# STATUS OF THE BLACK ROCKFISH RESOURCE NORTH OF CAPE FALCON, OREGON TO THE U.S.-CANADIAN BORDER IN 2006 



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## Executive Summary

In this document, we included model results from the STAR base model and results based on the "STAT best fit" model, where natural mortality for "old" females is assumed to be 0.24 compared to the assumption of 0.2 in the STAR base model. All other parameter settings remain the same in both models. The "STAT best fit" model is based largely on new and expanded analyses following the conclusion of the STAR Panel. We ran a grid search of natural mortality between 0.1 and 0.3 for "old" females and found that model with natural mortality of 0.24 for "old" females resulted in a better fit to the data with the largest negative change in log likelihood. The mortality of 0.24 agreed with a direct estimate of female natural mortality at $0.27(\mathrm{SE}=0.26)$ from historical catch, effort, and length frequency data. We felt compelled to integrate these results because the "Low Natural Mortality" model selected by the STAR panel to bracket model uncertainty does not appear plausible. Further, we believe that management should be based on the "STAT best fit" model because it represents the best fit to data, and the STAR base and "High Natural Mortality" models be used to bracket the uncertainty.

## Stock

This assessment applies to the Northern portion of the black rockfish (Sebastes melanops) stock found between Cape Falcon, Oregon and the U.S. border with Canada. This assessment treats these fish as a separate unit stock. The stock found South of Cape Falcon, Oregon is treated as another unit stock in a different assessment document. Black rockfish are not subjected to a targeted fishery in Canadian coastal waters and are not assessed.

## Catches

Little information exists on the historical landings of black rockfish prior to the early 1960's. Landings of "rockfish" peaked at nearly 25,000 mt in 1945 in support of the war effort; however, there is no known species composition estimates for these catches. Due to the nearshore habitat of this species it is likely that very little of this catch was black rockfish. Predominate harvesters of black rockfish between 1963 and 1983 were commercial line and trawl fishers. Black rockfish trawl landings typically came from directed tows on nearshore rocky reefs and shipwrecks with few landings incidental to other targeted fisheries. Peak landings in the trawl fishery reached 350 mt in 1976 and declined to less than 10 mt in recent years. Black rockfish comprised less than $1 \%$ of total rockfish landings by the trawl fishery during this period.

The "non-trawl" fishery is composed of three distinct line fisheries, and each differs in target species. Oregon and Washington fish receiving tickets show nominal rockfish catches as early as 1970 in the salmon troll fishery, during 1973 in the jig fishery, and during 1979 in the bottomfish troll fishery. Black rockfish are generally caught as bycatch in the commercial salmon troll fishery; landings peaked in the late 70 's ( 151 mt ) and steadily decreased coincident with losses in fishing opportunities for coastal salmon. The bottomfish troll fishery generally targeted lingcod; rockfish landings were small and estimated black rockfish catch never exceeded 2 mt . The jig fishery is primarily composed of small vessels less than 26 feet in length that generally fish near their port of
access. Black rockfish were targeted in nearshore areas and were a significant fraction of the nominal rockfish landings in the jig fishery. Black rockfish catch in the jig fishery was inconsequential prior to 1980, and peaked in 1982 at 272 mt . Since 1996, nominal rockfish landings have contained no black rockfish due to area restrictions that have forced jig fishers to target other rockfish species found farther offshore.

Black rockfish are the primary target of the coastal groundfish sport fishery, with small catches first reported in the late 1970's that steadily increased to over 300 ton per year by the mid 1990's. Due to the implementation of a 10 fish bag limit in 1995 (Figure 7), and longer salmon seasons, annual catches of black rock declined to 188 mt in 2001. In recent years, sport catches increased to more than 300 mt . The coastal recreational rockfish fishery generally competed with sport salmon, halibut and tuna fisheries, and this is reflected in year-to-year variability in black rockfish catch.

Discard of black rockfish in Washington waters in either the commercial or recreational fisheries is likely very small. "Sebastes complex" trip limits in the line fishery were nonrestrictive prior to 1999 since few landings ever achieved the trip limit, and there was no incentive to discard catch. Furthermore, Washington State waters (inside 3 miles) have been closed to directed non-trawl commercial fishing since 1996 and directed trawl fishing since 1999. Black rockfish represented only a small fraction of the nominal rockfish catch in the trawl fishery and it is unlikely they were discarded. Discard in the sport fishery is also insignificant since the vast majority of recreational fishers do not high-grade their rockfish catch. This is supported by recent sport fishery information that indicates discard is less than 16 mt on an annual basis.

Recent Black Rockfish Landings From Waters North of Cape Falcon, Oregon to the US-Canadian Border by Gear and Area

|  | Trawl Gear |  |  |  | Non-Trawl Gear |  |  | Recreational |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3A | 3B | 3CS | Total | 3A | 3B | Total | 3A | 3B | Total |
| 1995 | 2.9 | 0.1 | 0.3 | 3.3 | 2.7 | 63.1 | 65.8 | 209.3 | 55.5 | 264.8 |
| 1996 | 0.0 | 0.0 | 0.0 | 0.0 | 4.8 | 3.8 | 8.6 | 199.7 | 64.6 | 264.2 |
| 1997 | 0.7 | 8.2 | 0.1 | 9.0 | 14.5 | 0.5 | 15.0 | 179.7 | 54.4 | 234.1 |
| 1998 | 72.5 | 0.3 | 0.3 | 73.1 | 0.4 | 4.5 | 4.8 | 195.2 | 64.2 | 259.4 |
| 1999 | 0.0 | 0.0 | 0.0 | 0.0 | 3.4 | 0.9 | 4.3 | 166.0 | 55.6 | 221.6 |
| 2000 | 0.0 | 0.0 | 0.0 | 0.0 | 0.5 | 0.7 | 1.2 | 157.6 | 67.2 | 224.8 |
| 2001 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 0.5 | 1.1 | 133.7 | 55.0 | 188.7 |
| 2002 | 0.1 | 0.1 | 0.0 | 0.2 | 0.4 | 1.0 | 1.5 | 173.0 | 66.0 | 238.9 |
| 2003 | 0.1 | 0.0 | 0.0 | 0.1 | 0.0 | 0.2 | 0.2 | 166.7 | 70.4 | 237.1 |
| 2004 | 0.6 | 0.0 | 0.0 | 0.6 | 0.4 | 0.3 | 0.7 | 173.4 | 94.6 | 268.0 |
| 2005 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 0.6 | 0.9 | 217.5 | 114.2 | 331.7 |
| 2006 | 1.2 | 0.2 | 0.0 | 1.4 | 0.8 | 0.4 | 1.2 | 246.7 | 74.9 | 321.5 |
| Total | 78 | 9 | 1 | 88 | 29 | 77 | 105 | 2,218 | 837 | 3,055 |



## Data and Assessment

This portion of the U.S. black rockfish stock was last assessed in 1999 (Wallace et al. 1999) with a population dynamics model constructed with AD model builder software (Fournier1997).

The current assessment employed Stock synthesis 2 (SS2V2.00c, compiled 3/27/2006) to model the dynamics of the black rockfish population found between Cape Falcon, Oregon and North to the U.S./Canadian border in Coastal Waters. The model was specified to begin in 1915 to ensure population equilibrium at the start of the modeling time period. Catch data were decayed from the last reliable catch estimates (1965) to 0 by 1940 . Fisheries catch, size and age compositions were pooled into three fishery types including trawl, sport and non-trawl. The first size-age compositions were collected in the mid 1970's from the trawl fishery, but samples were not collected on a systematic basis until 1985. Growth (Lmin, Lmax and k) was estimated within the model to account for fishery selection of the larger individual fish at age. The population model was tuned to two fisheries-independent indices that include a tagging CPUE (1986-2007) and a tag abundance biomass index (2000-2007), both derived from WDFW black rockfish tagging information. Both STAT and STAR Panel members agreed that the available fishery dependent indices should not be incorporated due to potential bias resulting from bag limit changes and undocumented measures of fishing effort resulting from changes in search time across the time series.

## Unresolved Problems and Major Uncertainties

Natural mortality is confounded with fishing mortality and is therefore one of the most challenging biological parameters to estimate. It is also one of the most important parameters due to its affects on population dynamics, including stock rebuild time and the estimation of virgin fishery biomass. In this assessment, we explored direct methods to estimate natural mortality and compared it to estimates derived from indirect methods (from other biological parameters, e.g., the growth constant and fecundity) in previous assessments. The estimated $\hat{M}$ derived from direct methods was 0.223 ( $\mathrm{SE}=0.0071$ ) and 0.272 ( $\mathrm{SE}=0.061$ ) for males and females, respectively. Given the uncertainties, these estimates compared well with other existing indirect methods. The current base model assumes a female natural mortality rate to be age-specific for females using age at first and full maturity for inflections (10 and 15). A constant natural mortality rate of 0.16 was assumed for males and young females ( $<10$ years of age), and a rate of 0.2 was assumed for old females (>=15 years of age). This is higher than that used in the 2003 black rockfish assessment off Oregon and California (Ralston and Dick 2003) which used a natural mortality of 0.1 and 0.2 for males and old females, respectively. It is apparent from our analysis using both direct and indirect methods that our current assumptions on natural mortality in the base model are within our limits to estimate this parameter and that the low natural mortality rate model is likely too low. Model sensitivity analysis showed that model configurations using higher natural mortality for older females provided better overall fits to the data than the STAR base model.

Changes in total log likelihood relative to Base model


Tagging is not incorporated in the model as a tagging experiment, which is not possible within the current SS2 modeling framework. The index for tagging abundance is not fit well, and the model estimated effective $q$ for the tagging index was 0.83 . This is likely double what it should be based on STAT knowledge of available habitat off the Washington coast. Further, the north central Washington coast, where most of the nearshore rocky habitat exists, is inaccessible to most recreational fishers and is not part of the current tagging program. However, the estimation of $q$ is complicated by the fact that the SS2 value of $q$ is a function of selectivity that is strongly dome shaped for the fishery. Increasing the weighting on survey abundance shows that a better fit to the survey abundance index significantly improves our view of the current population status.


Without an objective evaluation of an informed prior on $q$, it is difficult to compare a prior conception of $q$ based on tagging and the one estimated by SS2. Other issues include the non-independence of the length/age compositions and non-independence of the tagging abundance and CPUE series.

## Reference Points

The Pacific Fisheries Management Council recommends that a default target fishing mortality rate of FSPR=0.5 be used for Council managed rockfish species. The current assessment uses this default for harvest projections for black rockfish and based on the Councils control rule for groundfish would not be considered overfished. The "STAR base" represents results from the STAR base model and the "best fit" model represents results from the best fit model incorporated by the STAT in the decision matrix postSTAR.

STAR Base Model Reference Points

| Unfished Stock | Value |
| :---: | :---: |
| Age 3+ Biomass $\left(\mathrm{B}_{0}\right)(\mathrm{mt})$ | 10,813 |
| Spawning Biomass $\mathrm{SB}\left({ }_{0}\right)(\mathrm{mt})$ | 2,429 |
| SPBio/Recruit $(\mathrm{kg} / \mathrm{fish})$ | 0.780 |
| Age1 Recruitment $\left(\mathrm{R}_{0}\right)(1,000 ' \mathrm{~s})$ | 3,113 |
| Steepness_R0_S0 | 0.6 |


| Exploited Stock | Reference points based on |  |  |
| :---: | :---: | :---: | :---: |
|  | Estimated MSY | SB ${ }_{40 \%}$ | SPR (SB ${ }_{0.5}$ ) |
| SPR (Spawning Biomass/Recruit) | 0.413 | 0.400 | 0.400 |
| F (Fishing Mortality Rate) | 0.132 | 0.101 | 0.101 |
| Exploitation Rate (Yield/Bsmry) | 0.076 | 0.060 | 0.060 |
| MSY (mt) or MSY proxy (mt) | 377 | 361 | 361 |
| Yield (mt) | 718 | 972 | 972 |
| SPBIO/SB(0) | 29.6\% | 40.0\% | 40.0\% |
| Age 3+ Biomass | 4,947 | 6,012 | 6,012 |

## STAT Best Fit Model Reference Points

| Unfished Stock | Value |
| :---: | :---: |
| Age 3+ Biomass $\left(\mathrm{B}_{0}\right)(\mathrm{mt})$ | 11,390 |
| Spawning Biomass $\mathrm{SB}\left({ }_{0}\right)(\mathrm{mt})$ | 2,321 |
| SPBio/Recruit $(\mathrm{kg} / \mathrm{fish})$ | 0.687 |
| Age1 Recruitment $\left(\mathrm{R}_{0}\right)(1,000$ 's $)$ | 3,377 |
| Steepness_R0_S0 | 0.6 |


|  | Reference points based on |  |  |
| :--- | :---: | ---: | ---: |
| Exploited Stock | Estimated MSY | SB $_{40 \%}$ | SPR (SB $_{\mathbf{0 . 5}}$ ) |
| SPR (Spawning Biomass/Recruit) | 0.418 | 0.400 | 0.40 |
| F (Fishing Mortality Rate) | 0.141 | 0.110 | 0.110 |
| Exploitation Rate (Yield/Bsmry) | 0.081 | 0.065 | 0.065 |
| MSY (mt) or MSY proxy (mt) | 423 | 408 | 408 |
| Yield (mt) | 700 | 928 | 928 |
| SPBIO/SB(0) | $30.1 \%$ | $40.0 \%$ | $40.0 \%$ |
| Age 3+ Biomass | 5,218 | 6,264 | 6,264 |

## Stock Biomass

The estimated current spawning biomass resulting from the STAR base model was 1,034 mt and unexploited spawning biomass is $2,429 \mathrm{mt}$, resulting in a current stock level that is $42.6 \%$ of the unfished. The STAT best fit model estimates current spawning biomass as being $1,239 \mathrm{mt}$ and unexploited spawning biomass at $2,321 \mathrm{mt}$, resulting in a current stock level that is $53.4 \%$ of the unfished. In both models spawning biomass and age $3+$ biomass reached the lowest levels in 1995, following poor recruitment and intense fishing in the late 1980's.

STAR Base Model Results

| Year | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Spawning Biomass | 612 | 652 | 701 | 754 | 809 | 880 | 938 | 985 | 1016 | 1034 |
| \% of Virgin | 0.252 | 0.268 | 0.289 | 0.310 | 0.333 | 0.362 | 0.386 | 0.405 | 0.418 | 0.426 |
| Age 3+ Biomass | 5069 | 5107 | 5146 | 5433 | 5594 | 6133 | 6178 | 6143 | 6204 | 6180 |



STAT "Best Fit" Model Results

| Year | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Spawning Biomass | 707 | 762 | 826 | 891 | 959 | 1043 | 1114 | 1171 | 1211 |
| \% of Virgin | 0.304 | 0.328 | 0.356 | 0.384 | 0.413 | 0.449 | 0.480 | 0.505 | 0.522 |
| Age 3+ Biomass | 5977 | 6066 | 6147 | 6516 | 6739 | 7405 | 7485 | 7470 | 7564 |



## Recruitment

Recent increases in biomass are the result of two prominent year classes in 1994 and in 1999. The 1999-year class is estimated to be the largest year class since the beginning of the estimation phase.

STAR Base Model Results

| Year | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age 1 Recruits (1,000's) | 2,614 | 2,239 | 4,478 | 1,997 | 1,696 | 2,414 | 2,468 | 2,509 | 2,535 | 2,550 |



## STAT "Best Fit" Model Results

| Year | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age 1 Recruits (1,000's) | 3,129 | 2,732 | 5,410 | 2,444 | 2,075 | 2,826 | 2,882 | 2,924 | 2,951 | 2,970 |



## Exploitation Status

Exploitation of black rockfish reached a peak in 1988 of 13\% of the Age 3+ biomass and remained near that level for 7 years, dropping precipitously between 1995 and 2000. In recent years exploitation has been relatively low (4-6\%).

STAR Base Model Results

| Recent trends in black rockfish exploitation <br> Year$\quad 1997$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Total Exploitation Rate | 0.0501 | 0.0418 | 0.0326 | 0.0323 | 0.027 | 0.0334 | 0.033 | 0.0368 | 0.0448 |



## STAT "Best Fit" Model Results

| Recent Trends in black rockfish exploitaion <br> Year | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ |
| :--- | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total Exploitation Rate | 0.042 | 0.035 | 0.027 | 0.027 | 0.022 | 0.028 | 0.027 | 0.030 | 0.037 |



Black rockfish stock abundance has been below the Councils' management target and results from the STAR base model indicates that it has dipped below the Councils’ minimum stock size threshold in the last decade. The stock is currently above the management target of B40\% in both the STAR base and STAT best fit models.

STAR Base Model


STAT "Best Fit" Model


Exploitation rate relative to spawning biomass indicate that harvest rates exceeded management targets between the mid 1980's through the mid 1990's. The STAT best fit model indicates a slightly improved exploitation time series.

STAR Base Model


STAT "Best Fit" Model


## Management Performance

Harvest has remained well below the harvest guideline of 517 mt (1997-1999) and the 577 mt (+- 2CV's 523-632 mt’s) equilibrium catch following the 1994 (Wallace et al., 1994) and the 1999 assessment (Wallace et al., 1999), respectively. The 1999 assessment estimated the 2001 spawning biomass of 646 mt (+- 2CV's 601-687 mt's) with an equilibrium spawning biomass of 451 mt (+- 2CV's 401-501 mt’s) equating to a 2001 $\mathrm{SB}_{2001} / \mathrm{SB}_{\text {Equil }}$ of $143 \%$. The catch time series includes discard when existing, ABC is constant and changes in spawning biomass across the time series is not available.

There were no explicit ABC’s for the northern area until 2004. Prior this time (for the period 2000 -2003), yield from the northern assessment was added to catches from the southern, unassessed area to produce a coastwide ABC of $1,115 \mathrm{mt}$ ( 615 mt from the N . assessment plus 500 mt of catch from the south). In 2004, a management line was implemented at the Columbia River, separating Washington and Oregon. Since the assessment extended to Cape Falcon, the GMT transferred a portion of the yield from the northern assessed area to the south to account for the portion of the stock (yield) from the Columbia River to Cape Falcon, $88 \%$ to the north, $12 \%$ to the south. This resulted in an ABC for Washington (Columbia River to the Canadian Border) of 540 mt . This has been (will be) constant from 2004 through 2008. With regard to management performance, catches have remained below both the northern portion of the coastwide ABC assumed from the assessment as well as the explicit northern ABC beginning in 2004
Total black rockfish catch by all fisheries

|  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 1997 | 1998 | 1999 | 2000 | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| Total Catch (mt) | 258.1 | 337.3 | 225.9 | 226 | 189.8 | 240.6 | 237.4 | 269.3 | 332.6 | 324.1 |

## Forecasts

Projections of future catches were based on a $\mathrm{F}_{\text {SPR } 50 \%}$ rate of fishing mortality. We also assumed that the sport fishery would account for $100 \%$ of the catch and that selectivity would remain unchanged from that estimated within the model in the final year. For the STAR Base model only, beginning in 2013, there is a slight downward adjustment in ABC of $\sim 1 \%$ to account for 40:10 harvest Control rule adjustments.

STAR Base Model

| Year | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 9}$ | $\mathbf{2 0 1 0}$ | $\mathbf{2 0 1 1}$ | $\mathbf{2 0 1 2}$ | $\mathbf{2 0 1 3}$ | $\mathbf{2 0 1 4}$ | $\mathbf{2 0 1 5}$ | $\mathbf{2 0 1 6}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| ABC (mt) | 394 | 377 | 361 | 350 | 345 | 344 | 346 | 350 | 354 | 357 |
| Spawning Biomass (mt) | 1064 | 1071 | 1060 | 1036 | 1005 | 977 | 956 | 944 | 940 | 943 |
| \% of Virgin | 0.438 | 0.441 | 0.436 | 0.426 | 0.414 | 0.402 | 0.394 | 0.389 | 0.387 | 0.388 |

STAT "Best Fit" Model

| Year | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 9}$ | $\mathbf{2 0 1 0}$ | $\mathbf{2 0 1 1}$ | $\mathbf{2 0 1 2}$ | $\mathbf{2 0 1 3}$ | $\mathbf{2 0 1 4}$ | $\mathbf{2 0 1 5}$ | $\mathbf{2 0 1 6}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| ABC (mt) | 535 | 503 | 474 | 453 | 440 | 433 | 431 | 432 | 434 | 436 |
| Spawning Biomass (mt) | 1281 | 1267 | 1233 | 1182 | 1126 | 1074 | 1033 | 1005 | 989 | 984 |
| \% of Virgin | 0.552 | 0.546 | 0.531 | 0.509 | 0.485 | 0.463 | 0.445 | 0.433 | 0.426 | 0.424 |

## Decision Table

The decision table matrix was developed through STAR Panel and STAT discussions. Three states of nature were defined in terms of natural mortality: 1) M equals to 0.12 for all males and females $<=10$ years of age, and $M$ linearly increases from 0.12 to 0.16 for females age 11 to 15 then remains constant at 0.12 after age $15 ; 2$ ) M equals to 0.16 for males and females $<=10$ years of age, and M linearly increases from 0.16 to 0.20 for females age 11 to15 then remains constant at 0.20 after age 15 ; and 3) $\mathrm{M}=0.19$ for males and females <=10 years of age, and M linearly increases from 0.19 to 0.23 for females age 11 to age 15 , then remains constant at 0.23 after age 15 . To assess the affect of alternative management actions, harvest was forecast with alternative catch levels derived from each state of nature.

In addition to the above three states of nature, we included model results in the decision matrix that are based on the "best fit" model where $\mathrm{M}=0.16$ for males and females <=10 years of age, and M for females linearly increasing from age 11 to age 15 to 0.24 , and then constant. The STAT feels compelled to integrate these results into the decision matrix (post STAR) because the "Low Natural Mortality" model does not appear plausible. Further, we consider the STAR base model as a very conservative representation of the current population. The STAT recommends that the "Best Fit" model be used for management recommendations and the "STAR Base Model" and the "High Natural Mortality Model" be used to bracket the uncertainty. Our evaluation is based on sensitivity analysis, comparison of model results to the tagging study, and general observations we have made in the fishery that include:

1) the assumed rate of natural mortality in the "Low Natural Mortality" state of nature is lower than any previous assessment for the "Northern" population, and is lower than any external estimation by direct and indirect methods,
2) biomass results from the "Low Natural Mortality" indicate that the population declined to less than $13 \%$ of the unfished population in the mid-1990's yet we have no indication from the fishery or from our tagging study that there was localized depletion during this time period,
3) sensitivity analyses indicate "Low Natural Mortality" model fit to the data is very poor relative to other model results that assume a higher rates of natural mortality,
4) the estimated q for the survey is likely double what it should be based on STAT knowledge of available habitat off the Washington coast,
5) tagging data are not fit well and tagging estimates external to the model indicate that the population is larger and fishing mortality is lower compared to STAR base model run results,
6) other model runs with higher steepness and Sigma R fit the data better and improved our view of the current population status above both the STAR base and "Best Fit" model runs and finally,
7) compared to the STAT best fit model, a model with high natural mortality for females (where $\mathrm{M}=0.16$ for males and females $<=10$ years of age and M for females linearly increasing from age 11 to age 15 to 0.26 ) fit the data equally well. This model resulted in an improved view of current population status above both the STAR base and "Best Fit" model runs. However, results from this model
were not incorporated in the decision table because the higher natural mortality on females (0.26) fell outside the range considered at the STAR Panel.

STAR and STAT decision matrix based on a range of natural mortality rates where rows represent results from the assumed natural mortality model given catch rates that resulted from alternative states of nature (columns).

Decision Table - 2007 Northern Black Rockfish Assessment

| State of Nature | Year | ABC | $\begin{aligned} & M=0.12 \text { Males } \\ & M=0.16 \text { Females } \end{aligned}$ <br> Low Natural Mortality SpawnBio Depletion |  | $\begin{aligned} & M=0.16 \text { Males } \\ & M=0.20 \text { Females } \\ & \text { STAR Base Model } \\ & \text { SpawnBio Depletion } \end{aligned}$ |  | $\begin{aligned} & \hline M=0.16 \text { Males } \\ & M=0.24 \text { Females } \\ & \text { Best Fit } \\ & \text { SpawnBio Depletion } \end{aligned}$ |  | $\begin{aligned} & \hline M=0.19 \text { Males } \\ & M=0.23 \text { Females } \\ & \text { High Natural Mortality } \\ & \text { SpawnBio Depletion } \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2007 | 108 | 320 | 14.1\% | 320 | 14.1\% | 320 | 14.1\% | 320 | 14.1\% |
| 근 | 2008 | 96 | 287 | 12.6\% | 287 | 12.6\% | 287 | 12.6\% | 287 | 12.6\% |
| Ti | 2009 | 86 | 246 | 10.8\% | 246 | 10.8\% | 246 | 10.8\% | 246 | 10.8\% |
| ${ }^{\circ}$ | 2010 | 100 | 279 | 12.3\% | 194 | 8.5\% | 163 | 7.1\% | 99 | 4.4\% |
| \% | 2011 | 115 | 316 | 13.9\% | 152 | 6.7\% | 96 | 4.2\% | 26 | 1.1\% |
| $\stackrel{1}{5}$ | 2012 | 129 | 359 | 15.8\% | 120 | 5.3\% | 48 | 2.1\% | 13 | 0.6\% |
| 艺 | 2013 | 140 | 403 | 17.7\% | 96 | 4.2\% | 18 | 0.8\% | 10 | 0.4\% |
| 3 | 2014 | 148 | 447 | 19.6\% | 77 | 3.4\% | 11 | 0.5\% | 9 | 0.4\% |
| 인 | 2015 | 153 | 486 | 21.4\% | 58 | 2.6\% | 9 | 0.4\% | 9 | 0.4\% |
|  | 2016 | 156 | 518 | 22.8\% | 39 | 1.7\% | 8 | 0.4\% | 8 | 0.4\% |
|  | 2007 | 394 | 1064 | 43.8\% | 1064 | 43.8\% | 1064 | 43.8\% | 1064 | 43.8\% |
|  | 2008 | 382 | 1088 | 44.8\% | 1088 | 44.8\% | 1088 | 44.8\% | 1088 | 44.8\% |
| \% | 2009 | 370 | 1092 | 44.9\% | 1092 | 44.9\% | 1092 | 44.9\% | 1092 | 44.9\% |
| $\sum^{0}$ | 2010 | 358 | 1139 | 46.9\% | 1065 | 43.8\% | 1030 | 42.4\% | 959 | 39.5\% |
| $\stackrel{0}{0}$ | 2011 | 351 | 1175 | 48.4\% | 1032 | 42.5\% | 965 | 39.7\% | 833 | 34.3\% |
| ¢ | 2012 | 349 | 1204 | 49.6\% | 1000 | 41.2\% | 906 | 37.3\% | 724 | 29.8\% |
| ~ | 2013 | 350 | 1232 | 50.7\% | 976 | 40.2\% | 860 | 35.4\% | 637 | 26.2\% |
|  | 2014 | 352 | 1260 | 51.8\% | 959 | 39.5\% | 825 | 34.0\% | 571 | 23.5\% |
| $\omega$ | 2015 | 356 | 1289 | 53.1\% | 952 | 39.2\% | 803 | 33.0\% | 524 | 21.6\% |
|  | 2016 | 360 | 1321 | 54.4\% | 952 | 39.2\% | 790 | 32.5\% | 490 | 20.2\% |
|  | 2007 | 535 | 1281 | 55.2\% | 1281 | 55.2\% | 1281 | 55.2\% | 1281 | 55.2\% |
|  | 2008 | 521 | 1317 | 56.7\% | 1317 | 56.7\% | 1317 | 56.7\% | 1317 | 56.7\% |
|  | 2009 | 505 | 1328 | 57.2\% | 1328 | 57.2\% | 1328 | 57.2\% | 1328 | 57.2\% |
| $\pm$ | 2010 | 478 | 1376 | 59.3\% | 1304 | 56.2\% | 1270 | 54.7\% | 1202 | 51.8\% |
| 4 | 2011 | 459 | 1406 | 60.6\% | 1268 | 54.6\% | 1204 | 51.9\% | 1076 | 46.4\% |
| $\stackrel{\rightharpoonup}{0}$ | 2012 | 448 | 1425 | 61.4\% | 1230 | 53.0\% | 1140 | 49.1\% | 964 | 41.5\% |
| $\infty$ | 2013 | 443 | 1440 | 62.0\% | 1198 | 51.6\% | 1087 | 46.8\% | 873 | 37.6\% |
|  | 2014 | 441 | 1456 | 62.7\% | 1174 | 50.6\% | 1048 | 45.2\% | 805 | 34.7\% |
|  | 2015 | 442 | 1474 | 63.5\% | 1162 | 50.1\% | 1023 | 44.1\% | 756 | 32.6\% |
|  | 2016 | 443 | 1498 | 64.6\% | 1159 | 49.9\% | 1010 | 43.5\% | 725 | 31.2\% |
|  | 2007 | 827 | 2075 | 71.8\% | 2075 | 71.8\% | 2075 | 71.8\% | 2075 | 71.8\% |
| $\underset{=}{\lambda}$ | 2008 | 804 | 2137 | 73.9\% | 2137 | 73.9\% | 2137 | 73.9\% | 2137 | 73.9\% |
| 픚 | 2009 | 775 | 2161 | 74.8\% | 2161 | 74.8\% | 2161 | 74.8\% | 2161 | 74.8\% |
| $\frac{\overline{0}}{\Sigma}$ | 2010 | 714 | 2206 | 76.3\% | 2132 | 73.8\% | 2096 | 72.5\% | 2025 | 70.1\% |
| T | 2011 | 671 | 2221 | 76.8\% | 2079 | 71.9\% | 2012 | 69.6\% | 1880 | 65.1\% |
| 年 | 2012 | 642 | 2219 | 76.8\% | 2019 | 69.9\% | 1926 | 66.7\% | 1744 | 60.4\% |
| $\stackrel{\pi}{\pi}$ | 2013 | 624 | 2210 | 76.5\% | 1963 | 67.9\% | 1850 | 64.0\% | 1629 | 56.4\% |
|  | 2014 | 615 | 2204 | 76.3\% | 1919 | 66.4\% | 1790 | 61.9\% | 1539 | 53.3\% |
| 흦 | 2015 | 610 | 2204 | 76.3\% | 1889 | 65.4\% | 1747 | 60.5\% | 1474 | 51.0\% |
| $\pm$ | 2016 | 607 | 2212 | 76.5\% | 1872 | 64.8\% | 1721 | 59.6\% | 1431 | 49.5\% |

Note:

1. The natural mortality rate of "young" females <= 10 years of age and males are equal. The natural mortality rate for "old" females between the ages of 11 and 15 is linearly increasing and then remains at the constant rate listed above. Assumed catch of 325 mt in 2007 and 2008.
2. $A B C$ for 2007 and 2008 in the current annual management specifications is 540 mt for the area north of the Columbia River. Since the assessment extends south to Cape Falcon, Oregon, the $A B C$ in regulation is a result of apportioning the $615 \mathrm{mt} A B C$ from the previous assessment north and south of the Columbia River.

## Research and Data Needs

In order to objectively evaluate a prior on q for the tagging, information on habitat distribution within the stock boundary is necessary. A nearshore assessment should be completed using side-scan, backscatter and multi beam methods. This has already been completed for some portions of the coast and new information can be integrated.

## Rebuilding Projections

None required.

## Regional Management Concerns

Black rockfish is highly resident to specific reefs and are therefore susceptible to localized depletion especially during times of population decline. Because of this, relatively higher levels of abundance may be needed to meet recreational fishery objectives. For example, the recreational fishery industries need to maintain a sufficient success rate to be economically feasible.

### 1.0 INTRODUCTION

The status of stocks for the "Northern" black rockfish stock found between Cape Falcon, Oregon and the U.S. Canadian border was last determined in 1999 (Wallace et al, 1999). The population was assessed using an AD model configuration where tag recovery was modeled explicitly. The population was regarded as healthy, stock abundance was estimated to be slightly increasing after passing through a low in the late 1980's and early 1990's. The recommended allowable annual yield was 577 mt based on an F45\% exploitation strategy and a tag recovery rate of $50 \%$. The estimated stock biomass ranged between 9,500-10,100 mt , depending on assumptions on tag reporting rates. The current analysis reprises estimates based on the 1999 model that uses an improved stock synthesis program (SS2) (Methot, 2006) and presents a completely new model specification. This assessment is distinguished from other more southerly black rockfish population assessment(s) by identifying it as the "northern" stock. However, we have no indication that there is any stock divide at the U.S.-Canadian border just that this assessment includes information only as far north as the U.S.-Canadian border.

Throughout the document we include model results that are based on both the "STAT best fit" model and the STAR base model. STAT best fit model natural mortality for "old" females is assumed to be 0.24 versus 0.20 in the STAR base model and all other parameter settings remain the same. Results in the STAT best fit model are based largely on new and expanded analyses following conclusion of the STAR Panel. We felt compelled to integrate these results because the "Low Natural Mortality" model used to bracket model uncertainty does not appear plausible and the STAT best fit model provided a better fit to the tagging and age composition data.

### 1.1 Species Distribution, Stock Structure, and Management Units

Black rockfish (Sebastes melanops) are widely distributed along the Pacific coast from central California to the Gulf of Alaska inhabiting nearshore areas at bottom depths of less than 50 fathoms (Miller and Lea, 1972). Adults are schooling and associated with irregular, rocky bottom or underwater structures, though at times may be found actively feeding on the surface.

Washington tagging data suggest that Cape Flattery and Cape Falcon may represent area bounds for a coastal Washington-northern Oregon black rockfish stock. From over 54,000 tag releases in this area, no fish were recovered north of the Strait of Juan de Fuca and only 6 were recovered south of Cape Falcon in the 15-year study (Figure 1). To corroborate these results, a genetic stock identification study of coastal black rockfish populations was conducted from 1995-1997 \{\}(WDFW report in progress). Horizontal starch-gel electrophoresis was used to examine 10 black rockfish collections from northern California, Oregon, Washington and southern British Columbia. Significant heterogeneity occurred among Oregon collections, while less heterogeneity was found among Washington collections. Dendrogram and multidimensional scaling (MDS) analysis of genetic distances (Nei, 1978) revealed three major geographical groupings (Figure 2). The groups include samples from: 1) north of Cape Falcon, 2) south of Cape

Falcon off the Oregon coast, and 3) a single collection from northern California (Port Albion). The study concluded that there is an apparent large-scale geographical clustering of coastal black rockfish populations and there does not appear to be any geographical pattern to clustering of populations within each group. For this assessment, we assume that black rockfish distributed between Cape Falcon, Oregon and Cape Flattery, Washington represent a unit stock. All biological parameters, data analysis and yield projections presented in this assessment are intended to describe this portion of black rockfish coast-wide distribution.

It is interesting to note that although no black rockfish tags were recovered from southern British Columbia during the 15 year tagging study, fish collected just 20 km north of Cape Flattery in Barkley Sound, B.C. were genetically similar to the coastal Washington collections. The lack of recoveries from across the Strait of Juan de Fuca is likely due to a lack of any target fisheries in coastal B.C. waters or may indicate that the Strait provides an effective physical boundary, which few if any adult black rockfish will cross. Nearshore and oceanic drift likely influence gene flow during the three to four month planktonic stages. Survival during the early life stages is strongly influenced by oceanic processes and recruitment may be dependent upon the health of black rockfish populations both north and south of the Strait of Juan de Fuca.

### 1.2 Life History Overview

Like many rockfish species, black rockfish are slow growing, long lived and mature late in life. Black rockfish are recruited into the commercial and sport fishery at 4 years of age; age composition of the catch can span three decades. Early recruitment, delayed maturity and schooling behavior make black rockfish susceptible to over-exploitation. Furthermore, WDFW found evidence that, in at least one year, a number (approximately $10 \%$ ) of mature females examined during parturition did not spawn during year of collection. Ovarian characteristics derived from histological preparations on these specimens indicated that although they had spawned in prior seasons, they had not advanced beyond the early yolk accumulation stage and were re-absorbing their oocytes. Thus, some fraction of the mature population may not spawn annually. If this behavior were common from year to year production would be accordingly reduced.

Another important aspect of black rockfish life history is differences in growth and apparent natural mortality rates between sexes. Composition sampling data show that the sex ratio before age 10 is nearly equal and then the percent female declines sharply thereafter (Figure 3). For the purposes of this assessment we interpret the loss of females due to increased natural mortality at age, which coincides with female transition into sexual maturity.

### 1.3 Review of Fishery

Recreational and commercial fishers have harvested black rockfish in nearshore areas off the Washington coast since the early 1960's. Commercial fisheries include salmon and bottomfish troll, jig and groundfish trawl. The recreational fishery is divided between
charter and private boat operations. Due to restrictive regulations black rockfish landings have steadily declined for commercial fisheries since the mid 1980's. Recreational landings peaked in the late 1980's and declined slightly in the 1990's and have increased slightly in the most recent years (Table 1 and Figure 4).

### 1.3.1 Catch

Little information exists on the historical landings of black rockfish prior to the early 1960's. The first black rockfish catch of 151.5 mt was recorded in 1952 for trawl gear. Landings of rockfish peaked at nearly $25,000 \mathrm{mt}$ in 1945 in support of the war effort, however, there is no known species composition estimates for these catches (Table 2). Due to the nearshore habitat of this species it is likely that very little of this catch was black rockfish. Catches prior to known estimates were decayed to zero back to 1940 within the model and these catches are presented in Table 3.

Predominate harvesters of black rockfish between 1963 and 1983 were commercial line and trawl fishers. Black rockfish trawl landings typically came from directed tows on nearshore rocky reefs and shipwrecks with few landings incidental to other targeted fisheries. Catch information has been updated since the 1999 assessment to reflect changes in species composition estimates derived from port sampling. These changes resulted in a slightly lower catch during the early part of the time series (Figure 5). Peak landings in the trawl fishery reached 350 mt in 1976 and declined to less than 10 mt in recent years due to area and catch restrictions (Figures 6-8).

The "non-trawl" fishery is composed of three distinct line fisheries and each differs in target species. Oregon and Washington fish receiving tickets show nominal rockfish catch as early as 1970 in the salmon troll fishery, during 1973 in the jig fishery and during 1979 in the bottomfish troll fishery. Black rockfish are generally caught as bycatch in the commercial salmon troll fishery; landings peaked in the late 70's (151 mt) and steadily decreased coincident with losses in fishing opportunities for coastal salmon. The bottomfish troll fishery generally targeted lingcod; rockfish landings are small and estimated black rockfish catch never exceeded 2 mt . The jig fishery is primarily composed of small vessels less than 26 feet in length that generally fish near their port of access. Black rockfish were targeted in nearshore areas and are a significant fraction of the nominal rockfish landings in the jig fishery. Black rockfish catch in the jig fishery was inconsequential prior to 1980 and peaked in 1982 at 272 mt. Since 1996 nominal rockfish landings contain no black rockfish due to area restrictions that have forced jig fishers to target other rockfish species found farther offshore.

Black rockfish have become the primary target of the coastal groundfish sport fishery since the mid 1980’s (Table 1 and Figure 4). Small black rockfish catches were reported in the late 1970 's and steadily increased to over 300 ton per year in the mid 1990's. Due to the implementation of a 10 fish bag limit in 1995 (Figure 7) and longer salmon seasons, annual catch of black rock declined to 188 mt in 2001. In recent years, sport catches increased to more than 300 mt . The coastal recreational rockfish fishery generally competed with sport salmon, halibut and tuna fisheries, and this is reflected in year-to-year variability in black rockfish catch.

### 1.3.2 Discard

Discard of black rockfish in Washington waters in either the commercial or recreational fisheries is likely very small. "Sebastes complex" trip limits in the line fishery were nonrestrictive prior to 1999 since few landings ever achieved the trip limit, and there was no incentive to discard catch. Furthermore, Washington State waters (inside 3 miles) have been closed to directed commercial fishing since 1995. Black rockfish represented only a small fraction of the nominal rockfish catch in the trawl fishery and it is unlikely they were discarded. Discard in the sport fishery is also insignificant since the vast majority of recreational fishers do not high-grade their rockfish catch. This is supported by recent sport fishery information that indicates discard is less than 16 mt on an annual basis (Table 4).

### 1.3.3 Effort

Coastal Washington recreational effort has steadily increased since the early 1980's with some declines in the late 1990’s (Table 5). Increase in the popularity of bottomfish fishing has been coincident with loss of salmon fishing opportunities and a genuine increase in public interest in recreational groundfish fishing. Though a multiple target strategy may be used by sport fishermen, the bottomfish-only trips consisted about $15 \%$ $20 \%$ of the total activities in the Washington recreational fisheries.

### 1.4 Fishery Management

1.4.1 ABC/HG and Management Performance

The black rockfish resource was first assessed in 1994 (Wallace and Tagart, 1994). Estimated biomass declined to 60\% and female egg production decreased to $43 \%$ of the unfished level. The 1995 forecasted yield ( $\mathrm{F}_{45 \%}$ ) and harvest guideline (HG) for combined fisheries was 517 mt . Black rockfish harvest has remained below the HG at 298, 244, 242 and 309 mt for 1995, 19961997 and 1998, respectively. Harvest has also remained well below the harvest guideline of 577 mt that was established by the Council following the 1999 assessment (Wallace et al., 1999).

### 1.4.2 Review of Regulatory Changes

In recognition of the recreational fishery dependence on black rockfish and to address concerns over localized declines in availability, state and federal regulations have significantly restricted commercial and recreational harvest over the last decade (Figures 5-7). In 1992, the recreational bag limit was reduced from 15 to 12 rockfish off most of Washington, and commercial line fisheries were limited to 100 lbs of black rockfish or $30 \%$ of total catch on board except when fishing in the area between Destruction Island and Cape Alava or south of Leadbetter Point. The area restrictions were intended to reduce commercial harvest of black rockfish in areas heavily utilized by recreational fishers. WDFW imposed further restrictions in 1995 that prohibit commercial line harvest (except for bycatch in the salmon troll fishery) inside state waters, imposed trawl gear restrictions and reduced the recreational bag limit to 10 fish. These regulations are still in effect today.

### 1.5 Sampling Regime

Oregon and Washington routinely collect commercial and sport groundfish biological samples at various ports of landing (Tables 6 and 7). ODFW sampling is stratified by port complex, gear, market category, and quarter and generally follow methodology describe by Sen (1984). Oregon samples of interest for this assessment include only those samples collected from the sport fishery fishing north of Cape Falcon and landing into Garibaldi. WDFW black rockfish age composition sampling is stratified by time (year) and area (3B and 3A). Washington samples are collected from the trawl fishery throughout the year, and between March and October for the sport and commercial line fisheries concurring with the spring to fall fishing season.

Both Oregon and Washington regularly collect species composition samples for mixed rockfish market categories in the trawl fishery. Samples are used to derive catch estimates for various species including black rockfish and are available from PacFIN. WDFW periodically collected species composition samples from nominal rockfish landings in the commercial line fishery and these were used to estimate black rockfish catch in mixed rockfish categories.

### 2.0 Data

### 2.1 Catch

Black rockfish catch data are compiled from a variety of sources including PacFIN, agency reports, fish ticket information and communication with agency personnel. Rockfish landings from the domestic trawl fishery are routinely sampled for species composition by coastal port samplers. Revised estimates of catch for Washington and Oregon were obtained from PacFIN, fish tickets, and species composition sampling in coastal ports in Oregon and Washington (Tables 1-2). Revised catch estimates were slightly smaller in most years prior to 1983 (Figure 8).

Estimates of Washington coastal sport catch and effort is produced from creel and exit count data collected by WDFW's Ocean Sampling Program (OSP). WDFW instituted the OSP in the 70 's to estimate catch. The program was later refined to provide necessary information to meet the goals of the Magnuson Fishery Conservation and Management Act of 1976. Estimation procedures for sport groundfish landings are not well documented in earlier years, but species-specific catches were reported in a series of WDFW technical publications since the 1970's. Lai, et al. (1991) describes estimation methodologies beginning in the late 1980's. Variance estimates for catch are available since 1990. Black rockfish discard data in sport fisheries are available since 2002 (Table 3). Proportion-at-size and proportion-at-age by sex and fisheries where derived from biological samples collected from coastal Washington and Oregon landings north of Cape Falcon (Figures 9 and 10).

### 2.2 Biology

### 2.2.1 Sampling

Biological sampling of fisheries for black rockfish age and length compositions goes back as far as 1976 in the trawl fishery. Coverage of the commercial fisheries in the last 10 years is nil due to restrictive management. The sport fishery has been relatively well sampled over the last two decades (Tables 6 and 7).

### 2.2.2 Length weight relationship

Random samples were collected in 1984 (1,157 fish) and from1988-2001 (1,397 fish), with fork length (cm) and weight (kg) measurements. We modeled the length weight relationship as $W=a L^{b}+\varepsilon$, where $W$ and $L$ were the weight and fork length, $\varepsilon \sim \mathrm{N}\left(0, \sigma^{2}\right)$ and the parameters $a$ and $b$ were to be determined. For male black rockfish, $\hat{a}$ and $\hat{b}$ were $3.796 \times 10^{-5} \mathrm{~kg} \mathrm{~cm}^{-2.782}\left(3.303 \times 10^{-6} \mathrm{~kg} \mathrm{~cm}^{-2.782}\right)$ and 2.782 ( $\mathrm{SE}=0.02309$ ), respectively (Figure 11). For female black rockfish, $\hat{a}$ and $\hat{b}$ were $4.030 \times 10^{-5} \mathrm{~kg} \mathrm{~cm}^{-2.768}\left(3.334 \times 10^{-6}\right)$ and 2.768 ( $\mathrm{SE}=0.02188$ ), respectively (Figure 9). There was no statistical difference ( $\mathrm{P}>0.05$ ) between the male and female length weight relationships.

### 2.2.2 Growth

Random samples of black rockfish (14,919 male, 12,304 female, and 213 of unknown sex) were collected in 1984 and from 1988-2006 with age and fork length measurements. Most of the fish with unknown sex were juveniles with the smallest age equal to one. , The Schnute (1981) three-parameter von Bertalanffy growth model was used to model growth, with the assumption of no variation in growth among years. The growth equation is
$l_{t}=L_{\infty}+\left(L_{1}-L_{\infty}\right)\left(1-e^{-K(t-1)}\right)+\varepsilon$,
where $l_{t}$ is the fork length at age $t, L_{1}, L_{\infty}$ and $K$ are unknown to be determined, $\varepsilon \sim \mathrm{N}\left(0, \sigma^{2}\right) . L_{1}$ is the length at age one; $L_{\infty}$ and $K$ are von Bertalanffy growth parameters, limited size and growth constant.

Due to the restriction of $L_{1}$, which was assumed to be the same in male and female fish, we proposed the use of a dummy variable in the Schnute three parameters growth model. The use of dummy variable was to test the growth difference between male and female fish. The proposed model was
$l_{t}=L_{\infty}+D_{L} z+\left(L_{1}-L_{\infty}-D_{L} z\right)\left(1-e^{-\left(K+D_{k}\right)(t-1)}\right)+\varepsilon$,
where $z$ was a dummy variable (male $=0$, female $=1$ ), $D_{L}$ and $D_{k}$ were two additional unknown variables to be determined. In this model, males and females have the same growth curves before age 1 .

The parameters $\hat{L}_{\infty}, \hat{K}, \hat{L}_{1}, \hat{D}_{L}$ and $\hat{D}_{K}$ were $46.370 \mathrm{~cm}($ SE. $=0.215 \mathrm{~cm}), 0.194 / \mathrm{yr}$ ( $\mathrm{SE}=0.00628 / \mathrm{yr}$ ), $20.123 \mathrm{~cm}(\mathrm{SE}=0.583 \mathrm{~cm}$ ), 3.903 cm ( $\mathrm{SE}=0.347 \mathrm{~cm}$ ) and $-0.0299 / \mathrm{yr}$ ( $\mathrm{SE}=0.00472 / \mathrm{yr}$ ), respectively. In Figure 12, there are plots of the expected age fork
length relationships of male and female black rockfish. Both $\hat{D}_{L}$ and $\hat{D}_{K}$ were highly significant ( $\mathrm{P}<0.001$ ).; this implied the expected $\hat{L}_{\infty}$ and $\hat{K}$ were different between the sexes. The expected limited size of male rockfish and the growth constant were 46.370 cm and $0.194 / \mathrm{yr}$. The expected limited size of female rockfish and the growth constant were 50.273 cm and $0.164 /$ year. The estimated expected limited sizes of male and female black rockfish were similar to the expected limited sizes of male ( 46.611 cm ) and female ( 51.225 cm ) estimated by the capture-recapture data with time at large and size measurements but not the growth constants. The difference in the growth constants estimation might be due to the assumption of age at zero and the aging of fish with an integer scale.

### 2.2.3 Aging error

Since 1992, 3,147 black rockfish were sequentially selected and aged with two age readers independently. We modeled the aging error with a simple regression with no intercept. The estimated slope was 0.9977 (s.e. $=0.001858$ ). The CV of the aging error was small ( $0.18 \%$ ). Figure 13, shows a scatter plot of the age data from the two readers. Figure 14 shows the between reader age specific variation that was used for data input in the SS2 stock assessment.

### 2.2.4 Age weight conversion errors

There were aging errors, age to length conversion errors, and length weight conversion errors in age to weight conversion:

$$
W=a\left[\left(L_{\infty}+D_{L} z+\left(L_{1}-L_{\infty}-D_{L} z\right)\left(1-e^{-\left(K+D_{k} z\right)(t+1)}\right)\right]^{b} .\right.
$$

We assumed all these errors were independently normally distributed. The Delta method was employed to estimate the overall standard errors. The estimated male and female black rockfish age to weight and standard errors are presented in Table 8.

### 2.2.5 Age-length relationship and maturity

A random sample of 352 female black rockfish captured in 1998 was selected for the estimation of black rockfish maturity (Table 9). A generalized linear model with a binomial (logit link) was used to model the age of $50 \%$ maturity. Bootstrapping was used to estimate the $95 \%$ confidence intervals of the age of $50 \%$ maturity. The estimated age of $50 \%$ maturity was 10.31 year and the $95 \%$ confidence intervals by bootstrapping was ( 9.72 year, 11.24 year). The estimated probability of maturity with age was

$$
\hat{\pi}=\frac{e^{-7.13+0.69 t}}{1+e^{-7.13+0.69 t}}
$$

The estimated probability of sex maturity curve with age for females is plotted in Figure. 15. Females with fork length (l) 25-49 cm captured in 1998 ( 391 fish) were randomly selected for the estimation of maturity of rockfish (Table 10). The estimated length of $50 \%$ was 42.15 cm and the $95 \%$ confidence intervals by bootstrapping was ( 41.49 cm , $42.87 \mathrm{~cm})$. The estimated probability of maturity with fork length was

$$
\hat{\pi}=\frac{e^{-17.05+0.401}}{1+e^{-17.05+0.401}}
$$

The estimated probability of sex maturity curve with fork length is plotted in Figure. 16.
Fecundity estimates are based on 47 mature black rockfish ovaries collected during parturition between 1989 and 1991 off the central Washington coast. Estimated fecundity ranged from 117,550 eggs for a 37 mm fish to 1.2 million eggs for a 490 mm fish. Fecundity at a mean size of 41 cm is 544,528 . There is a significant relationship between fecundity and length ( $\mathrm{M}, \mathrm{E}+6 \mathrm{larvae} / \mathrm{cm}$ ) $=0.0634 \mathrm{~L}-2.0586$ and fecundity and weight $(\mathrm{M}, \mathrm{E}+6 \mathrm{larvae} / \mathrm{kg})=0.7674 \mathrm{~W}-0.3657$ (Figure 17). Fecundity at weight parameters are provided as data input to synthesis and since larval output are in 1.0E+6 units, spawning biomass from the model should be multiplied by 106 to obtain the absolute spawning output. An increasing larval output by older, larger fish has a significant impact on the population dynamics such that a lightly exploited population with and age structure shifted towards older fish, would have greater spawning potential than a population shifted towards younger fish even if the biomass of spawning females were the same. This effect is significantly amplified in the black rockfish populations because it appears that larvae from larger, older black rockfish appear to be more viable than those from younger fish (Berkeley, 2004). This further implies that maintaining a black rockfish population that preserves the older segment of the population may be very important for reproductive success of this species.

### 2.2.6 Total mortality

The mortality model we used assumes direct density dependence. If the population at time $t$ is $N(t)$, then

$$
\frac{d N(t)}{d t}=-Z N(t),
$$

where $M$ is the termed the instantaneous coefficient of total mortality. This model is popular for fish stock assessment because it is simple, because data are usually not available to support more complex representations, and because it often makes reasonable assumptions for the exploited age classes. The population size at time $t$ is

$$
N(t)=N(0) e^{-Z t} .
$$

We assume that fishing mortality ( $F$ ) and natural mortality ( $M$ ) sum to total mortality (Z) , where $\mathrm{Z}=F+M$.

Taking the log of both sides of the equation, we get

$$
\log (N(t))=\log (N(0))-Z t .
$$

For the rockfish length frequency composition data, we need to convert the fork length into age. The inverse von Bertalanffy growth equation is
$t(L)=t_{0}-\frac{\log \left(1-\frac{L}{L_{\infty}}\right)}{K}$.We set $t_{0}=0$ for simplicity. Now,

$$
\log \left(\frac{N(t)}{t(L+\Delta L)-t(L-\Delta L)}\right)=\log \left(\frac{N(0)}{2(t(L+\Delta L)-t(L))}\right)-Z t(L) .
$$

The length interval for the frequency data is 3 cm . Assuming errors in the data, we can fit a regression line with

$$
y=\log \left(\frac{N(t)}{t(L+\Delta L)-t(L-\Delta L)}\right) \text { and } x=t(L) \text { with } \Delta L=1.5 \mathrm{~cm} \text {. }
$$

The above method is equivalent to the method of Pauly (1983).. who derived it by using the Baranov catch equation,

$$
C\left(t_{1}, t_{2}\right)=N\left(t_{1}\right) \frac{F}{Z}\left\{1-\exp \left[-Z\left(t_{2}-t_{1}\right)\right]\right\},
$$

where $C\left(t_{1}, t_{2}\right)$ is the total between age class $t_{1}$ and $t_{2}$. He approximated part of the catch equation

$$
\log C\left(t_{1}, t_{2}\right)=d-Z t_{1}+\log \left\{1-\exp \left[-Z\left(t_{2}-t_{1}\right)\right]\right\}
$$

with $\log (1-\exp (-\Delta t)) \approx \log (\Delta t)-\frac{\Delta t}{2}$. Both results are similar.
Black rockfish length frequency data have been collected from port sampling and recreational surveys since 1984. Both male and female black rockfish length frequency data show peaks near 36 cm , presumably due to fishery selectivity. Thus, for the purposes of this analysis, black rockfish with size greater than 36 cm were used to estimate the total mortality. We estimated the total mortalities of black rockfish by sex. The estimated male and female total mortality coefficients from 1984 to 2006 and number of samples are listed in Table 11. Plot of expected male and female estimated total mortality coefficients against total fishing effort are shown in Figure 18. The estimated intercept ( $\sim 0.2$ for males and $\sim 0.26$ for females) in each sub graph is the estimated natural mortality (where effort=0) using the mortality model described above and assuming direct density dependence. The estimated female total mortality coefficients were greater than the estimated male total mortality coefficients from 1984 to 2006 and beginning in 2000 there was a decreasing trend observed in both male and female black rockfish total mortality (Figure 19).

### 2.2.7 Natural mortality

Fish natural mortality is confounded with fishing mortality, so it is one of the most challenging fish biological parameters to be estimated. It significantly affects the stock rebuild time and the estimation of virgin fishery biomass. There are both direct and
indirect methods to estimate the natural mortality of fish species. Indirect methods are derived from other biological parameters, e.g., the growth constant and fecundity (Wallace et al., 1994 and Wallace et al., 1999). It is difficult to estimate the uncertainties from indirect methods.

In this assessment, we attempted to estimate the natural mortality of black rockfish with a direct method. We assumed $F=q E$, where $q$ was catchability coefficient and $E$ was fishing effort. Natural mortality could be estimated with the relationship

$$
Z=F+M .
$$

After 1995, the bag limit for recreational catch dropped to 10; thus, we only included the recreational rockfish trip effort (fish/angler) and the total catch in this analysis. We assumed constant $M$ with annual variation and the total fishing effort at year $t$ would result in the total mortality in year $t+1$. The proposed model was

$$
Z_{t+1}=q E_{t}+M+\varepsilon_{t},
$$

where $\varepsilon_{t} \sim \operatorname{NID}\left(0, \sigma^{2}\right), q$ and $M$ were the unknown to be determined.
Plot of expected male and female estimated total mortality coefficients against total fishing effort where the intercept was the estimated natural mortality is shown in Figure 18. The estimated linear relationship between male and female black rockfish is shown in Figure 19 and a time series plot of the estimated male and female black rockfish total mortalities is shown in Figure 20. The estimated $\hat{M}$ of male and female black rockfish were 0.223 ( $\mathrm{SE}=0.0071$ ) and 0.272 ( $\mathrm{SE}=0.061$ ). The relationship of $\hat{M} \approx \hat{K}$ was observed in male black rockfish, while $M=1.6 \hat{K}$ was observed in female black rockfish. All these values agreed with other existing indirect methods.

### 2.3 Abundance Indices

### 2.3.1 Bottom trawl surveys

The NMFS has conducted the West Coast triennial bottom trawl survey of groundfish resources since 1977. Survey depth range in most years has been from $30-200 \mathrm{fm}$ (Wilkins et al., 1995). This is outside the normal depth range of black rockfish and only 233 fish in 27 tows have been captured to date. Therefore, we incorporated no triennial trawl survey data into this assessment.

### 2.3.2 Recreational CPUE

Abundance indices are assumed to be proportional to population abundance. The catchability coefficient $(\mathrm{Q})$ is the factor that relates the units of the index to the abundance of the population. Random variability in the coefficient may occur, but if there is a trend over time or if the coefficient varies with population abundance, then the assessment may be biased. Sport fishery catch rates will be influenced by undocumented search time at sea, and the observed decline in CPUE indices would be underestimated. There is no information to evaluate annual differences in effort for specific individual target species such as black rockfish. April-September estimates of catch and effort (by trip type) for the sport fishery from coastal Washington ports are available from the WDFW Ocean Sampling Program since 1990. Black rockfish abundance trends were
explored using methods described below, but not used in the current assessment due to 1) changes in bag limits, 2) a switch to bait in the early -mid 1990's, and 3) a bag limit that may have capped the trend since the mid-late 1990's that may have biased the population trend.

## Delta lognormal model

A delta lognormal model (Lo et al. 1992) has been commonly used for modeling the abundance of marine species from trawling data. It uses generalized linear models GLMs in both stages, where $P_{i j}$ is the probability of abundance of observation $j$ in year $i$ and $C_{i j}$ is the catch per unit effort (CPUE). CPUE can be catch per angler hr, catch per trip, or catch per angler. The distribution of $C_{i j}>0$ usually follows a lognormal distribution.
The distribution of $P_{i}$ follows a binomial distribution. The modeling of $P_{i j}$ and $C_{i j}$ through a two stage process with other predictor variables is commonly called delta lognormal model (Lo et al., 1992). This approach affords the opportunity to investigate the probability of abundance on a spatial scale with other predictor variables, which include both geographical information and environmental variables. Problems associated with zero values in catch data can be avoided by using the delta lognormal model, which only fits the positive catch data. There is, however, a possible bias induced by using a two stage model process. Lo et. al. (1992) and Syrjala (2000) attempted to estimate the bias of the estimated variance in this model using both simulation and approximation techniques. Both $P_{i j}$ and $C_{i j}$ do not assume normally distributed (binomial, lognormal) in the two stages model process and there is possible correlation between them. Also, the use of the delta lognormal method to estimate the variance of the final estimate is questionable. This can be overcome by non-parametric bootstrapping.

## First stage model

The response variable $P_{i j}$ is a Bernoulli component (presence-absence) of CPUE $j$ in year i. The choice of the logit link function is standard (McCullagh and Nelder, 1989; Cheng and Gallinat, 2004). The link function is

$$
g\left(P_{i j}\right)=\log \left(\frac{P_{i j}}{1-P_{i j}}\right)=x_{i},
$$

where $x_{i}$ is a factor variable (annual effect).

## Second stage model

We model $C_{i j}>0$ in terms of the covariates $x_{i j}$. It is a truncated Poisson distribution.

## Bootstrapping method and non-parametric coefficient of variation

The nonparametric bootstrap method (Efron 1982; Hall 1992; Jackson and Cheng 2001) was used to estimate the $95 \%$ confidence intervals for the mean CPUE estimates obtained from average CPUE and from the delta lognormal model. Due to the computational intensity required when applying GLMs and a large data set, $K=200$ to 1000 samples
have been used. We have rerun the bootstrapping three times and compared the precision of the estimates of the $2.5 \%, 15.87 \%, 84.13 \%$, and $97.5 \%$ quantiles. The estimates of the quantiles are correct to the first 3 significant places due to the very large dataset. The coefficient of variation of a data $X$,

$$
C V_{X}=\frac{\sigma_{X}}{\mu_{X}} \approx \frac{\hat{\sigma}_{X}}{\bar{X}},
$$

is commonly used to describe the variation (one standard deviation) of the data compared with the mean of the data. The parameters $\sigma_{X}$ and $\hat{\sigma}_{X}$ are population $X$ standard deviation and estimate population $X$ standard deviation. Letting $q_{X, 0.025}$ be the $2.5 \%$ quantile of data $X$, we define the ad hoc CV for the non-normal distribution as

$$
C V_{X}=\frac{q_{X, 0.8413}-q_{X, 0.1587}}{2 \mu_{X}} \approx \frac{\hat{q}_{X, 0.8413} \hat{q}_{X, 0.1587}}{2 \bar{X}} .
$$

For the sample mean, we use

$$
C V_{\bar{X}}=\frac{q_{X, 0.8413-} q_{X, 0.1587}}{2 \sqrt{n} \mu_{X}} \approx \frac{\hat{q}_{X, 0.8413-} \hat{q}_{X, 0.1587}}{2 \sqrt{n} \bar{X}}
$$

where $n$ is the sample mean.
The sample mean of the CPUE in each year was compared with delta lognormal model results. Black rockfish length frequency data have been collected since 1990 in both recreational and commercial fisheries. Plots of the estimated recreational fishery CPUEs from mean estimators and the delta lognormal model for all areas combined is shown in Figure 21 and Figure 22 shows results from Area 2 only, A summary of the number of recreational data recorded in all areas (Areas 1,2 3, 74 and 84) and the proportion of these from 1990 to 2006 is given in Table 12. Area 2 was the major fishing area and the fishing effort was roughly proportional to the catch. Area 2 was the major rockfish area. Tables 13 and 14 provide summaries of the estimated CPUEs from the mean estimator and the delta lognormal model for all areas combined and area 2, respectively. Undoubtedly, Area 2 had a higher CPUEs compared with the other areas. Although the bag limit changed from 15, to 12 to 10 during 1990 to 1995, the estimated CPUEs reflected the changes from 1990 to 1993. From 1995 onwards, there was an increasing trend in CPUEs with a constant bag limit (Figures 19-20).

### 2.3.3 Tagging CPUE

Since the start of the coastal Washington black rockfish tagging program in 1981 information on catch and rod hours have been recorded. These data represent the total number of fish caught and angler hours at each specific fishing location during a trip. The number of fish tagged (and released) was typically less because of mortality from hooking or barotraumas. The tag CPUE in the current assessment represents the mean annual CPUE across all trips (by year) for the Central Washington Coast between Grays Harbor and Sea Lion Rock since 1985 (Table 15 and Figure 24).
2.3.4 Mark-recapture tagging study

From 1981 to 1990 and resuming again in 1998, black rockfish has been the subject of multiple tagging experiments along the coast of Northern Oregon and Washington Since

1998, internally implanted coded wire tags (CWT) were employed to ensure that tag recovery was not dependent upon tag reporting by fishers. Information from the first two years of recovery was suspect and was dropped from the tag abundance index.

### 2.3.5 CWT tag loss rate

Double CWT tagging experiments were conducted between 1998 and 2006 to estimate the tag loss rates. The estimated tag loss rates were used to adjust the number of tag returns. In 1998, 2262 black rockfish were released with double CWT tags on both the left and right sides of the fish in order to estimate the instant CWT tag loss rate per year. In total, there were 2209 fish returned with double CWT tags, 58 fish returned with left CWT tag loss and 66 fish returned with right CWT tag loss (Table 16 and Figure 23). The estimate the instant rate of tag loss per year was -0.0017 (st. err $=0.0003, \mathrm{P}=0.0035$ ).

### 2.3.6 Population estimate

Petersen's method (Chapman, 1951) was used to estimate the population size from capture and recapture data. The method requires only two survey periods; the first survey involves the initial marking of $n_{1}$ fish, of which $m$ tagged fish are recovered from $n_{2}$ fish sampled in the second survey. The estimated population size is

$$
\hat{N}=\frac{\left(n_{1}+\right)\left(n_{2}+1\right)}{m+1}-1
$$

and

$$
\operatorname{Var}(\hat{N})=\frac{\left(n_{1}+1\right)\left(n_{2}+1\right)\left(n_{1}-m\right)\left(n_{2}-m\right)}{(m+1)^{2}(m+2)} .
$$

The assumptions are
i) The tags are not lost and always identified on recapture.
ii) The population is closed.
iii) Every individual has the same probability of recapture.

When the tag loss rate is known, the new estimate is $\hat{\tilde{m}}=\frac{m}{1+\hat{\beta}}$.
The estimated population sizes for years 2000 to 2006 are given in Table 17.

### 3.0 Modeling History

In the 1994 stock synthesis model configuration, two auxiliary data sets were used as black rockfish abundance indicators: tagging CPUE, and coastal recreational bottomfish directed effort (Wallace and Tagart, 1994).

In 1999 we constructed an assessment model by using the AD model builder software (Fournier1997) to assess black rockfish abundance (Wallace et al 1999). The three key features of the 1999 model were (1) the parameterization of the expected catches at age, (2) the definitions of the sampling unit for the different types of data input, and (3) the integration of tagging data explicitly. The parameterization chosen mostly affected parameter bias whereas the sampling unit designation mostly affected estimator variance.

Both bias and variance were components of overall parameter uncertainty. The parameterization and the sampling unit definitions were both designed to conform to the actual sampling protocol used, thereby propagating sampling uncertainty through to the final biomass estimates.

### 4.0 Current Model Description

The current assessment employed Stock synthesis 2 (SS2V2.00c, compiled 3/27/2006) to model the dynamics of black rockfish population found between Cape Falcon, Oregon and North to the U.S./Canadian border in Coastal Waters. This model is a forward projecting, separable, age-structured model developed by Methot (2006). The convergence criterion for maximum gradient was set to $1.0 \mathrm{e}-5$. The model was specified to begin in 1915 to ensure population equilibrium at the start of the modeling time period. Catch data were decayed from the last reliable catch estimates (1965) to 0 by 1940. Fisheries catch, size and age compositions were pooled into three fishery types including trawl, sport and non-trawl. The first size-age compositions were collected in the mid 1970's from the trawl fishery, but samples were not collected on a systematic basis until 1985. Growth (Lmin, Lmax and k) was estimated within the model to account for fishery selection of the larger individual fish at age. The population model was tuned to two fisheries-independent indices that include a tagging CPUE (1986-2007) and a tag abundance biomass index (2000-2007) both derived from WDFW black rockfish tagging information. Both STAT and STAR Panel members agreed that the available fishery dependent indices should not be incorporated due to potential bias resulting from bag limit changes and undocumented measures of fishing effort resulting from changes in search time across the time series. The black rockfish STAR base and STAT best fit models down-weights size composition for all fisheries (emphasis=0.1) to improve model fit to the age composition and indices rather than length. Given that length compositions are from all samples including aged samples, down weighting mitigates effects that may be contributed to the model by "double-counting" composition sampling.

There are 10 likelihood components for data including: 1) tag abundance, 2) tag CPUE, 3) trawl size compositions, 4) sport size compositions, 5) non-trawl size compositions, 6) tag survey size compositions, 7) trawl age compositions, 8) sport age compositions, 9) non-trawl age compositions and 10) mean size at age (Table 18).

There are a total of 76 parameters estimated within the base model and assumptions on priors are listed in Table 19. We modeled the black rockfish spawner-recruit relationship using the Beverton-Holt curve. The key steepness parameter (h), which determines overall productivity of a stock, is fixed at 0.6 and the prior on $h$ is set to 0.35 in the STAR base model and in the STAT best fit model. Based on Dorn's (personal communication) Bayesian meta-analysis of productivity for west coast rockfish stocks, steepness parameter (h) for black rockfish should be in the 0.6-0.7 range and variation about the stock-recruit curve (Sigma R) would be near 0.57. The natural mortality for females is assumed to be constant ( 0.16 ) for ages $<=10$ and then increasing to 0.2 by age 15 , and males were assumed to have a constant natural mortality of 0.16 for all ages in the STAR base model. The natural mortality for females is assumed to be constant (0.16) for ages
<=10 and then increasing to 0.24 by age 15, and males were assumed to have a constant natural mortality of 0.16 for all ages in the STAT best fit model.

## Sample size and effective sample size

Initial sample size inputs were based upon methods presented at the NMFS 2006 Stock Assessment Data and Modeling workshop that incorporates objective weighting for length- and age-frequency data for West coast groundfish fisheries where:
Fishery data:
Effective $\mathrm{N}=((0.138 *$ FPS $)+1) *$ NS where: FPS $<44$
Effective $\mathrm{N}=7.06^{*}$ NS $\quad$ where: FPS $>44$
Survey data:
Effective N = ((0.070*FPS)+1)*NS where: FPS < 55
Effective $\mathrm{N}=4.89^{*} \mathrm{NS} \quad$ where: FPS $>55$
NS = Number of samples
FPS = Average number of fish per sample
Comparison of input sample size and the effective sample sizes estimated by the STAR base model are provided in Tables 20 and 21.

### 5.0 Model Selection and Evaluation

A large number of model structures were initially explored prior to establishing a base black rockfish model. Our primary goal in model selection was to ensure fit to the tag abundance index and age composition data while minimizing the overall likelihood. This is because we have confidence in the study design and methodology of our current tagging program and the resulting abundance estimates. In addition, we have collected numerous age samples from the fisheries during the last two decades that likely represents the underlying age structure of the population.

Natural mortality for mature females (>10 years of age) was assumed to be 0.20 (STAR base) and 0.24 (STAT best fit) and constant 0.16 for males and females < age 11. These rates are within the range of natural mortality rates estimated external to synthesis. Both male and female natural mortality rates are lower than that estimated in the 1999 assessment (Figure 25) and somewhat lower than the 1982 catch curve estimate of 0.265 (Wallace et al., 1994). The natural mortality in the current assessment is higher than that used in the 2003 assessment for black rockfish populations off Oregon and California (Ralston and E.J. Dick, 2003), which used a natural mortality of 0.1 and 0.2 for males and females, respectively.

Results of the model sensitivity analysis on natural mortality (Table 24) indicate that the STAT best fit model provided a better overall fit to the data compared to the STAR base model and estimates of fishing mortality is closer to tagging study results (Figure 26). We conclude that the STAR base model should be used to base management decisions and set allowable harvest. A list of supporting information include: 1) the assumed rate of natural mortality in the "Low Natural Mortality" state of nature is lower than any previous assessment for the "Northern" population, is lower that any external estimation
by direct and indirect methods, 2) biomass results from the "Low Natural Mortality" indicate that the population declined to less than $13 \%$ of the unfished population in the mid-1990's yet we have no indication from the fishery or from our tagging study that there was localized depletion during this time period, 3) sensitivity analyses indicate "Low Natural Mortality" model fit to the data is very poor relative to other model results that assume a higher rates of natural mortality, 4) the estimated q for the survey is likely double what it should be based on STAT knowledge of available habitat off the Washington coast, 5) tagging data are not fit well and tagging estimates external to the model indicate that the population is larger and fishing mortality is lower compared to STAR base model run results, 6) other model runs with higher steepness and Sigma R fit the data better and improved our view of the current population status above both the STAR base and "Best Fit" model runs and finally, 7) compared to the STAT best fit model a model with high natural mortality for females (where $\mathrm{M}=0.16$ for males and females $<=10$ years of age and M for females linearly increasing from age 11 to age 15 to 0.26 ) fit the data equally well. This model resulted in an improved view of current population status above both the STAR base and STAT best-fit model runs. However, results from this model were not incorporated in the decision table because the higher natural mortality on females (0.26) fell outside the range considered at the STAR Panel.

Convergence properties using a parameter jitter of 0.001 was also explored in the base model and results indicate no other local minima (Figure 27). Growth was assumed linear to age 5 where variation in fork-length at age was stabilized across ages (Figure 28). Growth was fully (Lmin, Lmax and k ) estimated within the model to account for fishery selection that favors the largest individuals at age. Model estimates of growth compared reasonably to external estimates and there were no apparent differences in estimates of growth between STAR and STAT models (Figure 29).

Both the STAR and STAT models underestimated the increasing trend in tag abundance and tag CPUE indices in most recent years (Figure 30). We believe this is due to several factors including that tagging is not incorporated in the model as a tagging experiment, which is not possible within the current SS2 modeling framework. Further, the model estimated effective q for the tagging index was 0.83 and this is likely double what it should be based on STAT knowledge of available habitat off the Washington coast. The north central Washington coast, where most of the nearshore rocky habitat exists, is inaccessible to most recreational fishers and is not part of the current tagging program. However, the estimation of $q$ is complicated by the fact that the $S S 2$ value of $q$ is a function of selectivity that is strongly dome shaped for the fishery. Increasing the weighting on survey abundance demonstrates that a better fit to the survey abundance index significantly improves our view of the current population status (Figure 31). Additionally, a retrospective analysis of the STAR base and STAT best-fit models shows that the indices strongly influence population trends and that the population trajectory in most recent years was highly influenced by the large (estimated) 1999 year-class (Figure 32).

Without an objective evaluation of an informed prior on $q$ it is difficult to compare a prior conception of $q$ based on tagging and the one estimated by SS2. Other issues include the
non-independence of the length/age compositions and non-independence of the tagging abundance and CPUE series. However, both STAR and STAT conclude the current model configuration(s) represented the "best available" scientific information and should be used for management.

There appears to be some pattern in the size composition residuals such that model estimates for small fish were much higher than that observed in the trawl fishery fit. However, forcing the model to fit the size compositions degraded the fit to the age composition. Fit to size compositions in the sport, non-trawl and survey showed little trend. Overall, fit (or lack of) to the size composition data did not draw significant debate at the STAR panel and model fit to size compositions is likely within the uncertainty (Figures 33-36). Model fit to the age composition data showed relatively inconsistent patterns and was considered to be within model uncertainty (Figures 37-39). There was an obvious trade off where forcing the model to fit the age data degraded the fit to the size composition data. This was not fully resolved and is discussed below in the uncertainty section.

### 6.0 Base-run Results

Comparison of STAR base model recruitment estimates to the previous assessments and the STAT best fit model indicates similar estimated recruitment patterns (Figure 40). It is apparent that the large estimated recruitment in 1994 and 1999 is highly influential in determining current stock status. Due to lack of good recruitments and intense fishing by multiple fisheries, highest fishing mortality rates occurred in the late 1980’s (Figure 41). Selectivity was domed-shaped (STAR and STAT models) in both the tagging survey and sport fishery and asymptotic in the trawl and non-trawl commercial fisheries (Figure 42). Comparison of STAR base model spawning biomass estimates to the previous assessments indicate a similar declining trend through the mid 1990's and then sharply increasing to $43 \%$ of the unfished biomass by 2006, though the trend is lower in the current model (Figure 43). The STAT best fit model resulted in a slightly smaller unfished biomass and a larger ending biomass compared to the STAR base model, biomass estimates show little difference in population trend (Figure 44).

Black rockfish stock abundance was below the Councils’ management target a number of years and also dipped below the Councils' minimum stock size threshold in the STAR base model. The STAT best fit model population trajectory remained just above minimum stock size threshold. Both models indicate that the stock is currently well above the management target of B40\% (Figure 45). The corresponding exploitation rate relative to spawning biomass shows similar trend and harvest rates have exceeded management targets between the mid 1980's through the mid 1990's (Figure 46).

### 7.0 Uncertainty and Sensitivity

Natural mortality is confounded with fishing mortality making it one of the most difficult biological parameters to estimate. In this assessment we explored direct methods to
estimate natural mortality and compared to estimates derived from indirect methods (from other biological parameters, e.g., the growth constant and fecundity) in previous assessments. It is apparent from our analysis using both direct and indirect methods that our current assumptions on natural mortality in the STAR base model are within our limits to estimate this parameter and that the "low natural mortality rate" model used to bracket uncertainty is likely too low. Model sensitivity analysis show that other model configurations using higher natural mortality assumptions such as the STAT best fit model provides a better overall fit to the data (Figure 47).

Tagging is not incorporated in the model as a tagging experiment, which is not possible within the current SS2 modeling framework. The index for tagging abundance is not fit well and the model estimated effective $q$ for the tagging index was 0.83 . This is likely double what it should be based on STAT knowledge of available habitat off the Washington coast. Further, the north central Washington coast, where most of the nearshore rocky habitat exists, is inaccessible to most recreational fishers and is not part of the current tagging program. However, the estimation of $q$ is complicated by the fact that the SS2 value of $q$ is a function of selectivity that is strongly dome shaped for the fishery. Increasing the weighting on survey abundance shows that a better fit to the survey abundance index significantly improves our view of the current population status (Figure 31). Without an objective evaluation of an informed prior on $q$ it is difficult to compare a prior conception of $q$ based on tagging and the one estimated by SS2. Other issues include the non-independence of the length/age compositions and nonindependence of the tagging abundance and CPUE series.

Likelihood profile of the STAR base assessment model for different fixed values of the Beverton-Holt steepness parameter (h) and Sigma R show that higher values (STAR base and STAT best-fit model had the steepness fixed at 0.6 and Sigma R tuned to 0.35) of both parameters improved the overall fit to the data (Figure 48). Our assumption on h is well within the uncertainties based on the Dorn meta-analysis, but assumptions on Sigma R may be too low (Dorn personal communication).

Changes in likelihood profile for various components of the base assessment model following changes in the emphasis (weight) of the recruitment Dev and Dev time series indicate very modest changes in fit for weighting between 0.1 to 100 with fit improving to age compositions and declining fit to size compositions with increasing emphasis (Figure 49).

Likelihood profile for various components of the base assessment model following changes in the emphasis (weight) on the abundance and tag CPUE indices indicate a slight improvement in fit by increasing emphasis to 10 on most components with the exception to the fit to sport size and age that declined (Figure 50). Increasing emphasis on the age composition for all fisheries above 1 improves fit to the abundance indices but increased likelihood for the fishery size components (Figure 51). The model was very sensitive to increasing emphasis on the size compositions and declined fit to all age and index components substantially (Figure 52).

### 8.0 Reference Points and Forecast

The Pacific Fisheries Management Council recommends that a default target fishing mortality rate of FSPR=0.5 be used for Council managed rockfish species. The current assessment uses this default for harvest projections and based on the Councils control rule for groundfish would not be considered overfished. Reference points and benchmark fishing mortality rates are shown in Table 23. Forecast ABC's, Spawning biomass and depletion is shown for both the "STAR base" and STAR base model and the STAT best fit model in Table 24.

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## Tables

Table 1. Black rockfish catch North of Cape Falcon, Oregon by gear and year 1963-2006 (blanks indicate no data).

Coastal black rockfish catch North of Cape Falcon, Oregon

| Year | Catch by Gear |  |  | Catch by PMFC Area |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Trawl | Sport | Ion-Trawl | 3A | 3B | 3C-S | TOTAL |
| 1963 | 19.0 |  |  | 19.0 | 0.0 | 0.0 | 19.0 |
| 1964 | 8.0 |  |  | 8.0 | 0.0 | 0.0 | 8.0 |
| 1965 | 108.0 |  |  | 108.0 | 0.0 | 0.0 | 108.0 |
| 1966 | 186.0 |  |  | 186.0 | 0.0 | 0.0 | 186.0 |
| 1967 | 234.0 |  |  | 234.0 | 0.0 | 0.0 | 234.0 |
| 1968 | 122.0 |  |  | 122.0 | 0.0 | 0.0 | 122.0 |
| 1969 | 261.0 |  |  | 261.0 | 0.0 | 0.0 | 261.0 |
| 1970 | 303.0 |  | 20.5 | 320.4 | 3.1 | 0.0 | 323.5 |
| 1971 | 134.1 |  | 17.5 | 147.6 | 4.0 | 0.0 | 151.6 |
| 1972 | 116.0 |  | 29.3 | 137.7 | 7.5 | 0.0 | 145.3 |
| 1973 | 48.0 |  | 26.8 | 63.7 | 11.1 | 0.0 | 74.8 |
| 1974 | 75.0 |  | 51.2 | 106.8 | 19.4 | 0.0 | 126.2 |
| 1975 | 156.0 | 62.3 | 36.9 | 216.7 | 38.5 | 0.0 | 255.2 |
| 1976 | 347.2 | 36.8 | 32.7 | 384.5 | 32.3 | 0.0 | 416.7 |
| 1977 | 15.0 | 76.0 | 52.2 | 96.1 | 47.2 | 0.0 | 143.2 |
| 1978 | 96.0 | 94.2 | 89.8 | 185.2 | 94.8 | 0.0 | 280.1 |
| 1979 | 321.3 | 150.7 | 104.0 | 500.5 | 75.5 | 0.0 | 576.0 |
| 1980 | 64.6 | 144.8 | 70.5 | 228.9 | 51.0 | 0.0 | 279.9 |
| 1981 | 213.0 | 213.8 | 81.8 | 436.6 | 72.0 | 0.0 | 508.6 |
| 1982 | 185.1 | 135.7 | 128.9 | 364.8 | 84.9 | 0.0 | 449.7 |
| 1983 | 327.5 | 244.3 | 134.1 | 458.2 | 247.7 | 0.0 | 705.9 |
| 1984 | 218.9 | 302.2 | 145.8 | 513.2 | 153.8 | 0.0 | 666.9 |
| 1985 | 127.3 | 305.3 | 272.0 | 407.8 | 296.8 | 0.0 | 704.6 |
| 1986 | 158.6 | 391.1 | 103.0 | 534.0 | 118.8 | 0.0 | 652.8 |
| 1987 | 82.0 | 389.3 | 220.1 | 494.3 | 197.0 | 0.0 | 691.3 |
| 1988 | 129.0 | 414.2 | 129.3 | 521.1 | 151.5 | 0.0 | 672.5 |
| 1989 | 124.4 | 369.6 | 165.3 | 469.3 | 188.0 | 2.0 | 659.3 |
| 1990 | 43.3 | 387.2 | 119.4 | 386.9 | 163.0 | 0.0 | 549.9 |
| 1991 | 46.2 | 332.3 | 83.4 | 320.3 | 139.6 | 1.9 | 461.9 |
| 1992 | 71.4 | 342.9 | 132.3 | 327.2 | 219.3 | 0.0 | 546.5 |
| 1993 | 46.8 | 316.9 | 88.4 | 298.3 | 152.9 | 1.0 | 452.2 |
| 1994 | 1.0 | 358.6 | 106.3 | 323.9 | 141.6 | 0.4 | 465.9 |
| 1995 | 3.3 | 264.8 | 65.8 | 214.9 | 118.7 | 0.3 | 333.9 |
| 1996 | 0.0 | 264.2 | 8.6 | 204.4 | 68.4 | 0.0 | 272.8 |
| 1997 | 9.0 | 234.1 | 15.0 | 194.8 | 63.2 | 0.1 | 258.1 |
| 1998 | 73.1 | 259.4 | 4.8 | 268.1 | 69.0 | 0.3 | 337.4 |
| 1999 | 0.0 | 221.6 | 4.3 | 169.4 | 56.5 | 0.0 | 225.9 |
| 2000 | 0.0 | 224.8 | 1.2 | 158.2 | 67.9 | 0.0 | 226.1 |
| 2001 | 0.0 | 188.7 | 1.1 | 134.3 | 55.5 | 0.0 | 189.8 |
| 2002 | 0.2 | 238.9 | 1.5 | 173.5 | 67.1 | 0.0 | 240.6 |
| 2003 | 0.1 | 237.1 | 0.2 | 166.8 | 70.6 | 0.0 | 237.4 |
| 2004 | 0.6 | 268.0 | 0.7 | 174.4 | 94.9 | 0.0 | 269.3 |
| 2005 | 0.0 | 331.7 | 0.9 | 217.8 | 114.7 | 0.0 | 332.5 |
| 2006 | 1.4 | 321.5 | 1.2 | 248.7 | 75.4 | 0.0 | 324.1 |
| Mean | 101.7 | 253.8 | 68.8 | 261.5 | 82.6 | 0.1 | 344.2 |
| Last 10 y | 8.4 | 252.6 | 3.1 | 190.6 | 73.5 | 0.0 | 264.1 |
| Last 5 y | 0.5 | 279.4 | 0.9 | 196.2 | 84.5 | 0.0 | 280.8 |

Table 2. Historical catch of rockfish by known rockfish catch categories between 1936 and 1969.

| Historical catch of rockfish by rockfish catch categories for coastal <br> Washington Waters <br> Year | Black Rockfish | Rockfish |
| :---: | :---: | ---: |
| 1936 | - | 0.1 |
| 1937 | - | 219.0 |
| 1938 | - | 273.7 |
| 1939 | - | 290.8 |
| 1940 | - | 330.2 |
| 1941 | - | 554.7 |
| 1942 | - | $1,925.0$ |
| 1943 | - | $5,811.7$ |
| 1944 | - | $9,084.7$ |
| 1945 | - | $25,969.7$ |
| 1946 | - | $11,322.2$ |
| 1947 | - | $2,970.8$ |
| 1948 | - | $5,192.1$ |
| 1949 | - | $5,943.5$ |
| 1950 | - | 151.1 |
| 1951 | - | $6,777.8$ |
| 1952 | 151.5 | - |
| 1953 | 8.0 | 153.1 |
| 1954 | 16.1 | 2.8 |
| 1955 | 5.0 | 76.5 |
| 1956 | 7.8 | - |
| 1957 | 19.1 | 76.5 |
| 1958 | 71.8 | 33.1 |
| 1959 | 26.6 | 36.2 |
| 1960 | 96.2 | 32.7 |
| 1961 | 40.7 | 40.5 |
| 1962 | 12.5 | 22.5 |
| 1963 | - | 279.9 |
| 1964 | 3.4 | 38.7 |
| 1965 | - | 347.8 |
| 1966 | 1.0 | 36.6 |
| 1967 | - | 167.7 |
| 1968 | - | 130.9 |
| 1969 | - | 151.4 |
|  |  |  |

Note: Data from WDFW annual catch reports.

Table 3. Assumed catch and data input of black rockfish between 1940 and 1975. Bold italics represents catch assumptions and normal italics indicate the actual catch estimate based on fish ticket and species composition data.

|  | Catch by Gear |  |  |
| ---: | ---: | ---: | ---: |
| Year | Trawl | Sport | Non-Trawl |
| Initial | 2.0 | 2.0 | 1.0 |
| 1940 | 3.2 | 2.8 | 0.0 |
| 1941 | 9.2 | 4.6 | 0.0 |
| 1942 | 15.2 | 6.3 | 0.0 |
| 1943 | 21.2 | 8.1 | 0.0 |
| 1944 | 27.2 | 9.8 | 0.0 |
| 1945 | 33.2 | 11.6 | 0.0 |
| 1946 | 39.2 | 13.3 | 0.0 |
| 1947 | 52.0 | 15.1 | 0.0 |
| 1948 | 51.2 | 16.8 | 0.0 |
| 1949 | 57.2 | 18.6 | 0.0 |
| 1950 | 63.2 | 20.3 | 1.5 |
| 1951 | 69.2 | 22.1 | 2.5 |
| 1952 | 75.2 | 23.8 | 3.5 |
| 1953 | 81.2 | 25.6 | 4.5 |
| 1954 | 87.2 | 27.3 | 5.5 |
| 1955 | 93.2 | 29.1 | 6.5 |
| 1956 | 99.2 | 30.8 | 7.5 |
| 1957 | 105.2 | 32.6 | 8.5 |
| 1958 | 111.2 | 34.3 | 9.5 |
| 1959 | 117.2 | 36.1 | 10.5 |
| 1960 | 123.2 | 37.8 | 11.5 |
| 1961 | 129.2 | 39.6 | 12.5 |
| 1962 | 135.2 | 41.3 | 13.5 |
| 1963 | 141.2 | 43.1 | 14.5 |
| 1964 | 108.0 | 44.8 | 15.5 |
| 1965 | 186.0 | 46.6 | 16.5 |
| 1966 | 234.0 | 48.3 | 17.5 |
| 1967 | 122.0 | 50.1 | 18.5 |
| 1968 | 261.0 | 51.8 | 19.5 |
| 1969 | 303.0 | 53.6 | 20.5 |
| 1970 | 134.1 | 55.3 | 17.5 |
| 1971 | 116.0 | 57.1 | 29.3 |
| 1972 | 48.0 | 58.8 | 26.8 |
| 1973 | 75.0 | 60.6 | 51.2 |
| 1974 | 156.0 | 62.3 | 36.9 |
| 1975 | 347.2 | 62.3 | 32.7 |
| Presumed Catch SS2 input in bold italics. |  |  |  |
|  |  |  |  |

Table 4. Estimated black rockfish discard in the Washington recreational sport fishery. Black Rockfish Discard in the Washington Sport Fishery

| Year | \# of Fish | Mean Weight (kg) | Assumed Mortality | Catch Weight (mt) |
| :---: | :---: | :---: | :---: | :---: |
| $\mathbf{2 0 0 2}$ | 5,719 | 1.17 | $90 \%$ | 6.0 |
| $\mathbf{2 0 0 3}$ | 4,554 | 1.21 | $90 \%$ | 5.0 |
| $\mathbf{2 0 0 4}$ | 9,764 | 1.18 | $90 \%$ | 10.4 |
| $\mathbf{2 0 0 5}$ | 15,085 | 1.19 | $90 \%$ | 16.2 |
| $\mathbf{2 0 0 6}$ | 8,733 | 1.22 | $90 \%$ | 9.6 |

Note: Discard not availible prior to 2002

Table 5. Total effort (expanded) in Washington sport fisheries.

|  | All Trip Types |  | Bottom-Fish-Only Trips |  |
| :---: | ---: | ---: | ---: | ---: |
|  | ANGLERS |  | BOAT TRIPS | ANGLERS | BOAT TRIPS

Table 6. Summary of size composition data collected from commercial and recreational fisheries during 1976 - 2006.

| Year | Number of field samples |  |  | Number of length measurements |  |  | Mean size (cm) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sport | Trawl | Non-Trawl | Sport | Trawl | Non-Trawl | Sport | Trawl | Non-Trawl |
| 1976 |  | 4 |  |  | 782 |  |  | 47.5 |  |
| 1977 |  |  |  |  |  |  |  |  |  |
| 1978 |  |  |  |  |  |  |  |  |  |
| 1979 | 7 |  |  | 508 |  |  | 46.4 |  |  |
| 1980 | 8 | 2 | 2 | 703 | 206 | 96 | 45.4 | 46.1 | 45.9 |
| 1981 | 23 | 4 |  | 1468 | 400 |  | 43.3 | 46.8 |  |
| 1982 | 9 | 4 | 1 | 263 | 400 | 29 | 40.7 | 44.5 | 40.1 |
| 1983 | 1 | 8 | 2 | 10 | 800 | 124 | 36.9 | 45.3 | 41.2 |
| 1984 | 21 | 3 | 1 | 835 | 300 | 100 | 40.4 | 44.7 | 40.9 |
| 1985 | 2 |  |  | 160 |  |  | 43.1 |  |  |
| 1986 | 21 | 13 | 27 | 512 | 322 | 527 | 41.8 | 45.5 | 44.4 |
| 1987 | 23 | 16 | 25 | 645 | 401 | 722 | 43.3 | 47.3 | 43.2 |
| 1988 | 18 | 4 | 17 | 451 | 100 | 424 | 41.9 | 47.3 | 43.8 |
| 1989 | 16 | 9 | 12 | 397 | 225 | 299 | 42.2 | 49.1 | 44.7 |
| 1990 | 11 | 10 | 4 | 290 | 249 | 125 | 41.6 | 47.1 | 36.7 |
| 1991 | 22 | 12 | 19 | 720 | 302 | 500 | 40.8 | 47.1 | 40.2 |
| 1992 | 34 | 8 | 11 | 890 | 200 | 275 | 41.3 | 46.3 | 40.0 |
| 1993 | 35 | 5 | 13 | 866 | 125 | 325 | 40.6 | 46.9 | 40.4 |
| 1994 | 35 | 2 | 9 | 868 | 49 | 250 | 40.9 | 46.2 | 38.1 |
| 1995 | 32 | 2 | 9 | 814 | 50 | 225 | 40.5 | 45.7 | 39.6 |
| 1996 | 33 |  |  | 834 |  |  | 39.5 |  |  |
| 1997 | 36 | 2 |  | 900 | 31 |  | 40.5 | 46.6 |  |
| 1998 | 37 | 2 |  | 1327 | 85 |  | 39.8 | 43.6 |  |
| 1999 | 34 |  |  | 1673 |  |  | 39.5 |  |  |
| 2000 | 33 | 1 |  | 1650 | 3 |  | 40.0 | 47.3 |  |
| 2001 | 36 | 1 |  | 1777 | 1 |  | 40.2 | 53.0 |  |
| 2002 | 56 | 1 |  | 2629 | 50 |  | 40.9 | 47.8 |  |
| 2003 | 58 | 1 |  | 2323 | 3 |  | 41.4 | 45.7 |  |
| 2004 | 44 | 2 |  | 2002 | 15 |  | 41.0 | 51.7 |  |
| 2005 | 61 |  | 1 | 2228 |  | 1 | 41.2 |  | 43.0 |
| 2006 | 152 |  | 2 | 2854 |  | 20 | 41.1 |  | 48.1 |

Table 7. Summary of age composition data collected from commercial and recreational fisheries during 1976 - 2006.

| Year | Number of field samples |  |  | Number of age |  |  | Mean age |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sport | Trawl | Non-Trawl | Sport | Trawl | Non-Trawl | Sport | Trawl | Non-Trawl |
| 1976 | 2 |  |  | 238 |  |  | 11.3 |  |  |
| 1977 |  |  |  |  |  |  |  |  |  |
| 1978 |  |  |  |  |  |  |  |  |  |
| 1979 |  |  |  |  |  |  |  |  |  |
| 1980 | 4 | 2 |  | 364 | 195 |  | 14.3 | 13.0 |  |
| 1981 | 2 | 4 |  | 71 | 394 |  | 10.6 | 11.8 |  |
| 1982 |  | 3 |  |  | 295 |  |  | 10.2 |  |
| 1983 |  | 8 | 1 |  | 794 | 100 |  | 10.2 | 11.8 |
| 1984 | 20 | 3 | 1 | 828 | 298 | 99 | 9.7 | 12.6 | 11.3 |
| 1985 | 2 |  |  | 160 |  |  | 10.8 |  |  |
| 1986 | 21 | 13 | 27 | 506 | 321 | 525 | 9.3 | 10.1 | 11.9 |
| 1987 | 23 | 16 | 25 | 642 | 401 | 720 | 11.5 | 11.3 | 10.9 |
| 1988 | 18 | 4 | 17 | 448 | 99 | 416 | 10.0 | 12.1 | 10.8 |
| 1989 | 16 | 9 | 12 | 395 | 224 | 297 | 9.3 | 10.5 | 10.8 |
| 1990 | 11 | 10 | 4 | 289 | 249 | 125 | 9.4 | 11.2 | 7.3 |
| 1991 | 22 | 12 | 19 | 717 | 301 | 500 | 9.2 | 12.2 | 8.7 |
| 1992 | 34 | 8 | 11 | 889 | 200 | 275 | 9.7 | 10.1 | 9.0 |
| 1993 | 35 | 5 | 13 | 863 | 125 | 324 | 9.0 | 10.9 | 8.5 |
| 1994 | 35 | 2 | 9 | 866 | 48 | 250 | 9.6 | 13.4 | 7.7 |
| 1995 | 32 | 2 | 9 | 813 | 49 | 225 | 8.6 | 12.0 | 7.7 |
| 1996 | 33 |  |  | 829 |  |  | 8.5 |  |  |
| 1997 | 36 |  |  | 893 |  |  | 9.6 |  |  |
| 1998 | 37 |  |  | 1323 |  |  | 9.4 |  |  |
| 1999 | 34 |  |  | 1655 |  |  | 9.1 |  |  |
| 2000 | 33 |  |  | 1644 |  |  | 9.6 |  |  |
| 2001 | 36 |  |  | 1773 |  |  | 9.7 |  |  |
| 2002 | 38 |  |  | 1894 |  |  | 9.8 |  |  |
| 2003 | 37 |  |  | 1841 |  |  | 9.6 |  |  |
| 2004 | 33 |  |  | 1645 |  |  | 9.4 |  |  |
| 2005 | 33 |  |  | 1603 |  |  | 9.6 |  |  |
| 2006 | 30 |  | 1 | 1484 |  | 19 | 9.5 |  | 14.3 |

Table 8. Summary of male and female black rockfish age to weight data with estimated errors in each conversion.

| $\begin{aligned} & \text { Age } \\ & \hat{t} \text { (st. err.) } \end{aligned}$ | Male |  | Female |  |
| :---: | :---: | :---: | :---: | :---: |
|  | $\hat{L}$ (st.err.) | $\hat{W}$ (st.err.) | $\hat{L}$ (st.err.) | $\hat{W}$ (st.err.) |
| 1(0.002) | 20.123(0.583) | 0.161(0.019) | 20.123(0.583) | 0.163(0.019) |
| 2(0.004) | 24.754(0.406) | 0.286(0.029) | 24.689(0.396) | 0.288(0.028) |
| 3(0.006) | 28.568(0.281) | 0.426(0.040) | 28.564(0.262) | 0.431(0.039) |
| 4(0.007) | 31.709(0.195) | 0.569(0.052) | 31.851(0.170) | 0.583(0.051) |
| 5(0.009) | 34.296(0.138) | 0.708(0.064) | 34.641(0.113) | 0.735(0.063) |
| 6(0.011) | 36.427(0.105) | 0.838(0.076) | 37.008(0.083) | 0.883(0.076) |
| 7(0.013) | 38.181(0.092) | 0.955(0.086) | 39.017(0.070) | 1.022(0.088) |
| 8(0.015) | 39.626(0.094) | 1.059(0.096) | 40.722(0.064) | 1.150(0.099) |
| 9(0.017) | 40.816(0.106) | 1.149(0.104) | 42.168(0.062) | 1.267(0.109) |
| 10(0.019) | 41.796(0.122) | 1.228(0.111) | 43.396(0.063) | 1.371(0.118) |
| 11(0.020) | 42.603(0.140) | 1.295(0.117) | 44.437(0.068) | 1.465(0.126) |
| 12(0.022) | 43.268(0.159) | 1.352(0.123) | 45.321(0.077) | 1.547(0.133) |
| 13(0.024) | 43.815(0.177) | 1.400(0.127) | 46.071(0.090) | 1.618(0.139) |
| 14(0.026) | 44.266(0.195) | 1.441(0.131) | 46.707(0.106) | 1.681(0.144) |
| 15(0.028) | 44.637(0.211) | 1.474(0.134) | 47.247(0.123) | 1.735(0.149) |
| 16(0.030) | 44.943(0.227) | 1.503(0.137) | 47.706(0.141) | 1.782(0.153) |
| 17(0.032) | 45.195(0.241) | 1.526(0.139) | 48.095(0.158) | 1.823(0.157) |
| 18(0.033) | 45.402(0.253) | 1.546(0.141) | 48.425(0.175) | 1.858(0.160) |
| 19(0.035) | 45.573(0.265) | 1.562(0.143) | 48.705(0.192) | 1.888(0.163) |
| 20(0.037) | 45.714(0.275) | 1.576(0.144) | 48.942(0.207) | 1.913(0.165) |
| 21(0.039) | 45.829(0.284) | 1.587(0.146) | 49.144(0.221) | 1.935(0.167) |
| 22(0.041) | 45.925(0.292) | 1.596(0.147) | 49.315(0.234) | 1.954(0.169) |
| 23(0.043) | 46.003(0.299) | 1.603(0.147) | 49.460(0.246) | 1.970(0.171) |
| 24(0.045) | 46.068(0.306) | 1.610(0.148) | 49.583(0.257) | 1.983(0.172) |
| 25(0.046) | 46.121(0.311) | 1.615(0.149) | 49.688(0.267) | 1.995(0.173) |
| 26(0.048) | 46.165(0.316) | 1.619(0.149) | 49.776(0.275) | 2.005(0.174) |
| 27(0.050) | 46.201(0.320) | 1.623(0.150) | 49.852(0.283) | 2.013(0.175) |
| 28(0.052) | 46.231(0.323) | 1.626(0.150) | 49.916(0.291) | 2.020(0.176) |
| 29(0.054) | 46.256(0.326) | 1.628(0.150) | 49.970(0.297) | 2.026(0.177) |
| 30(0.056) | 46.276(0.329) | 1.630(0.150) | 50.016(0.303) | 2.032(0.177) |
| 31(0.058) | 46.293(0.331) | 1.632(0.151) | 50.055(0.308) | 2.036(0.178) |
| 32(0.059) | 46.306(0.333) | 1.633(0.151) | 50.088(0.312) | 2.040(0.178) |
| 33(0.061) | 46.318(0.335) | 1.634(0.151) | 50.116(0.316) | 2.043(0.179) |
| 34(0.063) | 46.327(0.336) | 1.635(0.151) | 50.140(0.320) | 2.046(0.179) |
| 35(0.065) | 46.334(0.338) | 1.636(0.151) | 50.160(0.323) | 2.048(0.179) |
| 36(0.067) | 46.341(0.339) | 1.636(0.151) | 50.177(0.325) | 2.050(0.179) |
| 37(0.069) | 46.346(0.339) | 1.637(0.151) | 50.192(0.328) | 2.051(0.180) |
| 38(0.071) | 46.350(0.340) | 1.637(0.151) | 50.204(0.330) | 2.053(0.180) |
| 39(0.072) | 46.354(0.341) | 1.638(0.151) | 50.215(0.332) | 2.054(0.180) |
| 40(0.074) | 46.357(0.341) | 1.638(0.151) | 50.224(0.333) | 2.055(0.180) |
| 41(0.076) | 46.359(0.342) | 1.638(0.151) | 50.231(0.335) | 2.056(0.180) |
| 42(0.078) | 46.361(0.342) | 1.638(0.152) | 50.238(0.336) | 2.057(0.180) |
| 43(0.080) | 46.363(0.343) | 1.639(0.152) | 50.243(0.337) | 2.057(0.180) |
| 44(0.082) | 46.364(0.343) | 1.639(0.152) | 50.248(0.338) | 2.058(0.180) |
| 45(0.084) | 46.365(0.343) | 1.639(0.152) | 50.252(0.339) | 2.058(0.180) |
| 46(0.085) | 46.366(0.343) | 1.639(0.152) | 50.255(0.340) | 2.059(0.180) |
| 47(0.087) | 46.367(0.343) | 1.639(0.152) | 50.258(0.340) | 2.059(0.181) |


| $48(0.089)$ | $46.367(0.344)$ | $1.639(0.152)$ | $50.260(0.341)$ | $2.059(0.181)$ |
| :--- | :--- | :--- | :--- | :--- |
| $49(0.091)$ | $46.368(0.344)$ | $1.639(0.152)$ | $50.262(0.341)$ | $2.059(0.181)$ |
| $50(0.093)$ | $46.368(0.344)$ | $1.639(0.152)$ | $50.264(0.342)$ | $2.060(0.181)$ |
| $51(0.095)$ | $46.369(0.344)$ | $1.639(0.152)$ | $50.265(0.342)$ | $2.060(0.181)$ |
| $52(0.097)$ | $46.369(0.344)$ | $1.639(0.152)$ | $50.267(0.342)$ | $2.060(0.181)$ |
| $53(0.098)$ | $46.369(0.344)$ | $1.639(0.152)$ | $50.268(0.343)$ | $2.060(0.181)$ |
| $54(0.100)$ | $46.369(0.344)$ | $1.639(0.152)$ | $50.269(0.343)$ | $2.060(0.181)$ |
| $55(0.102)$ | $46.369(0.344)$ | $1.639(0.152)$ | $50.269(0.343)$ | $2.060(0.181)$ |
| $56(0.104)$ | $46.370(0.344)$ | $1.639(0.152)$ | $50.270(0.343)$ | $2.060(0.181)$ |
| $57(0.106)$ | $46.370(0.344)$ | $1.639(0.152)$ | $50.270(0.343)$ | $2.060(0.181)$ |
| $58(0.108)$ | $46.370(0.344)$ | $1.639(0.152)$ | $50.271(0.343)$ | $2.060(0.181)$ |
| $59(0.110)$ | $46.370(0.344)$ | $1.639(0.152)$ | $50.271(0.344)$ | $2.060(0.181)$ |
| $60(0.112)$ | $46.370(0.344)$ | $1.639(0.152)$ | $50.272(0.344)$ | $2.061(0.181)$ |

Table 9. Summary of the number of black rockfish fish sampled with age in maturity study and the expected probability of maturity with age.

| Age | No. of immature <br> fish | No. of mature fish | Expected <br> probability of <br> maturity |
| :--- | :--- | :--- | :--- |
| 4 | 1 | 0 | 0.01 |
| 5 | 12 | 0 | 0.02 |
| 6 | 50 | 1 | 0.05 |
| 7 | 73 | 7 | 0.09 |
| 8 | 65 | 13 | 0.17 |
| 9 | 38 | 22 | 0.29 |
| 10 | 22 | 12 | 0.45 |
| 11 | 6 | 15 | 0.62 |
| 12 | 2 | 5 | 0.76 |
| 13 | 2 | 2 | 0.87 |
| 14 | 0 | 2 | 0.93 |
| 15 | 0 | 0 | 0.96 |
| 16 | 0 | 0 | 0.98 |
| 17 | 0 | 0 | 0.99 |
| 18 | 0 | 0 | 1.00 |
| 19 | 0 | 0 | 1.00 |
| 20 | 0 | 2 | 1.00 |

Table 10. Summary of the number of black rockfish fish sampled with fork length in maturity study and the expected probability of maturity with fork length.

| Fork length (cm) | No. of immature <br> fish | No. of mature fish | Expected <br> probability of <br> maturity |
| ---: | :--- | :--- | :--- |
| 25 | 1 | 0 | 0 |
| 26 | 1 | 0 | 0 |
| 27 | 1 | 0 | 0 |
| 28 | 2 | 0 | 0 |
| 29 | 3 | 0 | 0 |
| 30 | 7 | 0 | 0.01 |
| 31 | 3 | 1 | 0.01 |
| 32 | 5 | 0 | 0.02 |
| 33 | 13 | 0 | 0.02 |
| 34 | 18 | 0 | 0.03 |
| 35 | 30 | 0 | 0.05 |
| 36 | 32 | 3 | 0.07 |
| 37 | 37 | 4 | 0.11 |
| 38 | 30 | 8 | 0.15 |
| 39 | 35 | 10 | 0.22 |
| 40 | 27 | 12 | 0.29 |
| 41 | 20 | 13 | 0.38 |
| 42 | 14 | 9 | 0.48 |
| 43 | 8 | 10 | 0.59 |
| 44 | 4 | 11 | 0.68 |
| 45 | 2 | 2 | 0.76 |
| 46 | 0 | 7 | 0.83 |
| 47 | 1 | 3 | 0.88 |
| 48 | 0 | 2 | 0.92 |
| 49 | 0 | 2 | 0.94 |

Table 11. Summary of the estimated total mortality coefficients of male and female black rockfish from 1984 to 2006.

| Year | Male |  | Female |  |
| :---: | ---: | ---: | ---: | ---: |
|  | N | $\hat{Z}$ (st. err.) | n | $\hat{Z}$ (st. err.) |
| 1984 | 267 | $0.162(0.068)$ | 429 | $0.267(0.005)$ |
| 1988 | 128 | $0.169(0.098)$ | 148 | $0.341(0.207)$ |
| 1989 | 180 | $0.256(0.112)$ | 217 | $0.205(0.071)$ |
| 1990 | 132 | $0.200(0.044)$ | 158 | $0.407(0.129)$ |
| 1991 | 326 | $0.213(0.050)$ | 394 | $0.259(0.031)$ |
| 1992 | 424 | $0.187(0.080)$ | 457 | $0.325(0.011)$ |
| 1993 | 364 | $0.270(0.048)$ | 495 | $0.277(0.028)$ |
| 1994 | 399 | $0.244(0.013)$ | 465 | $0.348(0.016)$ |
| 1995 | 372 | $0.304(0.009)$ | 440 | $0.370(0.039)$ |
| 1996 | 399 | $0.394(0.080)$ | 432 | $0.387(0.014)$ |
| 1997 | 437 | $0.298(0.079)$ | 438 | $0.361(0.031)$ |
| 1998 | 947 | $0.315(0.043)$ | 874 | $0.400(0.013)$ |
| 1999 | 851 | $0.320(0.034)$ | 822 | $0.353(0.013)$ |
| 2000 | 741 | $0.316(0.071)$ | 909 | $0.406(0.056)$ |
| 2001 | 800 | $0.353(0.026)$ | 974 | $0.427(0.053)$ |
| 2002 | 783 | $0.324(0.064)$ | 1066 | $0.298(0.057)$ |
| 2003 | 793 | $0.290(0.055)$ | 1009 | $0.327(0.069)$ |
| 2004 | 731 | $0.254(0.066)$ | 922 | $0.297(0.032)$ |
| 2005 | 681 | $0.238(0.092)$ | 982 | $0.339(0.069)$ |
| 2006 | 806 | $0.220(0.074)$ | 802 | $0.323(0.035)$ |

Table 12. Summary of the proportion by area and the number of recreational observations taken from 1990 to 2006.

| Year | Area |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 |  |  |  |  |
| 2 | 3 | 74 | 84 |  |  |
| 1990 | $5102(2.87 \%)$ | $159462(89.83 \%)$ | $2202(1.24 \%)$ | $5601(3.16 \%)$ | $5144(2.90 \%)$ |
| 1991 | $2156(1.43 \%)$ | $138150(91.69 \%)$ | $2602(1.73 \%)$ | $3122(2.07 \%)$ | $4643(3.08 \%)$ |
| 1992 | $3422(2.82 \%)$ | $97598(80.29 \%)$ | $4159(3.42 \%)$ | $10128(8.33 \%)$ | $6252(5.14 \%)$ |
| 1993 | $5636(5.13 \%)$ | $88923(81.01 \%)$ | $3153(2.87 \%)$ | $6115(5.57 \%)$ | $5942(5.41 \%)$ |
| 1994 | $7754(4.37 \%)$ | $148419(83.69 \%)$ | $7552(4.26 \%)$ | $7275(4.10 \%)$ | $6340(3.58 \%)$ |
| 1995 | $3442(2.42 \%)$ | $112959(79.57 \%)$ | $5118(3.61 \%)$ | $10172(7.17 \%)$ | $10271(7.24 \%)$ |
| 1996 | $5018(3.02 \%)$ | $133094(80.22 \%)$ | $4179(2.52 \%)$ | $8263(4.98 \%)$ | $15349(9.25 \%)$ |
| 1997 | $5771(4.67 \%)$ | $100816(81.61 \%)$ | $1729(1.40 \%)$ | $5814(4.71 \%)$ | $9400(7.61 \%)$ |
| 1998 | $8048(5.79 \%)$ | $110960(79.78 \%)$ | $2711(1.95 \%)$ | $4645(3.34 \%)$ | $12720(9.15 \%)$ |
| 1999 | $1951(1.77 \%)$ | $93642(84.92 \%)$ | $2801(2.54 \%)$ | $4412(4.00 \%)$ | $7470(6.77 \%)$ |
| 2000 | $3524(3.09 \%)$ | $93927(82.31 \%)$ | $3125(2.74 \%)$ | $6625(5.81 \%)$ | $6918(6.06 \%)$ |
| 2001 | $3814(4.01 \%)$ | $77415(81.37 \%)$ | $2232(2.35 \%)$ | $5322(5.59 \%)$ | $6355(6.68 \%)$ |
| 2002 | $4610(4.54 \%)$ | $79168(77.89 \%)$ | $2823(2.78 \%)$ | $8967(8.82 \%)$ | $6079(5.98 \%)$ |
| 2003 | $6589(7.25 \%)$ | $68067(74.87 \%)$ | $2735(3.01 \%)$ | $6757(7.43 \%)$ | $6766(7.44 \%)$ |
| 2004 | $4599(4.66 \%)$ | $74905(75.93 \%)$ | $3706(3.76 \%)$ | $6047(6.13 \%)$ | $9399(9.53 \%)$ |
| 2005 | $4136(3.43 \%)$ | $84719(70.28 \%)$ | $7052(5.85 \%)$ | $9351(7.76 \%)$ | $15280(12.68 \%)$ |
| 2006 | $5769(4.31 \%)$ | $106803(79.75 \%)$ | $4558(3.40 \%)$ | $6307(4.71 \%)$ | $10492(7.83 \%)$ |

Table 13. Summary of the recreational fishery CPUEs estimated from mean estimator and delta lognormal model for all areas.

| Year | Total catch/total anglers |  |  |  | Delta lognormal model |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Estimates | $q_{\bar{X}, 2.5 \%}$ | $q_{\bar{X}, 97.5 \%}$ | $C V_{\bar{X}}$ | Estimates | $q_{\bar{X}, .5 \%}$ | $q_{\bar{X}, 97.5 \%}$ | $C V_{\bar{X}}$ |
| 1990 | 8.58 | 8.33 | 8.85 | 0.02 | 5.73 | 5.52 | 5.92 | 0.02 |
| 1991 | 7.37 | 7.18 | 7.60 | 0.02 | 5.43 | 5.24 | 5.61 | 0.02 |
| 1992 | 6.14 | 5.92 | 6.37 | 0.02 | 4.77 | 4.63 | 4.92 | 0.02 |
| 1993 | 5.83 | 5.61 | 6.11 | 0.02 | 4.24 | 4.13 | 4.42 | 0.02 |
| 1994 | 6.87 | 6.70 | 7.04 | 0.01 | 4.43 | 4.29 | 4.56 | 0.01 |
| 1995 | 5.94 | 5.75 | 6.10 | 0.01 | 4.07 | 3.94 | 4.18 | 0.02 |
| 1996 | 6.37 | 6.22 | 6.53 | 0.01 | 4.57 | 4.45 | 4.69 | 0.01 |
| 1997 | 5.78 | 5.64 | 5.94 | 0.02 | 3.93 | 3.81 | 4.05 | 0.02 |
| 1998 | 6.35 | 6.17 | 6.50 | 0.01 | 4.80 | 4.66 | 4.91 | 0.01 |
| 1999 | 6.93 | 6.73 | 7.07 | 0.01 | 4.86 | 4.70 | 4.99 | 0.02 |
| 2000 | 6.83 | 6.63 | 6.98 | 0.01 | 5.03 | 4.87 | 5.18 | 0.02 |
| 2001 | 6.46 | 6.25 | 6.66 | 0.01 | 4.29 | 4.13 | 4.44 | 0.02 |
| 2002 | 7.03 | 6.86 | 7.20 | 0.01 | 5.01 | 4.86 | 5.17 | 0.02 |
| 2003 | 6.93 | 6.75 | 7.12 | 0.01 | 4.95 | 4.75 | 5.14 | 0.02 |
| 2004 | 7.14 | 6.94 | 7.33 | 0.01 | 5.57 | 5.41 | 5.73 | 0.02 |
| 2005 | 6.98 | 6.80 | 7.13 | 0.01 | 5.36 | 5.21 | 5.48 | 0.01 |
| 2006 | 7.29 | 7.15 | 7.42 | 0.01 | 5.20 | 5.06 | 5.33 | 0.01 |

Table 14. Summary of the recreational sport CPUEs estimated from mean estimator and delta lognormal model for Area 2.

| Year | Total catch/total anglers |  |  |  | Delta lognormal model |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
|  | Estimates | $q_{\bar{X}, 2.5 \%}$ | $q_{\bar{X}, 97.5 \%}$ | $C V_{\bar{X}}$ | Estimates | $q_{\bar{X}, .5 \%}$ | $q_{\bar{X}, 97.5 \%}$ | $C V_{\bar{X}}$ |  |
| 1990 | 10.98 | 10.66 | 11.29 | 0.02 | 10.84 | 10.51 | 11.18 | 0.01 |  |
| 1991 | 8.75 | 8.54 | 8.96 | 0.01 | 8.35 | 8.11 | 8.56 | 0.01 |  |
| 1992 | 7.35 | 7.01 | 7.63 | 0.02 | 6.85 | 6.53 | 7.11 | 0.02 |  |
| 1993 | 7.52 | 7.24 | 7.85 | 0.02 | 7.16 | 6.92 | 7.48 | 0.02 |  |
| 1994 | 9.64 | 9.43 | 9.86 | 0.01 | 9.33 | 9.13 | 9.55 | 0.01 |  |
| 1995 | 8.31 | 8.16 | 8.46 | 0.01 | 7.81 | 7.64 | 8.01 | 0.01 |  |
| 1996 | 8.03 | 7.83 | 8.23 | 0.01 | 7.63 | 7.45 | 7.81 | 0.01 |  |
| 1997 | 7.23 | 7.01 | 7.44 | 0.01 | 6.37 | 6.12 | 6.59 | 0.02 |  |
| 1998 | 7.44 | 7.20 | 7.63 | 0.01 | 6.76 | 6.57 | 6.95 | 0.01 |  |
| 1999 | 8.54 | 8.35 | 8.72 | 0.01 | 7.70 | 7.46 | 7.91 | 0.02 |  |
| 2000 | 8.36 | 8.18 | 8.58 | 0.01 | 7.80 | 7.60 | 7.99 | 0.01 |  |
| 2001 | 8.25 | 8.03 | 8.47 | 0.01 | 7.08 | 6.81 | 7.35 | 0.02 |  |
| 2002 | 8.85 | 8.63 | 9.05 | 0.01 | 7.95 | 7.68 | 8.24 | 0.02 |  |
| 2003 | 8.46 | 8.24 | 8.68 | 0.01 | 6.83 | 6.51 | 7.11 | 0.02 |  |
| 2004 | 8.10 | 7.86 | 8.31 | 0.01 | 6.86 | 6.58 | 7.12 | 0.02 |  |
| 2005 | 8.77 | 8.60 | 8.93 | 0.01 | 7.80 | 7.60 | 8.03 | 0.02 |  |
| 2006 | 8.92 | 8.78 | 9.05 | 0.01 | 8.16 | 7.96 | 8.33 | 0.01 |  |

Table 15. Central Washington coastal tagging mean catch per trip (catch/hours fished).

| Central Washington Coast Tagging CPUE <br> Mean Catch Per Hour <br> Year <br> Across All Trips | $\operatorname{In}(\mathbf{1 + c v})$ |  |
| :--- | :---: | :---: |
| 1981 | 4.8 | 0.666 |
| 1986 | 2.3 | 0.5993 |
| 1987 | 1.2 | 0.6344 |
| 1988 | 0.8 | 0.5539 |
| 1989 | 1.2 | 0.9771 |
| 1990 | 1.0 | 0.8439 |
| 1998 | 2.5 | 0.813 |
| 1999 | 3.1 | 0.7407 |
| 2000 | 2.2 | 0.5684 |
| 2001 | 4.7 | 0.6076 |
| 2002 | 5.5 | 0.5034 |
| 2003 | 6.2 | 0.5913 |
| 2004 | 9.4 | 0.5149 |
| 2005 | 10.2 | 0.7579 |
| 2006 | 10.5 | 0.4205 |

Table 16. Summary of the return of tagged fish from the CWT double tags experiment.

| Year |  | i | No. of one tag return ( $r_{\text {si }}$ ) |  | No. of two tags return ( $r_{d i}$ ) |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | left | right |  |
| 1998 | 0 |  | 8 | 17 | 691 |
| 1999 | 1 |  | 14 | 11 | 542 |
| 2000 | 2 |  | 14 | 18 | 433 |
| 2001 | 3 |  | 14 | 8 | 276 |
| 2002 | 4 |  | 6 | 8 | 160 |
| 2003 | 5 |  | 2 | 2 | 73 |
| 2004 | 6 |  | 0 | 2 | 34 |

Table 17. Summary of the year, the no. of fish tagged, no. of fish sampled, the numbers of fish return with tags, tag on the right, tag on the left, double tag, the estimated population size and variance, the adjusted no. of tag return with tag loss, the estimated population size with tag loss adjustment and variance.

| Year | $n_{1}$ | $n_{2}$ | $m$ | $m_{r}$ | $m_{l}$ | $m_{d}$ | $\hat{N}$ | $\operatorname{Var}(\hat{N})$ | $\hat{\ddot{m}}$ | $\hat{N}$ |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1998 | 2456 | 46951 | 14 | 1 | 1 | 12 | $7.69 \mathrm{E}+06$ | $3.67 \mathrm{E}+12$ | 14.08 | $7.65 \mathrm{E}+06$ | $4.53 \mathrm{E}+12$ |
| 1999 | 3479 | 66253 | 43 | 1 | 0 | 42 | $5.24 \mathrm{E}+06$ | $6.02 \mathrm{E}+11$ | 43.01 | $5.24 \mathrm{E}+06$ | $6.46 \mathrm{E}+11$ |
| 2000 | 2789 | 65276 | 130 | 3 | 5 | 122 | $1.39 \mathrm{E}+06$ | $1.39 \mathrm{E}+10$ | 130.13 | $1.39 \mathrm{E}+06$ | $1.53 \mathrm{E}+10$ |
| 2001 | 3210 | 64440 | 68 | 2 | 1 | 65 | $3.00 \mathrm{E}+06$ | $1.26 \mathrm{E}+11$ | 68.03 | $3.00 \mathrm{E}+06$ | $1.35 \mathrm{E}+11$ |
| 2002 | 4089 | 68475 | 143 | 1 | 1 | 141 | $1.94 \mathrm{E}+06$ | $2.51 \mathrm{E}+10$ | 143.01 | $1.94 \mathrm{E}+06$ | $2.66 \mathrm{E}+10$ |
| 2003 | 6747 | 77622 | 246 | 1 | 8 | 237 | $2.12 \mathrm{E}+06$ | $1.74 \mathrm{E}+10$ | 246.09 | $2.12 \mathrm{E}+06$ | $1.86 \mathrm{E}+10$ |
| 2004 | 4209 | 53385 | 74 | 1 | 1 | 72 | $3.00 \mathrm{E}+06$ | $1.16 \mathrm{E}+11$ | 74.01 | $3.00 \mathrm{E}+06$ | $1.23 \mathrm{E}+11$ |
| 2005 | 3913 | 70482 | 54 | 0 | 0 | 54 | $5.02 \mathrm{E}+06$ | $4.43 \mathrm{E}+11$ | 54.00 | $5.02 \mathrm{E}+06$ | $4.66 \mathrm{E}+11$ |

Table 18. Likelihood components from the STAR base (top) and STAT best-fit (bottom) northern black rockfish models.

| STAR Base Model |  |  |
| :--- | :---: | ---: |
| Likelihood Components | Emphasis | Likelihood |
| indices |  |  |
| $\quad$ Tag Abundance | 1.0 | 43.4 |
| Tag CPUE | 1.0 | 11.7 |
| discard | 0.0 | 0.0 |
| length_comps |  |  |
| $\quad$ Trawl | 0.1 | 67.6 |
| Sport | 0.1 | 32.3 |
| Non-Trawl | 0.1 | 38.1 |
| Tag | 0.1 | 18.3 |
| age_comps |  |  |
| Trawl | 1.0 | 187.2 |
| Sport | 1.0 | 395.3 |
| Non-Trawl | 1.0 | 187.0 |
| size-at-age | 0.0 | 105.9 |
| mean_body_wt | 0.0 | 0.0 |
| Equil_catch | 1.0 | 0.0 |
| Recruitment | 0.1 | 14.5 |
| Parm_priors | 1.0 | 0.0 |
| Parm_devs | 0.1 | 0.0 |
| penalties | 0.0 | 0.0 |
| Forecast_Recruitment | 0.0 | 0.2 |
|  |  | 1101.6 |


| STAT Best Fit Model <br> Likelihood Components | Emphasis | Likelihood |
| :--- | :---: | ---: |
| indices |  |  |
| $\quad$ Tag Abundance | 1.0 | 41.5 |
| Tag CPUE | 1.0 | 10.4 |
| discard | 0.0 | 0.0 |
| length_comps |  |  |
| $\quad$ Trawl | 0.1 | 69.2 |
| Sport | 0.1 | 32.5 |
| $\quad$ Non-Trawl | 0.1 | 39.4 |
| $\quad$ Tag |  | 19.0 |
| age_comps | 1.0 |  |
| $\quad 1.0$ | 180.6 |  |
| Trawl | 1.0 | 386.8 |
| Sport | 0.0 | 185.7 |
| Non-Trawl | 0.0 | 106.5 |
| size-at-age | 1.0 | 0.0 |
| mean_body_wt | 0.1 | 0.0 |
| Equil_catch | 1.0 | 15.4 |
| Recruitment | 0.1 |  |
| Parm_priors | 0.0 | 0.0 |
| Parm_devs | 0.0 | 0.0 |
| penalties |  | 0.2 |
| Forecast_Recruitment |  | 1087.15 |

Table 19. Assumptions and Priors used in the Northern black rockfish STAR base model. The only change in the STAT Best Fit model is an increase in the "old" female natural mortality rate from 0.20 to 0.24 .


Table 19. Continued.


Table 20. Average Pearson residual by fishery (Trawl=1, Sport=2, Non-Trawl=3) by likelihood component.

| Used | 1 |
| :--- | :--- |
| year | (All) |


| kind fleet \|season |mkt |  |  |  | Data |  |  |  |  |  | For Age \& Len: effN/inputN |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Average of Average of effN |  | Average ( Min of Pe Max of P¢StdDev of Pearson |  |  |  |  |
| AGE | 1 | 1 | 0 | 66 | 103 | -0.026 | -2.39 | 4.86 | 0.823880087 | 1.57 |
|  | 2 | 1 | 0 | 312 | 343 | 0.027 | -2.67 | 3.84 | 0.885559936 | 1.10 |
|  | 3 | 1 | 0 | 116 | 157 | 0.004 | -2.44 | 8.05 | 0.988462445 | 1.35 |
| LEN | 1 | 1 | 0 | 110 | 44 | 0.785 | -3.78 | 14.38 | 2.878351254 | 0.40 |
|  | 2 | 1 | 0 | 135 | 340 | 0.161 | -3.08 | 7.96 | 1.137126017 | 2.52 |
|  | 3 | 1 | 0 | 199 | 139 | 0.430 | -3.72 | 9.64 | 2.065073952 | 0.70 |
|  | 6 | 1 | 0 | 217 | 204 | 0.001 | -2.71 | 4.79 | 1.166505276 | 0.94 |
| L@A | 2 | 1 | 0 | 42 | 1.2326 | -0.454 | -4.37 | 3.95 | 1.501360074 | 0.03 |

Table 21. Average Pearson residual by fishery (Trawl=Top 2 rows, Sport=Middle 2 rows, Non-Trawl Bottom 2 rows) , age and sex.

Mean Pearson residule by Age and Feet (TraM=1, Sport =2 and Non-TraN=3)

| Average of Pearson |  | bin |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| fleet | gender | 34 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 30 | Total |
| 1 | 1 | 0.33-0.35-0.22-0.38-0.06-0.02 |  |  |  |  | 0.29 | -0.10 | 0.11 | 0.20 | 0.08 | -0.30 | -0.35 | 0.40 | -0.24 | -0.45 |  | -0.44 | -0.49 | -0.62 | -0.31 | -0.29 | -0.63 | -0.16 | -0.20 |
|  | 2 | $195-0.190 .00$ |  | 0.21 | 0.64 | 101 | 0.82 | 0.43 | -0.20 | 0.24 | 0.24 | -0.13 | -0.04 | -0.08 | -0.20 | -0.26 | -0.27 | -0.19 | 0.23 | 0.00 | -0.13 | -0.03 | -0.04 | 0.03 | 0.12 |
| 2 | 1 | $\begin{array}{cc} -0.22 & 0.00 \\ -0.63 & -0.25 \end{array}$ | $\begin{array}{llll}  & 0.17 & 0.05 & 0.16 \\ 5 & 0.04 & -0.02 & -0.09 \end{array}$ |  |  | 0.59 | 0.46 | 0.56 | 0.53 | 0.46 | 0.32 |  | 0.26 | -0.01 | -0.08 | -0.05 | -0.18 | -0.27 | -0.25 | -0.43 | -0.29 | -0.42 | -0.33 | 0.60 | 0.10 |
|  | 2 |  |  |  |  | -0.09 | -0.68 | -0.90 | -0.64 | -0.81 | -0.43 | -0.28 | 0.02 | -0.05 | 0.26 | 0.23 | 0.25 | 0.58 | 0.45 | 0.25 | 0.36 | 0.40 | 0.47 | 0.47 | -0.04 |
| 3 | 1 | $\begin{array}{rr} -0.15 & 0.73 \\ 0.00 & -0.24 \end{array}$ | $\begin{array}{rrr} 0.80 & 0.24 & 0.25 \\ 0.58 & -0.05 & -0.31 \end{array}$ |  |  | 0.09 | 0.22 | -0.36 | -0.09 | 0.21 | 0.48 |  | 0.30 | 0.28 | -0.10 | 0.00 | -0.29 | -0.32 | -0.56 | 0.09 | -0.57 | -0.40 | -0.47 | 125 | 0.10 |
|  | 2 |  |  |  |  | -0.29 | -0.58 | -0.25 | -0.68 | -0.24 | -0.05 | -0.10 | -0.10 | -0.30 | 0.05 | -0.16 | -0.02 | 0.19 | 0.11 | 0.02 | 0.15 | -0.01 | 0.05 | 0.82 | -0.07 |
| Grand Total |  | -0.20-0.02 | 0.24 | 0.00 | 0.09 | 0.25 | 0.08 | -0.10 | -0.13 | -0.02 | 0.07 | -0.12 | 0.02 | -0.09 | -0.03 | -0.08 | -0.12 | 0.01 | 0.02 | -0.08 | -0.02 | 0.00 | 0.06 | 0.47 | 0.01 |

Table 22. Model sensitivity by likelihood component to rates of natural mortality relative to that assumed in the STAR base model. Values represent the change in likelihood relative to the base models such that negative values indicate a better fit. Female natural mortality rates for ages less than 11 are assumed to be equal to that assumed for males.


Note: Square indcates the Base Model and bold font indicates best fit.

Table 22. Continued.


[^0]Table 23. Comparison of Councils' default target fishing mortality rates and reference points between the STAR base model and STAT best fit model. The default target fishing mortality rate of FSPR=0.5 is used in this assessment for both models and that used for other Council managed rockfish species.

STAR Base Model Results

| Unfished Stock | Value |  |  |
| :---: | :---: | :---: | :---: |
| Age 3+ Biomass ( $\mathrm{B}_{0}$ ) (mt) | 10,813 |  |  |
| Spawning Biomass $\mathrm{SB}_{(0)}$ (mt) | 2,429 |  |  |
| SPBio/Recruit (kg/fish) | 0.780 |  |  |
| Age1 Recruitment ( $\mathrm{R}_{0}$ ) (1,000's) | 3,113 |  |  |
| Steepness_R0_S0 | 0.6 |  |  |
|  | Reference points based on |  |  |
| Exploited Stock | Estimated MSY | SB ${ }_{40 \%}$ | SPR ( $\mathrm{SB}_{0.5}$ ) |
| SPR (Spawning Biomass/Recruit) | 0.413 | 0.400 | 0.400 |
| F (Fishing Mortality Rate) | 0.132 | 0.101 | 0.101 |
| Exploitation Rate (Yield/Bsmry) | 0.076 | 0.060 | 0.060 |
| MSY (mt) or MSY proxy (mt) | 377 | 361 | 361 |
| Yield (mt) | 718 | 972 | 972 |
| SPBIO/SB(0) | 29.6\% | 40.0\% | 40.0\% |
| Age 3+ Biomass | 4,947 | 6,012 | 6,012 |

STAT Best Fit Model Results

| Unfished Stock | Value |
| :---: | :---: |
| Age 3+ Biomass $\left(B_{0}\right)(m t)$ | 11,390 |
| Spawning Biomass $S B\left({ }_{0}\right)(m t)$ | 2,321 |
| SPBio/Recruit $(k g / f i s h)$ | 0.687 |
| Age1 Recruitment $\left(R_{0}\right)(1,000$ 's) | 3,377 |
| Steepness_RO_SO | 0.6 |


|  | Reference points based on <br> Exploited Stock |  |  |
| :--- | :---: | ---: | ---: |
| Sstimated MSY | SB $_{40 \%}$ | SPR (SB $\left._{0.5}\right)$ |  |
| SPR (Spawning Biomass/Recruit) | 0.418 | 0.400 | 0.40 |
| Exploitation Rattality Rate) | 0.141 | 0.110 | 0.110 |
| MSY (mt) or MSY proxy (mt) | 0.081 | 0.065 | 0.065 |
| Yield (mt) | 423 | 408 | 408 |
| SPBIO/SB(0) | 700 | 928 | 928 |
| Age 3+ Biomass | $30.1 \%$ | $40.0 \%$ | $40.0 \%$ |
|  | 5,218 | 6,264 | 6,264 |

Table 24. Comparison of ABC's, Spawning biomass and depletion between the STAR base (top) and STAT best fit model (bottom).

STAR Base Model

| Year | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 9}$ | $\mathbf{2 0 1 0}$ | $\mathbf{2 0 1 1}$ | $\mathbf{2 0 1 2}$ | $\mathbf{2 0 1 3}$ | $\mathbf{2 0 1 4}$ | $\mathbf{2 0 1 5}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| ABC (mt) | 394 | 377 | 361 | 350 | 345 | 344 | 346 | 350 | 354 |
| Spawning Biomass (mt) | 1064 | 1071 | 1060 | 1036 | 1005 | 977 | 956 | 944 | 940 |
| \% of Virgin | 0.438 | 0.441 | 0.436 | 0.426 | 0.414 | 0.402 | 0.394 | 0.389 | 0.387 |

## STAT "Best Fit" Model

| Year | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 9}$ | $\mathbf{2 0 1 0}$ | $\mathbf{2 0 1 1}$ | $\mathbf{2 0 1 2}$ | $\mathbf{2 0 1 3}$ | $\mathbf{2 0 1 4}$ | $\mathbf{2 0 1 5}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| ABC (mt) | 535 | 503 | 474 | 453 | 440 | 433 | 431 | 432 | 434 |
| Spawning Biomass (mt) | 1281 | 1267 | 1233 | 1182 | 1126 | 1074 | 1033 | 1005 | 989 |
| \% of Virgin | 0.552 | 0.546 | 0.531 | 0.509 | 0.485 | 0.463 | 0.445 | 0.433 | 0.426 |

Figures


Figure 1. Location of black rockfish tag release area (top) and tag recovery locations (bottom).


Figure 2. Dendogram showing results of cluster analysis of ten black rockfish collections using Nei's (1978) unbiased genetic distance at 20 polymorphic loci.


Figure 3. Relative abundance of females with age in pooled age data for Washington fisheries.


Figure 4. Total black rockfish catch by gear and year for areas North of Cape Falcon to the U.S. Canadian border.


Figure 5. Regulation changes in commercial fisheries


Figure 6. Maximum retainable rockfish catch per trip for the sport fisheries.


Figure 7. Maximum retainable black for the limited entry commercial fisheries.


Figure 8. Comparison of catch estimates between the 1999 and the current assessment.


Figure 9. Proportion at size by sex and fisheries from 1984 to 2006.


Figure 10. Proportion at age by sex and fisheries from 1984 to 2006.



Figure 11. Scatter plot of fork length and weight of male (top panel) and female (bottom panel) black rockfish and the expected length weight relationship.


Figure 12. Scatter plots of male (top panel) and female (bottom panel) age and fork length data and the estimated growth curves.


Figure 13. Scatter plot of age reading from two independent age readers and the expected relationship of age reading between the two age readers.


Figure 14. Standard deviation of ageing error.


Figure 15. Plot of the estimated probability of maturity against the estimated age of female black rockfish. The intervals are the $95 \%$ confidence intervals estimated by bootstrapping.


Figure 16. Plot of the estimated probability of maturity against the fork length of female black rockfish. The intervals are the $95 \%$ confidence intervals estimated by bootstrapping.



Figure 17. Relationship between fecundity and size (top panel) and fecundity and body weight (bottom panel).


Figure 18. Plot of expected male (top panel) and female (bottom panel) estimated total mortality coefficients against total fishing effort. The estimated intercept in each sub graph was the estimated natural mortality.


Figure 19. Scatter plot of estimated female black rockfish mortality coefficients versus estimated male black rockfish mortality coefficients, and the estimated linear relationship.


Figure 20. Time series plot of the estimated male and female black rockfish total mortalities.


Figure 21. Time series plot of the estimated CPUEs of recreational survey data in all areas from 1990 to 2006. The estimated CPUEs were done by mean estimator and delta lognormal model.


Figure 22. Time series plot of the estimated CPUEs of recreational survey data in Area 2 from 1990 to 2006. The estimated CPUEs were done by mean estimator and delta lognormal model.


Figure 23. Plot of accumulated CWT tag lost rate with time.


Figure 24. Time series of the tagging CPUE of the central Washington coast.


Figure 25. Comparison of natural mortality rates between males and females as defined in the STAT Best Fit Model. In the STAR base model Female natural mortality asymptotes at 0.20 at age 15 instead of 0.24 in the STAT Best Fit model.


Figure 26. Comparison of fishing mortality rates estimated from STAR Base, STAT Best Fit model and the tagging model (assuming $\mathrm{M}=0.2$ for both sexes).



Figure 27. Convergence properties of the STAR base model.


Figure 28. Variation in fork-length at age by sex.


Figure 29. Comparison of growth curves estimated from STAR base, STAT best-fit model and external estimates to the mean size at observed age.


Figure 30. STAR base and STAT "best fit" model fit to tagging abundance (top panel) and tagging CPUE (bottom panel) data by fishery.


Figure 31. STAR base model sensitivity to increased weight on the tagging CPUE and tagging population estimates of abundance.


Figure 32. Retrospective analysis of the northern black rockfish STAR base (top panel) and STAT best-fit (bottom panel) models. Observation data are ignored and there is no recruitment deviations estimated beyond retrospective year

Female whole catch length fits for fleet 1


Male whole catch length fits for fleet 1


Figure 33. STAR base model fit to female (top) and male (bottom) length composition samples collected from the trawl fishery.


Figure 34. STAR base model fit to female (top) and male (bottom) length composition samples collected from the sport fishery.

Female whole catch length fits for fleet 3


Male whole catch length fits for fleet 3


Figure 35. STAR base model fit to female (top panel) and male (bottom panel) length composition samples collected from the non-trawl fishery.

## Combined sex whole catch length fits for fleet 6



Figure 36. STAR base model fit to male length composition samples (combined sex) collected from the trawl fishery.

Female whole catch age fits for fleet 1


Male whole catch age fits for fleet 1


Figure 37. STAR base model fit to female (top panel) and male (bottom panel) age composition samples collected from the trawl fishery.

Female whole catch age fits for fleet 2


Male whole catch age fits for fleet 2


Figure 38. STAR base model fit to female (top panel) and male (bottom panel) age composition samples collected from the sport fishery.


Figure 39. STAR base model fit to female (top panel) and male (bottom panel) age composition samples collected from the non-trawl fishery.

$\rightarrow$ Current Assessment - - 1999 Assessment - - 1994 Assessment


Figure 40. Comparison of STAR base model recruitment estimates to the previous northern black rockfish assessments (top panel) and to the STAT best-fit model recruitment estimates (bottom panel).


Figure 41. Northern black rockfish STAR base model estimated fishing mortality rates by year and fishery (top panel) and cumulative fishing mortality by year and fishery (bottom panel).


Figure 42. Trawl, sport, non-trawl, and tagging survey selectivity estimated by the STAR base model.


Figure 43. Comparison of STAR base model spawning biomass estimates to the previous two assessments for the northern black rockfish assessment.


Figure 44. Comparison of STAR base model spawning biomass estimates to the STAT best-fit spawning biomass estimates for the Northern black rockfish assessment.


Figure 45. Comparison of stock abundance resulting from the STAT base model (left panel) and the STAT best fit model (right panel) to the Councils’ minimum stock size threshold and management target.


Figure 46. Plot of population status in relation to fishery management benchmarks for the STAR base (left panels) and STAT best fit models (right panels). Fspr versus (Bunfished/B40) in top panel and spawning depletion in relation to management target of B40\% and B $25 \%$ in bottom panel.

Changes in total log likelihood relative to Base model


Figure 47. Change in total likelihood relative to the STAR base model. Negative values indicate a better fit.


Figure 48. Likelihood profile of the STAR base assessment model for different fixed values of the Beverton-Holt steepness parameter (h) and Sigma R. The STAR base and STAT best-fit model had the steepness fixed at 0.6 and Sigma R tuned to 0.35.


Figure 49. Likelihood profile for various components following simultaneous changes in the emphasis (weight) of the Recruitment Dev and Recruitment Dev time series for the STAR base (top panel) and STAT best-fit models (bottom panel).


Figure 50. Likelihood profile for various components of the STAR base model (top panel) and STAT best-fit model (bottom panel) following changes in the emphasis (weight) on the tagging abundance and tagging CPUE indices.


Figure 51. Likelihood profile for various components of the STAR base model and the STAT best-fit model following changes in the emphasis (weight) on the age composition for all fisheries.


Figure 52. Likelihood profile for various components of the STAR base model and STAT best-fit model following changes in the emphasis (weight) on the length composition for all fisheries.

## Appendix A: SS2 2.00c Control and Data Files




3 \#_SR_function



| 1 | 19 | 19 | 5 | -1 | 0.05 | -3 | 0 | 0 | 0 | $\bigcirc$ | 0.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | \#MirrorTagCPUE |  |  |  |  |  |  |  |  |
| \#_size_sel:6 |  |  |  |  |  |  |  |  |  |  |  |
| \#1 | 19 | 30 | 5 | -1 | 0.05 | 2 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | $\bigcirc$ | \# |  |  |  |  |  |  |  |  |
| \#1 | 19 | 19 | 5 | -1 | 0.05 | 3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \# |  |  |  |  |  |  |  |  |
| 20 | 60 | 39.513 | 41.2 | -1 | 0.05 | 2 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| -6 | 4 | -3.41 | -2.6 | -1 | 0.05 | 3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| -1 | 9 | 3.7 | 5.2 | -1 | 0.05 | 3 | 0 | 0 | 0 | 0 | 0.5 |
|  | $\bigcirc$ | 0 |  |  |  |  |  |  |  |  |  |
| -1 | 9 | 3.5 | 6 | -1 | 0.05 | 3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| -8 | 9 | -4.69 | -3.7 | -1 | 0.05 | 2 | 0 | 0 | 0 | $\bigcirc$ | 0.5 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| -5 | 9 | -3.95 | 0.1 | -1 | 0.05 | 2 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| \#_age_sel:1 |  |  |  |  |  |  |  |  |  |  |  |
| \#0 | 40 | 1 | 5 | 0 | 99 | 2 | 0 | 0 | 0 | $\bigcirc$ | 0.5 |
|  | 0 | 0 | \# | 3 |  |  |  |  |  |  |  |
| \#0. 01 | 10 | 2 | 2 | 0 | 99 | 3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \# | 4 |  |  |  |  |  |  |  |
| \#1 | 20 | 7 | 5 | 0 | 99 | 2 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \# | 5 |  |  |  |  |  |  |  |
| \#1 | 25 | 17 | 2 | 0 | 99 | 3 | 0 | 0 | 0 | $\bigcirc$ | 0.5 |
|  | 0 | 0 | \# | 6 |  |  |  |  |  |  |  |
| \#_age_sel:2 |  |  |  |  |  |  |  |  |  |  |  |
| \#1 | 20 | 7 | 5 | 0 | 99 | 2 | 0 | $\bigcirc$ | 0 | 0 | 0.5 |
|  | 0 | 0 | \# | 5 |  |  |  |  |  |  |  |
| \#1 | 25 | 17 | 2 | 0 | 99 | 3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \# | 6 |  |  |  |  |  |  |  |
| \#_age_sel:3 |  |  |  |  |  |  |  |  |  |  |  |
| \#0 | 40 | 1 | 5 | 0 | 99 | 2 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \# | 7 |  |  |  |  |  |  |  |
| \#0. 01 | 10 | 2 | 2 | 0 | 99 | 3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \# | 8 |  |  |  |  |  |  |  |
| \#1 | 20 | 7 | 5 | 0 | 99 | 2 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \# | 5 |  |  |  |  |  |  |  |
| \#1 | 25 | 17 | 2 | 0 | 99 | 3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \# | 6 |  |  |  |  |  |  |  |
| \#_age_sel:4 |  |  |  |  |  |  |  |  |  |  |  |
| \#0 | 40 | 1 | 5 | 0 | 99 | 2 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \# | 9 |  |  |  |  |  |  |  |
| \#0.01 | 10 | 2 | 2 | 0 | 99 | 3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \# | 10 |  |  |  |  |  |  |  |
| 1 | \#_Selparm_Adjust_Method |  |  |  |  |  |  |  |  |  |  |
| 0 \#_custom_sel-env_setup |  |  |  |  |  |  |  |  |  |  |  |
| 0 \#_custom_sel-block_setup |  |  |  |  |  |  |  |  |  |  |  |
| \#-6 | $\# \#$     <br> 60 44 -2.6 1  <br> \#     |  |  |  |  |  |  |  |  |  |  |
| \#-10 | 10 | . 0 | . 1 | 1 | 0.05 | 4 |  |  |  |  |  |
| \#-10 | 10 | . 0 | . 1 | 1 | 0.05 | 4 |  |  |  |  |  |
| -1 \#_selparmdev-phase |  |  |  |  |  |  |  |  |  |  |  |
| \#_Variance_adjustments_to_input_values |  |  |  |  |  |  |  |  |  |  |  |
| \#_1 2 |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | \#_add_to_survey_cV |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | \#_add_to_discard_CV |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | \#_add_to_bodywt_CV |  |  |  |  |  |
| 3 | 4 | 3 | 1 | 1 | 3 | \#_m | by | mp |  |  |  |




| 0 | 0 | 0 | \# | 1915 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | $\bigcirc$ | \# | 1916 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1917 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1918 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1919 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1920 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1921 |  |  |  |  |  |  |
| $\bigcirc$ | 0 | 0 | \# | 1922 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1923 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1924 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1925 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1926 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1927 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1928 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1929 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1930 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1931 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1932 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1933 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1934 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1935 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1936 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1937 |  |  |  |  |  |  |
| 0 | 0 | 0 | \# | 1938 |  |  |  |  |  |  |
| 0 | 1 | 0 | \# | 1939 |  |  |  |  |  |  |
| 0 | 2.8 | 0 | \# | 1940 | Landings | MT | 1.4 | 312 | 0.4 | 2 |
|  | 2 | 1 |  |  |  |  |  |  |  |  |
| 3.2 | 4.6 | 0 | \# | 1941 | Landings | MT | 1.4 | 312 | 1.2 | 3.2 |
|  | 2.8 | 0 |  |  |  |  |  |  |  |  |
| 9.2 | 6.3 | 0 | \# | 1942 | Landings | MT | 0 | 315.5 | 0.9 | 9.2 |
|  | 4.6 | 0 |  |  |  |  |  |  |  |  |
| 15.2 | 8.1 | 0 | \# | 1943 | Landings | MT | 0.6 | 257.6 | 0.7 | 15.2 |
|  | 6.3 | 0 |  |  |  |  |  |  |  |  |
| 21.2 | 9.8 | 0 | \# | 1944 | Landings | MT | 0.1 | 232.2 | 0.2 | 21.2 |
|  | 8.1 | 0 |  |  |  |  |  |  |  |  |
| 27.2 | 11.6 | 0 | \# | 1945 | Landings | MT | 0.2 | 232.9 | 1.5 | 27.2 |
|  | 9.8 | 0 |  |  |  |  |  |  |  |  |
| 33.2 | 13.3 | 0 | \# | 1946 | Landings | MT | 0 | 188.7 | 1.1 | 33.2 |
|  | 11.6 | 0 |  |  |  |  |  |  |  |  |
| 39.2 | 15.1 | 0 | \# | 1947 | Landings | MT | 0 | 224.8 | 1.2 | 39.2 |
|  | 13.3 | 0 |  |  |  |  |  |  |  |  |
| 52 | 16.8 | $\bigcirc$ | \# | 1948 | Landings | MT | 0 | 221.6 | 4.3 | 5.2 |
|  | 15.1 | 0 |  |  |  |  |  |  |  |  |
| 51.2 | 18.6 | 0 | \# | 1949 | Landings | MT | 73.1 | 259.4 | 4.8 | 51.2 |
|  | 16.8 | 0 |  |  |  |  |  |  |  |  |
| 57.2 | 20.3 | 0 | \# | 1950 | Landings | MT | 9 | 234.1 | 15 | 57.2 |
|  | 18.6 | 0 |  |  |  |  |  |  |  |  |
| 63.2 | 22.1 | 1.5 | \# | 1951 | Landings | MT | 0 | 264.2 | 8.6 | 63.2 |
|  | 20.3 | 1.5 |  |  |  |  |  |  |  |  |
| 69.2 | 23.8 | 2.5 | \# | 1952 | Landings | MT | 3.3 | 264.8 | 65.8 | 69.2 |
|  | 22.1 | 2.5 |  |  |  |  |  |  |  |  |
| 75.2 | 25.6 | 3.5 | \# | 1953 | Landings | MT | 1 | 358.6 | 106.3 | 75.2 |
|  | 23.8 | 3.5 |  |  |  |  |  |  |  |  |
| 81.2 | 27.3 | 4.5 | \# | 1954 | Landings | MT | 46.8 | 316.9 | 88.4 | 81.2 |
|  | 25.6 | 4.5 |  |  |  |  |  |  |  |  |
| 87.2 | 29.1 | 5.5 | \# | 1955 | Landings | MT | 71.4 | 342.9 | 132.3 | 87.2 |
|  | 27.3 | 5.5 |  |  |  |  |  |  |  |  |
| 93.2 | 30.8 | 6.5 | \# | 1956 | Landings | MT | 46.2 | 332.3 | 83.4 | 93.2 |
|  | 29.1 | 6.5 |  |  |  |  |  |  |  |  |
| 99.2 | 32.6 | 7.5 | \# | 1957 | Landings | MT | 43.3 | 387.2 | 119.4 | 99.2 |
|  | 30.8 | 7.5 |  |  |  |  |  |  |  |  |
| 105.2 | 34.3 | 8.5 | \# | 1958 | Landings | MT | 124.4 | 369.6 | 165.3 | 105.2 |
|  | 32.6 | 8.5 |  |  |  |  |  |  |  |  |
| 111.2 | 36.1 | 9.5 | \# | 1959 | Landings | MT | 129 | 414.2 | 129.3 | 111.2 |
|  | 34.3 | 9.5 |  |  |  |  |  |  |  |  |
| 117.2 | 37.8 | 10.5 | \# | 1960 | Landings | MT | 82 | 389.3 | 220.1 | 117.2 |
|  | 36.1 | 10.5 |  |  |  |  |  |  |  |  |
| 123.2 | 39.6 | 11.5 | \# | 1961 | Landings | MT | 158.6 | 391.1 | 103 | 123.2 |
|  | 37.8 | 11.5 |  |  |  |  |  |  |  |  |
| 129.2 | 41.3 | 12.5 | \# | 1962 | Landings | MT | 127.3 | 305.3 | 272 | 129.2 |
|  | 39.6 | 12.5 |  |  |  |  |  |  |  |  |


| 135.2 | 43.1 | 13.5 | \# | 1963 | Landings | MT | 218.9 | 302.2 | 145.8 | 135.2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 41.3 | 13.5 |  |  |  |  |  |  |  |  |
| 141.2 | 44.8 | 14.5 | \# | 1964 | Landings | MT | 327.5 | 244.3 | 134.1 | 141.2 |
|  | 43.1 | 14.5 |  |  |  |  |  |  |  |  |
| 108 | 46.6 | 15.5 | \# | 1965 | Landings | MT | 185.1 | 135.7 | 128.9 | 108 |
|  | 44.8 | 15.5 |  |  |  |  |  |  |  |  |
| 186 | 48.3 | 16.5 | \# | 1966 | Landings | MT | 213 | 213.8 | 81.8 | 186 |
|  | 46.6 | 16.5 |  |  |  |  |  |  |  |  |
| 234 | 50.1 | 17.5 | \# | 1967 | Landings | MT | 64.6 | 144.8 | 70.5 | 234 |
|  | 48.3 | 17.5 |  |  |  |  |  |  |  |  |
| 122 | 51.8 | 18.5 | \# | 1968 | Landings | MT | 321.3 | 150.7 | 104 | 122 |
|  | 50.1 | 18.5 |  |  |  |  |  |  |  |  |
| 261 | 53.6 | 19.5 | \# | 1969 | Landings | MT | 96 | 94.2 | 89.8 | 261 |
|  | 51.8 | 19.5 |  |  |  |  |  |  |  |  |
| 303 | 55.3 | 20.5 | \# | 1970 | Landings | MT | 15 | 76 | 52.2 | 303 |
|  | 53.6 | 20.5 |  |  |  |  |  |  |  |  |
| 134.1 | 57.1 | 17.5 | \# | 1971 | Landings | MT | 347.2 | 36.8 | 32.7 | 134.1 |
|  | 55.3 | 17.5 |  |  |  |  |  |  |  |  |
| 116 | 58.8 | 29.3 | \# | 1972 | Landings | MT | 156 | 62.3 | 36.9 | 116 |
|  | 57.1 | 29.3 |  |  |  |  |  |  |  |  |
| 48 | 60.6 | 26.8 | \# | 1973 | Landings | MT | 75 | 60.6 | 51.2 | 48 |
|  | 58.8 | 26.8 |  |  |  |  |  |  |  |  |
| 75 | 62.3 | 51.2 | \# | 1974 | Landings | MT | 48 | 58.8 | 26.8 | 75 |
|  | 60.6 | 51.2 |  |  |  |  |  |  |  |  |
| 156 | 62.3 | 36.9 | \# | 1975 | Landings | MT | 116 | 57.1 | 29.3 | 156 |
|  | 62.3 | 36.9 |  |  |  |  |  |  |  |  |
| 347.2 | 36.8 | 32.7 | \# | 1976 | Landings | MT | 134.1 | 55.3 | 17.5 | 347.2 |
|  | 36.8 | 32.7 |  |  |  |  |  |  |  |  |
| 15 | 76 | 52.2 | \# | 1977 | Landings | MT | 303 | 53.6 | 20.5 | 15 |
|  | 76 | 52.2 |  |  |  |  |  |  |  |  |
| 96 | 94.2 | 89.8 | \# | 1978 | Landings | MT | 261 | 51.8 | 19.5 | 96 |
|  | 94.2 | 89.8 |  |  |  |  |  |  |  |  |
| 321.3 | 150.7 | 104 | \# | 1979 | Landings | MT | 122 | 50.1 | 18.5 | 321.3 |
|  | 150.7 | 104 |  |  |  |  |  |  |  |  |
| 64.6 | 144.8 | 70.5 | \# | 1980 | Landings | MT | 234 | 48.3 | 17.5 | 64.6 |
|  | 144.8 | 70.5 |  |  |  |  |  |  |  |  |
| 213 | 213.8 | 81.8 | \# | 1981 | Landings | MT | 186 | 46.6 | 16.5 | 213 |
|  | 213.8 | 81.8 |  |  |  |  |  |  |  |  |
| 185.1 | 135.7 | 128.9 | \# | 1982 | Landings | MT | 108 | 44.8 | 15.5 | 185.1 |
|  | 135.7 | 128.9 |  |  |  |  |  |  |  |  |
| 327.5 | 244.3 | 134.1 | \# | 1983 | Landings | MT | 141.2 | 43.1 | 14.5 | 327.5 |
|  | 244.3 | 134.1 |  |  |  |  |  |  |  |  |
| 218.9 | 302.2 | 145.8 | \# | 1984 | Landings | MT | 135.2 | 41.3 | 13.5 | 218.9 |
|  | 302.2 | 145.8 |  |  |  |  |  |  |  |  |
| 127.3 | 305.3 | 272 | \# | 1985 | Landings | MT | 129.2 | 39.6 | 12.5 | 127.3 |
|  | 305.3 | 272 |  |  |  |  |  |  |  |  |
| 158.6 | 391.1 | 103 | \# | 1986 | Landings | MT | 123.2 | 37.8 | 11.5 | 158.6 |
|  | 391.1 | 103 |  |  |  |  |  |  |  |  |
| 82 | 389.3 | 220.1 | \# | 1987 | Landings | MT | 117.2 | 36.1 | 10.5 | 82 |
|  | 389.3 | 220.1 |  |  |  |  |  |  |  |  |
| 129 | 414.2 | 129.3 | \# | 1988 | Landings | MT | 111.2 | 34.3 | 9.5 | 129 |
|  | 414.2 | 129.3 |  |  |  |  |  |  |  |  |
| 124.4 | 369.6 | 165.3 | \# | 1989 | Landings | MT | 105.2 | 32.6 | 8.5 | 124.4 |
|  | 369.6 | 165.3 |  |  |  |  |  |  |  |  |
| 43.3 | 387.2 | 119.4 | \# | 1990 | Landings | MT | 99.2 | 30.8 | 7.5 | 43.3 |
|  | 387.2 | 119.4 |  |  |  |  |  |  |  |  |
| 46.2 | 332.3 | 83.4 | \# | 1991 | Landings | MT | 93.2 | 29.1 | 6.5 | 46.2 |
|  | 332.3 | 83.4 |  |  |  |  |  |  |  |  |
| 71.4 | 342.9 | 132.3 | \# | 1992 | Landings | MT | 87.2 | 27.3 | 5.5 | 71.4 |
|  | 342.9 | 132.3 |  |  |  |  |  |  |  |  |
| 46.8 | 316.9 | 88.4 | \# | 1993 | Landings | MT | 81.2 | 25.6 | 4.5 | 46.8 |
|  | 316.9 | 88.4 |  |  |  |  |  |  |  |  |
| 1 | 358.6 | 106.3 | \# | 1994 | Landings | MT | 75.2 | 23.8 | 3.5 | 1 |
|  | 358.6 | 106.3 |  |  |  |  |  |  |  |  |
| 3.3 | 264.8 | 65.8 | \# | 1995 | Landings | MT | 69.2 | 22.1 | 2.5 | 3.3 |
|  | 264.8 | 65.8 |  |  |  |  |  |  |  |  |
| 0 | 264.2 | 8.6 | \# | 1996 | Landings | MT | 63.2 | 20.3 | 1.5 | 0 |
|  | 264.2 | 8.6 |  |  |  |  |  |  |  |  |
| 9 | 234.1 | 15 | \# | 1997 | Landings | MT | 57.2 | 18.6 | 0 | 9 |
|  | 234.1 | 15 |  |  |  |  |  |  |  |  |


| 73.1 | 259.4 | 4.8 | \# | 1998 | Landings | MT | 51.2 | 16.8 | 0 | 73.1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 259.4 | 4.8 |  |  |  |  |  |  |  |  |
| 0 | 221.6 | 4.3 | \# | 1999 | Landings | MT | 5.2 | 15.1 | 0 | 0 |
|  | 221.6 | 4.3 |  |  |  |  |  |  |  |  |
| 0 | 224.8 | 1.2 | \# | 2000 | Landings | MT | 39.2 | 13.3 | 0 | 0 |
|  | 224.8 | 1.2 |  |  |  |  |  |  |  |  |
| 0 | 188.7 | 1.1 | \# | 2001 | Landings | MT | 33.2 | 11.6 | 0 | 0 |
|  | 188.7 | 1.1 |  |  |  |  |  |  |  |  |
| 0.2 | 238.9 | 1.5 | \# | 2002 | Landings | MT | 27.2 | 9.8 | 0 | 0.2 |
|  | 232.9 | 1.5 |  |  |  |  |  |  |  |  |
| 0.1 | 237.1 | 0.2 | \# | 2003 | Landings | MT | 21.2 | 8.1 | 0 | 0.1 |
|  | 232.2 | 0.2 |  |  |  |  |  |  |  |  |
| 0.6 | 268 | 0.7 | \# | 2004 | Landings | MT | 15.2 | 6.3 | 0 | 0.6 |
|  | 257.6 | 0.7 |  |  |  |  |  |  |  |  |
| 0 | 331.7 | 0.9 | \# | 2005 | Landings | MT | 9.2 | 4.6 | 0 | 0 |
|  | 315.5 | 0.9 |  |  |  |  |  |  |  |  |
| 1.4 | 321.5 | 1.2 | \# | 2006 | Landings | MT | 3.2 | 2.8 | $\bigcirc$ | 1.4 |
|  | 312 | 1.2 |  |  |  |  |  |  |  |  |
| \# | 1.4 | 312 | 0.4 |  |  |  |  |  |  |  |
| \# | Survey |  |  |  |  |  |  |  |  |  |
| \#CPUE | rom_Are | _2_R | Means |  |  |  |  |  |  |  |
| \#Year | Season | Type | Value | $\ln (1$ |  |  |  |  |  |  |
| 28 |  |  |  |  |  |  |  |  |  |  |
| 1990 | 1 | 4 | 5.73 | 0.728 | 59186 |  |  |  |  |  |
| 1991 | 1 | 4 | 5.426 | 0.703 | 59282 |  |  |  |  |  |
| 1992 | 1 | 4 | 4.768 | 0.695 | 33036 |  |  |  |  |  |
| 1993 | 1 | 4 | 4.242 | 0.759 | 57379 |  |  |  |  |  |
| 1994 | 1 | 4 | 4.426 | 0.740 | 46527 |  |  |  |  |  |
| 1995 | 1 | 4 | 4.069 | 0.705 | 9139 |  |  |  |  |  |
| \#1996 | 1 | 4 | 4.569 | 0.646 | 20543 |  |  |  |  |  |
| \#1997 | 1 | 4 | 3.932 | 0.699 | 88754 |  |  |  |  |  |
| \#1998 | 1 | 4 | 4.805 | 0.622 | 19705 |  |  |  |  |  |
| \#1999 | 1 | 4 | 4.856 | 0.620 | 1093 |  |  |  |  |  |
| \#2000 | 1 | 4 | 5.028 | 0.604 | 28452 |  |  |  |  |  |
| \#2001 | 1 | 4 | 4.288 | 0.673 | 16624 |  |  |  |  |  |
| \#2002 | 1 | 4 | 5.01 | 0.607 | 0313 |  |  |  |  |  |
| \#2003 | 1 | 4 | 4.946 | 0.607 | 24035 |  |  |  |  |  |
| \#2004 | 1 | 4 | 5.571 | 0.5531 | 2333 |  |  |  |  |  |
| \#2005 | 1 | 4 | 5.355 | 0.562 | 3981 |  |  |  |  |  |
| \#2006 | 1 | 4 | 5.201 | 0.586 | 1481 |  |  |  |  |  |
| \#Tag_A | undance | from | ea_2_R | , Mean |  |  |  |  |  |  |
| \#Year | Season | Type | Value | $\ln (1+$ |  |  |  |  |  |  |
| 2000 | 1 | 5 | 1389 | 0.085 |  |  |  |  |  |  |
| 2001 | 1 | 5 | 2997 | 0.115 |  |  |  |  |  |  |
| 2002 | 1 | 5 | 1944 | 0.080 |  |  |  |  |  |  |
| 2003 | 1 | 5 | 2119 | 0.062 |  |  |  |  |  |  |
| 2004 | 1 | 5 | 2996 | 0.110 |  |  |  |  |  |  |
| 2005 | 1 | 5 | 5015 | 0.127 |  |  |  |  |  |  |
| 2006 | 1 | 5 | 3464 | 0.08 |  |  |  |  |  |  |
| \#TagCP | E_from | Area_2 | aw_Mea |  |  |  |  |  |  |  |
| \#Year | Season | Type | Value | $\ln (1+$ |  |  |  |  |  |  |
| 1981 | 1 | 6 | 4.75 | 0.666 |  |  |  |  |  |  |
| 1986 | 1 | 6 | 2.337 | 0.599 |  |  |  |  |  |  |
| 1987 | 1 | 6 | 1.172 | 0.634 |  |  |  |  |  |  |
| 1988 | 1 | 6 | 0.826 | 0.553 |  |  |  |  |  |  |
| 1989 | 1 | 6 | 1.236 | 0.977 |  |  |  |  |  |  |
| 1990 | 1 | 6 | 0.991 | 0.843 |  |  |  |  |  |  |
| 1998 | 1 | 6 | 2.46 | 0.813 |  |  |  |  |  |  |
| 1999 | 1 | 6 | 3.061 | 0.740 |  |  |  |  |  |  |
| 2000 | 1 | 6 | 2.203 | 0.568 |  |  |  |  |  |  |
| 2001 | 1 | 6 | 4.657 | 0.607 |  |  |  |  |  |  |
| 2002 | 1 | 6 | 5.486 | 0.503 |  |  |  |  |  |  |
| 2003 | 1 | 6 | 6.245 | 0.591 |  |  |  |  |  |  |


| 2004 | 1 | 6 | 9.414 | 0.5149 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 1 | 6 | 10.192 | 0.7579 |  |  |  |  |  |  |  |
| 2006 | 1 | 6 | 10.543 | 0.4205 |  |  |  |  |  |  |  |
| \# | Discards |  |  |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |
| 2 |  |  |  |  |  |  |  |  |  |  |  |
| -1 |  |  |  |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |
| \# | Mean | Body | Weight |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |
| 0 |  |  |  |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |
| \# | Composi | tion | Conditi | oners |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |
| 0.0001 |  |  |  |  |  |  |  |  |  |  |  |
| 0.0001 |  |  |  |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |
| \# | Length | Composi | tion |  |  |  |  |  |  |  |  |
| , |  |  |  |  |  |  |  |  |  |  |  |
| \#Yr | Seas | Flt/Svy | Gender | Part | Nsamp | datavec | or (fem | le-male |  |  |  |
| 19 le |  |  |  |  |  |  |  |  |  |  |  |
| 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 |
|  | 44 | 46 | 48 | 50 | 52 | 54 | 56 |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |
| 67 |  |  |  |  |  |  |  |  |  |  |  |
| 1976 | 1 | 1 | 3 | 0 | 28 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.0051 | 0.0102 | 0.0332 | 0.0652 | 0.0985 | 0.0806 | 0.0793 | 0.0678 |
|  | 0.0358 | 0.0396 | 0 | 0 | 0 | 0 | 0 | 0.0013 | 0 | $\bigcirc$ | 0 |
|  | 0.0026 | 0.0307 | 0.0729 | 0.1036 | 0.1189 | 0.0844 | 0.0473 | 0.0179 | 0.0026 | 0.0026 |  |
| 1980 | 1 | 1 | 3 | 0 | 14 | $\bigcirc$ | 0 | 0 | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ |
|  | 0 | 0.0049 | 0.0049 | 0.0388 | 0.0194 | 0.034 | 0.0485 | 0.0631 | 0.0825 | 0.0194 | 0.0146 |
|  | 0.0097 | 0.0097 | 0 | 0 | 0 | 0 | 0 | 0 | $\bigcirc$ | $\bigcirc$ | 0.0049 |
|  | 0.0243 | 0.034 | 0.0922 | 0.1165 | 0.1359 | 0.1165 | 0.0631 | 0.034 | 0.0243 | 0.0049 |  |
| 1981 | 1 | 1 | 3 | 0 | 28 | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 |
|  | 0.0025 | 0 | 0.0025 | 0.0075 | 0.0225 | 0.035 | 0.0625 | 0.08 | 0.0875 | 0.055 | 0.0475 |
|  | 0.0175 | 0.015 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0025 | 0.005 |
|  | 0.0175 | 0.0575 | 0.05 | 0.1 | 0.12 | 0.1 | 0.0625 | 0.035 | 0.01 | 0.005 |  |
| 1982 | 1 | 1 | 3 | 0 | 28 | $\bigcirc$ | 0 | 0 | 0 | $\bigcirc$ | $\bigcirc$ |
|  | 0 | 0.0025 | 0.01 | 0.025 | 0.0325 | 0.075 | 0.08 | 0.0625 | 0.0575 | 0.0225 | 0.025 |
|  | 0.0025 | 0.005 | 0 | 0 | 0 | 0 | 0 | 0 | 0.005 | 0.0025 | 0.01 |
|  | 0.0275 | 0.0875 | 0.1475 | 0.1175 | 0.1 | 0.0625 | 0.0275 | 0.0125 | $\bigcirc$ | 0 |  |
| 1983 | 1 | 1 | 3 | 0 | 56 | $\bigcirc$ | 0 | 0 | 0 | $\bigcirc$ | 0 |
|  | 0 | 0 | 0.0063 | 0.015 | 0.0363 | 0.0625 | 0.0788 | 0.0675 | 0.0738 | 0.05 | 0.0388 |
|  | 0.01 | 0.0038 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0025 | 0.0025 | 0.0075 |
|  | 0.0163 | 0.07 | 0.1138 | 0.13 | 0.0963 | 0.0763 | 0.0338 | 0.0075 | $\bigcirc$ | 0.0013 |  |
| 1984 | 1 | 1 | 3 | 0 | 21 | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 |
|  | 0 | 0 | 0.0033 | 0.02 | 0.0367 | 0.0367 | 0.0733 | 0.1 | 0.0633 | 0.0333 | 0.02 |
|  | 0.0167 | 0 | $0$ | 0 | 0 | 0 | 0 | 0 | $0$ | 0.0133 | 0.02 |
|  | 0.0433 | 0.0767 | 0.12 | 0.1267 | 0.09 | 0.0567 | 0.0333 | 0.01 | 0.0067 | 0 |  |
| 1986 | 1 | 1 | 3 | 0 | 57 | $\bigcirc$ | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0.0031 | 0.0062 | 0.0093 | 0.059 | 0.0497 | 0.087 | 0.059 | 0.0932 | 0.0994 | 0.0559 |
|  | 0.0093 | 0.0155 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0031 | 0.0124 |
|  | 0.0217 | 0.0683 | 0.0963 | 0.1056 | 0.0994 | 0.0311 | 0.0124 | 0.0031 | 0 | 0 |  |
| 1987 | 1 | 1 | 3 | 0 | 71 | $\bigcirc$ | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.0025 | 0.01 | 0.0299 | 0.0623 | 0.0823 | 0.1471 | 0.0873 | 0.0599 |
|  | 0.0399 | 0.0299 | 0 | 0 | $\bigcirc$ | 0 | 0 | 0 | 0 | 0 | 0.0025 |
|  | 0.005 | 0.0249 | 0.0698 | 0.1446 | 0.1172 | 0.0499 | 0.0224 | 0.0125 | 0 | 0 |  |
| 1988 | 1 | 1 | 3 | 0 | 18 | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 |
|  | 0 | 0 | 0 | 0 | 0.02 | 0.02 | 0.08 | 0.1 | 0.12 | 0.12 | 0.03 |
|  | 0.01 | 0.01 | 0 | 0 | $\bigcirc$ | 0 | 0.01 | 0 | 0 | 0 | 0.01 |
|  | 0 | 0.01 | 0.03 | 0.08 | 0.19 | 0.09 | 0.03 | 0.03 | 0 | 0.01 |  |
| 1989 | 1 | 1 | 3 | 0 | 40 | $\bigcirc$ | 0 | 0 | 0 | $\bigcirc$ | 0 |
|  | 0 | 0 | 0 | 0.0044 | 0 | 0.0311 | 0.0489 | 0.08 | 0.1067 | 0.12 | 0.1022 |
|  | 0.0889 | 0.0578 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0044 | 0 |
|  | 0.0044 | 0.0178 | 0.0222 | 0.0533 | 0.0889 | 0.0978 | 0.04 | 0.0178 | 0.0044 | 0.0089 |  |
| 1990 | 1 | 1 | 3 | 0 | 44 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.004 | 0.008 | 0.0241 | 0.0402 | 0.0281 | 0.0723 | 0.0723 | 0.0803 | 0.0562 |
|  | 0.0482 | 0.0281 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0161 |
|  | 0.0201 | 0.0442 | 0.0683 | 0.0803 | 0.1044 | 0.1365 | 0.0522 | 0.0161 | 0 | $\bigcirc$ |  |
| 1991 | 1 | 1 | 3 | 0 | 54 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0.0033 | 0.0066 | 0.0066 | 0.0132 | 0.0728 | 0.0695 | 0.0861 | 0.0695 | 0.0662 | 0.0232 |


|  | 0.0464 | 0.0166 | 0 | 0 | 0 | 0 | $\bigcirc$ | $\bigcirc$ | 0 | 0.0033 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.0099 | 0.0397 | 0.0695 | 0.0662 | 0.1026 | 0.0795 | 0.0828 | 0.0464 | 0.0199 | 0 |  |
| 1992 | 1 | 1 | 3 | 0 | 36 | 0 | 0 | $\bigcirc$ | 0 | 0 | $\bigcirc$ |
|  | $\bigcirc$ | 0 | 0 | 0.01 | 0.025 | 0.055 | 0.06 | 0.085 | 0.09 | 0.075 | 0.035 |
|  | 0.03 | 0.03 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.005 | 0.01 |
|  | 0.01 | 0.065 | 0.065 | 0.16 | 0.085 | 0.07 | 0.035 | 0 | 0 | $\bigcirc$ |  |
| 1993 | 1 | 1 | 3 | 0 | 22 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.008 | 0.024 | 0.048 | 0.032 | 0.064 | 0.056 | 0.064 | 0.08 |
|  | 0.048 | 0.008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0.056 | 0.024 | 0.072 | 0.088 | 0.128 | 0.104 | 0.08 | 0.008 | 0.008 | 0 |  |
| \#1994 | 1 | 1 | 3 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.0612 | 0 | 0 | 0.0612 | 0.1224 | 0 | 0.0408 | 0.0612 | 0.0204 |
|  | 0.0408 | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 | 0 | 0 | $\bigcirc$ |
|  | $\bigcirc$ | 0.0408 | 0.102 | 0.102 | 0.1224 | 0.102 | 0.102 | 0.0204 | 0 | 0 |  |
| \#1995 | 1 | 1 | 3 | 0 | 9 | $\bigcirc$ | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.04 | 0 | 0.06 | 0.02 | 0.04 | 0.02 | 0.04 | 0 |
|  | $\bigcirc$ | 0.02 | 0 | 0 | 0 | $\bigcirc$ | 0.02 | 0.02 | 0 | $\bigcirc$ | 0 |
|  | 0.02 | 0.02 | 0.06 | 0.2 | 0.14 | 0.12 | 0.1 | 0.04 | 0.02 | 0 |  |
| \#1998 | 1 | 1 | 3 | 0 | 14 | 0 | $\bigcirc$ | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.0235 | 0.0235 | 0.0941 | 0.1059 | 0.0941 | 0.0588 | 0.0118 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 | 0.0353 |
|  | 0.0235 | 0.1294 | 0.1765 | 0.1412 | 0.0588 | 0.0235 | 0 | 0 | 0 | 0 |  |
| \#2002 | 1 | 1 | 3 | 0 | 8 | 0 | $\bigcirc$ | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0.14 | 0.16 | 0.18 | 0.08 |
|  | 0.02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\bigcirc$ |
|  | 0.02 | 0 | 0.04 | 0.12 | 0.16 | 0.06 | $\bigcirc$ | 0 | 0 | 0 |  |
| \#1980 | 1 | 2 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0.002 | 0.0039 | 0.0236 | 0.0394 | 0.0571 | 0.0886 | 0.1575 | 0.1594 | 0.25 | 0.1496 | 0.0413 |
|  | 0.0276 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.0039 | 0.0236 |
|  | 0.0394 | 0.0571 | 0.0886 | 0.1575 | 0.1594 | 0.25 | 0.1496 | 0.0413 | 0.0276 | 0 |  |
| 1980 | 1 | 2 | 3 | 0 | 10 | 0 | 0 | 0 | 0 | 0 | 0.0014 |
|  | $\bigcirc$ | 0.0071 | 0.0085 | 0.0299 | 0.0341 | 0.0327 | 0.0284 | 0.0356 | 0.0512 | 0.027 | 0.0199 |
|  | 0.0028 | 0.0014 | 0 | 0 | 0 | 0 | 0 | 0.0014 | 0 | 0.0057 | 0.0284 |
|  | 0.0341 | 0.0612 | 0.0512 | 0.1124 | 0.1607 | 0.1579 | 0.0811 | 0.0256 | 0 | 0 |  |
| 1981 | 1 | 2 | 3 | 0 | 30 | 0 | 0 | 0 | 0.0103 | 0.0103 | 0 |
|  | 0.0206 | 0.0206 | 0 | 0.0206 | 0.0103 | 0 | 0.1031 | 0.0515 | 0.0619 | 0.0103 | 0 |
|  | $\bigcirc$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0206 | 0.0309 | 0.0309 |
|  | 0.0309 | 0.0722 | 0.1134 | 0.1134 | 0.1443 | 0.0619 | 0.0412 | 0.0206 | 0 | 0 |  |
| \#1982 | 1 | 2 | 0 | 0 | 22 | 0.0088 | 0.0109 | 0.0044 | 0.0022 | 0.0073 | 0.008 |
|  | 0.0146 | 0.0248 | 0.0518 | 0.0904 | 0.1196 | 0.1276 | 0.1488 | 0.1145 | 0.1371 | 0.0795 | 0.0343 |
|  | 0.0131 | 0.0022 | 0.0088 | 0.0109 | 0.0044 | 0.0022 | 0.0073 | 0.008 | 0.0146 | 0.0248 | 0.0518 |
|  | 0.0904 | 0.1196 | 0.1276 | 0.1488 | 0.1145 | 0.1371 | 0.0795 | 0.0343 | 0.0131 | 0.0022 |  |
| 1982 | 1 | 2 | 3 | 0 | 12 | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 |
|  | 0 | 0.0177 | 0.0177 | 0.0442 | 0.0796 | 0.0619 | 0.0796 | 0.0531 | 0.0265 | 0.0088 | 0 |
|  | $\bigcirc$ | $\bigcirc$ | 0 | 0 | 0 | $\bigcirc$ | $\bigcirc$ | 0 | 0.0088 | 0.0088 | 0.0354 |
|  | 0.115 | 0.0973 | 0.0619 | 0.115 | 0.0973 | 0.0531 | 0.0177 | 0 | 0 | $\bigcirc$ |  |
| \#1983 | 1 | 2 | 0 | 0 | 1 | $\bigcirc$ | 0 | 0 | 0 | 0 | 0.04 |
|  | 0.0733 | 0.12 | 0.1133 | 0.1733 | 0.1733 | 0.12 | 0.08 | 0.0467 | 0.0133 | 0.0267 | 0.02 |
|  | $\bigcirc$ | 0 | 0 | 0 | 0 | 0 | 0 | 0.04 | 0.0733 | 0.12 | 0.1133 |
|  | 0.1733 | 0.1733 | 0.12 | 0.08 | 0.0467 | 0.0133 | 0.0267 | 0.02 | 0 | 0 |  |
| \#1984 | 1 | 2 | 0 | 0 | 2 | 0 | 0 | 0 | 0.1 | 0 | 0 |
|  | 0.2 | 0.1 | 0.2 | $\bigcirc$ | 0 | 0.4 | 0 | $\bigcirc$ | 0 | $\bigcirc$ | 0 |
|  | $\bigcirc$ | 0 | 0 | 0 | 0 | 0.1 | 0 | 0 | 0.2 | 0.1 | 0.2 |
|  | $\bigcirc$ | 0 | 0.4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1984 | 1 | 2 | 3 | 0 | 25 | 0 | 0 | 0 | 0.0029 | 0.0014 | 0.0115 |
|  | 0.033 | 0.0445 | 0.0603 | 0.0848 | 0.0905 | 0.0876 | 0.0647 | 0.0546 | 0.0417 | 0.0158 | 0.0129 |
|  | 0.0043 | 0.0057 | 0.0014 | 0 | 0 | 0 | 0.0014 | 0.0029 | 0.0029 | 0.0172 | 0.0302 |
|  | 0.0388 | 0.0704 | 0.0718 | 0.0503 | 0.0431 | 0.0201 | 0.023 | 0.0086 | 0.0014 | 0 |  |
| \#1985 | $1$ | $2$ | 0 | 0 |  | 0.1884 | 0.0072 | $\bigcirc$ | 0.0072 | $0.0217$ | $0.0435$ |
|  | 0.1594 | 0.1377 | 0.2391 | 0.1014 | 0.058 | 0.0145 | 0.0072 | 0 | 0.0072 | $0$ | $0$ |
|  | 0 | 0.0072 | 0.1884 | 0.0072 | 0 | 0.0072 | 0.0217 | 0.0435 | 0.1594 | 0.1377 | 0.2391 |
|  | 0.1014 | 0.058 | 0.0145 | 0.0072 | 0 | 0.0072 | 0 | 0 | 0 | 0.0072 |  |
| 1985 | 1 | 2 | 3 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0.0127 | 0.019 | 0.0316 | 0.0633 | 0.0759 | 0.0443 | 0.0759 | 0.0633 | 0.0443 | 0.0127 | 0.019 |
|  | 0.0063 | 0.0063 | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 | 0.0063 | 0.038 |
|  | 0.0759 | 0.0633 | 0.0759 | 0.0886 | 0.1076 | 0.038 | 0.019 | 0.0127 | 0 | $\bigcirc$ |  |
| 1986 | 1 | 2 | 3 | 0 | 17 | 0 | 0 | 0 | 0 | 0.002 | 0.0039 |
|  | 0.0098 | 0.0117 | 0.0391 | 0.0469 | 0.0723 | 0.0801 | 0.0938 | 0.0488 | 0.0313 | 0.0234 | 0.0098 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0039 | 0.0117 | 0.0293 | 0.0547 |
|  | 0.0957 | 0.0957 | 0.1055 | 0.0566 | 0.043 | 0.0176 | 0.0078 | 0.0059 | 0 | 0 |  |


| 1987 | 1 | 2 | 3 | 0 | 21 | 0 | 0 | 0 | 0 |  | 0.0047 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.014 | 0.0171 | 0.029 | 0.048 | 0.05 | 0.069 | 0.08 | 0.0713 | 0.0543 | 0.034 | 0.0202 |
|  | 0.0124 | 0.0016 | 0 | 0 | 0 |  |  | 0.014 | 0.014 | 0.015 | 0.0357 |
|  | 0.0326 | 0.0651 | 0.0 | 0.07 | 0.0 | 0.06 | 0.0 | 0.0124 | 0.0031 | 0 |  |
| 1988 | 1 | 2 | 3 | 0 |  | 0 | 0 | 0 | 0 | 0.00 | 0.0177 |
|  | 0.0 | 0.0288 | 0.0 | 0.06 | 0.05 | 0.0909 | 0.0909 | 0.057 | 0.0443 | 0.0244 | 0.0089 |
|  | 0 | 0.0022 | 0 | 0 | 0 | 0.0022 | 0.0022 | 0.006 | 0.0222 | 0.011 | 0.0377 |
|  | 0.06 | 0.0798 | 0. | 0.04 | 0.0 | 0.0288 | 0.0222 | 0.0022 | 0.0022 | 0 |  |
| 89 | 1 | 2 | 3 | 0 | 13 | 0 | 0 | 0 | 0 | 0 | 0.00 |
|  | 0. | 0.0 | 0. | 0.05 | 0.08 | 0.09 | 0.09 | 0.0705 | 0.0605 | 0.04 | 0.00 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0025 | 0.0101 | 0.03 | 0.0 |
|  | 0. | 0. | 0. | 0.04 | 0.0 | 0.0 | 0. |  | 0 |  |  |
| 990 | 1 | 2 | 3 | 0 | 9 | 0 | 0 | 0 | 0 | 0.0 | 0.0034 |
|  | 0.0 | 0.01 | 0.0 | 0.08 | 0.12 | 0.0 | 0.0966 | 0.0 | 0.0 |  | 0.0034 |
|  | 0 | 0 | 0 | 0 | 0.003 | 0 | 0.0034 | 0 | 0 | 0.0 | 0.0483 |
|  | 0.0 | 0.089 | 0.0 | 0.05 | 0.034 | 0.0 | 0.0276 | 0 | 0 | 0 |  |
| 1991 | 1 | 2 | 3 | 0 | 22 | 0 | 0 |  | 0 | 0.01 | 粏 |
|  | 0.018 | 0.0403 | 0.0 | 0.0 | 0.07 | 0.0903 | 0.061 | 0.061 | 0.027 | 0.019 | 0.0056 |
|  | 0.0028 | 0.0014 | 0 | 0 | 0 | 0.0014 | 0.0056 | 0.0125 | 0.0153 | 0.030 | 0.0542 |
|  | 0.0708 | 0.0833 | 0.05 | 0.055 | 0.02 | 0.0236 | 0.0097 | 0.0014 | 0.0014 | 0 |  |
| 1992 | 1 | 2 | 3 | ${ }^{0}$ | 29 | 0 | 0 | 0.0011 | 0.0023 | 0.00 | 0.0091 |
|  | 0.013 | 0.027 | 0.039 | 0.061 | 0.096 | 0.0942 | 0.0658 | 0.044 | 0.0409 | 0.012 | 0.0045 |
|  | 0.0023 | 0.0011 | 0 | 0 | 0 | 0.0011 | 0.0045 | 0.0102 | 0.0159 | 0.02 | 0.0522 |
|  | 0.059 | 0.0942 | 0. | 0.06 | 0.0 | 0.0193 | 0.0125 | 0.0057 | 0.0011 |  |  |
| 1993 | 1 | 2 | 3 | 0 |  | 0 | 0 | 0 | 0 | 0.0023 | 0.0163 |
|  | 0.019 | 0.0431 | 0.0 | 0.07 | 0.09 | 0.0 | 0.0 | 0.0536 | 0.03 | 0.0198 | 0.0081 |
|  | 0.008 | 0.0012 | 0 | 0 | 0 |  | 0 | 0.0116 | 0.021 | 0.0442 | 0.0547 |
|  | 0.066 | 0.071 | 0. | 0. | 0.0 | 0. | 0. | 0.0035 | 0. |  |  |
| 1994 | 1 | 2 | 3 | 0 | 28 | 0 | 0 | 0 | 0 | 0.00 | 035 |
|  | 0.020 | 0. | 0. | 0. | 0.0 | 0. | 0.0567 | 0.03 | 0.0197 | 0.030 | 0.0116 |
|  | 0.004 | 0 | 0 | 0 | 0 | 0 | 0.0012 | 0.0093 | 0.0162 | 0.05 |  |
|  | 0.075 | 0.072 | 0.0 | 0. | 0. | 0.0 | 0.0116 | 0.0058 | 0 | 0 |  |
| 1995 | 1 | 2 | 3 | 0 |  |  | 0 |  | 0 | 0.001 | 037 |
|  | 0.00 | 0.03 | 0.06 | 0.09 | 0.09 | 0.0 | 0.0702 | 0.03 | 0.0222 | 0.016 | 0.0062 |
|  | 0.003 | 0 | 0 | ${ }^{0}$ | 0 | 0 | 0.0012 | 0.004 | 0.0185 | 0.035 | 0.0776 |
|  | 0.093 | 0.0 | 0. | 0.0 | 0.02 | 0.00 | 0.0025 | 0.001 | 0.0025 | 0 |  |
| 1996 | 1 | 2 | 3 | 0 | 27 | 0 | 0 | 0 | 0.0012 | 0.001 | 0.006 |
|  | 0.0 | 0.03 | 0.07 | 0.10 | 0.10 | 0.06 | 0.047 | 0.0277 | 0.0181 | 0.00 | 0.0012 |
|  | － |  |  | 0 |  | 0.0024 | 0.002 | 0.0024 | 0.020 | 0.0507 | 0.0881 |
|  | 0.092 | 0.102 | 0.0 | 0.024 | 0.02 | 0.0084 | 0 |  | 0 | 0.0012 |  |
| 1997 | 1 | 2 | 3 | 0 | 29 | 0 | 0 | 0 | 0 | 0 | ． 0022 |
|  | 0.01 | 0.04 | 0.06 | 0.0 | 0.08 | 0.0756 | 0.0533 | 0.0344 | 0.02 | 0.02 | 0.0056 |
|  | 0.004 | 0.002 | 0 | 0 | 0 | 0.0033 | 0.0022 | 0.004 | 0.023 | 0.03 | 0.0678 |
|  | 0.088 | 0.11 | 0. | 0.0 | 0.0 | 0.0122 | 0.008 | 0.002 | 0 | 0 |  |
| 1998 | 1 | 2 | 3 | 0 | 40 | 0 | 0 | 0 | 0 | 0.00 | ． 0106 |
|  | 0.02 | 0.0355 | 0.05 | 0.08 | 0.08 | 0.0 | 0.0544 | 0.028 | 0.0144 | 0.013 | 0.0008 |
|  | 0 | 0.0008 | 0 | 0 | 0 | 0 | 0.0023 | 0.0136 | 0.0333 | 0.04 |  |
|  | 0.09 | 0.0998 | 0.0 | 0.05 | 0.02 | 0.01 | 0.0023 | 0.0008 | 0.0008 |  |  |
| 99 | 1 | 2 | 3 | 0 | 44 |  | 0 | 0.0006 | 0.0012 | 0.004 | ． 0126 |
|  | 0.0263 | 0.04 | 0.05 | 0.081 | 0.08 | 0.0699 | 0.0562 | 0.0359 | 0.0132 | 0.003 | 0.0012 |
|  | 0.0012 | 0.0006 | 0 | 0 | 0 | 0.0006 | 0.0012 | 0.0185 | 0.0281 | 0.0466 | 0.0831 |
|  | 0.095 | 0.0932 | 0.06 | 0.0 | 0.01 | 0.0048 | 0.003 | 0.0024 | 0.0012 | 0.0006 |  |
| 2000 | 1 | 2 | 3 | 0 |  |  | 0 |  | 0.0012 | 0.0042 | 0.0079 |
|  | 0.0176 | 0.038 | 0.05 | 0.08 | 0.10 | 0.08 | 0.0812 | 0.0303 | 0.0212 | 0.003 | 0.0018 |
|  | 0.001 | 0 |  | 0 | 0 |  | 0.0036 | 0.0091 | 0.0188 | 0.047 | 0.0661 |
|  | 0.08 | 0.09 | 0.07 | 0.0 | 0.0 | 0.0 | 0.00 | 0 | 0.0 | 0 |  |
| 2001 | 1 | 2 | 3 | 0 | 47 | 0 | 0 | 0 | 0 | 0 | 0．0028 |
|  | 0.013 | 0.035 | 0.05 | 0.09 | 0.11 | 0.10 | 0.0614 | 0.0271 | 0.018 | 0.009 | 0.0023 |
|  | 0.0017 | 0 | 0 | 0 | 0 | 0 | 0.0011 | 0.0056 | 0.0152 | 0.0355 | 0.071 |
|  | 0.1071 | 0.089 | 0.0 | 0.0 | 0.01 | 0.00 | 0 | 0.0023 | 0.0011 | 0.0006 |  |
| 20 | 1 | 2 | 3 | 0 | 72 | 0 | 0 | 0.0005 | 0.0005 | 0.0005 |  |
|  | 0.0049 | 0.0168 | 0.0 | 0.08 | 0.102 | 0.1141 | 0.0984 | 0.0606 | 0.0292 | 0.0059 | 0.0027 |
|  | 0.0011 | 0.0005 | 0 | 0 | 0 | 0.0011 | 0.0016 | 0.0027 | 0.0087 | 0.0276 | 0.0573 |
|  | 0.0957 | 0.1001 | 0.0757 | 0.0335 | 0.013 | 0.0043 | 0.0016 | 0.0005 | 0 | 0 |  |
| 200 | 1 | 2 | 0 |  | 9 | 0 | 0 |  | 0.0013 | 0.0077 | 0.0103 |
|  | 0.0333 | 0.0833 | 0.1308 | 0.1487 | 0.2064 | 0.1603 | 0.1038 | 0.0641 | 0.0282 | 0.0128 | 0.0051 |
|  | 0.0038 | 0 | 0 | 0 | 0 | 0.0013 | 0.0077 | 0.0103 | 0.0333 | 0.083 | 308 |
|  | 0.1487 | 0.2064 | 0.1603 | 0.1038 | 0.06 | 0.0282 | 0.0128 | 0.0051 | 0.0038 |  |  |
| 03 | 1 | 2 | 3 | 0 | 70 | 0 | 0 | 0 | 0 | 0.0005 |  |
|  | 0.005 |  |  |  |  | 0.12 |  |  |  | 0.0151 |  |



| 1992 | 1 | 3 | 3 | $\bigcirc$ | 49 | 0 | 0 | 0 | 0 | 0.011 | 0.011 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.0037 | 0.0733 | 0.0989 | 0.0696 | 0.0989 | 0.0586 | 0.0586 | 0.0366 | 0.0293 | 0.0073 | 0.0073 |
|  | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 | 0.0073 | 0.011 | 0.022 | 0.044 | 0.0513 |
|  | 0.0733 | 0.0513 | 0.0659 | 0.0696 | 0.0256 | 0.011 | $\bigcirc$ | 0.0037 | 0 | 0 |  |
| 1993 | 1 | 3 | 3 | 0 | 58 | 0 | 0 | 0 | 0.0031 | 0.0031 | 0.0123 |
|  | 0.0031 | 0.0617 | 0.0895 | 0.0772 | 0.0802 | 0.0525 | 0.071 | 0.0432 | 0.0247 | 0.0216 | 0 |
|  | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 | 0.0031 | 0.0031 | 0.0123 | 0.0494 | 0.0494 |
|  | 0.0648 | 0.0864 | 0.071 | 0.0772 | 0.037 | 0.0031 | $\bigcirc$ | 0 | 0 | 0 |  |
| 1994 | 1 | 3 | 3 | 0 | 44 | 0 | 0 | 0 | 0.004 | 0.012 | 0.044 |
|  | 0.068 | 0.068 | 0.068 | 0.08 | 0.08 | 0.048 | 0.04 | 0.008 | 0.008 | 0.004 | 0 |
|  | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 | 0.016 | 0.016 | 0.032 | 0.052 | 0.072 |
|  | 0.056 | 0.072 | 0.068 | 0.052 | 0.012 | 0.02 | 0 | 0 | 0 | 0 |  |
| 1995 | 1 | 3 | 3 | 0 | 40 | 0 | 0 | 0 | 0 | 0 | 0.0089 |
|  | 0.0268 | 0.0446 | 0.067 | 0.0714 | 0.0625 | 0.0759 | 0.0893 | 0.0268 | 0.0134 | 0.0045 | 0.0089 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0.0089 | 0.0357 | 0.0268 | 0.0357 | 0.0759 |
|  | 0.0893 | 0.0804 | 0.058 | 0.0536 | 0.0268 | 0.0089 | 0 | 0 | 0 | 0 |  |
| 1981 | 1 | 6 | 0 | 0 | 29 | $\bigcirc$ | $\bigcirc$ | 0.0004 | 0.0006 | 0.0006 | 0.0045 |
|  | 0.0159 | 0.0416 | 0.0855 | 0.1166 | 0.1255 | 0.1247 | 0.137 | 0.1569 | 0.117 | 0.0492 | 0.0142 |
|  | 0.0023 | 0.0004 | 0 | $\bigcirc$ | 0.0004 | 0.0006 | 0.0006 | 0.0045 | 0.0159 | 0.0416 | 0.0855 |
|  | 0.1166 | 0.1255 | 0.1247 | 0.137 | 0.1569 | 0.117 | 0.0492 | 0.0142 | 0.0023 | 0.0004 |  |
| 1982 | 1 | 6 | 0 | $\bigcirc$ | 24 | 0 | $\bigcirc$ | 0.0024 | 0.0024 | 0.0044 | 0.0047 |
|  | 0.0142 | 0.0305 | 0.0665 | 0.1322 | 0.1535 | 0.1516 | 0.1504 | 0.1417 | 0.0926 | 0.0372 | 0.0115 |
|  | 0.0008 | 0.0004 | 0 | 0 | 0.0024 | 0.0024 | 0.0044 | 0.0047 | 0.0142 | 0.0305 | 0.0665 |
|  | 0.1322 | 0.1535 | 0.1516 | 0.1504 | 0.1417 | 0.0926 | 0.0372 | 0.0115 | 0.0008 | 0.0004 |  |
| 1983 | 1 | 6 | 0 | 0 | 29 | 0 | $\bigcirc$ | 0.0005 | 0.0015 | 0.0045 | 0.0198 |
|  | 0.06 | 0.1011 | 0.1269 | 0.1477 | 0.1472 | 0.1487 | 0.1259 | 0.0962 | 0.0407 | 0.0183 | 0.004 |
|  | 0.0005 | 0 | 0 | 0 | 0.0005 | 0.0015 | 0.0045 | 0.0198 | 0.06 | 0.1011 | 0.1269 |
|  | 0.1477 | 0.1472 | 0.1487 | 0.1259 | 0.0962 | 0.0407 | 0.0183 | 0.004 | 0.0005 | 0 |  |
| 1984 | 1 | 6 | 0 | 0 | 24 | 0 | 0.0015 | 0.0089 | 0.0193 | 0.0297 | 0.0565 |
|  | 0.0996 | 0.1441 | 0.1694 | 0.1441 | 0.1441 | 0.0966 | 0.0609 | 0.0416 | 0.0282 | 0.003 | 0.003 |
|  | 0 | 0 | 0 | 0.0015 | 0.0089 | 0.0193 | 0.0297 | 0.0565 | 0.0996 | 0.1441 | 0.1694 |
|  | 0.1441 | 0.1441 | 0.0966 | 0.0609 | 0.0416 | 0.0282 | 0.003 | 0.003 | 0 | $\bigcirc$ |  |
| \#1985 | 1 | 6 | 0 | 0 | 64 | 0.0002 | 0.0002 | 0.0031 | 0.0025 | 0.006 | 0.0151 |
|  | 0.0501 | 0.1035 | 0.166 | 0.1766 | 0.1708 | 0.1387 | 0.0987 | 0.0735 | 0.036 | 0.0141 | 0.0029 |
|  | 0.0004 | 0 | 0.0002 | 0.0002 | 0.0031 | 0.0025 | 0.006 | 0.0151 | 0.0501 | 0.1035 | 0.166 |
|  | 0.1766 | 0.1708 | 0.1387 | 0.0987 | 0.0735 | 0.036 | 0.0141 | 0.0029 | 0.0004 | 0 |  |
| 1986 | 1 | 6 | 0 | 0 | 103 | 0.0002 | 0.0002 | 0.0007 | 0.0017 | 0.005 | 0.0133 |
|  | 0.0302 | 0.067 | 0.1064 | 0.1577 | 0.1761 | 0.1616 | 0.1546 | 0.1116 | 0.0395 | 0.0135 | 0.0041 |
|  | 0.0007 | 0 | 0.0002 | 0.0002 | 0.0007 | 0.0017 | 0.005 | 0.0133 | 0.0302 | 0.067 | 0.1064 |
|  | 0.1577 | 0.1761 | 0.1616 | 0.1546 | 0.1116 | 0.0395 | 0.0135 | 0.0041 | 0.0007 | 0 |  |
| 1987 | 1 | 6 | 0 | 0 | 122 | $\bigcirc$ | 0.0009 | 0.0025 | 0.007 | 0.0101 | 0.0216 |
|  | 0.0363 | 0.0877 | 0.1338 | 0.1631 | 0.1739 | 0.1853 | 0.134 | 0.0723 | 0.018 | 0.0059 | 0.0021 |
|  | 0.0008 | 0.0002 | 0 | 0.0009 | 0.0025 | 0.007 | 0.0101 | 0.0216 | 0.0363 | 0.0877 | 0.1338 |
|  | 0.1631 | 0.1739 | 0.1853 | 0.134 | 0.0723 | 0.018 | 0.0059 | 0.0021 | 0.0008 | 0.0002 |  |
| 1988 | 1 | 6 | 0 | 0 | 103 | 0.0003 | 0.0004 | 0.0051 | 0.0105 | 0.016 | 0.0326 |
|  | 0.0465 | 0.0603 | 0.0869 | 0.1227 | 0.1433 | 0.1745 | 0.1622 | 0.1071 | 0.0416 | 0.0131 | 0.0045 |
|  | 0.0023 | 0.0006 | 0.0003 | 0.0004 | 0.0051 | 0.0105 | 0.016 | 0.0326 | 0.0465 | 0.0603 | 0.0869 |
|  | 0.1227 | 0.1433 | 0.1745 | 0.1622 | 0.1071 | 0.0416 | 0.0131 | 0.0045 | 0.0023 | 0.0006 |  |
| 1989 | 1 | 6 | 0 | 0 | 103 | 0 | 0.0006 | 0.0033 | 0.0081 | 0.0215 | 0.047 |
|  | 0.0695 | 0.0993 | 0.1085 | 0.1265 | 0.1362 | 0.1476 | 0.1311 | 0.095 | 0.0288 | 0.0105 | 0.0018 |
|  | 0.0007 | 0.0006 | 0 | 0.0006 | 0.0033 | 0.0081 | 0.0215 | 0.047 | 0.0695 | 0.0993 | 0.1085 |
|  | 0.1265 | 0.1362 | 0.1476 | 0.1311 | 0.095 | 0.0288 | 0.0105 | 0.0018 | 0.0007 | 0.0006 |  |
| 1990 | 1 | 6 | 0 | 0 | 108 | 0.0004 | 0.0026 | 0.0116 | 0.0211 | 0.026 | 0.0464 |
|  | 0.0766 | 0.1349 | 0.1533 | 0.1321 | 0.126 | 0.1061 | 0.093 | 0.0684 | 0.0268 | 0.0099 | 0.0026 |
|  | 0.0012 | 0.0007 | 0.0004 | 0.0026 | 0.0116 | 0.0211 | 0.026 | 0.0464 | 0.0766 | 0.1349 | 0.1533 |
|  | 0.1321 | 0.126 | 0.1061 | 0.093 | 0.0684 | 0.0268 | 0.0099 | 0.0026 | 0.0012 | 0.0007 |  |
| 1998 | 1 | 6 | 0 | 0 | 83 | 0 | 0.0019 | 0.0023 | 0.0034 | 0.0129 | 0.0278 |
|  | 0.0636 | 0.12 | 0.2034 | 0.2224 | 0.171 | 0.107 | 0.0468 | 0.0129 | 0.0038 | 0.0008 | 0 |
|  | 0 | $\bigcirc$ | 0 | 0.0019 | 0.0023 | 0.0034 | 0.0129 | 0.0278 | 0.0636 | 0.12 | 0.2034 |
|  | 0.2224 | 0.171 | 0.107 | 0.0468 | 0.0129 | 0.0038 | 0.0008 | 0 | 0 | 0 |  |
| 1999 | 1 | 6 | 0 | $\bigcirc$ | 93 | 0 | 0.0003 | 0.0003 | 0.0029 | 0.0063 | 0.0173 |
|  | 0.0434 | 0.0811 | 0.157 | 0.2105 | 0.1915 | 0.1432 | 0.0894 | 0.0408 | 0.0109 | 0.004 | 0.0009 |
|  | 0 | 0.0003 | 0 | 0.0003 | 0.0003 | 0.0029 | 0.0063 | 0.0173 | 0.0434 | 0.0811 | 0.157 |
|  | 0.2105 | 0.1915 | 0.1432 | 0.0894 | 0.0408 | 0.0109 | 0.004 | 0.0009 | 0 | 0.0003 |  |
| 2000 | 1 | 6 | 0 | 0 | 78 | 0 | 0.0007 | 0.0011 | 0.0011 | 0.0093 | 0.0237 |
|  | 0.0714 | 0.1104 | 0.166 | 0.2302 | 0.1775 | 0.1233 | 0.0567 | 0.0183 | 0.0093 | 0.0011 | 0 |
|  | 0 | $\bigcirc$ | 0 | 0.0007 | 0.0011 | 0.0011 | 0.0093 | 0.0237 | 0.0714 | 0.1104 | 0.166 |
|  | 0.2302 | 0.1775 | 0.1233 | 0.0567 | 0.0183 | 0.0093 | 0.0011 | 0 | 0 | 0 |  |
| 2001 | 1 | 6 | 0 | 0 | 78 | 0.0003 | 0 | 0.0003 | 0.0016 | 0.0041 | 0.0062 |
|  | 0.0212 | 0.0614 | 0.1156 | 0.1911 | 0.2347 | 0.1911 | 0.1141 | 0.0396 | 0.0128 | 0.0041 | 0.0016 |


|  | 0.0003 | 0 | 0.0003 | 0 | 0.0003 | 0.0016 | 0.0041 | 0.0062 | 0.0212 | 0.0614 | 0.1156 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.1911 | 0.2347 | 0.1911 | 0.1141 | 0.0396 | 0.0128 | 0.0041 | 0.0016 | 0.0003 | 0 |  |
| 2002 | 1 | 6 | 0 | 0 | 49 | 0 | 0 | 0 | 0.0012 | 0.0017 | 0.0113 |
|  | 0.0237 | 0.0614 | 0.1177 | 0.1955 | 0.2214 | 0.1781 | 0.115 | 0.0521 | 0.0135 | 0.0049 | 0.0015 |
|  | 0.0007 | 0.0005 | 0 | 0 | 0 | 0.0012 | 0.0017 | 0.0113 | 0.0237 | 0.0614 | 0.1177 |
|  | 0.1955 | 0.2214 | 0.1781 | 0.115 | 0.0521 | 0.0135 | 0.0049 | 0.0015 | 0.0007 | 0.0005 |  |
| 2003 | 1 | 6 | 0 | 0 | 78 | 0.0007 | 0 | 0.0006 | 0.001 | 0.0013 | 0.0043 |
|  | 0.0196 | 0.0444 | 0.1013 | 0.1739 | 0.2486 | 0.221 | 0.1182 | 0.0505 | 0.0123 | 0.0015 | 0.0004 |
|  | 0.0003 | 0 | 0.0007 | 0 | 0.0006 | 0.001 | 0.0013 | 0.0043 | 0.0196 | 0.0444 | 0.1013 |
|  | 0.1739 | 0.2486 | 0.221 | 0.1182 | 0.0505 | 0.0123 | 0.0015 | 0.0004 | 0.0003 | 0 |  |
| 2004 | 1 | 6 | 0 | 0 | 68 | 0.0005 | $\bigcirc$ | 0.0005 | 0.0028 | 0.0065 | 0.0138 |
|  | 0.0242 | 0.0615 | 0.136 | 0.2066 | 0.2167 | 0.1753 | 0.0969 | 0.0399 | 0.0137 | 0.0036 | 0.0013 |
|  | 0.0002 | 0.0002 | 0.0005 | 0 | 0.0005 | 0.0028 | 0.0065 | 0.0138 | 0.0242 | 0.0615 | 0.136 |
|  | 0.2066 | 0.2167 | 0.1753 | 0.0969 | 0.0399 | 0.0137 | 0.0036 | 0.0013 | 0.0002 | 0.0002 |  |
| 2005 | 1 | 6 | 0 | $\bigcirc$ | 49 | 0.0005 | 0.0005 | 0.0003 | 0.001 | 0.0043 | 0.0205 |
|  | 0.0352 | 0.0777 | 0.1372 | 0.197 | 0.2051 | 0.1752 | 0.0972 | 0.0337 | 0.0104 | 0.003 | 0.0005 |
|  | 0.0005 | 0 | 0.0005 | 0.0005 | 0.0003 | 0.001 | 0.0043 | 0.0205 | 0.0352 | 0.0777 | 0.1372 |
|  | 0.197 | 0.2051 | 0.1752 | 0.0972 | 0.0337 | 0.0104 | 0.003 | 0.0005 | 0.0005 | 0 |  |
| 2006 | 1 | 6 | 0 | 0 | 64 | 0 | 0.0005 | 0.0017 | 0.0025 | 0.0035 | 0.0153 |
|  | 0.038 | 0.0824 | 0.1454 | 0.1829 | 0.2063 | 0.1624 | 0.0953 | 0.0445 | 0.0146 | 0.003 | 0.001 |
|  | 0.0003 | 0.0002 | 0 | 0.0005 | 0.0017 | 0.0025 | 0.0035 | 0.0153 | 0.038 | 0.0824 | 0.1454 |
|  | 0.1829 | 0.2063 | 0.1624 | 0.0953 | 0.0445 | 0.0146 | 0.003 | 0.001 | 0.0003 | 0.0002 |  |
| \# Age Compo |  |  |  |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \text { \#Yr } \\ & \text { male) } \end{aligned}$ | Seas | Flt/Svy Gender |  | Part | Ageerr | Lbin_lo Lbin_hi Nsamp |  |  | datavector(female- |  |  |
| 24 |  |  |  |  |  |  |  |  |  |  |  |
| 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
|  | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
|  | 30 |  |  |  |  |  |  |  |  |  |  |
| \# | number | of | unique | ageing | error | matrice | S | to | generat |  |  |
| 1 lor |  |  |  |  |  |  |  |  |  |  |  |
| 0.5 | 1.5 | 2.5 | 3.5 | 4.5 | 5.5 | 6.5 | 7.5 | 8.5 | 9.5 | 10.5 | 11.5 |
|  | 12.5 | 13.5 | 14.5 | 15.5 | 16.5 | 17.5 | 18.5 | 19.5 | 20.5 | 21.5 | 22.5 |
|  | 23.5 | 24.5 | 25.5 | 26.5 | 27.5 | 28.5 | 29.5 | 30.5 | 31.5 | 32.5 | 33.5 |
|  | 34.5 | 35.5 | 36.5 | 37.5 | 38.5 | 39.5 | 40.5 |  |  |  |  |
| 0.4817 | 0.5149 | 0.5481 | 0.5813 | 0.6145 | 0.6477 | 0.6809 | 0.7141 | 0.7473 | 0.7805 | 0.8137 | 0.8469 |
|  | 0.8801 | 0.9133 | 0.9465 | 0.9797 | 1.0129 | 1.0461 | 1.0793 | 1.1125 | 1.1457 | 1.1789 | 1.2121 |
|  | 1.2453 | 1.2785 | 1.3117 | 1.3449 | 1.3781 | 1.4113 | 1.4445 | 1.4777 | 1.5109 | 1.5441 | 1.5773 |
|  | 1.6105 | 1.6437 | 1.6769 | 1.7101 | 1.7433 | 1.7765 | 1.7765 |  |  |  |  |
| \#Sampson Below |  |  |  |  |  |  |  |  |  |  |  |
| \#0. 5 | 1.5 | 2.5 | 3.5 | 4.5 | 5.5 | 6.5 | 7.5 | 8.5 | 9.5 | 10.5 | 11.5 |
|  | 12.5 | 13.5 | 14.5 | 15.5 | 16.5 | 17.5 | 18.5 | 19.5 | 20.5 | 21.5 | 22.5 |
|  | 23.5 | 24.5 | 25.5 | 26.5 | 27.5 | 28.5 | 29.5 | 30.5 | 31.5 | 32.5 | 33.5 |
|  | 34.5 | 35.5 | 36.5 | 37.5 | 38.5 | 39.5 | 40.5 |  |  |  |  |
| \#0. 062 | 0.186 | 0.310 | 0.435 | 0.559 | 0.683 | 0.807 | 0.931 | 1.056 | 1.180 | 1.304 | 1.428 |
|  | 1.552 | 1.676 | 1.801 | 1.925 | 2.049 | 2.173 | 2.297 | 2.422 | 2.546 | 2.670 | 2.794 |
|  | 2.918 | 3.043 | 3.167 | 3.291 | 3.415 | 3.539 | 3.663 | 3.788 | 3.912 | 4.036 | 4.160 |
|  | 4.284 | 4.409 | 4.533 | 4.657 | 4.781 | 4.905 | 5.029 |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |
| 53 |  |  |  |  |  |  |  |  |  |  |  |
| 1976 | 1 | 1 | 3 | $\bigcirc$ | 1 | -1 | -1 | 14 | 0 | 0 | 0 |
|  | 0.0084 | 0.021 | 0.0924 | 0.0504 | 0.0714 | 0.0672 | 0.0588 | 0.0336 | 0.0336 | 0.021 | 0.0126 |
|  | 0.0042 | 0.0084 | 0.0042 | 0 | 0.0084 | 0 | $\bigcirc$ | $\bigcirc$ | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.0294 | 0.0882 | 0.0798 | 0.0672 | 0.042 | 0.0504 | 0.0504 | 0.021 |
|  | $\begin{aligned} & 0.0168 \\ & 0 \end{aligned}$ | 0.0126 | 0.0084 | 0.0042 | 0.0042 | 0 | 0 | 0.0126 | 0 | 0.0042 | 0.0126 |
| 1980 | 1 | 1 | 3 | $\bigcirc$ | 1 | -1 | -1 | 14 | 0 | 0 | 0 |
|  | 0.0205 | 0.0256 | 0.041 | 0.0462 | 0.0051 | 0.0359 | 0.041 | 0.0513 | 0.0359 | 0.0051 | 0 |
|  | 0.0103 | 0.0154 | 0 | 0 | 0.0103 | 0 | 0 | 0.0051 | 0 | 0.0051 | 0 |
|  | 0 | 0 | 0.0256 | 0.0564 | 0.041 | 0.0462 | 0.0462 | 0.041 | 0.041 | 0.0769 | 0.0615 |
|  | $\begin{aligned} & 0.0308 \\ & 0.0103 \end{aligned}$ | 0.0103 | 0.0462 | 0.0154 | 0.0103 | 0.0103 | $\bigcirc$ | 0.0103 | 0.0103 | 0 | 0.0564 |
| 1981 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 28 | 0 | 0 | 0.0025 |
|  | 0.0127 | 0.0406 | 0.0457 | 0.0457 | 0.066 | 0.0635 | 0.0457 | 0.0355 | 0.0178 | 0.0228 | 0.0127 |
|  | 0.0102 | 0.0051 | 0.0025 | 0 | 0.0025 | 0.0025 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0.0025 | 0.0152 | 0.0508 | 0.0787 | 0.0457 | 0.0558 | 0.0279 | 0.0381 | 0.0635 | 0.0533 |


|  | 0.0152 | 0.0152 | 0.0228 | 0.0178 | 0.0076 | 0.0051 | 0.0051 | 0.0102 | 0.0025 | 0.0127 | 0.0127 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.0076 |  |  |  |  |  |  |  |  |  |  |
| 1982 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 21 | 0 | $\bigcirc$ | 0.0034 |
|  | 0.0576 | 0.061 | 0.0814 | 0.0373 | 0.0373 | 0.0169 | 0.0305 | 0.0305 | 0.0102 | 0.0034 | 0.0169 |
|  | 0.0034 | 0.0102 | $\bigcirc$ | 0 | $\bigcirc$ | 0.0034 | 0 | 0 | 0 | $\bigcirc$ | 0 |
|  | 0 | 0.0034 | 0.0814 | 0.0814 | 0.1186 | 0.061 | 0.0339 | 0.0305 | 0.0203 | 0.0305 | 0.0305 |
|  | 0.0237 | 0.0169 | 0.0068 | 0.0102 | 0.0169 | 0 | 0 | 0.0068 | 0.0068 | 0.0034 | 0.0034 |
|  | 0.0034 |  |  |  |  |  |  |  |  |  |  |
| 1983 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 56 | 0.0013 | 0 | 0.0101 |
|  | 0.0277 | 0.0806 | 0.0529 | 0.0856 | 0.0516 | 0.0428 | 0.029 | 0.0189 | 0.0151 | 0.0101 | 0.0063 |
|  | 0.0025 | 0 | $\bigcirc$ | 0.0013 | $\bigcirc$ | 0 | 0.0025 | 0.0038 | $\bigcirc$ | 0.0013 | 0 |
|  | 0 | 0.0126 | 0.0416 | 0.0957 | 0.0642 | 0.0844 | 0.0592 | 0.0302 | 0.0327 | 0.0277 | 0.0227 |
|  | 0.0189 | 0.0176 | 0.0113 | 0.0063 | 0.0038 | 0.005 | 0.005 | 0.0038 | 0.0038 | 0.0013 | 0.0038 |
|  | 0.005 |  |  |  |  |  |  |  |  |  |  |
| 1984 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 21 | 0 | $\bigcirc$ | 0 |
|  | 0.0101 | 0.0101 | 0.0303 | 0.0471 | 0.0976 | 0.064 | 0.037 | 0.0236 | 0.0168 | 0.0135 | 0.0101 |
|  | 0.0269 | 0.0067 | 0.0034 | 0.0034 | 0.0034 | 0 | 0 | 0 | $\bigcirc$ | $\bigcirc$ | 0 |
|  | 0 | 0 | 0.0101 | 0.037 | 0.0606 | 0.0539 | 0.0842 | 0.0673 | 0.0404 | 0.037 | 0.0303 |
|  | 0.0168 | 0.0202 | 0.0269 | 0.0135 | 0.0135 | 0.0101 | 0.0269 | 0.0034 | 0 | 0.0101 | 0.0168 |
|  | 0.0168 |  |  |  |  |  |  |  |  |  |  |
| 1986 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 57 | 0 | $\bigcirc$ | 0.0031 |
|  | 0.028 | 0.053 | 0.0841 | 0.0872 | 0.0841 | 0.0405 | 0.0592 | 0.0249 | 0.0249 | 0.0156 | 0.0062 |
|  | 0.0156 | 0.0156 | 0.0062 | 0 | $\bigcirc$ | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 |
|  | 0 | 0.0093 | 0.0218 | 0.081 | 0.0997 | 0.0561 | 0.0498 | 0.0249 | 0.0312 | 0.0218 | 0.0125 |
|  | 0.0218 | 0.0031 | 0 | 0.0031 | 0.0031 | 0.0031 | 0.0031 | 0.0031 | $\bigcirc$ | 0 | 0.0031 |
|  | 0 |  |  |  |  |  |  |  |  |  |  |
| 1987 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 71 | 0 | $\bigcirc$ | 0.0025 |
|  | 0.0075 | 0.0249 | 0.0723 | 0.0848 | 0.0574 | 0.0673 | 0.0648 | 0.0499 | 0.0324 | 0.0349 | 0.0125 |
|  | 0.0125 | 0.0025 | 0 | 0.01 | 0.005 | 0.005 | 0.0025 | 0.0025 | 0 | 0 | 0 |
|  | 0 | $\bigcirc$ | 0.01 | 0.02 | 0.0698 | 0.0873 | 0.0499 | 0.0374 | 0.0549 | 0.0324 | 0.0175 |
|  | 0.02 | 0.015 | 0.01 | 0.0025 | 0.005 | 0.0025 | 0.0025 | 0 | 0 | 0.0025 | 0.0075 |
|  | 0.0025 |  |  |  |  |  |  |  |  |  |  |
| 1988 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 18 | 0 | $\bigcirc$ | 0 |
|  | 0 | 0.0202 | 0.0202 | 0.1212 | 0.0505 | 0.0808 | 0.0505 | 0.0303 | 0.0404 | 0.0303 | 0.0202 |
|  | 0.0202 | 0 | 0.0202 | 0 | $\bigcirc$ | 0.0101 | 0 | 0 | 0 | $\bigcirc$ | 0 |
|  | 0 | 0 | 0.0101 | 0.0101 | 0.0404 | 0.0707 | 0.0707 | 0.0303 | 0.0808 | 0.0303 | 0.0101 |
|  | 0.0404 | 0.0101 | 0.0202 | 0 | 0.0101 | 0.0101 | 0.0202 | 0 | 0.0202 | 0 | 0 |
|  | 0 |  |  |  |  |  |  |  |  |  |  |
| 1989 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 40 | 0 | $\bigcirc$ | 0 |
|  | 0.0179 | 0.0268 | 0.0848 | 0.0938 | 0.0982 | 0.1027 | 0.067 | 0.0625 | 0.0357 | 0.0179 | 0.0179 |
|  | 0.0045 | 0 | 0.0089 | 0.0045 | $\bigcirc$ | 0 | 0 | 0 | $\bigcirc$ | $\bigcirc$ | 0 |
|  | 0 | 0.0089 | 0.0089 | 0.0357 | 0.0179 | 0.0893 | 0.0804 | 0.0357 | 0.0134 | 0.0313 | 0.0134 |
|  | 0.0134 | 0 | 0.0089 | 0 | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 |
|  | 0 |  |  |  |  |  |  |  |  |  |  |
| 1990 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 44 | $\bigcirc$ | $\bigcirc$ | 0 |
|  | 0.004 | 0.0361 | 0.0482 | 0.0482 | 0.0562 | 0.0602 | 0.0643 | 0.0402 | 0.0241 | 0.012 | 0.0281 |
|  | 0.0201 | 0.008 | 0.004 | 0.004 | $\bigcirc$ | 0 | 0 | 0 | 0.004 | $\bigcirc$ | 0 |
|  | 0 | 0.008 | 0.0201 | 0.0442 | 0.0683 | 0.0803 | 0.0562 | 0.0763 | 0.0643 | 0.012 | 0.012 |
|  | 0.0241 | 0.0321 | 0.008 | 0.012 | $\bigcirc$ | 0 | 0.0161 | 0 | $\bigcirc$ | 0 | 0.004 |
|  | 0 |  |  |  |  |  |  |  |  |  |  |
| 1991 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 54 | 0 | $\bigcirc$ | 0 |
|  | 0.0133 | 0.0365 | 0.0664 | 0.0797 | 0.0565 | 0.0532 | 0.0565 | 0.0365 | 0.0266 | 0.0199 | 0.0133 |
|  | 0.0133 | 0 | 0 | 0.0033 | $\bigcirc$ | 0.0033 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.0066 | 0.0299 | 0.0631 | 0.0399 | 0.0532 | 0.0399 | 0.0465 | 0.0399 | 0.01 |
|  | 0.0199 | 0.0332 | 0.0199 | 0.0133 | 0.01 | 0.0299 | 0.0166 | 0.0166 | 0.01 | 0.0066 | 0.0166 |
|  | 0 |  |  |  |  |  |  |  |  |  |  |
| 1992 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 36 | $\bigcirc$ | 0 | 0 |
|  | 0.015 | 0.055 | 0.08 | 0.135 | 0.05 | 0.04 | 0.025 | 0.05 | 0.015 | 0.005 | 0.005 |
|  | 0.01 | 0.01 | $\bigcirc$ | 0 | $\bigcirc$ | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 |
|  | 0 | 0 | 0.02 | 0.03 | 0.11 | 0.115 | 0.06 | 0.015 | 0.04 | 0.03 | 0.045 |
|  | 0.015 | 0.01 | 0.015 | 0 | $\bigcirc$ | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 |
|  | 0 |  |  |  |  |  |  |  |  |  |  |
| 1993 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 22 | 0 | 0 | 0 |
|  | 0.016 | 0.056 | 0.04 | 0.072 | 0.024 | 0.056 | 0.08 | 0.048 | 0.008 | 0.016 | 0.008 |
|  | 0.008 | 0 | $\bigcirc$ | 0 | $\bigcirc$ | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 |
|  | 0 | 0 | $\bigcirc$ | 0.024 | 0.112 | 0.088 | 0.088 | 0.048 | 0.04 | 0.048 | 0.024 |
|  | $\begin{aligned} & 0.024 \\ & 0 \end{aligned}$ | 0.032 | 0.016 | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | 0.016 | $\bigcirc$ | 0.008 | $\bigcirc$ | $\bigcirc$ |
| 1994 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 9 | 0 | 0 | 0 |
|  | 0.0208 | 0.0625 | 0.0208 | 0.0208 | 0.0417 | 0.0625 | 0.0625 | 0.0417 | 0 | 0.0208 | 0 |
|  | 0.0208 | 0.0208 | 0 | 0.0208 | 0 | $\bigcirc$ | 0 | 0 | $\bigcirc$ | $\bigcirc$ | 0 |


|  | 0 | $\bigcirc$ | 0 | 0 | 0.0417 | 0.0833 | 0.0625 | 0 | 0.0625 | 0.0625 | 0.0208 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.0208 | 0.0625 | 0.0625 | 0 | 0 | 0 | 0 | 0.0208 | 0.0208 | 0 | 0.0417 |
|  | 0.0208 |  |  |  |  |  |  |  |  |  |  |
| 1995 | 1 | 1 | 3 | 0 | 1 | -1 | -1 | 9 | 0 | 0 | 0 |
|  | 0.0204 | 0.0612 | 0 | 0.0612 | 0 | 0.0204 | 0 | 0.0408 | 0.0204 | 0 | 0 |
|  | $\bigcirc$ | 0.0204 | 0 | $\bigcirc$ | 0 | $\bigcirc$ | 0 | $\bigcirc$ | $\bigcirc$ | 0 | 0.0204 |
|  | 0 | 0 | 0 | 0.0204 | 0.0612 | 0.0612 | 0.0612 | 0.0816 | 0.102 | 0.102 | 0.0612 |
|  | $\begin{aligned} & 0.0612 \\ & 0 \end{aligned}$ | 0 | 0.0204 | 0.0408 | 0.0204 | 0 | 0.0204 | 0 | $\bigcirc$ | 0.0204 | 0 |
| 1980 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 28 | 0 | 0.0027 | 0.0027 |
|  | 0.0192 | 0.022 | 0.0275 | 0.0412 | 0.022 | 0.0165 | 0.0467 | 0.0192 | 0.011 | 0.0275 | 0.011 |
|  | 0.011 | 0.0082 | 0.0055 | 0.0055 | 0.0027 | 0.0027 | 0 | 0.0027 | 0.0055 | 0.0027 | 0 |
|  | $\bigcirc$ | 0.0055 | 0.0247 | 0.033 | 0.0687 | 0.0412 | 0.033 | 0.0412 | 0.0467 | 0.0357 | 0.011 |
|  | $\begin{aligned} & 0.033 \\ & 0.0275 \end{aligned}$ | 0.0495 | 0.0275 | 0.0275 | 0.033 | 0.0302 | 0.0165 | 0.0247 | 0.011 | 0.011 | 0.0522 |
| 1981 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 12 | 0.0141 | 0.0141 | 0 |
|  | 0.0423 | 0.0141 | 0.0141 | 0.0282 | 0.0423 | 0.0563 | 0.0282 | 0.0141 | 0.0282 | 0 | 0 |
|  | $\bigcirc$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 |
|  | 0 | 0.0282 | 0.0282 | 0.0141 | 0.1408 | 0.0986 | 0.0423 | 0.1268 | 0.0282 | 0.0423 | 0.0141 |
|  | $\begin{aligned} & 0.0423 \\ & 0 \end{aligned}$ | 0.0141 | 0.0141 | 0.0282 | 0 | 0.0282 | 0.0141 | 0 | $\bigcirc$ | 0 | $\bigcirc$ |
| 1984 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 134 | 0 | 0.0086 | 0.0375 |
|  | 0.0893 | 0.1081 | 0.0634 | 0.0605 | 0.0879 | 0.049 | 0.0288 | 0.0245 | 0.0115 | 0.013 | 0.013 |
|  | 0.0072 | 0.0072 | 0.0029 | 0 | 0.0014 | 0.0029 | 0 | 0 | 0 | 0 | 0 |
|  | 0.0029 | 0.0029 | 0.0303 | 0.0346 | 0.0331 | 0.0418 | 0.049 | 0.0346 | 0.0173 | 0.0159 | 0.0144 |
|  | 0.0202 | 0.013 | 0.0173 | 0.0029 | 0.0043 | 0.0029 | 0.0086 | 0.0043 | 0.0086 | 0.0058 | 0.013 |
|  | 0.0058 |  |  |  |  |  |  |  |  |  |  |
| 1985 | 1 | 2 | 3 | $\bigcirc$ | 1 | -1 | -1 | 14 | $\bigcirc$ | 0.0063 | 0.0253 |
|  | 0.0633 | 0.038 | 0.0886 | 0.0443 | 0.0316 | 0.0316 | 0.0506 | 0.0316 | 0.0253 | 0.0127 | 0 |
|  | 0.0063 | 0 | 0.0063 | 0.0063 | 0.0063 | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 |
|  | 0 | 0.0127 | 0.0316 | 0.0633 | 0.0759 | 0.0316 | 0.0253 | 0.0253 | 0.0253 | 0.0633 | 0.038 |
|  | 0 | 0.0316 | 0.0316 | 0.0063 | 0.0127 | 0.038 | 0 | 0.0063 | 0 | 0 | 0.0063 |
|  | 0 |  |  |  |  |  |  |  |  |  |  |
| 1986 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | $91$ | 0 | 0.0079 | 0.0237 |
|  | 0.0632 | 0.0731 | 0.0751 | 0.0593 | 0.0514 | 0.0296 | 0.0296 | $0.0217$ | 0.0119 | 0.0079 | 0.0059 |
|  | 0.004 | 0.0059 | 0.002 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0 | 0 |
|  | 0.0079 | 0.0356 | 0.0711 | 0.0909 | 0.0889 | 0.0534 | 0.0356 | 0.0237 | 0.0138 | 0.0198 | 0.0138 |
|  | $0.0099$ | 0.0138 | 0.0119 | 0.0138 | 0.002 | 0.0059 | 0.0059 | $\bigcirc$ | 0.002 | 0.002 | 0.004 |
| 1987 | 1 | 2 | 3 | $\bigcirc$ | 1 | -1 | -1 | 112 | 0 | 0.0093 | 0.0125 |
|  | 0.0327 | 0.0545 | 0.0654 | 0.0607 | 0.0607 | 0.0327 | 0.0452 | 0.028 | 0.0405 | 0.0171 | 0.0156 |
|  | 0.0109 | 0.0078 | 0.0125 | 0.0016 | 0.0016 | 0.0016 | 0 | 0 | 0.0031 | 0 | 0 |
|  | 0.0078 | 0.0171 | 0.0234 | 0.0498 | 0.0452 | 0.0374 | 0.0312 | 0.0405 | 0.0249 | 0.0249 | 0.0389 |
|  | 0.0296 | 0.0093 | 0.0156 | 0.0171 | 0.0171 | 0.014 | 0.0125 | 0.0016 | 0.0031 | 0.0031 | 0.0156 |
|  | 0.0062 |  |  |  |  |  |  |  |  |  |  |
| 1988 | $1$ | 2 | 3 | $0$ | $1$ |  |  | $80$ | $0$ | $0.0067$ | $0.0201$ |
|  | $0.0446$ | 0.0804 | 0.0759 | 0.0625 | 0.0848 | 0.0513 | 0.0201 | $0.0179$ | 0.0268 | $0.0112$ | $0.0112$ |
|  | 0.0022 | 0.0022 | 0.0112 | 0.0022 | 0 | 0 | 0.0045 | 0.0022 | 0 | 0 | 0 |
|  | 0.0022 | 0.0223 | 0.0424 | 0.067 | 0.0513 | 0.0469 | 0.058 | 0.0424 | 0.0201 | 0.0179 | 0.0156 |
|  | $\begin{aligned} & 0.029 \\ & 0 \end{aligned}$ | 0.0045 | 0.0089 | 0.0045 | 0.0067 | 0.0067 | 0.0045 | 0.0045 | 0.0022 | 0 | 0.0045 |
| 1989 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 71 | $\bigcirc$ | 0 | 0.0101 |
|  | 0.0658 | 0.0759 | 0.0911 | 0.0709 | 0.0608 | 0.0405 | 0.0456 | 0.0228 | 0.0177 | 0.0228 | 0.0076 |
|  | 0.0051 | 0.0025 | 0.0025 | 0 | 0.0025 | 0 | 0 | 0 | 0.0025 | 0 | 0 |
|  | 0.0152 | 0.0203 | 0.0759 | 0.0861 | 0.0734 | 0.0481 | 0.0278 | 0.0278 | 0.0127 | 0.0101 | 0.0152 |
|  | $\begin{aligned} & 0.0101 \\ & 0 \end{aligned}$ | $\bigcirc$ | 0.0076 | 0.0051 | 0.0025 | 0.0076 | 0.0025 | 0.0025 | 0.0025 | 0 | 0 |
| 1990 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 51 | 0 | 0.0035 | 0.0035 |
|  | 0.0588 | 0.128 | 0.0934 | 0.09 | 0.0519 | 0.0519 | 0.0208 | 0.0173 | 0.0104 | 0.0104 | 0.0035 |
|  | $\bigcirc$ | 0 | 0 | $\bigcirc$ | 0 | 0 | 0 | 0 | 0 | 0 | 0.0035 |
|  | 0 | 0.0208 | 0.0554 | 0.0623 | 0.0761 | 0.0415 | 0.0484 | 0.0242 | 0.0242 | 0.0208 | 0.0173 |
|  | $\begin{aligned} & 0.0104 \\ & 0 \end{aligned}$ | 0.0104 | 0.0104 | 0 | 0.0069 | $\bigcirc$ | 0.0104 | 0.0035 | 0.0035 | 0 | 0.0069 |
| 1991 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 121 | 0.0042 | 0.0195 | 0.0432 |
|  | 0.0669 | 0.0628 | 0.1144 | 0.0753 | 0.0474 | 0.0223 | 0.0237 | 0.0237 | 0.0153 | 0.0112 | 0.0056 |
|  | 0.0084 | 0.0014 | 0.0014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0028 |
|  | 0.0181 | 0.0181 | 0.0474 | 0.0669 | 0.0725 | 0.053 | 0.0404 | 0.0181 | 0.0209 | 0.0181 | 0.0098 |
|  | 0.0139 | 0.0056 | 0.0126 | 0.0084 | 0.0056 | 0.0014 | 0.0042 | 0.0042 | 0.0042 | 0.0014 | 0.0028 |
|  | 0.0028 |  |  |  |  |  |  |  |  |  |  |
| 1992 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 157 | 0.0011 | 0.0125 | 0.0227 |
|  | 0.0409 | 0.0727 | 0.0864 | 0.0841 | 0.067 | 0.033 | 0.0307 | 0.0216 | 0.0136 | 0.0148 | 0.0045 |


|  | 0.0034 | 0.0057 | 0.0011 | 0.0011 | 0 | 0.0011 | 0.0011 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.0114 | 0.0318 | 0.0295 | 0.067 | 0.0693 | 0.0614 | 0.0398 | 0.0295 | 0.0227 | 0.025 | 0.0182 |
|  | 0.0182 | 0.0091 | 0.008 | 0.0125 | 0.0034 | 0.0034 | 0.0011 | 0.0023 | 0.0023 | 0.0057 | 0.0068 |
|  | 0.0023 |  |  |  |  |  |  |  |  |  |  |
| 1993 | 1 | 2 | 3 | $\bigcirc$ | 1 | -1 | -1 | 154 | 0 | 0.0082 | 0.0397 |
|  | 0.0806 | 0.0771 | 0.0993 | 0.0993 | 0.0561 | 0.0409 | 0.0129 | 0.0175 | 0.0117 | 0.014 | 0.007 |
|  | 0.0023 | 0.0023 | 0.0047 | 0.0012 | 0.0012 | 0 | $\bigcirc$ | 0 | 0.0012 | 0 | 0 |
|  | 0.0082 | 0.0234 | 0.0666 | 0.0537 | 0.0806 | 0.0584 | 0.035 | 0.0199 | 0.0187 | 0.0129 | 0.0023 |
|  | 0.0058 | 0.007 | 0.0117 | 0.0023 | 0.0035 | 0.0035 | 0.0035 | 0 | 0 | 0.0012 | 0.0035 |
|  | 0.0012 |  |  |  |  |  |  |  |  |  |  |
| 1994 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 155 | 0 | 0.007 | 0.0348 |
|  | 0.0615 | 0.0893 | 0.08 | 0.058 | 0.0545 | 0.0394 | 0.0244 | 0.0174 | 0.0162 | 0.0128 | 0.0151 |
|  | 0.0081 | 0.0058 | 0.0046 | 0.0035 | 0.0012 | 0 | $\bigcirc$ | 0 | 0.0012 | 0.0023 | 0.0012 |
|  | 0.007 | 0.0383 | 0.0638 | 0.0615 | 0.0464 | 0.0487 | 0.0441 | 0.0302 | 0.022 | 0.0232 | 0.022 |
|  | 0.0128 | 0.0058 | 0.0046 | 0.0023 | 0.0093 | 0.0023 | 0.0046 | 0.0046 | 0.0023 | 0 | 0.0035 |
|  | 0.0023 |  |  |  |  |  |  |  |  |  |  |
| 1995 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 144 | 0.0012 | 0.0025 | 0.0395 |
|  | 0.0875 | 0.0912 | 0.1036 | 0.0666 | 0.0518 | 0.0259 | 0.021 | 0.0197 | 0.0123 | 0.0074 | 0.0062 |
|  | 0.0012 | 0.0012 | 0.0025 | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 | 0 | 0 |
|  | 0.0049 | 0.0456 | 0.0691 | 0.0937 | 0.0654 | 0.0469 | 0.0358 | 0.0222 | 0.0222 | 0.0037 | 0.0086 |
|  | 0.0062 | 0.0086 | 0.0062 | 0.0049 | 0.0037 | 0.0025 | 0.0037 | 0 | 0.0012 | 0.0012 | 0 |
|  | 0.0025 |  |  |  |  |  |  |  |  |  |  |
| 1996 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 147 | 0 | 0.0036 | 0.0364 |
|  | 0.0836 | 0.1358 | 0.0727 | 0.0764 | 0.0364 | 0.0255 | 0.0255 | 0.0097 | 0.0024 | 0.0012 | 0.0048 |
|  | 0.0036 | 0 | 0 | 0 | 0 | 0 | 0.0012 | 0 | 0 | $\bigcirc$ | 0 |
|  | 0.0097 | 0.0339 | 0.08 | 0.1067 | 0.0764 | 0.0533 | 0.0206 | 0.0267 | 0.0121 | 0.0121 | 0.0085 |
|  | 0.0073 | 0.0061 | 0.0097 | 0.0048 | 0.0024 | 0.0024 | $\bigcirc$ | 0.0012 | 0.0012 | 0.0012 | 0.0036 |
|  | 0.0012 |  |  |  |  |  |  |  |  |  |  |
| 1997 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 159 | 0 | 0 | 0.0246 |
|  | 0.047 | 0.0761 | 0.0873 | 0.0672 | 0.0571 | 0.0493 | 0.0246 | 0.0202 | 0.0112 | 0.0123 | 0.0056 |
|  | 0.0056 | 0.0056 | 0.0022 | 0 | 0.0022 | 0 | $\bigcirc$ | 0 | 0.0011 | 0.0022 | 0 |
|  | 0.0067 | 0.0269 | 0.0437 | 0.0672 | 0.1019 | 0.0649 | 0.0414 | 0.0302 | 0.0224 | 0.0235 | 0.0123 |
|  | 0.0078 | 0.0101 | 0.0067 | 0.0078 | 0.0022 | 0.0056 | 0.0045 | 0.0022 | 0.0022 | 0.0034 | 0.0022 |
|  | 0.0022 |  |  |  |  |  |  |  |  |  |  |
| 1998 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 220 | 0 | 0.0182 | 0.0174 |
|  | 0.0477 | 0.0795 | 0.0811 | 0.0795 | 0.0462 | 0.0311 | 0.0364 | 0.0159 | 0.0106 | 0.0061 | 0.0038 |
|  | 0.0038 | 0.0015 | 0 | 0.0015 | 0.0008 | 0.0008 | 0.0008 | 0 | 0.0008 | 0 | 0 |
|  | 0.0159 | 0.0265 | 0.0386 | 0.0644 | 0.0909 | 0.0742 | 0.0508 | 0.0326 | 0.0197 | 0.0189 | 0.0129 |
|  | $\begin{aligned} & 0.0167 \\ & 0 \end{aligned}$ | 0.0083 | 0.0076 | 0.0076 | 0.0076 | 0.0053 | 0.0045 | 0.0053 | 0.0015 | 0.0023 | 0.0045 |
| 1999 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 240 | 0.0024 | 0.0127 | 0.0453 |
|  | 0.0375 | 0.0689 | 0.0943 | 0.0647 | 0.058 | 0.029 | 0.0314 | 0.0169 | 0.0115 | 0.0085 | 0.0048 |
|  | 0.0006 | 0.0018 | 0.0012 | 0 | 0.0012 | 0.0006 | 0.0006 | 0.0006 | 0 | 0 | 0.0006 |
|  | 0.0133 | 0.0538 | 0.0417 | 0.0749 | 0.0725 | 0.0616 | 0.0508 | 0.0302 | 0.0242 | 0.0169 | 0.0139 |
|  | 0.0109 | 0.006 | 0.0066 | 0.0018 | 0.006 | 0.006 | 0.0048 | 0.0036 | 0.0006 | 0.0018 | 0.0036 |
|  | 0.0012 |  |  |  |  |  |  |  |  |  |  |
| 2000 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 233 | 0.0006 | 0.003 | 0.0231 |
|  | 0.0669 | 0.0681 | 0.0748 | 0.0809 | 0.0596 | 0.0663 | 0.0383 | 0.0255 | 0.0152 | 0.0085 | 0.0043 |
|  | 0.0061 | 0.003 | 0.003 | 0.0006 | 0 | 0.0018 | 0 | 0 | 0.0006 | 0 | 0.0006 |
|  | 0.0043 | 0.017 | 0.062 | 0.0669 | 0.0724 | 0.0408 | 0.0408 | 0.0347 | 0.028 | 0.0152 | 0.0158 |
|  | 0.0073 | 0.0091 | 0.0049 | 0.0073 | 0.0036 | 0.0024 | 0.0043 | 0.0006 | 0.0043 | 0.0024 | 0.0036 |
|  | 0.0012 |  |  |  |  |  |  |  |  |  |  |
| 2001 | 1 | 2 | 3 | $\bigcirc$ | 1 | -1 | -1 | 254 | 0 | 0.0034 | 0.0141 |
|  | 0.0401 | 0.1073 | 0.0915 | 0.0554 | 0.0446 | 0.0542 | 0.0475 | 0.0362 | 0.0158 | 0.0136 | 0.0073 |
|  | 0.0068 | 0.0045 | 0.0023 | 0.0023 | 0.0017 | 0.0006 | 0 | 0 | 0 | 0 | 0 |
|  | 0.0056 | 0.0119 | 0.0418 | 0.0927 | 0.0678 | 0.0463 | 0.0407 | 0.0316 | 0.0305 | 0.0203 | 0.0119 |
|  | 0.0102 | 0.0079 | 0.0045 | 0.004 | 0.0051 | 0.0045 | 0.0017 | 0.0011 | 0.0023 | 0.0023 | 0.0034 |
|  | 0.0028 |  |  |  |  |  |  |  |  |  |  |
| 2002 | 1 | 2 | 3 | $\bigcirc$ | 1 | -1 | -1 | 268 | 0.0005 | 0.0022 | 0.0146 |
|  | 0.0304 | 0.0781 | 0.1204 | 0.0765 | 0.0613 | 0.051 | 0.0418 | 0.0385 | 0.0211 | 0.0125 | 0.0108 |
|  | 0.0049 | 0.0043 | 0.0016 | 0.0033 | 0.0005 | 0 | 0.0011 | 0 | 0.0005 | 0 | 0.0011 |
|  | 0.0016 | 0.0179 | 0.0396 | 0.0667 | 0.0743 | 0.0521 | 0.0445 | 0.0233 | 0.0222 | 0.0184 | 0.0152 |
|  | 0.0119 | 0.0081 | 0.0065 | 0.0054 | 0.0016 | 0.0022 | 0.0016 | 0.0016 | 0.0011 | 0.0016 | 0.0033 |
|  | 0.0022 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 2 | 3 | $\bigcirc$ | 1 | -1 | -1 | 261 | 0 | 0.006 | 0.0136 |
|  | 0.0462 | 0.062 | 0.1011 | 0.0989 | 0.0609 | 0.0527 | 0.0326 | 0.0304 | 0.0207 | 0.0141 | 0.0092 |
|  | 0.0038 | 0.0033 | 0.0022 | 0.0011 | 0.0022 | 0 | 0.0005 | 0 | 0 | $\bigcirc$ | 0.0016 |
|  | 0.0168 | 0.0125 | 0.0543 | 0.0554 | 0.0723 | 0.0609 | 0.0321 | 0.0321 | 0.0201 | 0.0168 | 0.0158 |
|  | 0.0109 | 0.0071 | 0.0043 | 0.0065 | 0.0016 | 0.0043 | 0.0016 | 0.0016 | 0.0027 | 0.0016 | 0.0038 |
|  | 0.0016 |  |  |  |  |  |  |  |  |  |  |


| 2004 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 233 | 0.0006 | 0.0116 | 0.0469 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.0451 | 0.084 | 0.0572 | 0.0767 | 0.0682 | 0.0627 | 0.0329 | 0.0238 | 0.0171 | 0.0122 | 0.0073 |
|  | 0.0043 | 0.0024 | 0.0018 | 0.0006 | 0 | 0 | 0.0006 | 0 | 0.0012 | - | 0.0006 |
|  | 0.0079 | 0.0359 | 0.0438 | 0.0688 | 0.0542 | 0.0487 | 0.039 | 0.0451 | 0.0183 | 0.0164 | 0.0152 |
|  | 0.014 | 0.0091 | 0.0049 | 0.0037 | 0.0024 | 0.0024 | 0.0018 | 0.0024 | 0.003 | 0.0006 | 0.0018 |
|  | 0.0024 |  |  |  |  |  |  |  |  |  |  |
| 2005 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 233 | 0 | 0.0037 | 0.0331 |
|  | 0.0893 | 0.06 | 0.1118 | 0.055 | 0.0687 | 0.0537 | 0.0331 | 0.0175 | 0.0162 | 0.0156 | 0.0119 |
|  | 0.0094 | 0.0056 | 0.0019 | 0 | 0.0012 | 0.0012 | 0.0006 | , | 0.0006 | 0.0006 | - |
|  | 0.0012 | 0.0131 | 0.0562 | 0.05 | 0.0687 | 0.045 | 0.0294 | 0.0437 | 0.0231 | 0.0131 | 0. 0144 |
|  | 0.0081 | 0.0112 | 0.0056 | 0.0044 | 0.0025 | 0.0044 | 0.0031 | 0.0012 | 0.0025 | 0.0012 | 0.005 |
|  | 0.0019 |  |  |  |  |  |  |  |  |  |  |
| 2006 | 1 | 2 | 3 | 0 | 1 | -1 | -1 | 212 | 0 | 0.002 | 0.0229 |
|  | 0.0647 | 0.1005 | 0.0742 | 0.062 | 0.0432 | 0.0512 | 0.0297 | 0.0169 | 0.0088 | 0.0108 | 0.0054 |
|  | 0.002 | 0.0047 | 0 | 0.0034 | 0 | 0.0007 | 0 | 0 | - | 0 | - |
|  | 0.0027 | 0.0121 | 0.0512 | 0.0829 | 0.0722 | 0.0762 | 0.0418 | 0.0384 | 0.0236 | 0.0236 | 0.0128 |
|  | 0.0108 | 0.0088 | 0.0088 | 0.0061 | 0.0067 | 0.0061 | 0.0027 | 0.0013 | 0.0007 | 0.0007 | 0.0047 |
|  | 0.002 |  |  |  |  |  |  |  |  |  |  |
| 1983 | 1 | 3 | 3 | $\bigcirc$ | 1 | -1 | -1 | 7 | 0 |  | 0.03 |
|  | 0.03 | 0.1 | 0.07 | 0.09 | 0.03 | 0.03 | 0.01 | 0.01 | 0.02 | 0.01 |  |
|  | $0$ | 0.01 | 0 | 0 | 0 | 0 | 0 | - | - | 0 | - |
|  | $0$ | 0.03 | 0.03 | 0 | 0.05 | 0.03 | 0.02 | 0.07 | 0.05 | 0.02 | 0.01 |
|  | $0.04$ | 0.02 | 0.04 | 0.03 | 0.02 | 0.03 | 0.01 | 0 | 0.02 | 0 | 0.02 |
|  | 0.0 |  |  |  |  |  |  |  |  |  |  |
| 1984 | 1 | 3 | 3 | 0 | 1 | -1 | -1 | 7 | 0 | 0 | 0.0101 |
|  | $0.0404$ | 0.0505 | 0.0808 | 0.0404 | 0.0404 | $0.0202$ | $0.0505$ | 0.0303 | 0.0101 | 0.0202 | $0.0101$ |
|  | 0 | 0.0101 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0.0202 | 0.0303 | 0.0404 | 0.0303 | 0.0808 | 0.0808 | 0.0404 | 0.0202 | 0.0505 | 0.0303 |
|  | 0.0303 | 0.0303 | 0.0101 | 0 | 0 | 0.0101 | 0.0202 | 0.0202 | 0.0101 | 0.0202 | 0.0101 |
|  | 0.0101 |  |  |  |  |  |  |  |  |  |  |
| 1986 | 1 | 3 | 3 | 0 | 1 | -1 | -1 | 99 | - | 0.0057 | 0.0133 |
|  | 0.0267 | 0.0552 | 0.0629 | 0.059 | 0.04 | 0.019 | 0.0686 | 0.0838 | 0.061 | 0.0457 | 0.0362 |
|  | 0.019 | 0.0229 | 0.0114 | 0.0019 | 0 | 0 | 0 | 0 | 0.0019 | 0.0057 | 0 |
|  | 0 | 0.0152 | 0.0171 | 0.04 | 0.04 | 0.0362 | 0.0343 | 0.0057 | 0.0419 | 0.019 | 0.021 |
|  | 0.0114 | 0.0152 | 0.0057 | 0.0095 | 0.0114 | 0.0038 | 0.0057 | 0.0038 | 0.0038 | 0.0019 | 0.0114 |
|  | 0.0114 |  |  |  |  |  |  |  |  |  |  |
| 1987 | 1 | 3 | 3 | 0 | 1 | -1 | -1 | 124 | 0 | 0.0042 | 0.0139 |
|  | 0.0473 | 0.0445 | 0.0612 | 0.0695 | 0.0376 | 0.0362 | 0.0403 | 0.0292 | 0.0389 | 0.0209 | 0.0223 |
|  | 0.0111 | 0.007 | 0.0028 | 0.0042 | 0 | 0.0014 | 0.0028 | - | 0 | - | - |
|  | 0.0042 | 0.0209 | 0.032 | 0.0626 | 0.0584 | 0.0487 | 0.0431 | 0.0292 | 0.0334 | 0.0445 | 0.0292 |
|  | 0.0236 0.0111 | 0.0056 | 0.0139 | 0.0181 | 0.0056 | 0.007 | 0.007 | 0.0028 | 0.0028 | 0 | 0.0083 |
| 1988 | 1 | 3 | 3 | 0 | , | -1 | -1 | 74 | 0 | 0 | 0.0313 |
|  | 0.0264 | 0.0553 | 0.0625 | 0.0721 | 0.0673 | 0.0769 | 0.0505 | 0.0337 | 0.024 | 0.024 | 0.0144 |
|  | 0.0096 | 0.0048 | 0.0024 | 0 | 0 | 0.0024 | 0 | - | 0.0024 | - | 0 |
|  | 0.0024 | 0.0264 | 0.0216 | 0.0481 | 0.0457 | 0.0457 | 0.0505 | 0.0361 | 0.0313 | 0.0216 | 0.0192 |
|  | 0.0168 | 0.0096 | 0.012 | 0.0048 | 0.0096 | 0.0144 | 0.0024 | 0.0024 | 0.0024 | 0.0072 | 0.0048 |
|  | 0.0072 |  |  |  |  |  |  |  |  |  |  |
| 1989 | 1 | 3 | 3 | 0 | 1 | -1 | -1 | 53 | 0 | 0.0034 | 0.0135 |
|  | 0.064 | 0.0505 | 0.0471 | 0.0539 | 0.0438 | 0.0606 | 0.0471 | 0.0337 | 0.0168 | 0.0236 | 0.0135 |
|  | 0.0067 | 0.0101 | 0.0067 | 0.0067 | 0 | 0.0101 | 0 | 0.0034 | - | - | 0 |
|  | 0.0067 | 0.0168 | 0.0438 | 0.0303 | 0.0505 | 0.0673 | 0.0572 | 0.0236 | 0.0438 | 0.037 | 0.0236 |
|  | 0.0135 | 0.0101 | 0.0101 | 0.0101 | 0.0101 | 0.0034 | 0.0034 | 0.0067 | 0.0034 | 0.0034 | 0.0101 |
|  | 0.0034 |  |  |  |  |  |  |  |  |  |  |
| 1990 | 1 | 3 | 3 | 0 | 1 | -1 | -1 | 21 | 0.008 | 0.12 | 0.176 |
|  | 0.032 | 0.064 | 0.016 | 0.032 | 0.048 | 0.032 | 0.008 | 0.016 | 0.008 | 0.008 | 0.024 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.016 |
|  | 0.032 | 0.112 | 0.064 | 0.048 | 0.032 | 0.04 | 0.024 | 0 | 0 | 0.008 | 0.008 |
|  | $\begin{aligned} & 0.008 \\ & 0.008 \end{aligned}$ | 0 | 0 | 0 | 0 | 0 | 0.008 | 0 | 0 | 0 | - |
| 1991 | 1 | 3 | 3 | 0 | 1 | -1 | -1 | 88 | 0.0021 | 0.0042 | 0.04 |
|  | 0.0989 | 0.0926 | 0.0905 | 0.0463 | 0.0379 | 0.0358 | 0.0232 | 0.0211 | 0.0021 | 0.0042 | 0.0105 |
|  | 0.0021 | 0 | 0 | 0 | 0.0021 | 0 | 0 | 0 | 0.0021 | 0 | 0.0021 |
|  | 0.0211 | 0.0337 | 0.0674 | 0.0947 | 0.0905 | 0.0295 | 0.0337 | 0.0147 | 0.0189 | 0.0168 | 0.0168 |
|  | $\begin{aligned} & 0.0105 \\ & 0.0084 \end{aligned}$ | 0.0063 | 0.0021 | 0 | 0.0021 | 0.0021 | 0.0042 | 0.0042 | 0.0042 | 0 | 0.0021 |
| 1992 | 1 | 3 | 3 | 0 | 1 | -1 | -1 | 49 | 0.0037 | 0.0147 | 0.0586 |
|  | 0.0476 | 0.0989 | 0.1209 | 0.0879 | 0.0293 | 0.022 | 0.0256 | 0.0293 | 0.0037 | 0.0073 | 0.0037 |
|  | 0.0037 | 0 | 0 | 0 | 0 | 0.0037 | 0 | 0.0037 | 0 | 0 | 0.0037 |
|  | 0.0073 | 0.0476 | 0.0623 | 0.0586 | 0.0513 | 0.0366 | 0.033 | 0.0183 | 0.0147 | 0.033 | 0.0073 |


|  | $\begin{aligned} & 0.0073 \\ & 0 \end{aligned}$ | 0.0147 | 0.0073 | 0.0073 | 0 | 0.011 | 0.0037 | 0.0037 | 0 | 0.0073 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 1 | 3 | 3 | 0 | 1 | -1 | -1 | 58 | 0 | 0.0062 | 0.0341 |
|  | 0.1053 | 0.096 | 0.0805 | 0.1053 | 0.0402 | 0.0341 | 0.0155 | 0.0124 | 0.0031 | 0.0031 | 0 |
|  | 0 | 0.0062 | 0.0031 | 0 | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 | 0 |
|  | 0.0031 | 0.0217 | 0.0836 | 0.065 | 0.1022 | 0.0433 | 0.0433 | 0.0279 | 0.0186 | 0.0031 | 0.0217 |
|  | 0.0031 | 0.0031 | 0.0062 | 0 | 0.0031 | 0.0031 | 0 | 0 | 0 | 0 | 0.0031 |
|  | 0.0031 |  |  |  |  |  |  |  |  |  |  |
| 1994 | 1 | 3 | 3 | 0 | 1 | -1 | -1 | 44 | 0.004 | 0.04 | 0.084 |
|  | 0.084 | 0.16 | 0.052 | 0.056 | 0.016 | 0.016 | 0.004 | 0.008 | 0.008 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.008 |
|  | 0.008 | 0.064 | 0.048 | 0.076 | 0.08 | 0.048 | 0.048 | 0.028 | 0.02 | 0.008 | 0.004 |
|  | 0.012 | 0 | 0.008 | 0.008 | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 | $\bigcirc$ |
|  | $\bigcirc$ |  |  |  |  |  |  |  |  |  |  |
| 1995 | 1 | 3 | 3 | 0 | 1 | -1 | -1 | 40 | 0 | 0.0045 | 0.0625 |
|  | 0.1161 | 0.0938 | 0.1205 | 0.0536 | 0.0223 | 0.0045 | 0.0045 | 0.0089 | 0.0045 | 0 | 0 |
|  | 0 | 0.0045 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0.0179 | 0.0982 | 0.1116 | 0.067 | 0.0446 | 0.0446 | 0.0402 | 0.0268 | 0 | 0.0134 | 0.0045 |
|  | $\begin{aligned} & 0.0045 \\ & 0 \end{aligned}$ | 0 | 0.0134 | 0 | 0.0045 | 0 | 0 | 0 | 0.0045 | 0.0045 | 0 |
| \# |  |  |  |  |  |  |  |  |  |  |  |
| \# | Mean | Size | at | Age |  |  |  |  |  |  |  |
| 2 |  |  |  |  |  |  |  |  |  |  |  |
| \#_Year | Season | Type | Gender | Partiti | on | Age-Err | Nsamp | 3 | 4 | 5 | 6 |
|  | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 |
|  | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 30 | 3 | 4 |
|  | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
|  | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 30 |
| \#1986 | 1 | 2 | 3 | 0 | 1 | 1 | 27.5009 | 6098 | 31.2838 | 9375 |  |
|  | 34.4412 | 3171 | 37.0764 | 3134 | 39.2758 | 402 | 41.1115 | 2629 | 42.6436 | 3953 |  |
|  | 43.9223 | 8267 | 44.9896 | 5626 | 45.8804 | 3165 | 46.6238 | 9688 | 47.244 | 1306 |  |
|  | 47.7623 | 1264 | 48.1945 | 6563 | 48.5553 | 3566 | 48.8564 | 4408 | 49.1077 | 5731 |  |
|  | 49.3175 | 1013 | 49.4925 | 755 | 49.6386 | 8979 | 49.7606 | 4073 | 49.8624 | 2428 |  |
|  | 49.9473 | 7558 | 50.0182 | 7824 | 28.1226 | 401 | 31.4506 | 9993 | 34.2013 | 0247 |  |
|  | 36.4746 | 4329 | 38.3535 | 329 | 39.9064 | 1319 | 41.1898 | 5059 | 42.2505 | 9659 |  |
|  | 43.1272 | 9075 | 43.8518 | 6823 | 44.4507 | 2304 | 44.9456 | 6953 | 45.3547 | 3703 |  |
|  | 45.6928 | 2652 | 45.9722 | 5354 | 46.2031 | 967 | 46.3940 | 6852 | 46.5518 | 2185 |  |
|  | 46.6822 | 0313 | 46.7899 | 6174 | 46.8790 | 2296 | 46.9526 | 3103 | 47.0134 | 6723 |  |
|  | 47.0637 | 4764 |  |  |  |  |  |  |  |  |  |
| \# | 0 | 0 | 1 | 9 | 17 | 27 | 28 | 27 | 13 | 19 | 8 |
|  | 8 | 5 | 2 | 5 | 5 | 2 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | $\bigcirc$ | 3 | 7 | 26 | 32 | 18 | 16 | 8 |
|  | 10 | 7 | 4 | 7 | 1 | 0 | 1 | 1 | 1 | 1 | 1 |
|  | 0 | 0 | 1 | 0 |  |  |  |  |  |  |  |
| \#1986 | 1 | 1 | 3 | 0 | 1 | 1 | 28 | 31.5 | 37 | 41.2 | 42.1 |
|  | 44.6 | 46.9 | 49 | 49.6 | 48.4 | 48.9 | 48.1 | 50.6 | 52 | 52.2 | 50 |
|  | 49 | 51 | 51 | 51 | 51 | 51 | 51 | 51 | 28 | 31.5 | 42 |
|  | 40.9 | 41.4 | 42.2 | 43.6 | 45.1 | 45.1 | 47.1 | 46.6 | 47 | 46.7 | 46 |
|  | 47.3 | 46 | 49 | 52 | 46 | 50 | 47.3 | 47.3 | 49 | 47.3 |  |
| \# | 0 | 0 | 1 | 9 | 17 | 27 | 28 | 27 | 13 | 19 | 8 |
|  | 8 | 5 | 2 | 5 | 5 | 2 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 3 | 7 | 26 | 32 | 18 | 16 | 8 |
|  | 10 | 7 | 4 | 7 | 1 | 0 | 1 | 1 | 1 | 1 | 1 |
|  | 0 | 0 | 1 | 0 |  |  |  |  |  |  |  |
| \#1987 | 1 | 1 | 3 | 0 | 1 | 1 | 28 | 31.5 | 44 | 41.3 | 47.1 |
|  | 46.5 | 47 | 48 | 47.8 | 50.4 | 50.4 | 50.4 | 51.2 | 51.6 | 53.8 | 56 |
|  | 51 | 56 | 54.5 | 53 | 58 | 60 | 50 | 50 | 28 | 31.5 | 36 |
|  | 41.3 | 43.4 | 43.5 | 44.3 | 45.4 | 44.2 | 45.8 | 46.1 | 46.9 | 47.1 | 49 |
|  | 48.8 | 48 | 47 | 51 | 51 | 47.3 | 47.3 | 53 | 52 | 53 |  |
| \# | 0 | 0 | 1 | 3 | 10 | 29 | 34 | 23 | 27 | 26 | 20 |
|  | 13 | 14 | 5 | 5 | 1 | 0 | 4 | 2 | 2 | 1 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 4 | 8 | 28 | 35 | 20 | 15 |
|  | 22 | 13 | 7 | 8 | 6 | 4 | 1 | 2 | 1 | 1 | 0 |
|  | 0 | 1 | 3 | 1 |  |  |  |  |  |  |  |
| \#1984 | 1 | 2 | 3 | 0 | 1 | 1 | 28 | 32.3 | 34.4 | 36.5 | 38.9 |
|  | 40.6 | 42.8 | 43.2 | 45.9 | 46 | 48.4 | 48.7 | 49.9 | 48.5 | 52.4 | 51.3 |
|  | 55.9 | 50 | 41 | 47.1 | 50 | 50 | 50 | 50 | 28 | 30.3 | 36.1 |
|  | 37.1 | 39.1 | 40.2 | 40.7 | 42.4 | 42.6 | 43.4 | 44.4 | 45.8 | 47.3 | 46.7 |
|  | 46.5 | 45.2 | 47.7 | 44.9 | 51.6 | 51.1 | 45.3 | 49 | 49.5 | 50.3 |  |


| \# | 0 | 6 | 26 | 62 | 75 | 44 | 42 | 61 | 34 | 20 | 17 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 8 | 9 | 9 | 5 | 5 | 2 | 0 | 1 | 2 | 0 | 0 |
|  | 0 | 0 | 0 | 2 | 2 | 21 | 24 | 23 | 29 | 34 | 24 |
|  | 12 | 11 | 10 | 14 | 9 | 12 | 2 | 3 | 2 | 6 | 3 |
|  | 6 | 4 | 5 | 4 |  |  |  |  |  |  |  |
| \#1993 | 1 | 2 | 3 | 0 | 1 | 1 | 28 | 32.6 | 33.9 | 36.5 | 39 |
|  | 40.6 | 41.7 | 44.3 | 44.5 | 47 | 49.7 | 49.8 | 47.2 | 46.8 | 50 | 47 |
|  | 51.8 | 55 | 52 | 50 | 50 | 50 | 54 | 50 | 28 | 35.4 | 33.1 |
|  | 36.4 | 38 | 38.6 | 40.8 | 41.5 | 43.6 | 44.1 | 45.3 | 44.5 | 46.2 | 45.2 |
|  | 45.9 | 50.5 | 45.3 | 48.3 | 46.7 | 47.3 | 47.3 | 48 | 48 | 54 |  |
| \# | 0 | 7 | 34 | 69 | 66 | 85 | 85 | 48 | 35 | 11 | 15 |
|  | 10 | 12 | 6 | 2 | 2 | 4 | 1 | 1 | 0 | 0 | 0 |
|  | 1 | 0 | 0 | 7 | 20 | 57 | 46 | 69 | 50 | 30 | 17 |
|  | 16 | 11 | 2 | 5 | 6 | 10 | 2 | 3 | 3 | 3 | 0 |
|  | 0 | 1 | 3 | 1 |  |  |  |  |  |  |  |
| 2001 | 1 | 2 | 3 | 0 | 1 | 1 | 28 | 34.2 | 35.8 | 36.8 | 38.2 |
|  | 39.7 | 40.2 | 41.1 | 42.6 | 43.8 | 43.8 | 45.3 | 45.6 | 47.1 | 44.1 | 48.4 |
|  | 46 | 42.5 | 43.3 | 44 | 50 | 50 | 50 | 50 | 28 | 32.3 | 33.5 |
|  | 36.4 | 37.5 | 38.5 | 39.3 | 40.3 | 41.3 | 42.4 | 42.3 | 42.1 | 41.8 | 42.9 |
|  | 43.9 | 44 | 44.7 | 45.5 | 42.3 | 44 | 47.3 | 46.8 | 44.5 | 51 |  |
|  | 0 | 6 | 25 | 71 | 190 | 162 | 98 | 79 | 96 | 84 | 64 |
|  | 28 | 24 | 13 | 12 | 8 | 4 | 4 | 3 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 10 | 21 | 74 | 164 | 120 | 82 | 72 | 56 |
|  | 54 | 36 | 21 | 18 | 14 | 8 | 7 | 9 | 8 | 3 | 2 |
|  | 4 | 4 | 6 | 5 |  |  |  |  |  |  |  |
| 2002 | 1 | 2 | 3 | 0 | 1 | 1 | 25 | 31 | 35.5 | 37.6 | 39.1 |
|  | 40.8 | 41.6 | 42.6 | 43.8 | 44.6 | 45.1 | 45.4 | 46.8 | 46.3 | 44.4 | 46.9 |
|  | 49.3 | 49.3 | 47 | 50 | 45 | 50 | 48 | 50 | 27 | 30 | 34.3 |
|  | 36.4 | 38.2 | 39.4 | 39.5 | 41.1 | 41.1 | 41.6 | 42.5 | 42.3 | 43.4 | 43.1 |
|  | 44.4 | 44.7 | 44 | 45.8 | 44 | 43.7 | 41 | 45.7 | 46.3 | 48.5 |  |
|  | 1 | 4 | 27 | 56 | 144 | 222 | 141 | 113 | 94 | 77 | 71 |
|  | 39 | 23 | 20 | 9 | 8 | 3 | 6 | 1 | 0 | 2 | 0 |
|  | 1 | 0 | 2 | 3 | 33 | 73 | 123 | 137 | 96 | 82 | 43 |
|  | 41 | 34 | 28 | 22 | 15 | 12 | 10 | 3 | 4 | 3 | 3 |
|  | 2 | 3 | 6 | 4 |  |  |  |  |  |  |  |
| \#1987 | 1 | 3 | 3 | 0 | 1 | 1 | 28 | 35.7 | 35.3 | 36.6 | 38.8 |
|  | 41.2 | 43.2 | 44.4 | 45 | 46.6 | 47.7 | 48.6 | 48.9 | 51.6 | 50.5 | 50.8 |
|  | 50.5 | 52 | 50 | 57 | 51 | 50 | 50 | 50 | 28 | 30.7 | 34.4 |
|  | 37.9 | 38.6 | 40.3 | 41.5 | 41.6 | 44 | 45 | 45.5 | 45.9 | 44.6 | 48.3 |
|  | 46.9 | 49.4 | 50.3 | 47 | 47.6 | 49 | 52.5 | 47.3 | 49.43 | 51 |  |
| \# | 0 | 3 | 10 | 34 | 32 | 44 | 50 | 27 | 26 | 29 | 21 |
|  | 28 | 15 | 16 | 8 | 5 | 2 | 3 | 0 | 1 | 2 | 0 |
|  | 0 | 0 | 0 | 3 | 15 | 23 | 45 | 42 | 35 | 31 | 21 |
|  | 24 | 32 | 21 | 17 | 4 | 10 | 13 | 4 | 5 | 5 | 2 |
|  | 2 | $\bigcirc$ | 6 | 3 |  |  |  |  |  |  |  |
| \#1991 | 1 | 3 | 3 | 0 | 1 | 1 | 29 | 32 | 34.1 | 36.1 | 39 |
|  | 41.4 | 42.7 | 44.9 | 46.8 | 47.5 | 50.5 | 52 | 50.8 | 49.3 | 54 | 50 |
|  | 50 | 50 | 47 | 50 | 50 | 50 | 51 | 50 | 34 | 30.8 | 33.8 |
|  | 35.7 | 38.4 | 40.9 | 41.8 | 44.5 | 45.6 | 46.5 | 45.6 | 45.8 | 45.2 | 49.6 |
|  | 47.5 | 48 | 51 | 49.2 | 48.3 | 49.3 | 49.5 | 50 | 50.7 | 47.3 |  |
| \# | 1 | 2 | 19 | 49 | 49 | 48 | 27 | 21 | 24 | 19 | 13 |
|  | 7 | 5 | 7 | 3 | 0 | 0 | 0 | 1 | 0 | $\bigcirc$ | 0 |
|  | 1 | 0 | 1 | 10 | 16 | 33 | 50 | 51 | 19 | 27 | 12 |
|  | 17 | 14 | 9 | 6 | 7 | 2 | 2 | 1 | 5 | 4 | 6 |
|  | 4 | 1 | 3 | 3 |  |  |  |  |  |  |  |
| \# | Environmental Data |  |  |  |  |  |  |  |  |  |  |
| 0 |  |  |  |  |  |  |  |  |  |  |  |
| 0 |  |  |  |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |
| 999 |  |  |  |  |  |  |  |  |  |  |  |


[^0]:    Note: Square indcates the Base Model and bold font indicates best fit.

