

APPENDIX 3 - MRLs AND EMEGs FOR TCDD

CURRENT MRLs

ATSDR published the *Toxicological Profile for TCDD* (ATSDR, 1989). Minimal risk levels (MRLs) listed in the profile were for acute, intermediate-duration, and chronic oral exposures (see Table 3-1).

Acute Oral MRL

The acute oral MRL of 100 pg/kg/day was based on hepatotoxic effects in guinea pigs that were observed following administration of a single gavage dose of 0.1 µg/kg TCDD (Turner and Collins, 1983).

An uncertainty factor of 10 was used for extrapolation from animals to humans, a factor of 10 for human variability, and a factor of 10 for the use of a lowest-observed-adverse-effect level (LOAEL).

Intermediate Oral MRL

The LOAEL of 0.001 µg/kg/day was considered for derivation of the intermediate-duration oral MRL of 1 pg/kg/day. At this exposure level, dilated pelvises and changes in gestational index were observed in rats (Murray et al., 1979) and abortions were reported in monkeys (Allen et al., 1979).

An uncertainty factor of 10 was used for extrapolation from animals to humans, a factor of 10 for human variability, and a factor of 10 for the use of a LOAEL.

Chronic Oral MRL

The intermediate-duration oral MRL of 1 pg/kg/day was also adopted as the chronic oral MRL.

PROPOSED MRLs

The *Toxicological Profile for CDDs* was in a draft stage in 1993/1994. The internal MRL workgroup proposed oral MRLs for TCDD (see Table 3-1).

Acute Oral MRL

The acute oral MRL of 20 pg/kg/day was based on the LOAEL of 0.01 µg/kg/day TCDD that induced suppressed serum complement activity in B6C3F1 mice exposed to 14 daily doses administered by gavage-in-oil vehicle (White et al., 1986).

An uncertainty factor of 10 was used for extrapolation from animals to humans, a factor of 10 for human variability, and a factor of 10 for the use of a LOAEL. Furthermore, a modifying factor of 0.5 was applied to adjust for the difference in higher bioavailability of TCDD from gavage-in-oil vehicle than from food or soil.

Intermediate Oral MRL

The intermediate-duration oral MRL of 7 pg/kg/day was based on a no-observed-adverse-effect level (NOAEL) of 0.0007 µg/kg/day TCDD for decreased thymus weight in guinea pigs exposed for 90 days in their feed (DeCaprio et al., 1986). The LOAEL in the study was 0.005 µg/kg/day.

An uncertainty factor of 10 was used for interspecies extrapolation and a factor of 10 for human variability. The NOAEL for deriving an intermediate-duration exposure MRL is also supported by the same level NOAEL for liver effects in the DeCaprio et al. study. The liver effects reported at higher levels consisted of hepatocellular inclusions and hypertriglyceridemia.

Chronic Oral MRL

A chronic oral MRL of 0.7 pg/kg/day was based on a LOAEL of 0.0002 µg/kg/day TCDD in the feed of monkeys that resulted in mild learning and behavioral impairment in their offspring (Bowman et al., 1989).

An uncertainty factor of 3 was used for the use of a minimal LOAEL, a factor of 10 was used for interspecies extrapolation, and a factor of 10 for human variability.

Environmental media evaluation guides (EMEGs) are media-specific comparison values that are used to select contaminants of concern at hazardous waste sites.

EMEGs are derived for air, water, and soil environmental media. They are based on inhalation and oral MRLs for air and water/soil exposures, respectively. The methodology and formula for derivation of EMEGs are described in ATSDR's Public Health Assessment Guidance Manual (ATSDR, 1992).

EMEGs are estimates of external dose. They do not provide data on how much of the dose is actually absorbed. No EMEGs are available for the dermal exposure route.

EMEGs based on these MRLs are presented in Tables 3-2a and 3-2b.

TABLE 3-1. MRLs* for TCDD

Year	Exposure duration	MRL* in pg/kg /day	UF LOAEL /NOAEL	UF inter- species	UF sensitivity	MF**	End point	Study
1989	acute	100	10	10	10		LOAEL for hepatotoxicity guinea pigs	Turner and Collins, 1983
1989	inter- mediate	1	10	10	10		LOAEL for abortions and other reproductive, developmental effects rats, monkeys	Murray et al., 1979 Allen et al., 1979
1989	chronic	1	10	10	10		LOAEL for abortions and other reproductive, developmental effects rats, monkeys	Murray et al., 1979 Allen et al., 1979
1994	acute	20	10	10	10	0.5	LOAEL for suppressed serum complement activity mice	White et al., 1986
1994	inter- mediate	7		10	10		NOAEL for decreased thymus weight: liver toxicity guinea pigs	DeCaprio et al., 1986
1994	chronic	0.7	3	10	10		LOAEL for mild learning and behavioral impairment monkey offspring	Bowman et al., 1989

*The MRL is calculated as $MRL = (NOAEL \text{ or } LOAEL) / (UF \times MF)$, where MRL = minimal risk level (mg/kg/day), NOAEL = no-observed-adverse-effect level (mg/kg/day), LOAEL = lowest-observed-adverse-effect level (mg/kg/day), UF = uncertainty factor (unitless), MF = modifying factor (unitless)

**MF for bioavailability was used in the derivation of an acute MRL (1994)

TABLE 3-2a. EMEGs (in ppb) Based on 1989 TCDD MRLs

Exposure duration	Child	Adult
acute	5	70
intermediate	0.05	0.7
chronic	0.05	0.7

TABLE 3-2b. EMEGs (in ppb) Based on 1994 TCDD MRLs

Exposure duration	Child	Adult
acute	1	14
intermediate	5	5
chronic	0.04	0.5

*The EMEG is calculated as $EMEG = (MRL)(BW) / IR$, where EMEG = environmental media evaluation guide (mg/kg), BW = body weight in kg (adult = 70 kg; child = 10 kg), IR = soil ingestion rate (mg/day) (adult = 100 mg/day; child = 200 mg/day)

APPENDIX 4 - RECENT HEALTH EFFECTS STUDIES

Introduction

A significant number of toxicological studies have been conducted since the development of the 1 ppb action level for dioxin and dioxin-like compounds in residential soil. Many of these studies have examined human health effects after known or suspected exposure. In addition, in these intervening years, analytical techniques have been perfected to permit determination of very low levels of dioxin and dioxin-like compounds in environmental and biologic media. Significant advances have also been made in assessing possible health effects associated with exposure. This appendix is a synopsis of this more recent information.

Mechanism of Action

Recent studies have indicated that dioxin and dioxin-like compounds act through the same mechanism of action mediated by the Ah receptor, and that responses to their toxicity have been shown to be similar in several species (Birnbaum, 1994; DeVito et al., 1995).

Human Studies

Direct exposure information is generally not available in human studies, and so body burden is used as a surrogate. In this approach, the exposure is estimated from measured body burden, the elimination rate for humans, and the time since the exposure incident. Positive correlations have been observed between dioxin exposure and cancer (Fingerhut et al., 1991; Zober et al., 1990; Manz et al., 1991). More recent studies on cohorts investigated previously confirmed the association between dioxin exposure and higher cancer mortality (Flesch-Janys et al., 1995; Becher et al., 1996; Ott and Zober, 1996). The correlation was dose-dependent and increased with the latency period. IARC (1997) classified TCDD as a Group I carcinogen (carcinogenic to humans).

For health end points other than cancer, epidemiologic studies suggest a positive correlation between exposure to TCDD and development of chloracne (Mocarelli et al., 1986; Pazderova-Vejlupkova et al., 1981; Reggiani, 1980; Zober et al., 1990), dermal hyperpigmentation and hirsutism (Poland et al., 1971; Jirasek et al., 1974), elevated hepatic enzyme levels, mainly γ -glutamyl transferase (Mocarelli et al., 1986; May, 1982), and increased risk of diabetes (Sweeney et al., 1992; Table 4-1).

Other studies showed an association between development of subtle health effects (e.g., lower vitamin K levels, mild changes in liver enzymes, decreased neurologic optimality, and subtle changes in hormonal levels) in infants and their exposure to dioxin and dioxin-like chemicals from maternal milk (Pluim et al., 1992, 1994a, 1994b; Huisman et al., 1995; Koopman-Esseboom et al., 1994; Table 4-2). It is important to note that in reviewing the issues surrounding breastfeeding, the World Health Organization has concluded that the risks to infants do not outweigh the positive biologic and psychologic aspects of breastfeeding (Johnson, 1992a).

It has been suggested that dioxin and dioxin-like chemicals have the ability to disrupt endocrine function at low levels of exposure. A recent study of the cohort of people exposed during the Seveso accident indicated an alteration of the human sex ratio in their offspring (Mocarelli et al., 1996). In the 7-year period following the exposure, 26 males versus 48 females were born, but the study was limited by not providing information on sex-related spontaneous abortions in the cohort. A study of occupationally exposed workers reported altered reproductive hormone levels (Egeland et al., 1992). Other studies indicate low-exposure contamination of maternal milk with dioxin and dioxin-like compounds may have an impact on the hypothalamic-pituitary-thyroid regulatory system in newborns (Pluim et al., 1992; Koopman-Esseboom et al., 1994).

Animal Studies

Studies in animals demonstrated a wide range of effects associated with CDDs exposure including mortality, cancer, wasting, and hepatic, immunologic, neurologic, reproductive, and developmental effects (ATSDR, 1989). In support of the findings that showed endocrine system disruption in humans, studies in animals reported that TCDD affects the adrenal (DiBartolomeis et al., 1987; Gorski et al., 1988a, 1988b) and thyroid glands (Hermansky et al., 1988; Hong et al., 1987; Lu et al., 1986; Henry and Gasiewicz, 1987; Rozman et al., 1985) and also alters estradiol (Umbreit et al., 1987), testosterone, and dihydrotestosterone levels (Mebus et al., 1987; Moore et al., 1985). TCDD decreased responsiveness of the ventral prostate to testosterone in male offspring of exposed female rats and inhibited sexual differentiation in the central nervous system without altering sexual dimorphism in estrogen-receptor concentrations (Bjerke et al., 1994; Bjerke and Peterson, 1994). In animal studies, effects have been seen with the lowest doses evaluated, with the most sensitive end point being neurobehavioral changes in the offspring of dioxin-exposed monkeys (Schantz et al., 1992). A summary of critical study results and observed effect levels is presented in Table 4-3.

Body Burdens and Associated Health Effects

Health effects reported from human studies and associated body burdens of TCDD are listed in Table 4-1; these body burdens range from concentrations of 18 to 2,357 ng/kg. As can be seen from a comparison of animal and human studies shown in Table 4-3, body burden concentrations calculated for effect dosage rates in animal studies are in the same range as body burden concentrations associated with health effects in human studies. These results underscore the need for research to elucidate the toxicity of dioxin at low doses to human populations (CCEHRP, 1992) and to evaluate exposures in at-risk populations in view of total body burdens of dioxin and dioxin-like compounds.

Based on this review of more recent data, ATSDR has determined that its MRL of 1 pg/kg/day for TCDD is approximately two orders of magnitude below the health effect levels observed in recent studies. This is also true of cancer effect levels (Kociba et al., 1978). Independently, the Health Council of the Netherlands (1996) reassessed the risk associated with dioxin and dioxin-like compounds based on recent studies and recommended a health-based exposure limit equal to 1 pg/kg/day total TEQs.

ATSDR concludes that the chronic oral MRL of 1 pg/kg/day TCDD is protective of public health based on the fact that the MRL is approximately two orders of magnitude below the effect levels demonstrated experimentally and in epidemiologic studies.

TABLE 4-1. Health Effects Associated with Exposure to TCDD and Body Burdens in Humans

Duration of exposure	System	Effect	Body burdens ng/kg body weight	Reference
< 1 year	Dermal	Chloracne in children	2357 ^a	Mocarelli et al., 1991
< 1 year	Reproductive	No increased risk of spontaneous abortion	> 24 ^b	Wolfe et al., 1995
≥ 15 years	Gastrointestinal	No increased risk of clinical gastrointestinal disease	418 ^c	Calvert et al., 1992
≥ 15 years	Hepatic	No increased risk of clinical hepatic disease	418 ^c	Calvert et al., 1992
Not specified	Dermal	Chloracne in 5/7 subjects	80.5 ^d 18 ^e	Schechter et al., 1993
11 years	Dermal	Chloracne	646 ^f	Jansing and Korff, 1994
6.5 years	Immunologic	Immunosuppression	156–176 ^g	Tonn et al., 1996
≥ 15 years	Neurologic	No increased risk for peripheral neuropathy	418 ^c	Sweeney et al., 1993
≥ 15 years	Reproductive	Increased prevalence of high luteinizing hormone and low testosterone levels	31 ^h	Egeland et al., 1994
Not specified	Genotoxicity	No chromosome aberrations or sister chromatid exchanges	63-833 ⁱ	Zober et al., 1993
≥ 1 year	Cancer	Increased cancer mortality risk	124–459 ^j	Fingerhut et al., 1991
≥ 20 years	Cancer	Increased cancer mortality rate	69–461 ^k	Manz et al., 1991

TABLE 4-1. Health Effects Associated with Exposure to TCDD and Body Burdens in Humans (cont'd)

^aCalculated using serum TCDD levels measured shortly after exposure. Body burdens were calculated using body weights of 13 kg for 1–3 year olds, 20 kg for 4–6 year olds, 28 kg for 7–10 year olds, 45 kg for 11-year-old males, and 55 kg for 16-year-old females and body fat percentages of 15% for 0–10 year olds, 15% for 11-year-old males, and 20% for 16-year-old females (ICRP, 1981).

^bCalculated using the reported mean half-life adjusted serum TCDD level of > 110 ppt and assuming the average worker weighed 70 kg with 22% body fat (DeVito et al., 1995). The authors calculated the half-life adjusted serum TCDD level using a half-life of 7.1 years.

^cCalculated using the reported mean half-life adjusted serum TCDD level of 1900 pg/g lipid and assuming the average worker weighed 70 kg with 22% body fat (DeVito et al., 1995).

^dCalculated by averaging the reported individual body burdens divided by the reference body weight of 75 kg for males and 65 kg for females. The authors calculated half-life adjusted serum TCDD levels using the assumption of 75 kg and 65 kg body weights for male and female workers, respectively, and a half-life of 5 years.

^eSame as footnoted, but using a half-life of 10 years.

^fCalculated using the reported mean half-life adjusted serum TCDD level of 2935 pg/g blood fat and assuming the average worker weighed 70 kg with 22% body fat (DeVito et al., 1995). The authors calculated the half-life adjusted serum TCDD level using a half-life of 7 years.

^gCalculated using the reported mean current serum TCDD level of 329.5 pg/g blood lipid. Half-life adjusted serum TCDD level was calculated using a half-life of 11.6 years (Wolfe et al., 1994), background TCDD concentration of 5 ng/kg lipid, and 13–15 years elapsed time. Body burdens were calculated assuming the average worker weighed 70 kg with 22% body fat (DeVito et al., 1995).

^hCalculated using the reported mean current serum TCDD level of 15 ppt. Half-life adjusted serum TCDD levels were calculated using a half-life of 11.6 years (Wolfe et al., 1994), background TCDD concentration of 5 ng/kg lipid, and 34 years of elapsed time. Body burdens were calculated assuming the average worker weighed 70 kg with 22% body fat (DeVito et al., 1995).

Calculated using the reported mean of current serum TCDD levels of 340–472 ppt (based on lipid content of blood). Half-life adjusted serum TCDD levels were calculated using a half-life of 11.6 years (Wolfe et al., 1994), background TCDD concentration of 5 ng/kg lipid, and 35 years of elapsed time. Body burdens were calculated assuming the average worker weighed 70 kg with 22% body fat (DeVito et al., 1995).

ⁱCalculated using the reported mean current serum TCDD level of 233 pg/g lipid. Half-life adjusted serum TCDD levels were calculated using a half-life of 11.6 years (Wolfe et al., 1994), background TCDD concentration of 5 ng/kg lipid, and 35 years of elapsed time. Body burdens were calculated assuming the average worker weighed 70 kg with 22% body fat (DeVito et al., 1995).

^kCalculated using the reported mean current adipose tissue TCDD level of 296 ng/kg. Half-life adjusted adipose TCDD levels were calculated using a half-life of 11.6 years (Wolfe et al., 1994), background TCDD concentration of 5 ng/kg lipid, and 1–33 years of elapsed time.

TABLE 4-2. Breast Milk Levels of Total TEQs Associated with Health Effects in Human Infants

Number of Children	Breast milk levels (mean levels in pg of TEQs per g of milk fat)	Health Effects	References
17	(29.85-92.88)	Late-type hemorrhagic disease of newborns correlated with increased TCDD levels in maternal milk	Koppe et al., 1991
32	29.4 (13.7-62.6)	Decreased vitamin K ₁ and decarboxylated prothrombin levels in infants correlated with increased 2,3,7,8-tetraCDF and 1,2,3,6,7,8-hexaCDF levels, respectively, in breast milk at 11 weeks of age	Pluim et al., 1994a
78	> 30.75	Higher CDD and CDF levels in breast milk correlated with higher plasma levels of TSH in infants in 2nd week and 3rd month postnatally	Koopman-Esseboom et al., 1994
104	30.19	Higher CDD and CDF levels were related to reduced neonatal neurologic optimality	Huisman et al., 1995
48	not specified	Higher exposure to CDDs in breast milk was associated with increase in total T cells and lower monocyte and granulocyte counts	Weisglas-Kuperus et al., 1995
35	28.1 (8.7-62.7)	Cumulative intake correlated with ALT and AST plasma activities; inverse correlation was found between cumulative intake and number of platelets at 11 weeks of age	Pluim et al., 1994b
19	37.5 (29.2-62.7) high exposure group	Increased thyroxine levels and increased thyroxine/thyroid binding globulin ratios in a group with higher breast milk exposure as compared to lower exposure group	Pluim et al., 1992
19	18.6 (8.7-28.0) low exposure group	Baseline control values	Pluim et al., 1992

AST = aspartate aminotransferase; ALT = alanine aminotransferase; TEQs = toxicity equivalents; TSH = thyroid-stimulating hormone

TABLE 4-3. Human Body Burdens and Animal Body Burdens Associated with Health Effects

Duration of exposure	System	Effect	Body burdens ng/kg body weight	Reference
Studies in humans				
< 1 year	Dermal	Chloracne in children	2357 ^a	Mocarelli et al., 1991
Not specified	Dermal	Chloracne in 5/7 subjects	80.5 ^b 18 ^c	Schecter et al., 1993
11 years	Dermal	Chloracne	646 ^d	Jansing and Korff, 1994
6.5 years	Immunologic	Immunosuppression	156-176 ^e	Tonn et al., 1996
≥ 15 years	Reproductive	Increased prevalence of high luteinizing hormone and low testosterone levels	31 ^f	Egeland et al., 1994
≥ 1 year	Cancer	Increased cancer mortality risk	124-459 ^g	Fingerhut et al., 1991
> 20 years	Cancer	Increased cancer mortality rate	69-461 ^h	Manz et al., 1991
Studies in animals				
14 days	Immunologic	Suppressed serum complement in mice	74 ⁱ	*White et al., 1986
90 days	Reproductive	Decreased litter size in rats	26 ^j	*Murray et al., 1979
90 days	Immunologic	Decreased thymus weight in guinea pigs	164 ^k	*DeCaprio et al., 1986
16 months	Developmental	Behavioral alterations in offspring in monkeys	32 ^l	Schantz et al., 1992
2 years	Cancer	Liver, lung carcinoma in rats	2976 ^m	Kociba et al., 1978
2 years	Cancer	Liver carcinoma in mice	944 ⁿ	NTP, 1972

TABLE 4-3. Human Body Burdens and Animal Body Burdens Associated with Health Effects (cont'd)

*Studies which serve as the basis for ATSDR's health guidance values

^aCalculated using serum TCDD levels measured shortly after exposure. Body burdens were calculated using body weights of 13 kg for 1-3 year olds, 20 kg for 4-6 year olds, 28 kg for 7-10 year olds, 45 kg for 11-year-old males, and 55 kg for 16-year-old females and body fat percentages of 15% for 0-10 year olds, 15% for 11-year-old males, and 20% for 16-year-old females (ICRP, 1981).

^bCalculated by averaging the reported individual body burdens divided by the reference body weight of 75 kg for males and 65 kg for females. The authors calculated half-life adjusted serum TCDD levels using the assumption of 75 kg and 65 kg body weights for male and female workers, respectively, and a half-life of 5 years.

^cSame as footnote d but using a half-life of 10 years.

^dCalculated using the reported mean half-life adjusted serum TCDD level of 2935 pg/g blood fat and assuming the average worker weighed 70 kg with 22% body fat (DeVito et al., 1995). The authors calculated the half-life adjusted serum TCDD level using a half-life of 7 years.

^eCalculated using the reported mean current serum TCDD level of 329.5 pg/g blood lipid. Half-life adjusted serum TCDD level was calculated using a half-life of 11.6 years (Wolfe et al., 1994), background TCDD concentration of 5 ng/kg lipid, and 13-15 years elapsed time. Body burdens were calculated assuming the average worker weighed 70 kg with 22% body fat (DeVito et al., 1995).

^fCalculated using the reported half-life adjusted serum TCDD level of > 140 pg/g blood lipid and assuming the average worker weighed 70 kg with 22% body fat (DeVito et al., 1995). The authors calculated the adjusted serum dioxin level using a dioxin half-life of 7.1 years and background dioxin level of 6.08 pg/g blood lipid.

^gCalculated using the reported mean current serum TCDD level of 233 pg/g lipid. Half-life adjusted serum TCDD levels were calculated using a half-life of 11.6 years (Wolfe et al., 1994), background TCDD concentration of 5 ng/kg lipid, and 35 years of elapsed time. Body burdens were calculated assuming the average worker weighed 70 kg with 22% body fat (DeVito et al., 1995).

^hCalculated using the reported mean current adipose tissue TCDD level of 296 ng/kg. Half-life adjusted adipose TCDD levels were calculated using a half-life of 11.6 years (Wolfe et al., 1994), background TCDD concentration of 5 ng/kg lipid, and 1-33 years of elapsed time. Body burdens were calculated assuming the average worker weighed 70 kg with 22% body fat (DeVito et al., 1995).

ⁱAcute exposure study in mice (White et al., 1986). Assumed parameter values $a = 0.8$ (Curtis et al., 1990), $t_{1/2} = 11$ days (Birbaum, 1986).

^jIntermediate-duration exposure study in rats (Murray et al., 1979) Assumed parameter values $a = 0.8$ (Curtis et al., 1990), $t_{1/2} = 24$ days (Van den Berg et al., 1994).

^kAssumed parameter values for guinea pigs in DeCaprio et al. (1986) study: $a = 0.5$ (Van den Berg et al., 1994), $t_{1/2} = 94$ days (Olson, 1986).

^lThe lowest effect level in the current database for chronic-duration exposure. Assumed parameter values for monkeys in Schantz et al. (1992) study: $a = 0.8$ (value for rats from Van den Berg et al., 1994), $t_{1/2} = 391$ days (Bowman et al., 1989).

^mA cancer study in rats. Body burdens calculated in De Vito et al., 1995.

ⁿA cancer study in mice. Body burdens calculated in De Vito et al., 1995.

APPENDIX 5 - CHRONOLOGY FOR DIOXIN AND DIOXIN-LIKE COMPOUNDS:
HEALTH GUIDANCE VALUES AND POLICY STATEMENTS

- 1984 R. Kimbrough, H. Falk, and P. Stehr (1984) recommended 1 ppb of TCDD in soil as a level of concern for human health. They also concluded that "One ppb of 2,3,7,8-TCDD in soil is a reasonable level at which to begin consideration of action to limit human exposure for contaminated soil" (emphasis added) (p. 47). However, the authors cautioned not to use this level for every site, but rather to estimate the risk associated with each site according to specific circumstances at the site.
- The estimated risk dose was 1.4 pg/kg/day TCDD (a 95% upper bound for a one-in-a-million risk estimate for cancer). The calculations were based on cancer studies in laboratory animals.
- 1985 EPA derived oral slope factor, q_1^* , of 1.56×10^5 (mg/kg/day)⁻¹ for TCDD (EPA, 1985) that represents the mean 95% upper-limit carcinogenic potency factor for humans. Based on this factor, a risk-specific dose of 0.006 pg/kg/day TCDD was calculated.
- 1989 ATSDR published the *Toxicological Profile for TCDD*. The profile describes the use of toxicity equivalents (TEQs) for assessing exposure to dioxin and dioxin-like compounds. MRLs for TCDD listed in the profile for the acute, intermediate-duration, and chronic exposures were 100 pg/kg/day, 1 pg/kg/day, and 1 pg/kg/day, respectively. Developmental and reproductive end points were the bases for intermediate and chronic duration MRLs. Based on the chronic MRL, the EMEG of 50 ppt is typically used in public health assessments for dioxin contaminated soil.
- 1990 The Food and Drug Administration (1990) introduced a risk-specific dose of 0.057 pg/kg/day TCDD (a 95% upper bound for a one-in-a-million risk estimate for cancer). The number was based on a linear low-dose extrapolation from the Kociba et al. (1978) cancer study in rats. The value applied to consumption of contaminated food, specifically fish.
- 1992 The Public Health Service Committee to Coordinate Environmental Health and Related Programs (CCEHRP) recommended, in the Interim Statement on Dioxins, to adopt the FDA risk-specific dose (0.057 pg/kg/day) as the risk-specific level for TCDD equivalents (TEQs).
- 1992 In a memo to ATSDR senior management, B. Johnson stated that "The Interim Statement, while mentioning FDA's tolerable daily intake of dioxin as 0.057 pg/kg/day, should not be understood to supplant ATSDR's position of 1 ppb of dioxin in residential soil as a soil action level." Consistent with the CCEHRP statement, ATSDR's practice incorporates the TEQ approach.

- 1993 The *Toxicological Profile for CDDs* was in a draft stage. The internal MRL workgroup met with representatives of other ATSDR divisions and proposed MRLs for TCDD for the acute, intermediate-duration, and chronic exposures as 20 pg/kg/day, 7 pg/kg/day, and 0.7 pg/kg/day, respectively. Developmental effects were the bases for derivation of the chronic MRL.
- 1995 Pohl et al. (1995) published the "Public health assessment for dioxins exposure from soil" paper.
- This paper reviewed more recent findings on the potential health effects of dioxin. Based upon this review, Pohl et al. presented a proposed chronic MRL for TCDD of 0.7 pg/kg/day and a corresponding EMEG of 40 ppt for children.
- From a health risk assessment perspective, the EMEG of 40 ppt is not substantially different from the current EMEG of 50 ppt based on the 1 pg/kg/day MRL (ATSDR, 1989). The MRL of 1 pg/kg/day is approximately two orders of magnitude below effect levels demonstrated experimentally or in epidemiologic studies.

APPENDIX 6 - REFERENCES

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