide with the arrival of fish released in the tailrace at The Dalles Dam. The timing of the releases was determined using a regression equation based on Zabel et al. (1997) to estimate travel times. Specifically travel time for Bonneville reservoir was estimated such that:

$$
\text { Travel rate }(\mathrm{km} / \mathrm{d})=49.902+0.1309 *(\text { Discharge })
$$

The relation between travel time and river discharge was developed by evaluating travel time and discharge data collected during previous studies. For the 2004 migration season, we used predicted discharge data obtained from Kyle Dittmer of Columbia River Inter-Tribal Fish Commission (Kyle Dittmer, personal communication). We then consider the output of this exercise in combination with crew logistics and the dam operation and test treatment schedule at each study site to determine our release times. Since fish released from The Dalles Dam tailrace were also part of a survival study at The Dalles Dam, further coordination was necessary to accommodate releases made in the John Day Dam tailrace. During the summer, we also assigned releases equally to spill block treatments at Bonneville Dam. Four-day spill blocks were designed by randomly assigning a 2-day continuous block of 24 h spill with 56 kcfs during the day ( 0400 to 2200 hrs ) and night spill ( 2200 to 0400 hrs ) until the $125 \%$ total dissolved gas cap was reached in the tailrace alternating with 23 kcfs for 24 hr for 2-days. Treatment spill conditions were initiated at the dam prior to the start of releases and continued until after the last fish passed Bonneville Dam.

## Converting radio signals into detection histories

After data collection, radio signals have to be interpreted and converted into detection histories. Aerial and underwater antennas attached to data logging equipment will often record spurious radio signals or "noise" and designate them as such, or misinterpret other radio signals (e.g., from cars or trucks) and label them with fish channel and code designations. We performed automated data processing using Statistical Analysis System (SAS) software to separate spurious radio signals from true radio signals and assign passage and location designators. The following criteria were used to classify data records as noise:

1. Records composed of invalid channel and code combinations, typically a result of erroneous radio transmissions (noise) that overlap with the radio frequencies that we are monitoring.
2. Records logged before a fish's release.
3. Records below an empirically determined signal strength threshold for each aerial and underwater array at the dam.
4. Single records recorded within a 20 min period on an array of nearby adjacent antennas (e.g. entrance, forebay, tailrace, or survival gate arrays).
5. A group of fewer than 3 records within a $60-\mathrm{min}$ interval on an individual entrance, forebay, tailrace, or exit station receiver previously not classified as noise by criteria 1 through 4, that are unsupported by at least two other valid records among these areas during the hour interval of detection or the hour before and after detection.
6. Records not classified as noise by criteria 1 through 4 and detected on an array of nearby antennas in the forebay or tailrace that were recorded more than an hour after the previous valid record at the same antenna array.
7. Records on the MITAS aerial tailrace array over a 3-h interval, not classified as noise by the above criteria, and unsupported by any other valid entrance, forebay, tailrace, or exit detections during the same time period.

Once all times and locations of interest (events) were electronically assigned, individual fish histories were verified using criteria derived from manually-proofed radio-telemetry data obtained in past years for the same species. A fish's event history was considered potentially suspect if 1) the travel time between release and first forebay, tailrace, or exit detection, or travel time between sequential events was less than the $5^{\text {th }}$ or greater than the $95^{\text {th }}$ percentiles of past data from a similar flow year, 2) forebay, tailrace, and exit residence times exceeded the $95^{\text {th }}$ percentile of similar past year's metrics, or 3) a fish's events were chronologically or geographically out of order. Fish whose event histories were suspect because of one or more of the above criteria were flagged to be manually proofed and reconciled with the electronic proof prior to further analyses. In addition to the flagged files, a random $10 \%$ of the fish from nonflagged files were manually examined by separate proofing staff and then reconciled by another staff member if any disagreement in either the time of passage or passage location were noted between the electronically assigned events and the manually assigned events. Once individual fish histories were verified the capture histories were generated for each passage scenario, indicating detection at the release location, detection at the dam, and detection down stream of the dam by assigning a 1 for detection and a 0 for not detected at antenna arrays.

## Statistical methods

## Paired release-recapture model

We used the paired release-recapture models of Burnham et al. (1987) to estimate the survival of yearling Chinook salmon and steelhead trout through an MGR turbine unit and to estimate the survival of yearling and subyearling Chinook salmon through the ice and trash sluiceway at Bonneville Dam's powerhouse 1. The paired release-recapture model was also used to estimate the survival of fish passing via spillbays equipped with flow deflectors located at 7 feet above mean sea level and at 14 feet above mean sea level at Bonneville Dam during the different dam operations. Paired release groups for the spillbay analysis were formed post-hoc from radio-tagged fish released at The Dalles Dam and passing Bonneville Dam spillway (treatment groups) with fish released below the powerhouse 2 JBS outfall (control groups). Spillbays with deflectors at 7 feet above mean sea level were located on the north (bays 1-3) and
south (bays 16-18) end of the spillway and spillbays with deflectors at 14 feet above sea level were located in the middle, bays 4-15.

## Model Assumptions

There are assumptions associated with using the paired release-recapture model to estimate survival; some are biological and some pertain to the statistical models (Burnham et al. 1987, Skalski et al. 1998, Skalski 1999a). The validity of some of the assumptions listed below can be evaluated using statistical tests and others can be met through careful consideration of fish collection, holding, tagging, and detection techniques. The assumptions are the following:

A1. Individuals marked for the study are a representative sample from the population of interest.

A2. Survival and capture probabilities are not affected by tagging or sampling (i.e., tagged animals have the same probabilities as untagged animals).

A3. All sampling events are "instantaneous" (i.e. sampling occurs over a short time relative to the length of the intervals between sampling events).

A4. The fate of each tagged individual is independent of the fate of all others.
A5. All individuals alive at a sampling location have the same probability of surviving until the end of that event.

A6. All tagged individuals alive at a sampling location have the same probability of being detected on that event.

A7. All tags are correctly identified and the status of fish (i.e. alive or dead) is correctly identified.

To evaluate assumption A1, we monitor the timing and lengths of run-of-river fish sampled at the John Day Dam and Bonneville Dam smolt monitoring facilities. We compare this to our sampling dates and lengths of radio-tagged fish to assess how representative the radiotagged fish are to run-of -river fish. We also conduct statistical tests to evaluate assumptions A5 and A6 using tests developed by Burnham et al. (1987). Burnham et al. (1987) presents a series of tests of assumptions named Test 2 that examine whether upstream or downstream detections affect downstream survival and/or detection. To examine whether upstream capture histories affect downstream survival and/or capture, Burnham et al. (1987) present a series of tests called Test 3. Another factor that may have implications pertaining to these assumptions is the fact that radio-tags have a limited and varied battery life. Therefore, the tag failure rate will affect detection probabilities, depending on travel time and amount of time tags are operational prior to release. To address the probability of tag failure at detection arrays, we performed a tag life study (Appendix 2) to determine the potential for bias in survival estimates caused by tag failure. The assumption A7 is evaluated by releasing dead radio-tagged fish throughout the season with live radio-tagged fish.

Survival was estimated from paired releases by the expression:

$$
\begin{equation*}
\hat{S}=\frac{\hat{S}_{11}}{\hat{S}_{21}} \tag{1}
\end{equation*}
$$

with a variance estimate based on the Delta method (Seber 1982) of:

$$
\begin{align*}
& \operatorname{Var}\left(\hat{S}_{W}\right) \cong\left(\frac{\hat{S}_{11}}{\hat{S}_{21}}\right)^{2}\left[\frac{\operatorname{Var}\left(\hat{S}_{11}\right)}{\hat{S}_{11}^{2}}+\frac{\operatorname{Var}\left(\hat{S}_{21}\right)}{\hat{S}_{21}^{2}}\right]  \tag{2}\\
& \cong \hat{S}_{W}^{2}\left[\hat{C} V\left(\hat{S}_{11}\right)^{2}+\hat{C} V\left(\hat{S}_{21}\right)^{2}\right]
\end{align*}
$$

where $\hat{S}_{11}$ = survival estimates for fish released above the project of interest or into a particular route and $\hat{S}_{22}=$ fish released below the project, and where

$$
\hat{C} V(\hat{\theta})=\frac{\sqrt{\operatorname{Var}(\hat{\theta})}}{\theta} .
$$

In order to estimate $S$, the survival $S_{11}$ is assumed to be of the form:

$$
S_{11}=S \cong S_{21}
$$

leading to the relationship

$$
\begin{equation*}
\frac{S_{11}}{S_{21}}=\frac{S \cdot S_{21}}{S_{21}}=S \tag{3}
\end{equation*}
$$

The equality (3) suggests two additional assumptions for valid survival estimation using the paired release-recapture protocol.

A8. Survival in the upriver segment $(S)$ is conditionally independent of survival in the lower river segment.

A9. Releases $\left(R_{1}\right)$ and $\left(R_{2}\right)$ have the same survival probability in the lower river segment $\left(S_{21}\right)$.

The assumption of downstream mixing was tested at each downstream array. An $\mathrm{R} \times \mathrm{C}$ contingency table test of homogenous recoveries over time was performed using a table of the form:


For each paired-release ( $\mathrm{R}_{1}$ and $\mathrm{R}_{2}$ ), a chi-square test of homogeneity was performed at each downstream array. Tests were performed at $\alpha=0.10$. Because there were multiple releases and tests across paired releases, the Type I error rates were adjusted for an overall experimentalwise error rate pertaining specifically to each paired release-recapture evaluation conducted at Bonneville Dam (Dunn-Sidak method, Sokal and Rohlf, 1995).

Inferences regarding mixing will be largely based on the sequential use of likelihood ratio tests. In any given survival estimation scenario, a number of potential models will be generated and subsequently evaluated (Burnham et al. 1987, Lebreton et al. 1992). Forward-sequential and reverse-sequential procedures will be used to find the most parsimonious statistical model that adequately describes the downstream survival and capture processes of the paired release. The most efficient estimate of survival will be based on the statistical model for the paired releases that properly share all common parameters between release groups.

We evaluated $t$-tests to compare the estimated survival of subyearling Chinook salmon released into the ice and trash sluiceway at powerhouse 1, Bonneville Dam during both the 23 kcfs 24 h spill and 56 kcfs Day/TDG Night spill operations. The specific hypothesis tested was:

$$
\begin{aligned}
& H_{0}: S_{\text {Ice and trash } 24 \mathrm{~h} 23 \mathrm{kcfs}}=S_{\text {ice and trash } 56 \mathrm{kcfs} \text { Day/TDG Night }} \\
& H_{A}: S_{\text {Ice and trash } 24 \mathrm{~h} 23 \mathrm{kcfs}} \neq S_{\text {ice and trash } 56 \mathrm{kcfs} \text { Day/TDG Night }}
\end{aligned}
$$

Bartlett's, Brown-Forsythe, and Levene's tests for equal variance were evaluated for each comparison and where suggested by the results of these tests, variance weighted $t$-tests were used.

## Estimable Parameters

The release and detection schemes used during 2004 allowed us to generate the survival and capture probabilities shown in Figure 4 for all site-specific releases at Bonneville Dam.


Survival through the ice and trash sluiceway and MGR: $\hat{S}_{\text {IandTsluice }}=\frac{\hat{S}_{11}}{\hat{S}_{21}}$

Figure 4. Schematic of estimable capture and survival probabilities ( $S=$ survival estimate, $p=$ capture probability, and $\lambda=\mathrm{S} \cdot \mathrm{p}$ ) from site-specific releases ( $R_{\text {ROUTE }}$ ) at Bonneville Dam and in the Bonneville Dam tailrace.

## Route-Specific Survival Model

## Model Assumptions

The assumptions associated with the Route-Specific Survival Model (RSSM) are described in detail in Skalski et al. (2002) and are similar to those for the paired release-recapture model of Burnham et al. (1987). Assumptions of the RSSM are:

A1. Individuals marked for the study are a representative sample from the population of interest.

A2. Survival and capture probabilities are not affected by tagging or sampling (i.e., tagged animals have the same probabilities as untagged animals).

A3. All sampling events are "instantaneous" (i.e., sampling occurs over a short time relative to the length of the intervals between sampling events).

A4. The fate of each tagged individual is independent of the fate of all others.
A5. All tagged individuals alive at a sampling location have the same probability of surviving until the end of that event.

A6. All tagged individuals alive at a sampling location have the same probability of being detected.

A7. All tags are correctly identified and the status of fish (i.e., alive or dead) is correctly identified.

A8. Survival in the upriver segment $(S)$ is conditionally independent of survival in the lower river segment.

A9. Both the upstream and downstream release groups, within a paired release, experience the same survival probability in the segment of the river that they travel together.

Skalski et al. (2002) identified two additional assumptions associated with the RSSM:
A10. Routes taken by the radio-tagged fish are known without error.
A11. Detections in the primary and secondary antenna arrays within a passage route are independent.

Skalski et al. (2002) suggest that assumption A10 can be qualitatively assessed by examining radio telemetry detection histories to determine whether inconsistencies in individual fish detection histories exist. Skalski et al. (2002) use an example of a situation where a radiotagged fish is detected in the upstream array of a route and then in the downstream array of another route, resulting in uncertainty in the route taken. That is, they used aerial antennas that
monitored the tailrace area to help determine passage. Similar to the radio-telemetry system used in Skalski et al. (2002), the double array we employed at Bonneville Dam consisted of aerial and underwater telemetry systems that interrogated fish in the immediate forebay area of each particular route, with the exception of the JBS and corner collector where underwater antennas were placed at two locations within these structures. However, while we did have a radio-telemetry system monitoring the tailrace area of each route, we did not consider detections in the tailrace when determining passage routes.

Skalski et al. (2002) determined that while assumption A11 is necessary for valid estimation of in-route detection probabilities, the assumption cannot be empirically assessed with the data collected during this type of study. Rather, they suggest that the detection fields of the primary and secondary arrays should be located in a way that fish detected in one array does not have a higher or lower probability of being detected in the secondary array than the primary array. Further, they suggest that this is best accomplished by having independent receivers for each antenna array and by having the detection field of at least one array encompass the entire passage route. The arrays we deployed at Bonneville Dam powerhouses 1 and 2, the JBS, corner collector, and spillway adhere to these requirements.

## Parameter Estimation

The double radio-telemetry array systems that we deployed at Bonneville Dam allowed us to estimate route-specific detection probabilities. In turn, these route-specific detection probabilities can be incorporated into a statistical analysis that will extract route-specific passage and survival (Skalski et al. 2002). The following parameters were defined for the construction of the RSSM used at Bonneville Dam: $S_{\text {POOL }}$, survival from the release location at The Dalles Dam; E, probability that fish will pass via the spillway; PH2, conditional probability of passing via the second powerhouse, given that fish were going to either the first or second powerhouse; B2CC, conditional probability of passing via the corner collector, given that fish were going to powerhouse $2 ; \mathrm{B}_{2} \mathrm{JBS}$, conditional probability of passing via the JBS, given that fish were going to powerhouse $2 ; P_{B 2 C C}$, the corner collector primary array detection probability ( $q_{B 2 C C}=1-\mathrm{P}$ $\left.{ }_{B 2 C C}\right) ; P^{\prime}{ }_{B 2 C C}$, the corner collector secondary array detection probability $\left(q^{\prime}{ }_{B 2 C C}=1-\mathrm{P}^{\prime}{ }_{B 2 C C}\right)$; $P_{B 2 J B S}$, the JBS primary array detection probability $\left(q_{B 2 J B S}=1-P_{B 2 J B S}\right) ; P^{\prime}{ }_{B 2 J B S}$, the JBS secondary array detection probability ( $\left.q^{\prime}{ }_{B 2 J B S}=1-P^{\prime}{ }_{B 2 J B S}\right) ; P_{B 2 T u r b}$, the second powerhouse turbines primary array detection probability $\left(q_{B 2 T u r b}=1-P_{B 2 T u r b}\right)$; $P^{\prime}{ }_{B 2 T u r b}$, the second powerhouse turbines secondary array detection probability ( $\left.q^{\prime}{ }_{B 2 T u r b}=1-P^{\prime}{ }_{B 2 T u r b}\right) ; P_{P H I}$, the first powerhouse primary array detection probability $\left(q_{P H I}=1-P_{P H I}\right) ; P_{P H I}$, the first powerhouse secondary array detection probability $\left(q^{\prime}{ }_{P H I}=1-P^{\prime}{ }_{P H I}\right) ; P_{S P I L L}$, spillway primary array detection probability $\left(q_{\text {SPILL }}=1-P_{S P I L L}\right) ; P^{\prime}{ }_{S P I L L}$, spillway secondary array detection probability $\left(q^{\prime}{ }_{S P I L L}=1-P^{\prime}{ }_{S P I L L}\right) ; S_{S P I L L}$, spillway survival probability; $S_{B 2 C C}$, the corner collector survival probability; $S_{B 2 J B S}$, the JBS survival probability; $S_{B 2 T u r b}$, the second powerhouse turbines survival probability; $S_{P H I}$, the first powerhouse survival probability; $\lambda$, the joint probability of surviving and being detected at the arrays below Bonneville Dam. The releases made at The Dalles $\operatorname{Dam}\left(R_{l}\right)$ and the releases made below the second powerhouse JBS outfall $\left(R_{2}\right)$ were interrogated at three arrays below Bonneville Dam, the furthest downriver being an array deployed on the I-205 Bridge (Figure 1). A branching process was used to model the migration and survival of releases $R_{1}$ and $R_{2}$ (Figure 5). Additional details regarding the methodology used in the formulation of the RSSM and the estimation of the associated
parameters can be found in Skalski et al. (2002). For the RSSM survival probabilities both standard errors and profile likelihood $95 \%$ confidence intervals are reported (Skalski et al. 2002).

The route-specific survival and passage probabilities can be combined using maximum likelihood estimation to estimate survival through the dam. The survival through Bonneville dam was estimated from the expression:

$$
\begin{aligned}
& \widehat{S}_{D A M}=(1-\widehat{E})(1-\widehat{P} H 2) \widehat{S}_{P H 1}+\widehat{E} \widehat{S}_{S P I L L}+(1-\widehat{E})(\hat{P} H 2)(\widehat{B} 2 C C) \widehat{S}_{B 2 C C}+ \\
& (1-\widehat{E})\left(\widehat{P} H_{2}\right)(1-\widehat{B} 2 C C)(1-\widehat{B} 2 J B S) \widehat{S}_{B 2 T_{u r b}}+(1-\widehat{E})\left(\widehat{P} H_{2}\right)(1-\widehat{B} 2 C C)(\widehat{B} 2 J B S) \widehat{S}_{B 2 J B S}
\end{aligned}
$$

The variance for the dam survival estimate was estimated using the delta method (Seber 1982, pp 7-9). All of the route-specific survival and passage probabilities were estimated with the USER (User Specified Estimation Routine) software developed at the University of Washington (Lady et al. 2003; see: http://www.cqs.washington.edu/paramEst/USER/).

## Comparisons of RSSM estimates between treatments

Z-tests were performed to assess the differences of route-specific survival estimates of subyearling Chinook salmon between treatments (i.e. spill operations) for each passage route (John Skalski and Jim Lady, University of Washington, personal communication). The hypotheses tested for subyearling Chinook salmon passing via the powerhouse 1, powerhouse 2 turbines, spillway, powerhouse 2 corner collector and the JBS at powerhouse 2, and through all routes were:

$$
\begin{aligned}
& H_{O}: S_{23 \mathrm{kcfs} \text { Day/23 kcfs Night }}=S_{56 \mathrm{kcfs} \text { Day/TDG Night }} \\
& H_{A}: S_{23 \mathrm{kcfs} \text { Day/23 kcfs Night }} \neq S_{56 \mathrm{kcfs} \text { Day/TDG Night }}
\end{aligned}
$$



Figure 5. The estimable parameters (see p. 49 for definitions) for the route-specific survival model using the proposed release and detection schemes for 2004. Included in the detection scheme is a double radio-telemetry array at Bonneville Dam.

## Results

## Releases of dead radio-tagged fish

We determined that the standard operating procedure (SOP) for euthanizing fish for planned dead fish releases was not properly implemented. An SOP has been established for the killing and handling of dead fish that requires an overdose of MS-222 for 30 minutes while fish are confined below the water surface with a screen, followed by pithing, prior to release. During the field season there was turnover in the crew that implemented this procedure and the proper steps were not taken to communicate the SOP to all staff thus, the SOP was not consistently implemented. In particular, the screening procedure that prevented fish from gulping air while being subjected to MS-222 was not implemented. Consequently, though several dead fish were detected at the downstream detection arrays, we are unable to determine whether these detections were an artifact of the protocol breach, a result of particular environmental conditions that were more conducive to transporting dead radio-tagged fish past the downstream arrays, or reflect the estimated probability that a dead radio-tagged fish is detected at our downstream arrays. In past years, and for the majority of the release during 2004, dead fish were not detected at the arrays downstream of Bonneville Dam.

As called for in our Quality Assurance Plan, we are implementing corrective measures to ensure this protocol breach does not occur again. Specifically, we have stipulated the nature of staff that implements the protocol (i.e., only shift leaders), we have increased the rigor of the training program for this procedure, have separated the training from the general tagging procedure to be a stand-alone component, and have implemented quality control checks in the field to ensure the procedure is properly implemented. We will evaluate the validity of this assumption during 2005 and if warranted will provide an errata to this report describing our observations.

The detections of dead radio-tagged fish can be grouped into two categories. The first category consists of the detected dead radio-tagged fish with long travel times (Table 1). One of the five dead radio-tagged yearling Chinook salmon had a travel time that was $\geq$ the $99.8^{\text {th }}$ percentile of travel times for fish detected at our furthest downstream detection array and another had the longest travel time ( $100^{\text {th }}$ percentile) to the first and second detection arrays. The second category is comprised of two dead yearling Chinook ( $10.1^{\text {th }}$ and $2.9^{\text {th }}$ percentile) and the only dead subyearling Chinook salmon ( $6.7^{\text {th }}$ percentile) detected that had travel times on the opposite (e.g., fast) end of the distribution (Table 1). That the dead radio-tagged fish could be transported downstream at a slower rate than live fish and eventually be detected seems plausible. However, developing the rationale where the dead radio-tagged fish travel at a rate faster than the live fish through this river reach is more difficult.

To assess the potential effect of this assumption violation on the estimates generated during 2004, we took two approaches. First, to account for the possibility that our detections of radio-tagged fish with long travel times could be dead fish, we eliminated all fish with travel times $>$ the $99.7^{\text {th }}$ percentile and recalculated the estimates (Appendix 2). Further, we generated survival estimates with and without releases that occurred 24 h before or after releases associated with the dead fish detections (see Appendix 2). The rationale for this approach is: since dead fish were not detected consistently throughout the season, then the conditions (i.e., discharge, water velocities, dam operations) present during releases where dead fish were detected (and
conservatively $\pm 24 \mathrm{~h}$ of the release) may have been conducive for allowing the transport of dead radio-tagged fish to our detection sites below Bonneville Dam. Thus, all fish passing through the study area during this time (i.e., $\pm 24 \mathrm{~h}$ of the release) were not included in the data set evaluated in the alternate analysis.

Table 1. Summary of the release dates, times and travel times to the radio telemetry detection arrays (Gates) for radio-tagged dead fish detections. The percentile ranking of the dead fish, with respect to the travel time distribution for all live radio-tagged fish released into the Bonneville Dam tailrace below the powerhouse 2 juvenile bypass outfall 2004, is presented for each detection array.

| Release |  | Species | Gate 1 <br> (hours) | Rank <br> (\%) | Gate 2 <br> (hours) | Rank <br> (\%) | Gate 3 (hours) | Rank <br> (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date | Time |  |  |  |  |  |  |  |
| 8-May | 10:11 | Yearling Chinook | 111.0 | 100.0 | 114.0 | 100.0 |  | NA |
| 8-May | 21:46 | Yearling Chinook | NA | NA | NA | A | 73.7 | 99.8 |
| 9-May | 10:11 | Yearling Chinook | NA | NA | NA | NA | 19.0 | 80.1 |
| 14-May | 21:59 | Yearling Chinook | 7.0 | 14.4 | 9.1 | 11.4 | 12.3 | 10.1 |
| 31-May | 00:01 | Yearling Chinook | NA | NA | NA | NA | 10.7 | 2.9 |
| 22-May | 10:55 | Steelhead trout | NA | NA | NA | NA | 153.0 | 100.0 |
| 5-July | 12:59 | Subyearling Chinook | 7.1 | 2.9 | 9.6 | 4.7 | 13.7 | 6.7 |

## Run timing and radio telemetry tagging dates

One assumption of the mark-recapture models used in this study is that individuals marked constitute a representative sample from the population of interest. However, there are technological (i.e. tag size and battery life) or logistical (i.e. availability of fish of appropriate sizes, inexperience of conducting radio-telemetry studies during late-August and at the temperatures commonly seen during this month) limitations dictating the size of fish tagged and the timing of the study. Fish size criteria were established such that the radio tag weight in air would not exceed $6.5 \%$ of a fish weight in air. For yearling Chinook salmon and steelhead trout implanted with a Lotek Wireless model 3 KM tag (weight $=1.4 \mathrm{~g}$ in air) the minimum weight for tagging was 21.5 g (corresponding estimated length of 130 mm ), and for subyearling Chinook salmon implanted with a Lotek Model NTC 3-1, "nanotag" (weight $=0.85 \mathrm{~g}$ in air) the minimum weight was 13 g (corresponding estimated length of 110 mm ). Due to these limitations the resultant data needs to be viewed critically in the context of these assumptions.

In an effort to fulfill this assumption, radio-telemetry tagging dates are designed to encompass the run timing for run-of-river fish. The Fish Passage Center (see: www.fpc.org) maintains passage index data for fish passing Bonneville Dam powerhouse 2. The passage index is the number of fish sampled divided by the sample rate divided by the proportion of water passing through the sampling system. For yearling Chinook salmon radio telemetry tagging started at approximately $35 \%$ of the run and ended at $90 \%$ (Figure 6), and for steelhead trout tagging started at about $18 \%$ of the run and ended at $95 \%$ of the run (Figure 7). For subyearling Chinook salmon radio telemetry tagging started at approximately $55 \%$ of the run and ended near $98 \%$ of the run (Figure 8). Approximately $40 \%$ of the subyearling Chinook salmon run was composed of hatchery releases which occurred prior to 10 May.


Figure 6. Yearling Chinook salmon daily passage index at Bonneville Dam, powerhouse 2. The vertical bars represent the passage index (see: www.fpc.org) for a given day. Vertical lines represent the start and end dates for radio telemetry tagging.


Figure 7. Steelhead trout daily passage index at Bonneville Dam, powerhouse 2 The vertical bars represent the passage index (see: www.fpc.org) for a given day. Vertical lines represent the start and end dates for radio telemetry tagging.


Figure 8. Subyearling Chinook salmon daily passage index at Bonneville Dam, powerhouse 2. The vertical bars represent the passage index (see: www.fpc.org) for a given day. Vertical lines represent the start and end dates for radio telemetry tagging.

## Radio-tagged fish size relative to run-of-river fish

We obtained fork length data for run-of-river fish sampled at the John Day Dam and Bonneville Dam smolt monitoring facilities and compared it to fork length data for radio-tagged fish obtained from each of these sites. For yearling Chinook salmon the radio-tagged fish were of very similar sizes to the run-of-river fish (Figure 9). We observed that less than $10 \%$ of the sampled run was below the 130 mm size criteria throughout the season. For steelhead trout very few run-of-river fish fell below the 130 mm size criteria (Figure 10). The average radio-tagged steelhead trout was 12 to 20 mm larger in size than the run-of-river fish. The run-of-river subyearling Chinook salmon were unusually small during the summer of 2004. The mean length at the sampling facilities ranged from 94 to 97 mm from 20 June to 22 July, while the mean radio-tagged fish length was 116 to 117 mm (Figure 11). Furthermore, less than $10 \%$ of the run-of-river fish sampled at the facilities were larger than 110 mm .


Figure 9. The distribution of fork lengths of run-of-river yearling Chinook salmon that were sampled at the John Day Dam and Bonneville Dam smolt monitoring facilities and fish tagged with MCFT-3KM radio transmitters (Lotek Engineering, Newmarket, Ontario) and released at The Dalles Dam and Bonneville Dam during 2004. Based on length to weight regression equations and tag weight to fish weight criterion, fish to the left of the dashed lines were too small to be tagged with the transmitters.


Figure 10. The distribution of fork lengths of run-of-river steelhead trout that were sampled at the John Day Dam and Bonneville Dam smolt monitoring facilities and fish tagged with MCFT3KM radio transmitters (Lotek Engineering, Newmarket, Ontario) and released at The Dalles Dam and Bonneville Dam during 2004. Based on length to weight regression equations and tag weight to fish weight criterion, fish to the left of the dashed lines were too small to be tagged with the transmitters.


Figure 11. The distribution of fork lengths of run-of-river subyearling Chinook salmon that were sampled at the John Day Dam and Bonneville Dam smolt monitoring facilities and fish tagged with MCFT-3KM radio transmitters (Lotek Engineering, Newmarket, Ontario) and released at The Dalles Dam and Bonneville Dam during 2004. Based on length to weight regression equations and tag weight to fish weight criterion, fish to the left of the dashed lines were too small to be tagged with the transmitters.

## Tag-life performance for determining potential bias of survival estimates

An assumption of release-recapture models used to estimate survival is that all live tagged individuals have the same probability of being detected at downstream detection arrays. A factor that may influence this assumption is that radio-tags have a limited and varied battery life. Therefore, the tag failure rate will affect detection probabilities, depending on travel time of a tagged fish and the time a tag is on prior to release. Thus, survival estimates may be biased if the radio-tag expires prior to a fish exiting all the detection arrays. Radio-tags may expire before fish exit the study area due to equipment malfunction, extended travel time of fish during periods of low discharge, or extended length of time tag was on prior to release. Information obtained by a tag-life study (see Appendix 3) can be used to adjust survival estimates using the probability that a tag will expire prior to fish exiting the study area (Townsend et al. 2004, Cowen and Schwarz 2005).

We determined that the probability of a tag being operational at downstream arrays was high, with all probabilities greater than $99.9 \%$ (Table 2). Probabilities were higher for the summer study than for the spring study. The cumulative arrival distributions plotted with the Gompertz model over time shows that tagged juvenile salmonids passed through downstream detection arrays several days before tag-failure was substantial for both fish released from The Dalles Dam tailrace and Bonneville Dam tailrace (Appendix 3, Figure A3.3). Townsend et al. (2004) found that the probability of a tag being operational at downstream detection arrays was quite high ( $>98 \%$ ), therefore, the adjusted survival estimate ( 0.9387 ) changed very little from the unadjusted estimate $(0.9339)$ having a difference of just 0.0048 . Our probabilities being greater than this indicates our survival estimates would change even less after correction. Since the probability of a tag being operational at the downstream detection arrays for our survival studies were very close to one (Table 2), thus we did not adjust our survival estimates.

Table 2. Estimated probabilities (mean, SE in parentheses) a radio-tag was operational at Bonneville Dam and the downstream detection arrays for yearling Chinook salmon, steelhead trout, and subyearling Chinook salmon, during 2004.

| Release Site | Detection Array Locations |  |
| :---: | :---: | :---: |
|  | Bonneville Dam | Survival Gates |
|  | Yearling Chinook sa |  |
| The Dalles Dam | 0.9996 (3.634×10 ${ }^{-6}$ ) | $0.9992\left(7.036 \times 10^{-5}\right)$ |
| Bonneville Dam | NA | $0.9999\left(3.474 \times 10^{-6}\right)$ |
|  | Steelhead tro |  |
| The Dalles Dam | $0.9996\left(2.152 \times 10^{-5}\right)$ | 0.9993 (3.795 ${ }^{10} 0^{-5}$ ) |
| Bonneville Dam | NA | $0.9998\left(1.122 \times 10^{-4}\right)$ |
|  | Subyearling Chinoo |  |
| The Dalles Dam | $0.9999\left(1.604 \times 10^{-5}\right)$ | $0.9999\left(2.030 \times 10^{-5}\right)$ |
| Bonneville Dam | NA | $1.0000\left(3.285 \times 10^{-7}\right)$ |

## River Discharge and Project Operations

In July of 2004, the U.S. Army Corps of Engineers (ACOE) discovered that the amount of water reported to be spilled at Bonneville Dam was incorrect. An error in the calibration of spill gate openings installed in the early 1970's resulted in up to $30 \%$ less water discharged through the spillway than was reported to regional fish and water management officials. In this report we have included the corrected spill discharges and summary discharge statistics as per the revised data set by the Army Corps of Engineers (see: Bonneville Lock and Dam, Revision to Fish Passage Plan Spill Patterns FEB2005, Memorandum; Appendix 6).

During spring 2004 (29 April to 7 June), mean river discharge at Bonneville Dam was 218.4 kcfs , and ranged from 147.6 to 302.2 kcfs (Table 3). Allocation of mean river discharge among dam areas during the spring was $14.3 \%$ through powerhouse $1,35.7 \%$ through the spillway and $46.7 \%$ through powerhouse 2. Spillway operations evaluated during the spring were supposed to be the NMFS Biological Opinion spill of 75 kcfs during the day with spill up to the total dissolved gas cap at night. However, due to the miscalibration errors, the spill operations were approximately 56 kcfs during the day ( 0500 to 1959 hrs ) and spill up to $125 \%$ of the total dissolved gas cap at night ( 2000 to 0459 hours). During spring day operations the majority of the river discharge passed through powerhouse 2 ( $51 \%$ ) with only $27 \%$ passing via the spillway, while at night $54 \%$ of the mean river discharge passed through the spillway with $37 \%$ passing via powerhouse 2 (Table 4).

During summer ( 21 June to 22 July), mean river discharge was 155.9 kcfs , with a range from 88.4 to 249.4 kcfs (Table 3). The majority of the mean river discharge passed through the powerhouse $2(56 \%)$, with $34 \%$ passing via the spillway and $6 \%$ through powerhouse 1 . Two different spill operations were evaluated during the summer. The NMFS Biological Opinion spill, which was actually 55.8 kcfs day ( 0400 to 2159 hours), with spill up to the total dissolved gas cap at night and 32 kcfs spill for 24 hours. During summer day operations $61 \%$ of the mean river discharge passed via the powerhouse 2 , with $29 \%$ passing via the spillway, while at night $43 \%$ of the river discharge was passed via powerhouse 2 and $49 \%$ via the spillway (Table 5).

Table 3. Descriptive statistics for discharge (kcfs) at Bonneville Dam during spring (29 April to 7 June) 2004. Values have been rounded to the nearest tenth and are based on hourly averages.

| Spring |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | :---: |
| Dam area | Percentage | Mean | Median | Minimum | Maximum |  |
| Powerhouse 1 | 14.3 | 33.3 | 27.75 | 0.0 | 100.9 |  |
| Sluiceway | 0.6 | 1.3 | 1.2 | 1.1 | 1.4 |  |
| Powerhouse 2 | 46.7 | 101.5 | 103.8 | 20.3 | 139.3 |  |
| Corner collector | 2.6 | 5.6 | 5.6 | 5.1 | 5.9 |  |
| Spillway | 35.7 | 76.6 | 56.8 | 35.3 | 145.8 |  |
| Total Discharge | 100.0 | 218.4 | 219.4 | 147.6 | 302.2 |  |
|  |  |  |  |  |  |  |
|  | Summer |  |  |  |  |  |
| Powerhouse 1 | 5.7 | 10.5 | 0.0 | 0.0 | 64.9 |  |
| Sluiceway | 0.7 | 1.1 | 1.2 | 0.2 | 1.4 |  |
| Powerhouse 2 | 56.1 | 86.7 | 94.2 | 23.3 | 127.7 |  |
| Corner collector | 3.6 | 5.3 | 5.4 | 4.1 | 5.9 |  |
| Spillway | 33.9 | 52.4 | 33.1 | 13.5 | 161.4 |  |
| Total Discharge | 100.0 | 155.9 | 156.0 | 88.4 | 249.4 |  |

Table 4. Descriptive statistics for discharge (kcfs) during day (0500 to 1959) and night (2000 to 0459 hours) by dam area at Bonneville Dam during spring (29 April to 7 June) 2004. Dam operations were 56 kcfs spill during the day and night spill up to $125 \%$ of the total dissolved gas cap (TDG) at night

| Period and dam area | Percent <br> (of period) | Mean | Median | Minimum | Maximum |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Day, 56 kcfs |  |  |  |  |  |
| Powerhouse 1 | 18.8 | 43.9 | 43.9 | 0.0 | 100.9 |
| Sluiceway | 0.6 | 1.3 | 1.3 | 1.1 | 1.4 |
| Powerhouse 2 | 51.4 | 110.7 | 114.0 | 77.9 | 139.3 |
| Corner collector | 2.6 | 5.6 | 5.6 | 5.1 | 5.9 |
| Spillway | 26.6 | 56.4 | 56.5 | 35.3 | 69.5 |
| Total Discharge | 100.0 | 217.9 | 219.1 | 147.6 | 302.2 |
| Night, TDG |  |  |  |  |  |
| Powerhouse 1 | 5.5 | 12.4 | 8.4 | 0.0 | 62.5 |
| Sluiceway | 0.6 | 1.3 | 1.2 | 1.1 | 1.4 |
| Powerhouse 2 | 37.4 | 83.3 | 85.8 | 20.3 | 133.1 |
| Corner collector | 2.6 | 5.6 | 5.6 | 5.2 | 5.9 |
| Spillway | 53.9 | 116.8 | 114.6 | 57.3 | 145.8 |
| Total Discharge | 100.0 | 219.3 | 221.2 | 157.0 | 295.9 |
|  |  |  |  |  |  |

Table 5. Descriptive statistics for discharge (kcfs) during day (0400 to 2159) and night (2200 to 0359 hours) by dam area at Bonneville Dam during summer (21 June to 22 July) 2004 during. Dam operations were either 56 kcfs day spill with night spill up to $125 \%$ of the total dissolved gas cap (TDG) at night or 32 kcfs spill (day and night).

| Percent | Mean | Median | Minimum | Maximum |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Period and dam area | Pof period) |  |  |  |  |
| Day, 32 kcfs | 9.2 | 16.6 | 15.6 | 0.0 | 62.1 |
| Powerhouse 1 | 0.8 | 1.1 | 1.2 | 0.4 | 1.4 |
| Sluiceway | 64.6 | 97.9 | 104.4 | 55.5 | 127.7 |
| Powerhouse 2 | 3.7 | 5.4 | 5.4 | 4.4 | 5.9 |
| Corner collector | 21.7 | 31.8 | 31.8 | 13.5 | 56.6 |
| Spillway | 100.0 | 152.8 | 157.5 | 93.0 | 224.5 |
| Total Discharge |  |  |  |  |  |
| Night, 32 kcfs | 8.3 | 14.3 | 8.0 | 0.0 | 56.8 |
| Powerhouse 1 | 0.8 | 1.2 | 1.2 | 0.4 | 1.4 |
| Sluiceway | 66.5 | 99.0 | 101.4 | 69.9 | 114.9 |
| Powerhouse 2 | 3.7 | 5.5 | 5.6 | 4.4 | 5.9 |
| Corner collector | 20.7 | 30.7 | 31.8 | 16.1 | 32.6 |
| Spillway | 100.0 | 150.7 | 148.0 | 106.3 | 196.6 |
| Total Discharge |  |  |  |  |  |
| Day, 56 kcfs | 2.9 | 6.2 | 0.0 | 0.0 | 64.9 |
| Powerhouse 1 | 0.7 | 1.1 | 1.2 | 0.2 | 1.4 |
| Sluiceway | 56.3 | 89.4 | 91.5 | 25.4 | 125.4 |
| Powerhouse 2 | 3.4 | 5.3 | 5.4 | 4.1 | 5.9 |
| Corner collector | 36.7 | 55.8 | 56.4 | 33.5 | 68.3 |
| Spillway |  |  |  |  | 88.4 |
| Total Discharge | 100.0 | 157.7 | 153.6 | 88.4 | 249.4 |
| Night, TDG | 0.0 | 6.3 | 0.0 |  | 0.0 |
| Powerhouse 1 | 1.2 | 1.2 | 0.5 | 5.9 |  |
| Sluiceway | 0.7 | 3.2 | 26.8 | 23.3 | 1.4 |
| Powerhouse 2 | 19.7 | 3.2 | 59.4 |  |  |
| Corner collector | 3.4 | 5.4 | 5.4 | 4.5 | 5.9 |
| Spillway | 76.1 | 126.9 | 136.0 | 56.4 | 161.4 |
| Total Discharge | 100.0 | 165.7 | 174.8 | 88.4 | 224.8 |

## Yearling Chinook salmon

## Paired Release-recapture Model

Throughout the migration season the MGR and the ice and trash sluiceway, at powerhouse 1 were not operating continuously due to a policy that prioritized the passage of water through powerhouse 2 . Because of this policy, powerhouse 1 was only operated sporadically for short time intervals before and after the releases of radio tagged fish associated with this study.

## Assumption tests for MGR turbine unit and ice and trash sluiceway

## Burnham Tests

The results of the Burnham Tests 2 and 3 testing assumptions A5 and A6 for the following yearling Chinook salmon paired releases: the MGR turbine unit and the control at the MGR front roll, the MGR turbine unit and the control downstream of the outfall of the juvenile bypass at powerhouse 2, and the ice and trash sluiceway with the control downstream of the outfall of the juvenile bypass at powerhouse 2 were mostly incalculable because of the presence of all zeroes in either rows or columns of the contingency table. The results of these tests can be found in Appendix 4 (Tables A4.1, A4.2, and A4.3).

## Tests of the assumption of mixing of the treatment and control groups

The chi-square tests of homogeneity testing for the similarity in arrival times of the following paired releases of yearling Chinook salmon: the MGR turbine unit and the control at the MGR front roll, the MGR turbine unit and the control downstream of the outfall of the juvenile bypass at powerhouse 2 , and the ice and trash sluiceway with the control downstream of the outfall of the juvenile bypass at powerhouse 2 indicated that there were no significant differences in arrival times between the two release groups at the downstream radio telemetry arrays (Appendix 5 Tables A5.1, A5.2, A5.3).

## Minimum Gap Runner Turbine Unit Survival Estimation

## Control group released directly downstream of front roll turbine unit

We estimated that the survival of yearling Chinook salmon released into the MGR turbine unit at Bonneville Dam's powerhouse 1 with the control group released in the tailrace below the MGR unit front roll ranged from 0.830 to 1.042 during 2004 (Table 6). The average survival was estimated to be $0.956(\mathrm{SE}=0.016,95 \%$ confidence interval [0.924, 0.988]).

Control group released below the outfall of the powerhouse 2 juvenile bypass
We estimated that the survival of yearling Chinook salmon released into the MGR turbine unit at Bonneville Dam's powerhouse 1 with the control group released in the tailrace below the powerhouse 2 juvenile bypass outfall ranged from 0.83 to 1.02 during 2004 (Table 6). The average survival was estimated to be 0.944 ( $\mathrm{SE}=0.015,95 \%$ confidence interval [0.913, 0.976]).

Table 6. The estimated survival and standard error (SE) of yearling Chinook salmon released into the minimum gap runner (MGR) turbine unit at Bonneville Dam's powerhouse 1 during 2004. Treatment releases were made directly into the MGR turbine unit and two control releases were evaluated; one directly below the front roll of the turbine unit and one below the powerhouse 2 juvenile bypass system (JBS) outfall at Bonneville Dam. Survival estimates are presented for both paired-release groupings (e.g., the MGR and the control group directly below the turbine unit and MGR and the control group below the powerhouse 2 juvenile bypass outfall). The survival estimates presented are the estimated survival of the release group into the MGR to the release location of the tailrace release group. The specific dates and times of the releases can be referenced in Appendix 1.

|  | Treatment $=$ MGR, Control $=$ directly <br> below the front roll of the turbine unit | Treatment $=$ MGR, Control = below <br> the powerhouse 2 JBS outfall |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Release | S | SE | S | SE |
| 1 | 0.97 | 0.08 | 1.02 | 0.08 |
| 2 | a | a | 0.88 | 0.10 |
| 3 | 1.00 | 0.06 | 0.96 | 0.04 |
| 4 | 0.88 | 0.08 | 0.94 | 0.11 |
| $5^{\text {c }}$ | 1.01 | 0.06 | 1.01 | 0.07 |
| $6^{\mathrm{c}}$ | 0.99 | 0.09 | 0.96 | 0.08 |
| 7 | 1.00 | 0.10 | 1.00 | 0.10 |
| $8^{\mathrm{c}}$ | 0.91 | 0.08 | 0.88 | 0.07 |
| 9 | 1.04 | 0.09 | 1.00 | 0.12 |
| 10 | b | b | b | b |
| 11 | 0.97 | 0.07 | 0.95 | 0.07 |
| 12 | 0.97 | 0.08 | 0.93 | 0.07 |
| 13 | 0.83 | 0.09 | 0.83 | 0.08 |
| 14 | 0.94 | 0.07 | 0.92 | 0.09 |
| 15 | 0.93 | 0.09 | 0.93 | 0.06 |
| $16^{\text {c }}$ | b | b | b | b |

[^1]
## Powerhouse 1 Ice and Trash Sluiceway Survival Estimation

We estimated that the survival of yearling Chinook salmon released into the ice and trash sluiceway at Bonneville Dam's powerhouse 1 (control release in the tailrace below the outfall of powerhouse 2 juvenile bypass outfall) during 2004 ranged from 0.84 to 1.15 (Table 7). The average survival was estimated to be 0.947 ( $\mathrm{SE}=0.018,95 \%$ confidence interval [0.908, 0.986]).

Table 7. The estimated survival and standard error (SE) of yearling Chinook salmon released into the ice and trash sluiceway at Bonneville Dam's powerhouse 1 during 2004. Treatment releases were made into the ice and trash sluiceway and the control release was released below the powerhouse 2 , juvenile bypass (JBS) outfall at Bonneville Dam. The survival estimates presented are the estimated survival of the release group into the ice and trash sluiceway to the release location of the tailrace release group. The specific dates and times of the releases can be referenced in Appendix 1.

|  | Powerhouse 1 Ice and trash sluiceway and downstream tailrace control |  |
| :---: | :---: | :---: |
| Release | S | SE |
| 1 | 0.97 | 0.08 |
| 2 | 0.93 | 0.09 |
| 3 | 0.89 | 0.07 |
| 4 | 0.94 | 0.11 |
| $5^{\text {a }}$ | 1.05 | 0.06 |
| $6^{\text {a }}$ | 1.01 | 0.07 |
| 7 | 0.93 | 0.06 |
| $8^{\text {a }}$ | 0.90 | 0.06 |
| 9 | 0.92 | 0.05 |
| 10 | 1.15 | 0.10 |
| 11 | 0.91 | 0.08 |
| 12 | 0.97 | 0.08 |
| 13 | 0.93 | 0.05 |
| 14 | 0.92 | 0.08 |
| 15 | 0.89 | 0.07 |
| $16^{\text {a }}$ | 0.84 | 0.09 |

[^2]
## Spillway Flow Deflectors

## 7-ft Spillbay Flow Deflectors

## Assumption Tests

## Burnham Tests

The results of the Burnham Tests 2 and 3 testing assumptions A5 and A6 for the yearling Chinook passing via spillbays with 7 -ft deflectors at Bonneville Dam's spillway and the corresponding tailrace releases below the powerhouse 2 JBS outfall were mostly incalculable because of the presence of all zeroes in either rows or columns of the contingency table. The results of these tests can be found in Appendix 4, Table A4.8.

## Tests of the assumption of mixing of the treatment and control groups

The chi-square tests of homogeneity testing for the similarity in arrival times of post-hoc paired releases of yearling Chinook salmon passing via 7 -ft deflectors at Bonneville Dam's spillway and the corresponding tailrace releases below the powerhouse 2 JBS outfall indicated that there were few significant differences ( $P<0.0009$, Dunn-Sidak experimentwise error rate) in arrival times between the two release groups (Appendix 5, Table A5.8.).

## Survival estimation

We estimated the survival of yearling Chinook salmon passing via spillbays with 7-ft deflectors at Bonneville Dam spillway during daytime 56 kcfs spill operations ranged from 0.780 to 1.010 ( $\bar{X}=0.937 ; S E=0.018 ; 95 \% C I=[0.898,0.976]$; Table 8 ). Survival estimates for yearling Chinook salmon passing via spillbays with 7 -ft deflectors during total dissolved gas cap spill operations ranged from 0.780 to $1.077(\bar{X}=0.943 ; S E=0.026 ; 95 \% C I=[0.886,1.000]$; Table 8).

Table 8. The estimated survival probabilities (S), standard errors (SE), $95 \%$ confidence intervals (CI), number released ( N ; number of fish), and dates associated with paired releases of radiotagged yearling Chinook salmon passing via spillbays with deflectors located 7 feet above mean sea level during 56 kcfs day (05:00-19:59 hrs) spill operations and spill at night (20:00-04:59 hrs) up to the total dissolved gas cap of $125 \%$ in the tailrace at the Bonneville Dam spillway. The paired release groups were formed post-hoc from fish released into The Dalles Dam tailrace (treatment) and in the Bonneville Dam tailrace (control) during 2004.

| Release | 56 kcfs day spill operations |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | S | SE |  | CI | Treatment N | Control N | Start <br> Date | End date |
| 1 | 0.940 | 0.072 | 0.798, | 1.082 | 14 | 75 | 29-Apr | 01-May |
| 2 | 0.959 | 0.055 | 0.852, | 1.066 | 28 | 78 | 01-May | 03-May |
| 3 | 0.990 | 0.061 | 0.871, | 1.108 | 32 | 77 | 03-May | 05-May |
| 4 | 0.938 | 0.046 | 0.847, | 1.028 | 66 | 125 | 05-May | 08-May |
| 5 | 1.010 | 0.038 | 0.937, | 1.084 | 54 | 98 | 08-May | 10-May |
| 6 | 0.938 | 0.046 | 0.847, | 1.028 | 48 | 95 | 10-May | 13-May |
| 7 | 0.908 | 0.032 | 0.845, | 0.971 | 85 | 155 | 13-May | 17-May |
| 8 | 0.837 | 0.112 | 0.617, | 1.057 | 17 | 82 | 17-May | 19-May |
| 9 | 0.956 | 0.063 | 0.832, | 1.080 | 45 | 117 | 19-May | 22-May |
| 10 | 0.780 | 0.091 | 0.601, | 0.959 | 22 | 82 | 22-May | 24-May |
| 11 | 0.990 | 0.037 | 0.916, | 1.063 | 19 | 81 | 24-May | 26-May |
| 12 | 0.980 | 0.101 | 0.783, | 1.177 | 12 | 39 | 26-May | 27-May |
| 13 | 0.959 | 0.045 | 0.870, | 1.048 | 48 | 172 | 27-May | 31-May |
| Total dissolved gas cap night operations |  |  |  |  |  |  |  |  |
| 1 | 0.920 | 0.072 | 0.778, | 1.062 | 21 | 75 | 29-Apr | 01-May |
| 2 | 0.990 | 0.045 | 0.902, | 1.078 | 28 | 78 | 01-May | 03-May |
| 3 | 0.875 | 0.088 | 0.703, | 1.047 | 23 | 77 | 03-May | 05-May |
| 4 | 1.042 | 0.105 | 0.836, | 1.248 | 16 | 125 | 05-May | 08-May |
| 5 | 0.918 | 0.085 | 0.752, | 1.083 | 16 | 98 | 08-May | 10-May |
| 6 | 0.896 | 0.046 | 0.806, | 0.985 | 23 | 95 | 10-May | 13-May |
| 7 | 0.827 | 0.092 | 0.646, | 1.007 | 21 | 155 | 13-May | 17-May |
| 8 | 1.077 | 0.075 | 0.930, | 1.224 | 22 | 82 | 17-May | 19-May |
| 9 | 1.055 | 0.065 | 0.927, | 1.182 | 21 | 117 | 19-May | 22-May |
| 10 | 0.890 | 0.072 | 0.748, | 1.032 | 18 | 82 | 22-May | 24-May |
| 11 | 1.052 | 0.076 | 0.903, | 1.201 | 24 | 81 | 24-May | 26-May |
| 12 | 0.780 | 0.112 | 0.560 , | 1.000 | 17 | 39 | 26-May | 27-May |
| 13 | 0.939 | 0.074 | 0.794 , | 1.084 | 19 | 172 | 27-May | 31-May |

## 14-ft Spillbay Flow Deflectors

## Assumption Tests

## Burnham Tests

The results of the Burnham Tests 2 and 3 testing assumptions A5 and A6 for the yearling Chinook salmon passing via spillbays with 14 -ft deflectors at Bonneville Dam's spillway and the corresponding tailrace releases below the powerhouse 2 JBS outfall were mostly incalculable because of the presence of all zeroes in either rows or columns of the contingency table. The results of these tests can be found in Appendix 4, Table A4.8.

## Tests of the assumption of mixing of the treatment and control groups

The chi-square tests of homogeneity testing for the similarity in arrival times of post-hoc paired releases of yearling Chinook salmon passing via $14-\mathrm{ft}$ deflectors at Bonneville Dam's spillway and the corresponding tailrace releases below the powerhouse 2 JBS outfall indicated that there were some significant differences ( $P<0.0009$, Dunn-Sidak experimentwise error rate) in arrival times between the two release groups (Appendix 5, Table A5.8).

## Survival estimation

We estimated the survival of yearling Chinook salmon passing via spillbays with $14-\mathrm{ft}$ deflectors at Bonneville Dam spillway during 56 kcfs spill operations ranged from 0.643 to 1.020 ( $\bar{X}=0.773 ; S E=0.045 ; 95 \% C I=[0.667,0.879]$; Table 9). Survival estimates for yearling Chinook salmon passing via spillbays with 14-ft deflector during total dissolved gas cap spill operations ranged from 0.845 to 1.087 ( $\bar{X}=0.946 ; S E=0.018 ; 95 \% C I=[0.907,0.985]$; Table 9).

## Comparison of yearling Chinook salmon survival estimates for $\mathbf{7 - f t}$ and $\mathbf{1 4 - f t}$ Spillbay deflector during two spill operations

Survival estimates for yearling Chinook salmon passing through spillbays with 7-ft and $14-\mathrm{ft}$ deflectors were higher during the total dissolved gas cap night spill operations where flows are typically higher and more fish tend to pass than during the 56 kcfs day spill operations (Figure 12). At the lower flow spill operation of 56 kcfs day survival estimates of yearling Chinook salmon were much higher for fish passing through spillbays with the 7 - ft deflectors than through spillbays with the 14 -ft deflectors.

Table 9. The estimated survival probabilities (S), standard errors (SE), 95\% confidence intervals (CI), number released ( N ; number of fish), and dates associated with paired releases of radiotagged yearling Chinook salmon passing via spillbays with deflectors located 14 feet above mean sea level during 56 kcfs day (05:00-19:59 hrs) spill operations and spill at night (20:00-04:59 hrs ) up to the total dissolved gas cap of $125 \%$ in the tailrace at Bonneville Dam spillway. The paired release groups were formed post-hoc from fish released into The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

| Release | 56 kcfs day spill operations |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | S | SE | 95\% |  | Treatment N | Control N | Start date | End <br> date |
| 2 | 1.020 | 0.121 | 0.784, | 1.257 | 13 | 78 | 01-May | 03-May |
| 4 | 0.688 | 0.095 | 0.502 , | 0.873 | 26 | 125 | 05-May | 08-May |
| 5 | 0.784 | 0.135 | 0.519 , | 1.048 | 12 | 98 | 08-May | 10-May |
| 7 | 0.816 | 0.092 | 0.636, | 0.997 | 20 | 155 | 13-May | 17-May |
| 8 | 0.728 | 0.154 | 0.426, | 1.030 | 12 | 82 | 17-May | 19-May |
| 9 | 0.648 | 0.112 | 0.429 , | 0.868 | 24 | 117 | 19-May | 22-May |
| 11 | 0.854 | 0.095 | 0.667, | 1.041 | 20 | 81 | 24-May | 26-May |
| 13 | 0.643 | 0.093 | 0.461, | 0.825 | 32 | 172 | 27-May | 31-May |
| Total dissolved gas cap night spill operations |  |  |  |  |  |  |  |  |
| 1 | 0.910 | 0.072 | 0.768, | 1.052 | 21 | 75 | 29-Apr | 01-May |
| 2 | 0.950 | 0.058 | 0.837, | 1.063 | 29 | 78 | 01-May | 03-May |
| 3 | 0.917 | 0.078 | 0.763, | 1.070 | 23 | 77 | 03-May | 05-May |
| 4 | 0.958 | 0.066 | 0.830, | 1.087 | 30 | 125 | 05-May | 08-May |
| 5 | 1.042 | 0.068 | 0.909, | 1.174 | 28 | 98 | 08-May | 10-May |
| 6 | 0.865 | 0.065 | 0.737, | 0.992 | 46 | 95 | 10-May | 13-May |
| 7 | 0.845 | 0.093 | 0.663, | 1.028 | 28 | 155 | 13-May | 17-May |
| 8 | 0.924 | 0.082 | 0.764, | 1.084 | 26 | 82 | 17-May | 19-May |
| 9 | 1.087 | 0.078 | 0.935, | 1.239 | 27 | 117 | 19-May | 22-May |
| 10 | 0.960 | 0.041 | 0.879, | 1.041 | 24 | 82 | 22-May | 24-May |
| 11 | 0.917 | 0.065 | 0.789, | 1.045 | 34 | 81 | 24-May | 26-May |
| 12 | 0.970 | 0.083 | 0.808, | 1.132 | 15 | 39 | 26-May | 27-May |
| 13 | 0.949 | 0.055 | 0.842, | 1.056 | 37 | 172 | 27-May | 31-May |

Note- releases were not included where too few fish passed (i.e. $<10$ )


Figure 12. The estimated survival probabilities ( $95 \%$ confidence intervals error bars) from paired releases of radio-tagged yearling Chinook salmon passing through spillbays with deflectors located at either 7 - ft or $14-\mathrm{ft}$ above mean sea level at Bonneville Dam under two spill operations ( 56 kcfs day or total dissolved gas cap(TDG) night) by release. The paired release groupings were formed post-hoc from fish released into The Dalles Dam tailrace and in the Bonneville Dam tailrace during 2004.

## Route-specific Survival Model

## Survival estimation

Capture histories were generated for each passage scenario, indicating detection at the release location, detection at the dam, and detection downstream of the dam by assigning a 1 for detection and a 0 for not detected at antenna arrays. Using capture histories from the detections of radio-tagged yearling Chinook salmon released at The Dalles Dam and below the Bonneville Dam powerhouse 2 JBS outfall (Table 10), we generated maximum likelihood estimates of the route-specific passage and survival probabilities through Bonneville Dam (Figure 13). The survival of yearling Chinook salmon through Bonneville Dam spillway was estimated to be 0.910 ( $\mathrm{SE}=0.011$, profile likelihood $95 \%$ confidence interval [ $0.888,0.931]$ ). For yearling Chinook salmon passing via powerhouse 1 , the estimated survival was 0.913 ( $\mathrm{SE}=0.019$, profile likelihood $95 \%$ confidence interval [0.872, 0.949]) and for yearling Chinook salmon passing via powerhouse 2 turbines the estimated survival was 0.951 ( $\mathrm{SE}=0.011$, profile likelihood $95 \%$ confidence interval $[0.929,0.972])$. For yearling Chinook salmon passing via the JBS the estimated survival was $0.970(\mathrm{SE}=0.013$, profile likelihood $95 \%$ confidence interval [0.943, $0.994]$ ) and passing via the corner collector at powerhouse 2 the estimated survival was 1.016 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval [0.999, 1.032]). Yearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.951 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval [0.937, 0.966]) and project survival was estimated to be 0.883 ( $\mathrm{SE}=0.008$, profile likelihood confidence interval [0.868, 0.898]).

Table 10. Counts of radio-tagged yearling Chinook salmon for the releases from The Dalles Dam $\left(R_{l}\right)$ and in the tailrace of Bonneville Dam $\left(R_{2}\right)$ used in the route-specific survival model during 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $R_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only $(01)$ within the passage route. B 1 is powerhouse 1 and B 2 is powerhouse 2 at Bonneville Dam.

| Release | Detection History | Route | Counts | Within-route histories Bonneville Dam |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 11 | 01 | 10 |
| $\mathrm{R}_{1}=4486$ | 100 |  | 326 |  |  |  |
|  | 101 |  | 70 |  |  |  |
|  | 110 | Spillway | 172 | 1126 | 9 | 203 |
|  | 111 |  | 1166 |  |  |  |
|  | 110 | B1 | 43 | 216 | 121 | 5 |
|  | 111 |  | 299 |  |  |  |
|  | 110 | B2 Turbines | 91 | 511 | 50 | 448 |
|  | 111 |  | 918 |  |  |  |
|  | 110 | B2 Juvenile bypass | 37 | 518 | 2 | 2 |
|  | 111 |  | 485 |  |  |  |
|  | 110 | B2 Corner collector | 24 | 878 | 0 | 1 |
|  | 111 |  | 855 |  |  |  |
| $\mathrm{R}_{2}=1276$ | 010 |  | 54 |  |  |  |
|  | 011 |  | 1222 |  |  |  |



Figure 13. Schematic of estimated route-specific passage and survival parameters for yearling Chinook salmon through Bonneville Dam during 56 kcfs day/total dissolved gas cap night spill operations. Estimated standard errors are in parentheses.

## 56 kcfs day spill operations

Using capture histories generated from the detections of radio-tagged yearling Chinook salmon released at The Dalles Dam and passing Bonneville Dam between 0500 and 2000 hrs during 56 kcfs spill operations (Table 11), we generated maximum likelihood estimates of the route-specific passage and survival probabilities (Figure 14). The survival of yearling Chinook salmon through Bonneville Dam spillway during 56 kcfs day spill operations was estimated to be 0.861 ( $\mathrm{SE}=0.015$, profile likelihood $95 \%$ confidence interval [ $0.831,0.889]$ ). For yearling Chinook salmon passing via powerhouse 1 , the estimated survival was 0.894 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval [ $0.853,0.930]$ ) and for yearling Chinook salmon passing via powerhouse 2 turbines, the estimated survival was 0.925 ( $\mathrm{SE}=0.012$, profile likelihood $95 \%$ confidence interval [0.902, 0.948]). For yearling Chinook salmon passing via the powerhouse 2 JBS the estimated survival was 0.938 ( $\mathrm{SE}=0.015$, profile likelihood $95 \%$ confidence interval [ $0.906,0.966]$ ). The survival of yearling Chinook salmon passing via the corner collector at powerhouse 2 was estimated to be 0.993 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval [0.977, 1.008]). Yearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.925 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval [ $0.912,0.941]$ ) and project survival was estimated to be 0.860 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval [0.844, 0.876]).

## Total dissolved gas cap night spill operations

Using capture histories generated from the detections of radio-tagged yearling Chinook salmon released at The Dalles Dam and passing Bonneville Dam between 2000 and 0500 hrs during total dissolved gas cap night spill operations (Table 11), we generated maximum likelihood estimates of the route-specific passage and survival probabilities (Figure 15). The survival of yearling Chinook salmon through Bonneville Dam spillway during total dissolved gas cap night spill operations was estimated to be 0.964 ( $\mathrm{SE}=0.016$, profile likelihood $95 \%$ confidence interval [0.932, 0.996]). For yearling Chinook salmon passing via powerhouse 1 , the estimated survival was 0.903 ( $\mathrm{SE}=0.060$, profile likelihood $95 \%$ confidence interval [0.768, $0.999]$ ) and for yearling Chinook salmon passing via powerhouse 2 turbines, the estimated survival was 0.981 ( $\mathrm{SE}=0.021$, profile likelihood $95 \%$ confidence interval [0.937, 1.020]). For yearling Chinook salmon passing via the powerhouse 2 JBS the estimated survival was 1.013 ( $\mathrm{SE}=0.022$, profile likelihood $95 \%$ confidence interval [ $0.964,1.053]$ ). The survival of yearling Chinook salmon passing via the corner collector at powerhouse 2 was estimated to be 1.028 (SE $=0.021$, profile likelihood $95 \%$ confidence interval [ $0.980,1.065]$ ). Yearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.979 ( $\mathrm{SE}=0.015$, profile likelihood $95 \%$ confidence interval $[0.953,1.007]$ ) and project survival was estimated to be 0.908 ( $\mathrm{SE}=$ 0.015 , profile likelihood $95 \%$ confidence interval [0.881, 0.937]).

Table 11. Counts of radio-tagged yearling Chinook salmon for releases from The Dalles Dam $\left(R_{1}\right)$ and in the tailrace of Bonneville Dam $\left(R_{2}\right)$ used in the route-specific survival model during two spill operations: 56 kcfs day ( $0500-2000$ ) and total dissolved gas cap (TDG) at night in 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $R_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B1 is powerhouse 1 and B 2 is powerhouse 2 at Bonneville Dam.

| 56 kcfs day spill operations |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Release | Detection History | Route | Counts | Within-route histories Bonneville Dam |  |  |
|  |  |  |  | 11 | 01 | 10 |
| $\mathrm{R}_{1}=3100$ | 100 |  | 222 |  |  |  |
|  | 101 |  | 52 |  |  |  |
|  | 110 | Spillway | 106 | 622 | 2 | 63 |
|  | 111 |  | 581 |  |  |  |
|  | 110 | B1 | 36 | 185 | 107 | 5 |
|  | 111 |  | 261 |  |  |  |
|  | 110 | B2 Turbines | 68 | 355 | 38 | 353 |
|  | 111 |  | 678 |  |  |  |
|  | 110 | B2 Juvenile bypass | 28 | 357 | 2 | 1 |
|  | 111 |  | 332 |  |  |  |
|  | 110 | B2 Corner collector | 18 | 735 | 0 | 1 |
|  | 111 |  | 718 |  |  |  |
| $\mathrm{R}_{2}=640$ | 010 |  | 11 |  |  |  |
|  | 011 |  | 629 |  |  |  |
| TDG night spill operations |  |  |  |  |  |  |
| Release |  |  |  | Within-route histories Bonneville Dam |  |  |
|  | Detection History | Route | Counts | 11 | 01 | 10 |
| $\mathrm{R}_{1}=1386$ | 100 |  | 101 |  |  |  |
|  | 101 |  | 22 |  |  |  |
|  | 110 | Spillway | 66 | 504 | 7 | 140 |
|  | 111 |  | 585 |  |  |  |
|  | 110 | B1 | 7 | 31 | 14 | 0 |
|  | 111 |  | 38 |  |  |  |
|  | 110 | B2 Turbines | 23 | 156 | 12 | 95 |
|  | 111 |  | 240 |  |  |  |
|  | 110 | B2 Juvenile bypass | $9$ | 161 | 0 | 1 |
|  | 111 |  | $153$ |  |  |  |
|  | 110 | B2 Corner collector | 6 | 143 | 0 | 0 |
|  | 111 |  | 137 |  |  |  |
| $\mathrm{R}_{2}=636$ | 010 |  | 43 |  |  |  |
|  | 011 |  | 593 |  |  |  |



Figure 14. Schematic of estimated route-specific passage and survival parameters for yearling Chinook salmon through Bonneville Dam during 56 kcfs day spill operations. Estimated standard errors are in parentheses.


Figure 15. Schematic of estimated route-specific passage and survival parameters for yearling Chinook salmon through Bonneville Dam during total dissolved gas cap night spill operations. Estimated standard errors are in parentheses.

## Comparison of survival during 56 kcfs day and TDG night spill operations

The estimated survival probabilities for yearling Chinook salmon passing via the powerhouse 2 turbines, the powerhouse 2 JBS, and the spillway were significantly different between the two spill operations (Table 12). The dam survival was also found to be significantly different between the 56 kcfs day and TDG night spill operations.

Table 12. Summary table of estimated route-specific survival probabilities (S) and their associated standard errors (SE) of yearling Chinook salmon survival through Bonneville Dam (Dam survival) generated from the route-specific survival model. The results of $Z$-tests (i.e., $Z$ statistic) structured to assess whether the estimated survival probabilities during the 56 kcfs day spill operations were different than the estimated survival probabilities during the total dissolved gas cap (TDG) night spill operations. Significant results are indicated where $\mathrm{Z} \geq 1.645$ given a two-tailed test and $\alpha=0.10$. The JBS refers to the juvenile bypass system at powerhouse 2 .

|  | 56 kcfs day |  | TDG night |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Passage route | S | SE | S | SE | $Z$ |
| Powerhouse 1 | 0.894 | 0.020 | 0.903 | 0.060 | 0.142 |
| Powerhouse 2 | 0.925 | 0.012 | 0.981 | 0.021 | 2.315 |
| Corner Collector | 0.993 | 0.008 | 1.028 | 0.021 | 1.557 |
| JBS | 0.938 | 0.015 | 1.013 | 0.022 | 2.817 |
| Spillway | 0.861 | 0.015 | 0.964 | 0.016 | 4.696 |
| Dam Survival | 0.925 | 0.008 | 0.979 | 0.015 | 3.177 |

## Steelhead trout

## Paired Release-recapture Model

## Assumption tests for the MGR turbine unit and the ice and trash sluiceway

## Burnham tests

The results of the Burnham Tests 2 and 3 testing assumptions A5 and A6 for the following steelhead trout paired releases: the MGR turbine unit and the control at the MGR front roll, the MGR turbine unit and the control downstream of the outfall of the juvenile bypass at powerhouse 2, and the ice and trash sluiceway with the control downstream of the outfall of the juvenile bypass at powerhouse 2 were mostly incalculable because of the presence of all zeroes in either rows or columns of the contingency table. The results of these tests can be found in Appendix 4 (Tables A4.4, A4.5, and A4.6).

## Tests of the assumption of mixing of the treatment and control groups

The chi-square tests of homogeneity testing for the similarity in arrival times of the following paired releases of steelhead trout: the MGR turbine unit and the control at the MGR front roll, the MGR turbine unit and the control downstream of the outfall of the juvenile bypass at powerhouse 2 , and the ice and trash sluiceway with the control downstream of the outfall of the juvenile bypass at powerhouse 2 indicated that there were no significant differences in arrival
times between the two release groups at the downstream radio telemetry arrays (Appendix 5 Tables A5.4, A5.5, A5.6).

## Minimum Gap Runner Turbine Unit Survival Estimation

## Control group released directly downstream of front roll below turbine unit

We estimated that the survival of steelhead trout released into the MGR turbine unit at Bonneville Dam's powerhouse 1 with the control group released in the tailrace below the MGR unit front roll ranged from 0.830 to 1.136 (Table 13) during 2004. The average survival was estimated to be 0.952 ( $\mathrm{SE}=0.024,95 \%$ confidence interval [0.900, 1.003]).

Control group released below the outfall of the powerhouse 2 juvenile bypass We estimated that the survival of steelhead trout released into the MGR turbine unit at Bonneville Dam's powerhouse 1 with the control group released in the tailrace below the outfall of the powerhouse 2 juvenile bypass outfall ranged from 0.74 to1.09 (Table 13) during 2004. The average survival was estimated to be 0.926 ( $\mathrm{SE}=0.030,95 \%$ confidence interval [0.861, 0.992]).

Table 13. The estimated survival (S) and standard error (SE) of steelhead trout released into the Minimum Gap Runner (MGR) Turbine Unit at Bonneville Dam's powerhouse 1 during 2004. Treatment releases were made directly into the MGR turbine unit and two control releases were evaluated; one directly below the front roll of the turbine unit and one below the powerhouse 2 juvenile bypass system (JBS) outfall at Bonneville Dam. Survival estimates are presented for both paired-release groupings (e.g., the MGR and the control group directly below the turbine unit and MGR and the control group below the powerhouse 2 juvenile bypass outfall). The survival estimates presented are the estimated survival of the release group into the MGR to the release location of the tailrace release group. The specific dates and times of the releases can be referenced in Appendix 1.

|  | Treatment $=$ MGR, Control $=$ directly <br> below the front roll of the turbine unit | Treatment $=$ MGR, Control = below <br> the powerhouse 2 JBS outfall |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Release | S | SE | S | SE |
| 1 | 1.14 | 0.15 | 1.00 | 0.14 |
| 2 | 0.83 | 0.12 | 0.78 | 0.10 |
| 3 | 0.87 | 0.10 | 0.82 | 0.08 |
| 4 | 0.92 | 0.08 | 0.89 | 0.06 |
| 5 | 0.91 | 0.06 | 0.91 | 0.08 |
| 6 | 0.88 | 0.13 | 1.06 | 0.21 |
| 7 | 0.90 | 0.08 | 0.91 | 0.09 |
| 8 | a | a | a | a |
| 9 | 0.93 | 0.08 | 0.88 | 0.07 |
| 10 | 0.93 | 0.16 | 0.83 | 0.07 |
| 11 | 1.14 | 0.12 | 1.09 | 0.07 |
| 12 | 0.99 | 0.18 | 0.74 | 0.10 |
| 13 | 0.96 | 0.10 | 1.06 | 0.09 |
| 14 | a | a | a | a |
| 15 | 0.95 | 0.09 | 0.96 | 0.09 |
| 16 | 1.00 | 0.09 | 1.06 | 0.13 |

[^3]
## Powerhouse 1 Ice and Trash Sluiceway Survival Estimation

We estimated that the survival of steelhead trout released into the ice and trash sluiceway at Bonneville Dam's powerhouse 1 (control release in the tailrace below the outfall of powerhouse 2 juvenile bypass outfall) during 2004 ranged from 0.76 to 1.07 (Table 14). The average survival was estimated to be 0.935 ( $\mathrm{SE}=0.024,95 \%$ confidence interval [0.884, 0.985]).

Table 14. The estimated survival and standard error (SE) of steelhead trout released into the ice and trash sluiceway at Bonneville Dam's powerhouse 1 during 2004. Treatment releases were made into the ice and trash sluiceway and the control release was released below the powerhouse 2, juvenile bypass (JBS) outfall at Bonneville Dam. The survival estimates presented are the estimated survival of the release group into the ice and trash sluiceway to the release location of the tailrace release group.

|  | Powerhouse 1 Ice and trash sluiceway and downstream tailrace control |  |
| :---: | :---: | :---: |
| Release | S | SE |
| 1 | 0.78 | 0.11 |
| 2 | 1.00 | 0.004 |
| 3 | 0.87 | 0.07 |
| 4 | 0.79 | 0.08 |
| 5 | 0.90 | 0.08 |
| 6 | 0.90 | 0.22 |
| 7 | 0.76 | 0.10 |
| 8 | 1.01 | 0.07 |
| 9 | 0.92 | 0.06 |
| 10 | 1.02 | 0.11 |
| 11 | 1.05 | 0.08 |
| 12 | 1.00 | 0.06 |
| 13 | 0.95 | 0.11 |
| 14 | 1.00 | 0.16 |
| 15 | 0.94 | 0.10 |
| 16 | 1.07 | 0.13 |

## Spillbay Flow Deflectors

## 7-ft Spillbay Flow Deflectors

## Assumption Tests

## Burnham tests

The results of the Burnham Tests 2 and 3 testing assumptions A5 and A6 for the steelhead trout passing via spillbays with 7-ft deflectors at Bonneville Dam's spillway and the corresponding tailrace releases below the powerhouse 2 JBS outfall were mostly incalculable because of the presence of all zeroes in either rows or columns of the contingency table. Of the tests that were calculated, only 2 of the 49 tests (Test 2 ) indicated lack of fit ( $P<0.0004$, DunnSidak experimentwise error rate). The results of these tests can be found in Appendix 4, Table A4.12.

Tests of the assumption of mixing of the treatment and control groups
The chi-square tests of homogeneity testing for the similarity in arrival times of post-hoc paired releases of steelhead trout indicated that there were some significant differences ( $P<$ 0.0009 , Dunn-Sidak experimentwise error rate) in arrival times between the two release groups (Appendix 5, Table A5.12.).

## Survival estimation

We estimated the survival of steelhead trout passing via spillbays with 7 -ft deflectors at Bonneville Dam spillway during 56 kcfs spill operations ranged from 0.667 to 1.064 ( $\bar{X}=0.927 ; S E=0.046 ; 95 \% C I=[0.818,1.036]$; Table 15). Survival estimates for steelhead trout passing via spillbays with 7 -ft deflectors during total dissolved gas cap spill operations ranged from 0.926 to 1.143 ( $\bar{X}=1.013 ; S E=0.016 ; 95 \% C I=[0.979,1.047]$; Table 15) .

Table 15. The estimated survival probabilities (S), standard errors (SE), 95\% confidence intervals, number released ( N ; number of fish), and dates associated with paired releases of radio-tagged steelhead trout passing via spillbays with deflectors located 7 feet above mean sea level during 56 kcfs day ( $05: 00-19: 59$ hours) spill operations and spill at night (20:00-04:59 hrs) up to the total dissolved gas cap of $125 \%$ in the tailrace at the Bonneville Dam spillway. The paired release groups were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

| 56 kcfs day spill operations |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Release | S | SE | 95\% CI | Treatment N | Control N | Start <br> date | End date |
| 5 | 1.042 | 0.088 | 0.869, 1.215 | 20 | 90 | 08-May | 12-May |
| 6 | 0.856 | 0.115 | 0.631, 1.080 | 17 | 81 | 12-May | 14-May |
| 7 | 1.064 | 0.134 | 0.802, 1.326 | 12 | 114 | 14-May | 17-May |
| 8 | 0.990 | 0.056 | 0.880, 1.099 | 21 | 136 | 17-May | 20-May |
| 9 | 0.667 | 0.157 | 0.359, 0.974 | 11 | 96 | 20-May | 22-May |
| 10 | 0.989 | 0.077 | 0.838, 1.141 | 14 | 132 | 22-May | 25-May |
| 13 | 0.948 | 0.075 | 0.802, 1.095 | 12 | 88 | 29-May | 31-May |
| 15 | 0.863 | 0.107 | 0.654, 1.073 | 16 | 119 | 01-Jun | 04-Jun |
| Total dissolved gas cap night spill operations |  |  |  |  |  |  |  |
| 1 | 1.075 | 0.116 | 0.848, 1.302 | 16 | 42 | 28-Apr | 02-May |
| 2 | 1.000 | 0.101 | 0.803, 1.197 | 20 | 30 | 02-May | 04-May |
| 3 | 1.020 | 0.094 | 0.836, 1.204 | 18 | 47 | 04-May | 06-May |
| 4 | 0.979 | 0.066 | 0.850, 1.108 | 16 | 78 | 06-May | 08-May |
| 5 | 1.042 | 0.105 | 0.836, 1.248 | 16 | 90 | 08-May | 12-May |
| 6 | 1.044 | 0.066 | 0.916, 1.173 | 18 | 81 | 12-May | 14-May |
| 7 | 0.968 | 0.067 | 0.837, 1.100 | 23 | 114 | 14-May | 17-May |
| 8 | 0.927 | 0.065 | 0.799, 1.055 | 26 | 136 | 17-May | 20-May |
| 9 | 0.938 | 0.075 | 0.790, 1.085 | 20 | 96 | 20-May | 22-May |
| 10 | 1.021 | 0.048 | 0.928, 1.115 | 26 | 132 | 22-May | 25-May |
| 11 | 0.926 | 0.080 | 0.769, 1.083 | 23 | 95 | 25-May | 27-May |
| 12 | 1.143 | 0.072 | 1.001, 1.285 | 25 | 90 | 27-May | 29-May |
| 13 | 1.031 | 0.078 | 0.878, 1.184 | 23 | 88 | 29-May | 31-May |
| 14 | 1.087 | 0.128 | 0.837, 1.337 | 15 | 36 | 31-May | 01-Jun |
| 15 | 1.000 | 0.077 | $0.850,1.150$ | 15 | 119 | 01-Jun | 04-Jun |

Note- releases were not included where too few fish passed (i.e. $<10$ )

## 14-ft Spillbay Flow Deflectors

## Assumption Tests

## Burnham tests

The results of the Burnham Tests 2 and 3 testing assumptions A5 and A6 for the steelhead trout passing via spillbays with 14-ft deflectors at Bonneville Dam's spillway and the corresponding tailrace releases below the powerhouse 2 JBS outfall were mostly incalculable
because of the presence of all zeroes in either rows or columns of the contingency table. Of the tests (Test 2) that were calculated, 2 of 47 tests indicated lack of fit ( $P<0.0004$, Dunn-Sidak experimentwise error rate). The results of these tests can be found in Appendix 4, Table A4.11.

Tests of the assumption of mixing of the treatment and control groups
The chi-square tests of homogeneity testing for the similarity in arrival times of post-hoc paired releases of steelhead trout indicated that there were some significant differences ( $P<$ 0.0009 , Dunn-Sidak experimentwise error rate) in arrival times between the two release groups (Appendix 5, Table A5.11).

## Survival estimation

We estimated the survival of steelhead trout passing via spillbays with $14-\mathrm{ft}$ deflectors at Bonneville Dam spillway during 56 kcfs spill operations ranged from 0.684 to 0.959 ( $\bar{X}=0.850 ; S E=0.063 ; 95 \% C I=[0.650,1.050]$; Table 16). Survival estimates for steelhead trout passing via spillbays with $14-\mathrm{ft}$ deflectors during total dissolved gas cap spill operations ranged from 0.878 to $1.143(\bar{X}=1.012 ; S E=0.015 ; 95 \% C I=[0.980,1.044]$; Table 16).

## Comparison of steelhead trout survival estimates for 7-ft and 14-ft Spillbay deflector during two spill operations

The point estimates of survival for steelhead trout passing through spillbays with 7 -ft and $14-\mathrm{ft}$ deflectors were higher during the total dissolved gas cap night spill operations when flows are typically higher and more fish tend to pass than during the 56 kcfs day spill operations (Figure 16). For the 56 kcfs day spill, the point estimates of survival for steelhead trout were higher for fish passing through spillbays with the 7 -ft deflectors than through spillbays with the $14-\mathrm{ft}$ deflectors.

Table 16. The estimated survival probabilities (S), standard errors (SE), $95 \%$ confidence intervals, number released (N; number of fish), and dates associated with paired releases of radio-tagged steelhead trout passing via spillbays with deflectors located 14 feet above mean sea level during 56 kcfs day ( $05: 00-19: 59 \mathrm{hrs}$ ) spill operations and spill at night (20:00-04:59 hrs) up to the total dissolved gas cap of $125 \%$ in the tailrace at the Bonneville Dam spillway. The paired release groups were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

| 56 kcfs day spill operations |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Release | S | SE | 95\% |  | Treatment N | Control <br> N | Start date | End date |
| 8 | 0.938 | 0.096 | 0.750, | 1.125 | 10 | 136 | 17-May | 20-May |
| 10 | 0.819 | 0.129 | 0.567, | 1.072 | 13 | 132 | 22-May | 25-May |
| 13 | 0.959 | 0.228 | 0.513, | 1.405 | 10 | 88 | 29-May | 31-May |
| 15 | 0.684 | 0.159 | 0.373, | 0.995 | 11 | 119 | 01-Jun | 04-Jun |
| Total dissolved gas cap night spill operations |  |  |  |  |  |  |  |  |
| 1 | 1.022 | 0.069 | 0.885, | 1.158 | 19 | 42 | 28-Apr | 02-May |
| 2 | 1.000 | 0.101 | 0.803, | 1.197 | 20 | 30 | 02-May | 04-May |
| 3 | 0.878 | 0.084 | 0.714, | 1.041 | 21 | 47 | 04-May | 06-May |
| 4 | 1.042 | 0.082 | 0.881, | 1.202 | 22 | 78 | 06-May | 08-May |
| 5 | 1.000 | 0.047 | 0.909, | 1.091 | 25 | 90 | 08-May | 12-May |
| 6 | 1.067 | 0.057 | 0.955 , | 1.178 | 24 | 81 | 12-May | 14-May |
| 7 | 1.021 | 0.048 | 0.928, | 1.115 | 28 | 114 | 14-May | 17-May |
| 8 | 0.979 | 0.023 | 0.934, | 1.025 | 32 | 136 | 17-May | 20-May |
| 9 | 0.958 | 0.056 | 0.849, | 1.068 | 26 | 96 | 20-May | 22-May |
| 10 | 0.989 | 0.057 | 0.877, | 1.101 | 28 | 132 | 22-May | 25-May |
| 11 | 1.011 | 0.053 | 0.906, | 1.115 | 35 | 95 | 25-May | 27-May |
| 12 | 1.143 | 0.072 | 1.001, | 1.285 | 27 | 90 | 27-May | 29-May |
| 13 | 1.031 | 0.059 | 0.916, | 1.146 | 33 | 88 | 29-May | 31-May |
| 14 | 1.043 | 0.071 | 0.903, | 1.184 | 25 | 36 | 31-May | 01-Jun |
| 15 | 0.989 | 0.047 | 0.897, | 1.082 | 32 | 119 | 01-Jun | 04-Jun |

Note- releases were not included where too few fish passed (i.e. $<10$ )


Figure 16. The estimated survival probabilities ( $95 \%$ confidence interval error bars) from paired releases of radio-tagged steelhead trout passing through spillbays with either deflectors located 7 feet or 14 feet above mean sea level at Bonneville Dam under two spill operations (TDG is total dissolved gas cap), by release. The paired release groupings were formed post-hoc from fish released at The Dalles Dam and in the Bonneville Dam tailrace during 2004.

## Route-specific Survival Model

## Survival estimation

Using capture histories generated from the detections of radio-tagged steelhead trout released at The Dalles Dam and in the tailrace of Bonneville Dam (Table 17), we generated maximum likelihood estimates of the route-specific passage and survival probabilities for steelhead trout through Bonneville Dam (Figure 17). The survival of steelhead trout through the Bonneville Dam spillway was estimated to be 0.979 ( $\mathrm{SE}=0.011$, profile likelihood $95 \%$ confidence interval $=[0.956,1.002])$. For steelhead trout passing via powerhouse 1 the estimated survival was 0.965 ( $\mathrm{SE}=0.019$, profile likelihood $95 \%$ confidence interval [0.926, $0.999]$ ) and for steelhead trout passing via powerhouse 2 turbines the estimated survival was 0.889 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval [ $0.848,0.927$ ]). For steelhead trout passing via the powerhouse 2 JBS the estimated survival was 0.951 ( $\mathrm{SE}=0.021$, profile likelihood $95 \%$ confidence interval [0.907, 0.989]). The survival of steelhead trout passing via the powerhouse 2 corner collector was estimated to be 1.030 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval $[1.014,1.047]$ ). Steelhead trout dam survival through Bonneville Dam was estimated to be 0.991 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval [0.975, 1.008]) and the project survival was estimated to be $0.897(\mathrm{SE}=0.009$, profile likelihood $95 \%$ confidence interval [0.881, 0.915]).

## 56 kefs day spill operations

Using capture histories generated from the detections of radio-tagged steelhead trout released at The Dalles Dam and passing Bonneville Dam between 0500 and 2000 hrs during 56 kcfs spill operations (Table 18), we generated maximum likelihood estimates of the routespecific passage and survival probabilities (Figure 18). The survival of steelhead trout through the Bonneville Dam spillway during 56 kcfs day spill operations was estimated to be 0.891 ( $\mathrm{SE}=$ 0.024 , profile likelihood $95 \%$ confidence interval [ $0.840,0.936]$ ). For steelhead trout passing via powerhouse 1, the estimated survival was 0.966 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval $[0.922,1.003])$ and for steelhead trout passing via powerhouse 2 turbines the estimated survival was 0.863 ( $\mathrm{SE}=0.028$, profile likelihood $95 \%$ confidence interval [0.804, 0.915]). For steelhead trout passing via the powerhouse 2 JBS , the estimated survival was 0.904 ( $\mathrm{SE}=0.031$, profile likelihood $95 \%$ confidence interval [0.837, 0.960]). For steelhead trout passing via the powerhouse 2 , corner collector, the estimated survival was 1.018 ( $\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval [0.9998, 1.039]). Steelhead trout dam survival through Bonneville Dam was estimated to be 0.980 ( $\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval [ 0.962 , $1.001]$ ) and project survival was estimated to be $0.888(\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval [0.868, 0.909].

## Total dissolved gas cap night spill operations

Using capture histories generated from the detections of radio-tagged steelhead trout released at The Dalles Dam and passed Bonneville Dam between 2000 and 0500 hrs during total dissolved gas cap spill operations (Table 18), we generated maximum likelihood estimates of the route-specific passage and survival probabilities (Figure 19). The survival of steelhead trout through Bonneville Dam spillway during total dissolved gas cap spill operations was estimated to be 1.020 ( $\mathrm{SE}=0.015$, profile likelihood $95 \%$ confidence interval [0.992, 1.050]). For steelhead trout passing via powerhouse 1, the estimated survival was $0.940(\mathrm{SE}=0.041$, profile
likelihood $95 \%$ confidence interval [0.850, 1.009]) and for steelhead trout passing via powerhouse 2 turbines the estimated survival was 0.917 ( $\mathrm{SE}=0.029$, profile likelihood $95 \%$ confidence interval $[0.857,0.970]$ ). For steelhead trout passing via the powerhouse 2 JBS, the estimated survival was $1.003(\mathrm{SE}=0.027$, profile likelihood $95 \%$ confidence interval [0.944, 1.050]). For steelhead trout passing via the powerhouse 2, corner collector the estimated survival was 1.028 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval [0.985, 1.066]). Steelhead trout dam survival through Bonneville Dam was estimated to be 0.998 ( $\mathrm{SE}=0.014$, $95 \%$ profile likelihood confidence interval [0.973, 1.027]) and project survival was estimated to be 0.904 ( $\mathrm{SE}=0.015$, profile likelihood confidence interval [0.876, 0.933]).

Table 17. Counts of radio-tagged steelhead trout for the releases from The Dalles Dam $\left(R_{l}\right)$ and in the tailrace of Bonneville $\operatorname{Dam}\left(R_{2}\right)$ used in the route-specific survival model during 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B1 is powerhouse 1 and B 2 is powerhouse 2 at Bonneville Dam.

|  |  |  | Within-route histories <br> Release |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Detection History | Route | Counts | 11 | 01 | 10 |
| $\mathrm{R}_{1}=4398$ | 100 |  | 420 |  |  |  |
|  | 101 | Spillway | 87 | 80 | 792 | 19 |

Table 18. Counts of radio-tagged steelhead trout for releases from The Dalles Dam $\left(R_{l}\right)$ and in the tailrace of Bonneville $\operatorname{Dam}\left(R_{2}\right)$ used in the route-specific survival model during two spill operations: 56 kcfs day (0500-1959) and total dissolved gas cap (TDG) at night in 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only ( 01 ) within the passage route. B1 is powerhouse 1 and B2 is powerhouse 2 at Bonneville Dam.

| 56 kcfs day spill operations |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Release | Detection History | Route | Counts | Within-route histories Bonneville Dam |  |  |
|  |  |  |  | 11 | 01 | 10 |
| $\mathrm{R}_{1}=2893$ | 100 |  | 276 |  |  |  |
|  | 101 |  | 31 |  |  |  |
|  | 110 | Spillway | 40 | 235 | 3 | 25 |
|  | 111 |  | 223 |  |  |  |
|  | 110 | B1 | 19 | 177 | 59 | 0 |
|  | 111 |  | 217 |  |  |  |
|  | 110 | B2 Turbines | 38 | 135 | 18 | 45 |
|  | 111 |  | 160 |  |  |  |
|  | 110 | B2 Juvenile bypass | 20 | 141 | 2 | 1 |
|  | 111 |  | 124 |  |  |  |
|  | 110 | B2 Corner collector | 54 | 1738 | 3 | 4 |
|  | 111 |  | 1691 |  |  |  |
| $\mathrm{R}_{2}=671$ | 010 |  | 32 |  |  |  |
|  | 011 |  | 639 |  |  |  |
| TDG night spill operations |  |  |  |  |  |  |
| Release | Detection History | Route | Counts | Within-route histories Bonneville Dam |  |  |
|  |  |  |  | 11 | 01 | 10 |
| $\mathrm{R}_{1}=1505$ | 100 |  | 144 |  |  |  |
|  | 101 |  | $16$ |  |  |  |
|  | 110 | Spillway | 40 | 557 | 16 | 160 |
|  | 111 |  | 693 |  |  |  |
|  | 110 | B1 | 11 | 37 | 48 | 0 |
|  | 111 |  | 74 |  |  |  |
|  | 110 | B2 Turbines | $31$ | 125 | 13 | 67 |
|  | 111 |  | 174 |  |  |  |
|  | 110 | B2 Juvenile bypass | 9 | 128 | 0 | 1 |
|  | 111 |  | 120 |  |  |  |
|  | 110 | B2 Corner collector | 9 | 192 | 1 | 0 |
|  | 111 |  | 184 |  |  |  |
| $\mathrm{R}_{2}=603$ | 010 |  | 44 |  |  |  |
|  | 011 |  | 559 |  |  |  |



Figure 17. Schematic of estimated route-specific passage and survival parameters for steelhead trout through Bonneville Dam during 2004. Estimated standard errors are in parentheses.


Figure 18. Schematic of estimated route-specific passage and survival parameters for steelhead trout passing during daytime 56 kcfs spill operations through Bonneville Dam 2004. Estimated standard errors are in parentheses.


Figure 19. Schematic of estimated route-specific passage and survival parameters for steelhead trout passing during total dissolved gas cap nighttime spill operations through Bonneville Dam 2004. Estimated standard errors are in parentheses.

## Comparison of survival during 56 kcfs day and TDG night spill operations

The estimated survival probabilities for steelhead trout passing via the powerhouse 2 JBS and the spillway were found to be significantly different between the two spill operations (Table 19). The dam survival for steelhead trout was not found to be significantly different between 56 kcfs day and TDG night spill operations.

Table 19. Summary table of estimated route-specific survival probabilities (S) and their associated standard errors (SE) of steelhead trout survival through Bonneville Dam (Dam survival) generated from the route-specific survival model. The results of $Z$-tests (i.e., $Z$ statistic) structured to assess whether the estimated survival probabilities during the 56 kcfs day spill operations were different than the estimated survival probabilities during the total dissolved gas cap (TDG) night spill operations. Significant results are indicated where $\mathrm{Z} \geq 1.645$ given a two-tailed test and $\alpha=0.10$. The JBS refers to the juvenile bypass system at powerhouse 2 .

|  | 56 kcfs day |  | TDG night |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Passage route | S | SE | S | SE | $Z$ |
| Powerhouse 1 | 0.966 | 0.020 | 0.940 | 0.041 | 0.570 |
| Powerhouse 2 | 0.863 | 0.028 | 0.917 | 0.029 | 1.340 |
| Corner Collector | 1.018 | 0.010 | 1.028 | 0.020 | 0.447 |
| JBS | 0.904 | 0.031 | 1.003 | 0.027 | 2.408 |
| Spillway | 0.891 | 0.024 | 1.020 | 0.015 | 4.558 |
| Dam Survival | 0.980 | 0.010 | 0.998 | 0.014 | 1.046 |

## Subyearling Chinook salmon

## Paired Release-recapture Model

## Powerhouse 1 Ice and Trash Sluiceway

Throughout the migration season the MGR and the ice and trash sluiceway, at powerhouse 1 were not operating continuously due to a policy that prioritized the passage of water through powerhouse 2 . Because of this policy, powerhouse 1 was only operated sporadically for short time intervals before and after the releases of radio tagged fish associated with this study.

## Assumption Tests

## Burnham Tests

The results of the Burnham Tests 2 and 3 testing assumptions A5 and A6 for subyearling Chinook salmon released into the ice and trash sluiceway at Bonneville Dam's powerhouse 1 and the corresponding tailrace releases below the powerhouse 2 JBS outfall were mostly incalculable because of the presence of all zeroes in either rows or columns of the contingency table. The results of these tests can be found in Appendix 4, Table A4.9.

## Tests of the assumption of mixing of treatment and control groups

The chi-square tests of homogeneity testing for the similarity in arrival times of paired releases of subyearling Chinook salmon indicated that there were no significant differences in arrival times between the two release groups at the downstream radio-telemetry arrays (Appendix 5, Table A5.9).

## Survival Estimation

We estimated the survival of subyearling Chinook salmon released into the ice and trash sluiceway from 21 June to 22 July at Bonneville Dam's powerhouse 1. The control group for this evaluation was released in the tailrace below the outfall of the powerhouse 2 juvenile bypass outfall. The estimated survival of subyearling Chinook salmon released into the ice and trash sluiceway ranged from 0.67 to 1.11 (Table 20), during 2004. The average survival was estimated to be 0.925 ( $\mathrm{SE}=0.019,95 \%$ confidence interval [0.887, 0.963]).

## Comparison of 56 kefs day/TDG night and 23 kefs spill operations

The survival of subyearling Chinook salmon released into the ice and trash sluiceway during two spill operations, 56 kcfs during the day and night spill until the total dissolved gas cap of $125 \%$ is reached in the tailrace of the dam ( 56 kcfs day/TDG night) and an alternate operation of 23 kcfs for 24 h ( 23 kcfs ), was also estimated. The average survival of subyearling Chinook salmon through the ice and trash sluiceway during the 56 kcfs day/TDG night spill operations was estimated to be 0.916 ( $\mathrm{SE}=0.025,95 \%$ confidence interval [0.862, 0.969]). For subyearling Chinook salmon passing via the ice and trash sluiceway during the 23 kcfs spill operations the average survival was estimated to be 0.934 ( $\mathrm{SE}=0.028,95 \%$ confidence interval [0.875, $0.994]$ ). The average survival of subyearling Chinook salmon during 56 kcfs day/TDG night spill operations was not significantly different than during 23 kcfs spill operations (two-tailed $t$ test, $P=0.6261, \beta=0.076$ ).

## Comparison of 56 kcfs day and 23 kefs day spill operations

We also compared the survival of subyearling Chinook salmon during daytime passage between 0400 and 2200 hrs during the two spill operations. Survival of subyearling Chinook salmon through the ice and trash sluiceway during the 56 kcfs spill operations was estimated to be 0.944 ( $\mathrm{SE}=0.034,95 \%$ confidence interval [0.863, 1.024]). For subyearling Chinook salmon passing via the ice and trash sluiceway during the 23 kcfs daytime spill operations the estimated survival was 0.944 ( $\mathrm{SE}=0.034,95 \%$ confidence interval [ $0.863,1.025$ ]). The average survival of subyearling Chinook salmon during 56 kcfs spill operations was not significantly different than the survival during 23 kcfs day spill operations (two-tailed $t$-test, $P=0.9950, \beta=0.050$ ).

## Comparison of TDG night and 23 kcfs night spill operations

We also compared the survival of subyearling Chinook salmon during nighttime passage between 2200 and 0400 hrs of TDG spill operations to those during 23 kcfs spill operations. Survival of subyearling Chinook salmon through the ice and trash sluiceway during the TDG spill operations was estimated to be 0.888 ( $\mathrm{SE}=0.037,95 \%$ confidence interval [ $0.801,0.974]$ ). For subyearling Chinook salmon passing via the ice and trash sluiceway during the 23 kcfs night spill operations, the estimated survival was 0.924 ( $\mathrm{SE}=0.046,95 \%$ confidence interval [0.815, 1.033]). The average survival of subyearling Chinook salmon during TDG spill operations was
not significantly different than during 23 kcfs night spill operations (two-tailed $t$-test, $P=$ $0.5433, \beta=0.090$ ).

Table 20. The estimated survival (S) and standard error (SE) of subyearling Chinook salmon released into the ice and trash sluiceway at Bonneville Dam's powerhouse 1 during two dam operations, summer 2004. Dam operations were 56 kcfs during the day with total dissolved gas cap at night ( $56 \mathrm{kcfs} / \mathrm{TDG}$ ) or 23 kcfs for 24 h . Releases were made directly into the ice and trash sluiceway with the control release below the powerhouse 2 juvenile bypass (JBS) outfall at Bonneville Dam. The survival estimates are for the fish released directly into the ice and trash sluiceway to the release location of the tailrace release group. The specific dates and times of the releases can be referenced in Appendix 1.

| Release | Powerhouse 1 Ice and trash sluiceway and downstream tailrace control |  |  |
| :---: | :---: | :---: | :---: |
|  | Spill Operations | S | SE |
| 1 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 1.01 | 0.07 |
| 2 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 1.05 | 0.08 |
| 3 | 23 kcfs | 0.87 | 0.08 |
| 4 | 23 kcfs | 1.04 | 0.16 |
| 5 | 23 kcfs | 1.04 | 0.12 |
| 6 | 23 kcfs | 0.80 | 0.09 |
| 7 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.97 | 0.06 |
| 8 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.86 | 0.09 |
| 9 | 23 kcfs | 1.01 | 0.09 |
| 10 | 23 kcfs | 0.72 | 0.09 |
| 11 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 1.07 | 0.09 |
| 12 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.89 | 0.09 |
| 13 | 23 kcfs | 0.92 | 0.06 |
| 14 | 23 kcfs | 0.90 | 0.06 |
| $15^{\text {a }}$ | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.85 | 0.08 |
| $16^{\text {a }}$ | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.89 | 0.09 |
| 17 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 1.06 | 0.14 |
| 18 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.91 | 0.10 |
| 19 | 23 kcfs | 0.75 | 0.12 |
| 20 | 23 kcfs | 0.86 | 0.11 |
| 21 | 23 kcfs | 0.99 | 0.04 |
| 22 | 23 kcfs | 0.93 | 0.08 |
| 23 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.83 | 0.10 |
| 24 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.89 | 0.09 |
| 25 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.89 | 0.10 |
| 26 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.94 | 0.04 |
| 27 | 23 kcfs | 1.01 | 0.06 |
| 28 | 23 kcfs | 1.03 | 0.02 |
| 29 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.86 | 0.09 |
| 30 | $56 \mathrm{kcfs} / \mathrm{TDG}$ | 0.67 | 0.12 |
| 31 | 23 kcfs | 0.96 | 0.09 |
| 32 | 23 kcfs | 1.11 | 0.08 |

[^4]
## Spillway Flow Deflectors

## 7-ft spillbay flow deflectors during two spill operations

## Assumption Tests

## Burnham Tests

The results of the Burnham Tests 2 and 3 testing assumptions A5 and A6 for the yearling Chinook passing via spillbays with 7 -ft deflectors at Bonneville Dam's spillway and the corresponding tailrace releases below the powerhouse 2 JBS outfall were inconclusive due to the number of tests that were incalculable because of the presence of all zeroes in either rows or columns of the contingency table. Of the 78 tests (Test 2) that were calculated 5 indicated lack of fit ( $P<0.0003$, Dunn-Sidak experimentwise error rate), while goodness of fit (Test 3) was not rejected in any of the tests. The results of these tests can be found in Appendix 4, Table A4.12.

## Tests of the assumption of mixing of the treatment and control groups

The chi-square tests of homogeneity testing for the similarity in arrival times of post-hoc paired releases of subyearling Chinook salmon passing via 7 - ft deflectors at Bonneville Dam's spillway and the corresponding tailrace releases below the powerhouse 2 JBS outfall, indicated that there were no significant differences ( $P<0.0009$, Dunn-Sidak experimentwise error rate) in arrival times between the two groups (Appendix 5, Table A5.12).

## Survival estimation

The survival of subyearling Chinook salmon passing via spillbays with 7-ft deflectors at Bonneville Dam spillway during 56 kcfs day and total dissolved gas cap night spill operations ranged from 0.832 to $1.055(\bar{X}=0.920 ; S E=0.010 ; 95 \% C I=[0.899,0.941]$; Table 21). The estimated survival of subyearling Chinook salmon passing via spillbays with 7 - ft deflectors during 23 kcfs spill operations ranged from 0.583 to 1.000
( $\bar{X}=0.822 ; S E=0.033 ; 95 \% C I=[0.758,0.886] ;$ Table 22).

Table 21. The estimated survival probabilities (S), standard errors (SE), $95 \%$ confidence intervals, the number released ( N ), and dates and times associated with paired releases of radiotagged subyearling Chinook salmon passing via spillbays with deflectors located 7 feet above mean sea level during 24 h spill with 56 kcfs during the day and night spill until the total dissolved gas cap of $125 \%$ is reached in the tailrace. The paired release groups were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

| Release | S | SE | 95\% CI |  | Treatment N | Control N | Start date | time | End date | time |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1.055 | 0.048 | 0.961, | 1.149 | 50 | 112 | 21-Jun | 07:00 | 22-Jun | 07:00 |
| 2 | 0.890 | 0.041 | 0.810, | 0.970 | 55 | 175 | 25-Jun | 19:00 | 27-Jun | 07:00 |
| 3 | 0.980 | 0.032 | 0.917, | 1.042 | 35 | 115 | 27-Jun | 07:00 | 28-Jun | 07:00 |
| 4 | 0.938 | 0.056 | 0.828, | 1.047 | 38 | 189 | 29-Jun | 19:00 | 01-Jul | 07:00 |
| 5 | 0.894 | 0.077 | 0.743, | 1.044 | 31 | 112 | 01-Jul | 07:00 | 02-Jul | 07:00 |
| 6 | 0.889 | 0.051 | 0.788, | 0.989 | 53 | 234 | 03-Jul | 19:00 | 05-Jul | 19:00 |
| 7 | 0.926 | 0.076 | 0.777, | 1.076 | 24 | 111 | 05-Jul | 19:00 | 06-Jul | 19:00 |
| 8 | 0.883 | 0.067 | 0.753, | 1.013 | 40 | 182 | 06-Jul | 19:00 | 08-Jul | 07:00 |
| 9 | 0.897 | 0.055 | 0.790, | 1.004 | 44 | 168 | 11-Jul | 19:00 | 13-Jul | 07:00 |
| 10 | 0.832 | 0.066 | 0.703, | 0.960 | 56 | 120 | 13-Jul | 07:00 | 14-Jul | 07:00 |
| 11 | 0.896 | 0.065 | 0.768, | 1.024 | 36 | 138 | 14-Jul | 07:00 | 15-Jul | 07:00 |
| 12 | 0.837 | 0.102 | 0.638, | 1.036 | 21 | 91 | 15-Jul | 07:00 | 16-Jul | 07:00 |
| 13 | 1.000 | 0.038 | 0.926, | 1.074 | 48 | 150 | 17-Jul | 19:00 | 18-Jul | 19:00 |
| 14 | 0.947 | 0.061 | 0.827, | 1.067 | 36 | 80 | 18-Jul | 19:00 | 19-Jul | 07:00 |
| 15 | 0.948 | 0.056 | 0.839 , | 1.057 | 41 | 159 | 19-Jul | 07:00 | 20-Jul | 07:00 |
| 16 | 0.923 | 0.059 | 0.808, | 1.038 | 61 | 253 | 23-Jul | 19:00 | 24-Jul | 19:00 |
| 17 | 0.879 | 0.062 | 0.757, | 1.001 | 82 | 145 | 24-Jul | 19:00 | 25-Jul | 07:00 |
| 18 | 0.925 | 0.068 | 0.792, | 1.057 | 35 | 145 | 25-Jul | 07:00 | 25-Jul | 19:00 |
| 19 | 0.862 | 0.086 | 0.694, | 1.030 | 40 | 141 | 25-Jul | 19:00 | 26-Jul | 07:00 |
| 20 | 0.979 | 0.057 | 0.867, | 1.091 | 34 | 143 | 26-Jul | 07:00 | 26-Jul | 19:00 |
| 21 | 0.913 | 0.072 | 0.773, | 1.054 | 36 | 121 | 26-Jul | 19:00 | 27-Jul | 07:00 |
| 22 | 0.950 | 0.052 | 0.848, | 1.052 | 32 | 121 | 27-Jul | 07:00 | 27-Jul | 19:00 |
| 23 | 0.909 | 0.085 | 0.742, | 1.076 | 34 | 142 | 27-Jul | 19:00 | 28-Jul | 07:00 |
| 24 | 0.927 | 0.063 | 0.803, | 1.051 | 31 | 293 | 29-Jul | 19:00 | 30-Jul | 19:00 |
| 25 | 0.920 | 0.086 | 0.753, | 1.088 | 34 | 147 | 30-Jul | 19:00 | 01-Aug | 07:00 |

Table 22. The estimated survival probabilities (S), standard errors (SE), $95 \%$ confidence intervals, number released ( N ), and dates and times associated with paired releases of radiotagged subyearling Chinook salmon passing via spillbays with deflectors located 7 feet above mean sea level during 23 kcfs 24 h spill operations at Bonneville Dam. The paired release groups were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

| Release | S | SE | 95\% CI |  | Treatment N | $\begin{gathered} \text { Control } \\ \mathrm{N} \end{gathered}$ | Start date | time | End date | time |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.824 | 0.048 | 0.731, | 0.917 | 16 | 168 | 21-Jun | 19:00 | 23-Jun | 07:00 |
| 2 | 0.755 | 0.102 | 0.555, | 0.956 | 22 | 176 | 23-Jun | 07:00 | 24-Jun | 19:00 |
| 3 | 1.000 | 0.030 | 0.941, | 1.059 | 30 | 180 | $24-$ Jun | 19:00 | 26-Jun | 07:00 |
| 4 | 0.866 | 0.114 | 0.643 , | 1.089 | 12 | 242 | 27-Jun | 19:00 | 30-Jun | 07:00 |
| 5 | 0.583 | 0.125 | 0.338 , | 0.829 | 18 | 231 | 01-Jul | 19:00 | 04-Jul | 07:00 |
| 6 | 0.594 | 0.115 | 0.368 , | 0.820 | 23 | 184 | 08-Jul | 07:00 | 09-Jul | 19:00 |
| 7 | 0.698 | 0.104 | 0.493 , | 0.903 | 24 | 294 | 09-Jul | 19:00 | 12-Jul | 07:00 |
| 8 | 0.917 | 0.063 | 0.793, | 1.041 | 36 | 299 | 16-Jul | 07:00 | 18-Jul | 07:00 |
| 9 | 0.863 | 0.086 | 0.694 , | 1.032 | 27 | 161 | 19-Jul | 19:00 | 20-Jul | 19:00 |
| 10 | 0.813 | 0.106 | 0.606, | 1.019 | 18 | 150 | 20-Jul | 19:00 | 21-Jul | 19:00 |
| 11 | 0.906 | 0.111 | 0.689 , | 1.123 | 21 | 136 | 21-Jul | 19:00 | 22-Jul | 19:00 |
| 12 | 0.859 | 0.089 | 0.684, | 1.033 | 30 | 252 | 22-Jul | 19:00 | 24-Jul | 07:00 |
| 13 | 0.909 | 0.093 | 0.726, | 1.092 | 24 | 435 | 27-Jul | 19:00 | 29-Jul | 07:00 |
| 14 | 0.916 | 0.086 | 0.746 , | 1.085 | 21 | 285 | 29-Jul | 07:00 | 30-Jul | 07:00 |

## 14-ft Flow Spillbay Deflectors during two spill operations

## Assumption Tests

## Burnham Tests

The results of the Burnham Tests 2 and 3 testing assumptions A5 and A6 for the subyearling Chinook salmon passing via spillbays with 14 -ft deflectors at Bonneville Dam's spillway and the corresponding tailrace releases below the powerhouse 2 JBS outfall were mostly incalculable because of the presence of all zeroes in either rows or columns of the contingency table. Of the 74 tests (Test 2 ) that were calculated 3 tests indicated lack of fit ( $P<$ 0.0003 , Dunn-Sidak experimentwise error rate) while goodness of fit was not rejected for any of the tests (Test 3). The results of these tests can be found in Appendix 4, Table A4.12.

## Tests of the assumption of mixing of the treatment and control groups

The chi-square tests of homogeneity testing for the similarity in arrival times of post-hoc paired releases of subyearling Chinook salmon passing via 14-ft deflectors at Bonneville Dam's spillway and the corresponding tailrace releases below the powerhouse 2 JBS outfall, indicated that there were no significant ( $P<0.0007$, Dunn-Sidak experimentwise error rate) in arrival times between the two groups (Appendix 5, Table A5.12).

## Survival estimation

The survival of subyearling Chinook salmon passing via spillbays with 14-ft deflectors at Bonneville Dam spillway during 56 kcfs day and total dissolved gas cap night spill operations ranged from 0.552 to 1.066 ( $\bar{X}=0.803 ; S E=0.026 ; 95 \% C I=[0.749,0.857]$; Table 23). The estimated survival of subyearling Chinook salmon passing via spillbays with $14-\mathrm{ft}$ deflectors during 23 kcfs spill operations ranged from 0.553 to 0.913
( $\bar{X}=0.741 ; S E=0.027 ; 95 \% C I=[0.683,0.799] ;$ Table 24) .
Table 23. The estimated survival probabilities (S), standard errors (SE), 95\% confidence intervals, number released ( N ), and dates and times associated with paired releases of radiotagged subyearling Chinook salmon passing via spillbays with deflectors located 14 feet above mean sea level during 24 h spill with 56 kcfs during the day and night spill until the total dissolved gas cap of $125 \%$ is reached in the tailrace. The paired release groups were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

| Release | S | SE | 95\% CI |  | Treatment N | $\begin{gathered} \text { Control } \\ \mathrm{N} \\ \hline \end{gathered}$ | Start <br> date | time | End date | time |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1.066 | 0.048 | 0.971, | 1.160 | 32 | 112 | 21-Jun | 07:00 | 22-Jun | 07:00 |
| 2 | 0.880 | 0.061 | 0.761, | 0.999 | 30 | 175 | 25-Jun | 19:00 | 27-Jun | 07:00 |
| 3 | 0.616 | 0.101 | 0.418, | 0.815 | 25 | 115 | 27Jun | 07:00 | 28-Jun | 07:00 |
| 4 | 0.760 | 0.116 | 0.534, | 0.987 | 27 | 189 | 29-Jun | 19:00 | 01-Jul | 07:00 |
| 5 | 0.947 | 0.067 | 0.816, | 1.078 | 27 | 112 | 01-Jul | 07:00 | 02-Jul | 07:00 |
| 6 | 0.566 | 0.101 | 0.367, | 0.764 | 27 | 234 | 03-Jul | 19:00 | 05-Jul | 19:00 |
| 7 | 0.779 | 0.086 | 0.611, | 0.947 | 27 | 111 | 05-Jul | 19:00 | 06-Jul | 19:00 |
| 8 | 0.872 | 0.077 | 0.722, | 1.023 | 27 | 182 | 06-Jul | 19:00 | 08-Jul | 07:00 |
| 9 | 0.845 | 0.064 | 0.719, | 0.971 | 38 | 168 | 11-Jul | 19:00 | 13-Jul | 07:00 |
| 10 | 0.716 | 0.106 | 0.507, | 0.924 | 21 | 120 | 13-Jul | 07:00 | 14-Jul | 07:00 |
| 11 | 0.833 | 0.075 | 0.686, | 0.980 | 34 | 138 | 14-Jul | 07:00 | 15-Jul | 07:00 |
| 12 | 1.033 | 0.073 | 0.889, | 1.176 | 17 | 91 | 15-Jul | 07:00 | 16-Jul | 07:00 |
| 13 | 0.552 | 0.115 | 0.326, | 0.778 | 21 | 150 | 17-Jul | 19:00 | 18-Jul | 19:00 |
| 14 | 0.862 | 0.089 | 0.686, | 1.037 | 26 | 80 | 18-Jul | 19:00 | 19-Jul | 07:00 |
| 15 | 0.896 | 0.065 | 0.768, | 1.024 | 35 | 159 | 19-Jul | 07:00 | 20-Jul | 07:00 |
| 16 | 0.728 | 0.078 | 0.576, | 0.881 | 49 | 253 | 23-Jul | 19:00 | 24-Jul | 19:00 |
| 17 | 0.835 | 0.082 | 0.675, | 0.995 | 41 | 145 | 24-Jul | 19:00 | 25-Jul | 07:00 |
| 18 | 0.946 | 0.078 | 0.793, | 1.099 | 23 | 145 | 25-Jul | 07:00 | 25-Jul | 19:00 |
| 19 | 0.782 | 0.107 | 0.572, | 0.991 | 28 | 141 | 25-Jul | 19:00 | 26-Jul | 07:00 |
| 20 | 0.745 | 0.087 | 0.575, | 0.914 | 32 | 143 | 26-Jul | 07:00 | 26-Jul | 19:00 |
| 21 | 0.903 | 0.061 | 0.783, | 1.023 | 48 | 121 | 26-Jul | 19:00 | 27-Jul | 07:00 |
| 22 | 0.610 | 0.130 | 0.355, | 0.865 | 15 | 121 | 27-Jul | 07:00 | 27-Jul | 19:00 |
| 23 | 0.830 | 0.095 | 0.643, | 1.016 | 33 | 142 | 27-Jul | 19:00 | 28-Jul | 07:00 |
| 24 | 0.708 | 0.084 | 0.544, | 0.872 | 33 | 293 | 29-Jul | 19:00 | 30-Jul | 19:00 |
| 25 | 0.773 | 0.084 | 0.608, | 0.937 | 42 | 174 | 30-Jul | 19:00 | 01-Aug | 07:00 |

Table 24. The estimated survival probabilities (S), standard errors (SE), $95 \%$ confidence intervals, number released ( N ; number of fish), and dates and times associated with paired releases of radio-tagged subyearling Chinook salmon passing via spillbays with deflectors located 14 feet above mean sea level during $23 \mathrm{kcfs}, 24 \mathrm{~h}$ spill operations at Bonneville Dam. The paired release groups were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

| Release | S | SE | Treatment |  |  | $\begin{gathered} \text { Control } \\ \mathrm{N} \end{gathered}$ | Start date | time $\begin{aligned} & \text { End } \\ & \text { date }\end{aligned}$ |  | time |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | CI | N |  |  |  |  |  |
| 1 | 0.846 | 0.111 | 0.628, | 1.065 | 17 | 168 | 21-Jun | 19:00 | 23-Jun | 07:00 |
| 2 | 0.643 | 0.123 | 0.403 , | 0.883 | 16 | 176 | 23-Jun | 07:00 | 24-Jun | 19:00 |
| 3 | 0.553 | 0.086 | 0.385 , | 0.722 | 39 | 180 | 24-Jun | 19:00 | 26-Jun | 07:00 |
| 4 | 0.680 | 0.103 | 0.478, | 0.883 | 23 | 242 | 27-Jun | 19:00 | 30-Jun | 07:00 |
| 5 | 0.833 | 0.073 | 0.689 , | 0.977 | 29 | 286 | 01-Jul | 19:00 | 04-Jul | 07:00 |
| 6 | 0.844 | 0.085 | 0.677, | 1.011 | 25 | 184 | 08-Jul | 07:00 | 09-Jul | 19:00 |
| 7 | 0.781 | 0.094 | 0.597 , | 0.966 | 24 | 294 | 09-Jul | 19:00 | 12-Jul | 07:00 |
| 8 | 0.760 | 0.115 | 0.535 , | 0.986 | 18 | 299 | 16-Jul | 07:00 | 18-Jul | 07:00 |
| 9 | 0.674 | 0.117 | 0.445 , | 0.902 | 19 | 161 | 19-Jul | 19:00 | 20-Jul | 19:00 |
| 10 | 0.688 | 0.095 | 0.502 , | 0.873 | 26 | 150 | 20-Jul | 19:00 | 21-Jul | 19:00 |
| 11 | 0.779 | 0.097 | 0.589, | 0.969 | 35 | 136 | 21-Jul | 19:00 | 22-Jul | 19:00 |
| 12 | 0.913 | 0.058 | 0.800, | 1.026 | 62 | 252 | 22-Jul | 19:00 | 24-Jul | 07:00 |
| 13 | 0.648 | 0.092 | 0.467, | 0.828 | 41 | 435 | 27-Jul | 19:00 | 29-Jul | 07:00 |
| 14 | 0.726 | 0.086 | 0.559, | 0.894 | 33 | 285 | 29-Jul | 07:00 | 30-Jul | 07:00 |

## Comparison of subyearling Chinook salmon survival estimates for $\mathbf{7}$ - $\mathbf{f t}$ and $\mathbf{1 4} \mathbf{- f t}$ spillbay deflectors during two spill operations

The point estimates of survival for subyearling Chinook salmon passing through the 7 - ft deflectors were consistently higher than the survival point estimates for fish passing through the 14 -ft deflectors (Figures 20 and 21) for both spill conditions. Survival point estimates for subyearling Chinook salmon passing through both the 7 - ft and the 14 - ft spillbay deflectors were higher during the 56 kcfs day and total dissolved gas cap night spill operations than during the 23 kcfs 24 h spill operations.


Release
Figure 20. The estimated survival probabilities ( $95 \%$ CI error bars) from paired releases of radio-tagged subyearling Chinook salmon passing through spillbays with either deflectors located at 7 feet or 14 feet above mean sea level at Bonneville Dam under the 56 kcfs day/total dissolved gas night spill operation, by release. The paired release groupings were formed posthoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.


Figure 21. The estimated survival probabilities ( $95 \%$ CI error bars) from paired releases of radio-tagged subyearling Chinook salmon passing through spillbays with either deflectors located at 7 feet or 14 feet above mean sea level at Bonneville Dam under $24 \mathrm{~h}, 23 \mathrm{kcfs}$ spill, by release. The paired release groupings were formed post-hoc from fish released at The Dalles Dam (treatment) and in the Bonneville Dam tailrace (control) during 2004.

## Route-specific Survival Model

Using capture histories generated from the detections of radio-tagged subyearling Chinook salmon (Table 25) released at The Dalles Dam and in the tailrace of Bonneville Dam, we generated maximum likelihood estimates of the route-specific passage and survival probabilities for subyearling Chinook salmon through Bonneville Dam during two spill operations from 20 June to 22 July (Figures 22 and 23). The two spill operations tested were 56 kcfs during the day ( 0400 to 2200 hrs ) and night spill until $125 \%$ total dissolved gas cap was reached in the tailrace ( 56 kcfs day/TDG night) and 24 hr spill operations of 23 kcfs ( 23 kcfs ).

## 56 kcfs day/TDG night spill operations

During the 56 kcfs day/TDG night spill operations, the survival of subyearling Chinook salmon through the Bonneville Dam spillway was estimated to be 0.877 ( $\mathrm{SE}=0.013$ profile likelihood $95 \%$ confidence interval [0.848, 0.902]). For subyearling Chinook salmon passing via powerhouse 1, the estimated survival was 0.827 ( $\mathrm{SE}=0.061$, profile likelihood $95 \%$ confidence interval [0.694, 0.937]) and for subyearling Chinook salmon passing via powerhouse 2 turbines, the estimated survival was 0.824 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval [0.782, $0.864]$ ). For subyearling Chinook salmon passing via the powerhouse 2 JBS during the 56 kcfs day/TDG night dam operations, the estimated survival was 0.927 ( $\mathrm{SE}=0.027$, profile likelihood $95 \%$ confidence interval $[0.863,0.976])$. For subyearling Chinook salmon passing via the powerhouse 2 corner collector the estimated survival was 0.981 ( $\mathrm{SE}=0.013$, profile likelihood $95 \%$ confidence interval [ $0.951,1.005]$ ). Subyearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.891 ( $\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval $[0.871,0.910$ ) and project survival was estimated to be 0.768 ( $\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval [0.747, 0.788].

## 56 kefs day spill operations

During the 56 kcfs spill operations the survival of subyearling Chinook salmon through the Bonneville Dam spillway was estimated to be 0.851 ( $\mathrm{SE}=0.016$ profile likelihood $95 \%$ confidence interval [0.819, 0.883]). For subyearling Chinook salmon passing via powerhouse 1, the estimated survival was 0.845 ( $\mathrm{SE}=0.059$, profile likelihood $95 \%$ confidence interval [0.715, $0.943]$ ) and for subyearling Chinook salmon passing via powerhouse 2 turbines during the 56 kcfs spill operations the estimated survival was 0.834 ( $\mathrm{SE}=0.021$, profile likelihood $95 \%$ confidence interval [0.792, 0.873]). For subyearling Chinook salmon passing via the powerhouse 2 JBS the estimated survival was 0.900 ( $\mathrm{SE}=0.031$, profile likelihood $95 \%$ confidence interval [ $0.833,0.955]$ ). For subyearling Chinook salmon passing via the powerhouse 2 corner collector the estimated survival was 0.966 ( $\mathrm{SE}=0.014$, profile likelihood $95 \%$ confidence interval [ $0.937,0.991]$ ). Subyearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.882 ( $\mathrm{SE}=0.011,95 \%$ confidence interval $[0.861,0.903$ ) and project survival was estimated to be 0.763 ( $\mathrm{SE}=0.012$, profile likelihood $95 \%$ confidence interval [0.741, 0.785]).

## Total dissolved gas night spill operations

During the TDG spill operations, the survival of subyearling Chinook salmon through the Bonneville Dam spillway was estimated to be 0.913 ( $\mathrm{SE}=0.021$ profile likelihood $95 \%$ confidence interval [0.869, 0.953]). For subyearling Chinook salmon passing via powerhouse 1,
the estimated survival was 0.519 ( $\mathrm{SE}=0.266$, profile likelihood $95 \%$ confidence interval [0.109, $0.947]$ ) and for subyearling Chinook salmon passing via powerhouse 2 turbines the estimated survival was 0.696 ( $\mathrm{SE}=0.062$, profile likelihood $95 \%$ confidence interval [0.570, 0.810]). For subyearling Chinook salmon passing via the powerhouse 2 JBS the estimated survival was 0.998 ( $\mathrm{SE}=0.048$, profile likelihood $95 \%$ confidence interval [0.868, 1.060]). For subyearling Chinook salmon passing via the powerhouse 2 corner collector the estimated survival was 1.009 ( $\mathrm{SE}=0.041$, profile likelihood $95 \%$ confidence interval [0.898, 1.063]). Subyearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.887 ( $\mathrm{SE}=0.022$, profile likelihood $95 \%$ confidence interval [ $0.847,0.925$ ) and project survival was estimated to be 0.757 $(\mathrm{SE}=0.023$, profile likelihood $95 \%$ confidence interval [0.715, 0.797]).

## 23 kcfs 24 h spill operations

During the 23 kcfs spill operations the survival of subyearling Chinook salmon through the Bonneville Dam spillway was estimated to be 0.744 ( $\mathrm{SE}=0.022$, profile likelihood $95 \%$ confidence interval [ $0.700,0.786]$ ). For subyearling Chinook salmon passing via powerhouse 1 the estimated survival was 0.829 ( $\mathrm{SE}=0.030$, profile likelihood $95 \%$ confidence interval [0.767, $0.884]$ ) and for subyearling Chinook salmon passing via powerhouse 2 turbines the estimated survival was 0.833 ( $\mathrm{SE}=0.014$, profile likelihood $95 \%$ confidence interval [0.805, 0.860]). For subyearling Chinook salmon passing via the JBS the estimated survival was 0.958 ( $\mathrm{SE}=0.019$, profile likelihood $95 \%$ confidence interval $[0.918,0.991]$ ) and passing via the corner collector at powerhouse 2 the estimated survival was 0.954 ( $\mathrm{SE}=0.013$, profile likelihood $95 \%$ confidence interval $[0.926,0.978]$ ). Subyearling Chinook salmon survival through Bonneville Dam was estimated to be 0.858 ( $\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval [0.840, 0.876]) and project survival was estimated to be 0.736 ( $\mathrm{SE}=0.010$, profile likelihood $95 \%$ confidence interval [0.717, 0.754]).

## 23 kcfs day spill operations

For subyearling Chinook salmon passing during day ( 0400 to 2200 hrs ) 23 kcfs spill operations, survival through the Bonneville Dam spillway was estimated to be 0.725 ( $\mathrm{SE}=$ 0.025 , profile likelihood $95 \%$ confidence interval [0.675, 0.773]). For subyearling Chinook salmon passing via powerhouse 1 the estimated survival was 0.826 ( $\mathrm{SE}=0.033$, profile likelihood $95 \%$ confidence interval [ $0.758,0.886]$ ) and for subyearling Chinook salmon passing via powerhouse 2 turbines, the estimated survival was 0.820 ( $\mathrm{SE}=0.017$, profile likelihood $95 \%$ confidence interval $[0.785,0.853]$ ). For subyearling Chinook salmon passing via the powerhouse 2 JBS , the estimated survival was 0.957 ( $\mathrm{SE}=0.024$, profile likelihood $95 \%$ confidence interval [0.906, 0.999]). For subyearling Chinook salmon passing via the powerhouse 2 corner collector the estimated survival was 0.955 ( $\mathrm{SE}=0.015$, profile likelihood $95 \%$ confidence interval [ $0.926,0.983]$ ). Subyearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.852 ( $\mathrm{SE}=0.012$, profile likelihood $95 \%$ confidence interval $[0.831,0.875]$ ) and project survival was estimated to be 0.731 ( $\mathrm{SE}=0.012$, profile likelihood $95 \%$ confidence interval [0.710, 0.753]).

## 23 kefs night spill operations

For subyearling Chinook salmon passing during the night (2200 to 0400 hrs ) 23 kcfs spill operations, survival through the Bonneville Dam spillway was estimated to be 0.830 ( $\mathrm{SE}=$ 0.045 , profile likelihood $95 \%$ confidence interval [0.735, 0.909]). For subyearling Chinook
salmon passing via powerhouse 1 the estimated survival was 0.863 ( $\mathrm{SE}=0.077$, profile likelihood $95 \%$ confidence interval [ $0.687,0.982$ ]) and for subyearling Chinook salmon passing via powerhouse 2 turbines, the estimated survival was 0.864 ( $\mathrm{SE}=0.023$, profile likelihood $95 \%$ confidence interval [0.818, 0.907]). For subyearling Chinook salmon passing via the powerhouse 2 JBS the estimated survival was 0.960 ( $\mathrm{SE}=0.030$, profile likelihood $95 \%$ confidence interval [0.891, 1.010]). For subyearling Chinook salmon passing via the powerhouse 2, corner collector the estimated survival was 0.959 ( $\mathrm{SE}=0.045$, profile likelihood $95 \%$ confidence interval [ $0.847,1.025]$ ). Subyearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.881 ( $\mathrm{SE}=0.019$, profile likelihood $95 \%$ confidence interval [ $0.847,0.914]$ ) and project survival was estimated to be $0.755(\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval [0.719, 0.791]).

Table 25. Counts of radio-tagged subyearling Chinook salmon for the releases from The Dalles Dam $\left(R_{l}\right)$ and in the tailrace of Bonneville Dam $\left(R_{2}\right)$ during two spill operations: 56 kcfs day/total dissolve gas cap at night (TDG) and 23 kcfs for 24 h used in the route-specific survival model during 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only ( 01 ) within the passage route. B1 is powerhouse 1 and B 2 is powerhouse 2 at Bonneville Dam.

| 56 kcfs day/TDG night spill operations |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Release | Detection History | Route | Counts | Within-route histories Bonneville Dam |  |  |
|  |  |  |  | 11 | 01 | 10 |
| $\mathrm{R}_{1}=2714$ | 100 |  | 385 |  |  |  |
|  | 101 |  | 35 |  |  |  |
|  | 110 | Spillway | 176 | 863 | 11 | 168 |
|  | 111 |  | 866 |  |  |  |
|  | 110 | B1 | 11 | 40 | 12 | 0 |
|  | 111 |  | 41 |  |  |  |
|  | 110 | B2 Turbines | 108 | 251 | 55 | 190 |
|  | 111 |  | 388 |  |  |  |
|  | 110 | B2 Juvenile bypass | 20 | 162 | 1 | 2 |
|  | 111 |  | 145 |  |  |  |
|  | 110 | B2 Corner collector | 38 | 530 | 2 | 7 |
|  | 111 |  | 501 |  |  |  |
| $\mathrm{R}_{2}=1835$ | 010 |  | 94 |  |  |  |
|  | 011 |  | 1741 |  |  |  |
| 23 kcfs for 24 h spill operations |  |  |  |  |  |  |
| Release | Detection History | Route | Counts | With | oute <br> evill |  |
|  |  |  |  | 11 | 01 | 10 |
| $\mathrm{R}_{1}=3269$ | 100 |  | 483 |  |  |  |
|  | 101 |  | 61 |  |  |  |
|  | 110 | Spillway | 146 | 410 | 15 | 71 |
|  | 111 |  | 350 |  |  |  |
|  | 110 | B1 | 46 | 148 | 67 | 0 |
|  | 111 |  | 169 |  |  |  |
|  | 110 | B2 Turbines | 225 | 504 | 93 | 478 |
|  | 111 |  | 850 |  |  |  |
|  | 110 | B2 Juvenile bypass | 27 | 289 | 1 | 3 |
|  | 111 |  | 266 |  |  |  |
|  | 110 | B2 Corner collector | 62 | 625 | 4 | 17 |
|  | 111 |  | 584 |  |  |  |
| $\mathrm{R}_{2}=1960$ | $010$ |  | $102$ |  |  |  |
|  | 011 |  | 1858 |  |  |  |



Figure 22. Schematic of estimated route-specific passage and survival parameters for subyearling Chinook salmon through Bonneville dam during 20 June through 22 July, during 56 kcfs day and total dissolved gas cap at night spill operations in 2004. Estimated standard errors are in parentheses.


Figure 23. Schematic of estimated route-specific passage and survival parameters for subyearling Chinook salmon through Bonneville dam during 20 June through 22 July, during 23 kcfs 24 h spill operations in 2004. Estimated standard errors are in parentheses.

## Comparison of survival during 56 kcfs/TDG and 23 kcfs for $\mathbf{2 4} \mathbf{h}$ spill operations

The estimated survival probabilities for subyearling Chinook salmon passing via the powerhouse 1 and 2 turbines, and the powerhouse 2 corner collector and JBS were not found to be significantly different between the two spill operations. However, the estimated survival of subyearling Chinook salmon passing via the spillway was significantly different for fish passing during the 56 kcfs day/TDG night spill operations and the 23 kcfs spill operations (Table 26). The estimated dam survival for fish passing during the two spill operations was also found to be significantly different.

Table 26. Summary table of estimated route-specific survival probabilities (S) and their associated standard errors (SE) of subyearling Chinook salmon survival through Bonneville Dam (Dam survival) generated from the route-specific survival model at Bonneville Dam during 20 June to 22 July 2004. The results of $Z$-tests (i.e., $Z$-statistic and $P$-value) structured to assess whether the estimated survival probabilities during the 56 kcfs day/TDG night spill operations were different than the estimated survival probabilities during the 23 kcfs spill operations. Significant results are indicated where $Z \geq 1.645$ given a two-tailed test and $\alpha=0.10$. The JBS refers to the juvenile bypass system at powerhouse 2 .

|  | 56 kcfs day/TDG night |  |  |  | 23 kcfs |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Passage route | S | SE |  | S | SE |  |  |  |

## Discussion

During our study, the Army Corps of Engineers identified a discrepancy between the reported inflow (The Dalles outflow + tributary inflow) and outflow from Bonneville Dam during times of spill. The reported spillway discharge was greater than the actual discharge (as measured downstream of the dam). Two issues resulted in this discrepancy. The spillway gate hoist mechanism had been miscalibrated and the spillway rating curve had not been updated when the spill gate lip changed from a rounded to a sharp edged design in the 1970's. The greatest impact of the gate miscalibration was on discharges with small gate openings, magnified by the spill pattern adopted in 2002 that utilized more gates at smaller openings for a given flow. The magnitude of this discrepancy varied but was on the order of 20 kcfs less than what had been reported (see: Bonneville Lock and Dam, Revision to Fish Passage Plan Spill Patterns FEB2005, Memorandum; Appendix 6).

During our evaluation of survival through the MGR and the ice and trash sluiceway at powerhouse 1 , the powerhouse 1 was not continuously operated because of a policy that prioritized the passage of water through powerhouse 2 . Instead it was operated sporadically for short time intervals before and after the time of our site-specific releases into the MGR turbine
unit and the ice and trash sluiceway. Consequently, $14 \%$ and $6 \%$ of the total discharge during spring and summer study periods passed through powerhouse 1.

Route-specific survival estimates for yearling Chinook salmon indicated that the estimated survival through the powerhouse 2 , corner collector (1.016, [0.999, 1.032] $95 \%$ profile likelihood confidence interval) was higher than for all other passage routes at Bonneville Dam, followed by the powerhouse $2 \mathrm{JBS}(0.970,[0.943,0.994] 95 \%$ profile likelihood confidence interval) and powerhouse 2 turbines ( $0.951,[0.929,0.972] 95 \%$ profile likelihood confidence interval). Survival estimates through the spillway were the lowest ( $0.910,[0.888,0.931] 95 \%$ profile likelihood confidence interval) of all routes. Reagan et al. (2005) demonstrated that the passage route was influenced by discharge. For the spring migration season $51 \%$ of the overall discharge was passed through powerhouse 2 , with $27 \%$ discharge through the spillway and $19 \%$ through powerhouse 1. The passage results from Reagan et al. (2005) indicated that $59 \%$ of yearling Chinook salmon passed via powerhouse $2,33 \%$ passed via the spillway, and only $8 \%$ passed via the powerhouse 1. For yearling Chinook salmon passage within powerhouse 2, 43\% passed via the turbines, $36 \%$ via the corner collector, and $21 \%$ via the JBS. These results were consistent with our route-specific survival estimates and with the dam survival of 0.951 ([0.936, $0.966] 95 \%$ confidence interval), which was likely influenced by the large proportion of fish passing through powerhouse 2 .

The point estimate of dam survival during 2004 is lower than the estimated dam survival during 2002. The priority to pass water through the powerhouse 2 during 2004 could have effected overall survival through Bonneville Dam. Further, in 2002, an additional day spill condition allowing discharge until the total dissolved gas cap of $125 \%$ was reached in the tailrace was evaluated. In 2002, $46 \%$ of the overall discharge passed through the spillway, $40 \%$ via powerhouse 2 , and $14 \%$ through powerhouse 1 (Evans et al. 2003). The miscalibration error resulting in spillway discharge during 2004 being less than the recommended BIOP spill may have influenced the relatively low survival observed at the spillway. Despite the higher survival for fish passing via the new powerhouse 2 corner collector, $43 \%$ of yearling Chinook salmon at powerhouse 2 still passed unguided via the turbines.

Similar to yearling Chinook salmon, the estimated survival of steelhead trout was greatest through the powerhouse 2 corner collector ( 1.030 [1.014, 1.047] 95\% profile likelihood confidence interval), followed by the spillway, powerhouse 1 , the powerhouse 2 JBS, and finally the powerhouse 2 turbines. Reagan et al. (2005) evaluated passage routes of steelhead trout released at The Dalles Dam and passed at Bonneville Dam. Their results indicated $66 \%$ of the steelhead trout passed via powerhouse $2,25.2 \%$ passed via the spillway, and $8.5 \%$ passed via the powerhouse 1. These results are again consistent with the overall discharge proportions through each route. Steelhead trout passage via the corner collector was very high at $74 \%$ of fish passing powerhouse 2, the other $16 \%$ passed through the turbines and $10 \%$ through the JBS. The high dam survival estimate of 0.991 ( $[0.974,1.007] 95 \%$ confidence interval) is likely a result of the high passage proportions and survival estimates through the corner collector.

Route-specific survival for subyearling Chinook salmon at Bonneville Dam was evaluated during two spill operations. In general, the route-specific survival estimates were higher during the higher spill operation of 56 kcfs day/TDG night spill than for the 23 kcfs spill for 24 h . In particular, we observed significant differences between the survival estimates for the powerhouse 2 corner collector, the spillway, and the overall dam survival between the two spill operations. The differences we observed are likely a result of the different proportions of total discharge through the various routes (i.e. during the 56 kcfs day/TDG night spill operations more
fish pass via the spillway, and during the 23 kcfs spill operations more fish pass via powerhouse 2 turbines and the corner collector).

Evans et al. (2005) evaluated passage of subyearling Chinook salmon released from John Day Dam and The Dalles Dam and passed at Bonneville Dam, their results indicated $60 \%$ fish passed at powerhouse 2, $35 \%$ at the spillway, and $5 \%$ at powerhouse 1 for the season. This follows the proportion of total discharge through each route of $56 \%$ at powerhouse $2,34 \%$ at the spillway, and $6 \%$ through powerhouse 1 . It was also noted that during the 56 kcfs day/TDG night spill operations $50 \%$ of the subyearling Chinook salmon passed via the spillway, where $47 \%$ of the total discharge was apportioned. During the 23 kcfs spill operation, $72 \%$ of the subyearling Chinook salmon passed via the powerhouse 2 , which operated at $65 \%$ of the total discharge. Of the subyearling Chinook salmon that passed powerhouse 2 throughout the season, $49 \%$ passed via the turbines, $37 \%$ passed via the corner collector and $14 \%$ passed via the JBS. For subyearling Chinook salmon passing during the 56 kcfs day/TDG night spill operations, the survival estimate through the powerhouse 2 , corner collector was highest, followed by the powerhouse 2 JBS, the spillway, powerhouse 1, and powerhouse 2 turbines. During the 23 kcfs spill operations the powerhouse 2 JBS route had the highest survival estimate, followed by the powerhouse 2 corner collector, powerhouse 2 turbines, powerhouse 1, and the lowest survival estimate was through the spillway.

The 2004 dam survival estimate during the $56 \mathrm{kcfs} /$ TDG spill operations $(0.891,[0.871$, 0.910 ] profile likelihood $95 \%$ confidence interval) is very similar to the 2001 paired releaserecapture estimate of 0.902 ( $[0.831,0.973] 95 \%$ confidence interval) for fish released near Hood River and in the tailrace of Bonneville Dam. The 2004 dam survival estimate for the 23 kcfs spill operations was considerably lower at 0.858 ([0.840, 0.876] profile likelihood $95 \%$ confidence interval). During the 2001 study year very little discharge ( $2.4 \%$ ) passed through the spillway and the majority of the discharge and fish passed via powerhouse 2 (Evans et al. 2001).

In general we observed that spillway survival was typically lower than the survival estimates through the guided powerhouse 2 routes (corner collector and JBS) and that survival estimates through the spillway were lowest during the lower spill discharge operating conditions. Similarly, it was observed that survival of fish passing via the $7-\mathrm{ft}$ and 14 - ft deflector spillbays was higher during higher flow conditions (TDG spill operations versus 48 and 23 kcfs ). At the lower spill conditions of 56 kcfs or 23 kcfs the 7 -ft spillbay deflector consistently had higher survival estimates than for the 14 ft -spillbay deflector. During the spring migration the mean tail water elevation was 18.1 ft with a range between 14 and 23 ft , and for the summer migration the mean tail water elevation was 13.9 ft , with a range between 9.1 and 19.5 ft above mean sea level. It may be that at the lower tail water elevations and in conjunction with the lower spill discharge operations, the shallower 14 ft deflector was less effective.

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## References

Beeman, J. W., H. H. Hansel, P. V. H. Haner, and J. Hardiman. 2001a. Estimates of fish and spill passage efficiency of radio-tagged juvenile steelhead and yearling Chinook salmon at John Day Dam, 2000. Prepared by U. S. Geological Survey, Cook, Washington for the U.S. Army Corps of Engineers, Portland, Oregon, contract W66QKZ00381910.

Beeman, J. W., H. H. Hansel, P. V. H. Haner, and J. Hardiman. 2001b. Estimates of fish-, spilland sluiceway-passage efficiencies of radio-tagged juvenile steelhead and yearling Chinook salmon at The Dalles Dam, 2000. Prepared by U. S. Geological Survey, Cook, Washington for the U.S. Army Corps of Engineers, Portland, Oregon, contract W66QKZ00392015.

Bickford, S. A. and J. R. Skalski. 2000. Reanalysis and interpretation of 25 years of SnakeColumbia River juvenile salmonid survival studies. North American Journal of Fisheries Management, 20:53-68.

Burnham, K.P., D.R Anderson, G.C. White, C. Brownie, and K.H. Pollock. 1987. Design and analysis methods for fish survival estimates based on release-recapture. American Fisheries Society Monograph No. 5.

Counihan, T. D., J. H. Petersen, N. S. Adams, R. S. Shively, and H. C. Hansel. 2001. Feasibility of Extracting Survival Information from Radio-Telemetry Studies at the John Day Dam. Annual report prepared by U. S. Geological Survey, Cook, Washington for the U.S. Army Corps of Engineers, Portland, Oregon.

Counihan, T. D., J. H. Petersen, and K. J. Felton. 2002a. Survival Estimates of migrant Juvenile Salmonids in the Columbia River from John Day Dam through Bonneville Dam using RadioTelemetry, 2000. Annual report prepared by U. S. Geological Survey, Cook, Washington for the U.S. Army Corps of Engineers, Portland, Oregon.

Counihan, T. D., K. J. Felton, G. S. Holmberg, and J. H. Petersen. 2002b. Survival Estimates of migrant Juvenile Salmonids through Bonneville Dam using Radio-Telemetry. Annual report of research for 2001. Prepared by the U. S. Geological Survey, Cook, Washington for the U.S. Army Corps of Engineers, Portland, Oregon. Contract No. W66QKZ1019057.

Counihan, T. D., G. S. Holmberg, and J. H. Petersen. 2003. Survival estimates of migrant juvenile salmonids through Bonneville Dam using radio-telemetry, 2002. Annual report prepared by U. S. Geological Survey, Cook, Washington for the U.S. Army Corps of Engineers, Portland, Oregon. Contract No. W66QKZ20303679.

Cowen, L. and C. J. Schwarz. 2005. Capture-recapture studies using radio telemetry with premature radio-tag failure. Biometrics 61:657-664.

Dawley, E.M., L.G. Gilbreath, E.P. Nunnallee, and B. P. Sandford. 1998. Survival of juvenile salmon passing through the spillway of The Dalles Dam, 1997. Annual report prepared for the U.S. Army Corps of Engineers, Portland, OR. Contract MIPR E96970020.

Evans, S. D., L. S. Wright, R. E. Reagan, N. S. Adams, and D. W. Rondorf. 2005. Passage behavior of radio-tagged subyearling Chinook salmon at Bonneville Dam, 2004. Draft Annual report prepared for the U.S. Army Corps of Engineers, Portland, OR. Contract No. W66QKZ40238289.

Evans, S.D., L.S. Wright, C.D. Smith, R.E. Wardell, N.S. Adams, and D.W. Rondorf. 2003. Passage behavior of radio-tagged yearling chinook salmon and steelhead at Bonneville Dam, 2002. Annual report prepared for the U.S. Army Corps of Engineers, Portland, OR. Contract No. W66QKZ20303685.

Evans, S. D., C. D. Smith, N. S. Adams, and D. W. Rondorf. 2001. Passage behavior of radiotagged subyearling Chinook salmon at Bonneville Dam, 2001. Annual report prepared for the U.S. Army Corps of Engineers, Portland, OR. Contract No. W66QKZ10442576.

Giorgi, A.E., M. Miller, J. Stevenson. 2002. Mainstem passage strategies in the Columbia River System: transportation, spill and flow augmentation. Technical Report Northwest Power Planning Council, Portland, Oregon.

Hockersmith, E. E., W. D. Muir, S. G. Smith, B. P. Sandford, R. W. Perry, N. S. Adams, and D. W. Rondorf. 2003. Comparison of migration rate and survival between radio-tagged and PITtagged migrant yearling Chinook salmon in the Snake and Columbia rivers. North American Journal of Fisheries Management 23: 404-413.

Iwamato, R.N., Muir, W.D., Sandford, McIntyre, Frost, D.A., Williams, J.G., Smith, S.G., and J.R. Skalski. 1994. Survival estimates for the passage of juvenile chinook salmon through Snake River Dams and reservoirs, 1993. Report prepared for the U.S. Department of Energy, Bonneville Power Administration. Division of Fish and Wildlife, Contract DE-A17993BP10891, Project 93-29. 139 pp.

Lady, J.M., P. Westhagen, and J.R. Skalski. 2003. USER 2., User Specified Estimation Routine. U.S. Department of Energy, Bonneville Power Administration, Division of Fish and Wildlife, Portland, OR, Project No. 198910700

Lebreton, J.D., K.P. Burnham, J. Clobert, and D. R. Anderson. 1992. Modeling survival and testing biological hypotheses: using marked animals: A unified approach. Ecological Monographs 62:67-118.

Martinelli, T. L., H. C. Hansel, and R. S. Shively. 1998. Growth and physiological responses to surgical and gastric radio transmitter implantation techniques in subyearling chinook salmon (Oncorhynchus tshawytscha). Hydrobiologia 371/372:79-87.

Muir, W.D., and eleven coauthors. 1995. Survival estimates or the passage of juvenile salmonids through Snake River Dams and reservoirs, 1994. Annual report prepared for the Bonneville Power Administration, Portland, OR and U.S. Army Corps of Engineers, Walla Walla, WA. Contract DE93-29A179-93B101891, Project 93-29. 187 p.
NMFS 2000 FCRPS Biological Opinion .
Normandeau Associates, Inc., J.R. Skalski, and Parametrix, Inc. 1998. Feasibility of estimating smolt survival with radio telemetry through the Priest Rapids hydroelectric project, Columbia River, Washington. Report prepared for Grant County Public Utility District No. 2.

National Marine Fisheries Service 2000 Federal Columbia River Power System Biological Opinion.

Peven, C., A. Giorgi, J. Skalski, M. Langeslay, A. Grassell, S. G. Smith, T. Counihan, R. Perry, and S. Bickford. 2005. Guidelines and recommended protocols for conducting, analyzing, and reporting juvenile salmonid survival studies in the Columbia River Basin. http://www.cbr.washington.edu/papers/Guidelines_and_Protocols_FINAL_DRAFT.pdf

Prentice, E. F., T. A. Flagg, and C. S. McCutcheon. 1990. Feasibility of using implantable passive integrated transponder (PIT) tags in salmonids. American Fisheries Society Symposium 7:317-322.

Raymond H. L. 1979. Effects of dams and impoundments on migrations of juvenile chinook salmon and steelhead from the Snake River, 1966-1975. Transactions of the American Fisheries Society 90:58-72.

Reagan, R. E., S. D. Evans, L. S. Wright, M. J. Farley, N. S. Adams, and D. W. Rondorf. 2005. Movement, distribution, and passage behavior of radio-tagged yearling Chinook salmon and steelhead trout at Bonneville Dam, 2004. Draft Annual report prepared for the U.S. Army Corps of Engineers, Portland, OR. Contract No.W66QKZ240238289.

Seber, G. A. F. 1982. The estimation of animal abundance and related parameters. Macmillan, New York.

Skalski, J. R. 1992. Sample size calculations for normal variates under binomial censoring. Biometrics 48:877-882.

Skalski, J. R., S.G. Smith, R.N. Iwamato, J.G. Williams, and A, Hoffman. 1998. Use of passive integrated transponder tags to estimate survival of migrant juvenile salmonids in the Snake and Columbia rivers. Canadian Journal of Fisheries and Aquatic Sciences. 55:1484-1493.

Skalski, J.R. 1999a. Statistical methods to extract survival information from the John Day and The Dalles dams radiotag studies. Report to the Army Corps of Engineers.

Skalski, J.R. 1999b. Sample Size Calculations for conducting John Day Project smolt survival studies. Report to the Army Corps of Engineers.

Skalski, J. R., J. Lady, R. Townsend, A. E. Giorgi, J. R. Stevenson, C. M. Peven, and R. D. McDonald. 2001. Estimating inriver survival of migrating salmonid smolts using radiotelemetry. Canadian Journal of Fisheries and Aquatic Sciences 58:1987-1997.

Skalski, J. R., R. Townsend, J. Lady, A.E. Giorgi, J.R. Stevenson, and R.D. McDonald. 2002. Estimating route-specific passage and survival probabilities at a hydroelectric project from smolt radio-telemetry studies. Canadian Journal of Fisheries and Aquatic Sciences. 59:1385-1393.

Smith, S. G., J.R. Skalski, J.W. Schlechte, A. Hoffman, and V. Cassen. 1996. Introduction to SURPH. 1 analysis of release-recapture data for survival studies. Report prepared for Bonneville Power Administration, Environment, Fish, and Wildlife, Portland, OR. DOE/BP-02341-3, October 1996.

Smith, S.G., W.D. Muir, E.E. Hockersmith, S. Achord, M.B. Eppard, T.E. Ruehle, and J.G. Williams. 1998. Survival estimates or the passage of juvenile salmonids through Snake River Dams and reservoirs, 1996. Annual report prepared for the Bonneville Power Administration, Portland, OR. Contract DE93-29A179-93B101891, Project 93-29. 197 p.

Sokal, R.R., and F.J. Rolf. 1995. Biometry: the principles and practice of statistics in biological research. $3^{\text {rd }}$ ed. W.H. Freeman and Company, New York. 850 pp.

Stier, D.J. and B. Kynard. 1986. Use of radio telemetry to determine the mortality of Atlantic salmon smolts passed through a 17-MW Kaplan turbine at a low-head hydroelectric dam. Transactions of the American Fisheries Society. 115:771-775.

Townsend, R.L., J. R. Skalski, P. Dillingham, and T. W. Steig. 2004. Correcting Bias in Survival Estimation Resulting from Tag Failure in Acoustic and Radiotelemetry Studies. Report prepared for the Bonneville Power Administration, Contract No. 00012494.

## Appendix 1: Release Dates, Times, Fork Lengths and Weights

Table A1.1. Summary of yearling Chinook salmon releases at Bonneville Dam ice and trash sluiceway during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | 30-Apr | 12:28 | 24 | 1 | 0 | 146 | 10.8 | 130-179 | 30.7 | 7.0 | 23.6-51.9 |
| 2 | 3-May | 00:05 | 24 | 1 | 0 | 142 | 6.3 | 130-155 | 29.1 | 4.5 | 22.9-40.1 |
| 3 | 4-May | 11:19 | 25 | 1 | 0 | 141 | 5.3 | 132-154 | 32.9 | 4.1 | 26.9-43.5 |
| 4 | 6-May | 23:48 | 26 | 0 | 0 | 153 | 12.2 | 133-179 | 35.8 | 9.8 | 25.0-65.6 |
| 5 | 8-May | 12:14 | 25 | 0 | 0 | 147 | 7.6 | 140-165 | 31.1 | 5.2 | 25.2-44.8 |
| 6 | 10-May | 23:08 | 23 | 1 | 1 | 147 | 13.4 | 134-183 | 30.8 | 9.4 | 21.5-57.5 |
| 7 | 12-May | 12:33 | 24 | 1 | 0 | 153 | 13.9 | 138-185 | 37.1 | 11.2 | 25.6-64.1 |
| 8 | 14-May | 23:50 | 27 | 0 | 0 | 155 | 16.7 | 134-192 | 33.2 | 10.9 | 21.7-61.7 |
| 9 | 16-May | 12:44 | 25 | 0 | 0 | 167 | 16.8 | 135-200 | 45.0 | 13.6 | 27.3-80.0 |
| 10 | 19-May | 01:10 | 26 | 0 | 0 | 166 | 17.2 | 141-204 | 42.1 | 13.2 | 25.5-71.0 |
| 11 | 20-May | 11:58 | 24 | 0 | 1 | 164 | 16.3 | 142-194 | 41.1 | 13.2 | 25.8-66.0 |
| 12 | 22-May | 22:34 | 24 | 0 | 1 | 163 | 11.9 | 142-186 | 40.7 | 8.8 | 29.0-59.5 |
| 13 | 24-May | 10:05 | 26 | 0 | 0 | 164 | 11.7 | 147-191 | 41.2 | 10.2 | 27.7-68.0 |
| 14 | 26-May | 22:13 | 26 | 0 | 0 | 164 | 14.7 | 144-210 | 33.2 | 10.4 | 22.3-69.4 |
| 15 | 28-May | 10:41 | 24 | 0 | 0 | 162 | 14.5 | 146-210 | 34.5 | 12.4 | 22.9-83.3 |
| 16 | 30-May | 22:50 | 26 | 0 | 0 | 165 | 16.0 | 142-212 | 39.1 | 14.1 | 24.8-87.2 |
| Overall |  |  | 399 | 5 | 3 | 156 | 15.8 | 130-212 | 36.2 | 11.2 | 21.5-87.2 |

Table A1.2. Summary of yearling Chinook salmon releases at the Bonneville Dam powerhouse 1 minimum gap runner (MGR) turbine unit during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | 30-Apr | 12:06 | 24 | 1 | 0 | 144 | 6.5 | 130-156 | 29.7 | 3.9 | 22.7-36.7 |
| 2 | 2-May | 23:53 | 24 | 1 | 0 | 140 | 7.0 | 130-154 | 27.2 | 3.6 | 21.6-35.2 |
| 3 | 4-May | 11:04 | 23 | 1 | 0 | 144 | 12.8 | 130-180 | 34.5 | 9.2 | 24.4-65.7 |
| 4 | 6-May | 23:25 | 25 | 1 | 0 | 144 | 6.7 | 133-161 | 29.3 | 5.1 | 23.5-48.0 |
| 5 | 8-May | 11:56 | 26 | 0 | 0 | 148 | 11.5 | 135-194 | 32.0 | 8.0 | 22.7-64.8 |
| 6 | 10-May | 22:54 | 24 | 0 | 0 | 150 | 16.0 | 131-190 | 33.1 | 13.4 | 21.5-66.7 |
| 7 | 12-May | 12:19 | 25 | 0 | 0 | 154 | 14.9 | 140-194 | 37.8 | 12.2 | 27.1-72.1 |
| 8 | 14-May | 23:19 | 24 | 1 | 0 | 153 | 14.8 | 134-188 | 31.0 | 9.3 | 21.9-54.9 |
| 9 | 16-May | 12:20 | 24 | 1 | 0 | 164 | 14.0 | 146-185 | 41.3 | 10.1 | 28.1-59.0 |
| 10 | 19-May | a | a | a | a | a | a | a | a | a | a |
| 11 | 20-May | 11:39 | 25 | 0 | 0 | 152 | 17.1 | 134-188 | 31.4 | 11.1 | 21.7-56.2 |
| 12 | 22-May | 22:06 | 25 | 0 | 0 | 161 | 12.2 | 139-183 | 36.6 | 8.1 | 23.0-52.1 |
| 13 | 24-May | 9:55 | 24 | 0 | 0 | 165 | 12.9 | 138-192 | 41.6 | 10.1 | 24.9-67.4 |
| 14 | 26-May | 21:54 | 24 | 0 | 0 | 155 | 12.2 | 135-182 | 38.0 | 9.3 | 23.6-58.0 |
| 15 | 28-May | 10:28 | 25 | 0 | 0 | 159 | 12.9 | 142-186 | 32.1 | 8.8 | 22.4-54.5 |
| 16 | 30-May | a | a | a | a | a | a | a | a | a | a |
| Overall |  |  | 342 | 6 | 0 | 153 | 14.6 | 130-194 | 34.0 | 9.9 | 21.5-72.1 |

a - Dam operations were different then the specified treatment conditions.

Table A1.3. Summary of yearling Chinook salmon releases at Bonneville Dam in the tailrace directly below the front roll of the powerhouse 1 MGR turbine unit during spring 2004. Dates, times, numbers of tagged fish released ( N ), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | 30-Apr | 12:11 | 22 | 2 | 0 | 145 | 9.0 | 133-165 | 29.8 | 5.9 | 22.3-46.1 |
| 2 | 3-May | 00:08 | 25 | 0 | 0 | 141 | 5.6 | 134-155 | 28.5 | 3.5 | 23.1-35.6 |
| 3 | 4-May | 11:09 | 27 | 0 | 0 | 140 | 8.8 | 121-164 | 32.9 | 6.4 | 24.8-52.9 |
| 4 | 6-May | 23:29 | 25 | 0 | 0 | 147 | 10.1 | 137-175 | 31.0 | 7.1 | 24.3-53.8 |
| 5 | 8-May | 12:08 | 24 | 0 | 0 | 145 | 9.4 | 135-179 | 29.0 | 6.3 | 22.0-52.2 |
| 6 | 10-May | 23:01 | 25 | 0 | 0 | 150 | 15.6 | 134-190 | 31.0 | 11.6 | 21.5-64.9 |
| 7 | 12-May | 13:01 | 25 | 0 | 0 | 159 | 15.3 | 143-198 | 41.2 | 12.5 | 26.2-75.5 |
| 8 | 14-May | 23:30 | 23 | 0 | 0 | 163 | 21.3 | 136-220 | 39.3 | 18.6 | 22.1-104.7 |
| 9 | 16-May | 12:33 | 25 | 0 | 0 | 162 | 18.5 | 143-216 | 39.7 | 16.4 | 24.2-97.9 |
| 10 | 19-May | 00:37 | 25 | 0 | 0 | 156 | 12.7 | 138-176 | 34.3 | 9.3 | 22.4-50.2 |
| 11 | 20-May | 11:48 | 22 | 3 | 0 | 158 | 12.4 | 137-184 | 31.0 | 7.2 | 21.7-47.6 |
| 12 | 22-May | 22:16 | 23 | 1 | 0 | 166 | 13.3 | 144-187 | 43.6 | 9.5 | 29.9-60.7 |
| 13 | 24-May | 09:51 | 28 | 0 | 0 | 174 | 14.2 | 150-206 | 43.5 | 12.0 | 28.0-78.5 |
| 14 | 26-May | 21:58 | 26 | 0 | 0 | 159 | 11.5 | 140-189 | 29.9 | 6.2 | 21.6-47.5 |
| 15 | 28-May | 10:34 | 25 | 0 | 0 | 157 | 9.2 | 138-179 | 31.8 | 5.4 | 22.0-48.6 |
| 16 | 30-May | 22:41 | 26 | 0 | 0 | 173 | 16.7 | 150-206 | 48.0 | 14.7 | 33.7-83.5 |
| Overall |  |  | 396 | 6 | 0 | 156 | 16.6 | 121-220 | 35.4 | 11.9 | 21.5-104.7 |

Table A1.4. Summary of yearling Chinook salmon releases at The Dalles Dam ice and trash sluiceway (Sluiceway), and the tailrace 550 m downstream of the spillway beneath the I-197 bridge (Tailrace) during spring 2004. Dates, times, numbers of tagged fish released $(\mathrm{N}), 24 \mathrm{~h}$ post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Location | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | Tailrace | 28-Apr | 01:25 | 41 | 6 | 0 | 148 | 10.5 | 130-188 | 34.8 | 9.2 | 23.8-76.4 |
| 2 | Tailrace | 28-Apr | 12:23 | 45 | 0 | 0 | 152 | 7.4 | 139-169 | 33.2 | 5.2 | 25.9-44.7 |
| 3 | Sluiceway | 28-Apr | 12:29 | 32 | 0 | 0 | 156 | 8.8 | 142-180 | 36.3 | 6.7 | 26.0-52.1 |
| 4 | Tailrace | 29-Apr | 01:49 | 40 | 6 | 1 | 155 | 12.1 | 131-185 | 36.7 | 9.0 | 24.5-58.2 |
| 5 | Tailrace | 29-Apr | 12:30 | 44 | 0 | 0 | 156 | 13.0 | 132-200 | 38.2 | 9.3 | 23.4-62.2 |
| 6 | Sluiceway | 29-Apr | 19:09 | 44 | 4 | 0 | 154 | 8.7 | 142-175 | 36.1 | 6.5 | 27.9-53.8 |
| 7 | Tailrace | 30-Apr | 01:33 | 38 | 6 | 2 | 152 | 9.9 | 135-180 | 35.1 | 8.0 | 23.0-62.6 |
| 8 | Tailrace | 30-Apr | 13:06 | 47 | 1 | 0 | 153 | 8.4 | 133-178 | 35.1 | 5.6 | 22.5-51.6 |
| 9 | Tailrace | 1-May | 00:10 | 48 | 0 | 0 | 150 | 7.0 | 136-167 | 32.9 | 4.7 | 24.2-50.5 |
| 10 | Sluiceway | 1-May | 01:59 | 43 | 0 | 0 | 154 | 10.9 | 136-179 | 35.7 | 7.9 | 24.2-56.2 |
| 11 | Sluiceway | 1-May | 07:05 | 42 | 3 | 1 | 151 | 10.0 | 130-173 | 33.8 | 6.6 | 23.5-49.7 |
| 12 | Tailrace | 1-May | 13:15 | 45 | 1 | 0 | 153 | 9.8 | 137-180 | 35.5 | 7.4 | 24.2-55.6 |
| 13 | Tailrace | 2-May | 00:30 | 47 | 0 | 0 | 149 | 7.8 | 136-170 | 33.0 | 5.6 | 24.3-48.1 |
| 14 | Tailrace | 2-May | 13:38 | 47 | 0 | 0 | 148 | 12.4 | 130-178 | 35.0 | 7.8 | 25.4-57.5 |
| 15 | Sluiceway | 2-May | 13:00 | 43 | 3 | 1 | 147 | 12.1 | 131-186 | 32.7 | 6.5 | 22.9-55.9 |
| 16 | Tailrace | 3-May | 00:48 | 46 | 2 | 0 | 153 | 10.5 | 135-186 | 34.9 | 7.7 | 25.1-60.0 |
| 17 | Sluiceway | 3-May | 07:00 | 47 | 1 | 0 | 147 | 10.5 | 125-180 | 33.7 | 7.9 | 21.7-58.0 |
| 18 | Tailrace | 3-May | 12:30 | 45 | 1 | 0 | 150 | 14.2 | 127-192 | 36.2 | 10.8 | 21.5-69.5 |
| 19 | Tailrace | 4-May | 00:37 | 44 | 2 | 0 | 149 | 9.3 | 137-190 | 33.0 | 6.4 | 22.5-56.0 |
| 20 | Tailrace | 4-May | 12:46 | 47 | 1 | 0 | 150 | 9.8 | 136-176 | 32.5 | 6.5 | 23.9-54.9 |
| 21 | Tailrace | 5-May | 00:33 | 46 | 0 | 1 | 147 | 10.2 | 130-176 | 30.7 | 8.0 | 21.8-63.3 |
| 22 | Sluiceway | 5-May | 01:00 | 48 | 0 | 0 | 149 | 11.2 | 134-179 | 33.1 | 8.7 | 22.4-55.5 |
| 23 | Tailrace | 5-May | 12:28 | 47 | 1 | 0 | 149 | 10.8 | 131-191 | 31.6 | 7.1 | 21.7-60.8 |
| 24 | Sluiceway | 5-May | 19:09 | 46 | 1 | 0 | 147 | 9.6 | 136-174 | 30.6 | 6.9 | 21.6-52.5 |

Table A1.4 (continued). Summary of yearling Chinook salmon releases at The Dalles Dam ice and trash sluiceway (Sluiceway), and the tailrace 550 m downstream of the spillway beneath the I-197 bridge (Tailrace) during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Location | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 25 | Tailrace | 6-May | 00:33 | 47 | 0 | 0 | 154 | 13.9 | 133-184 | 35.4 | 10.3 | 21.5-59.5 |
| 26 | Sluiceway | 6-May | 07:00 | 47 | 0 | 0 | 151 | 13.6 | 131-192 | 33.5 | 9.9 | 21.7-63.1 |
| 27 | Tailrace | 6-May | 12:35 | 46 | 1 | 0 | 152 | 12.2 | 131-183 | 34.0 | 8.5 | 21.7-58.7 |
| 28 | Tailrace | 7-May | 00:19 | 41 | 6 | 0 | 149 | 11.7 | 132-177 | 35.5 | 8.0 | 21.8-54.1 |
| 29 | Tailrace | 7-May | 12:04 | 45 | 1 | 0 | 152 | 14.2 | 130-180 | 34.6 | 10.2 | 22.5-55.8 |
| 30 | Tailrace | 8-May | 00:03 | 46 | 2 | 0 | 154 | 10.1 | 139-174 | 38.9 | 6.7 | 27.4-54.5 |
| 31 | Sluiceway | 8-May | 01:00 | 46 | 1 | 0 | 158 | 13.6 | 136-189 | 43.1 | 12.1 | 28.4-69.5 |
| 32 | Tailrace | 8-May | 12:33 | 43 | 2 | 0 | 156 | 10.9 | 140-184 | 36.0 | 7.2 | 25.7-55.1 |
| 33 | Sluiceway | 8-May | 19:00 | 43 | 0 | 5 | 152 | 14.9 | 122-185 | 35.6 | 11.1 | 22.0-60.1 |
| 34 | Tailrace | 9-May | 00:06 | 47 | 0 | 0 | 147 | 12.5 | 132-181 | 31.8 | 10.5 | 21.5-77.5 |
| 35 | Tailrace | 9-May | 13:01 | 47 | 0 | 0 | 147 | 12.5 | 132-205 | 30.1 | 9.8 | 22.0-75.1 |
| 36 | Sluiceway | 9-May | 13:00 | 48 | 0 | 0 | 149 | 9.8 | 135-178 | 31.8 | 7.2 | 23.7-58.7 |
| 37 | Tailrace | 10-May | 01:30 | 48 | 0 | 0 | 157 | 13.3 | 133-192 | 37.5 | 9.8 | 23.6-70.2 |
| 38 | Tailrace | 10-May | 12:47 | 47 | 0 | 0 | 154 | 12.1 | 134-191 | 35.9 | 9.3 | 23.5-66.7 |
| 39 | Sluiceway | 10-May | 19:00 | 47 | 1 | 0 | 152 | 14.8 | 133-205 | 35.0 | 13.6 | 21.9-98.6 |
| 40 | Tailrace | 11-May | 00:43 | 48 | 0 | 0 | 149 | 12.4 | 133-181 | 32.2 | 8.5 | 23.3-58.9 |
| 41 | Tailrace | 11-May | 12:05 | 48 | 0 | 0 | 147 | 11.9 | 131-188 | 30.3 | 8.6 | 21.6-62.0 |
| 42 | Sluiceway | 11-May | 13:00 | 47 | 1 | 0 | 150 | 18.3 | 130-210 | 33.1 | 14.1 | 22.2-79.8 |
| 43 | Tailrace | 12-May | 00:26 | 46 | 1 | 0 | 159 | 18.1 | 134-200 | 39.4 | 13.5 | 22.8-74.7 |
| 44 | Sluiceway | 12-May | 07:08 | 48 | 0 | 0 | 156 | 16.6 | 129-210 | 37.6 | 13.6 | 21.8-95.6 |
| 45 | Tailrace | 12-May | 12:06 | 45 | 2 | 0 | 159 | 14.7 | 136-189 | 39.2 | 10.7 | 23.9-64.5 |
| 46 | Tailrace | 13-May | 00:31 | 48 | 0 | 0 | 154 | 16.9 | 133-190 | 36.2 | 12.4 | 22.6-65.4 |
| 47 | Tailrace | 13-May | 12:10 | 48 | 0 | 0 | 163 | 17.3 | 131-197 | 42.3 | 13.8 | 22.3-71.2 |
| 48 | Tailrace | 14-May | 00:55 | 48 | 0 | 0 | 155 | 15.2 | 135-199 | 36.1 | 12.2 | 21.7-76.6 |

Table A1.4 (continued). Summary of yearling Chinook salmon releases at The Dalles Dam ice and trash sluiceway (Sluiceway), and the tailrace 550 m downstream of the spillway beneath the I-197 bridge (Tailrace) during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Location | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 49 | Sluiceway | 14-May | 01:00 | 44 | 0 | 1 | 160 | 17.8 | 134-203 | 40.0 | 13.6 | 23.4-77.4 |
| 50 | Tailrace | 14-May | 13:06 | 48 | 0 | 0 | 159 | 17.6 | 136-198 | 38.4 | 12.4 | 23.5-69.3 |
| 51 | Tailrace | 15-May | 01:04 | 47 | 0 | 0 | 160 | 18.8 | 136-218 | 40.7 | 15.9 | 22.5-94.1 |
| 52 | Sluiceway | 15-May | 01:12 | 48 | 0 | 0 | 166 | 19.8 | 132-207 | 45.1 | 16.4 | 21.9-79.0 |
| 53 | Sluiceway | 15-May | 07:00 | 47 | 0 | 1 | 161 | 21.0 | 132-220 | 41.8 | 18.1 | 21.6-97.3 |
| 54 | Tailrace | 15-May | 12:29 | 47 | 1 | 0 | 161 | 18.0 | 135-208 | 40.8 | 14.8 | 21.5-83.0 |
| 55 | Tailrace | 16-May | 00:21 | 47 | 1 | 0 | 160 | 19.3 | 136-205 | 39.7 | 14.7 | 21.8-72.2 |
| 56 | Tailrace | 16-May | 12:40 | 47 | 0 | 0 | 159 | 18.6 | 137-200 | 38.7 | 14.6 | 21.8-76.9 |
| 57 | Sluiceway | 16-May | 19:00 | 46 | 1 | 0 | 160 | 17.9 | 130-190 | 39.4 | 13.5 | 22.0-69.1 |
| 58 | Tailrace | 17-May | 00:39 | 45 | 1 | 2 | 161 | 18.2 | 130-200 | 41.0 | 13.7 | 21.6-72.9 |
| 59 | Tailrace | 17-May | 12:28 | 48 | 0 | 0 | 161 | 16.0 | 134-189 | 40.0 | 13.1 | 21.9-68.8 |
| 60 | Sluiceway | 17-May | 13:00 | 48 | 0 | 0 | 164 | 16.9 | 136-205 | 42.3 | 14.7 | 22.8-91.1 |
| 61 | Tailrace | 18-May | 00:20 | 46 | 1 | 1 | 160 | 17.9 | 138-195 | 40.0 | 14.3 | 24.1-76.0 |
| 62 | Tailrace | 18-May | 12:47 | 43 | 4 | 0 | 165 | 17.8 | 135-198 | 42.7 | 13.9 | 22.2-73.5 |
| 63 | Tailrace | 19-May | 00:10 | 48 | 0 | 0 | 155 | 15.6 | 135-194 | 36.0 | 11.4 | 22.2-69.6 |
| 64 | Sluiceway | 19-May | 01:00 | 47 | 1 | 0 | 160 | 15.7 | 132-194 | 38.5 | 11.0 | 22.5-67.3 |
| 65 | Tailrace | 19-May | 12:19 | 47 | 1 | 0 | 156 | 14.6 | 131-191 | 35.6 | 9.5 | 22.0-60.3 |
| 66 | Sluiceway | 19-May | 13:01 | 47 | 1 | 0 | 158 | 14.8 | 134-195 | 37.5 | 11.3 | 22.1-64.5 |
| 67 | Tailrace | 20-May | 00:04 | 53 | 0 | 0 | 160 | 15.9 | 130-197 | 39.9 | 11.7 | 21.6-75.9 |
| 68 | Sluiceway | 20-May | 07:03 | 54 | 0 | 0 | 165 | 14.6 | 137-200 | 42.1 | 11.2 | 23.3-73.9 |
| 69 | Tailrace | 20-May | 12:40 | 51 | 1 | 0 | 163 | 15.3 | 131-205 | 41.5 | 11.4 | 25.2-70.3 |
| 70 | Tailrace | 21-May | 00:59 | 52 | 0 | 1 | 160 | 15.5 | 133-200 | 39.7 | 11.1 | 24.3-69.9 |
| 71 | Tailrace | 21-May | 12:03 | 53 | 0 | 0 | 158 | 15.3 | 130-185 | 37.1 | 10.5 | 21.6-58.6 |
| 72 | Sluiceway | 21-May | 19:00 | 54 | 1 | 0 | 156 | 16.1 | 130-200 | 36.7 | 11.3 | 22.4-71.0 |

Table A1.4 (continued). Summary of yearling Chinook salmon releases at The Dalles Dam ice and trash sluiceway (Sluiceway), and the tailrace 550 m downstream of the spillway beneath the I-197 bridge (Tailrace) during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Location | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 73 | Tailrace | 22-May | 00:55 | 53 | 0 | 0 | 160 | 15.2 | 131-200 | 39.3 | 11.7 | 22.9-81.1 |
| 74 | Tailrace | 22-May | 12:26 | 52 | 0 | 0 | 159 | 13.4 | 134-190 | 37.4 | 9.8 | 22.3-67.6 |
| 75 | Sluiceway | 22-May | 13:00 | 55 | 0 | 0 | 161 | 15.5 | 133-202 | 39.3 | 12.0 | 22.9-78.4 |
| 76 | Tailrace | 23-May | 00:04 | 51 | 2 | 0 | 162 | 14.7 | 139-216 | 40.2 | 11.8 | 24.6-91.0 |
| 77 | Tailrace | 23-May | 12:20 | 53 | 0 | 0 | 161 | 13.6 | 133-191 | 40.0 | 9.9 | 22.3-65.8 |
| 78 | Sluiceway | 23-May | 19:00 | 53 | 0 | 0 | 160 | 13.4 | 132-185 | 38.0 | 9.4 | 22.5-59.5 |
| 79 | Tailrace | 24-May | 00:36 | 52 | 0 | 0 | 162 | 13.8 | 141-205 | 40.0 | 11.6 | 26.4-83.6 |
| 80 | Tailrace | 24-May | 12:40 | 48 | 0 | 0 | 160 | 15.6 | 135-191 | 39.0 | 11.5 | 21.7-62.5 |
| 81 | Tailrace | 25-May | 00:43 | 51 | 2 | 0 | 159 | 16.6 | 136-199 | 37.7 | 12.6 | 23.9-70.8 |
| 82 | Sluiceway | 25-May | 01:01 | 52 | 1 | 0 | 166 | 17.5 | 135-214 | 43.0 | 15.6 | 21.5-92.0 |
| 83 | Sluiceway | 25-May | 07:01 | 55 | 0 | 0 | 164 | 15.3 | 134-200 | 42.7 | 12.5 | 23.4-81.2 |
| 84 | Tailrace | 25-May | 15:47 | 51 | 1 | 0 | 162 | 18.5 | 135-209 | 42.5 | 16.0 | 24.3-94.1 |
| 85 | Tailrace | 26-May | 00:18 | 51 | 1 | 1 | 162 | 18.3 | 136-215 | 41.0 | 16.7 | 22.0-100.6 |
| 86 | Tailrace | 26-May | 12:40 | 52 | 0 | 1 | 161 | 16.8 | 137-205 | 41.1 | 14.9 | 24.7-91.7 |
| 87 | Tailrace | 27-May | 00:57 | 53 | 0 | 0 | 163 | 17.8 | 135-214 | 41.8 | 15.7 | 21.9-96.7 |
| 88 | Sluiceway | 27-May | 01:02 | 52 | 0 | 2 | 165 | 20.0 | 132-215 | 44.1 | 18.4 | 22.3-99.0 |
| 89 | Tailrace | 27-May | 12:04 | 53 | 0 | 0 | 161 | 16.4 | 135-202 | 40.9 | 14.2 | 22.1-87.9 |
| 90 | Sluiceway | 27-May | 13:03 | 53 | 1 | 0 | 167 | 19.1 | 134-214 | 46.8 | 17.9 | 23.7-96.8 |
| 91 | Tailrace | 28-May | 00:47 | 44 | 0 | 0 | 168 | 20.5 | 134-210 | 48.8 | 19.5 | 26.4-99.8 |
| 92 | Sluiceway | 28-May | 07:00 | 45 | 1 | 0 | 161 | 18.2 | 138-210 | 40.4 | 16.4 | 24.6-87.6 |
| 93 | Tailrace | 28-May | 12:08 | 47 | 0 | 0 | 169 | 21.8 | 133-224 | 48.5 | 20.9 | 21.8-112.6 |
| 94 | Tailrace | 29-May | 01:10 | 28 | 0 | 0 | 166 | 21.1 | 141-202 | 46.5 | 18.9 | 26.0-84.7 |
| 95 | Tailrace | 29-May | 12:10 | 29 | 2 | 0 | 169 | 19.3 | 145-212 | 47.1 | 19.1 | 26.5-88.0 |
| 96 | Sluiceway | 29-May | 19:00 | 20 | 0 | 0 | 166 | 20.5 | 137-202 | 46.3 | 18.2 | 25.6-85.6 |
| Overall |  |  |  | 4486 | 82 | 21 | 157 | 16.0 | 122-224 | 38.0 | 12.6 | 19.7-112.6 |

Table A1.5. Summary of yearling Chinook salmon releases at Bonneville Dam tailrace during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | 29-Apr | 09:54 | 20 | 0 | 0 | 148 | 14.6 | 135-205 | 31.8 | 10.9 | 23.0-76.0 |
| 2 | 29-Apr | 21:43 | 18 | 0 | 0 | 145 | 6.9 | 135-158 | 30.9 | 5.0 | 24.0-41.7 |
| 3 | 30-Apr | 10:19 | 17 | 1 | 1 | 147 | 12.9 | 135-180 | 31.6 | 10.1 | 21.5-61.8 |
| 4 | 30-Apr | 21:56 | 20 | 1 | 0 | 148 | 7.5 | 135-163 | 32.2 | 5.1 | 24.5-40.5 |
| 5 | 1-May | 10:08 | 22 | 0 | 0 | 148 | 8.7 | 135-172 | 31.8 | 6.8 | 23.3-54.1 |
| 6 | 1-May | 21:45 | 17 | 2 | 0 | 149 | 8.6 | 136-163 | 32.3 | 5.2 | 24.6-41.1 |
| 7 | 2-May | 09:42 | 21 | 0 | 0 | 148 | 10.2 | 132-170 | 32.2 | 6.8 | 23.2-47.0 |
| 8 | 2-May | 22:07 | 18 | 2 | 0 | 142 | 6.7 | 131-152 | 29.3 | 4.1 | 23.7-38.9 |
| 9 | 3-May | 10:00 | 20 | 0 | 0 | 145 | 9.0 | 134-166 | 30.6 | 6.4 | 24.7-46.0 |
| 10 | 3-May | 22:00 | 19 | 3 | 0 | 147 | 8.6 | 136-167 | 31.2 | 5.7 | 23.9-45.8 |
| 11 | 4-May | 09:53 | 18 | 2 | 0 | 151 | 10.9 | 139-179 | 33.9 | 7.8 | 26.2-56.7 |
| 12 | 4-May | 22:01 | 20 | 0 | 0 | 144 | 9.1 | 134-160 | 29.1 | 6.3 | 23.0-42.3 |
| 13 | 5-May | 09:46 | 20 | 0 | 0 | 142 | 6.4 | 130-154 | 27.6 | 3.4 | 21.6-33.5 |
| 14 | 5-May | 21:55 | 22 | 0 | 1 | 148 | 13.0 | 135-192 | 31.3 | 9.4 | 22.6-65.1 |
| 15 | 6-May | 10:11 | 22 | 0 | 0 | 145 | 12.3 | 133-181 | 30.3 | 9.3 | 22.0-62.3 |
| 16 | 6-May | 21:39 | 20 | 0 | 0 | 151 | 9.2 | 138-171 | 33.6 | 7.5 | 24.7-51.1 |
| 17 | 7-May | 10:25 | 20 | 0 | 0 | 151 | 8.0 | 137-165 | 33.4 | 5.2 | 27.1-46.7 |
| 18 | 7-May | 21:51 | 21 | 0 | 0 | 145 | 6.4 | 136-157 | 30.5 | 5.6 | 23.9-43.6 |
| 19 | 8-May | 10:07 | 19 | 0 | 1 | 145 | 8.9 | 131-171 | 29.5 | 6.7 | 22.1-54.1 |
| 20 | 8-May | 21:42 | 20 | 0 | 0 | 151 | 12.8 | 134-175 | 34.3 | 9.4 | 23.3-58.4 |
| 21 | 9-May | 10:12 | 20 | 1 | 0 | 152 | 14.3 | 136-199 | 34.5 | 10.7 | 24.1-66.1 |
| 22 | 9-May | 21:43 | 19 | 0 | 0 | 145 | 10.2 | 130-165 | 28.9 | 7.1 | 21.8-51.0 |
| 23 | 10-May | 09:53 | 20 | 0 | 0 | 149 | 11.7 | 135-180 | 30.6 | 8.9 | 22.1-53.1 |
| 24 | 10-May | 21:59 | 20 | 0 | 0 | 148 | 13.2 | 132-187 | 31.1 | 11.1 | 21.7-70.2 |

Table A1.5 (continued). Summary of yearling Chinook salmon releases at Bonneville Dam tailrace during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 25 | 11-May | 10:24 | 19 | 1 | 1 | 148 | 9.9 | 135-169 | 30.5 | 7.7 | 21.5-48.3 |
| 26 | 11-May | 22:00 | 18 | 1 | 0 | 156 | 10.4 | 139-181 | 36.5 | 7.6 | 26.0-57.5 |
| 27 | 12-May | 09:23 | 22 | 0 | 0 | 159 | 14.7 | 139-190 | 40.2 | 14.3 | 23.2-75.4 |
| 28 | 12-May | 22:07 | 16 | 2 | 0 | 168 | 22.0 | 141-208 | 46.2 | 18.6 | 26.6-86.8 |
| 29 | 13-May | 10:01 | 20 | 0 | 0 | 161 | 17.7 | 142-199 | 35.7 | 11.3 | 23.7-63.6 |
| 30 | 13-May | 22:00 | 20 | 1 | 0 | 163 | 12.7 | 148-191 | 35.9 | 8.9 | 25.2-56.8 |
| 31 | 14-May | 10:01 | 19 | 1 | 0 | 167 | 13.8 | 147-194 | 38.5 | 9.9 | 24.7-60.7 |
| 32 | 14-May | 21:54 | 22 | 0 | 0 | 157 | 12.8 | 140-180 | 38.3 | 10.2 | 24.1-57.7 |
| 33 | 15-May | 10:25 | 19 | 0 | 1 | 153 | 11.9 | 140-178 | 33.8 | 7.8 | 24.9-52.3 |
| 34 | 15-May | 21:52 | 19 | 0 | 0 | 157 | 11.8 | 141-190 | 34.9 | 6.8 | 26.2-50.1 |
| 35 | 16-May | 10:16 | 17 | 3 | 0 | 160 | 11.9 | 141-185 | 35.6 | 7.2 | 24.4-51.6 |
| 36 | 16-May | 22:08 | 19 | 0 | 1 | 168 | 15.0 | 145-199 | 44.1 | 12.5 | 29.1-71.9 |
| 37 | 17-May | 09:55 | 21 | 1 | 0 | 174 | 17.0 | 146-230 | 50.3 | 21.2 | 28.4-132.0 |
| 38 | 17-May | 22:15 | 20 | 0 | 0 | 168 | 11.0 | 152-188 | 44.6 | 7.6 | 31.0-61.4 |
| 39 | 18-May | 09:52 | 19 | 1 | 0 | 169 | 14.0 | 150-198 | 45.6 | 11.5 | 32.4-71.1 |
| 40 | 18-May | 22:33 | 22 | 0 | 0 | 157 | 14.9 | 139-189 | 35.7 | 10.8 | 24.2-58.7 |
| 41 | 19-May | 10:05 | 20 | 0 | 0 | 167 | 14.4 | 139-190 | 43.4 | 12.1 | 22.0-62.4 |
| 42 | 19-May | 22:35 | 20 | 0 | 0 | 151 | 10.5 | 133-168 | 33.7 | 7.8 | 24.7-47.6 |
| 43 | 20-May | 10:11 | 21 | 0 | 0 | 155 | 12.8 | 135-183 | 33.1 | 8.6 | 22.3-53.5 |
| 44 | 20-May | 21:51 | 20 | 0 | 0 | 166 | 10.6 | 148-184 | 44.3 | 9.4 | 28.3-61.9 |
| 45 | 21-May | 10:01 | 19 | 1 | 0 | 171 | 14.0 | 152-200 | 45.9 | 11.5 | 31.6-73.7 |
| 46 | 21-May | 21:49 | 17 | 4 | 0 | 174 | 11.6 | 151-195 | 47.5 | 9.4 | 33.3-67.1 |
| 47 | 22-May | 10:55 | 20 | 0 | 0 | 172 | 11.5 | 148-188 | 47.8 | 9.0 | 31.2-65.1 |
| 48 | 22-May | 23:49 | 18 | 2 | 0 | 159 | 8.2 | 144-172 | 38.1 | 6.3 | 28.9-49.6 |
| 49 | 23-May | 10:22 | 20 | 0 | 0 | 167 | 15.7 | 146-212 | 42.2 | 12.8 | 28.2-81.6 |
| 50 | 23-May | 22:14 | 24 | 0 | 0 | 166 | 12.2 | 144-193 | 43.1 | 11.1 | 29.5-75.3 |
| 51 | 24-May | 11:21 | 20 | 0 | 0 | 172 | 10.7 | 150-191 | 46.8 | 9.5 | 32.1-69.7 |

Table A1.5 (continued). Summary of yearling Chinook salmon releases at Bonneville Dam tailrace during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 52 | 24-May | 22:15 | 22 | 0 | 0 | 167 | 12.4 | 149-198 | 43.7 | 10.6 | 31.5-73.8 |
| 53 | 25-May | 09:26 | 20 | 0 | 0 | 173 | 13.4 | 151-197 | 46.5 | 11.3 | 29.0-67.9 |
| 54 | 25-May | 22:06 | 19 | 1 | 0 | 160 | 12.5 | 142-185 | 37.5 | 8.9 | 26.0-56.6 |
| 55 | 26-May | 10:06 | 20 | 1 | 0 | 165 | 13.2 | 144-200 | 39.6 | 10.1 | 28.6-69.1 |
| 56 | 26-May | 23:50 | 19 | 1 | 0 | 155 | 18.2 | 135-210 | 38.7 | 17.0 | 24.0-98.4 |
| 57 | 27-May | 10:35 | 20 | 0 | 0 | 161 | 11.0 | 147-185 | 38.7 | 8.5 | 29.4-66.7 |
| 58 | 27-May | 21:54 | 20 | 1 | 0 | 156 | 13.2 | 141-184 | 36.0 | 11.7 | 24.1-66.2 |
| 59 | 28-May | 12:05 | 20 | 0 | 0 | 161 | 11.8 | 139-180 | 38.4 | 7.0 | 27.2-54.3 |
| 60 | 28-May | 22:08 | 22 | 0 | 0 | 163 | 17.8 | 145-225 | 38.7 | 14.4 | 22.9-83.0 |
| 61 | 29-May | 10:50 | 21 | 0 | 0 | 166 | 16.0 | 143-200 | 43.4 | 13.6 | 27.8-77.3 |
| 62 | 29-May | 22:27 | 22 | 0 | 1 | 164 | 12.9 | 146-200 | 38.9 | 13.2 | 26.3-78.8 |
| 63 | 30-May | 10:59 | 24 | 0 | 0 | 176 | 17.4 | 155-215 | 46.0 | 16.6 | 27.2-83.7 |
| 64 | 31-May | 00:01 | 23 | 0 | 0 | 174 | 14.0 | 150-202 | 50.4 | 13.0 | 32.8-80.1 |
| Overall |  |  | 1276 | 34 | 7 | 158 | 15.8 | 130-230 | 37.2 | 9.6 | 21.5-132.0 |

Table A1.6. Summary of steelhead trout releases at Bonneville Dam ice and trash sluiceway during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | 5-May | 23:50 | 14 | 3 | 0 | 225 | 18.8 | 198-260 | 93.2 | 23.2 | 63.9-153.2 |
| 2 | 6-May | 23:48 | 18 | 1 | 1 | 224 | 24.1 | 194-265 | 104.5 | 34.6 | 63.0-171.0 |
| 3 | 8-May | 12:18 | 23 | 1 | 0 | 227 | 23.2 | 200-275 | 101.4 | 34.6 | 64.4-179.3 |
| 4 | 12-May | 12:38 | 24 | 3 | 0 | 215 | 26.5 | 134-268 | 93.4 | 26.6 | 49.9-155.8 |
| 5 | 13-May | 12:10 | 28 | 0 | 0 | 235 | 18.5 | 191-265 | 90.2 | 23.5 | 42.3-139.3 |
| 6 | 14-May | 23:42 | 13 | 0 | 1 | 223 | 15.3 | 208-260 | 91.5 | 20.6 | 74.0-150.5 |
| 7 | 16-May | 12:39 | 25 | 0 | 0 | 235 | 21.1 | 193-275 | 110.3 | 31.9 | 57.6-172.2 |
| 8 | 19-May | 01:01 | 23 | 1 | 1 | 224 | 18.3 | 196-259 | 90.8 | 20.3 | 60.4-138.1 |
| 9 | 20-May | 12:05 | 23 | 2 | 0 | 211 | 21.5 | 161-252 | 73.1 | 21.5 | 30.2-127.2 |
| 10 | 22-May | 22:30 | 15 | 0 | 0 | 237 | 24.4 | 198-289 | 114.4 | 40.5 | 63.2-221.5 |
| 11 | 24-May | 10:12 | 25 | 0 | 0 | 231 | 27.9 | 177-295 | 90.4 | 33.6 | 38.4-182.8 |
| 12 | 26-May | 22:07 | 23 | 2 | 0 | 239 | 23.4 | 179-273 | 95.3 | 30.2 | 36.5-147.6 |
| 13 | 28-May | 10:45 | 22 | 3 | 0 | 231 | 17.8 | 187-260 | 96.1 | 28.4 | 51.6-165.0 |
| 14 | 30-May | 22:53 | 11 | 0 | 0 | 233 | 27.9 | 190-274 | 104.2 | 39.4 | 45.3-175.7 |
| 15 | 1-Jun | 10:20 | 18 | 1 | 0 | 229 | 24.4 | 172-282 | 86.3 | 32.6 | 31.2-180.3 |
| 16 | 2-Jun | 22:05 | 16 | 0 | 0 | 242 | 34.7 | 183-293 | 131.3 | 52.4 | 45.8-221.5 |
| Overall |  |  | 321 | 17 | 3 | 229 | 24.1 | 134-295 | 96.8 | 32.8 | 30.2-221.5 |

Table A1.7. Summary of steelhead trout releases at the Bonneville Dam powerhouse 1 minimum gap runner (MGR) turbine unit during spring 2004. Dates, times, numbers of tagged fish released ( N ), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | 5-May | 23:35 | 16 | 1 | 0 | 227 | 23.7 | 193-270 | 102.1 | 34.0 | 65.6-174.2 |
| 2 | 6-May | 23:25 | 18 | 1 | 1 | 217 | 11.0 | 199-241 | 93.7 | 20.3 | 66.8-155.9 |
| 3 | 8-May | 12:01 | 23 | 1 | 1 | 235 | 17.6 | 201-270 | 111.6 | 29.0 | 68.0-174.4 |
| 4 | 12-May | 12:22 | 26 | 0 | 0 | 231 | 21.0 | 188-270 | 104.4 | 28.2 | 53.2-175.5 |
| 5 | 13-May | 11:58 | 23 | 1 | 0 | 224 | 15.9 | 201-273 | 79.1 | 16.9 | 54.2-130.2 |
| 6 | 14-May | 23:24 | 16 | 0 | 0 | 242 | 25.4 | 200-305 | 130.5 | 53.2 | 59.5-276.1 |
| 7 | 16-May | 12:25 | 22 | 2 | 0 | 235 | 20.2 | 210-278 | 107.2 | 30.5 | 74.4-175.1 |
| 8 | 19-May | a | a | a | a | a | a | a | a | a | a |
| 9 | 20-May | 11:35 | 25 | 0 | 0 | 229 | 27.1 | 177-280 | 104.3 | 39.1 | 44.5-186.0 |
| 10 | 22-May | 22:01 | 17 | 0 | 0 | 226 | 22.5 | 186-276 | 100.7 | 32.8 | 51.9-178.9 |
| 11 | 24-May | 09:52 | 25 | 0 | 0 | 228 | 23.1 | 193-267 | 99.7 | 30.5 | 56.4-161.1 |
| 12 | 26-May | 21:47 | 24 | 0 | 1 | 232 | 24.5 | 180-275 | 107.6 | 33.2 | 46.8-177.2 |
| 13 | 28-May | 10:23 | 22 | 1 | 1 | 231 | 21.0 | 184-277 | 90.8 | 24.0 | 45.6-141.7 |
| 14 | 30-May | a | a | a | a | a | a | a | a | a | a |
| 15 | 1-Jun | 10:06 | 20 | 0 | 0 | 231 | 28.2 | 163-278 | 94.4 | 32.4 | 29.1-157.5 |
| 16 | 2-Jun | 21:54 | 15 | 0 | 0 | 233 | 18.2 | 205-265 | 108.7 | 30.9 | 66.6-163.2 |
| Overall |  |  | 292 | 7 | 4 | 230 | 22.1 | 163-305 | 102.0 | 32.8 | 29.1-276.1 |

a - Dam operations were different then the specified treatment conditions.

Table A1.8. Summary of steelhead trout releases at Bonneville Dam in the tailrace directly below the front roll of the powerhouse 1 MGR turbine unit during spring 2004. Dates, times, numbers of tagged fish released ( N ), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | 6-May | 00:14 | 17 | 0 | 0 | 227 | 20.8 | 179-260 | 100.4 | 27.7 | 50.3-143.8 |
| 2 | 6-May | 23:40 | 18 | 2 | 0 | 221 | 19.1 | 193-259 | 92.4 | 24.0 | 61.4-149.9 |
| 3 | 8-May | 12:03 | 20 | 0 | 0 | 220 | 14.6 | 188-241 | 91.4 | 19.8 | 57.5-126.7 |
| 4 | 12-May | 13:09 | 25 | 1 | 0 | 238 | 17.0 | 209-290 | 118.2 | 29.8 | 86.9-226.2 |
| 5 | 13-May | 12:07 | 25 | 1 | 0 | 234 | 14.5 | 199-268 | 91.4 | 17.6 | 69.1-136.5 |
| 6 | 14-May | 23:34 | 13 | 1 | 1 | 239 | 20.3 | 213-278 | 112.6 | 31.4 | 75.0-174.1 |
| 7 | 16-May | 12:41 | 23 | 2 | 0 | 229 | 23.7 | 190-283 | 99.4 | 33.4 | 51.4-173.1 |
| 8 | 19-May | 00:29 | 24 | 1 | 0 | 221 | 14.5 | 198-245 | 80.4 | 14.0 | 58.6-117.0 |
| 9 | 20-May | 11:56 | 20 | 5 | 0 | 226 | 14.3 | 202-251 | 88.0 | 17.1 | 62.3-123.8 |
| 10 | 22-May | 22:09 | 17 | 0 | 0 | 227 | 21.6 | 186-261 | 95.5 | 27.4 | 49.0-140.0 |
| 11 | 24-May | 10:01 | 24 | 1 | 0 | 237 | 34.9 | 118-280 | 97.3 | 33.2 | 39.7-168.4 |
| 12 | 26-May | 22:09 | 25 | 0 | 0 | 229 | 28.8 | 167-304 | 111.6 | 42.5 | 49.0-243.8 |
| 13 | 28-May | 10:39 | 25 | 0 | 0 | 238 | 25.9 | 182-290 | 97.8 | 31.0 | 43.0-154.5 |
| 14 | 30-May | 22:31 | 14 | 0 | 0 | 218 | 33.1 | 150-258 | 96.0 | 34.9 | 39.6-149.2 |
| 15 | 1-Jun | 10:10 | 20 | 0 | 0 | 226 | 31.8 | 159-285 | 86.0 | 39.6 | 25.1-169.5 |
| 16 | 2-Jun | 22:15 | 15 | 0 | 0 | 250 | 20.6 | 208-300 | 134.9 | 42.8 | 63.4-251.9 |
| Overall |  |  | 325 | 14 | 1 | 230 | 24.0 | 118-304 | 98.9 | 32.1 | 25.1-251.9 |

Table A1.9. Summary of steelhead trout releases at The Dalles Dam tailrace during spring 2004. Dates, times, numbers of tagged fish released ( N ), 24 h post-tagging tag loss and mortality, and means, standard deviations ( SD ) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | 28-Apr | 00:31 | 21 | 1 | 0 | 216 | 17.5 | 178-253 | 86.0 | 20.6 | 48.4-124.9 |
| 2 | 28-Apr | 12:23 | 17 | 1 | 0 | 218 | 17.1 | 190-260 | 80.3 | 12.6 | 62.7-111.4 |
| 3 | 29-Apr | 00:44 | 30 | 0 | 0 | 225 | 19.2 | 189-257 | 88.9 | 23.0 | 58.0-143.6 |
| 4 | 29-Apr | 13:05 | 27 | 1 | 0 | 227 | 21.1 | 193-271 | 90.7 | 23.9 | 54.4-166.4 |
| 5 | 30-Apr | 00:39 | 33 | 4 | 0 | 228 | 20.2 | 182-273 | 103.0 | 23.6 | 53.9-157.0 |
| 6 | 30-Apr | 12:45 | 40 | 10 | 0 | 221 | 21.7 | 144-258 | 83.6 | 24.0 | 26.0-135.2 |
| 7 | 1-May | 01:24 | 67 | 4 | 0 | 218 | 18.6 | 181-261 | 76.1 | 18.6 | 50.3-144.6 |
| 8 | 1-May | 13:12 | 34 | 4 | 0 | 227 | 17.3 | 187-250 | 91.7 | 24.8 | 54.4-131.1 |
| 9 | 2-May | 01:03 | 62 | 5 | 0 | 232 | 16.2 | 201-270 | 107.3 | 22.9 | 65.6-173.3 |
| 10 | 2-May | 13:20 | 56 | 4 | 0 | 216 | 22.5 | 124-264 | 99.3 | 24.1 | 50.8-181.5 |
| 11 | 3-May | 01:23 | 70 | 1 | 0 | 228 | 23.3 | 124-270 | 103.1 | 24.9 | 50.7-164.6 |
| 12 | 3-May | 12:49 | 64 | 6 | 0 | 218 | 22.3 | 129-260 | 101.7 | 24.0 | 59.4-157.7 |
| 13 | 4-May | 01:09 | 66 | 3 | 0 | 226 | 19.1 | 147-265 | 97.6 | 21.8 | 56.9-147.3 |
| 14 | 4-May | 12:29 | 35 | 3 | 0 | 223 | 16.9 | 193-262 | 93.1 | 21.4 | 63.2-148.5 |
| 15 | 5-May | 01:31 | 25 | 3 | 0 | 227 | 24.4 | 169-265 | 100.5 | 30.1 | 49.1-164.0 |
| 16 | 5-May | 12:28 | 28 | 0 | 0 | 220 | 16.9 | 183-256 | 90.7 | 22.2 | 51.1-138.7 |
| 17 | 6-May | 00:33 | 38 | 2 | 0 | 227 | 17.8 | 185-268 | 97.6 | 23.3 | 53.5-142.3 |
| 18 | 6-May | 13:06 | 40 | 3 | 0 | 226 | 16.6 | 190-264 | 98.8 | 26.5 | 56.7-173.9 |
| 19 | 7-May | 00:07 | 65 | 5 | 0 | 218 | 21.5 | 177-275 | 98.2 | 26.1 | 53.8-174.5 |
| 20 | 7-May | 12:37 | 61 | 8 | 1 | 228 | 15.7 | 190-260 | 101.0 | 22.3 | 55.1-158.4 |
| 21 | 8-May | 00:17 | 70 | 0 | 0 | 223 | 23.2 | 161-270 | 100.0 | 28.6 | 42.1-171.0 |
| 22 | 10-May | 00:29 | 41 | 0 | 0 | 225 | 23.4 | 125-258 | 99.6 | 24.3 | 53.5-150.4 |
| 23 | 10-May | 13:02 | 41 | 1 | 0 | 231 | 19.1 | 176-275 | 99.7 | 24.2 | 41.5-166.7 |
| 24 | 11-May | 01:13 | 70 | 0 | 0 | 230 | 19.4 | 173-270 | 100.5 | 27.8 | 44.9-164.9 |

Table A1.9 (continued). Summary of steelhead trout releases at The Dalles Dam tailrace during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 25 | 11-May | 12:05 | 68 | 3 | 1 | 232 | 19.9 | 176-284 | 102.3 | 29.7 | 32.1-191.6 |
| 26 | 12-May | 00:44 | 71 | 0 | 0 | 228 | 18.2 | 176-269 | 98.6 | 25.0 | 44.8-154.5 |
| 27 | 12-May | 12:26 | 67 | 2 | 1 | 226 | 19.0 | 181-275 | 95.7 | 26.6 | 48.4-188.7 |
| 28 | 13-May | 01:11 | 72 | 0 | 0 | 230 | 20.7 | 177-275 | 99.0 | 29.4 | 43.6-174.1 |
| 29 | 13-May | 12:16 | 67 | 4 | 0 | 226 | 19.0 | 194-270 | 95.7 | 27.9 | 56.2-197.3 |
| 30 | 14-May | 00:23 | 78 | 1 | 1 | 224 | 22.4 | 153-278 | 97.1 | 27.3 | 28.2-176.9 |
| 31 | 14-May | 12:45 | 75 | 4 | 1 | 223 | 19.2 | 176-270 | 89.5 | 23.2 | 42.4-136.5 |
| 32 | 15-May | 00:35 | 76 | 0 | 2 | 229 | 22.2 | 170-274 | 100.7 | 29.7 | 46.4-184.4 |
| 33 | 15-May | 12:08 | 79 | 3 | 0 | 229 | 19.2 | 163-268 | 98.9 | 25.7 | 30.0-165.9 |
| 34 | 16-May | 00:51 | 77 | 1 | 2 | 228 | 22.0 | 115-267 | 97.5 | 24.9 | 51.9-169.5 |
| 35 | 16-May | 12:15 | 80 | 2 | 0 | 232 | 23.1 | 179-288 | 103.2 | 32.4 | 47.3-195.9 |
| 36 | 17-May | 00:02 | 78 | 1 | 3 | 229 | 19.6 | 183-271 | 98.4 | 25.9 | 48.1-164.5 |
| 37 | 17-May | 12:02 | 70 | 0 | 0 | 221 | 22.0 | 126-265 | 90.1 | 23.1 | 43.5-161.1 |
| 38 | 18-May | 00:54 | 80 | 0 | 2 | 220 | 20.2 | 175-269 | 90.5 | 27.5 | 41.8-180.0 |
| 39 | 18-May | 12:11 | 80 | 2 | 0 | 220 | 22.0 | 146-277 | 91.3 | 27.7 | 45.1-191.4 |
| 40 | 19-May | 00:45 | 74 | 4 | 4 | 221 | 22.7 | 166-269 | 93.0 | 29.8 | 40.1-169.6 |
| 41 | 19-May | 12:41 | 81 | 1 | 0 | 221 | 21.5 | 172-286 | 91.1 | 29.2 | 40.9-217.5 |
| 42 | 20-May | 00:30 | 81 | 1 | 0 | 229 | 24.1 | 167-288 | 101.2 | 31.8 | 34.6-196.5 |
| 43 | 20-May | 12:08 | 77 | 4 | 0 | 219 | 22.1 | 143-269 | 91.9 | 26.6 | 51.9-146.3 |
| 44 | 21-May | 00:29 | 81 | 0 | 0 | 219 | 23.8 | 158-274 | 91.0 | 30.3 | 32.0-172.9 |
| 45 | 21-May | 12:18 | 77 | 2 | 2 | 221 | 25.2 | 172-300 | 94.5 | 36.0 | 43.1-235.8 |
| 46 | 22-May | 00:23 | 77 | 3 | 0 | 214 | 25.5 | 162-322 | 84.3 | 31.4 | 35.1-232.0 |
| 47 | 22-May | 12:08 | 79 | 3 | 0 | 213 | 20.1 | 158-256 | 81.1 | 22.8 | 33.1-142.4 |
| 48 | 23-May | 00:39 | 80 | 0 | 1 | 218 | 26.3 | 152-271 | 90.2 | 33.4 | 36.6-185.6 |
| 49 | 23-May | 12:54 | 77 | 2 | 0 | 217 | 24.6 | 174-277 | 90.5 | 34.4 | 35.6-199.9 |
| 50 | 24-May | 00:22 | 77 | 2 | 0 | 217 | 23.8 | 168-290 | 85.6 | 28.7 | 39.0-201.2 |
| 51 | 24-May | 13:07 | 80 | 2 | 0 | 219 | 25.2 | 172-298 | 88.6 | 29.8 | 44.9-169.3 |

Table A1.9 (continued). Summary of steelhead trout releases at The Dalles Dam tailrace during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 52 | 25-May | 00:10 | 80 | 1 | 1 | 218 | 24.1 | 158-278 | 85.7 | 29.3 | 32.2-195.3 |
| 53 | 25-May | 16:10 | 81 | 1 | 0 | 218 | 25.6 | 160-284 | 88.1 | 31.5 | 34.5-195.8 |
| 54 | 26-May | 00:18 | 39 | 1 | 1 | 226 | 27.7 | 162-288 | 94.9 | 36.6 | 31.5-200.5 |
| 55 | 26-May | 13:08 | 44 | 2 | 0 | 222 | 22.1 | 171-280 | 90.3 | 27.8 | 38.4-192.7 |
| 56 | 27-May | 00:17 | 79 | 2 | 1 | 229 | 24.8 | 133-274 | 99.7 | 28.6 | 37.5-179.5 |
| 57 | 27-May | 12:22 | 80 | 1 | 1 | 227 | 22.2 | 176-270 | 95.6 | 26.7 | 46.7-159.5 |
| 58 | 28-May | 00:14 | 80 | 0 | 2 | 234 | 20.9 | 178-281 | 106.3 | 28.4 | 45.4-170.6 |
| 59 | 28-May | 12:29 | 79 | 2 | 0 | 224 | 25.8 | 166-278 | 95.1 | 34.7 | 37.4-193.1 |
| 60 | 29-May | 00:25 | 71 | 1 | 4 | 223 | 24.6 | 128-270 | 95.8 | 27.5 | 40.8-162.4 |
| 61 | 29-May | 12:10 | 75 | 4 | 2 | 216 | 23.3 | 166-277 | 84.8 | 30.5 | 32.3-196.6 |
| 62 | 30-May | 00:24 | 78 | 1 | 3 | 220 | 25.4 | 160-293 | 91.0 | 37.1 | 24.5-215.3 |
| 63 | 30-May | 12:32 | 74 | 6 | 2 | 225 | 20.6 | 172-270 | 95.6 | 30.8 | 38.5-186.7 |
| 64 | 31-May | 00:31 | 75 | 5 | 2 | 228 | 25.6 | 173-300 | 101.7 | 38.0 | 41.9-252.2 |
| 65 | 31-May | 12:38 | 78 | 3 | 1 | 230 | 23.9 | 173-286 | 105.5 | 41.4 | 32.1-294.0 |
| 66 | 1-Jun | 00:28 | 76 | 1 | 0 | 227 | 23.8 | 179-289 | 99.0 | 33.4 | 47.6-210.7 |
| 67 | 1-Jun | 12:16 | 78 | 3 | 1 | 231 | 24.1 | 174-290 | 105.0 | 37.5 | 41.6-260.2 |
| 68 | 2-Jun | 00:16 | 71 | 3 | 0 | 227 | 26.5 | 168-295 | 97.0 | 35.2 | 37.7-217.9 |
| Overall |  |  | 4398 | 153 | 42 | 224 | 22.7 | 115-322 | 95.1 | 29.5 | 24.5-294.0 |

Table A1.10. Summary of steelhead trout releases at Bonneville Dam tailrace during spring 2004. Dates, times, numbers of tagged fish released ( N ), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | 30-Apr | 10:19 | 16 | 1 | 0 | 224 | 14.3 | 195-240 | 99.0 | 21.1 | 65.1-143.2 |
| 2 | 30-Apr | 21:56 | 13 | 1 | 0 | 234 | 20.6 | 205-273 | 112.5 | 35.4 | 73.1-188.3 |
| 3 | 1-May | 10:08 | 13 | 3 | 0 | 235 | 20.0 | 194-258 | 111.1 | 23.9 | 66.4-144.3 |
| 4 | 2-May | 22:07 | 20 | 1 | 0 | 227 | 28.0 | 189-320 | 107.3 | 47.9 | 58.6-283.6 |
| 5 | 3-May | 22:00 | 10 | 6 | 0 | 225 | 31.4 | 179-280 | 103.4 | 45.2 | 49.3-201.4 |
| 6 | 4-May | 09:53 | 10 | 0 | 0 | 234 | 15.2 | 203-259 | 114.9 | 23.2 | 76.0-154.8 |
| 7 | 4-May | 22:01 | 20 | 0 | 0 | 236 | 17.5 | 200-265 | 109.7 | 23.5 | 64.0-156.9 |
| 8 | 5-May | 21:55 | 17 | 0 | 0 | 234 | 17.6 | 205-265 | 106.0 | 26.4 | 69.2-152.5 |
| 9 | 6-May | 10:11 | 20 | 1 | 0 | 237 | 23.4 | 190-279 | 115.1 | 36.3 | 51.0-182.4 |
| 10 | 6-May | 21:39 | 21 | 0 | 0 | 221 | 18.2 | 198-257 | 93.6 | 23.5 | 62.4-144.5 |
| 11 | 7-May | 10:25 | 22 | 1 | 1 | 224 | 24.8 | 187-290 | 96.5 | 35.9 | 57.0-205.0 |
| 12 | 7-May | 21:51 | 15 | 3 | 0 | 220 | 16.5 | 192-247 | 94.3 | 24.5 | 53.5-132.9 |
| 13 | 8-May | 10:07 | 21 | 0 | 0 | 227 | 24.2 | 187-280 | 105.5 | 40.5 | 55.6-201.4 |
| 14 | 8-May | 21:42 | 26 | 1 | 0 | 227 | 16.7 | 200-276 | 99.2 | 20.3 | 64.2-151.1 |
| 15 | 9-May | 10:12 | 21 | 0 | 0 | 232 | 26.8 | 196-285 | 105.8 | 36.7 | 62.2-194.0 |
| 16 | 11-May | 22:00 | 22 | 1 | 0 | 232 | 20.8 | 191-266 | 105.8 | 27.8 | 52.3-159.7 |
| 17 | 12-May | 09:23 | 24 | 1 | 0 | 231 | 15.0 | 208-276 | 105.7 | 25.1 | 75.6-181.6 |
| 18 | 12-May | 22:07 | 20 | 2 | 0 | 229 | 21.1 | 201-274 | 99.0 | 28.2 | 64.6-167.8 |
| 19 | 13-May | 10:01 | 22 | 0 | 0 | 228 | 20.9 | 182-260 | 85.6 | 23.0 | 48.4-125.4 |
| 20 | 13-May | 22:00 | 15 | 2 | 0 | 222 | 22.9 | 184-288 | 80.1 | 31.0 | 40.2-180.6 |
| 21 | 14-May | 10:01 | 14 | 0 | 1 | 225 | 13.6 | 206-253 | 82.4 | 16.6 | 60.8-117.1 |
| 22 | 14-May | 21:54 | 13 | 2 | 1 | 223 | 20.7 | 190-255 | 98.2 | 25.5 | 55.2-138.8 |
| 23 | 15-May | 10:25 | 16 | 0 | 0 | 224 | 22.8 | 181-267 | 96.6 | 31.7 | 52.4-176.3 |
| 24 | 15-May | 21:52 | 26 | 1 | 0 | 234 | 21.9 | 190-290 | 110.0 | 30.3 | 66.9-206.5 |

Table A1.10 (continued). Summary of steelhead trout releases at Bonneville Dam tailrace during spring 2004. Dates, times, numbers of tagged fish released ( N ), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 25 | 16-May | 10:16 | 21 | 3 | 0 | 244 | 23.5 | 193-296 | 123.8 | 39.7 | 49.1-221.6 |
| 26 | 16-May | 22:08 | 24 | 0 | 0 | 242 | 15.6 | 214-278 | 112.4 | 27.3 | 70.3-177.1 |
| 27 | 17-May | 09:55 | 23 | 1 | 0 | 244 | 16.0 | 207-270 | 117.5 | 26.2 | 65.3-168.4 |
| 28 | 17-May | 22:15 | 24 | 1 | 0 | 238 | 14.3 | 216-266 | 109.5 | 25.7 | 77.9-163.0 |
| 29 | 18-May | 09:52 | 24 | 0 | 0 | 248 | 19.8 | 213-302 | 122.5 | 29.4 | 71.5-193.5 |
| 30 | 18-May | 22:33 | 21 | 3 | 0 | 229 | 27.0 | 178-275 | 97.1 | 31.9 | 40.8-149.9 |
| 31 | 19-May | 10:05 | 22 | 0 | 0 | 239 | 23.0 | 205-288 | 112.7 | 37.8 | 62.3-213.7 |
| 32 | 19-May | 22:35 | 22 | 3 | 1 | 236 | 21.3 | 200-274 | 112.8 | 34.1 | 63.1-177.5 |
| 33 | 20-May | 10:11 | 23 | 0 | 0 | 237 | 13.0 | 206-256 | 108.9 | 23.4 | 65.6-151.6 |
| 34 | 20-May | 21:51 | 28 | 1 | 0 | 219 | 24.3 | 174-267 | 85.4 | 30.5 | 27.2-163.6 |
| 35 | 21-May | 10:01 | 23 | 1 | 0 | 240 | 22.4 | 206-293 | 114.0 | 38.9 | 68.1-216.2 |
| 36 | 21-May | 21:49 | 22 | 2 | 0 | 236 | 18.3 | 205-273 | 108.5 | 27.2 | 71.5-163.5 |
| 37 | 22-May | 10:55 | 22 | 3 | 0 | 237 | 21.4 | 200-280 | 110.8 | 36.2 | 63.8-193.4 |
| 38 | 22-May | 23:49 | 12 | 0 | 0 | 225 | 22.2 | 175-255 | 96.6 | 27.8 | 42.4-140.7 |
| 39 | 23-May | 10:22 | 30 | 1 | 0 | 227 | 27.2 | 184-305 | 97.5 | 38.0 | 49.7-236.0 |
| 40 | 23-May | 22:14 | 20 | 3 | 1 | 234 | 29.8 | 178-310 | 109.6 | 46.4 | 47.2-261.6 |
| 41 | 24-May | 11:21 | 24 | 0 | 0 | 234 | 20.9 | 188-265 | 104.9 | 28.4 | 58.1-176.4 |
| 42 | 24-May | 22:15 | 24 | 0 | 0 | 231 | 30.9 | 174-304 | 110.4 | 45.3 | 44.5-241.8 |
| 43 | 25-May | 09:26 | 24 | 0 | 0 | 227 | 22.7 | 180-270 | 95.2 | 33.2 | 44.4-172.6 |
| 44 | 25-May | 22:06 | 23 | 0 | 0 | 221 | 27.1 | 157-263 | 104.3 | 39.9 | 46.4-162.5 |
| 45 | 26-May | 10:06 | 24 | 0 | 0 | 219 | 21.7 | 181-260 | 83.4 | 25.2 | 43.8-135.4 |
| 46 | 26-May | 23:49 | 24 | 0 | 0 | 232 | 31.3 | 176-289 | 88.8 | 33.7 | 43.3-158.4 |
| 47 | 27-May | 10:35 | 23 | 1 | 0 | 219 | 19.4 | 188-255 | 89.0 | 33.0 | 54.7-145.7 |
| 48 | 27-May | 21:54 | 22 | 2 | 0 | 229 | 25.5 | 190-283 | 98.5 | 37.3 | 52.3-197.6 |
| 49 | 28-May | 12:05 | 21 | 4 | 0 | 247 | 21.7 | 206-291 | 125.2 | 31.5 | 65.4-181.0 |
| 50 | 28-May | 22:08 | 24 | 0 | 0 | 229 | 27.7 | 183-274 | 85.2 | 30.7 | 38.5-137.6 |
| 51 | 29-May | 10:50 | 24 | 0 | 0 | 230 | 25.9 | 178-276 | 102.9 | 31.1 | 44.0-152.3 |

Table A1.10 (continued). Summary of steelhead trout releases at Bonneville Dam tailrace during spring 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 52 | 29-May | 22:27 | 23 | 3 | 0 | 224 | 24.8 | 190-283 | 81.3 | 30.0 | 44.2-173.7 |
| 53 | 30-May | 10:59 | 24 | 0 | 0 | 229 | 27.8 | 178-273 | 90.6 | 32.9 | 43.3-165.4 |
| 54 | 31-May | 00:01 | 17 | 0 | 0 | 229 | 32.5 | 167-300 | 114.9 | 51.0 | 43.2-262.4 |
| 55 | 31-May | 09:57 | 16 | 1 | 0 | 230 | 32.3 | 178-281 | 101.6 | 48.1 | 47.6-181.0 |
| 56 | 31-May | 22:00 | 20 | 0 | 0 | 228 | 30.6 | 159-270 | 104.8 | 37.3 | 31.4-150.4 |
| 57 | 1-Jun | 11:19 | 17 | 0 | 0 | 232 | 26.4 | 161-261 | 96.2 | 31.4 | 32.0-139.9 |
| 58 | 1-Jun | 21:50 | 32 | 0 | 0 | 236 | 27.9 | 178-295 | 114.3 | 45.6 | 46.6-231.4 |
| 59 | 2-Jun | 10:00 | 28 | 0 | 1 | 239 | 22.5 | 193-282 | 112.0 | 38.2 | 24.2-192.8 |
| 60 | 2-Jun | 22:59 | 15 | 2 | 0 | 254 | 22.7 | 212-284 | 142.5 | 45.7 | 71.4-206.2 |
| 61 | 3-Jun | 10:06 | 27 | 3 | 0 | 232 | 22.6 | 184-278 | 108.0 | 38.9 | 48.8-195.5 |
| Overall |  |  | 1274 | 65 | 6 | 231 | 22.6 | 157-320 | 104.0 | 32.6 | 24.2-283.6 |

Table A1.11. Summary of subyearling Chinook salmon releases at Bonneville Dam ice and trash sluiceway during summer 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | 21-Jun | 11:43 | 17 | 2 | 0 | 122 | 8.4 | 111-138 | 18.3 | 4.3 | 13.9-26.0 |
| 2 | 22-Jun | 00:24 | 16 | 3 | 0 | 124 | 12.8 | 112-161 | 20.1 | 6.3 | 13.9-37.7 |
| 3 | 23-Jun | 12:08 | 21 | 0 | 0 | 121 | 5.2 | 115-134 | 18.4 | 2.5 | 14.3-26.4 |
| 4 | 23-Jun | 23:50 | 20 | 1 | 0 | 120 | 6.9 | 108-135 | 17.6 | 2.7 | 13.9-23.7 |
| 5 | 25-Jun | 12:00 | 19 | 0 | 0 | 118 | 7.3 | 110-138 | 16.9 | 2.9 | 13.5-26.1 |
| 6 | 26-Jun | 00:00 | 20 | 0 | 0 | 119 | 3.9 | 110-127 | 17.9 | 1.7 | 15.3-20.8 |
| 7 | 27-Jun | 11:55 | 19 | 0 | 0 | 120 | 9.0 | 109-142 | 16.9 | 3.6 | 13.0-26.3 |
| 8 | 27-Jun | 23:48 | 19 | 0 | 0 | 116 | 6.7 | 110-137 | 15.4 | 3.7 | 13.2-28.8 |
| 9 | 29-Jun | 12:12 | 14 | 0 | 0 | 118 | 10.6 | 109-150 | 17.2 | 5.4 | 14.0-34.5 |
| 10 | 30-Jun | 00:09 | 24 | 0 | 0 | 116 | 5.6 | 110-134 | 16.8 | 2.5 | 13.6-23.4 |
| 11 | 1-Jul | 11:57 | 19 | 0 | 0 | 115 | 4.7 | 106-124 | 15.3 | 2.1 | 13.2-19.1 |
| 12 | 1-Jul | 23:58 | 19 | 0 | 0 | 115 | 5.2 | 109-132 | 15.3 | 2.4 | 13.0-23.3 |
| 13 | 3-Jul | 12:00 | 24 | 0 | 0 | 117 | 7.5 | 107-132 | 16.2 | 3.0 | 13.4-23.9 |
| 14 | 4-Jul | 00:16 | 19 | 0 | 0 | 116 | 5.9 | 105-129 | 15.8 | 2.8 | 13.0-22.6 |
| 15 | 5-Jul | 12:06 | 24 | 0 | 0 | 117 | 9.8 | 105-144 | 17.7 | 4.7 | 13.6-28.8 |
| 16 | 5-Jul | 23:51 | 19 | 0 | 0 | 114 | 5.1 | 106-127 | 15.3 | 2.4 | 13.0-23.6 |

Table A1.11 (continued). Summary of subyearling Chinook salmon releases at Bonneville Dam ice and trash sluiceway during summer 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 17 | 7-Jul | 12:07 | 18 | 0 | 1 | 115 | 5.2 | 105-125 | 15.6 | 2.0 | 13.0-19.5 |
| 18 | 7-Jul | 23:58 | 19 | 0 | 0 | 113 | 9.8 | 103-147 | 15.2 | 2.3 | 13.0-21.2 |
| 19 | $9-\mathrm{Jul}$ | 12:10 | 19 | 0 | 0 | 112 | 5.0 | 105-124 | 15.0 | 1.9 | 13.0-20.6 |
| 20 | 9 -Jul | 23:56 | 18 | 1 | 0 | 113 | 5.9 | 105-128 | 15.9 | 2.6 | 13.0-22.8 |
| 21 | 11-Jul | 12:04 | 18 | 1 | 0 | 114 | 10.9 | 105-145 | 17.2 | 6.7 | 13.3-39.7 |
| 22 | 12-Jul | 00:00 | 19 | 0 | 0 | 118 | 11.6 | 108-151 | 17.3 | 5.6 | 13.2-35.2 |
| 23 | 13-Jul | 12:00 | 19 | 0 | 0 | 120 | 13.5 | 108-149 | 18.6 | 6.3 | 13.5-32.3 |
| 24 | 14-Jul | 00:11 | 19 | 0 | 0 | 122 | 14.8 | 107-150 | 20.1 | 7.0 | 13.5-36.9 |
| 25 | 15-Jul | 11:55 | 19 | 0 | 0 | 116 | 14.0 | 105-166 | 17.0 | 7.6 | 13.0-46.2 |
| 26 | 16-Jul | 00:17 | 19 | 0 | 0 | 117 | 12.3 | 105-147 | 17.3 | 5.0 | 13.1-29.9 |
| 27 | 17-Jul | 12:00 | 19 | 0 | 0 | 116 | 7.6 | 106-134 | 17.1 | 3.3 | 13.2-25.2 |
| 28 | 18-Jul | 00:00 | 18 | 0 | 0 | 115 | 6.6 | 105-127 | 17.2 | 3.1 | 13.0-23.8 |
| 29 | 19-Jul | 12:00 | 19 | 0 | 0 | 125 | 13.0 | 108-149 | 21.0 | 5.7 | 14.3-35.1 |
| 30 | 19-Jul | 23:55 | 19 | 0 | 0 | 119 | 8.6 | 110-142 | 16.9 | 3.4 | 13.0-25.7 |
| 31 | 21-Jul | 11:54 | 18 | 1 | 0 | 115 | 6.4 | 108-131 | 16.0 | 2.9 | 13.0-24.1 |
| 32 | 22-Jul | 00:06 | 19 | 0 | 0 | 113 | 8.3 | 104-134 | 16.4 | 4.4 | 13.0-27.3 |
| Overall |  |  | 612 | 9 | 1 | 117 | 9.2 | 103-166 | 17.0 | 4.3 | 13.0-46.2 |

Table A1.12. Summary of subyearling Chinook salmon releases at The Dalles Dam ice and trash sluiceway (Sluiceway), and the tailrace 550 m downstream of the spillway beneath the I-197 bridge (Tailrace) during summer 2004. Dates, times, numbers of tagged fish released ( N ), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Location | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | Tailrace | 20-Jun | 00:41 | 117 | 2 | 0 | 119 | 8.8 | 106-149 | 18.6 | 4.1 | 14.0-34.6 |
| 2 | Sluiceway | 20-Jun | 07:00 | 61 | 1 | 0 | 119 | 7.3 | 110-141 | 17.8 | 3.6 | 13.6-27.9 |
| 3 | Tailrace | 20-Jun | 12:35 | 106 | 1 | 0 | 117 | 6.9 | 108-138 | 16.6 | 3.1 | 13.0-26.2 |
| 4 | Tailrace | 21-Jun | 00:31 | 117 | 2 | 0 | 118 | 6.9 | 109-137 | 18.1 | 3.5 | 13.2-28.2 |
| 5 | Tailrace | 21-Jun | 12:35 | 115 | 4 | 0 | 121 | 8.0 | 108-146 | 18.4 | 4.0 | 13.0-33.2 |
| 6 | Sluiceway | 21-Jun | 13:21 | 57 | 3 | 0 | 120 | 6.6 | 110-135 | 17.7 | 3.3 | 13.3-25.8 |
| 7 | Tailrace | 22-Jun | 00:23 | 114 | 3 | 0 | 120 | 7.6 | 104-141 | 18.0 | 3.5 | 14.1-31.0 |
| 8 | Sluiceway | 22-Jun | 09:12 | 61 | 1 | 0 | 117 | 6.1 | 108-134 | 16.5 | 2.7 | 13.1-25.4 |
| 9 | Tailrace | 22-Jun | 13:05 | 65 | 1 | 0 | 119 | 7.0 | 111-145 | 17.4 | 3.1 | 13.6-28.4 |
| 10 | Tailrace | 23-Jun | 01:05 | 116 | 2 | 0 | 119 | 7.4 | 109-140 | 18.5 | 3.2 | 13.4-27.6 |
| 11 | Tailrace | 23-Jun | 12:38 | 118 | 1 | 1 | 118 | 7.6 | 109-142 | 17.1 | 3.4 | 13.1-28.3 |
| 12 | Sluiceway | 23-Jun | 19:00 | 61 | 1 | 0 | 116 | 5.6 | 109-135 | 16.9 | 2.7 | 13.5-26.6 |
| 13 | Tailrace | 24-Jun | 00:43 | 117 | 2 | 0 | 115 | 4.9 | 108-132 | 16.5 | 2.4 | 13.2-24.9 |
| 14 | Tailrace | 24-Jun | 13:38 | 118 | 1 | 0 | 116 | 7.0 | 109-153 | 16.4 | 3.4 | 13.3-38.1 |
| 15 | Sluiceway | 24-Jun | 19:00 | 59 | 3 | 0 | 114 | 4.8 | 110-142 | 14.8 | 2.1 | 13.0-26.7 |
| 16 | Tailrace | 25-Jun | 00:35 | 88 | 2 | 0 | 115 | 6.3 | 109-142 | 15.3 | 3.0 | 13.0-30.9 |
| 17 | Tailrace | 25-Jun | 13:13 | 47 | 1 | 0 | 114 | 5.5 | 107-137 | 15.5 | 3.0 | 13.0-30.1 |
| 18 | Sluiceway | 25-Jun | 13:00 | 58 | 0 | 0 | 116 | 7.5 | 108-144 | 16.5 | 3.6 | 13.4-29.7 |
| 19 | Tailrace | 26-Jun | 00:56 | 89 | 4 | 0 | 115 | 6.7 | 108-140 | 15.6 | 3.2 | 13.1-29.3 |
| 20 | Sluiceway | 26-Jun | 07:00 | 58 | 0 | 0 | 115 | 8.0 | 109-154 | 16.3 | 4.7 | 13.0-42.4 |
| 21 | Tailrace | 26-Jun | 13:05 | 46 | 3 | 3 | 117 | 9.1 | 110-150 | 16.5 | 4.9 | 13.1-35.1 |
| 22 | Tailrace | 27-Jun | 01:07 | 19 | 1 | 0 | 117 | 6.3 | 110-132 | 17.2 | 2.9 | 13.4-24.1 |
| 23 | Sluiceway | 27-Jun | 01:02 | 28 | 0 | 0 | 118 | 9.6 | 108-145 | 17.8 | 4.7 | 13.7-32.8 |
| 24 | Tailrace | 27-Jun | 13:04 | 35 | 1 | 0 | 121 | 11.4 | 109-153 | 19.2 | 6.5 | 13.4-36.1 |

Table A1.12 (continued). Summary of subyearling Chinook salmon releases at The Dalles Dam ice and trash sluiceway (Sluiceway), and the tailrace 550 m downstream of the spillway beneath the I-197 bridge (Tailrace) during summer 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Location | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 25 | Tailrace | 28-Jun | 01:16 | 32 | 0 | 1 | 118 | 9.1 | 109-147 | 17.5 | 4.3 | 13.3-32.1 |
| 26 | Sluiceway | 28-Jun | 07:06 | 27 | 1 | 0 | 116 | 6.3 | 110-136 | 16.5 | 3.2 | 13.0-25.7 |
| 27 | Tailrace | 28-Jun | 13:12 | 14 | 0 | 0 | 116 | 6.1 | 111-135 | 16.2 | 3.3 | 13.0-26.3 |
| 28 | Tailrace | 29-Jun | 00:45 | 58 | 0 | 0 | 114 | 4.4 | 107-134 | 16.3 | 2.3 | 13.2-24.9 |
| 29 | Tailrace | 29-Jun | 12:53 | 81 | 0 | 0 | 115 | 7.0 | 107-146 | 16.1 | 3.7 | 13.0-32.3 |
| 30 | Sluiceway | 29-Jun | 19:04 | 25 | 0 | 1 | 114 | 5.7 | 109-133 | 15.7 | 2.7 | 13.0-25.0 |
| 31 | Tailrace | 30-Jun | 01:02 | 28 | 0 | 0 | 117 | 10.8 | 106-149 | 17.7 | 5.4 | 13.1-36.3 |
| 32 | Sluiceway | 30-Jun | 01:00 | 37 | 0 | 0 | 114 | 5.7 | 108-130 | 15.8 | 2.6 | 13.0-24.3 |
| 33 | Tailrace | 30-Jun | 12:49 | 165 | 2 | 2 | 115 | 7.9 | 107-146 | 16.2 | 3.9 | 13.0-34.2 |
| 34 | Tailrace | 1-Jul | 00:18 | 101 | 2 | 3 | 115 | 7.5 | 104-146 | 15.5 | 3.6 | 13.0-31.8 |
| 35 | Tailrace | 1-Jul | 12:36 | 125 | 1 | 2 | 115 | 7.4 | 107-148 | 16.5 | 3.8 | 13.0-35.2 |
| 36 | Sluiceway | 1-Jul | 13:16 | 60 | 1 | 0 | 114 | 6.9 | 107-148 | 15.6 | 3.5 | 13.1-33.0 |
| 37 | Tailrace | 2-Jul | 01:12 | 27 | 0 | 2 | 115 | 8.3 | 105-138 | 16.1 | 4.2 | 13.1-27.1 |
| 38 | Sluiceway | 2-Jul | 01:05 | 34 | 0 | 0 | 116 | 7.0 | 109-138 | 15.9 | 3.1 | 13.0-27.6 |
| 39 | Tailrace | 2-Jul | 12:48 | 53 | 2 | 0 | 115 | 7.1 | 107-136 | 16.0 | 3.5 | 13.1-26.6 |
| 40 | Tailrace | 3-Jul | 01:36 | 73 | 0 | 2 | 115 | 10.5 | 102-154 | 16.7 | 5.6 | 13.0-40.5 |
| 41 | Tailrace | 3-Jul | 13:08 | 50 | 0 | 0 | 115 | 7.5 | 107-142 | 16.1 | 3.9 | 13.0-31.0 |
| 42 | Sluiceway | 3-Jul | 19:01 | 39 | 1 | 0 | 115 | 6.4 | 108-133 | 16.0 | 2.9 | 13.1-27.4 |
| 43 | Tailrace | 4-Jul | 00:38 | 40 | 0 | 0 | 114 | 6.8 | 107-139 | 15.9 | 3.3 | 13.1-27.1 |
| 44 | Tailrace | 4-Jul | 13:30 | 18 | 0 | 0 | 117 | 8.5 | 109-142 | 17.3 | 4.1 | 13.0-28.1 |
| 45 | Sluiceway | 4-Jul | 13:05 | 30 | 2 | 0 | 117 | 10.6 | 107-142 | 17.5 | 5.2 | 13.2-30.1 |
| 46 | Tailrace | 5-Jul | 00:46 | 64 | 1 | 0 | 117 | 11.1 | 103-150 | 18.6 | 5.9 | 13.2-40.0 |
| 47 | Sluiceway | 5-Jul | 07:00 | 32 | 0 | 0 | 116 | 11.7 | 104-160 | 17.5 | 6.4 | 13.1-40.7 |
| 48 | Tailrace | 5-Jul | 13:01 | 27 | 0 | 0 | 116 | 8.8 | 107-150 | 16.8 | 5.1 | 13.0-37.8 |

Table A1.12 (continued). Summary of subyearling Chinook salmon releases at The Dalles Dam ice and trash sluiceway (Sluiceway), and the tailrace 550 m downstream of the spillway beneath the I-197 bridge (Tailrace) during summer 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Location | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 49 | Tailrace | 6-Jul | 00:55 | 33 | 0 | 0 | 118 | 11.8 | 107-151 | 18.7 | 6.1 | 13.5-38.5 |
| 50 | Sluiceway | 6-Jul | 01:00 | 34 | 1 | 1 | 119 | 11.7 | 105-145 | 19.4 | 6.4 | 13.2-34.2 |
| 51 | Tailrace | 6-Jul | 12:31 | 100 | 1 | 0 | 116 | 11.4 | 104-168 | 17.6 | 6.1 | 13.0-48.2 |
| 52 | Tailrace | 7-Jul | 00:46 | 43 | 0 | 0 | 114 | 6.9 | 107-137 | 16.2 | 3.1 | 13.0-27.4 |
| 53 | Tailrace | 7-Jul | 12:49 | 108 | 1 | 2 | 115 | 8.8 | 105-147 | 16.9 | 4.3 | 13.1-34.0 |
| 54 | Sluiceway | 7-Jul | 13:14 | 76 | 2 | 0 | 117 | 8.5 | 105-146 | 17.3 | 4.0 | 13.0-31.7 |
| 55 | Tailrace | 8 -Jul | 00:37 | 82 | 0 | 4 | 114 | 9.1 | 103-146 | 16.2 | 4.4 | 13.0-32.6 |
| 56 | Tailrace | 8 -Jul | 13:01 | 79 | 1 | 0 | 113 | 8.2 | 101-144 | 15.7 | 3.6 | 13.0-32.2 |
| 57 | Sluiceway | 8 -Jul | 19:00 | 26 | 0 | 0 | 116 | 10.6 | 104-148 | 17.1 | 5.2 | 13.1-35.7 |
| 58 | Tailrace | $9-\mathrm{Jul}$ | 01:00 | 11 | 0 | 1 | 122 | 13.0 | 108-145 | 19.6 | 6.7 | 13.5-30.1 |
| 59 | Sluiceway | 9 -Jul | 01:00 | 25 | 0 | 0 | 118 | 13.5 | 105-146 | 18.3 | 6.3 | 13.0-32.5 |
| 60 | Tailrace | $9-\mathrm{Jul}$ | 13:03 | 39 | 0 | 0 | 111 | 6.4 | 105-142 | 15.3 | 2.8 | 13.0-29.4 |
| 61 | Tailrace | 10-Jul | 00:49 | 78 | 0 | 0 | 115 | 10.7 | 103-149 | 17.6 | 5.1 | 13.1-36.1 |
| 62 | Tailrace | 10-Jul | 12:47 | 42 | 0 | 3 | 119 | 14.8 | 103-156 | 18.6 | 7.7 | 13.0-41.0 |
| 63 | Sluiceway | 10-Jul | 13:22 | 50 | 0 | 2 | 117 | 10.5 | 107-151 | 17.4 | 5.6 | 13.0-36.7 |
| 64 | Tailrace | 11-Jul | 00:34 | 58 | 1 | 1 | 124 | 13.9 | 106-159 | 21.3 | 7.5 | 13.1-49.8 |
| 65 | Sluiceway | 11-Jul | 07:00 | 55 | 0 | 0 | 124 | 12.6 | 106-151 | 20.3 | 6.4 | 13.0-37.3 |
| 66 | Tailrace | 11-Jul | 13:10 | 60 | 0 | 2 | 127 | 17.4 | 105-169 | 22.9 | 10.5 | 13.0-50.2 |
| 67 | Tailrace | 12-Jul | 00:38 | 50 | 0 | 1 | 120 | 12.9 | 105-158 | 19.6 | 6.9 | 13.0-42.8 |
| 68 | Sluiceway | 12-Jul | 01:02 | 51 | 1 | 1 | 120 | 13.8 | 104-156 | 20.0 | 7.4 | 13.0-45.0 |
| 69 | Tailrace | 12-Jul | 12:46 | 59 | 0 | 2 | 119 | 12.2 | 105-163 | 18.7 | 6.4 | 13.0-47.0 |
| 70 | Tailrace | 13-Jul | 00:28 | 63 | 1 | 1 | 125 | 15.8 | 105-162 | 21.6 | 9.2 | 13.0-48.8 |
| 71 | Tailrace | 13-Jul | 13:12 | 49 | 0 | 0 | 125 | 16.8 | 106-168 | 21.4 | 8.5 | 13.1-50.5 |
| 72 | Sluiceway | 13-Jul | 19:00 | 42 | 0 | 1 | 133 | 16.5 | 107-167 | 25.6 | 9.8 | 13.3-52.2 |

Table A1.12 (continued). Summary of subyearling Chinook salmon releases at The Dalles Dam ice and trash sluiceway (Sluiceway), and the tailrace 550 m downstream of the spillway beneath the I-197 bridge (Tailrace) during summer 2004. Dates, times, numbers of tagged fish released ( N ), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Location | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 73 | Tailrace | 14-Jul | 01:10 | 25 | 0 | 1 | 129 | 14.5 | 109-162 | 23.1 | 8.6 | 13.7-47.4 |
| 74 | Sluiceway | 14-Jul | 07:07 | 47 | 1 | 0 | 128 | 15.1 | 105-164 | 22.5 | 8.6 | 13.0-50.2 |
| 75 | Tailrace | 14-Jul | 13:10 | 21 | 0 | 0 | 120 | 11.8 | 104-142 | 18.1 | 4.9 | 13.1-27.9 |
| 76 | Tailrace | 15-Jul | 01:08 | 92 | 0 | 1 | 117 | 10.1 | 104-156 | 17.3 | 4.7 | 13.0-38.6 |
| 77 | Sluiceway | 15-Jul | 01:02 | 70 | 0 | 1 | 119 | 11.8 | 105-154 | 18.4 | 5.8 | 13.3-42.1 |
| 78 | Tailrace | 15-Jul | 12:46 | 62 | 0 | 0 | 119 | 11.3 | 107-145 | 17.6 | 4.6 | 13.0-30.1 |
| 79 | Tailrace | 16-Jul | 00:40 | 68 | 0 | 2 | 118 | 10.3 | 105-149 | 17.5 | 4.7 | 13.0-35.5 |
| 80 | Tailrace | 16-Jul | 12:54 | 59 | 0 | 2 | 118 | 12.2 | 104-150 | 18.1 | 5.5 | 13.2-36.9 |
| 81 | Sluiceway | 16-Jul | 13:11 | 51 | 2 | 3 | 118 | 10.0 | 105-150 | 17.6 | 4.8 | 13.0-35.6 |
| 82 | Tailrace | 17-Jul | 00:18 | 101 | 2 | 1 | 121 | 14.1 | 104-165 | 19.6 | 7.6 | 13.1-49.9 |
| 83 | Tailrace | 17-Jul | 12:40 | 106 | 1 | 1 | 121 | 12.2 | 105-153 | 19.1 | 5.9 | 13.0-41.9 |
| 84 | Sluiceway | 17-Jul | 19:00 | 56 | 1 | 0 | 115 | 6.7 | 105-129 | 16.5 | 2.5 | 13.0-21.2 |
| 85 | Tailrace | 18-Jul | 00:44 | 79 | 0 | 2 | 114 | 8.4 | 104-147 | 16.9 | 4.7 | 13.1-43.8 |
| 86 | Tailrace | 18-Jul | 13:08 | 108 | 4 | 1 | 118 | 10.5 | 105-161 | 17.8 | 5.6 | 13.1-47.2 |
| 87 | Sluiceway | 18-Jul | 13:10 | 59 | 0 | 0 | 117 | 9.1 | 103-149 | 17.4 | 4.3 | 13.1-35.6 |
| 88 | Tailrace | 19-Jul | 00:29 | 162 | 2 | 0 | 115 | 8.5 | 103-157 | 16.5 | 3.7 | 13.1-43.1 |
| 89 | Sluiceway | 19-Jul | 07:00 | 67 | 0 | 1 | 113 | 5.6 | 105-135 | 16.0 | 2.0 | 13.0-23.0 |
| 90 | Tailrace | 19-Jul | 13:11 | 158 | 0 | 2 | 116 | 9.5 | 103-160 | 17.2 | 5.0 | 13.1-46.0 |
| 91 | Tailrace | 20-Jul | 00:34 | 106 | 0 | 1 | 113 | 7.5 | 103-138 | 16.1 | 2.9 | 13.0-28.8 |
| 92 | Sluiceway | 20-Jul | 01:06 | 60 | 0 | 2 | 113 | 7.3 | 103-144 | 16.0 | 3.0 | 13.1-30.9 |
| Overall |  |  |  | 5980 | 75 | 60 | 117 | 10.0 | 101-169 | 17.4 | 5.1 | 13.0-52.2 |

Table A1.13. Summary of subyearling Chinook salmon releases at Bonneville Dam tailrace during summer 2004. Dates, times, numbers of tagged fish released (N), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 1 | 21-Jun | 13:04 | 56 | 4 | 0 | 119 | 6.2 | 106-134 | 17.5 | 2.7 | 14.2-26.4 |
| 2 | 22-Jun | 01:28 | 56 | 4 | 0 | 122 | 7.8 | 109-143 | 18.6 | 3.7 | 13.7-30.1 |
| 3 | 22-Jun | 13:29 | 55 | 3 | 0 | 120 | 6.8 | 109-145 | 17.9 | 3.2 | 14.3-29.7 |
| 4 | 23-Jun | 01:22 | 57 | 3 | 0 | 122 | 9.0 | 106-146 | 19.1 | 4.4 | 13.6-33.0 |
| 5 | 23-Jun | 13:21 | 58 | 2 | 0 | 123 | 8.3 | 110-141 | 18.7 | 4.0 | 14.3-30.9 |
| 6 | 24-Jun | 01:05 | 58 | 4 | 0 | 121 | 9.7 | 108-156 | 18.4 | 4.6 | 13.3-32.7 |
| 7 | 24-Jun | 13:13 | 60 | 4 | 0 | 119 | 5.6 | 109-136 | 17.3 | 2.3 | 13.6-25.0 |
| 8 | 25-Jun | 01:06 | 59 | 1 | 0 | 120 | 6.2 | 110-135 | 18.0 | 2.9 | 14.6-27.9 |
| 9 | 25-Jun | 13:01 | 62 | 0 | 0 | 119 | 6.9 | 108-145 | 17.2 | 3.2 | 13.8-34.3 |
| 10 | 26-Jun | 00:56 | 59 | 3 | 0 | 122 | 9.7 | 110-153 | 19.2 | 4.6 | 14.7-34.3 |
| 11 | 26-Jun | 13:01 | 59 | 2 | 0 | 123 | 8.4 | 112-143 | 18.5 | 3.8 | 14.0-28.5 |
| 12 | 27-Jun | 00:57 | 57 | 2 | 0 | 116 | 6.5 | 106-141 | 15.5 | 3.0 | 13.1-29.4 |
| 13 | 27-Jun | 12:55 | 59 | 0 | 0 | 118 | 7.4 | 107-140 | 16.0 | 3.2 | 13.0-28.5 |
| 14 | 28-Jun | 01:05 | 56 | 3 | 0 | 119 | 10.2 | 105-145 | 17.1 | 5.2 | 13.0-37.4 |
| 15 | 28-Jun | 13:05 | 60 | 1 | 0 | 118 | 8.0 | 109-155 | 16.6 | 3.8 | 13.1-32.1 |
| 16 | 29-Jun | 01:18 | 38 | 1 | 0 | 121 | 10.7 | 109-155 | 19.0 | 5.3 | 13.5-34.9 |
| 17 | 29-Jun | 13:07 | 38 | 1 | 0 | 119 | 8.4 | 106-146 | 17.8 | 4.9 | 13.1-41.0 |
| 18 | 30-Jun | 01:05 | 50 | 1 | 0 | 116 | 5.8 | 105-135 | 15.6 | 2.3 | 13.0-23.3 |
| 19 | 30-Jun | 13:00 | 76 | 0 | 0 | 117 | 7.5 | 106-145 | 15.8 | 3.2 | 13.2-29.0 |
| 20 | 1-Jul | 01:00 | 63 | 0 | 0 | 117 | 6.6 | 108-143 | 16.0 | 3.4 | 13.0-30.0 |
| 21 | 1-Jul | 12:49 | 57 | 1 | 0 | 118 | 9.4 | 107-151 | 16.7 | 4.4 | 13.0-33.5 |
| 22 | 2-Jul | 00:54 | 55 | 0 | 0 | 115 | 4.5 | 105-133 | 15.2 | 2.3 | 13.0-24.5 |
| 23 | 2-Jul | 12:52 | 71 | 0 | 0 | 116 | 6.9 | 108-139 | 16.3 | 3.5 | 13.0-30.9 |
| 24 | 3-Jul | 01:01 | 59 | 0 | 4 | 116 | 9.3 | 105-150 | 16.2 | 3.9 | 13.0-30.0 |

Table A1.13 (continued). Summary of subyearling Chinook salmon releases at Bonneville Dam tailrace during summer 2004. Dates, times, numbers of tagged fish released ( N ), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 25 | 3-Jul | 12:53 | 44 | 2 | 0 | 116 | 5.0 | 110-135 | 15.7 | 2.6 | 13.0-23.5 |
| 26 | 4-Jul | 01:10 | 57 | 1 | 2 | 114 | 4.6 | 107-132 | 15.2 | 2.5 | 13.0-26.8 |
| 27 | 4-Jul | 13:08 | 50 | 2 | 1 | 114 | 4.7 | 106-133 | 14.8 | 1.9 | 13.0-22.0 |
| 28 | 5-Jul | 00:55 | 68 | 1 | 1 | 114 | 6.1 | 105-143 | 15.4 | 2.8 | 13.0-30.0 |
| 29 | 5-Jul | 12:58 | 59 | 1 | 0 | 116 | 7.8 | 106-143 | 16.5 | 4.0 | 13.0-34.7 |
| 30 | 6-Jul | 00:56 | 61 | 3 | 1 | 116 | 8.0 | 109-152 | 15.9 | 3.8 | 13.0-34.2 |
| 31 | 6-Jul | 13:00 | 50 | 1 | 0 | 115 | 7.5 | 105-144 | 15.7 | 3.9 | 13.0-29.7 |
| 32 | 7-Jul | 00:56 | 64 | 1 | 0 | 114 | 8.1 | 104-143 | 16.4 | 4.2 | 13.0-35.1 |
| 33 | 7-Jul | 12:58 | 54 | 0 | 1 | 113 | 6.9 | 105-140 | 15.1 | 3.4 | 13.0-26.8 |
| 34 | 8-Jul | 00:56 | 64 | 1 | 0 | 114 | 8.4 | 102-143 | 16.6 | 3.9 | 13.0-30.2 |
| 35 | 8-Jul | 13:09 | 65 | 1 | 0 | 115 | 7.5 | 106-146 | 16.1 | 3.4 | 13.0-29.3 |
| 36 | 9-Jul | 00:53 | 59 | 1 | 0 | 116 | 8.8 | 104-146 | 16.8 | 4.3 | 13.0-31.4 |
| 37 | 9-Jul | 13:18 | 60 | 0 | 0 | 116 | 7.5 | 107-141 | 16.7 | 3.8 | 13.0-29.3 |
| 38 | 10-Jul | 00:54 | 60 | 1 | 0 | 113 | 5.6 | 105-138 | 15.1 | 2.2 | 13.0-24.3 |
| 39 | 10-Jul | 13:00 | 65 | 1 | 0 | 114 | 5.6 | 105-130 | 16.2 | 3.1 | 13.1-32.1 |
| 40 | 11-Jul | 00:57 | 51 | 0 | 1 | 118 | 8.1 | 106-145 | 17.5 | 3.9 | 13.0-31.8 |
| 41 | 11-Jul | 12:53 | 62 | 1 | 0 | 113 | 6.0 | 105-141 | 15.5 | 2.6 | 13.0-27.1 |
| 42 | 12-Jul | 00:52 | 56 | 2 | 0 | 117 | 8.3 | 106-145 | 16.8 | 3.9 | 13.2-30.5 |
| 43 | 12-Jul | 12:58 | 58 | 0 | 2 | 119 | 11.0 | 104-146 | 17.7 | 4.7 | 13.0-31.9 |
| 44 | 13-Jul | 01:05 | 54 | 1 | 1 | 118 | 11.9 | 105-143 | 18.2 | 5.1 | 13.2-31.6 |
| 45 | 13-Jul | 12:53 | 61 | 1 | 0 | 121 | 12.5 | 106-152 | 19.2 | 6.0 | 13.2-35.8 |
| 46 | 14-Jul | 01:05 | 59 | 0 | 0 | 121 | 11.3 | 107-145 | 19.5 | 5.1 | 13.8-30.3 |
| 47 | 14-Jul | 12:55 | 59 | 2 | 1 | 119 | 11.5 | 107-145 | 17.9 | 5.2 | 13.1-30.2 |
| 48 | 15-Jul | 00:45 | 79 | 1 | 1 | 115 | 9.8 | 105-150 | 16.3 | 4.5 | 13.0-36.3 |
| 49 | 15-Jul | 12:57 | 33 | 1 | 0 | 116 | 11.3 | 105-164 | 16.7 | 6.4 | 13.1-48.0 |
| 50 | 16-Jul | 01:23 | 58 | 3 | 1 | 117 | 9.8 | 106-146 | 17.1 | 4.5 | 13.0-32.7 |
| 51 | 16-Jul | 13:14 | 55 | 0 | 1 | 116 | 10.8 | 105-155 | 16.5 | 5.1 | 13.0-38.9 |

Table A1.13 (continued). Summary of subyearling Chinook salmon releases at Bonneville Dam tailrace during summer 2004. Dates, times, numbers of tagged fish released ( N ), 24 h post-tagging tag loss and mortality, and means, standard deviations (SD) and ranges for fork lengths and weights are presented. Release times are the start of releases and include fish released up to 1 hour later.

| Release | Date | Time | N | Tag loss | Mortality | Fork Length (mm) |  |  | Weight (g) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Mean | SD | Range | Mean | SD | Range |
| 52 | 17-Jul | 00:37 | 83 | 2 | 1 | 117 | 11.4 | 105-158 | 17.9 | 5.8 | 13.0-46.2 |
| 53 | 17-Jul | 13:05 | 82 | 0 | 0 | 117 | 9.0 | 106-152 | 17.8 | 4.2 | 13.2-38.9 |
| 54 | 18-Jul | 00:47 | 79 | 0 | 2 | 120 | 11.3 | 106-157 | 19.0 | 5.3 | 13.0-41.1 |
| 55 | 18-Jul | 13:00 | 71 | 0 | 1 | 116 | 9.2 | 105-150 | 17.2 | 4.1 | 13.1-33.0 |
| 56 | 19-Jul | 00:54 | 80 | 0 | 2 | 122 | 10.9 | 108-164 | 19.0 | 5.4 | 13.0-43.7 |
| 57 | 19-Jul | 12:58 | 78 | 1 | 1 | 123 | 14.7 | 105-172 | 20.6 | 8.3 | 13.3-57.4 |
| 58 | 20-Jul | 00:58 | 81 | 0 | 0 | 121 | 9.7 | 106-144 | 18.6 | 4.1 | 13.1-29.7 |
| 59 | 20-Jul | 13:05 | 80 | 1 | 0 | 121 | 11.3 | 105-157 | 20.4 | 6.4 | 13.0-44.5 |
| 60 | 21-Jul | 00:57 | 73 | 2 | 2 | 116 | 7.8 | 105-141 | 16.7 | 3.5 | 13.2-27.9 |
| 61 | 21-Jul | 12:55 | 77 | 0 | 1 | 117 | 9.9 | 105-155 | 17.1 | 4.7 | 13.0-35.5 |
| 62 | 22-Jul | 01:00 | 68 | 0 | 0 | 115 | 7.3 | 105-138 | 16.7 | 3.2 | 13.1-27.0 |
| Overall |  |  | 3795 | 75 | 28 | 118 | 9.2 | 102-172 | 17.2 | 4.4 | 13.0-57.4 |

## Appendix 2: Dead fish analysis

We generated survival estimates without releases that occurred 24 h before or after releases associated with the dead fish detections. The rationale for this approach is: since dead fish were not detected consistently throughout the season, then the conditions (i.e., discharge, water velocities, dam operations) present during releases where dead fish were detected (and conservatively $\pm 24 \mathrm{~h}$ of the release) may have been conducive for allowing the transport of dead radio-tagged fish to our detection sites below Bonneville. Thus, all fish passing through the study area during this time (i.e., $\pm 24 \mathrm{~h}$ of the release) were not included in the alternate analysis. In addition, to account for the possibility that our detections of radio-tagged fish with long travel times could be dead fish, we eliminated all fish with travel times $>$ the $99.7^{\text {th }}$ percentile and recalculated the estimates.

## Yearling Chinook salmon

## Paired Release-recapture Model

We generated survival estimates without releases that occurred 24 h before or after releases associated with the dead fish detections.

## Minimum Gap Runner Turbine Unit

## Control group released directly downstream of front roll below MGR turbine unit

## Survival Estimation with removal of releases within 24 h of detected dead fish

We removed releases 5, 6, 8, and 16 (Appendix 1, Table A1.2 and A1.3) and estimated that the average survival of yearling Chinook salmon into the MGR turbine unit 4A at Bonneville Dam's powerhouse 1 with the control group released in the tailrace below the MGR unit front roll was $0.952(\mathrm{SE}=0.012,95 \%$ confidence interval [0.907, 0.997].

Control group released below the outfall of the powerhouse $\mathbf{2}$ juvenile bypass outfall

## Survival estimation with removal of releases within $\mathbf{2 4} \mathbf{h}$ of detected dead fish

We removed releases 5, 6, 8, and 16 (Appendix 1, Table A1.2) and estimated that the average survival of yearling Chinook salmon into the MGR turbine unit 4A at Bonneville Dam's powerhouse 1 with the control group released in the tailrace below the powerhouse 2 juvenile bypass outfall was $0.995(\mathrm{SE}=0.020,95 \%$ confidence interval [0.951, 1.039].

## Powerhouse 1 Ice and Trash Sluiceway

## Survival estimation with removal of releases within $\mathbf{2 4} \mathbf{h}$ of detected dead fish

We removed releases 5, 6, 8, and 16 (Appendix 1, Table A1.1) and estimated that the average survival of yearling Chinook salmon released into the ice and trash sluiceway at

Bonneville Dam's powerhouse 1 (control release in the tailrace below the outfall of powerhouse 2 juvenile bypass outfall) was 1.004 ( $\mathrm{SE}=0.026,95 \%$ confidence interval $[0.947,1.061]$ ).

## Route-specific Survival Model

## Survival estimation with removal of releases within 24 h of detected dead fish

We also generated maximum likelihood estimates of the route-specific passage and survival probabilities through Bonneville Dam using capture histories (Table A2.1) with the removal of fish detected within 24 h of the detected radio-tagged dead fish (Figure A2.1). The survival of yearling Chinook salmon through Bonneville Dam spillway using this analysis method was estimated to be 0.910 ( $\mathrm{SE}=0.012$, profile likelihood $95 \%$ confidence interval [ $0.886,0.934]$ ). For yearling Chinook salmon passing via powerhouse 1 the estimated survival was 0.922 ( $\mathrm{SE}=0.021$, profile likelihood $95 \%$ confidence interval $[0.878,0.960]$ ) and for yearling Chinook salmon passing via powerhouse 2 turbines the estimated survival was 0.947 ( $\mathrm{SE}=0.012$, profile likelihood $95 \%$ confidence interval [ $0.922,0.970]$ ). For yearling Chinook salmon passing via the JBS the estimated survival was 0.968 ( $\mathrm{SE}=0.015$, profile likelihood $95 \%$ confidence interval $[0.937,0.995])$ and passing via the corner collector at powerhouse 2 the estimated survival was $1.019(\mathrm{SE}=0.009$, profile likelihood $95 \%$ confidence interval [1.001, 1.038]). Yearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.951 ( $\mathrm{SE}=0.009,95 \%$ confidence intervals $[0.933,0.968$ ].

## Survival estimation with removal of fish with travel times $>99.7^{\text {th }}$ percentile

We generated maximum likelihood estimates of the route-specific passage and survival probabilities through Bonneville Dam using capture histories (Table A2.2) with the removal of fish with travel times $>99.7^{\text {th }}$ percentile (Figure A2.2). The survival of yearling Chinook salmon through Bonneville Dam spillway using this analysis method was estimated to be 0.910 ( $\mathrm{SE}=$ 0.011 , profile likelihood $95 \%$ confidence interval [ $0.888,0.931$ ]). For yearling Chinook salmon passing via powerhouse 1 the estimated survival was 0.912 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval $[0.872,0.948]$ ) and for yearling Chinook salmon passing via powerhouse 2 turbines the estimated survival was 0.951 ( $\mathrm{SE}=0.011$, profile likelihood $95 \%$ confidence interval [0.929, 0.972]). For yearling Chinook salmon passing via the JBS the estimated survival was $0.970(\mathrm{SE}=0.013$, profile likelihood $95 \%$ confidence interval [0.943, 0.994]) and passing via the corner collector at powerhouse 2 the estimated survival was 1.016 ( $\mathrm{SE}=0.008$, profile likelihood $95 \%$ confidence interval [0.999, 1.032]). Yearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.951 ( $\mathrm{SE}=0.008,95 \%$ confidence intervals [0.936, 0.966].

Table A2.1. Counts of radio-tagged yearling Chinook salmon with the removal of fish within 24 h of detected dead fish for the releases from The Dalles Dam $\left(R_{I}\right)$ and in the tailrace of Bonneville Dam $\left(R_{2}\right)$ used in the route-specific survival model during 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $R_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B1 is powerhouse 1 and B2 is powerhouse 2 at Bonneville Dam.

| Release | Detection History | Route | Counts | Within-route histories Bonneville Dam |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 11 | 01 | 10 |
| $\mathrm{R}_{1}=3773$ | 100 |  | 274 |  |  |  |
|  | 101 |  | 59 |  |  |  |
|  | 110 | Spillway | 145 | 936 | 9 | 172 |
|  | 111 |  | 972 |  |  |  |
|  | 110 | B1 | 34 | 173 | 112 | 2 |
|  | 111 |  | 253 |  |  |  |
|  | 110 | B2 Turbines | 83 | 430 | 41 | 398 |
|  | 111 |  | 786 |  |  |  |
|  | 110 | B2 Juvenile bypass | 33 | 441 | 1 | 2 |
|  | 111 |  | 411 |  |  |  |
|  | 110 | B2 Corner collector | 18 | 722 | 0 | 1 |
|  | 111 |  | 705 |  |  |  |
| $\mathrm{R}_{2}=990$ | 010 |  | 43 |  |  |  |
|  | 011 |  | 947 |  |  |  |

Table A2.2. Counts of radio-tagged yearling Chinook salmon, with the removal of fish with travel times $>99.7^{\text {th }}$ percentile, from The Dalles Dam $\left(R_{l}\right)$ and in the tailrace of Bonneville Dam $\left(R_{2}\right)$ used in the route-specific survival model during 2004. Detection history recorded as: 1, detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B1 is powerhouse 1 and B 2 is powerhouse 2 at Bonneville Dam.

| Release | Detection History | Route | Counts | Within-route histories Bonneville Dam |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 11 | 01 | 10 |
| $\mathrm{R}_{1}=4481$ | 100 |  | 326 |  |  |  |
|  | 101 |  | 70 |  |  |  |
|  | 110 | Spillway | 172 | 1125 | 9 | 203 |
|  | 111 |  | 1165 |  |  |  |
|  | 110 | B1 | 43 | 214 | 121 | 5 |
|  | 111 |  | 297 |  |  |  |
|  | 110 | B2 Turbines | 91 | 511 | 50 | 448 |
|  | 111 |  | 918 |  |  |  |
|  | 110 | B2 Juvenile bypass | 37 | 516 | 2 | 2 |
|  | 111 |  | 483 |  |  |  |
|  | 110 | B2 Corner collector | 24 | 878 | 0 | 1 |
|  | 111 |  | 855 |  |  |  |
| $\mathrm{R}_{2}=1273$ | 010 |  | 54 |  |  |  |
|  | 011 |  | 1219 |  |  |  |



Figure A2.1. Schematic of estimated route-specific passage and survival parameters for yearling Chinook salmon through Bonneville Dam with the removal of fish with travel times > 99.7. Estimated standard errors are in parentheses.


Figure A2.2. Schematic of estimated route-specific passage and survival parameters for yearling Chinook salmon through Bonneville Dam with the removal of fish within 24 h of detected dead fish. Estimated standard errors are in parentheses.

## Subyearling Chinook salmon

## Paired Release-recapture Model

## Powerhouse 1 Ice and Trash Sluiceway

## Survival estimation with removal of releases within $\mathbf{2 4} \mathbf{h}$ of detected dead fish

We removed releases 15 and 16 (Appendix 1, Table A1.11) and estimated that the average survival of subyearling Chinook salmon released into the ice and trash sluiceway from 21 June to 22 July at Bonneville Dam's powerhouse 1 (control release in the tailrace below the outfall of powerhouse 2 JBS outfall) was $0.950(\mathrm{SE}=0.019,95 \%$ confidence interval [0.911, 0.989]).
$56 \mathrm{kcfs} /$ TDG spill operations with removal of fish within $\mathbf{2 4} \mathbf{h}$ of dead fish detections
We also generated maximum likelihood estimates of the route-specific passage and survival probabilities through Bonneville Dam using capture histories (Table A2.3) with the removal of fish detected within 24 h of the detected radio-tagged dead fish (Figure A2.3). The survival of subyearling Chinook salmon through Bonneville Dam spillway using this analysis method was estimated to be 0.887 ( $\mathrm{SE}=0.014$, profile likelihood $95 \%$ confidence interval [ $0.860,0.914]$ ). For subyearling Chinook salmon passing via powerhouse 1 the estimated survival was 0.837 ( $\mathrm{SE}=0.062$, profile likelihood $95 \%$ confidence interval [ $0.700,0.940]$ ) and for subyearling Chinook salmon passing via powerhouse 2 turbines the estimated survival was 0.835 ( $\mathrm{SE}=0.020$, profile likelihood $95 \%$ confidence interval [ $0.795,0.874]$ ). For subyearling Chinook salmon passing via the JBS the estimated survival was 0.925 ( $\mathrm{SE}=0.029$, profile likelihood $95 \%$ confidence interval $[0.861,0.976]$ ) and passing via the corner collector at powerhouse 2 the estimated survival was 0.976 ( $\mathrm{SE}=0.014$, profile likelihood $95 \%$ confidence interval [0.946, 1.003]). Subyearling Chinook salmon dam survival through Bonneville Dam was estimated to be 0.895 ( $\mathrm{SE}=0.011,95 \%$ confidence intervals [ $0.874,0.916$ ]. Since this analysis did not indicate any major changes in the survival estimates we did not generate survival estimates by day and night spill operations.

32 kcfs spill operations with removal of fish within 24 h of dead fish detections
We also generated maximum likelihood estimates of the route-specific passage and survival probabilities through Bonneville Dam using capture histories (Table A2.4) with the removal of fish detected within 24 h of the detected radio-tagged dead fish (Figure A2.4). The survival of subyearling Chinook salmon through Bonneville Dam spillway using this analysis method was estimated to be 0.744 ( $\mathrm{SE}=0.022$, profile likelihood $95 \%$ confidence interval [ $0.700,0.786]$ ). For subyearling Chinook salmon passing via powerhouse 1 the estimated survival was $0.829(\mathrm{SE}=0.030$, profile likelihood $95 \%$ confidence interval [ $0.767,0.884]$ ) and for subyearling Chinook salmon passing via powerhouse 2 turbines the estimated survival was 0.833 ( $\mathrm{SE}=0.014$, profile likelihood $95 \%$ confidence interval [ $0.805,0.859]$ ). For subyearling Chinook salmon passing via the JBS the estimated survival was 0.958 ( $\mathrm{SE}=0.019$, profile likelihood $95 \%$ confidence interval $[0.918,0.991]$ ) and passing via the corner collector at powerhouse 2 the estimated survival was 0.954 ( $\mathrm{SE}=0.013$, profile likelihood $95 \%$ confidence interval $[0.926,0.978])$. Subyearling Chinook salmon dam survival through Bonneville Dam was estimated to be $0.858(\mathrm{SE}=0.010,95 \%$ confidence intervals [0.838, 0.877]. Since this
analysis did not indicate any major changes in the survival estimates we did not generate survival estimates by day and night spill operations.

Table A2.3. Counts of radio-tagged subyearling Chinook salmon released from The Dalles Dam $\left(R_{l}\right)$ and the tailrace of Bonneville Dam $\left(R_{2}\right)$ during 56 kcfs day/TDG night spill operations with removal of fish within 24 h of deadfish detections used in the route-specific survival model during 2004. Detection history recorded as: 1 , detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only (01) within the passage route. B1 is powerhouse 1 and B 2 is powerhouse 2 at Bonneville Dam.

|  |  |  | Within-route histories |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Release | Detection History | Route | Counts | 11 | 01 | 10 |  |
| $\mathrm{R}_{1}=2440$ | 100 |  | 356 |  |  |  |  |
|  | 101 | 34 |  |  |  |  |  |
|  | 110 | Spillway | 149 | 772 | 8 | 151 |  |
|  | 111 |  | 782 |  |  |  |  |
|  | 110 | B1 | 10 | 37 | 11 | 0 |  |
|  | 111 |  | 38 |  |  |  |  |
|  | 110 | B2 Turbines | 100 | 243 | 53 | 182 |  |
|  | 111 |  | 378 |  |  |  |  |
|  | 110 | B2 Juvenile bypass | 18 | 142 | 1 | 2 |  |
|  | 111 |  | 127 |  |  |  |  |
|  | 110 | B2 Corner collector | 34 | 445 | 2 | 1 |  |
| $\mathrm{R}_{2}=1597$ | 111 |  | 414 |  |  |  |  |
|  | 010 |  | 85 |  |  |  |  |

Table A2.4. Counts of radio-tagged subyearling Chinook salmon released from The Dalles Dam $\left(R_{l}\right)$ and the tailrace of Bonneville Dam $\left(R_{2}\right)$ during 32 kcfs spill operations with removal of fish within 24 h of deadfish detections used in the route-specific survival model during 2004.
Detection history recorded as: 1 , detected; 0 , not detected. For $\mathrm{R}_{1}$, the first position indicates the release event, the second position indicates detection or not at Bonneville Dam, the third position indicates detection or not at at least one of the arrays below Bonneville Dam. For $\mathrm{R}_{2}$, the second position indicates the release event and the third position indicates detection or not at at least one of the arrays below Bonneville Dam. Within-route histories refer to whether a fish was detected on both antenna arrays (11), the first antenna array only (10), or the second antenna array only ( 01 ) within the passage route. B1 is powerhouse 1 and B 2 is powerhouse 2 at Bonneville Dam.

| Release | Detection History | Route | Counts | Within-route histories Bonneville Dam |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 11 | 01 | 10 |
| $\mathrm{R}_{1}=3252$ | 100 |  | 468 |  |  |  |
|  | 101 |  | 59 |  |  |  |
|  | 110 | Spillway | 146 | 410 | 15 | 71 |
|  | 111 |  | 350 |  |  |  |
|  | 110 | B1 | 46 | 148 | 67 | 1 |
|  | 111 |  | 169 |  |  |  |
|  | 110 | B2 Turbines | 225 | 504 | 93 | 478 |
|  | 111 |  | 850 |  |  |  |
|  | 110 | B2 Juvenile bypass | 27 | 289 | 1 | 3 |
|  | 111 |  | 266 |  |  |  |
|  | 110 | B2 Corner collector | 62 | 625 | 4 | 17 |
|  | 111 |  | 584 |  |  |  |
| $\mathrm{R}_{2}=1960$ | 010 |  | 102 |  |  |  |
|  | 011 |  | 1858 |  |  |  |



Figure A2.3. Schematic of estimated route-specific passage and survival parameters using an analysis with the removal of fish contacted within 24 h of detected dead fish, for subyearling Chinook salmon through Bonneville dam during 20 June through 22 July, at 56 kcfs day and total dissolved gas cap at night spill operations in 2004. Estimated standard errors are in parentheses.


Figure A2.4. Schematic of estimated route-specific passage and survival parameters using an analysis with the removal of fish contacted within 24 h of detected dead fish, for subyearling Chinook salmon through Bonneville dam during 20 June through 22 July, at 32 kcfs spill operations in 2004. Estimated standard errors are in parentheses.

# Appendix 3: Tag-Life Performance for Determining Potential Bias of Survival Estimates 

## Introduction

Survival estimates may be biased if the radio-tag expires prior to a fish exiting all the detection arrays. Radio-tags may expire before fish exit the study area due to a malfunction, extended travel time of fish during periods of low discharge, or extended length of time tag was on prior to release. Information obtained by a tag-life study can be used to adjust survival estimates using the probability that a tag will expire prior to fish exiting the study area (Townsend et al. 2004, Cowen and Schwarz 2005).

Several factors can affect the operational life of a radio-tag. For example, some tags lose a constant percentage (per unit time) of their battery life after the battery has been attached. Also, tag-life may be affected by water temperature and may vary among years or production batches. Thus, it is necessary to conduct the tag-life study concurrent with the survival study and under ambient conditions to emulate, as close as possible the source of the tags and the conditions they experience after they are released in fish.

To assess the probability of tag failure at detection arrays, a tag life study was performed. Our objectives were to: 1) estimate the probability a radio-tag was operational over time, 2) model the probability a radio-tag was operational, and 3 ) estimate the probability radio-tags were operational at detection arrays.

## Methods

The tag-life study entailed activating tags during spring and summer of 2004 at John Day Dam, and monitoring tag failure over time. A stratified random sub-sample of approximately the same number of tags from each frequency (channel) during early, middle, and late season for both spring ( $\mathrm{n}=65$ ) and summer $(\mathrm{n}=89)$ survival studies were taken. During the study, transmitters were set to emit a radio signal every 2 seconds and were held underwater at ambient water temperatures and monitored with a Lotek SRX-400 telemetry receiver. The receiver was programmed to scan all channels present for 15 s each hour with the gain set at zero. The receiver is checked daily to ensure that is working properly and the data is downloaded from the receiver at least once per week. The expiration time of each tag was noted at the time at which transmission ceased. Also, water temperature was recorded continuously at the study site with a recording thermograph. The Lotek Wireless Model 3 KM ( 7.3 mm in diameter x 18.0 mm in length and weighed 1.4 g in air) transmitters were used during the spring tag-life study and the Lotek model NTC 3-1 ( 6.3 mm wide $\times 14.5 \mathrm{~mm}$ length x 4.5 mm high and weigh 0.85 g in air) transmitters were used during the summer tag-life study corresponding to what was used for survival studies.

Our analytical approach was as per Townsend et al. (2004). Tag-life data was used to model tag survivorship and for calculating the probability of a tag being operational at detection arrays. The tag-life data was fit to a Gompertz distribution (Elandt-Johnson and Johnson 1980) for each season. A non-parametric form of the tag survival function was used because arrival times for radio-tagged salmonids had a non-normal distribution (Figure A3.1). This involved ranking tag-life data for calculating model parameters. Estimates for model parameters $\alpha$ and $\beta$
were generated for the tag survival function below and were used to calculate probabilities, where S is the probability the radio-tag is operational and t is time in days.

$$
\begin{equation*}
\left.S(t)=\mathrm{e}^{(\beta / \alpha)\left(1-\mathrm{e}^{\alpha} \mathrm{t}\right.}\right) \tag{1}
\end{equation*}
$$

Travel time to different detection arrays were then substituted into this function for estimating the probability a tag was operating when a fish arrived at a particular detection array. During our tagging procedures, tags were turned on prior to release ( $\approx 24$ hours), so the elapsed time a tag was operating before release was added to travel times.

## Results and Discussion

For spring, tag-failure was observed around 7-8 days and continued until day 12, at which all tags (model 3KM) were no longer operational. The average tag-life was estimated to be 9.87 days (Figure A3.2). For the summer tag-life study, the majority of radio-tags (model NTC 3-1) began to fail at day 7 and continued to day 12 averaging 8.96 days. Most tags were not operational by day 10 . There were two radio-tags for summer where transmission ceased around day 1 of the study at 25.44 and 27.36 hours. When these tags are excluded, which were probably defective, then the average tag-life was 9.14 days. For our tagging and release procedures, it is protocol to hold tagged fish at least 24 hours before release to reduce the possibility of releasing fish with defective tags. In 2004, tags were operating for about 30 hours prior to release at both dams for spring and summer. These tags would be recorded as not heard at time of release.

The tag-life studies for spring and summer were analyzed for generating model parameters of the Gompertz distribution and calculating probabilities that radio-tags were alive at detection arrays. Our tag-life data fit well with the Gompertz distribution for both the spring and summer tag-life studies allowing us to use this model for calculating probabilities (Figure A3.2, Table A3.1).

In our study, the probability a tag was operational at downstream arrays was high, with all probabilities greater than $99.9 \%$ (Table A3.2). Probabilities were higher for the summer study than for the spring study. The cumulative arrival distributions plotted with the Gompertz model over time shows that tagged juvenile salmonids passed through downstream detection arrays several days before tag-failure was substantial for both treatment and control fish at Bonneville Dam (Figure A3.3).

Townsend et al. (2004) found that the probability of a tag being operational at downstream detection arrays was quite high ( $>98 \%$ ), therefore, the adjusted survival estimate ( 0.9387 ) changed very little from the unadjusted estimate ( 0.9339 ) having a difference of just 0.0048 . Our probabilities being greater than this indicates our survival estimates would probably change even less after correction. Since the probability of a tag being operational at the downstream detection arrays for our survival studies were very close to one (Table A3.2), we did not adjust our survival estimates.

## References

Cowen, L. and C. J. Schwarz. 2005. Capture-recapture studies using radio telemetry with premature radio-tag failure. Biometrics 61:657-664.

Elandt-Johnson, R.C. and Johnson, N.L. 1980. Survival models and data analysis. Wiley, New York.

Townsend, R.L., J. R. Skalski, P. Dillingham, and T. W. Steig. 2004. Correcting Bias inSurvival Estimation Resulting from Tag Failure in Acoustic and Radiotelemetry Studies. Report prepared for the Bonneville Power Administration, Contract No. 00012494.

Table A3.1. Parameter estimates for tag-life using the Gompertz model during spring and

| Tag-life Study | N | $\alpha$ |  |  |  |  | $\beta$ | $\mathrm{R}^{2}$ | sum |
| :--- | :--- | :--- | :--- | :--- | :--- | :---: | :---: | :---: | :---: |
|  | mer |  |  |  |  |  |  |  |  |
| Suming | 65 | $1.0374(0.0259)$ | $2.600 \times 10^{-5}\left(5.995 \times 10^{-6}\right)$ | 0.9961 | duri |  |  |  |  |
| Summer | 89 | $1.6386(0.0256)$ | $3.405 \times 10^{-7}\left(7.59 \times 10^{-8}\right)$ | 0.9982 | ng |  |  |  |  |

2004, model estimate and (SE).

Table A3.2. Estimated probabilities (mean, SE in parentheses) a radio-tag was operational at Bonneville Dam and other downstream detection arrays for yearling Chinook salmon, hatchery steelhead trout, and subyearling Chinook salmon, during 2004.

| Release Site | Yearling Chinook salmon |  |
| :---: | :---: | :---: |
|  | Detection Array Locations |  |
|  | Bonneville Dam | Survival Gates |
| The Dalles Dam | 0.9996 (3.634×10 ${ }^{-6}$ ) | $0.9992\left(7.036 \times 10^{-5}\right)$ |
| Bonneville Dam | NA | $0.9999\left(3.474 \times 10^{-6}\right)$ |
|  | Hatchery steelhea |  |
| The Dalles Dam | $0.9996\left(2.152 \times 10^{-5}\right)$ | $0.9993\left(3.795 \times 10^{-5}\right)$ |
| Bonneville Dam | NA | $0.9998\left(1.122 \times 10^{-4}\right)$ |
|  | Subyearling Chinoo |  |
| The Dalles Dam | 0.9999 (1.604×10 ${ }^{-5}$ ) | $0.9999\left(2.030 \times 10^{-5}\right)$ |
| Bonneville Dam | NA | $1.0000\left(3.285 \times 10^{-7}\right)$ |



Figure A3.1. Arrival distributions of treatment fish for the, a) Bonneville Dam and b) The Dalles Dam survival assessment during summer of 2004 for subyearling Chinook salmon.


Figure A3.2. Fitted Gompertz model with tag-life data for a) spring and b) summer studies.
a) Yearling Chinook salmon, Spring


c) Subyearling Chinook salmon, Summer


Figure A3.3. Probability distributions (a-c) for radio-tags being operational over time with cumulative arrival distributions at downstream survival gates for the Bonneville Dam survival assessment during 2004.

## Appendix 4: Burnham Tests 2 and 3

Table A4.1. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for paired releases of yearling Chinook salmon, spring 2004. Treatment fish were released into the minimum gap runner (MGR) turbine 4A at Bonneville Dam powerhouse 1 and control fish were released directly downstream of the front roll below the (MGR) turbine unit at Bonneville Dam powerhouse 1.

| Release | Population | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 1 | Treatment | 1 | 0.00 | 0.95 | a | , | ${ }^{\text {a }}$ |
|  | Control | 1 | 0.05 | 0.83 | a | a | a |
| 2 | Treatment | 1 | 0.01 | 0.92 | a | a | a |
|  | Control | 1 | 1.47 | 0.23 | a | a | a |
| 3 | Treatment | 1 | 0.20 | 0.66 | a | a | a |
|  | Control | 1 | 0.54 | 0.46 | a | a | a |
| 4 | Treatment | 1 | 1.72 | 0.19 | 1 | 0.04 | 0.83 |
|  | Control | 1 | 2.08 | 0.15 | 1 | 0.77 | 0.38 |
| 5 | Treatment | 1 | 1.67 | 0.20 | 1 | 0.03 | 0.87 |
|  | Control | 1 | 0.32 | 0.57 | 1 | 0.98 | 0.32 |
| 6 | Treatment | 1 | 0.01 | 0.94 | 1 | 0.36 | 0.55 |
|  | Control | 1 | 0.34 | 0.56 | a | a | a |
| 7 | Treatment | 1 | 0.16 | 0.69 | a | a | a |
|  | Control | 1 | 0.07 | 0.79 | 1 | 0.00 | 0.95 |
| 8 | Treatment | 1 | 0.39 | 0.53 | 1 | 0.17 | 0.68 |
|  | Control | 1 | 0.35 | 0.56 | 1 | 0.41 | 0.52 |
| 9 | Treatment | 1 | 0.25 | 0.62 | a | a | , |
|  | Control | a | a | a | 1 | 0.00 | 1.00 |
| 10 | Treatment | b | b | b | b | , | b |
|  | Control | b | b | b | b | b | b |
| 11 | Treatment | 1 | 0.06 | 0.81 | 1 | 0.21 | 0.65 |
|  | Control | 1 | 0.29 | 0.59 | 1 | 0.00 | 0.96 |
| 12 | Treatment | 1 | 0.71 | 0.40 | a | a | a |
|  | Control | 1 | 0.39 | 0.53 | a | a | a |
| 13 | Treatment | 1 | 0.24 | 0.63 | a | a | a |
|  | Control | 1 | 3.25 | 0.07 | 1 | 0.06 | 0.81 |
| 14 | Treatment | 1 | 0.20 | 0.65 | 1 | 1.12 | 0.29 |
|  | Control | 1 | 0.10 | 0.75 | 1 | 0.22 | 0.64 |
| 15 | Treatment | 1 | 2.23 | 0.14 | 1 | 0.10 | 0.75 |
|  | Control | 1 | 0.02 | 0.90 | a | a | a |
| 16 | Treatment | b | b | b | b | b | b |
|  | Control | b | b | b | b | b | b |

[^5]Table A4.2. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for paired releases of yearling Chinook salmon, spring 2004. Treatment fish were released into the minimum gap runner turbine 4 A at Bonneville Dam powerhouse 1 and control fish were released below the Bonneville Dam powerhouse 2 juvenile bypass outfall.

| Release | Population | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | df | $\chi^{2}$ | P | Df | $\chi^{2}$ | P |
| 1 | Treatment | 1 | 0.00 | 0.95 | a | a | a |
|  | Control | 1 | 1.53 | 0.22 | a | a | a |
| 2 | Treatment | 1 | 0.01 | 0.92 | a | a | a |
|  | Control | 1 | 1.00 | 0.32 | a | a | a |
| 3 | Treatment | 1 | 0.20 | 0.66 | a | a | a |
|  | Control | 1 | 0.44 | 0.51 | 1 | 0.44 | 0.51 |
| 4 | Treatment | 1 | 1.72 | 0.19 | 1 | 0.04 | 0.83 |
|  | Control | 1 | 6.79 | 0.01 | a | a | a |
| 5 | Treatment | 1 | 1.67 | 0.20 | 1 | 0.03 | 0.87 |
|  | Control | 1 | 0.00 | 0.95 | a |  |  |
| 6 | Treatment | 1 | 0.01 | 0.94 | 1 | 0.36 | 0.55 |
|  | Control | 1 | 0.04 | 0.84 | 1 | 0.36 | 0.55 |
| 7 | Treatment | 1 | 0.16 | 0.69 | a | a | a |
|  | Control | 1 | 0.11 | 0.73 | 1 | 0.19 | 0.66 |
| 8 | Treatment | 1 | 0.39 | 0.53 | 1 | 0.17 | 0.68 |
|  | Control | 1 | 1.18 | 0.28 | 1 | 1.09 | 0.30 |
| 9 | Treatment | 1 | 0.25 | 0.62 | , | a | a |
|  | Control | 1 | 0.01 | 0.92 | 1 | 0.19 | 0.66 |
| 10 | Treatment | b | b | b | b | b | b |
|  | Control | b | b | b | b | b | b |
| 11 | Treatment | 1 | 0.06 | 0.81 | 1 | 0.21 | 0.65 |
|  | Control | 1 | 1.02 | 0.31 | 1 | 0.01 | 0.92 |
| 12 | Treatment | 1 | 0.71 | 0.40 | a | a | a |
|  | Control | 1 | 0.15 | 0.70 | a | a | a |
| 13 | Treatment | 1 | 0.24 | 0.63 | a | a | a |
|  | Control | 1 | 0.21 | 0.65 | a | a | a |
| 14 | Treatment | 1 | 0.20 | 0.65 | 1 | 1.12 | 0.29 |
|  | Control | 1 | 0.06 | 0.81 | a | a | a |
| 15 | Treatment | 1 | 2.23 | 0.14 | 1 | 0.10 | 0.75 |
|  | Control | 1 | 0.56 | 0.46 | 1 | 0.02 | 0.89 |
| 16 | Treatment | b | b | b | b | b | b |
|  | Control | b | b | b | b | b | b |

[^6]Table A4.3. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for paired releases of yearling Chinook salmon, spring 2004. Treatment fish were released at the top of the ice and trash sluiceway at Bonneville Dam powerhouse 1 and control fish were released below the Bonneville Dam powerhouse 2 juvenile bypass outfall.

| Release | Population | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | df | $\chi^{2}$ | P | Df | $\chi^{2}$ | P |
| 1 | Treatment | 1 | 0.59 | 0.44 | 1 | 0.10 | 0.76 |
|  | Control | 1 | 1.53 | 0.22 | a | a | a |
| 2 | Treatment | 1 | 0.44 | 0.51 | a | a | a |
|  | Control | 1 | 1.00 | 0.32 | a | a | a |
| 3 | Treatment | 1 | 1.06 | 0.30 | 1 | 0.10 | 0.75 |
|  | Control | 1 | 0.44 | 0.51 | 1 | 0.44 | 0.51 |
| 4 | Treatment | 1 | 0.04 | 0.84 | 1 | 0.10 | 0.76 |
|  | Control | 1 | 6.79 | 0.01 | a | a | a |
| 5 | Treatment | 1 | 0.13 | 0.72 | 1 | 0.00 | 0.96 |
|  | Control | 1 | 0.00 | 0.95 | a | a | a |
| 6 | Treatment | 1 | 0.37 | 0.54 | 1 | 0.04 | 0.85 |
|  | Control | 1 | 0.04 | 0.84 | 1 | 0.36 | 0.55 |
| 7 | Treatment | 1 | 0.00 | 0.98 | a | a | a |
|  | Control | 1 | 0.11 | 0.73 | 1 | 0.19 | 0.66 |
| 8 | Treatment | 1 | 0.02 | 0.90 | 1 | 0.17 | 0.68 |
|  | Control | 1 | 1.18 | 0.28 | 1 | 1.09 | 0.30 |
| 9 | Treatment | 1 | 0.42 | 0.52 | 1 | 0.03 | 0.87 |
|  | Control | 1 | 0.01 | 0.92 | 1 | 0.19 | 0.66 |
| 10 | Treatment | 1 | 0.33 | 0.57 | 1 | 0.04 | 0.83 |
|  | Control | 1 | 0.22 | 0.64 | 1 | 0.30 | 0.58 |
| 11 | Treatment | a | ${ }^{\text {a }}$ |  | a | , | a |
|  | Control | 1 | 1.02 | 0.31 | 1 | 0.01 | 0.92 |
| 12 | Treatment | 1 | 0.11 | 0.74 | a | a | a |
|  | Control | 1 | 0.15 | 0.70 | a | a | a |
| 13 | Treatment | 1 | 0.35 | 0.55 | 1 | 0.45 | 0.50 |
|  | Control | 1 | 0.21 | 0.65 | a | a | a |
| 14 | Treatment | 1 | 0.62 | 0.43 | 1 | 3.48 | 0.06 |
|  | Control | 1 | 0.06 | 0.81 | a | , | a |
| 15 | Treatment | 1 | 0.20 | 0.65 | 1 | 1.35 | 0.25 |
|  | Control | 1 | 0.56 | 0.46 | 1 | 0.02 | 0.89 |
| 16 | Treatment | 1 | 3.05 | 0.08 | 1 | 1.00 | 0.32 |
|  | Control | 1 | 0.82 | 0.37 | a | a | a |

${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.4. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for paired releases of hatchery steelhead trout, spring 2004. Treatment fish were released into the minimum gap runner (MGR) turbine 4A at Bonneville Dam powerhouse 1 and control fish were released directly downstream of the front roll below the (MGR) turbine unit at Bonneville Dam powerhouse 1.

| Release | Population | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 1 | Treatment | a | ${ }^{\text {a }}$ | a | a | , | a |
|  | Control | a | a | a | a | a | a |
| 2 | Treatment | a | a | a | a | a | a |
|  | Control | a | a | a | a | a | a |
| 3 | Treatment | 1 | 0.14 | 0.71 | a | a | a |
|  | Control | 1 | 0.04 | 0.84 | 1 | 0.68 | 0.41 |
| 4 | Treatment | 1 | 3.49 | 0.06 | 1 | 0.47 | 0.49 |
|  | Control | 1 | 0.13 | 0.71 | a | a | , |
| 5 | Treatment | 1 | 1.01 | 0.31 | 1 | 0.93 | 0.34 |
|  | Control | 1 | 1.34 | 0.25 | a | 0.28 | 0.60 |
| 6 | Treatment | a | a | a | a | a | a |
|  | Control | a | a | a | a | a | a |
| 7 | Treatment | 1 | 2.48 | 0.12 | a | a | a |
|  | Control | 1 | 0.19 | 0.66 | 1 | 00.0 | 0.96 |
| 8 | Treatment | b | b | b | b | b | b |
|  | Control | b | b | b | b | b | b |
| 9 | Treatment | 1 | 2.12 | 0.15 | 1 | 0.21 | 0.65 |
|  | Control | 1 | 2.73 | 0.10 | a | a | a |
| 10 | Treatment | a | a | a | a | a | a |
|  | Control | a | a | a | a | a | a |
| 11 | Treatment | 1 | 5.10 | 0.02 | 1 | 1.75 | 0.19 |
|  | Control | 1 | 1.87 | 0.17 | a | a | a |
| 12 | Treatment | a | a | a | a | a | a |
|  | Control | 1 | 3.99 | 0.05 | 1 | 1.50 | 0.22 |
| 13 | Treatment | 1 | 0.00 | 0.97 | 1 | 0.31 | 0.58 |
|  | Control | a | a | a | a | a | a |
| 14 | Treatment | b | b | b | b | b | b |
|  | Control | b | b | b | b | b | b |
| 15 | Treatment | 1 | 0.85 | 0.36 | a | a | a |
|  | Control | 1 | 0.00 | 0.96 | a | a | a |
| 16 | Treatment | a | a | a | a | a | a |
|  | Control | a | a | a | a | a | a |

[^7]Table A4.5. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for paired releases of hatchery steelhead trout, spring 2004. Treatment fish were released into the minimum gap runner turbine 4A at Bonneville Dam powerhouse 1 and control fish were released below the Bonneville Dam powerhouse 2 juvenile bypass outfall.

| Release | Population | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | df | $\chi^{2}$ | P | Df | $\chi^{2}$ | P |
| 1 | Treatment | a | , | a | a | a | a |
|  | Control | a | a | a | a | a | a |
| 2 | Treatment | a | a | a | a | a | a |
|  | Control | a | a | a | a | a | a |
| 3 | Treatment | 1 | 0.14 | 0.71 | a | a | a |
|  | Control | a | a | a | a | a | a |
| 4 | Treatment | 1 | 3.49 | 0.06 | 1 | 0.47 | 0.49 |
|  | Control | 1 | 0.83 | 0.36 | 1 | 0.77 | 0.38 |
| 5 | Treatment | 1 | 1.01 | 0.31 | 1 | 0.93 | 0.34 |
|  | Control | 1 | 2.00 | 0.16 | a | a | a |
| 6 | Treatment | a | a | a | a | a | a |
|  | Control | 1 | 1.98 | 0.16 | a | a | a |
| 7 | Treatment | 1 | 2.48 | 0.12 | a | a | a |
|  | Control | 1 | 0.01 | 0.93 | a | a | a |
| 8 | Treatment | b | b | b | b | b | b |
|  | Control | b | b | b | b | b | b |
| 9 | Treatment | 1 | 2.12 | 0.15 | 1 | 0.21 | 0.65 |
|  | Control | 1 | 0.77 | 0.38 | 1 | 1.18 | 0.28 |
| 10 | Treatment | a | a | a | a | a | a |
|  | Control | a | a | a | a | a | a |
| 11 | Treatment | 1 | 5.10 | 0.02 | 1 | 1.75 | 0.19 |
|  | Control | 1 | 2.12 | 0.15 | 1 | 2.00 | 0.16 |
| 12 | Treatment | a | a | a | a | a | a |
|  | Control | a | a | a | a | a | a |
| 13 | Treatment | 1 | 0.00 | 0.97 | 1 | 0.31 | 0.58 |
|  | Control | 1 | 0.17 | 0.68 | 1 | 0.13 | 0.72 |
| 14 | Treatment | b | b | b | , | b |  |
|  | Control | b | b | b | b | b | b |
| 15 | Treatment | 1 | 0.85 | 0.36 | a | a | a |
|  | Control | 1 | 0.30 | 0.58 | a | a | a |
| 16 | Treatment | a | a | a | a | a | a |
|  | Control | a | a | a | a | a | a |

[^8]Table A4.6. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for paired releases of hatchery steelhead trout, spring 2004. Treatment fish were released at the top of the sluiceway at Bonneville Dam powerhouse 1 and control fish were released below the Bonneville Dam powerhouse 2 juvenile bypass outfall.

| Release | Population | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | df | $\chi^{2}$ | P | Df | $\chi^{2}$ | P |
| 1 | Treatment | a | a | ${ }^{\text {a }}$ | a | a | a |
|  | Control | a | a | a | a | a | a |
| 2 | Treatment | a | a | a | a | a | a |
|  | Control | a | a | a | a | a | a |
| 3 | Treatment | 1 | 1.01 | 0.31 | 1 | 0.50 | 0.48 |
|  | Control | a | . | a | a | a | a |
| 4 | Treatment | 1 | 0.00 | 0.96 | 1 | 0.68 | 0.41 |
|  | Control | 1 | 0.83 | 0.36 | 1 | 0.77 | 0.38 |
| 5 | Treatment | 1 | 0.13 | 0.72 | a | a | a |
|  | Control | 1 | 2.00 | 0.16 | a | a | a |
| 6 | Treatment | 1 | - | , | a | a | a |
|  | Control | 1 | 1.98 | 0.16 | a | a | a |
| 7 | Treatment | 1 | 0.19 | 0.67 | a | a | a |
|  | Control | 1 | 0.01 | 0.93 | a | a | a |
| 8 | Treatment | a | , | , | a | a | a |
|  | Control | a | a | a | a | a | a |
| 9 | Treatment | 1 | 0.02 | 0.88 | 1 | 0.41 | 0.52 |
|  | Control | 1 | 0.77 | 0.38 | 1 | 1.18 | 0.28 |
| 10 | Treatment | a | a |  | a | a | a |
|  | Control | a | a | a | a | a | a |
| 11 | Treatment | 1 | 0.04 | 0.85 | 1 | 0.65 | 0.42 |
|  | Control | 1 | 2.12 | 0.15 | 1 | 2.00 | 0.16 |
| 12 | Treatment | 1 | 4.49 | 0.03 | a | a | a |
|  | Control | a | a | a | a | a | a |
| 13 | Treatment | a | ${ }^{\text {a }}$ | ${ }^{\text {a }}$ | 1 | 0.50 | 0.48 |
|  | Control | 1 | 0.17 | 0.68 | 1 | 0.13 | 0.72 |
| 14 | Treatment | a | a | a | , |  | a |
|  | Control | a | a | a | a | a | a |
| 15 | Treatment | 1 | 0.01 | 0.94 | a | a | a |
|  | Control | 1 | 0.30 | 0.58 | ${ }^{\text {a }}$ | a | ${ }^{\text {a }}$ |
| 16 | Treatment | a | a | a | ${ }^{\text {a }}$ | a | a |
|  | Control | a | a | a | a | a | a |

${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.7. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for each of 32 paired releases of subyearling Chinook salmon, summer 2004. Treatment fish were released at the ice and trash sluiceway of Bonneville Dam powerhouse 1 and control fish were released below the powerhouse 2 juvenile bypass outfall at Bonneville Dam.

| Release | Population | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 1 | Treatment | 1 | 1.37 | 0.24 | a | , | a |
|  | Control | 1 | 2.60 | 0.11 | 1 | 0.00 | 0.98 |
| 2 | Treatment | 1 | 0.13 | 0.71 | a | a | a |
|  | Control | 1 | 10.99 | 0.00 | 1 | 0.41 | 0.52 |
| 3 | Treatment | 1 | 0.05 | 0.83 | a | a | a |
|  | Control | 1 | 0.11 | 0.74 | 1 | 0.00 | 1.00 |
| 4 | Treatment | 1 | 0.83 | 0.36 | , | , | a |
|  | Control | 1 | 0.28 | 0.59 | 1 | 0.07 | 0.79 |
| 5 | Treatment | 1 | 0.17 | 0.68 | a | a | a |
|  | Control | 1 | 0.12 | 0.73 | 1 | 0.09 | 0.77 |
| 6 | Treatment | 1 | 0.57 | 0.45 | , | , |  |
|  | Control | 1 | 3.62 | 0.06 | 1 | 0.08 | 0.78 |
| 7 | Treatment | 1 | 0.10 | 0.75 | 1 | 0.00 | 1.00 |
|  | Control | 1 | 0.10 | 0.76 | 1 | 0.03 | 0.86 |
| 8 | Treatment | 1 | 0.06 | 0.81 | , | - |  |
|  | Control | 1 | 2.16 | 0.14 | 1 | 0.31 | 0.58 |
| 9 | Treatment | 1 | 0.36 | 0.55 | 1 | 0.21 | 0.65 |
|  | Control | 1 | 0.01 | 0.93 |  | a | a |
| 10 | Treatment | 1 | 0.23 | 0.63 | a | a | a |
|  | Control | 1 | 0.10 | 0.75 | a | a | a |
| 11 | Treatment | 1 | 0.03 | 0.85 | a | a | a |
|  | Control | 1 | 2.34 | 0.13 | 1 | 0.07 | 0.79 |
| 12 | Treatment | 1 | 0.60 | 0.44 | , | , |  |
|  | Control | 1 | 0.04 | 0.83 | 1 | 2.81 | 0.09 |
| 13 | Treatment | 1 | 0.02 | 0.88 | 1 | 0.06 | 0.81 |
|  | Control | 1 | 1.24 | 0.26 | 1 | 0.00 | 0.98 |
| 14 | Treatment | 1 | 0.11 | 0.74 | , | , | a |
|  | Control | 1 | 0.35 | 0.55 | 1 | 1.96 | 0.16 |
| 15 | Treatment | 1 | 0.00 | 1.00 | a | a | a |
|  | Control | 1 | 0.01 | 0.93 | 1 | 0.07 | 0.79 |
| 16 | Treatment | 1 | 0.37 | 0.55 | 1 | 0.04 | 0.85 |
|  | Control | 1 | 0.08 | 0.78 | 1 | 0.04 | 0.83 |

$\overline{{ }^{a}}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.7 (continued). Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for each of 32 paired releases of subyearling Chinook salmon, summer 2004. Treatment fish were released at the ice and trash sluiceway of Bonneville Dam powerhouse 1 and control fish were released below the powerhouse 2 juvenile bypass outfall at Bonneville Dam.

| Release | Population | $\underline{\text { Test } 2}$ |  |  | $\underline{\text { Test } 3}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 17 | Treatment | 1 | 0.58 | 0.45 | ${ }^{\text {a }}$ | , | a |
|  | Control | 1 | 1.00 | 0.32 | a | a | a |
| 18 | Treatment | 1 | 0.00 | 1.00 | 1 | 0.52 | 0.47 |
|  | Control | 1 | 0.11 | 0.75 | 1 | 0.55 | 0.46 |
| 19 | Treatment | 1 | 0.06 | 0.81 | 1 | 0.07 | 0.78 |
|  | Control | 1 | 10.42 | 0.00 | 1 | 0.06 | 0.81 |
| 20 | Treatment | 1 | 0.07 | 0.80 | 1 | 0.00 | 1.00 |
|  | Control | 1 | 0.01 | 0.94 | 1 | 0.00 | 0.97 |
| 21 | Treatment | 1 | 0.03 | 0.86 | 1 | 0.18 | 0.67 |
|  | Control | 1 | 0.02 | 0.88 | 1 | 0.01 | 0.93 |
| 22 | Treatment | 1 | 0.00 | 0.98 | a | a | , |
|  | Control | 1 | 0.03 | 0.86 | 1 | 0.07 | 0.78 |
| 23 | Treatment | 1 | 0.01 | 0.91 | 1 | 0.01 | 0.92 |
|  | Control | 1 | 0.19 | 0.67 | 1 | 0.52 | 0.47 |
| 24 | Treatment | 1 | 0.02 | 0.90 | a | a | a |
|  | Control | 1 | 6.13 | 0.01 | 1 | 0.02 | 0.88 |
| 25 | Treatment | a | a | a | 1 | 0.07 | 0.79 |
|  | Control | 1 | 0.01 | 0.93 | a | a | a |
| 26 | Treatment | 1 | 0.02 | 0.90 | 1 | 0.05 | 0.83 |
|  | Control | 1 | 0.42 | 0.51 | 1 | 0.86 | 0.36 |
| 27 | Treatment | 1 | 0.00 | 1.00 | a | a | a |
|  | Control | 1 | 1.20 | 0.27 | 1 | 0.10 | 0.75 |
| 28 | Treatment | 1 | 0.10 | 0.91 | 1 | 0.03 | 0.85 |
|  | Control | 1 | 0.24 | 0.62 | 1 | 0.13 | 0.72 |
| 29 | Treatment | 1 | 0.60 | 0.44 | a | a | a |
|  | Control | 1 | 0.00 | 0.98 | 1 | 0.03 | 0.85 |
| 30 | Treatment | 1 | 0.15 | 0.70 | a | a | a |
|  | Control | 1 | 0.13 | 0.72 | 1 | 0.34 | 0.56 |
| 31 | Treatment | 1 | 0.00 | 0.94 | 1 | 0.00 | 1.00 |
|  | Control | 1 | 0.63 | 0.43 | 1 | 0.42 | 0.52 |
| 32 | Treatment | 1 | 0.94 | 0.33 | 1 | 0.26 | 0.61 |
|  | Control | 1 | 8.99 | 0.00 | 1 | 0.11 | 0.75 |

${ }^{\bar{a}}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.8. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of yearling Chinook salmon through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, spring 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam.

| Release | Deflector |  |  | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Ht | Population | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| Releas | Day | 7 | Treatment | 1 | 0.36 | 0.55 | a | ${ }^{\text {a }}$ | a |
|  |  |  | Control | 1 | 0.17 | 0.68 | a | a | a |
|  |  | 14 | Treatment | 1 | 0.00 | 1.00 | a | a | a |
|  |  |  | Control | 1 | 0.17 | 0.68 | a | a | a |
|  | Night | 7 | Treatment | 1 | 0.00 | 0.98 | 1 | 0.06 | 0.81 |
|  |  |  | Control | 1 | 0.17 | 0.68 | a | a | a |
|  |  | 14 | Treatment | 1 | 0.17 | 0.68 | a | a | a |
|  |  |  | Control | 1 | 0.17 | 0.68 | a | a | a |
| 2 | Day | 7 | Treatment | 1 | 0.06 | 0.81 | 1 | 0.03 | 0.87 |
|  |  |  | Control | 1 | 1.23 | 0.27 | 1 | 0.00 | 0.99 |
|  |  | 14 | Treatment | 1 | 0.02 | 0.88 | , | a | - |
|  |  |  | Control | 1 | 1.23 | 0.27 | 1 | 0.00 | 0.99 |
|  | Night | 7 | Treatment | 1 | 0.07 | 0.79 | , | - | , |
|  |  |  | Control | 1 | 1.23 | 0.27 | 1 | 0.00 | 0.99 |
|  |  | 14 | Treatment | 1 | 0.51 | 0.47 | - | , | , |
|  |  |  | Control | 1 | 1.23 | 0.27 | 1 | 0.00 | 0.99 |
| 3 | Day | 7 | Treatment | 1 | 1.41 | 0.23 | 1 | 0.01 | 0.93 |
|  |  |  | Control | 1 | 0.00 | 0.96 | 1 | 0.00 | 0.96 |
|  |  | 14 | Treatment | 1 | 0.00 | 1.00 | , | , |  |
|  |  |  | Control | 1 | 0.00 | 0.96 | 1 | 0.00 | 0.96 |
|  | Night | 7 | Treatment | 1 | 0.44 | 0.51 | 1 | 0.09 | 0.77 |
|  |  |  | Control | 1 | 0.00 | 0.96 | 1 | 0.00 | 0.96 |
|  |  | 14 | Treatment | 1 | 0.04 | 0.84 | a | - | , |
|  |  |  | Control | 1 | 0.00 | 0.96 | 1 | 0.00 | 0.96 |
| 4 | Day | 7 | Treatment | 1 | 0.01 | 0.93 | 1 | 0.28 | 0.60 |
|  |  |  | Control | 1 | 7.87 | 0.01 | 1 | 0.09 | 0.77 |
|  |  | 14 | Treatment | 1 | 0.23 | 0.63 | a | a | a |
|  |  |  | Control | 1 | 7.87 | 0.01 | 1 | 0.09 | 0.77 |
|  | Night | 7 | Treatment | 1 | 0.27 | 0.61 | 1 | 0.00 | 1.00 |
|  |  |  | Control | 1 | 7.87 | 0.01 | 1 | 0.09 | 0.77 |
|  |  | 14 | Treatment | 1 | 0.22 | 0.64 | , | , |  |
|  |  |  | Control | 1 | 7.87 | 0.01 | 1 | 0.09 | 0.77 |
| 5 | Day | 7 | Treatment | 1 | 0.03 | 0.85 | 1 | 0.02 | 0.88 |
|  |  |  | Control | 1 | 0.26 | 0.61 | 1 | 0.13 | 0.72 |
|  |  | 14 | Treatment | 1 | 0.06 | 0.81 | a |  |  |

${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.8 (continued). Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of yearling Chinook salmon through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, spring 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam.

| Release | Deflector |  |  | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Ht | Population | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 6 | Day | 14 | Control | 1 | 0.26 | 0.61 | 1 | 0.13 | 0.72 |
|  | Night | 7 | Treatment | 1 | 0.29 | 0.59 | a | a | a |
|  |  |  | Control | 1 | 0.26 | 0.61 | 1 | 0.13 | 0.72 |
|  |  | 14 | Treatment | 1 | 2.10 | 0.15 | 1 | 0.00 | 0.96 |
|  |  |  | Control | 1 | 0.26 | 0.61 | 1 | 0.13 | 0.72 |
|  | Day | 7 | Treatment | 1 | 0.39 | 0.53 | 1 | 0.25 | 0.62 |
|  |  |  | Control | 1 | 0.03 | 0.87 | 1 | 0.37 | 0.54 |
|  |  | 14 | Treatment | a | a | a | a | a | a |
| 7 |  |  | Control | 1 | 0.03 | 0.87 | 1 | 0.37 | 0.54 |
|  | Night | 7 | Treatment | 1 | 0.01 | 0.92 | a |  | a |
|  |  |  | Control | 1 | 0.03 | 0.87 | 1 | 0.37 | 0.54 |
|  |  | 14 | Treatment | 1 | 2.31 | 0.13 |  |  |  |
|  |  |  | Control | 1 | 0.03 | 0.87 | 1 | 0.37 | 0.54 |
|  | Day | 7 | Treatment | 1 | 0.79 | 0.37 | a | a |  |
|  |  |  | Control | 1 | 1.00 | 0.32 | 1 | 2.77 | 0.10 |
|  |  | 14 | Treatment | 1 | 0.33 | 0.57 | 1 | 0.02 | 0.88 |
|  |  |  | Control | 1 | 1.00 | 0.32 | 1 | 2.77 | 0.10 |
| 8 | Night | 7 | Treatment | 1 | 0.00 | 0.97 | 1 | 0.14 | 0.71 |
|  |  |  | Control | 1 | 1.00 | 0.32 | 1 | 2.77 | 0.10 |
|  |  | 14 | Treatment | 1 | 0.12 | 0.73 | a | a | , |
|  |  |  | Control | 1 | 1.00 | 0.32 | , | 2.77 | 0.10 |
|  | Day | 7 | Treatment | 1 | 0.22 | 0.64 | a | a | a |
|  |  |  | Control | 1 | 0.51 | 0.47 | 1 | 0.30 | 0.58 |
|  |  | 14 | Treatment | 1 | 0.38 | 0.54 | a | a | a |
|  |  |  | Control | 1 | 0.51 | 0.47 | 1 | 0.30 | 0.58 |
| 9 | Night | 7 | Treatment | 1 | 4.70 | 0.03 | 1 | 0.36 | 0.55 |
|  |  |  | Control | 1 | 0.51 | 0.47 | 1 | 0.30 | 0.58 |
|  |  | 14 | Treatment | 1 | 1.44 | 0.23 | 1 | 0.26 | 0.61 |
|  |  |  | Control | 1 | 0.51 | 0.47 | 1 | 0.30 | 0.58 |
|  | Day | 7 | Treatment | 1 | 1.59 | 0.21 | a |  | , |
|  |  |  | Control | 1 | 6.65 | 0.01 | 1 | 0.12 | 0.73 |
|  |  | 14 | Treatment | 1 | 0.60 | 0.44 | a |  |  |
|  |  |  | Control | 1 | 6.65 | 0.01 | 1 | 0.12 | 0.73 |
|  | Night | 7 | Treatment | 1 | 0.01 | 0.93 | a |  | , |
|  |  |  | Control | , | 6.65 | 0.01 | 1 | 0.12 | 0.73 |
|  |  | 14 | Treatment | 1 | 0.02 | 0.90 | , | . | a |
|  |  |  | Control | 1 | 6.65 | 0.01 | 1 | 0.12 | 0.73 |

[^9]Table A4.8 (continued). Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of yearling Chinook salmon through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, spring 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam.

| Release | Deflector |  |  | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Ht | Population | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 10 | Day | 7 | Treatment | 1 | 0.01 | 0.92 | a | a | a |
|  |  |  | Control | 1 | 1.49 | 0.22 | 1 | 0.01 | 0.91 |
|  |  | 14 | Treatment | 1 | 0.31 | 0.58 | , | a | a |
|  |  |  | Control | 1 | 1.49 | 0.22 | 1 | 0.01 | 0.91 |
|  | Night | 7 | Treatment | 1 | 0.10 | 0.76 | a | a |  |
|  |  |  | Control | 1 | 1.49 | 0.22 | 1 | 0.01 | 0.91 |
|  |  | 14 | Treatment | 1 | 1.89 | 0.17 | a | a | , |
|  |  |  | Control | 1 | 1.49 | 0.22 | 1 | 0.01 | 0.91 |
| 11 | Day | 7 | Treatment | 1 | 4.43 | 0.04 | , | , |  |
|  |  |  | Control | 1 | 3.93 | 0.05 | a | a | a |
|  |  | 14 | Treatment | 1 | 0.27 | 0.61 | 1 | 0.08 | 0.78 |
|  |  |  | Control | 1 | 3.93 | 0.05 | a | a |  |
|  | Night | 7 | Treatment | 1 | 0.09 | 0.77 | a | a | a |
|  |  |  | Control | 1 | 3.93 | 0.05 | a | a | a |
|  |  | 14 | Treatment | 1 | 0.16 | 0.69 | a | a | a |
|  |  |  | Control | 1 | 3.93 | 0.05 | a | a | a |
| 12 | Day | 7 | Treatment | 1 | 0.33 | 0.56 | a | a | a |
|  |  |  | Control | 1 | 0.01 | 0.94 | 1 | 0.20 | 0.66 |
|  |  | 14 | Treatment | a | a | ${ }^{\text {a }}$ | a | a | a |
|  |  |  | Control | 1 | 0.01 | 0.94 | 1 | 0.20 | 0.66 |
|  | Night | 7 | Treatment | 1 | 0.17 | 0.68 | 1 | 0.75 | 0.39 |
|  |  |  | Control | 1 | 0.01 | 0.94 | 1 | 0.20 | 0.66 |
|  |  | 14 | Treatment | 1 | 0.01 | 0.94 | , | , |  |
|  |  |  | Control | 1 | 0.01 | 0.94 | 1 | 0.20 | 0.66 |
| 13 | Day | 7 | Treatment | 1 | 0.23 | 0.63 | 1 | 0.00 | 0.97 |
|  |  |  | Control | 1 | 7.17 | 0.01 | 1 | 0.00 | 0.97 |
|  |  | 14 | Treatment | 1 | 0.04 | 0.84 | 1 | 0.02 | 0.88 |
|  |  |  | Control | 1 | 7.17 | 0.01 | 1 | 0.00 | 0.97 |
|  | Night | 7 | Treatment | 1 | 0.51 | 0.47 | ${ }^{\text {a }}$ | ${ }^{\text {a }}$ | ${ }^{\text {a }}$ |
|  |  |  | Control | 1 | 7.17 | 0.01 | 1 | 0.00 | 0.97 |
|  |  | 14 | Treatment | 1 | 2.89 | 0.09 | a | a | a |
|  |  |  | Control | 1 | 7.17 | 0.01 | 1 | 0.00 | 0.97 |

${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.9. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of hatchery steelhead trout through spillbays with deflectors at 7 -ft and 14-ft, spring 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam.

| Release | Deflector |  |  | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Ht | Population | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 1 | Day | 7 | Treatment | 1 | 0.05 | 0.82 | ${ }^{\text {a }}$ | a | a |
|  |  |  | Control | 1 | 0.04 | 0.84 | a | a | a |
|  |  | 14 | Treatment | a | , | a | a | a | a |
|  |  |  | Control | 1 | 0.04 | 0.84 | a | a | a |
|  | Night | 7 | Treatment | a | a | a | a | a | a |
|  |  |  | Control | 1 | 0.04 | 0.84 | a | a | a |
|  |  | 14 | Treatment | 1 | 0.85 | 0.36 | a | a | a |
|  |  |  | Control | 1 | 0.04 | 0.84 | a | a | a |
| 2 | Day | 7 | Treatment | 1 | 0.00 | 1.00 | a | a | a |
|  |  |  | Control | a | a | a | a | a | a |
|  |  | 14 | Treatment | a | a | a | a | a | a |
|  |  |  | Control | a | a | a | a | a | a |
|  | Night | 7 | Treatment | 1 | $1.87$ | 0.17 | 1 | 1.75 | 0.19 |
|  |  |  | Control | a | a | a | , | a | a |
|  |  | 14 | Treatment | a | a | a | a | a | a |
|  |  |  | Control | a | a | a | a | a | a |
| 3 | Day | 7 | Treatment | 1 | 2.13 | 0.14 | a | a | a |
|  |  |  | Control | 1 | 5.00 | 0.03 | a | a | a |
|  |  | 14 | Treatment | 1 | 0.00 | 1.00 | a | a | a |
|  |  |  | Control | 1 | 5.00 | 0.03 | a | a | a |
|  | Night | 7 | Treatment | a | , | a | a | a | a |
|  |  |  | Control | 1 | 5.00 | 0.03 | a | a | a |
|  |  | 14 | Treatment | 1 | 0.08 | 0.77 | a | a | a |
|  |  |  | Control | 1 | 5.00 | 0.03 | a | a | a |
| 4 | Day | 7 | Treatment | 1 | 0.00 | 1.00 | a | a | a |
|  |  |  | Control | 1 | 0.20 | 0.65 | 1 | 1.36 | 0.24 |
|  |  | 14 | Treatment | a | ${ }^{\text {a }}$ | ${ }^{\text {a }}$ | a | , |  |
|  |  |  | Control | 1 | 0.20 | 0.65 | 1 | 1.36 | 0.24 |
|  | Night | 7 | Treatment |  |  | a | a | , | , |
|  |  |  | Control | 1 | 0.20 | 0.65 | 1 | 1.36 | 0.24 |
|  |  | 14 | Treatment | a | , |  | a | , | a |
|  |  |  | Control | 1 | 0.20 | 0.65 | 1 | 1.36 | 0.24 |
| 5 | Day | 7 | Treatment | 1 | 0.49 | 0.48 | 1 | 0.97 | 0.32 |

${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.9 (continued). Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of hatchery steelhead trout through spillbays with deflectors at 7 - ft and $14-\mathrm{ft}$, spring 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam.

${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.9 (continued). Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of hatchery steelhead trout through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, spring 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam.

| Release | Deflector |  |  | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Ht | Population | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 10 |  | 14 | Treatment | 1 | 0.19 | 0.67 | a | ${ }^{\text {a }}$ | a |
|  |  |  | Control | 1 | 0.11 | 0.74 | 1 | 0.13 | 0.72 |
|  | Night | 7 | Treatment | a | a | a | a | , | a |
|  |  |  | Control | 1 | 0.11 | 0.74 | 1 | 0.13 | 0.72 |
|  |  | 14 | Treatment | 1 | 0.79 | 0.37 | a |  | a |
|  |  |  | Control | 1 | 0.11 | 0.74 | 1 | 0.13 | 0.72 |
|  | Day | 7 | Treatment | 1 | 1.86 | 0.17 | a | , | , |
|  |  |  | Control | 1 | 9.50 | 0.00 | 1 | 3.93 | 0.05 |
|  |  | 14 | Treatment | 1 | 0.63 | 0.43 | , | , |  |
|  |  |  | Control | 1 | 9.50 | 0.00 | 1 | 3.93 | 0.05 |
|  | Night | 7 | Treatment | 1 | 0.13 | 0.72 | a | , | , |
| 11 |  |  | Control | 1 | 9.50 | 0.00 | 1 | 3.93 | 0.05 |
|  |  | 14 | Treatment | 1 | 0.96 | 0.33 | 1 | 1.42 | 0.23 |
|  |  |  | Control | 1 | 9.50 | 0.00 | 1 | 3.93 | 0.05 |
|  | Day | 7 | Treatment | 1 | 0.39 | 0.53 | - |  |  |
|  |  |  | Control | 1 | 0.02 | 0.89 | 1 | 0.10 | 0.75 |
|  |  | 14 | Treatment | a | , |  | , | , | , |
|  |  |  | Control | 1 | 0.02 | 0.89 | 1 | 0.10 | 0.75 |
| 12 | Night | 7 | Treatment | a | ${ }^{\text {a }}$ |  | 1 | 4.24 | 0.04 |
|  |  |  | Control | 1 | 0.02 | 0.89 | 1 | 0.10 | 0.75 |
|  |  | 14 | Treatment | 1 | 1.39 | 0.24 | 1 | 6.43 | 0.01 |
|  |  |  | Control | 1 | 0.02 | 0.89 | 1 | 0.10 | 0.75 |
|  | Day | 7 | Treatment | a | ${ }^{\text {a }}$ | a | , | , | , |
|  |  |  | Control | 1 | 0.01 | 0.91 | 1 | 0.03 | 0.87 |
|  |  | 14 | Treatment | 1 | 0.00 | 1.00 | a | a | a |
|  |  |  | Control | 1 | 0.01 | 0.91 | 1 | 0.03 | 0.87 |
| 13 | Night | 7 | Treatment | 1 | 1.34 | 0.25 | 1 | 2.25 | 0.13 |
|  |  |  | Control | 1 | 0.01 | 0.91 | 1 | 0.03 | 0.87 |
|  |  | 14 | Treatment | 1 | 0.38 | 0.54 | a | , | ${ }^{\text {a }}$ |
|  |  |  | Control | 1 | 0.01 | 0.91 | 1 | 0.03 | 0.87 |
|  | Day | 7 | Treatment | 1 | 0.14 | 0.71 | a | ${ }^{\text {a }}$ | a |
|  |  |  | Control | 1 | 0.76 | 0.38 | 1 | 2.76 | 0.10 |
|  |  | 14 | Treatment | 1 | 0.11 | 0.74 | 1 | 0.19 | 0.67 |
|  |  |  | Control | 1 | 0.76 | 0.38 | 1 | 2.76 | 0.10 |

[^10]Table A4.9 (continued). Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of hatchery steelhead trout through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, spring 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam.

| Release | Deflector |  |  | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Ht | Population | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 14 | Night | 7 | Treatment | a | a | a | a | x | a |
|  |  |  | Control | 1 | 0.76 | 0.38 | 1 | 2.76 | 0.10 |
|  |  | 14 | Treatment | 1 | 2.09 | 0.15 | 1 | 2.01 | 0.16 |
|  |  |  | Control | 1 | 0.76 | 0.38 | 1 | 2.76 | 0.10 |
|  | Day | 7 | Treatment | 1 | 0.19 | 0.67 | , | a | a |
|  |  |  | Control | 1 | 0.65 | 0.42 | a | a | a |
|  |  | 14 | Treatment | 1 | 0.00 | 1.00 | a | a | a |
|  |  |  | Control | 1 | 0.65 | 0.42 | a | a | a |
|  | Night | 7 | Treatment | , | a | , | a | a | a |
|  |  |  | Control | 1 | 0.65 | 0.42 | a | a | a |
|  |  | 14 | Treatment | 1 | 0.11 | 0.74 | a | a | a |
|  |  |  | Control | 1 | 0.65 | 0.42 | a | a | a |
| 15 | Day | 7 | Treatment | 1 | 0.22 | 0.64 | a | a | a |
|  |  |  | Control | 1 | 0.34 | 0.56 | 1 | 0.47 | 0.49 |
|  |  | 14 | Treatment | 1 | 0.26 | 0.61 | , | , | , |
|  |  |  | Control | 1 | 0.34 | 0.56 | 1 | 0.47 | 0.49 |
|  | Night | 7 | Treatment | 1 | 0.00 | 1.00 | , | , | a |
|  |  |  | Control | 1 | 0.34 | 0.56 | 1 | 0.47 | 0.49 |
|  |  | 14 | Treatment | a | ${ }^{\text {a }}$ | ${ }^{\text {a }}$ | 1 | 1.20 | 0.27 |
|  |  |  | Control | 1 | 0.34 | 0.56 | , | 0.47 | 0.49 |

${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.10. Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of subyearling Chinook through spillbays with deflectors at 7ft and $14-\mathrm{ft}$, summer 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam. BIOP spill $=56 \mathrm{kcfs}$ day $/$ total dissolved gas cap at night.

| Release | Deflector |  |  | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spill | Ht | Population | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 1 | BIOP | 7 | Treatment | 1 | 0.51 | 0.48 | 1 | 0.07 | 0.79 |
|  |  |  | Control | 1 | 12.66 | 0.00 | 1 | 0.13 | 0.72 |
|  |  | 14 | Treatment | 1 | 0.00 | 0.97 | a | a | a |
|  |  |  | Control | 1 | 12.66 | 0.00 | 1 | 0.13 | 0.72 |
| 2 | KCFS50 | 7 | Treatment | 1 | 0.27 | 0.60 | 1 | 0.05 | 0.83 |
|  |  |  | Control | 1 | 11.96 | 0.00 | 1 | 0.45 | 0.50 |
|  |  | 14 | Treatment | 1 | 0.01 | 0.94 | a | a | a |
|  |  |  | Control | 1 | 11.96 | 0.00 | 1 | 0.45 | 0.50 |
| 3 | KCFS50 | 7 | Treatment | 1 | 0.25 | 0.62 | 1 | 1.33 | 0.25 |
|  |  |  | Control | 1 | 0.27 | 0.60 | 1 | 0.45 | 0.50 |
|  |  | 14 | Treatment | 1 | 0.05 | 0.83 | a | a | a |
|  |  |  | Control | 1 | 0.27 | 0.60 | 1 | 0.45 | 0.50 |
| 4 | KCFS50 | 7 | Treatment | 1 | 0.00 | 0.96 | a | a | a |
|  |  |  | Control | 1 | 6.13 | 0.01 | 1 | 0.31 | 0.58 |
|  |  | 14 | Treatment | 1 | 0.36 | 0.55 | 1 | 0.00 | 1.00 |
|  |  |  | Control | 1 | 6.13 | 0.01 | 1 | 0.31 | 0.58 |
| 5 | BIOP | 7 | Treatment | 1 | 2.94 | 0.09 | a | a | a |
|  |  |  | Control | 1 | 4.56 | 0.03 | 1 | 0.09 | 0.77 |
|  |  | 14 | Treatment | 1 | 0.18 | 0.67 | a | a | a |
|  |  |  | Control | 1 | 4.56 | 0.03 | 1 | 0.09 | 0.77 |
| 6 | BIOP | 7 | Treatment | 1 | 0.18 | 0.67 | a | a | a |
|  |  |  | Control | 1 | 1.32 | 0.25 | 1 | 0.10 | 0.75 |
|  |  | 14 | Treatment | 1 | 0.04 | 0.85 | a | a | a |
|  |  |  | Control | 1 | 1.32 | 0.25 | 1 | 0.10 | 0.75 |
| 7 | KCFS50 | 7 | Treatment | 1 | 0.21 | 0.65 | 1 | 0.14 | 0.71 |
|  |  |  | Control | 1 | 1.15 | 0.28 | 1 | 0.00 | 0.95 |
|  |  | 14 | Treatment | 1 | 0.22 | 0.64 | a | a | a |
|  |  |  | Control | 1 | 1.15 | 0.28 | 1 | 0.00 | 0.95 |
| 8 | BIOP | 7 | Treatment | 1 | 1.66 | 0.20 | a | a | a |
|  |  |  | Control | 1 | 0.02 | 0.88 | 1 | 1.50 | 0.22 |
|  |  | 14 | Treatment | 1 | 0.00 | 1.00 | 1 | 0.27 | 0.60 |

${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.10 (continued). Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of subyearling Chinook through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, summer 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam. BIOP spill $=56 \mathrm{kcfs}$ day/total dissolved gas cap at night.

| Release | Deflector |  |  | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spill | Ht | Population | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 9 | BIOP | 7 | Control | 1 | 0.02 | 0.88 | 1 | 1.50 | 0.22 |
|  |  |  | Treatment | 1 | 0.35 | 0.56 | 1 | 0.13 | 0.72 |
|  |  |  | Control | 1 | 0.35 | 0.55 | 1 | 0.50 | 0.48 |
|  | KCFS50 | 14 | Treatment | 1 | 0.53 | 0.47 | 1 | 0.01 | 0.91 |
| 10 |  | 7 | Control | 1 | 0.35 | 0.55 | 1 | 0.50 | 0.48 |
|  |  |  | Treatment | 1 | 0.00 | 1.00 | a | a | a |
|  |  | 14 | Control | 1 | 0.08 | 0.78 | 1 | 1.07 | 0.30 |
|  |  |  | Treatment | 1 | 0.17 | 0.68 | 1 | 0.10 | 0.75 |
| 11 | BIOP | 7 | Control | 1 | 0.00 | 0.95 | 1 | 3.66 | 0.06 |
|  |  |  | Treatment | 1 | 1.73 | 0.19 | 1 | 0.24 | 0.62 |
|  |  |  | Control | 1 | 3.62 | 0.06 | 1 | 0.43 | 0.51 |
|  |  | 14 | Treatment | 1 | 0.02 | 0.90 | a | , |  |
| 12 | BIOP | 7 | Control | 1 | 3.62 | 0.06 | 1 | 0.43 | 0.51 |
|  |  |  | Treatment | 1 | 0.01 | 0.93 | 1 | 0.56 | 0.45 |
|  |  |  | Control | 1 | 0.08 | 0.78 | 1 | 0.00 | 1.00 |
|  |  | 14 | Treatment | 1 | 1.05 | 0.30 | 1 | 0.17 | 0.68 |
| 13 | BIOP | 7 | Control | 1 | 0.08 | 0.78 | 1 | 0.00 | 1.00 |
|  |  |  | Treatment | 1 | 0.36 | 0.55 | 1 | 0.00 | 0.97 |
|  |  |  | Control | 1 | 0.67 | 0.41 | 1 | 0.12 | 0.73 |
| 14 |  | 14 | Treatment | 1 | 0.58 | 0.45 | 1 | 0.03 | 0.87 |
|  | KCFS50 | 7 | Control | 1 | 0.67 | 0.41 | 1 | 0.12 | 0.73 |
|  |  |  | Treatment | 1 | 0.00 | 0.94 | a | ${ }^{\text {a }}$ | ${ }^{\text {a }}$ |
|  |  |  | Control | 1 | 6.36 | 0.01 | 1 | 0.31 | 0.58 |
|  |  | 14 | Treatment | a | $\mathrm{a}$ | a | 1 | 1.40 | 0.24 |
| 15 | KCFS50 | 7 | Control | 1 | 6.36 | 0.01 | 1 | 0.31 | 0.58 |
|  |  |  | Treatment | 1 | 0.17 | 0.68 | 1 | 0.01 | 0.92 |
|  |  |  | Control | 1 | 0.37 | 0.54 | 1 | 1.29 | 0.26 |
| 16 |  | 14 | Treatment | 1 | 0.22 | 0.64 | 1 | 0.00 | 0.94 |
|  | BIOP | 7 | Control | 1 | 0.37 | 0.54 | 1 | 1.29 | 0.26 |
|  |  |  | Treatment | 1 | 0.13 | 0.72 | a | , |  |
|  |  |  | Control | 1 | 0.21 | 0.65 | 1 | 0.03 | 0.86 |
|  |  | 14 | Treatment | 1 | 0.86 | 0.35 | 1 | 0.28 | 0.60 |
|  |  |  | Control | 1 | 0.21 | 0.65 | 1 | 0.03 | 0.86 |
| 17 | BIOP | 7 | Treatment | 1 | 0.02 | 0.88 | 1 | 0.14 | 0.70 |

${ }^{a}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.10 (continued). Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of subyearling Chinook through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, summer 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam. BIOP spill $=56 \mathrm{kcfs}$ day/total dissolved gas cap at night.

| Release | Deflector |  |  | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spill | Ht | Population | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 18 | BIOP | 14 | Control | 1 | 4.18 | 0.04 | 1 | 0.01 | 0.91 |
|  |  |  | Treatment | 1 | 1.07 | 0.30 | 1 | 0.86 | 0.35 |
|  |  | 7 | Control | 1 | 4.18 | 0.04 | 1 | 0.01 | 0.91 |
|  |  |  | Treatment | 1 | 0.16 | 0.69 | a | a | a |
|  |  |  | Control | 1 | 0.26 | 0.61 | 1 | 0.93 | 0.33 |
| 19 |  | 14 | Treatment | 1 | 1.71 | 0.19 | 1 | 0.00 | 0.96 |
|  | BIOP | 7 | Control | 1 | 0.26 | 0.61 | 1 | 0.93 | 0.33 |
|  |  |  | Treatment | 1 | 0.00 | 1.00 | 1 | 0.02 | 0.89 |
|  |  |  | Control | 1 | 0.51 | 0.48 | 1 | 0.85 | 0.36 |
|  |  | 14 | Treatment | 1 | 0.57 | 0.45 | 1 | 0.00 | 1.00 |
| 20 | KCFS50 | 7 | Control | 1 | 0.51 | 0.48 | 1 | 0.85 | 0.36 |
|  |  |  | Treatment | 1 | 1.95 | 0.16 | 1 | 0.05 | 0.82 |
|  |  |  | Control | 1 | 8.67 | 0.00 | , | 0.18 | 0.67 |
|  |  | 14 | Treatment | 1 | 0.43 | 0.51 | a | a | a |
| 21 | BIOP | 7 | Control | 1 | 8.67 | 0.00 | 1 | 0.18 | 0.67 |
|  |  |  | Treatment | 1 | 0.00 | 1.00 | 1 | 0.00 | 0.97 |
|  |  |  | Control | 1 | 1.20 | 0.27 | 1 | 0.13 | 0.72 |
|  |  | 14 | Treatment | a | a | a | 1 | 0.09 | 0.77 |
| 22 | BIOP | 7 | Control | 1 | 1.20 | 0.27 | 1 | 0.13 | 0.72 |
|  |  |  | Treatment | 1 | 0.07 | 0.79 | a | a | a |
|  |  |  | Control | 1 | 0.01 | 0.94 | 1 | 1.20 | 0.27 |
| 23 |  | 14 | Treatment | 1 | 0.07 | 0.79 | , | 0.37 | 0.54 |
|  | BIOP |  | Control | 1 | 0.01 | 0.94 | 1 | 1.20 | 0.27 |
|  |  | 7 | Treatment | 1 | 0.17 | 0.68 | 1 | 0.08 | 0.77 |
|  |  |  | Control | 1 | 0.18 | 0.67 | , | 1.20 | 0.27 |
|  |  | 14 | Treatment | 1 | 0.07 | 0.80 | a | a | a |
| 24 | KCFS50 | 7 | Control | 1 | 0.18 | 0.67 | 1 | 1.20 | 0.27 |
|  |  |  | Treatment | 1 | 0.23 | 0.63 | a | ${ }^{\text {a }}$ | ${ }^{\text {a }}$ |
|  |  |  | Control | 1 | 0.62 | 0.43 | 1 | 0.67 | 0.41 |
| 25 | KCFS50 | 14 | Treatment | 1 | 0.08 | 0.78 | , | 0.03 | 0.86 |
|  |  |  | Control | 1 | 0.62 | 0.43 | 1 | 0.67 | 0.41 |
|  |  | 7 | Treatment | 1 | 0.07 | 0.80 | a | a | a |
|  |  |  | Control | 1 | 0.56 | 0.45 | 1 | 0.41 | 0.52 |
|  |  | 14 | Treatment | 1 | 0.03 | 0.85 | 1 | 0.02 | 0.90 |
|  |  |  | Control | 1 | 0.56 | 0.45 | , | 0.41 | 0.52 |
| 26 | KCFS50 | 7 | Treatment | 1 | 0.48 | 0.49 | 1 | 0.29 | 0.59 |

${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.10 (continued). Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of subyearling Chinook through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, summer 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam. BIOP spill $=56 \mathrm{kcfs}$ day/total dissolved gas cap at night.

| Release | Deflector |  |  | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spill | Ht | Population | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 27 | KCFS50 | 14 | Control | 1 | 1.99 | 0.16 | 1 | 0.01 | 0.91 |
|  |  |  | Treatment | 1 | 2.27 | 0.13 | a | a | a |
|  |  | 7 | Control | 1 | 1.99 | 0.16 | 1 | 0.01 | 0.91 |
|  |  |  | Treatment | 1 | 0.11 | 0.74 | 1 | 1.80 | 0.18 |
|  |  |  | Control | 1 | 9.97 | 0.00 | 1 | 0.01 | 0.91 |
|  |  | 14 | Treatment | 1 | 0.88 | 0.35 | 1 | 0.53 | 0.47 |
| 28 | BIOP | 7 | Control | 1 | 9.97 | 0.00 | 1 | 0.01 | 0.91 |
|  |  |  | Treatment | 1 | 1.96 | 0.16 | 1 | 3.59 | 0.06 |
|  |  |  | Control | 1 | 7.30 | 0.01 | 1 | 0.01 | 0.90 |
|  |  | 14 | Treatment | 1 | 0.87 | 0.35 | 1 | 0.01 | 0.94 |
| 29 | BIOP | 7 | Control | 1 | 7.30 | 0.01 | 1 | 0.01 | 0.90 |
|  |  |  | Treatment | 1 | 0.11 | 0.75 | 1 | 1.02 | 0.31 |
|  |  |  | Control | 1 | 1.12 | 0.29 | 1 | 0.00 | 0.95 |
| 30 |  | 14 | Treatment | 1 | 0.25 | 0.62 | 1 | 0.24 | 0.63 |
|  | BIOP | 7 | Control | 1 | 1.12 | 0.29 | 1 | 0.00 | 0.95 |
|  |  |  | Treatment | 1 | 0.38 | 0.54 | 1 | 0.00 | 1.00 |
|  |  |  | Control | 1 | 0.24 | 0.63 | 1 | 0.09 | 0.76 |
|  |  | 14 | Treatment | a | a | a | 1 | 0.17 | 0.68 |
| 31 | BIOP | 7 | Control | 1 | 0.24 | 0.63 | 1 | 0.09 | 0.76 |
|  |  |  | Treatment | 1 | 0.06 | 0.81 | 1 | 2.07 | 0.15 |
|  |  |  | Control | 1 | 0.03 | 0.87 | 1 | 0.08 | 0.77 |
| 32 | BIOP | 14 | Treatment | 1 | 0.01 | 0.93 | 1 | 0.36 | 0.55 |
|  |  |  | Control | 1 | 0.03 | 0.87 | 1 | 0.08 | 0.77 |
|  |  | 7 | Treatment | 1 | 0.00 | 0.95 | a | a | a |
|  |  |  | Control | 1 | 0.00 | 0.98 | 1 | 0.69 | 0.41 |
|  |  | 14 | Treatment | 1 | 0.08 | 0.77 | 1 | 0.21 | 0.65 |
| 33 | BIOP | 7 | Control | 1 | 0.00 | 0.98 | 1 | 0.69 | 0.41 |
|  |  |  | Treatment | 1 | 0.38 | 0.54 | 1 | 0.59 | 0.44 |
|  |  |  | Control | 1 | 6.94 | 0.01 | 1 | 1.29 | 0.26 |
| 34 | BIOP | 14 | Treatment | 1 | 0.04 | 0.85 | 1 | 0.05 | 0.82 |
|  |  |  | Control | 1 | 6.94 | 0.01 | 1 | 1.29 | 0.26 |
|  |  | 7 | Treatment | 1 | 0.20 | 0.65 | a | a | a |
|  |  |  | Control | 1 | 0.01 | 0.92 | 1 | 0.18 | 0.67 |
|  |  | 14 | Treatment | a | a | a | a | a | a |
|  |  |  | Control | 1 | 0.01 | 0.92 | 1 | 0.18 | 0.67 |
| 35 | BIOP | 7 | Treatment | 1 | 2.22 | 0.14 | 1 | 0.31 | 0.58 |

${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

Table A4.10 (continued). Summary statistics for goodness-of-fit tests (tests 2 and 3, Burnham et al. 1987) for releases used to estimate survival of subyearling Chinook through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, summer 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam. BIOP spill $=56 \mathrm{kcfs}$ day/total dissolved gas cap at night.

| Release | Deflector |  |  | Test 2 |  |  | Test 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spill | Ht | Population | df | $\chi^{2}$ | P | df | $\chi^{2}$ | P |
| 36 | 14 |  | Control | 1 | 0.23 | 0.63 | 1 | 0.03 | 0.87 |
|  |  |  | Treatment | 1 | 0.75 | 0.39 | 1 | 0.23 | 0.63 |
|  | KCFS50 | 7 | Control | 1 | 0.23 | 0.63 | 1 | 0.03 | 0.87 |
|  |  |  | Treatment | 1 | 0.03 | 0.87 | 1 | 0.06 | 0.81 |
|  |  |  | Control | 1 | 1.59 | 0.21 | 1 | 0.04 | 0.85 |
| 37 |  | 14 | Treatment | 1 | 0.02 | 0.88 | a | , |  |
|  | KCFS50 | 7 | Control | 1 | 1.59 | 0.21 | 1 | 0.04 | 0.85 |
|  |  |  | Treatment | 1 | 1.60 | 0.21 | a |  | a |
|  |  |  | Control | 1 | 1.46 | 0.23 | 1 | 0.89 | 0.34 |
|  |  | 14 | Treatment | 1 | 0.06 | 0.80 | 1 | 0.04 | 0.83 |
| 38 | BIOP | 7 | Control | 1 | 1.46 | 0.23 | 1 | 0.89 | 0.34 |
|  |  |  | Treatment | 1 | 0.03 | 0.86 | a |  | ${ }^{\text {a }}$ |
|  |  |  | Control | 1 | 2.62 | 0.11 | 1 | 0.06 | 0.81 |
| 39 |  | 14 | Treatment | 1 | 0.37 | 0.54 | 1 | 0.01 | 0.92 |
|  | BIOP | 7 | Control | 1 | 2.62 | 0.11 | 1 | 0.06 | 0.81 |
|  |  |  | Treatment | 1 | 0.19 | 0.66 | a | a | a |
|  |  |  | Control | 1 | 0.89 | 0.35 | 1 | 1.23 | 0.27 |
|  |  | 14 | Treatment | 1 | 0.01 | 0.93 | 1 | 0.80 | 0.37 |
|  |  |  | Control | 1 | 0.89 | 0.35 | 1 | 1.23 | 0.27 |

${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

## Appendix 5: Homogeneity of Arrival Times

Table A5.1. Summary of chi-square tests for homogeneity of arrival times of yearling Chinook salmon released into the minimum gap runner (MGR) turbine unit at Bonneville Dam powerhouse 1 and directly downstream of the front roll below the MGR turbine unit and detected at river kilometers 200, 194, and 181, spring 2004.

| Release | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | DF | Chi- <br> square | P | DF | Chi- <br> square | P | DF | Chi- <br> square | P |
| 1 | 3 | 3.60 | 0.308 | 2 | 0.97 | 0.616 | 3 | 3.24 | 0.356 |
| 2 | 0 | 0 | a | 0 | 0 | a | 1 | 0.76 | 0.385 |
| 3 | 0 | 0 | a | 1 | 0.81 | 0.368 | 1 | 0.86 | 0.353 |
| 4 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 5 | 0 | 0 | a | 1 | 0.05 | 0.823 | 1 | 0.01 | 0.923 |
| 6 | 0 | 0 | a | 1 | 1.10 | 0.294 | 1 | 0.89 | 0.345 |
| 7 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 8 | 1 | 1.35 | 0.245 | 1 | 1.10 | 0.295 | 1 | 1.03 | 0.311 |
| 9 | 1 | 1.33 | 0.249 | 1 | 1.12 | 0.291 | 1 | 0.94 | 0.333 |
| $10^{\text {b }}$ |  |  |  |  |  |  |  |  |  |
| 11 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 12 | 0 | 0 | a | 1 | 1.20 | 0.274 | 1 | 1.02 | 0.311 |
| 13 | 0 | 0 | a | 0 | 0 | 1 | 0 | 0 | , |
| 14 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 15 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| $16^{\text {b }}$ |  |  |  |  |  |  |  |  |  |

Table A5.2. Summary of chi-square tests for homogeneity of arrival times of yearling Chinook salmon released from a minimum gap runner (MGR) turbine at Bonneville Dam powerhouse 1 and in the Bonneville Dam tailrace below the juvenile bypass outfall at powerhouse 2 and detected at river kilometers 200, 194, and 181, spring 2004.

| Release | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | DF | Chisquare | P | DF | Chisquare | P | DF | Chisquare | P |
| 1 | 3 | 3.52 | 0.318 | 3 | 1.49 | 0.685 | 3 | 3.45 | 0.327 |
| 2 | 0 | 0 | a | 0 | 0 | , | 1 | 1.80 | 0.180 |
| 3 | 0 | 0 | a | 1 | 1.31 | 0.253 | 0 | 0 | a |
| 4 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 5 | 0 | 0 | a | 1 | 0.76 | 0.383 | 1 | 0.77 | 0.381 |
| 6 | 1 | 0.97 | 0.325 | 0 | 0 |  | 1 | 0.92 | 0.336 |
| 7 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | , |
| 8 | 1 | 1.51 | 0.219 | 1 | 1.35 | 0.245 | 1 | 1.08 | 0.300 |
| 9 | 1 | 0.82 | 0.366 | 1 | 0.61 | 0.435 | 1 | 0.68 | 0.408 |
| $10^{\text {b }}$ |  |  |  |  |  |  |  |  |  |
| 11 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 12 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 13 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 14 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| $\begin{aligned} & 15 \\ & 16^{\mathrm{b}} \end{aligned}$ | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |

${ }^{a}$ - All fish arrived on the same day at this detection array.
${ }^{\mathrm{b}}$ - Release was not analyzed due to dam operations not as specified in test conditions.

Table A5.3. Summary of chi-square tests for homogeneity of arrival times of yearling Chinook salmon released into the ice and trash sluiceway of powerhouse 1 at Bonneville Dam and below Bonneville Dam juvenile bypass outfall at powerhouse 2 and detected at river kilometers 200, 194, and 181, spring 2004.

| Release | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | DF | $\begin{gathered} \hline \text { Chi- } \\ \text { square } \end{gathered}$ | P | DF | $\begin{gathered} \text { Chi- } \\ \text { square } \end{gathered}$ | P | DF | $\begin{gathered} \text { Chi- } \\ \text { square } \end{gathered}$ | P |
| 1 | 2 | 2.18 | 0.336 | 2 | 4.13 | 0.127 | 3 | 5.82 | 0.121 |
| 2 | 0 | 0 | a | 0 | 0 | a | 1 | 0.83 | 0.362 |
| 3 | 0 | 0 | a | 1 | 1.37 | 0.241 | 0 | 0 | , |
| 4 | 0 | 0 | a | 0 | 0 | , | 0 | 0 | a |
| 5 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 6 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 7 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 8 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 9 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 10 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 11 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 12 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 13 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | ${ }^{\text {a }}$ |
| 14 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 15 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 16 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |

[^11]Table A5.4. Summary of chi-square tests for homogeneity of arrival times of hatchery steelhead trout released into a minimum gap runner (MGR) turbine unit at Bonneville Dam powerhouse 1 and directly downstream of the front roll below the MGR turbine unit and detected at river kilometers 200, 194, and 181, spring 2004.

| Release | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | DF | $\begin{gathered} \text { Chi- } \\ \text { square } \\ \hline \end{gathered}$ | P | DF | Chisquare | P | DF | Chisquare | P |
| 1 | 1 | 2.00 | 0.157 | 1 | 2.00 | 0.157 | 2 | 2.14 | 0.343 |
| 2 | 1 | 1.25 | 0.263 | 1 | 1.35 | 0.245 | 1 | 1.35 | 0.245 |
| 3 | 2 | 1.64 | 0.441 | 2 | 1.10 | 0.576 | 2 | 2.43 | 0.296 |
| 4 | 1 | 4.17 | 0.041 | 2 | 3.56 | 0.169 | 2 | 4.02 | 0.134 |
| 5 | 0 | 0 | a | 0 | 0 | a | 1 | 1.19 | 0.276 |
| 6 | 1 | 1.14 | 0.286 | 1 | 1.13 | 0.288 | 1 | 1.13 | 0.288 |
| 7 | 3 | 2.16 | 0.539 | 3 | 2.01 | 0.571 | 2 | 1.33 | 0.513 |
| 8 | b | b | b | b | b | b | b | b | b |
| 9 | 1 | 0.72 | 0.396 | 1 | 0.83 | 0.362 | 0 | 0 | a |
| 10 | 0 | 0 | a | 0 | 0 |  | 0 | 0 | a |
| 11 | 2 | 3.26 | 0.196 | 2 | 3.26 | 0.196 | 2 | 2.68 | 0.261 |
| 12 | 2 | 2.13 | 0.346 | 2 | 2.26 | 0.323 | 2 | 2.13 | 0.344 |
| 13 | 1 | 1.51 | 0.220 | 1 | 1.28 | 0.258 | 1 | 1.17 | 0.280 |
| 14 | b | b | b | b | b | b | b | b | b |
| 15 | 1 | 0.97 | 0.326 | 1 | 1.03 | 0.310 | 1 | 1.08 | 0.298 |
| 16 | 1 | 1.04 | 0.309 | 1 | 1.04 | 0.309 | 1 | 1.04 | 0.309 |

${ }^{\text {a }}$ - All fish arrived on the same day at this detection array.
${ }^{\mathrm{b}}$ - Release did not occur due to failure with release mechanism.
Table A5.5. Summary of chi-square tests for homogeneity of arrival times of hatchery steelhead trout released from a minimum gap runner (MGR) turbine at Bonneville Dam powerhouse 1 and the Bonneville Dam tailrace below the juvenile bypass outfall at powerhouse 2 and detected at river kilometers 200, 194, and 181, spring 2004.

| Release | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | DF | Chisquare | P | DF | Chisquare | P | DF | Chisquare | P |
| 1 | 2 | 1.03 | 0.596 | 2 | 1.03 | 0.596 | 2 | 1.11 | 0.574 |
| 2 | 1 | 1.47 | 0.225 | 1 | 1.59 | 0.208 | 1 | 1.59 | 0.208 |
| 3 | 1 | 0.44 | 0.509 | 1 | 0.46 | 0.498 | 1 | 0.29 | 0.589 |
| 4 | 2 | 4.46 | 0.107 | 3 | 5.76 | 0.124 | 3 | 4.09 | 0.252 |
| 5 | 0 | 0 | a | 0 | 0 | a | 1 | 1.13 | 0.287 |
| 6 | 0 | 0 | a | 1 | 1.51 | 0.219 | 1 | 1.36 | 0.244 |
| 7 | 3 | 3.54 | 0.315 | 2 | 2.26 | 0.323 | 2 | 1.09 | 0.581 |
| 8 | b | b | b | b | 2.26 | $\mathrm{b}^{\text {b }}$ | b | 1.0 | b |
| 9 | 1 | 0.98 | 0.323 | 2 | 2.00 | 0.367 | 1 | 0.89 | 0.347 |
| 10 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 11 | 2 | 3.42 | 0.181 | 2 | 3.58 | 0.167 | 2 | 2.81 | 0.245 |
| 12 | 2 | 1.46 | 0.483 | 2 | 1.41 | 0.494 | 2 | 1.41 | 0.494 |
| 13 | 1 | 0.79 | 0.374 | 1 | 0.92 | 0.336 | 1 | 0.88 | 0.348 |
| 14 | b | b | b | b | b | b | b | b | b |
| 15 | 1 | 0.83 | 0.362 | 1 | 0.85 | 0.356 | 1 | 0.86 | 0.354 |
| 16 | 1 | 0.82 | 0.366 | 1 | 0.89 | 0.345 | 1 | 0.89 | 0.345 |

[^12]Table A5.6. Summary of chi-square tests for homogeneity of arrival times of hatchery steelhead trout released into the ice and trash sluiceway of powerhouse 1 at Bonneville Dam and below Bonneville Dam juvenile bypass outfall at powerhouse 2 and detected at river kilometers 200, 194, and 181, spring 2004.

| Release | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | DF | Chisquare | P | DF | Chisquare | P | DF | Chisquare | P |
| 1 | 1 | 0.71 | 0.398 | 1 | 0.71 | 0.398 | 1 | 0.71 | 0.398 |
| 2 | 2 | 2.65 | 0.266 | 2 | 2.35 | 0.309 | 2 | 2.35 | 0.309 |
| 3 | 1 | 0 | 1.000 | 1 | 0.31 | 0.579 | 1 | 0.26 | 0.612 |
| 4 | 3 | 3.73 | 0.293 | 3 | 3.57 | 0.311 | 3 | 3.69 | 0.297 |
| 5 | 2 | 2.00 | 0.367 | 2 | 2.23 | 0.328 | 2 | 2.81 | 0.245 |
| 6 | 1 | 1.20 | 0.274 | 1 | 0 | 1.00 | 1 | 0.53 | 0.466 |
| 7 | 1 | 0.67 | 0.413 | 0 | 0 | a | 1 | 0.92 | 0.336 |
| 8 | 0 | 0 | , | 0 | 0 | a | 0 | 0 | a |
| 9 | 0 | 0 | a | 1 | 0.79 | 0.373 | 1 | 0.89 | 0.347 |
| 10 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 11 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 12 | 1 | 0.93 | 0.334 | 1 | 0.93 | 0.335 | 1 | 0.98 | 0.323 |
| 13 | 0 | 0 | a | 0 | 0 |  | 0 | 0 | , |
| 14 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 15 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 16 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |

[^13]Table A5.7. Summary of chi-square tests for homogeneity of arrival times of subyearling Chinook salmon released into the ice and trash sluiceway of powerhouse 1 at Bonneville Dam and below Bonneville Dam juvenile bypass outfall at powerhouse 2 and detected at river kilometers 200, 194, and 181, summer 2004.

| Release | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | DF | Chisquare | P | DF | Chisquare | P | DF | Chisquare | P |
| 1 | 0 | 0 | ${ }^{\text {a }}$ | 0 | 0 | ${ }^{\text {a }}$ | 0 | 0 | a |
| 2 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 3 | 1 | 0.35 | 0.552 | 0 | 0 | a | 0 | 0 | a |
| 4 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 5 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 6 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 7 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 8 | 0 | 0 | ${ }^{\text {a }}$ | 0 | 0 | a | 0 | 0 | a |
| 9 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 10 | 0 | 0 | ${ }^{\text {a }}$ | 0 | 0 | a | 0 | 0 | a |
| 11 | 1 | 44.00 | $3.284 \times 10^{-11}$ | 1 | 79.00 | 0 | 0 | 0 | a |
| 12 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 13 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 14 | 0 | 0 | a | 0 | 0 | a | 1 | 0.31 | 0.576 |
| 15 | 0 | 0 | a | 0 | 0 | a | 1 | 2.89 | 0.089 |
| 16 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 17 | 1 | 0.505 | 0.477 | 1 | 0.85 | 0.357 | 1 | 0.99 | 0.320 |
| 18 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 19 | 1 | 3.85 | 0.050 | 1 | 3.90 | 0.048 | 0 | 0 | a |
| 20 | 0 | 0 | a | 0 | 0 | , | 0 | 0 | ${ }^{\text {a }}$ |
| 21 | 1 | 4.12 | 0.042 | 1 | 4.29 | 0.038 | 1 | 3.99 | 0.046 |
| 22 | 0 | 0 | , | 1 | 0.27 | 0.605 | 2 | 0.67 | 0.714 |
| 23 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 24 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 25 | 0 | 0 | a | 0 | 0 | ${ }^{\text {a }}$ | 0 | 0 | a |
| 26 | 0 | 0 | a | 1 | 3.88 | 0.049 | 1 | 3.12 | 0.077 |
| 27 | 0 | 0 | a | 0 | 0 | , | 0 | 0 | , |
| 28 | 0 | 0 | a | 1 | 0.26 | 0.612 | 1 | 0.24 | 0.623 |
| 29 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 30 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 31 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | ${ }^{\text {a }}$ |
| 32 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |

[^14]Table A5.8. Summary of chi-square tests for homogeneity of arrival times of yearling Chinook salmon released and used to estimate survival through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, spring 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam and detected at river kilometers 200, 194, and 181.

| Release |  | Deflector | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | DF | Chisquare | P | DF | Chisquare | P | DF | Chisquare | P |
| 1 | Day | 7 | 3 | 4.09 | 0.252 | 3 | 2.68 | 0.443 | 3 | 3.96 | 0.266 |
|  |  | 14 | 3 | 0.87 | 0.832 | 3 | 0.34 | 0.951 | 3 | 1.46 | 0.691 |
|  | Night | 7 | 3 | 1.23 | 0.745 | 3 | 11.46 | 0.009 | 3 | 21.32 | 0.000 |
|  |  | 14 | 3 | 2.10 | 0.551 | 3 | 7.85 | 0.049 | 3 | 15.87 | 0.001 |
| 2 | Day | 7 | 2 | 0.51 | 0.773 | 3 | 5.04 | 0.169 | 3 | 10.86 | 0.012 |
|  |  | 14 | 2 | 1.12 | 0.571 | 3 | 1.24 | 0.743 | 3 | 2.29 | 0.515 |
|  | Night | 7 | 2 | 4.10 | 0.129 | 3 | 6.29 | 0.099 | 3 | 5.03 | 0.170 |
|  |  | 14 | 2 | 3.19 | 0.203 | 3 | 6.65 | 0.084 | 3 | 6.12 | 0.106 |
| 3 | Day | 7 | 2 | 7.51 | 0.023 | 2 | 3.56 | 0.169 | 2 | 5.83 | 0.054 |
|  |  | 14 | 2 | 0.14 | 0.935 | 2 | 0.36 | 0.834 | 2 | 0.47 | 0.792 |
|  | Night | 7 | 2 | 10.78 | 0.005 | 2 | 15.86 | 0.000 | 2 | 17.68 | 0.000 |
|  |  | 14 | 2 | 10.78 | 0.005 | 2 | 15.86 | 0.000 | 2 | 17.68 | 0.000 |
| 4 | Day | 7 | 4 | 4.80 | 0.308 | 4 | 6.14 | 0.189 | 3 | 8.50 | 0.037 |
|  |  | 14 | 3 | 0.96 | 0.812 | 3 | 4.09 | 0.252 | 3 | 3.26 | 0.353 |
|  | Night | 7 | 3 | 17.53 | 0.001 | 3 | 15.27 | 0.002 | 3 | 21.26 | 0.000 |
|  |  | 14 | 3 | 18.63 | 0.000 | 3 | 15.23 | 0.002 | 3 | 15.48 | 0.001 |
| 5 | Day | 7 | 2 | 0.56 | 0.757 | 2 | 2.17 | 0.338 | 2 | 0.06 | 0.969 |
|  |  | 14 | 2 | 3.49 | 0.175 | 2 | 2.55 | 0.279 | 2 | 0.48 | 0.789 |
|  | Night | 7 | 2 | 3.66 | 0.160 | 2 | 9.26 | 0.010 | 2 | 4.66 | 0.098 |
|  |  | 14 | 2 | 7.86 | 0.020 | 2 | 8.82 | 0.012 | 2 | 7.22 | 0.027 |
| 6 | Day | 7 | 2 | 9.88 | 0.007 | 2 | 10.48 | 0.005 | 2 | 7.29 | 0.026 |
|  |  | 14 | 2 | 8.95 | 0.011 | 2 | 7.07 | 0.029 | 2 | 10.63 | 0.005 |
|  | Night | 7 | 2 | 11.70 | 0.003 | 2 | 8.41 | 0.015 | 2 | 9.61 | 0.008 |
|  |  | 14 | 2 | 11.73 | 0.003 | 2 | 10.52 | 0.005 | 2 | 11.52 | 0.003 |
| 7 | Day | 7 | 4 | 24.92 | 0.000 | 4 | 14.87 | 0.005 | 4 | 20.85 | 0.000 |
|  |  | 14 | 4 | 8.73 | 0.068 | 4 | 3.81 | 0.432 | 4 | 10.27 | 0.036 |
|  | Night | 7 | 4 | 19.99 | 0.001 | 4 | 27.78 | 0.000 | 4 | 36.40 | 0.000 |
|  |  | 14 | 4 | 30.09 | 0.000 | 4 | 29.16 | 0.000 | 4 | 41.62 | 0.000 |
| 8 | Day | 7 | 2 | 11.09 | 0.004 | 2 | 8.20 | 0.017 | 2 | 10.35 | 0.006 |
|  |  | 14 | 1 | 6.69 | 0.010 | 2 | 3.40 | 0.183 | 2 | 5.59 | 0.061 |
|  | Night | 7 | 1 | 7.31 | 0.007 | 2 | 6.84 | 0.033 | 2 | 8.84 | 0.012 |
|  |  | 14 | 1 | 7.72 | 0.005 | 2 | 4.03 | 0.133 | 2 | 7.53 | 0.023 |
| 9 | Day | $7$ | 2 | 3.82 | 0.148 | 2 | 1.86 | 0.395 | 2 | 8.63 | 0.013 |
|  |  | 14 | 2 | 8.45 | 0.015 | 2 | 2.94 | 0.230 | 2 | 5.07 | 0.079 |
|  | Night | 7 | 2 | 5.69 | 0.058 | 2 | 8.89 | 0.012 | 2 | 10.56 | 0.005 |
|  |  | 14 | 2 | 8.17 | 0.017 | 2 | 12.06 | 0.002 | 2 | 19.53 | 0.000 |
| 10 | Day | 7 | 1 | 0.65 | 0.422 | 1 | 0.10 | 0.750 | 1 | 2.01 | 0.156 |
|  |  | 14 | 1 | 0.20 | 0.651 | 1 | 2.62 | 0.106 | 1 | 3.53 | 0.060 |
|  | Night | 7 | 1 | 0.01 | 0.905 | 1 | 0.32 | 0.570 | 1 | 1.17 | 0.279 |
|  |  | 14 | 1 | 0.73 | 0.392 | 1 | 0.01 | 0.933 | 1 | 0.06 | 0.801 |
| 11 | Day | 7 | 1 | 1.79 | 0.181 | 1 | 5.43 | 0.020 | 1 | 1.64 | 0.200 |
|  |  | 14 | 1 | 1.79 | 0.181 | 1 | 3.36 | 0.067 | 1 | 3.28 | 0.070 |
|  | Night | 7 | 1 | 2.37 | 0.124 | 1 | 4.44 | 0.035 | 1 | 9.99 | 0.002 |
|  |  | 14 | 1 | 0.33 | 0.564 | 1 | 6.23 | 0.013 | 1 | 9.71 | 0.002 |

Table A5.8 (continued). Summary of chi-square tests for homogeneity of arrival times of yearling Chinook salmon released and used to estimate survival through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, spring 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were released into the tailrace of Bonneville Dam and detected at river kilometers 200,194 , and 181.

| Release |  | Deflector | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | DF | Chisquare | P | DF | Chisquare | P | DF | Chisquare | P |
| 12 | Day | 7 | 0 | 0.00 | a | 1 | 0.47 | 0.493 | 1 | 0.47 | 0.493 |
|  |  | 14 | 0 | 0.00 | a | 1 | 0.08 | 0.779 | 1 | 0.08 | 0.779 |
|  | Night | 7 | 0 | 0.00 | a | 1 | 0.68 | 0.411 | 1 | 0.68 | 0.411 |
|  |  | 14 | 0 | 0.00 | a | 1 | 0.68 | 0.411 | 1 | 0.68 | 0.411 |
| 13 | Day | 7 | 3 | 19.07 | 0.000 | 3 | 12.11 | 0.007 | 3 | 16.74 | 0.001 |
|  |  | 14 | 4 | 20.94 | 0.000 | 3 | 15.82 | 0.001 | 4 | 32.65 | 0.000 |
|  | Night | 7 | 3 | 37.32 | 0.000 | 3 | 41.01 | 0.000 | 3 | 57.79 | 0.000 |
|  |  | 14 | 3 | 37.17 | 0.000 | 3 | 39.37 | 0.000 | 3 | 63.76 | 0.000 |

[^15]Table A5.9. Summary of chi-square tests for homogeneity of arrival times of hatchery steelhead trout used to estimate survival through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, spring 2004.
Releases were in The Dalles Dam tailrace and in the tailrace of Bonneville Dam and detected at river kilometers 200, 194, and 181, spring 2004.

| Release | $\overline{\text { Day }}$ | Deflector | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | DF | Chisquare | P | DF | Chisquare | P | DF | Chisquare | P |
| 1 |  | 7 | 1 | 2.55 | 0.110 | 3 | 32.71 | 0.000 | 3 | 28.93 | 0.000 |
|  |  | 14 | 1 | 0.68 | 0.410 | 2 | 1.84 | 0.399 | 2 | 2.17 | 0.337 |
|  | Night | 7 | 2 | 12.52 | 0.002 | 3 | 13.65 | 0.003 | 3 | 19.47 | 0.000 |
|  |  | 14 | 2 | 12.27 | 0.002 | 3 | 23.36 | 0.000 | 3 | 27.85 | 0.000 |
| 2 | Day | 7 | 1 | 13.43 | 0.000 | 2 | 4.52 | 0.104 | 2 | 6.01 | 0.049 |
|  |  | 14 | 1 | 4.85 | 0.028 | 1 | 5.51 | 0.019 | 1 | 5.51 | 0.019 |
|  | Night | 7 | 1 | 3.06 | 0.080 | 1 | 1.73 | 0.189 | 1 | 3.39 | 0.066 |
|  |  | 14 | 1 | 3.06 | 0.080 | 1 | 3.06 | 0.080 | 1 | 4.25 | 0.039 |
| 3 | Day | 7 | 2 | 14.80 | 0.001 | 2 | 12.50 | 0.002 | 2 | 8.81 | 0.012 |
|  |  | 14 | 2 | 3.01 | 0.222 | 2 | 3.01 | 0.222 | 2 | 1.37 | 0.505 |
|  | Night | 7 | 2 | 4.85 | 0.089 | 2 | 4.85 | 0.089 | 2 | 4.85 | 0.089 |
|  |  | 14 | 2 | 4.85 | 0.089 | 2 | 4.85 | 0.089 | 2 | 4.85 | 0.089 |
| 4 | Day | 7 | 2 | 24.42 | 0.000 | 2 | 26.12 | 0.000 | 2 | 18.29 | 0.000 |
|  |  | 14 | 2 | 15.56 | 0.000 | 2 | 14.88 | 0.001 | 2 | 14.88 | 0.001 |
|  | Night | 7 | 2 | 1.13 | 0.568 | 2 | 0.54 | 0.763 | 2 | 0.86 | 0.652 |
|  |  | 14 | 2 | 3.32 | 0.190 | 2 | 2.20 | 0.332 | 2 | 2.62 | 0.269 |
| 5 | Day | 7 | 5 | 27.48 | 0.000 | 5 | 35.28 | 0.000 | 5 | 29.41 | 0.000 |
|  |  | 14 | 5 | 10.53 | 0.062 | 5 | 9.29 | 0.098 | 5 | 5.19 | 0.393 |
|  | Night | 7 | 5 | 11.40 | 0.044 | 5 | 10.62 | 0.060 | 5 | 10.14 | 0.071 |
|  |  | 14 | 5 | 18.94 | 0.002 | 5 | 26.76 | 0.000 | 5 | 28.22 | 0.000 |
| 6 | Day | 7 | 3 | 22.54 | 0.000 | 3 | 20.15 | 0.000 | 3 | 18.55 | 0.000 |
|  |  | 14 | 3 | 6.41 | 0.093 | 3 | 6.00 | 0.111 | 3 | 4.75 | 0.191 |
|  | Night | 7 | 3 | 1.65 | 0.648 | 3 | 1.36 | 0.716 | 3 | 1.12 | 0.772 |
|  |  | 14 | 3 | 1.77 | 0.622 | 3 | 1.81 | 0.614 | 3 | 1.69 | 0.640 |
| 7 | Day | 7 | 4 | 75.51 | 0.000 | 3 | 70.64 | 0.000 | 4 | 75.22 | 0.000 |
|  |  | 14 | 4 | 33.46 | 0.000 | 3 | 32.43 | 0.000 | 4 | 35.64 | 0.000 |
|  | Night | 7 | 4 | 12.07 | 0.017 | 3 | 14.78 | 0.002 | 4 | 15.70 | 0.003 |
|  |  | 14 | 4 | 17.21 | 0.002 | 3 | 14.46 | 0.002 | 4 | 17.37 | 0.002 |
| 8 | Day | 7 | 3 | 38.20 | 0.000 | 3 | 46.96 | 0.000 | 3 | 41.96 | 0.000 |
|  |  | 14 | 2 | 14.57 | 0.001 | 2 | 17.77 | 0.000 | 3 | 12.48 | 0.006 |
|  | Night | 7 | 2 | 18.93 | 0.000 | 2 | 24.65 | 0.000 | 3 | 24.18 | 0.000 |
|  |  | 14 | 2 | 23.26 | 0.000 | 2 | 25.63 | 0.000 | 3 | 28.43 | 0.000 |
| 9 | Day | 7 | 1 | 15.48 | 0.000 | 1 | 15.32 | 0.000 | 1 | 16.08 | 0.000 |
|  |  | 14 | 1 | 13.69 | 0.000 | 1 | 15.49 | 0.000 | 2 | 16.13 | 0.000 |
|  | Night | 7 | 1 | 3.80 | 0.051 | 1 | 2.27 | 0.132 | 1 | 2.61 | 0.106 |
|  |  | 14 | 1 | 3.94 | 0.047 | 1 | 5.89 | 0.015 | 1 | 5.20 | 0.023 |
| 10 | Day | 7 | 3 | 38.30 | 0.000 | 3 | 41.52 | 0.000 | 3 | 27.56 | 0.000 |
|  |  | 14 | 3 | 26.89 | 0.000 | 3 | 26.61 | 0.000 | 3 | 25.95 | 0.000 |
|  | Night | 7 | 3 | 21.28 | 0.000 | 3 | 20.78 | 0.000 | 3 | 21.45 | 0.000 |
|  |  | 14 | 3 | 19.32 | 0.000 | 3 | 21.37 | 0.000 | 3 | 22.04 | 0.000 |
| 11 | Day | 7 | 2 | 18.05 | 0.000 | 2 | 14.57 | 0.001 | 2 | 10.46 | 0.005 |
|  |  | 14 | 2 | 13.30 | 0.001 | 2 | 12.81 | 0.002 | 2 | 13.30 | 0.001 |
|  | Night | 7 | 2 | 10.76 | 0.005 | 2 | 11.77 | 0.003 | 2 | 12.52 | 0.002 |
|  |  | 14 | 2 | 14.35 | 0.001 | 2 | 9.14 | 0.010 | 2 | 12.75 | 0.002 |
| 12 | Day | 7 | 1 | 6.62 | 0.010 | 1 | 12.95 | 0.000 | 1 | 11.37 | 0.001 |
|  |  | 14 | 1 | 6.28 | 0.012 | 1 | 7.27 | 0.007 | 1 | 3.80 | 0.051 |
|  | Night | 7 | 1 | 0.10 | 0.756 | 1 | 0.02 | 0.881 | 1 | 0.35 | 0.553 |
|  |  | 14 | 1 | 1.54 | 0.215 | 1 | 2.77 | 0.096 | 1 | 0.84 | 0.358 |

Table A5.9 (continued). Summary of chi-square tests for homogeneity of arrival times of hatchery steelhead trout used to estimate survival through spillbays with deflectors at 7 - ft and $14-\mathrm{ft}$, spring 2004. Releases were in The Dalles Dam tailrace and in the tailrace of Bonneville Dam and detected at river kilometers 200, 194, and 181, spring 2004.

| Release |  | Deflector | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | DF | Chisquare | P | DF | Chisquare | P | DF | Chisquare | P |
| 13 | Day | 7 | 1 | 6.62 | 0.010 | 2 | 9.51 | 0.009 | 2 | 8.15 | 0.0017 |
|  |  | 14 | 1 | 11.48 | 0.001 | 2 | 11.15 | 0.004 | 2 | 9.37 | 0.009 |
|  | Night | 7 | 2 | 3.52 | 0.172 | 2 | 2.70 | 0.259 | 2 | 3.02 | 0.220 |
|  |  | 14 | 1 | 6.23 | 0.013 | 2 | 7.55 | 0.023 | 2 | 9.69 | 0.008 |
| 14 | Day | 7 | 2 | 0.23 | 0.890 | 2 | 0.23 | 0.890 | 2 | 0.23 | 0.890 |
|  |  | 14 | 2 | 0.35 | 0.839 | 2 | 0.35 | 0.839 | 2 | 0.35 | 0.839 |
|  | Night | 7 | 1 | 0.45 | 0.501 | 1 | 0.45 | 0.501 | 2 | 0.92 | 0.630 |
|  |  | 14 | 2 | 0.77 | 0.682 | 2 | 1.52 | 0.467 | 2 | 5.49 | 0.064 |
| 15 | Day | 7 | 2 | 1.75 | 0.417 | 2 | 2.46 | 0.293 | 2 | 2.09 | 0.352 |
|  |  | 14 | 2 | 2.76 | 0.252 | 2 | 4.87 | 0.088 | 2 | 4.27 | 0.118 |
|  | Night | 7 | 2 | 1.36 | 0.506 | 2 | 5.34 | 0.069 | 2 | 7.57 | 0.023 |
|  |  | 14 | 2 | 3.83 | 0.147 | 2 | 5.60 | 0.061 | 2 | 5.74 | 0.057 |

Table A5.10. Summary of chi-square tests for homogeneity of arrival times of subyearling Chinook salmon used to estimate survival through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, summer 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were in the tailrace of Bonneville Dam and detected at river kilometers 200, 194, and 181. BIOP spill $=56 \mathrm{kcfs}$ day $/$ total dissolved gas cap at night.

| Release | Spill | Deflector | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | DF | Chisquare | P | DF | Chisquare | P | DF | Chisquare | P |
| 1 | BIOP | 7 | 1 | 0.45 | 0.503 | 1 | 0.01 | 0.926 | I | 2.81 | 0.094 |
|  |  | 14 | 1 | 0.91 | 0.341 | 1 | 3.50 | 0.061 | 1 | 2.80 | 0.095 |
| 2 | KCFS 50 | 7 | 1 | 2.14 | 0.144 | 1 | 1.74 | 0.187 | 1 | 2.47 | 0.116 |
|  |  | 14 | 1 | 0.00 | 0.986 | 1 | 0.04 | 0.843 | 1 | 0.00 | 0.977 |
| 3 | KCFS 50 | 7 | 1 | 1.57 | 0.211 | 1 | 0.24 | 0.627 | 1 | 0.04 | 0.832 |
|  |  | 14 | 1 | 2.98 | 0.084 | 1 | 6.83 | 0.009 | 1 | 7.74 | 0.005 |
| 4 | KCFS 50 | 7 | 1 | 0.86 | 0.353 | 1 | 4.75 | 0.029 | 1 | 5.67 | 0.017 |
|  |  | 14 | 1 | 3.42 | 0.064 | 1 | 5.34 | 0.021 | 1 | 8.93 | 0.003 |
| 5 | BIOP | 7 | 1 | 6.49 | 0.011 | 1 | 0.86 | 0.353 | 1 | 4.73 | 0.030 |
|  |  | 14 | 1 | 0.65 | 0.420 | 1 | 9.39 | 0.002 | 1 | 9.12 | 0.003 |
| 6 | BIOP | 7 | 0 | 0 | a | 1 | 0.31 | 0.580 | 1 | 1.25 | 0.263 |
|  |  | 14 | 0 | 0 | a | 1 | 0.22 | 0.640 | 1 | 1.57 | 0.210 |
| 7 | KCFS 50 | 7 | 2 | 1.86 | 0.394 | 2 | 1.98 | 0.372 | 2 | 3.88 | 0.144 |
|  |  | 14 | 2 | 4.84 | 0.089 | 2 | 3.60 | 0.166 | 2 | 5.50 | 0.064 |
| 8 | BIOP | 7 | 1 | 2.97 | 0.085 | 1 | 0.47 | 0.494 | 1 | 0.09 | 0.760 |
|  |  | 14 | 1 | 0.02 | 0.897 | 2 | 7.11 | 0.029 | 2 | 7.43 | 0.024 |
| 9 | BIOP | 7 | 1 | 0.28 | 0.598 | 1 | 3.63 | 0.056 | 2 | 3.91 | 0.142 |
|  |  | 14 | 0 | 0 | a | 1 | 4.18 | 0.041 | 1 | 0.24 | 0.622 |
| 10 | KCFS 50 | 7 | 1 | 2.07 | 0.151 | 3 | 0.91 | 0.824 | 3 | 0.77 | 0.856 |
|  |  | 14 | 2 | 2.29 | 0.319 | 3 | 7.44 | 0.059 | 4 | 11.63 | 0.020 |
| 11 | BIOP | 7 | 1 | 1.49 | 0.222 | 1 | 4.34 | 0.037 | 2 | 3.81 | 0.149 |
|  |  | 14 | 1 | 0.56 | 0.455 | 1 | 1.51 | 0.220 | 2 | 2.89 | 0.236 |
| 12 | BIOP | 7 | 1 | 0.44 | 0.508 | 0 | 0 | ${ }^{\text {a }}$ | 1 | 0.22 | 0.641 |
|  |  | 14 | 1 | 1.19 | 0.274 | 1 | 4.14 | 0.042 | 1 | 1.19 | 0.274 |
| 13 | BIOP | 7 | 1 | 5.17 | 0.023 | 1 | 0.42 | 0.520 | 1 | 0.73 | 0.394 |
|  |  | 14 | 1 | 0.87 | 0.352 | 1 | 0.40 | 0.528 | 2 | 7.86 | 0.020 |
| 14 | KCFS 50 | 7 | 1 | 4.39 | 0.036 | 1 | 0.32 | 0.573 | 1 | 0.42 | 0.516 |
|  |  | 14 | 1 | 0.91 | 0.339 | 1 | 0.56 | 0.453 | 1 | 0.00 | 0.978 |
| 15 | KCFS 50 | 7 | 2 | 3.53 | 0.171 | 3 | 2.96 | 0.398 | 4 | 5.72 | 0.221 |
|  |  | 14 | 2 | 7.50 | 0.023 | 3 | 0.86 | 0.835 | 4 | 0.77 | 0.942 |
| 16 | BIOP | 7 | 1 | 0.01 | 0.912 | 2 | 1.09 | 0.580 | 2 | 1.35 | 0.508 |
|  |  | 14 | 1 | 3.25 | 0.071 | 2 | 3.81 | 0.149 | 2 | 5.64 | 0.060 |
| 17 | BIOP | 7 | 1 | 1.88 | 0.170 | 1 | 0.68 | 0.409 | 1 | 1.75 | 0.186 |
|  |  | 14 | 1 | 0.35 | 0.551 | 1 | 0.72 | 0.396 | 1 | 1.29 | 0.256 |
| 18 | BIOP | 7 | 2 | 0.53 | 0.768 | 1 | 0.80 | 0.372 | 1 | 1.62 | 0.203 |
|  |  | 14 | 2 | 0.75 | 0.687 | 1 | 1.53 | 0.216 | 1 | 0.02 | 0.891 |
| 19 | BIOP | 7 | 0 | 0 | ${ }^{\text {a }}$ | 1 | 0.71 | 0.399 | 1 | 0.96 | 0.328 |
|  |  | 14 | 1 | 0.19 | 0.664 | 1 | 1.40 | 0.237 | 1 | 1.28 | 0.259 |
| 20 | KCFS 50 | 7 | 2 | 2.03 | 0.362 | 3 | 1.18 | 0.757 | 3 | 1.87 | 0.600 |
|  |  | 14 | 1 | 1.54 | 0.214 | 2 | 5.14 | 0.077 | 3 | 4.10 | 0.251 |
| 21 | BIOP | 7 | 1 | 0.00 | 0.960 | 1 | 0.81 | 0.367 | 1 | 0.67 | 0.415 |
|  |  | 14 | 1 | 0.87 | 0.351 | 1 | 0.27 | 0.600 | 1 | 2.99 | 0.084 |
| 22 | BIOP | 7 | 0 | 0 | , | 0 | 0 | a | 0 | 0 | a |
|  |  | 14 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |

${ }^{\text {a }}$ - All fish arrived on the same day at this detection array.

Table A5.10 (continued). Summary of chi-square tests for homogeneity of arrival times of subyearling Chinook salmon used to estimate survival through spillbays with deflectors at $7-\mathrm{ft}$ and $14-\mathrm{ft}$, summer 2004. Treatment fish were released in The Dalles Dam tailrace and control fish were in the tailrace of Bonneville Dam and detected at river kilometers 200, 194, and 181. BIOP spill $=56 \mathrm{kcfs}$ day $/$ total dissolved gas cap at night.

| Release | Spill | Deflector | River Kilometer 200 |  |  | River Kilometer 194 |  |  | River Kilometer 181 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | DF | Chisquare | P | DF | Chisquare | P | DF | Chisquare | P |
| 23 | BIOP | 7 | 1 | 3.90 | 0.048 | 1 | 0.31 | 0.579 | 1 | 0.00 | 0.978 |
|  |  | 14 | 0 | 0 | a | 1 | 1.40 | 0.237 | 1 | 9.18 | 0.002 |
| 24 | KCFS 50 | 7 | 0 | 0 | a | 1 | 0.17 | 0.681 | 1 | 0.38 | 0.540 |
|  |  | 14 | 0 | 0 | a | 1 | 0.12 | 0.730 | 1 | 0.36 | 0.548 |
| 25 | KCFS 50 | 7 | 1 | 1.63 | 0.201 | 0 | 0 | a | 1 | 8.38 | 0.004 |
|  |  | 14 | 0 | 0 | a | 1 | 5.80 | 0.016 | 1 | 5.80 | 0.016 |
| 26 | KCFS 50 | 7 | 1 | 0.08 | 0.773 | 1 | 1.02 | 0.313 | 2 | 1.29 | 0.525 |
|  |  | 14 | 1 | 1.88 | 0.170 | 1 | 2.19 | 0.139 | 2 | 5.75 | 0.056 |
| 27 | BIOP | 7 | 2 | 1.23 | 0.542 | 1 | 3.94 | 0.047 | 2 | 5.22 | 0.074 |
|  |  | 14 | 2 | 1.50 | 0.471 | 1 | 0.25 | 0.620 | 2 | 0.27 | 0.874 |
| 28 | BIOP | 7 | 0 | 0 | a | 0 | 0 | a | 1 | 1.78 | 0.183 |
|  |  | 14 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 29 | BIOP | 7 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
|  |  | 14 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 30 | BIOP | 7 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
|  |  | 14 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 31 | BIOP | 7 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
|  |  | 14 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 32 | BIOP | 7 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
|  |  | 14 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 33 | BIOP | 7 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
|  |  | 14 | 0 | 0 | a | 0 | 0 | a | 0 | 0 | a |
| 34 | BIOP | $7$ | 0 | $0$ |  | 0 | 0 | ${ }^{\text {a }}$ | 0 | 0 |  |
|  |  | 14 | 1 | 4.33 | 0.037 | 1 | 4.33 | 0.037 | 1 | 4.33 | 0.037 |
| 35 | KCFS 50 | 7 | 1 | 0.02 | 0.897 | 2 | 0.41 | 0.815 | 2 | 0.45 | 0.797 |
|  |  | 14 | 2 | 12.14 | 0.002 | 2 | 2.13 | 0.345 | 2 | 0.10 | 0.951 |
| 36 | KCFS 50 | 7 | 2 | 0.30 | 0.861 | 1 | 0.92 | 0.337 | 2 | 0.23 | 0.891 |
|  |  | 14 | 2 | 0.47 | 0.791 | 1 | 7.89 | 0.005 | 2 | 7.34 | 0.025 |
| 37 | BIOP | 7 | 1 | 0.76 | 0.384 | 1 | 0.00 | 0.962 | 1 | 0.21 | 0.644 |
|  |  | 14 | 1 | 0.05 | 0.821 | 1 | 3.03 | 0.082 | 1 | 2.65 | 0.104 |
| 38 | BIOP | 7 | 0 | 0 | , | 1 | 4.35 | 0.037 | 1 | 8.74 | 0.003 |
|  |  | 14 | 1 | 3.52 | 0.061 | 1 | 3.52 | 0.061 | 1 | 3.52 | 0.061 |

${ }^{\text {a }}$ - All fish arrived on the same day at this detection array.

# Appendix 6: Bonneville Lock and Dam, Revision to Fish Passage Plan Spill Patterns FEB2005 

CENWP-EC-HD
MEMORANDUM FOR THE RECORD
Subject:
Bonneville Lock and Dam, Revision to Fish Passage Plan Spill Patterns FEB2005

## 1. Introduction/ Background:

Location: Bonneville Dam and Lake, Columbia River Basin


Cross Section View of Spillway

Since additional flow deflectors were installed at Bonneville Dam Spillway (immediately prior to the 2002 spill season) a discrepancy between the computed inflow (The Dalles Outflow + tributary inflow) and outflow from Bonneville Dam was identified. This discrepancy occurred during times of spill. The reported spillway discharge turned out to be greater than the actual discharge (as measured downstream of the project). The magnitude of this discrepancy varied but was on the order of 20 Kcfs .

It has been determined that there are two major issues that have resulted in this flow discrepancy. One is a mis-calibration in the gate opening mechanism. The other is an out of date rating curve that gives the relationship between gate opening and flow.

During July 2004, it was discovered that the spillway gate hoist controller (GDACS) at Bonneville had been mis-calibrated and actual gate openings were up to 4 inches less than was reported. The greatest impact of this mis-calibration was on discharges at smaller gate openings. This effect was magnified by the new spill pattern developed for the new flow deflectors, which utilizes a larger number of gates at smaller openings for a given total spillway flow as compared to previous patterns.

The calibration errors would be significant primarily when the project was trying to meet a target discharge such as the 75 Kcfs daytime spill. When the project discharges to the gas cap the gas concentration downstream determines the spill volume that can be passed. The actual volume may have been misreported but the volume was set to meet the water quality requirements downstream. When the total river flows exceed the powerhouse capacity, the excess flow is also discharged through the spillway, increasing the 75 Kcfs daytime spill. In this case the spill is governed by total inflow and not increasing the forebay elevation.

During the investigation of the flow discrepancy between The Dalles (TDA) and Bonneville (BON) the spillway rating curves for both projects were scrutinized. Upon review the original TDA spillway-rating curve is consistent with current EM guidance. In addition the TDA discharge is verified by using a USGS gauging station just downstream of the TDA project. The BON spillway-rating curve is based on the orifice equation with the discharge coefficient determined from the original design physical model studies. In the 1970s the gate lip design was changed to reduce gate vibration. The lip changed from a rounded to a sharp edge design that also reduced the gate effeciency, especially at lower discharges. However, it does not appear that the rating curve was updated, and operation continued with the original rating curve. With older spill patterns, this difference was not particularly noticeable.

This memorandum will document the recommended BON spillway-rating curve and provide a relationship between actual spill in 2002 through 2004 given the reported spill in the Columbia River Operational Hydromet Management System (CROHMS) database for that same period. The relationship will not be exact but will provide a reasonable estimate of the actual spill volumes during the 2002, 2003 and 2004 spill season.

## 2. Rating Curve Revision:

The original Bonneville Spillway Rating Curve is based on the following orifice equation (HDC 311-1):
$Q=C_{d} A \sqrt{2 g H}$
Where:
$Q$ is the discharge in cfs
$C_{d}$ is the discharge coefficient
$A$ is the area of the opening in $\mathrm{ft}^{2}$
$\mathrm{A}=\mathrm{B} *$ Height of opening
B is the width of the opening in ft
$\mathrm{g}=\operatorname{gravity}\left(32.2 \mathrm{ft} / \mathrm{sec}^{2}\right)$
$H$ is the height of water from centerline of the opening to the surface
Physical model work conducted during the original design phase has been used to compute the discharge coefficient, $\mathrm{C}_{\mathrm{d}}$. The discharge coefficient from the original model work ranged from 1.1 for small gate openings to 0.7 for large gate openings (greater than 5 ft ). For small gate openings the discharge coefficient appears to be unusually high. Typical discharge coefficients range from 0.65 to 0.75 for tainter gates, 0.7 to 0.85 for regulating gates and 1.0 for very efficient tube orifices. It is most likely that the original gate lip may have been extremely efficient for small gate openings where as the current gate lip may be more typical of vertical sluice gates. Another possible factor is that gate leakage in the model may have adversely affected the flow measurements.


Comparison of original and current Bonneville Spillway gate lips

Discharge under high head vertical lift gates can be computed using the standard orifice equation (HDC 311-1) or using a relationship between gate-controlled discharge to free discharge (HDC 312). A spillway-rating curve was developed using both methods and they are presented in Figure 1. Included in Figure 1 is the original spillway-rating curve. The discharge coefficient for small gate openings is set equal to 0.80 in the rating curve called "Corrected Orifice Equation". The original rating curve falls above the other two rating curves for small gate openings. For this application a discharge coefficient of 0.80 was assumed for small gate openings. For example, for a 74.0 ft forebay, a gate opening equivalent to 1 dog , or 1.06 ft , yield a coefficient of 1.01 under the old rating curve for a flow of 3047 cfs , while the new rating curve would fix the coefficient at 0.80 for a flow of 2411 cfs. For a single bay this is a difference of 636 cfs, and across 18 bays the difference would be on the order of 11000 cfs (assuming for this example that all bays are open 1.06 ft ). As the gate opening increases, the coefficients match up better, and the differences themselves become less signifigant.

Thus the recommended rating curve for the Bonneville Spillway is documented in Table 2, which details a full rating curve for a full range of forebays ( 70 ft NGVD to 77 ft NGVD) and gate openings in both dogs and feet up to 12 dogs, or 22.15 ft .

Note that this will allow GDACS to compute the spillway discharge given the reported gate openings. The spill patterns recommended in the Fish Passage Plan and incorporated into the GDACS system at Bonneville need to have the discharge associated with a specific set of gate openings (the pattern) updated using the revised rating curve.

## Relationship Between Reported and Actual Spill

Using the spill patterns detailed in the Fish Passage Plan and incorporating the revised rating curve and the actual gate openings, the following comparison can be made for reported versus actual spill volumes. This assumes a Bonneville forebay elevation of 74.0 feet NGVD. The relationship would vary slightly for different forebay elevations and the impact of forebay elevation can be seen in Table 2. The results in Table 1 are presented in Figure 3 and a trend line has been fitted through the data points where:

$$
\text { ActualSpill }=0.001 x^{2}+0.8788 x-23.45
$$

Where:

$$
x=\text { reported spill }
$$

For example, if the reported spill from the CROHMS database showed 89 kcfs , using the ActualSpill formula the spill would actually be 63 kcfs .

Table 1
Comparison for spill patterns used since 2002, both ratings corrected for gate opening
Assumed correction applied to all gates to determine "Actual Gate Opening"
All values based on a Bonneville Forebay Elevation of 74.0
ft NGVD

|  |  | Gate Corrected Only |  |  | Gate and Gate Coefficient Corrected |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Nominal Spill | Reporte <br> d Spill | PRE200 <br> 5 <br> Ratinge Curve Flow | Differenc e | \% Differenc e | FEB200 <br> 5 Rating Curve Flow | $\left\lvert\, \begin{gathered} \text { Differenc } \\ \text { e } \end{gathered}\right.$ | $\%$ Differenc e |
| kcfs | kcfs |  |  |  | kcfs | kcfs | \% |
| 50 | 49.9 | 33.4 | 16.5 | 33.0 | 23.1 | 26.8 | 53.8 |
| 75 | 74.6 | 62.8 | 11.8 | 15.9 | 47.6 | 27.0 | 36.2 |
| 100 | 100.2 | 91.1 | 9.1 | 9.1 | 74.8 | 25.4 | 25.4 |
| 125 | 125.2 | 117.1 | 8.1 | 6.5 | 102.4 | 22.8 | 18.2 |
| 150 | 150.2 | 142.1 | 8.1 | 5.4 | 131.1 | 19.1 | 12.7 |

For a given requested spill, the spill pattern that closest matched was selected from the Fish Passage Plan. Table 1 shows the impact of the gate opening correction and the gate opening correction in conjunction with the gate coefficient correction.

## 3. Recommendations:

As of this writing (FEB2005) the GDACS system has been properly calibrated (see MFR "Bonneville Spillway Recalibration Field Trip Report", 180CT2004) by the project and reports an accurate gate opening. As of this writing the rating curve in GDACS has not been updated.

Recommended Actions:

- GDACS system used to control the spillway needs to be updated with the revised rating curves
- The Fish Passage Plan needs to be updated with the revised rating curves in the Bonneville Spill Pattern, it is recommended that the FEB2005 rating curve for a 74.0 ft NGVD forebay be used
- GDACS gate calibration should be confirmed prior to spill season and documented. Due to mechanical issues the hoists in general and the gantry operated bays in particular may loose calibration over time, and may do so to varying degrees (see MFR "Bonneville Spillway Recalibration Field Trip Report", 180CT2004)
- Spill should be monitored during the 2005 spill season to determine if discrepancy has been corrected to an acceptable level, if not, a field test may be required to update the rating curve due to the non-standard lip design.


## 4. References:

"Spillway Flow Discrepancy, Executive Summary", 15 pgs, Prepared by David B. Smith, Bonneville Project, dated 27JUL2004.
"Bonneville Spillway Recalibration Field Trip Report", 18oct2004bonnswrecal-1, written by HIGA, Nathan T., dated 180CT2004, EC-HD files.
"Hydraulic Design Criteria" (HDC), US Army Corps of Engineers Waterways Experiment Station, 1988.

Written by
HIGA, Nathan T.
Engineer, Hydraulic Design Section

Technical Review, Approval:
Laurie L. Ebner, P.E.
CF: CENWP-EC-HD Files


Bonneville Spillbay 17, looking across top of gate towards right pier, the markings shown correspond to dogs. However the dogs match the old 50 ft tall gates, and not the current 60 ft tall gates. Calibrating to the dogs resulted in the gates being open approximately 0.3 ft lower than reported by the hoisting equipment.


Local gate control with front panel open during recalibration process.


Figure 1. Bonneville Spillway Rating Curve Single Gate, Bonneville Forebay Elevation 74 ft .


Figure 2. Checking Proposed Spill Correction.


Figure 3. Bonneville Spillway, CROHMS Data, Reported vs Actual Flow.

Backed checked and compared to work by NTkH on 09FEB2005

| Forebay Elevation <br> FB <br> ft | Gate Opening |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | dogs | 0 | 1 |  |  |  |  | 2 |  |  | 3 |  |  |  |  |  |  |
|  | feet | 0 | 0.5 | 1 | 1.06 | 1.5 | 2 | 2.5 | 2.9 | 3 | 3.5 | 4 | 4.5 | 4.89 | 5 | 5.5 | 6 |
| 70 |  | 0 | 1102 | 2179 | 2307 | 3233 | 4265 | 5275 | 6069 | 6265 | 7235 | 8187 | 9120 | 9837 | 10037 | 10938 | 11824 |
| 70.2 |  | 0 | 1104 | 2184 | 2312 | 3240 | 4274 | 5287 | 6082 | 6279 | 7251 | 8205 | 9141 | 9859 | 10060 | 10963 | 11851 |
| 70.4 |  | 0 | 1106 | 2188 | 2317 | 3247 | 4283 | 5298 | 6095 | 6293 | 7267 | 8223 | 9161 | 9881 | 10083 | 10988 | 11878 |
| 70.6 |  | 0 | 1109 | 2193 | 2322 | 3254 | 4293 | 5310 | 6109 | 6306 | 7283 | 8241 | 9182 | 9903 | 10105 | 11012 | 11904 |
| 70.8 |  | 0 | 1111 | 2198 | 2327 | 3261 | 4302 | 5321 | 6122 | 6320 | 7299 | 8260 | 9202 | 9925 | 10128 | 11037 | 11931 |
| 71 |  | 0 | 1113 | 2202 | 2331 | 3268 | 4311 | 5333 | 6135 | 6334 | 7315 | 8278 | 9222 | 9947 | 10150 | 11062 | 11958 |
| 71.2 |  | 0 | 1116 | 2207 | 2336 | 3275 | 4320 | 5344 | 6148 | 6347 | 7331 | 8296 | 9243 | 9969 | 10172 | 11086 | 11985 |
| 71.4 |  | 0 | 1118 | 2212 | 2341 | 3282 | 4330 | 5356 | 6162 | 6361 | 7347 | 8314 | 9263 | 9991 | 10195 | 11111 | 12011 |
| 71.6 |  | 0 | 1120 | 2216 | 2346 | 3289 | 4339 | 5367 | 6175 | 6375 | 7363 | 8332 | 9283 | 10013 | 10217 | 11135 | 12038 |
| 71.8 |  | 0 | 1123 | 2221 | 2351 | 3296 | 4348 | 5378 | 6188 | 6388 | 7378 | 8350 | 9303 | 10035 | 10239 | 11160 | 12064 |
| 72 |  | 0 | 1125 | 2226 | 2356 | 3303 | 4357 | 5390 | 6201 | 6402 | 7394 | 8368 | 9323 | 10057 | 10262 | 11184 | 12091 |
| 72.2 |  | 0 | 1127 | 2230 | 2361 | 3309 | 4366 | 5401 | 6214 | 6415 | 7410 | 8386 | 9343 | 10078 | 10284 | 11208 | 12117 |
| 72.4 |  | 0 | 1130 | 2235 | 2366 | 3316 | 4375 | 5412 | 6227 | 6429 | 7426 | 8403 | 9363 | 10100 | 10306 | 11232 | 12144 |
| 72.6 |  | 0 | 1132 | 2239 | 2371 | 3323 | 4384 | 5424 | 6240 | 6442 | 7441 | 8421 | 9383 | 10121 | 10328 | 11257 | 12170 |
| 72.8 |  | 0 | 1134 | 2244 | 2375 | 3330 | 4393 | 5435 | 6253 | 6456 | 7457 | 8439 | 9403 | 10143 | 10350 | 11281 | 12196 |
| 73 |  | 0 | 1137 | 2248 | 2380 | 3337 | 4402 | 5446 | 6266 | 6469 | 7472 | 8457 | 9423 | 10165 | 10372 | 11305 | 12222 |
| 73.2 |  | 0 | 1139 | 2253 | 2385 | 3344 | 4411 | 5457 | 6279 | 6483 | 7488 | 8474 | 9443 | 10186 | 10394 | 11329 | 12248 |
| 73.4 |  | 0 | 1141 | 2258 | 2390 | 3350 | 4420 | 5469 | 6292 | 6496 | 7504 | 8492 | 9462 | 10207 | 10416 | 11353 | 12274 |
| 73.6 |  | 0 | 1143 | 2262 | 2395 | 3357 | 4429 | 5480 | 6305 | 6509 | 7519 | 8510 | 9482 | 10229 | 10438 | 11377 | 12300 |
| 73.8 |  | 0 | 1146 | 2267 | 2400 | 3364 | 4438 | 5491 | 6318 | 6523 | 7534 | 8527 | 9502 | 10250 | 10459 | 11401 | 12326 |
| 74 |  | 0 | 1148 | 2271 | 2404 | 3371 | 4447 | 5502 | 6331 | 6536 | 7550 | 8545 | 9522 | 10271 | 10481 | 11424 | 12352 |
| 74.2 |  | 0 | 1150 | 2276 | 2409 | 3377 | 4456 | 5513 | 6344 | 6549 | 7565 | 8562 | 9541 | 10293 | 10503 | 11448 | 12378 |
| 74.4 |  | 0 | 1153 | 2280 | 2414 | 3384 | 4465 | 5524 | 6356 | 6562 | 7581 | 8580 | 9561 | 10314 | 10524 | 11472 | 12404 |
| 74.6 |  | 0 | 1155 | 2285 | 2419 | 3391 | 4474 | 5535 | 6369 | 6576 | 7596 | 8597 | 9580 | 10335 | 10546 | 11496 | 12430 |
| 74.8 |  | 0 | 1157 | 2289 | 2423 | 3398 | 4483 | 5546 | 6382 | 6589 | 7611 | 8615 | 9600 | 10356 | 10568 | 11519 | 12455 |
| 75 |  | 0 | 1159 | 2294 | 2428 | 3404 | 4492 | 5557 | 6395 | 6602 | 7627 | 8632 | 9619 | 10377 | 10589 | 11543 | 12481 |
| 75.2 |  | 0 | 1162 | 2298 | 2433 | 3411 | 4501 | 5568 | 6407 | 6615 | 7642 | 8649 | 9639 | 10398 | 10611 | 11566 | 12507 |
| 75.4 |  | 0 | 1164 | 2303 | 2438 | 3418 | 4509 | 5579 | 6420 | 6628 | 7657 | 8667 | 9658 | 10419 | 10632 | 11590 | 12532 |
| 75.6 |  | 0 | 1166 | 2307 | 2442 | 3424 | 4518 | 5590 | 6433 | 6641 | 7672 | 8684 | 9677 | 10440 | 10653 | 11613 | 12558 |
| 75.8 |  | 0 | 1168 | 2312 | 2447 | 3431 | 4527 | 5601 | 6445 | 6654 | 7687 | 8701 | 9697 | 10461 | 10675 | 11637 | 12583 |
| 76 |  | 0 | 1170 | 2316 | 2452 | 3437 | 4536 | 5612 | 6458 | 6667 | 7702 | 8718 | 9716 | 10482 | 10696 | 11660 | 12608 |
| 76.2 |  | 0 | 1173 | 2320 | 2456 | 3444 | 4545 | 5623 | 6471 | 6680 | 7718 | 8735 | 9735 | 10503 | 10717 | 11683 | 12634 |
| 76.4 |  | 0 | 1175 | 2325 | 2461 | 3451 | 4553 | 5634 | 6483 | 6693 | 7733 | 8753 | 9754 | 10524 | 10739 | 11707 | 12659 |
| 76.6 |  | 0 | 1177 | 2329 | 2466 | 3457 | 4562 | 5645 | 6496 | 6706 | 7748 | 8770 | 9773 | 10544 | 10760 | 11730 | 12684 |
| 76.8 |  | 0 | 1179 | 2334 | 2470 | 3464 | 4571 | 5656 | 6508 | 6719 | 7763 | 8787 | 9793 | 10565 | 10781 | 11753 | 12709 |
| 77 |  | 0 | 1182 | 2338 | 2475 | 3470 | 4579 | 5666 | 6521 | 6732 | 7778 | 8804 | 9812 | 10586 | 10802 | 11776 | 12734 |


| Table 2 Bo | ville Spill | a a sin | bay with | pect to | $\overline{\text { ning ir }}$ | (\& fe | d La | evatio |  | foot C |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Computed by | LE on F | uary 11, | $05$ |  |  |  |  |  |  |  |  |  |  |  |  |  | 2/3 |
| Backed check | d and co | red to | k by NT | on 09 | 2005 |  |  |  |  |  |  |  |  |  |  |  |  |
| Forebay Elevation | Gate | penin |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| FB | dogs |  | 4 |  |  |  |  | 5 |  |  |  |  | 6 |  |  |  |  |
| ft | feet | 6.5 | 6.81 | 7 | 7.5 | 8 | 8.5 | 8.73 | 9 | 9.5 | 10 | 10.5 | 10.64 | 11 | 11.5 | 12 | 12.5 |
| 70 |  | 12695 | 13228 | 13553 | 14398 | 15231 | 16053 | 16428 | 16865 | 17668 | 18462 | 19248 | 19467 | 20028 | 20801 | 21568 | 22331 |
| 70.2 |  | 12724 | 13259 | 13584 | 14431 | 15267 | 16091 | 16467 | 16905 | 17710 | 18506 | 19295 | 19514 | 20076 | 20851 | 21621 | 22386 |
| 70.4 |  | 12753 | 13289 | 13615 | 14465 | 15302 | 16129 | 16505 | 16945 | 17752 | 18550 | 19341 | 19561 | 20125 | 20902 | 21674 | 22441 |
| 70.6 |  | 12782 | 13319 | 13646 | 14498 | 15338 | 16166 | 16544 | 16985 | 17794 | 18594 | 19387 | 19608 | 20173 | 20952 | 21726 | 22495 |
| 70.8 |  | 12811 | 13350 | 13677 | 14531 | 15373 | 16204 | 16582 | 17024 | 17836 | 18638 | 19433 | 19654 | 20221 | 21003 | 21779 | 22550 |
| 71 |  | 12840 | 13380 | 13708 | 14564 | 15408 | 16241 | 16621 | 17064 | 17877 | 18682 | 19479 | 19701 | 20269 | 21053 | 21831 | 22605 |
| 71.2 |  | 12869 | 13410 | 13739 | 14597 | 15443 | 16278 | 16659 | 17103 | 17919 | 18726 | 19525 | 19747 | 20317 | 21103 | 21883 | 22659 |
| 71.4 |  | 12898 | 13440 | 13770 | 14630 | 15478 | 16316 | 16697 | 17143 | 17960 | 18769 | 19570 | 19793 | 20365 | 21153 | 21935 | 22713 |
| 71.6 |  | 12926 | 13470 | 13801 | 14663 | 15513 | 16353 | 16735 | 17182 | 18001 | 18813 | 19616 | 19840 | 20412 | 21203 | 21987 | 22767 |
| 71.8 |  | 12955 | 13500 | 13832 | 14696 | 15548 | 16390 | 16773 | 17221 | 18043 | 18856 | 19661 | 19886 | 20460 | 21252 | 22039 | 22821 |
| 72 |  | 12983 | 13530 | 13862 | 14729 | 15583 | 16427 | 16811 | 17260 | 18084 | 18899 | 19707 | 19932 | 20507 | 21302 | 22091 | 22875 |
| 72.2 |  | 13012 | 13560 | 13893 | 14761 | 15618 | 16464 | 16849 | 17299 | 18125 | 18942 | 19752 | 19977 | 20555 | 21351 | 22143 | 22929 |
| 72.4 |  | 13040 | 13589 | 13924 | 14794 | 15653 | 16500 | 16887 | 17338 | 18166 | 18985 | 19797 | 20023 | 20602 | 21401 | 22194 | 22983 |
| 72.6 |  | 13069 | 13619 | 13954 | 14827 | 15687 | 16537 | 16924 | 17377 | 18207 | 19028 | 19842 | 20069 | 20649 | 21450 | 22245 | 23036 |
| 72.8 |  | 13097 | 13649 | 13984 | 14859 | 15722 | 16574 | 16962 | 17415 | 18247 | 19071 | 19887 | 20114 | 20696 | 21499 | 22297 | 23090 |
| 73 |  | 13125 | 13678 | 14015 | 14891 | 15756 | 16610 | 16999 | 17454 | 18288 | 19114 | 19932 | 20160 | 20743 | 21548 | 22348 | 23143 |
| 73.2 |  | 13153 | 13708 | 14045 | 14924 | 15791 | 16647 | 17037 | 17492 | 18329 | 19157 | 19977 | 20205 | 20790 | 21597 | 22399 | 23196 |
| 73.4 |  | 13182 | 13737 | 14075 | 14956 | 15825 | 16683 | 17074 | 17531 | 18369 | 19199 | 20021 | 20250 | 20837 | 21646 | 22450 | 23249 |
| 73.6 |  | 13210 | 13767 | 14105 | 14988 | 15859 | 16719 | 17111 | 17569 | 18410 | 19242 | 20066 | 20295 | 20883 | 21695 | 22501 | 23302 |
| 73.8 |  | 13238 | 13796 | 14135 | 15020 | 15894 | 16756 | 17149 | 17607 | 18450 | 19284 | 20110 | 20340 | 20930 | 21743 | 22551 | 23355 |
| 74 |  | 13266 | 13825 | 14165 | 15052 | 15928 | 16792 | 17186 | 17646 | 18490 | 19326 | 20155 | 20385 | 20976 | 21792 | 22602 | 23407 |
| 74.2 |  | 13294 | 13854 | 14195 | 15084 | 15962 | 16828 | 17223 | 17684 | 18530 | 19369 | 20199 | 20430 | 21023 | 21840 | 22652 | 23460 |
| 74.4 |  | 13321 | 13883 | 14225 | 15116 | 15996 | 16864 | 17260 | 17722 | 18570 | 19411 | 20243 | 20475 | 21069 | 21888 | 22703 | 23512 |
| 74.6 |  | 13349 | 13912 | 14255 | 15148 | 16030 | 16900 | 17297 | 17760 | 18610 | 19453 | 20287 | 20520 | 21115 | 21936 | 22753 | 23565 |
| 74.8 |  | 13377 | 13941 | 14285 | 15180 | 16063 | 16936 | 17333 | 17798 | 18650 | 19495 | 20331 | 20564 | 21161 | 21984 | 22803 | 23617 |
| 75 |  | 13405 | 13970 | 14315 | 15212 | 16097 | 16971 | 17370 | 17835 | 18690 | 19536 | 20375 | 20609 | 21207 | 22032 | 22853 | 23669 |
| 75.2 |  | 13432 | 13999 | 14344 | 15243 | 16131 | 17007 | 17407 | 17873 | 18730 | 19578 | 20419 | 20653 | 21253 | 22080 | 22903 | 23721 |
| 75.4 |  | 13460 | 14028 | 14374 | 15275 | 16164 | 17043 | 17443 | 17911 | 18770 | 19620 | 20462 | 20697 | 21298 | 22128 | 22953 | 23773 |
| 75.6 |  | 13487 | 14057 | 14403 | 15307 | 16198 | 17078 | 17480 | 17948 | 18809 | 19661 | 20506 | 20741 | 21344 | 22176 | 23002 | 23824 |
| 75.8 |  | 13515 | 14085 | 14433 | 15338 | 16231 | 17114 | 17516 | 17986 | 18849 | 19703 | 20550 | 20785 | 21389 | 22223 | 23052 | 23876 |
| 76 |  | 13542 | 14114 | 14462 | 15369 | 16265 | 17149 | 17552 | 18023 | 18888 | 19744 | 20593 | 20829 | 21435 | 22271 | 23101 | 23928 |
| 76.2 |  | 13569 | 14143 | 14491 | 15401 | 16298 | 17184 | 17589 | 18060 | 18927 | 19786 | 20636 | 20873 | 21480 | 22318 | 23151 | 23979 |
| 76.4 |  | 13597 | 14171 | 14521 | 15432 | 16331 | 17220 | 17625 | 18098 | 18966 | 19827 | 20679 | 20917 | 21525 | 22365 | 23200 | 24030 |
| 76.6 |  | 13624 | 14200 | 14550 | 15463 | 16365 | 17255 | 17661 | 18135 | 19006 | 19868 | 20723 | 20961 | 21570 | 22412 | 23249 | 24081 |
| 76.8 |  | 13651 | 14228 | 14579 | 15494 | 16398 | 17290 | 17697 | 18172 | 19045 | 19909 | 20766 | 21004 | 21615 | 22459 | 23298 | 24133 |
| 77 |  | 13678 | 14256 | 14608 | 15525 | 16431 | 17325 | 17733 | 18209 | 19084 | 19950 | 20809 | 21048 | 21660 | 22506 | 23347 | 24184 |

Backed checked and compared to work by NTkH on 09FEB2005



[^0]:    ${ }^{\text {A }}$ - No releases were made into the MGR unit in the summer.

[^1]:    ${ }^{\mathrm{a}}$ - Release equipment was not properly working.
    ${ }^{\mathrm{b}}$ - Dam operations were not as specified for treatment conditions.
    ${ }^{\text {c }}$ - Releases within 24 h of detected radio-tagged dead fish

[^2]:    ${ }^{\mathrm{a}}$ - Releases within 24 h of detected radio-tagged dead fish

[^3]:    ${ }^{\text {a }}$ - Dam operations were not as specified for treatment conditions.

[^4]:    ${ }^{\text {a }}$ - Releases within 24 h of detected radio-tagged dead fish

[^5]:    ${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables. ${ }^{\mathrm{b}}$ - Dam operations were not appropriate for prescribed test conditions.

[^6]:    ${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.
    ${ }^{\mathrm{b}}$ - Dam operations were not appropriate for prescribed test conditions.

[^7]:    ${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.
    ${ }^{\mathrm{b}}$ - Dam operations were not appropriate for prescribed test conditions.

[^8]:    ${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.
    ${ }^{\mathrm{b}}$ - Dam operations were not appropriate for prescribed test conditions.

[^9]:    ${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

[^10]:    ${ }^{\text {a }}$ - Chi-square statistic was not calculable for these tests due to the presence of only zeroes in rows or columns in the contingency tables.

[^11]:    ${ }^{a}$ - All fish arrived on the same day at this detection array.

[^12]:    ${ }^{\text {a }}$ - All fish arrived on the same day at this detection array.
    ${ }^{\mathrm{b}}$ - Release was not analyzed due to dam operations not as specified in test conditions.

[^13]:    ${ }^{\text {a }}$ - All fish arrived on the same day at this detection array.

[^14]:    ${ }^{\text {a }}$ - All fish arrived on the same day at this detection array.

[^15]:    ${ }^{\text {a }}$ - All fish arrived on the same day at this detection array.

