

JUVENILE SALMONID DIRECT SURVIVAL/INJURY IN PASSAGE THROUGH THE ICE HARBOR DAM SPILLWAY, SNAKE RIVER

Contract No. DACW68-02-D-0002 Task Order 0006

April 2004

NORMANDEAU ASSOCIATES
ENVIRONMENTAL CONSULTANTS

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Prepared for

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EXECUTIVE SUMMARY

Releases of HI-Z tagged (HI-Z tag or balloon tag) chinook salmon smolts, *Oncorhynchus tshawytscha*, (and rainbow trout, *O. mykiss*, in summer) were made into Spillbay 5 and downstream of Spillbay 1 (spring controls) and through the juvenile fish facility bypass pipe (spring and summer controls) at Ice Harbor Dam from 23 April to 2 May 2003 (spring) to (1) estimate survival (direct effects) within $\leq \pm 3\%$, 90% of the time, in passage at shallow (7 ft above ogee) and deep (3 ft above ogee) sites at 50% spill, 100% spill, and a special spill pattern, and (2) better understand the injury mechanisms to assist in possible spillway/deflector modifications for enhanced fish survival. In summer (July 2003) direct effects of dispersed and bulk spill volumes were estimated; fish were released at a deep release site (dispersed spill) and at shallow and deep sites at a bulk spill pattern. The resulting survival estimates from these releases were to be within $\leq +5\%$, 90% of the time. Differences in survival or clean fish estimates between treatment conditions were statistically tested *a posteriori*; the initial study was not designed to detect statistical differences.

Average total fish length in spring was about 141 mm and 123 mm in summer. Water temperatures ranged from 10.0 to 11.5°C (50.0 to 52.7°F) in spring and from 19.5 to 20.5°C (67.1 to 68.9°F) in summer. Estimated impact velocities encountered by released fish in spring ranged from 70.2 to 71.6 ft/s; in the summer, impact velocities ranged from 70.1 to 73.2 ft/s. Laboratory studies suggest impact velocity exceeding about 58 ft/s may inflict injury/mortality to fish.

Recapture rates (physical retrieval of alive and dead fish) were higher in April and May than for the July releases. Recapture rates for treatment groups in April and May were 99.7 to 100%; for the controls recapture rates were 99.3%. Recapture rates of chinook salmon treatment groups in July ranged from 90.8% (bulk spill, shallow release) to 97.0% (dispersed spill, deep release); recapture rates for the two rainbow trout treatment groups (n=20 for each) ranged from 95.0% (bulk spill, shallow release) to 100% (bulk spill, deep release). July recapture rate for the control group released through the bypass pipe was 99.2%.

Estimated immediate survival rates (1 h) were higher in April and May (all ≥98.7%) than in July. For the summer releases they ranged from 88.9 (bulk spill, shallow release) to 97.8% (dispersed spill, deep release). The 48 h survival estimates for spring releases also exceeded 98% (range 98.7 to 99.0%) with little difference either between spill patterns or between release sites. Unacceptable holding mortality in both the treatment (32 to 42%) and control (about 31%) groups precluded reliable estimation of 48 h survival.

The pre-specified precision (ε) level of \leq +3%, 90% of the time was met on all survival estimates in spring; the pre-specified precision (within $\leq \pm 5\%$) was attained for two of the three treatment groups (shallow release at bulk spill and deep release at dispersed spill) in the summer investigation. Bacterial infections at the tag insertion site or fungal infections on gills were observed on fish in holding in July and were likely exacerbated by the higher summer water temperatures (19.5 to 20.5°C or 67.1 to 68.9 \degree F). Smaller fish size in July may have also affected the physical recapture rates to a certain extent due to the higher propensity for tag dislodgment (fish assumed dead) on smaller sized fish.

Survival estimates derived from detection of PIT-tagged fish released through Spillbays 1 and 3 at the same time in summer were reported at 96%; interestingly, these estimates are higher than the summer direct estimates at the bulk spill (88.9 to 92.3%; pooled 90.1%) and counterintuitive because estimates derived from the PIT tag-recapture technique contain the direct (immediate) and indirect (occurring over a longer time and distance). The difference between the two estimates at summer dispersed spill condition is small in favor of the direct estimate (97.8% for direct and 96% for PIT

tagged fish). Estimates from our investigation represent only the direct effects of passage. However, the proportion of the PIT tagged fish injured during passage is unknown.

"Clean fish" estimates (CFE) differed between spill patterns and passage locations within the spillbay. With respect to spill patterns, CFE for deep released fish (79.3%) at 50% spill pattern was significantly lower (P<0.05) than at 100% spill (87.8%) in spring. However, CFEs for deep released fish at special spill (78.3%) and 100% spill (87.8%) were not significantly different (P>0.05); the sample size was too small for this test to detect a difference. With respect to release locations within each spill pattern in spring, significant differences $(P<0.05)$ were noted with deep releases having lower CFEs at both 100% spill and bulk spill patterns than shallow released fish; the magnitude of difference between release locations was similar (about 12%) at both spill patterns.

The effect of passage location was also evident in the July test. CFE for deep released fish (81.3%) was significantly lower (P<0.05) than for shallow released fish (95.1%) at the bulk spill pattern; the only spill pattern at which fish were released at two locations. CFE for deep released fish at the dispersed spill pattern was estimated at 79.6%, comparable to that at the bulk spill pattern. Except for one clean fish estimate, 97.4% , in spring (shallow release site at 100% spill) the other seven estimates were lower than those reported for the summer released PIT tagged fish (96%); it appears that all maladies observed in our investigation were not lethal or the two groups of fish traveled different paths in exiting the spillbays.

Spill volume and passage location appeared to affect injury types and rates. In spring, higher injury rates (20.7%) occurred at the 50% spill volume (3.4 to 5.1 kcfs through Spillbay 5) than at the 100% spill, 8.5% (4.25 to 8.5 kcfs through Spillbay 5). At 100% spill volume, only 1.3% of the shallow pipe released fish were injured compared to 12.2% for the deep release pipe. For the special spill (8.5 kcfs spill through Spillbay 5), injury rates were 8.3 and 21.7% for the shallow and deep release pipes, respectively. In the summer, injury rate was high (18.4 to 20.0%) for fish released via the deep pipe at both bulk and dispersed spill; however it was only 4.9% for fish passed by the shallow pipe at dispersed spill. The differential injury rates relative to release location and spill volume through the test spillbay may be related to the unique hydraulic characteristics of the jet as it enters the tailrace.

The predominant injury type was hemorrhaged or damaged eye. The causative mechanisms for the observed injury types may have resulted from shear forces in the vicinity of the flow deflector. Fish passed via the deep pipe were much more likely to sustain eye damage (13% deep pipe, 2% shallow).

It is unknown whether the passage-related maladies are lethal on a long-term basis. Because eye damage was the dominant injury type, information is needed on the persistence and severity of this injury type on post spillway passage survival of emigrating fish, effects on vision, predator avoidance and feeding success. The finding that fish passed close to the ogee (deep release pipe) suffered more injuries regardless of spill rate also needs to be integrated with the proportion of the naturally emigrating population that would be subjected to the hydraulic conditions experienced by these fish. Preliminary hydroacoustic data from Ice Harbor in 2003 suggest that most fish pass higher above the ogee; however, at low tainter gate openings (2 to 3 ft) fish do not have much of an opportunity to orient higher in the discharge jet. Thus, at lower gate openings the fish may have a greater chance of being exposed to injurious hydraulic conditions particularly in the vicinity of the flow deflector. The concurrent sensor fish data collected at Ice Harbor indicate that injurious hydraulic conditions do exist in the immediate vicinity of the flow deflector especially for sensors passed via the deep pipe. Further research is needed to ascertain the relationship between potentially injurious hydraulic conditions and depth of sensor fish upon passing the flow deflector and submergence depth of the flow deflector.

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1.0 INTRODUCTION

Juvenile salmonids on their seaward journey encounter any or all of the following exit routes at hydro dams: turbines, spillways, and bypasses. There are two inter-related concerns associated with passage through any of these routes for overall survival. One is the proportion of fish utilizing any of these routes during emigration and the other is their subsequent post-passage condition and survival. Spill of varying magnitude and duration is used at most hydro dams on the Columbia River Basin to enhance passage effectiveness and overall survival of juvenile salmonids (Wilson *et al.* 1991). However, spill is expensive in terms of lost power generation and with some spillway configurations and flow patterns, potentially lethal levels of total dissolved gas (TDG) in the river can result. To alleviate the TDG supersaturation levels at Ice Harbor Dam (Ice Harbor) on the lower Snake River (Figure 1-1), flow deflectors were installed downstream of the spillway at Spillbays 1 through 10. Flow deflectors are concrete sills installed on the downstream face of a spillway to maximize the surface skimming effect of spilled water and prevent plunging to the bottom of the stilling basin thus reducing the pressure gradient that forces atmospheric gases into the solution; the installation depths of the flow deflectors varies with sites. Due to a variety of factors (*e.g.*, spill volume, spill pattern, passage location within a spillbay, and flow deflector elevation), fish passage survival and injury rates in passage through spillbays equipped with flow deflectors have not been ≥98% at all of the hydroelectric dams on the Columbia River Basin (Normandeau Associates *et al.* 2003a). Consequently, there is a need for a better understanding of mechanisms influencing passage survival of juvenile salmonids so that appropriate mitigative measures, if needed, can be implemented.

Some recent spillway passage studies on juvenile salmonids at Ice Harbor utilizing PIT tags and radio tags have indicated variation between seasons and years; passage survival estimates were approximately 98% in the spring of 1999 but lower in the summer of 1999 and spring of 2002 (Absolon *et al.* 2003; Eppard *et al.* 2003). Powerhouse survival (turbine and bypass combined) was estimated to be 98% in 2001. Due to low river flows in 2001, spillway operations were constrained. However, mechanisms for lower survival at Ice Harbor could not be delineated by these studies. It is thought that the survival of fish passing through the spillway at Ice Harbor may be linked to the relationship between the spillway deflector elevation and the tailwater elevation. Tailwater elevation is dependent primarily upon project discharge at Ice Harbor.

Surface oriented salmonid emigrants generally occupy the top 20 ft of the water column and must sound to deeper depths to exit the Ice Harbor spillway equipped with bottom opening tainter gates. The height of the bottom opening and associated spill volume create hydraulic passage conditions which may affect fish survival and injury rates.

The primary objective of the present study was to obtain estimates of direct spillway passage survival of juvenile salmon within $\leq \pm 3\%$, 90% of the time, at two spill volumes (50 and 100% of the river flow) under typical spring high tailwater condition. Fish were introduced at two depths in Spillbay 5, shallow (7 ft above the ogee) and deep (3 ft above the ogee). Additionally, a smaller scale experiment was conducted at a special spill pattern. Results from the spring study elicited a secondary experiment in summer (low tailwater) to obtain survival estimates within $\pm 5\%$, 90% of the time at two additional spill conditions, bulk spill and dispersed spill. An additional objective was to determine the extent and types of injuries incurred by spillway passed juvenile salmon under spring and summer tailwater conditions.

Some recent studies have indicated that the use of the HI-Z tag-recapture (HI-Z tag or balloon tag) technique (Heisey *et al.* 1992, 2002a) can provide useful information on potential mechanisms influencing the direct effects of passage through various exit routes at hydroelectric dams.

1.1 Project Description

Ice Harbor is the first dam on the Snake River upstream of the confluence with the Columbia River (Figures 1-1 and 1-2). It has six Kaplan type turbine units and 10 spillway bays, along with a navigation lock and an earthfill section. The stilling basin for the spillway is 59 ft long and 168 ft wide with a floor elevation of 304 ft msl. Each spillbay is equipped with a 53 ft high and 50 ft wide tainter gate that seals at an elevation of 389 ft msl. The crest elevation for the spillway is 391 ft msl. Standard-length submersible traveling screens (STSs) are present in all turbine intake bays. The present investigation was conducted at Spillbay 5.

In order to reduce the level of TDG supersaturation produced by water passing over a spillbay, the Corps had installed flow deflectors on the downstream face of all spillbays that direct the flow along the surface of the tailrace rather than allowing it to plunge to the bottom of the spilling basin; these flow deflectors are located at an elevation of 338 ft msl on Spillbays 2 through 9 and at an elevation of 334 ft msl at the outer spillbays. The shallow and deeper flow deflectors are submerged 7 and 11 ft, respectively, below a typical springtime tailwater elevation of 345 ft msl (Figures 1-1 and 1-2).

A 12 ft high concrete end sill and 8 ft high by 10 ft wide dentates extends across the spillway to aid in energy dissipation (Figure 1-2). These structures are submerged approximately 29 and 33 ft during the typical springtime tailwater elevation (345 ft msl). The end sill and dentates are approximately 170 and 128 ft downstream of the face of the spillway, respectively.

The Ice Harbor tailwater elevation is directly dependent upon river flow (project discharge). At low discharge (less than 50k cfs), the tailwater is generally below 341 ft. The tailwater is generally above 345 ft at discharges greater than 100k cfs. The normal operating range of the tailrace elevation is 339.5 to 345 ft.

2.0 STUDY DESIGN

There are two primary components which affect fish using any exit route: direct and indirect effects. Direct effects are manifested immediately after passage (*e.g.*, instantaneous fish mortality, injury, loss of equilibrium); indirect effects (*e.g.*, predation, disease, physiological stress) may occur over an extended period or distance after passage. The present study was designed to estimate the direct effects. Spillway fish passage survival and condition was measured by a straightforward approach of introducing a known number of HI-Z tagged alive fish into Spillbay 5 (treatment) and downstream of the spillway (control), recapturing them immediately after passage, enumerating the alive and dead fish, and then carefully examining the condition of each fish. The sample sizes needed to estimate survival within a prespecified precision (ϵ) level were based on estimating the direct effects of passage.

Treatment fish were released either 3 ft or 7 ft above the ogee and approximately 16 ft upstream of the tainter gate (Figures 2-1 and 2-2). Control fish were released to assess the effects of handling, transport, tagging, release, and recapture. Controls were released either downstream of Spillbay 1 (spring) or via the juvenile facility fish bypass pipe in spring (April and May) and summer (July) (Figure 1-1). Table 2-1 provides the daily fish release schedule.

Ambient river temperature during the study ranged from 10.0 to 11.5° C (50.0 to 52.7°F) in April/May and 19.5 to 20.5°C (67.1 to 68.9°F) in July (Table 2-1). Table 2-2 provides total project discharge and spill volumes at Spillbay 5 and over the entire spillway during the study. During fish releases, spill volume through Spillbay 5 ranged from 3.4 to 8.5k cfs in spring and from 3.4 to 13.6k cfs in summer.

Total project spill between the two periods ranged from 31.0 to 79.0k cfs during the spring experiment and 13.6 to 45.0k cfs in summer (Table 2-2). The respective total project discharge ranged from 60.6 to 101.4k cfs and 14.4 to 56.1k cfs. The 100% spill condition, in reality, was less because the Project required the operation of at least one turbine. Tailwater elevation ranged from 342.2 to 345.9 ft msl in spring and 338.2 to 341.1 ft msl in summer. Thus, the respective submergence depth of the flow deflector ranged from 4 to 8 ft and 0 to 3 ft. The characteristics of the discharge jet upon interception of the flow deflector are shown in Figure 2-3. Forebay elevations ranged from 438.0 to 438.9 ft in spring and 438.2 to 439.4 ft in summer (Table 2-2). The respective net heads were 93.5 to 96.5 and 97.7 to 100.3 ft. Hourly values for the above project operations data are presented in Appendix Table A-1.

Three spill conditions were tested in spring and two in summer (Appendix B). The spring conditions were designated 50%, 100%, and special spill. The summer conditions were designated bulk spill and dispersed spill. During the 50%, 100%, special, and dispersed spill conditions, the spill volume was spread nearly uniformly across all spillbays (Table 2-3). During the bulk spill, Spillbay 5 was set at 8 stops (13.6k cfs) and the most of the spill was passed via this spillbay.

2.1 Sample Size Requirement

Prior to initiating the fish survival investigation at Ice Harbor Dam, the sample size requirement was determined to fulfill the primary objective of the study: achieving a prespecified precision (ϵ) level (within $\leq \pm 0.03$ (April and May) or $\leq \pm 0.05$ (July), 90% of the time) on the individual estimates of passage survival ($\hat{\tau}$). The sample size is a function of the recapture rate (P_A), expected passage survival $(\hat{\tau})$ or mortality $(1-\hat{\tau})$, survival of control fish (S), and the desired precision (ϵ) at a given probability of significance (α) . In general, sample size requirements decrease with an increase in control survival and recapture rates. Only precision (ϵ) and α levels can be strictly controlled by an investigator. The expression to calculate sample sizes for achieving a prespecified precision (ε) level is given in Mathur *et al.* (1996).

In performing the sample size calculations, we assumed capture data from replicate releases could be pooled (*i.e.*, natural variability $\sigma_r^2 = 0$). We calculated that with the following assumptions: a recapture rate of 0.98, control survival rate (S) of 0.99, and spillbay survival $(\hat{\tau})$ of 0.97, a precision (ε) level of $\leq \pm 0.03$, 90% of the time might be achievable with releasing 264 fish per treatment; however, only 95 fish per treatment are needed at a precision (ε) level of $\leq \pm 0.05$ (Table 2-4).

Based on the results of several recent spillbay survival experiments from other sites on the Columbia River Basin (Table 2-5), a sample size of approximately 250 and 150 fish per treatment release was deemed sufficient to attain the two prespecified precision levels (ϵ) of $\leq \pm 0.03$ and $\leq \pm 0.05$, 90% of the time. Given the above assumptions, the projected number of fish allocated for the April/May and July study period were 945 (655 treatment and 290 controls) and 450 (300 treatment, 150 controls), respectively.

Past experience suggests that the sample sizes can be adjusted as a study progresses because the statistical results are available daily. If recapture and control survival rates are higher than initially assumed, sample size can be reduced. Conversely, if the values of these parameters are lower than initially assumed, then sample size can be increased to achieve the pre-specified statistical precision. However, under certain extenuating circumstances (*e.g.*, time, fish availability, or desired test condition constraints) sample size adjustments may not always be possible during the course of an experiment. Indeed, the full initial allocation of fish for the spring study was not utilized because the precision (ϵ) objective had been achieved by a release of fewer fish. Consequently, this allowed

releasing the remainder of the fish to assess another spill condition (special spill pattern), *albeit* on a smaller scale.

2.2 Source and Maintenance of Specimens

Juvenile chinook salmon smolts used in the spring study were obtained from the Leavenworth National Fish Hatchery, Washington. Fish for the summer experiment were collected at John Day Dam. Some 1,208 fish were transported from the hatchery and approximately 475 fish from John Day Dam via truck to a 600 gal circular holding pool on the tailrace deck at Ice Harbor Dam (Table 2-6). Due to the limited supply of in-river fish (475) during the July experiment, the summer tests were supplemented with 500 rainbow trout from the Trout Lodge Hatchery, Soap Lake, Washington. The transport tank for the Leavenworth fish was equipped with a recirculation system and supplemental oxygen supply. The approximate transport times from Leavenworth Hatchery and John Day Dam to the study site were 3.5 and 2 h, respectively. Approximately 24 h prior to tagging, 150 fish were transferred to a 200 gal holding tank on the upper spillway deck. Fish holding tanks, equipped with degassing units, were continuously supplied with ambient river water. Fish were held a minimum of 24 h prior to tagging to alleviate handling stress and to allow fish to acclimate to ambient river conditions. Water temperature in the holding pools ranged from 10.5 to 11.0 \degree C in the spring and 20.0 to 20.5 $^{\circ}$ C in the summer. The higher ambient river temperatures in July (19.5 to 20.5 $^{\circ}$ C) appeared to have imposed some stress on the sub-yearling chinook. Seven of the 475 sub-yearling chinook died in the holding tanks prior to testing and another 25 fish had excessive scale loss (>20% per side) or developed fungal infections (Table 2-6). None of the fish in the spring succumbed during the pre-test holding period or developed infections. Consequently, because of the poor fish condition and subsequent unacceptable level of holding mortality, the 48 h survival estimates for the July experiment were not generated.

2.3 Tagging and Release

Fish handling and tagging recapture techniques were identical to those previously used at other hydroelectric projects on the Columbia River Basin (Heisey *et al.* 1992; Mathur *et al.* 1996, 1999; Normandeau Associates *et al.* 1996a,b,c). Briefly, lots of 5 to 10 fish were removed with a water sanctuary equipped net from holding tanks (on the spillway deck) to the adjacent tagging site using a small tub full of water. Fish displaying abnormal behavior, severe injury, fungal infection, or descaling (>20% per side) were not used. The same fish selection criterion was applied to all treatment and control groups. Fish were anesthetized in a 0.5% MS 222 solution ($\lt 5$ min) and equipped with two uninflated HI-Z tags and a miniature radio tag. Table 2-1 shows the number of treatment and control fish released each day.

Figure 2-4 summarizes the length data of the treatment and control fish groups. Chinook salmon lengths averaged about 141 mm (range 122 to 181 mm) in April and May and 123 mm (range 113 to 157 mm) in July. Most fish were longer than 140 mm in spring, while in summer most fish measured less than 135 mm. The rainbow trout used in July averaged 144 mm (range 118 to 166 mm).

Tags were attached via a stainless steel pin inserted through the musculature beneath the dorsal and adipose fins. A radio tag was attached in combination with the dorsal HI-Z tag (Heisey *et al.* 1992). A uniquely numbered VI tag (Visual Implant, Northwest Marine Technology, Inc., Shaw Island, Washington) was also inserted in the postocular tissue for use in tracking 48 h survival of individual recaptured fish. Fish also received a fin clip to designate release location (test or control) in the event the VI tag became dislodged. HI-Z tagged fish were placed in a covered, 20 gal container continually supplied with ambient river water until fully recovered from anesthesia (generally 30 to 45 min, minimum 20 min). After full recovery from anesthesia, fish were individually placed into the induction system, tags were activated, and the fish was released. Inflation time of the tags was

partially regulated by the temperature and amount of water injected into the tags just prior to release and/or the ingredients within the tag.

All treatment and control fish were released through an induction apparatus that consisted of a small holding basin attached to a 4 in diameter flexible hose (Normandeau Associates and Skalski 1999, 2000a; Normandeau Associates and Mid Columbia Consulting 2001; Normandeau Associates *et al.* 1996a,b,c). The release hose was continuously supplied with river water to ensure fish were transported quickly to the desired release point.

At each treatment release site the 4 in diameter flexible hose was threaded through a 6 in diameter welded steel pipe (Figures 2-1 and 2-2). The steel pipe and hose was held in position by braces mounted on the spillway headworks and steel guide wires secured to the spillbay nose piers. The terminus of each treatment release hose was oriented downstream either 3 (deep) or 7 (shallow) ft above the ogee. The deep and shallow release pipes were offset 4 and 2 ft, respectively, from the middle of Spillbay 5 and approximately 16 ft upstream of the tainter gate (Figures 2-1 and 2-2).

One control release pipe with induction hose was positioned downstream of Spillbay 1 (Figure 1-1). The terminus of this pipe was approximately 170 ft downstream of the base of Spillbay 1 and approximately 24 ft and 29 ft above the tailrace, in spring and summer, respectively. The second control release site utilized the juvenile fish facility bypass pipe. Fish were released into the bypass pipe via an approximately 5 ft long section of the 4 in diameter induction hose. Fish were introduced 1,150 ft upstream of the end of the bypass pipe (Figure 1-1). The bypass pipe alternate control release site was chosen because dye releases through the Ice Harbor model at the Corps' Vicksburg, Mississippi facility indicated the potential for some fish released downstream of Spillbay 1 to be drawn upstream.

Procedures for handling, tagging, release, and recapture of fish were identical for treatment and control groups. Fish were randomly selected from each day's transport. All spill tests, except the 100% spill tests, were conducted during daylight hours (0700 to 1800 h). The 100% spill tests were conducted only in the evening. The Project did not commence 100% spill until 1800 h and terminated it at 0600 h.

Fish allocation for the spring 50%, 100%, and special spill was 310, 225, and 120 fish, respectively. The number of control fish released was 140 downstream of Spillbay 1 and 150 fish via the bypass pipe (Table 2-1). The summer releases consisted of 179 and 100 treatment chinook salmon for the bulk spill and dispersed spill, respectively, and 125 controls released only at the bypass pipe (Table 2-1). An additional 40 rainbow trout were released at the bulk spill; no control rainbow trout were released.

2.4 Fish Recapture

Upon passage, fish were tracked and retrieved when buoyed to the surface downstream of the spillbays by one of three or four recapture boat crews. Boat crews were notified of the radio tag frequency of each fish upon its release. Only crew members trained in fish handling were used to retrieve tagged fish. To minimize crew bias, no crew was specifically assigned to retrieve either control or treatment fish.

Radio signals were received on a 5-element Yagi antenna coupled to an Advanced Telemetry System receiver. The radio signal transmission enabled the boat crew(s) to follow the movement of each fish after passage and position the boats downstream for retrieval when the HI-Z tag buoyed the fish to the surface. Boats were required to remain a safe distance downstream of the turbulent discharge.

Occasionally during the summer testing period spill was temporarily curtailed after a group of 10 to 15 fish were released to permit the boats to recapture fish that had moved into eddy areas. Spill

curtailment was not possible in the spring and a few fish that became inaccessible to the boat crews were replaced with additional fish. Additionally, control fish could not be successfully released and recaptured during bulk spill conditions in July because the fish were drawn upstream from the exit of the bypass pipe into the turbulent Spillbay 5 discharge. Consequently, control fish were released during a period of no spill or dispersed spill.

Active radio tags which failed to surface were tracked for a minimum of 30 min and then periodically thereafter to ascertain if fish displayed movement patterns typical of emigrating smolts or that of a predator. Recaptured fish were placed into an on-board holding facility and tags were removed (Heisey *et al.*1992). Each fish was immediately examined for maladies consisting of injuries, descaling, and loss of equilibrium and assigned appropriate condition codes, if necessary, per the descriptions presented in Table 2-7. Tagging and data recording personnel were notified via a twoway radio system of each fish's recovery time and condition.

Each recaptured fish with a visible injury or scale loss was assigned a likely causal mechanism. Limited controlled experiments (Neitzel *et al.* 2000; PNNL *et al.* 2001) to replicate and correlate injury type and characteristic to a specific causative mechanism provides some indication of the cause of observed injuries in the field. Some injury symptoms can be manifested by two different sources which may lessen the probability of accurate delineation of a cause and effect relationship (Eicher Associates 1987).

All fish recaptured alive were transferred in 5 gal pails to 600 gal pools on the tailrace deck for assessment of delayed effects (48 h). Each pool equipped with a degassing unit was continuously supplied with ambient river water and shielded to prevent potential fish escape and/or avian predation. Each day's treatment and control fish were held together in the same pools for 48 h.

As a precautionary measure, the Corps secured the services of personnel from the U. S. Department of Agriculture to scare gulls from the tailrace. Past experience has shown that the hazing of gulls minimizes the potential loss of buoyed experimental fish to gulls, and thus maintains the use of prespecified sample sizes. However, predation by piscivores (*e.g.*, northern pikeminnow, smallmouth bass, or walleye) on tagged fish could not be controlled.

2.5 Classification of Recaptured Fish

As in the previous investigation at spillways and other experiments on the Columbia River Basin (Normandeau Associates *et al.* 1996a,b,c, 1997; Normandeau Associates and Skalski 1998, 1999, 2000a,b,c) the immediate post-passage status of an individual recaptured fish and recovery of inflated tags dislodged from fish was designated as alive, dead, tag and pin recovered, unknown, or predation. The following criteria have been established to make these designations: (1) alive--recaptured alive and remaining so for 1 h; (2) alive--fish does not surface but radio signals indicate movement patterns typical of emigrating juveniles; (3) dead--recaptured dead or dead within 1 h of release; (4) dead--only inflated dislodged tag(s) are recovered, and telemetric tracking or the manner in which inflated tags surfaced is not indicative of predation; (5) unknown--no fish or dislodged tags are recaptured, or radio signals are received only briefly, and the subsequent status cannot be ascertained; and (6) predation--fish are either observed being preyed upon, the predator is buoyed to the surface, or subsequent radio telemetric tracking indicates predation (*i.e.*, rapid movements of tagged fish in and out of turbulent waters or sudden appearance of fully inflated tags). Unrecovered preyed upon fish are assumed dead in the survival calculations; alive recaptured fish suspected of predator attack were included with the alive category.

Mortalities of recaptured fish occurring after 1 h were assigned 48 h post-passage effects although fish were observed at approximately 12 h intervals. Specimens were examined for descaling and injury, and those that died were necropsied to determine the probable cause of death. Additionally all

specimens alive at 48 h were re-anesthetized and closely examined for injury and descaling. The reexamination of immobilized fish minimizes the need for extensive handling and associated stress upon immediate recapture. The initial examination allows detection of some injuries, such as bleeding and minor bruising that may not be evident after 48 h due to natural healing processes (Normandeau Associates *et al.* 1996a,b,c). Injury and descaling were categorized by type, extent, and area of body.

Fish without visible injuries that were not actively swimming or swimming erratically at recapture were classified as "loss of equilibrium". This condition has been noted in most past studies and often disappears within 10 to 15 min after recapture if the fish is not injured (Normandeau Associates *et al.* 1996a,b,c). A malady category was established to include fish with visible injuries, scale loss (greater than 20% on either side), or loss of equilibrium. Dead fish without any of these symptoms were not included in this category. Fish without maladies were designated "clean fish". Detailed descriptions of maladies observed on each recaptured fish are presented in Appendix Tables B-1 and B-2.

This clean fish metric was established to provide a standard way to present a rate depicting how a specific passage route affected the condition of passed fish. Clean fish, the absence of maladies, was chosen so that this metric may be more comparable to survival; however, the clean fish metric is based solely on fish physically recaptured and examined. Additionally, the clean fish metric in concert with site-specific hydraulic and physical data can provide insight into what passage conditions may provide safer fish passage.

Visible injuries were also categorized as minor or major, based on laboratory studies by PNNL *et al.* (2001) and Normandeau's field observations. These are as follows:

- Minor Injuries that were visible but not life threatening and tended to heal and disappear over the post-exposure observation period. Hemorrhages that covered less than half an eye or small bruises (approximately 0.5 cm in diameter) with minor discoloration (most commonly observed at the dorsal insertion of the operculum) were given a minor injury rating because fish quickly recovered from such injuries and/or displayed no apparent ill effects.
- Major Any injury that was life threatening, or persisted throughout the post–exposure observation were rated major, except eye hemorrhages of less than 50%. For example, a large bruise (>0.5 cm in diameter), damage to the spinal column, cuts with visible bleeding, injured eyeballs (bulging, hemorrhaged, or missing), gill damage (inverted gill arches severe enough to result in bleeding).

2.6 Spill Volume and Impact Velocity

The spill volume through Spillbay 5 varied between 3.4 (2 stops) and 8.5k cfs (5 stops) during the spring investigation to maintain a 50 or 100% spillway discharge (Table 2-2). During the summer bulk spill study, only two spill volumes (3.4 and 13.6k cfs, 2 and 8 stops) were tested at Spillbay 5.

Impact velocities were estimated for deflector impact only because the flow of the discharge swept the deflector. The velocity of the discharge jet upon impact with the flow deflector was calculated by adding vertical and horizontal vectors of velocity. The vertical component was calculated based on the vertical distance from the center of the jet to the deflector. Estimated impact velocities averaged from 70.2 to 71.6 ft/s in spring and from 70.0 to 73.3 ft/s in the summer (Appendix A); the laboratory studies suggest these velocities exceed those capable of inflicting injury/mortality (approximately 58 ft/s) on fish (Neitzel *et al.* 2000).

2.7 Survival and Clean Fish Estimation and Data Analysis

Passage survival probabilities ($\hat{\tau}$) for each spill condition were estimated relative to the control fish survival (Heisey *et al.* 2002a; Mathur *et al.* 1996, 1999). Survival probabilities were also computed for approximate location of fish passage; shallow (released 7 ft above ogee) and deep (released 3 ft above ogee). Data from individual daily trials (Appendix Tables B-3 and B-4) were used in the analysis. However, excessive mortality of both treatment and control fish occurred during the 48 h delayed assessment period in the summer experiment and thus, immediate (1 h) survival was estimated. The treatment conditions and common controls were simultaneously analyzed and modeled by joint likelihood (Normandeau Associates *et al.* 1996a,b,c).

A likelihood ratio test was used to determine whether recapture probabilities were similar for alive (P_A) and dead (P_D) fish. The statistic tested the null hypothesis of the simplified model $(H_O: P_A = P_D)$ versus the alternative of the generalized model $(H_A: P_A \neq P_D)$. Depending upon the outcome of this analysis for the 1 h survival the parameters and their associated standard errors were calculated using that model.

As in previous studies (Normandeau Associates and Skalski 2000a), separate chi-square analyses (Appendix C) were performed to test for homogeneity (P=0.05) between daily treatment and control releases with respect to recapture frequencies of alive, dead, and non-recovered fish. Homogeneity (P>0.05) between daily control trials within each season (spring and summer) allowed pooling of data. Thus, data from all of the daily control releases within each season were pooled and survival for each spill condition and passage location was estimated relative to survival of the pooled control group. All statistical analyses were conducted using the Statistical Analysis System (SAS). The statistical outputs with exact probabilities are provided in Appendix C (output discussed in the report are highlighted). The disposition of individual fish is given in Appendix D and only summarized information is discussed in the main body of the report.

The clean fish estimate (CFE) was calculated separately for both spring and summer releases. It was based on recaptured fish without maladies (*i.e.*, no visible injuries, scale loss, or loss of equilibrium) or displayed maladies that were not attributable to passage, *i.e.*, injuries solely attributed to predator attack or tag induced (tear at tag site). Clean fish estimates for each spillbay were made relative to the probabilities of control fish that were free of any maladies. Data from individual daily trials (Appendix Tables B-5 and B-6) were used in the analysis.

The 90% confidence intervals on the survival for clean fish estimates were calculated using the profile likelihood method (Normandeau Associates *et al.* 1996a,b,c and Appendix C); these are deemed superior to those based on the assumption of normality. Although the study was not designed for hypothesis testing, differences in survival and clean fish estimates between spill rates or passage location were tested, *a posteriori*, by Z-statistics (see Appendix C).

2.8 Autonomous Sensor Fish

Sensor fish, an instrumented package designed to determine exposure histories to turbulence and pressure during passage (PNNL *et al.* 2001) were equipped with two or three HI-Z tags and a miniature radio tag and released using the identical induction release hose into the same spill conditions as for the live fish during the spring investigation. No sensor fish were released in summer. Sensor fish were also released at the control release sites. Generally, at least one sensor fish was released with each group of 10 fish. The results of sensor fish passage are to be provided by PNNL in a separate report. However, relevant portions of that report can be included when available to explain some of the observed results on alive fish releases. Preliminary results were presented in November 2003 at the annual AFEP meeting in Walla Walla, Washington (Carlson and Duncan 2003).

2.8.1 Sensor Fish on Adult Fish

The feasibility of attaching a sensor fish externally to an adult fish and recapturing them after spillway passage was evaluated. This procedure will provide a means to obtain the post-passage physical condition of a fish along with the hydraulic forces it encountered during passage. The basic procedures utilized to attach HI-Z tags to adult salmonids passed through a turbine at McNary Dam were followed (Normandeau Associates and Mid Columbia Consulting 2003). Three rainbow trout 406 to 508 mm long and 1 to 2 kg were obtained from Pacific Northwest National Laboratory (PNNL), Richland, Washington. These fish were held at Ice Harbor in a 200 gal tank on the spillway deck.

Prior to HI-Z tag and sensor fish attachment the adult fish were removed from the holding tank and placed into a 12 gal tub full of water. The fish were then guided into a fish stress reduction tube (Figure 2-5). This device covers the head and eyes while still providing room for opercular movement and also allows large sized fish to be tagged and the tags activated with minimal stress and without anesthesia (Heisey *et al.* 2002b). This stress reduction device is a foam-lined, split, hinged, 6 inch diameter PVC tube that can be securely closed around the test specimens. Several notches cut into this device allow access to the fish for attachment of tags. One large (20 by 55 mm pre-inflated) HI-Z tag and a radio tag was sutured near the dorsal fin insertion and another HI-Z tag was sutured near the adipose fin. The fish was then turned over and the sensor fish package sutured at the base of one pelvic fin and the anal fin. A sensor package included a radio tag and a small HI-Z tag attached to the sensor fish in the event that the sensor package were to become detached from the adult fish. Two large HI-Z tags were attached at the base of the other pelvic fin. The adult fish was then transferred to the release site, the HI-Z tags injected with catalyst, and the fish was removed from the restraining tube and hand released into the Spillbay 1 discharge. The recapture crew followed basic fish recapture methods used on juvenile fish (Section 2.4) to track and retrieve adult tagged fish with a sensor package. The boat crews retrieved the buoyed fish with a net equipped with a water sanctuary and placed it into a tub of water and then removed the tags and sensor fish. Each recaptured fish was examined for injuries.

3.0 RESULTS

3.1 Recapture Rates

Recapture rates (physical retrieval of both alive and dead fish) of treatment and control groups were generally higher in spring than for summer releases (Tables 3-1 and 3-2). Recapture rates of treatment groups in spring were 100% for all conditions except fish released through the deep pipe at 50% spill where all but 1 of the 305 fish released was recaptured. Recapture rates of the control group were also high, 99.3%; only 2 of the 290 fish released were not recaptured. Most of the recaptured treatment (98.8%) and control (100%) fish were alive.

Recapture rates of sub-yearling chinook salmon treatment groups in summer ranged from 90.8% (bulk spill, shallow release) to 97.0% (dispersed spill, deep release) (Table 3-2). Recapture rates for the rainbow trout treatment groups ranged from 95.0% (bulk spill, shallow release) to 100% (bulk spill, deep release). The recapture rate for the control group released through the bypass pipe was 99.2%.

The percentage of fish from all groups in July assigned to the dead category ranged from 0% (rainbow trout, bulk spill, deep) to 8.4% (chinook salmon, bulk spill, shallow; Table 3-2). Three (2.5%) sub-yearling salmon assigned dead from the bulk spill shallow experiment were preyed upon. This was the highest predation rate observed.

Some 4.5% of the chinook salmon released for the bulk spill pattern test in the summer had dislodged tags, while none were observed in the dispersed spill test, and only 0.3% was observed for the treatment fish released in the spring (Table 3-2). The higher incidence of tag dislodgment may have been partially due to the smaller sized fish used in the summer (average 123 mm) than in the spring (average 141 mm).

Chi-square analyses indicated homogeneity (P>0.05) between daily control trials, allowing for the pooling of data. Homogeneity (P>0.05) was also revealed between each daily treatment trial. Thus, survival for each treatment condition was estimated relative to survival of the pooled control data in each season. Appendix C provides the outputs of these analyses with the associated exact probabilities.

With one exception (the estimate for deep release at 100% spill in spring), likelihood ratio tests indicated no significant differences (P >0.05) between the simplified (H₀:P_A=P_D) and generalized $(H_A: P_A \neq P_D)$ models. For 48 h estimates, however, there were no exceptions. It was noted that the recovery probability for dead fish (P_D) could not be estimated by the embedded constraints in the data set and it was assumed 1.0. The standard error of the estimate could not be calculated as well for the 1 h survival. For this data set, of the 148 treatment fish released, 147 were classified alive and 1 dead at 1 h; at 48 h there were 146 fish alive and 2 dead. At 48 h, an additional fish was dead and all of the model calculations were possible. The likelihood ratio test indicated no difference (P>0.05) between the recapture probabilities of dead (P_D) and alive (P_A) fish. Consequently, the 1 h statistic was ignored for the purposes of delineating trends in survival. Thus, survival probabilities and their associated standard errors were calculated using the simplified model for all test conditions. These values are highlighted in Appendix C (statistical outputs).

3.2 Retrieval Times

Retrieval times (the interval between fish release through the induction system and physical retrieval) for various releases were short and similar (Figure 3-1). Average times were 6.3 to 7.6 min for treatment groups and 9.4 min for control groups in the spring. During the summer, average recapture times for treatment fish ranged from 6.2 to 9.8 min and 5.4 min for control fish.

3.3 Survival Estimates

The 1 h survival estimates during the spring were high (\geq 98.7%) (Table 3-3) and little variation between release locations within a spill pattern occurred. However, differences in immediate (1 h) survival between spill patterns at release locations were observed. Survival for deep released fish $(\hat{\tau} = 0.987)$ at 50% spill was significantly lower (P<0.05, one tailed Z-test) than at 100% spill $(\hat{\tau} = 1.00)$. No other differences were detected.

For the summer test, the estimated immediate survival estimates were lower than those in spring (Table 3-3). Survival ranged from 88.9 to 97.9%. The effects of spill pattern (bulk spill versus dispersed spill) for comparable deep release location were evident; the survival was substantially higher (97.8%, 90% CI=93.9 to 100.0%) at dispersed spill than at bulk spill (92.3%, 90% CI=85.7 to 93.9%). The pooled (deep and shallow) immediate survival for the bulk spill was 90.1%; this estimate is also substantially lower than at the dispersed spill.

Precision (ε) on all immediate survival estimates for spring was $\leq 1.5\%$, 90% of the time. Precision (ε) on all estimates in the summer was lower (3.0 $\leq \pm$ ε $\leq \leq \pm$ 6.0%). The lower precision in summer was primarily due to lower recapture rates, survival, and a relatively smaller sample size.

The 48 h survival estimates in the spring were nearly the same as the 1 h estimates (Table 3-3); 98.7% (50% and special spill conditions) and 99.4% (100% spill). Little difference occurred between treatments. The statistical difference noted at 1 h between the 50% spill and 100% spill patterns at

the deep release location was not detected at 48 h due to a loss of some treatment fish during holding. Reliable 48 h survival estimates for the summer releases could not be generated due to unacceptable high control fish mortality (>20%) during the delayed assessment period. Again, the prespecified precision (ε) level of \leq +3%, 90% of the time, was met on all 48 h estimates in spring (Table 3-3).

As stated above, mortality of experimental and control fish during the 48 h assessment period was higher than desirable during the summer tests. Some 31% of the control fish died during the delayed holding period and 32 to 41% of the treatment groups also died (Table 3-2). In contrast, none of the 288 control fish and only 2 of 647 treatment fish died during the 48 h delayed assessment period in the spring (Table 3-1).

3.4 Injury Classification, Rates, and Probable Causal Mechanisms

All recaptured fish were examined for types of external injuries and those that were recaptured dead were also examined for internal injuries. Detailed descriptions of all recaptured injured fish are presented in Appendix Tables B-1 and B-2. Injury rates given below are based on the total number of recaptured fish examined and not the total number of fish released and refer to only passage-related injuries, adjusted for controls. Injury rates for both control releases (Spillbay 1 and bypass pipe) were zero (0%) in the spring and 1.6% in the summer.

Passage-related injury rates during the spring experiment appeared to be a function of spill volume and passage location (Table 3-4). Injury rates were generally lower, particularly for shallow released fish, with a greater spill volume through Spillbay 5. Higher injury rates (20.0 to 20.7%) occurred at the 50% spill (3.4 to 5.1 kcfs through Spillbay 5) volume than at the 100% spill, 1.3 to 12.2% (4.25 to 8.5 kcfs spill through Spillbay 5); at special spill pattern (8.5 kcfs through Spillbay 5) the injury rate was 8.3 to 21.7%.

Passage location (shallow or deep) also affected injury rates (Table 3-4). Fish released at the shallow depth suffered lower injury rates (1.3 to 8.3%) than those released through the deep pipe (12.2 to 21.7%); only five fish were released at the shallow depth at 50% spill pattern to provide conclusive information. For the special spill, the injury rates for these respective release pipes were 8.3 and 21.7%. When the effect of passage location is examined across all spill conditions, the injury rate is 4.9 and 18.4% for fish that passed through the shallow and deep pipes, respectively.

The common injury types in spring were hemorrhaged or damaged eye (11.5%), bruises and/or scrapes (3.4%), and gill or operculum damage (3.2%) (Table 3-4 and Figures 3-2 to 3-4). Few treatment fish sustained lacerations (1.1%) or internal injuries (0.6%) . Fish passing via the deep pipe were more likely to sustain eye damage than those passed through the shallow pipe, 13.7 versus 3.5%. The incidence of gill/opercular damage and hemorrhaged body was higher for the deep pipe passed fish (3.7 to 3.9%) than for the shallow pipe released fish (1.4%).

The effect of passage location (shallow or deep) on injury rates was similar in the summer experiment (Table 3-5). The passage related injury rates for the bulk spill were 4.9 and 18.4% for the shallow and deep fish releases, respectively (Table 3-5). The dispersed spill flow pattern (deep release only) had an injury rate of 20.0%. The injury rate, 19.5%, for fish released via the deep pipe for the combined spill patterns was similar to that observed in the spring tests (18.4%). The injury rates were identical (4.9%) for fish released by the shallow pipe, combined spills, during both spring and summer.

One of 20 (5.0% unadjusted; no controls were released) rainbow trout released deep during bulk spill in the summer tests exhibited an injury. None of the 20 rainbow trout passed via the shallow pipe at bulk spill were injured. Rainbow trout also appeared to be immune to the infection that affected some of the subyearling chinook salmon during the delayed assessment period.

The common injury types in summer were again bruises and/or scrapes, hemorrhaged or damaged eyes, and gill or operculum damage (Table 3-5 and Figures 3-2 to 3-4). During bulk spill, the deep pipe released fish, when adjusted for controls, incurred 8.3% eye damage, 8.3% hemorrhaged body, 7.3% opercular damage, and 5.5% cuts to the body. Fish released through the deep pipe during the dispersed spill had a higher rate of eye (13.6%) and opercular (9.3%) damage, but lower body hemorrhaging (1.3%) and no lacerations.

Shear forces were the probable cause for most of the eye injuries (especially hemorrhage) observed among treatment fish. Other injury types were more likely the result of contact with spillway or tailrace structures, possibly the flow deflector.

When the incidence of equilibrium loss and/or scale loss are included with the visibly injured fish to obtain a malady rate the results follow the trend observed for visible injuries (Tables 3-6 and 3-7). The malady rates (adjusted for control fish) for the special spill, 50% spill, and 100% spill tests in the spring (deep and shallow combined) were 15.8, 21.4, and 10.3%, respectively (Table 3-6). Fish released through the shallow pipe had a lower rate (6.3%) than those passed through the deep pipe (19.4%) for all spill levels combined in the spring.

The shallow release site for the summer bulk spill also had the lowest malady rate (4.3%; Table 3-7). The malady rates for fish passed through the deep pipe at the bulk and dispersed spill during July were 19.6 and 21.8%, respectively.

3.5 Clean Fish Estimates (CFE)

CFEs were considerably lower (up to 20%) than the direct survival estimates (Tables 3-3 and 3-8). The CFEs for the 50%, 100%, and special spill releases in spring were 78.6 (90% CI=74.8 to 82.5%), 89.7 (90% CI=86.4 to 93.1%), and 84.2% (90% CI=78.7-89.6%), respectively. The 90% CI for the above estimates were between ≤±3 to 5%.

Although the study was not originally designed to detect differences between two treatments, Z-statistic (two-tailed test) was used *a posteriori* to assess statistical differences. With respect to spill patterns, significant differences ($P<0.05$) were noted between 50% and 100% spill patterns in spring. CFE was significantly lower (P<0.05, $Z=2.38$) for deep released fish (0.877) at 100% spill pattern than at 50% spill pattern (0.793). CFE for deep released fish at special spill (0.783) was not significantly lower (P>0.10) than for deep released fish at 100% spill (0.878); the sample size was too small to detect this difference. CFEs for deep released fish at 50% spill and special spill were virtually identical (0.786 and 0.783).

With respect to release locations within each spill pattern in spring, significant differences (P<0.05)were noted with deep released fish having lower CFEs at both 100% spill and bulk spill patterns (Table 3-8). The magnitude of difference was similar at both spill patterns (about 12%).

The same effect of passage location was also observed in the July test (Table 3-8). Clean fish estimates were significantly lower (P<0.05) for the fish passed through the deep pipe $(81.3\%, 90\%$ CI=72.2 to 90.5%) than the shallow pipe $(95.1\%, 90\%$ CI=90.7 to 99.4%) at the bulk spill condition. The clean fish estimate for fish passed through the deep pipe at dispersed spill (77.4%) was also significantly (P<0.05) lower than the estimate for fish released via the shallow pipe (95.1%, 90%) CI=89.7 to 100.0%) at bulk spill.

3.6 Adult Fish

No injuries were observed on the two adult fish equipped with a sensor package. The sensor package attachment method proved feasible. All balloon tags and radio tags, along with the sensor package,

were securely attached to the adult fish upon recapture. Recapture times for the two fish were 4 and 5 minutes.

4.0 DISCUSSION

The primary objectives and assumptions established for the experiment were met for the spring (April and May) high tailwater condition. However, the objectives for the summer (July) low tailwater condition could not be fully realized, particularly the 48 h direct survival estimate with a prespecified precision (ϵ) level. The summer investigation was affected by a relatively high mortality of experimental and control fish during holding (>20%). Consequently, reliable 48 h survival estimates could not be generated from these data.

The fish appeared to tolerate handling stress associated with tagging, release, and recapture considerably better at lower water temperatures (\leq 15°C or \leq 59°F) than at higher water temperatures (≥19°C or ≥67°F) in the summer. Less than 1% of the fish died during the 48 h holding period in spring compared to 35% in the summer. In the summer test, immediate (1 h) losses of treatment and control fish in the daily trials ranged from 0 to 15%. However, losses over the 48 h holding period ranged from 0 to 42% for treatment and 20 to 43% for control fish.

Control losses of greater than 10% are rare, particularly at water temperatures \leq 15.0°C (\leq 59.0°F); in only 1 of 37 (3%) similar investigations of estimating direct effects of passage elsewhere on the Columbia River Basin did losses exceed 10% (Normandeau Associates *et al.* 2003a,b). Muscle necrosis (bacterial), particularly at the tagging location, was observed, along with fungal infection. These observations were also corroborated by fish infections in the holding tanks prior to tagging, handling, and release. Some 5% of the fish developed fungal infections. The handling, tagging, and recapture methods were identical both in the spring and summer. These observations are similar to those made during the recent survival experiments conducted at The Dalles and Bonneville spillways at water temperatures ≥19°C or ≥67°F (Normandeau Associates *et al.* 2003a,b).

Although only 40 juvenile rainbow trout were tested during the summer, there is evidence that these hatchery fish were less susceptible to disease and mortality during the delayed assessment period than the subyearling chinook salmon collected from the river. Only 1 of the 37 trout died during the 48 h delayed assessment period. Additionally, none of the trout appeared to develop fungal infection.

Tag dislodgment from smaller fish size in the summer (average length 123 mm versus 141 mm in spring) may have also contributed to the lower estimate of survival to a certain extent (Normandeau Associates *et al.* 2003a,b). Tag dislodgment, perhaps exacerbated by passage through turbulent waters, was more common on fish <120 mm long than on fish >120 mm; tag dislodgment does not necessarily indicate fish mortality though the fish is assumed dead in survival estimation (Mathur *et al.* 1996). Of the 9 treatment tag dislodgment(s) observed in summer, 8 occurred on fish <120 mm long and the remainder one on the larger rainbow trout. In the spring, only one tag dislodgment occurred. The fish were longer (most >130 mm) in the spring.

Regardless of spill pattern tested in both spring and summer, the deeper releases, particularly at a lower spill volume (≤5 kcfs) through the test Spillbay 5, had a notably higher incidence of malady rates than the shallow releases. The overall malady rate for deep releases ranged from 14.3 to 22.6% versus 2.6 to 10.0% for shallow releases. It is likely that fish passing closer to the ogee would be prone to pass near the flow deflector increasing the potential for collision. The actual path traversed by each balloon tagged fish released in the present study and the unique hydraulic jet conditions (*e.g.*, plunging, skimming, or undulating) experienced in the tailrace are unknown. However, data from concurrent release of balloon tagged "sensor fish" by PNNL to simulate the hydraulic conditions experienced by alive released fish indicate potentially injurious hydraulic conditions in the immediate vicinity of the

flow deflector (Carlson and Duncan 2003). Although the present study suggests that fish passing close to the ogee have a greater chance of injury, the proportion of naturally entrained fish subjected to these conditions must be ascertained to estimate the potential injury rate to these fish. Hydroacoustic data from Ice Harbor in 2003 (Moursund *et al.* 2003) suggest that most fish would pass higher above the ogee than the deep release point. However, the proportion of entrained fish subjected to potentially injurious conditions could increase at lower tainter gate openings (2 to 3 ft) because they would have less opportunities to orient higher in the discharge jet.

One notable finding of the summer experiment was a discrepancy between survival estimates in the present experiment and that derived from the PIT tag detection technique. Intuitively, estimates derived by the HI-Z tag-recapture method, which reflects direct effects of passage, are expected to be higher than those derived from the PIT tag detection technique; the estimates from the latter methodology portray both the direct and indirect post-passage effects (which may occur over time and distance). The present study estimated 1h survival of 88.9% (bulk spill shallow release) and 92.3% (bulk spill deep release). The preliminary survival estimate from PIT tagged fish released through Spillbays 1 and 3 was 96% (Absolon *et al.* 2003). Possible explanations for the discrepant estimates may include: (1) The HI-Z tag estimates may be conservative because only the tags were recaptured on 4.2% of the bulk spill shallow releases and 5.0% of the bulk spill deep releases; these fish were assigned dead status, but in fact may have survived. If this were the case, then the PIT and HI-Z tag estimates would be comparable; (2) the fish were released at different locations with consequent different travel paths in exiting the spillbay. PIT tagged fish were released at a relatively shallow depth (approximately 10 ft) upstream of the project (Absolon *et al.* 2003) and were likely entrained higher in the discharge jet and could have passed over the ogee at a shallower depth; (3) HI-Z tagged fish were released through Spillbay 5 while PIT tagged fish were released through Spillbays 1 and 2; differences in passage survival between spillbays have been noted elsewhere (Normandeau Associates *et al.* 2003b); and (4) PIT tagged fish, if injured, did not succumb; injuries, if similar to those observed herein, were not lethal. Any potential adverse effect of the induction hose for the HI-Z tagged fish was ruled out because the same induction system has been used at many sites on the Columbia River Basin without injurious effects on the released fish; injury rates on control fish were negligible. Additionally, upon completion of the study, the induction hose was physically examined via camera for defects; none were observed.

The PIT tag derived summer survival estimates were also higher than those estimated by radio telemetry (Eppard *et al.* 2003). They reported relative spillway passage survival estimates of 94.8 and 92.8% for BiOp (100%) and 50% spill releases in the spring, respectively. As expected, these are lower than those estimated in the present study (>98%).

The assignment of causal mechanisms to individual injury types in the field though difficult, followed symptoms noted by Neitzel *et al.* (2000) in laboratory studies. They reported that localized shear forces caused a variety of injuries to the eyes, opercles, and body of juvenile salmonids. In the current study, most of the passage-related eye injuries and tears at the opercle attachment site were attributed to shear forces. However, some of the eye damage may have been due to physical contact with spillway structures, likely the flow deflector. Additionally, tears and abrasions on the opercle appeared to be contact related. Scrapes and swaths of scale loss on the head or body were attributed to contact or impact with the spillbay flow deflector or tailwater structures (baffles and end sill). Some bruises, however, can be caused by shear forces (Neitzel *et al.* 2000).

Unlike in some recent summer studies at other hydroelectric dams on the Columbia River Basin (Normandeau Associates *et al.* 2003a,b) predation on tagged fish had a minimal effect on the results; approximately 2% (bulk spill, shallow release) of the treatment and 0% of the control fish were lost to predation in summer. For the spring releases there was no loss to predation of treatment or control fish. At other sites, predation on HI-Z tagged fish had exceeded 10% in summer experiments.

5.0 CONCLUSIONS AND RECOMMENDATIONS

Some evidence exists to suggest that a higher spill volume through one to three spillbays may be more beneficial for passage survival and minimizing fish injury than dividing the same spill volume among all of the spillbays. Differences may be partially due to a greater "water cushion" for entrained fish at a higher spill volume. Additionally, when a divided spill volume is released through all of the spillbays rather than a few (one to three spillbays), the opening for each tainter gate is substantially less. For example, each gate would be opened approximately 3 ft for a total spill of 50k cfs apportioned to 10 spillbays versus a 10 ft tainter gate opening if only three spillbays were operated. Fish entrained higher in the water column are less likely to encounter solid downstream structures (*e.g.*, flow deflectors, rock outcrops, etc.).

Injury rates may be a better indicator of potentially adverse passage conditions relative to the spill pattern and passage location than absolute survival estimates *per se*. Fish entrained deeper within the discharge jet had higher visible injury rates (16.5 to 20.0%) than fish that likely passed higher in the jet. It is likely that the fish which passed via the shallower (7 ft above the ogee) release pipe were provided a better "water cushion" than those passed via the deeper pipe. Causal mechanisms of injuries were primarily shear and collisions with hard objects in the stilling basin.

Released fish experienced potentially injurious impact velocities during the study. Although estimated impact velocities encountered by released fish ranged from 70.1 to 73.2 ft/s and laboratory studies suggest that impact velocities exceeding approximately 58 ft/s may inflict injury and mortality on fish, the unique characteristics of the jet upon intercepting the tailwater with its associated elevation and submergence of flow deflectors may expose only a proportion of fish released.

Because eye damage was the dominant injury type, information is needed on the persistence and severity of eye injuries on post-spillbay survival of emigrating fish, effects on vision, predator avoidance, and ability to feed.

Although the effects of flow deflector *per se* on direct fish survival were minimal in the spring, it appears that injury/malady rates provided some evidence of the potential adverse effects of passage location at Ice Harbor. Regardless of the spill pattern, fish that passed through the deeper (3 ft above ogee) release pipe suffered more injuries than those released through the shallow release pipe (7 ft above ogee). These findings need to be integrated with the proportion of the emigrating juvenile salmonid population that is subjected to the prevailing hydraulic conditions. Further research is needed to ascertain the relationship between potentially injurious hydraulic conditions and the travel pattern of sensor fish and submergence depth of the flow deflector.

The susceptibility of juvenile salmonids to the stress associated with handling, tagging, and release at high water temperatures (>19.0°C or 66.2°F), resulted in high treatment and control losses (>10%) during the summer study. In contrast, losses attributed to handling stress in spring (water temperatures of 10.0 to 11.5 $^{\circ}$ C or 50.0 to 52.7 $^{\circ}$ F) were near 0%. To minimize losses from handling stress and improve the precision of the survival estimates, future studies at Ice Harbor should be conducted at water temperatures $\leq 15.0^{\circ}$ C ($\leq 59.0^{\circ}$ F).

6.0 LITERATURE CITED

Absolon, R. F., M. B. Eppard, B. P. Sandford. 2003. Relative survival of PIT tagged hatchery chinook salmon through Ice Harbor Dam, 2003. U. S. Army Corps of Engineers 2003 Anadromous Fish Evaluation Program (AFEP) Annual Review, Walla Walla, WA.

- Carlson, T. J., and J. P. Duncan. 2003. Characterization of the Ice Harbor spill environment. U.S. Army Corps of Engineers 2003 Anadromous Fish Evaluation Program (AFEP) Annual Review, Walla Walla, WA.
- Eicher Associates, Inc. 1987. Turbine-related fish mortality: review and evaluation of studies. Research Project 2694-4. Prepared for Electric Power Research Institute, Palo Alto, CA.
- Eppard, M. B., E. E. Hockersmith, G. A. Axel, and B. P. Sandford. 2003. Migrational behavior and survival of radio tagged hatchery yearling chinook salmon at Ice Harbor Dam, 2003. U. S. Army Corps of Engineers 2003 Anadromous Fish Evaluation Program (AFEP) Annual Review, Walla Walla, WA.
- Heisey, P. G., D. Mathur, and T. Rineer. 1992. A reliable tag-recapture technique for estimating turbine passage survival: application to young-of-the-year American shad (*Alosa sapidissima*). Can. Jour. Fish. Aquat. Sci. 49:1826-1834.
- Heisey, P. G., D. Mathur, J. R. Skalski, and R. C. McDonald. 2002a. Effects of spillway structural modifications on juvenile salmonids survival. Amer. Fish. Soc. $4th$ BioEngg. Symposium. (In review with editor)
- Heisey, P. G., D. Mathur, and J. L. Fulmer. 2002b. Turbine passage survival of late running American shad and its potential effects on population restoration. Amer. Fish. Soc. 4th BioEngg. Symposium. (In review with editor)
- Mathur, D., P. G. Heisey, E. T. Euston, J. R. Skalski, and S. Hays. 1996. Turbine passage survival estimation for chinook salmon smolts (*Oncorhynchus tshawytscha*) at a large dam on the Columbia River. Can. Jour. Fish. Aquat. Sci. 53:542-549.
- Mathur, D., P. G. Heisey, J. R. Skalski, and D. R. Kenney. 1999. Survival of chinook salmon smolts through the surface bypass collector at Lower Granite Dam, Snake River. Pages 119-127 in M. Odeh, editor. Innovations in fish passage technology. American Fisheries Society Bethesda, Maryland.
- Moursund, R. A., K. D. Ham, P. S. Titzler, and F. Khan. 2003. Hydroacoustic evaluation of the effects of spill treatments on fish passage at Ice Harbor Dam in 2003. U. S. Army Corps of Engineers 2003 Anadromous Fish Evaluation Program (AFEP) Annual Review, Walla Walla, WA.
- Neitzel, D. A., and nine co-authors. 2000. Laboratory studies of the effects of shear on fish, final report FY 1999. Prepared for Advanced Hydropower Turbine System Team, U. S. Department of Energy, Idaho Falls, ID.
- Normandeau Associates, Inc., and J. R. Skalski. 1998. Chinook salmon smolt passage survival through modified and unmodified spillbays at Rock Island Dam, Columbia River, Washington. Report prepared for Public Utility District No. 1 of Chelan County, Wenatchee, WA.
- Normandeau Associates, Inc., and J. R. Skalski. 1999. Evaluation of mortality and injury associated with smolt passage through spillbays with a sloped flow deflector and no flow deflector at the Wanapum Dam. Report prepared for the Grant County Public Utility District No. 2, Ephrata, WA.
- Normandeau Associates, Inc., and J. R. Skalski. 2000a. 1999 spillway passage survival investigation of juvenile chinook salmon at Rock Island Dam, Washington. Report prepared for Public Utility District No. 1 of Chelan County, Wenatchee, WA.
- Normandeau Associates, Inc., and J. R. Skalski. 2000b. Evaluation of prototype shallow-flat deflector at Wanapum Dam spillbay 5 relative to chinook salmon smolt passage survival, 1999. Report prepared for Grant County Public Utility District No. 2, Ephrata, WA.
- Normandeau Associates, Inc., and J. R. Skalski. 2000c. Passage survival investigation of juvenile chinook salmon through a bypass pipe on the Columbia River, Washington. Report prepared for Public Utility District No. 1 of Chelan County, Wenatchee, WA.
- Normandeau Associates, Inc. and Mid Columbia Consulting, Inc. 2001. Feasibility of estimating direct mortality and injury on juvenile salmonids passing The Dalles Dam spillway during high discharge. Report prepared for Department of the Army, Corps of Engineers, Portland District, Portland, OR.
- Normandeau Associates, Inc. and Mid Columbia Consulting, Inc. 2003. Feasibility of estimating passage survival of adult salmonids using the HI-Z tag-recapture technique. Report prepared for Department of the Army, Walla Walla District, Walla Walla, WA.
- Normandeau Associates, J. R. Skalski, and Mid Columbia Consulting, Inc. 1995. Turbine passage survival of juvenile chinook salmon (Oncorhynchus tshawytscha) at Lower Granite Dam, Snake River, Washington. Report prepared for U. S. Army Corps of Engineers, Walla Walla, WA.
- Normandeau Associates, J. R. Skalski, and Mid Columbia Consulting, Inc. 1996a. Potential effects of modified spillbays on fish condition and survival at Bonneville Dam, Columbia River. Report prepared for Department of the Army, Portland District COE, Portland, OR.
- Normandeau Associates, J. R. Skalski, and Mid Columbia Consulting, Inc. 1996b. Potential effects of modified spillbays on fish condition and survival at The Dalles Dam, Columbia River. Report prepared for Department of the Army, Portland District COE, Portland, OR.
- Normandeau Associates, J. R. Skalski, and Mid Columbia Consulting, Inc. 1996c. Fish survival in passage through the spillway and sluiceway at Wanapum Dam on the Columbia River, Washington. Report prepared for the Grant County Public Utility District No. 2, Ephrata, WA.
- Normandeau Associates, J. R. Skalski, and Mid Columbia Consulting, Inc. 1997. Juvenile steelhead passage survival through flow deflector spillbays versus a non-flow deflector spillbay at Little Goose Dam, Snake River, Washington. Report prepared for Department of the Army, Walla Walla District COE, Walla Walla, WA.
- Normandeau Associates, J. R. Skalski, and Mid Columbia Consulting, Inc. 2003a. Juvenile salmonid survival and condition in passage through modified spillbays at Bonneville Dam, Columbia River. Report prepared for Department of the Army, Portland District, Portland, OR.
- Normandeau Associates, J. R. Skalski, and Mid Columbia Consulting, Inc. 2003b. Estimated direct mortality and injury rates of juvenile salmonids in passage through The Dalles Dam spillway, Columbia River, in spring and summer 2002. Report prepared for Department of the Army, Corps of Engineers, Portland District, Portland, OR.
- Pacific Northwest National Laboratory (PNNL), BioAnalysts, ENSR International, Inc., and Normandeau Associates, Inc. 2001. Design guidelines for high flow smolt bypass outfalls: field, laboratory and modeling studies. Report prepared for the U. S. Army Corps of Engineers, Portland District, Portland, OR.

Wilson, J. W., A. E. Giorgi, and L. C. Stuehrenberg. 1991. A method for estimating spill effectiveness for passing juvenile salmon and its application at Lower Granite Dam on the Snake River. Can. J. Fish. Aquat. Sci. 48:1872-1876.

TABLES

Daily spring (23 April to 2 May) and summer (13 to 15 July) releases of juvenile chinook salmon smolts and juvenile rainbow trout into Spillbay 5 and downstream of Spillbay 1 (spring controls) and through the juvenile fish facility bypass pipe (spring and summer controls) at Ice Harbor Dam, 2003.

* Pretest and not enough fish to include in survival or clean fish calculations.

Summary of physical conditions during release of juvenile salmon through Spillbay 5 at Ice Harbor Dam, April and May (spring) and July (summer), 2003. Hourly data are presented in Appendix A.

Spillbay release pattern (stop openings) followed during Spillbay 5 passage survival experiments at Ice Harbor Dam, April and May (spring) and July (summer) 2003. Boxed, bold numbers indicate experimental conditions (special spill on 2 May, dispersed spill on 14 July) which deviated from the preestablished values. One stop at a spillbay equals approximately 1,700 cfs.

* Dispersed spill, July 14.

** Special spill, May 2.

Required sample sizes (R) if control survival (S) is 0.99, 0.98, or 0.95, recapture rate (P_A) is 0.99, **0.98, or 0.95, and expected survival probability (**τˆ **) of treatment fish passed is 0.95, 0.97, and 0.99 to achieve a precision level (**ε**) of** ≤**±0.03 or** ≤**±0.05, 90% of the time. Highlighted values are discussed within the text.**

	Expected Survival $(\hat{\tau})$										
Control Survival (S)	0.95	0.97	0.99								
Precision (ε) $\leq \pm 0.03$											
		Recapture Rate=0.99									
0.99	256	205	150								
0.98	314	264	212								
0.95	496	451	405								
Recapture Rate=0.98											
0.99	314	264	218								
0.98	373	325	274								
0.95	556	514	469								
Recapture Rate=0.95											
0.99	496	451	405								
0.98	556	514	469								
0.95	745	709	670								
		Precision (ε) $\leq \pm 0.05$									
		Recapture Rate=0.99									
0.99	113	95	76								
0.98	134	117	99								
0.95	200	185	169								
Recapture Rate=0.98											
0.99	134	117	99								
0.98	156	139	122								
0.95	222	208	192								
Recapture Rate=0.95											
0.99	178	162	146								
0.98	200	185	169								
0.95	268	255	241								

Station			Water Temperature (°C)	Sample Size	Head (f ^t)	Test Spill		Recapture Rates (%)	Control	Passage
	Exit Route	Species				Volume (kcfs)	Control	Treatment	Survival (%)	Survival $(\%)$
The Dalles, WA	Spillway	Chinook salmon	$15 - 17$	270	81	10.5	97.0	94.1	97.0	95.5
	Spillway ^b	Chinook salmon	$15 - 17$	271	81	10.5	97.0	97.4	97.0	99.3
	Spillway ^b	Chinook salmon	$15 - 17$	210	81	$4.5\,$	96.2	94.3	96.2	99.0
	Spillway	Chinook salmon	$10 - 14$	391	75-80	$7.5 - 10.5$	98.7	96.7	98.0	97.4
	Spillway	Chinook salmon	$10-14$	396	75-80	$4.5 - 7.5$	98.7	95.4	98.0	97.4
	Spillway	Chinook salmon	$10 - 14$	405	75-80	$3.0 - 6.0$	98.7	93.8	98.0	93.8
Wanapum, WA	Sluice	Chinook salmon	$5 - 8$	195	79	$2.0\,$	100.0	97.9	100.0	97.4
	Spillway	Chinook salmon	$5 - 8$	235	79	4.3	100.0	99.6	99.6	99.6
	Spillway ^a	Chinook salmon	$5 - 8$	235	79	4.3	100.0	97.9	99.6	95.7
	Spillway ^b	Chinook salmon	$5 - 8$	155	79	$2.0\,$	100.0	97.4	100.0	92.0
	Spillway ^b	Chinook salmon	$5 - 8$	160	79	$4.0\,$	96.7	98.8	96.7	96.9
	Spillway	Chinook salmon	$17 - 18$	180	$82\,$	$2.8\,$	100.0	100.0	94.5	100.0
	Spillway	Chinook salmon	$17 - 18$	244	$82\,$	$6.0\,$	100.0	99.6	95.8	99.3
	Spillway	Chinook salmon	$17 - 18$	130	$82\,$	$11.5\,$	98.4	99.2	94.3	94.6
	Spillway ^a	Chinook salmon	$17 - 18$	200	$82\,$	$2.8\,$	100.0	100.0	96.5	99.0
	Spillway ^a	Chinook salmon	$17 - 18$	199	$82\,$	$6.0\,$	100.0	98.5	95.3	97.6
	Spillway ^a	Chinook salmon	$17 - 18$	191	$82\,$	11.5	98.4	96.7	94.3	92.8
	Spillway	Chinook salmon	16	180	$82\,$	$2.8\,$	100.0	100.0	97.5	99.4
	Spillway	Chinook salmon	$16\,$	169	$82\,$	$6.0\,$	100.0	100.0	95.8	97.6
	Spillway	Chinook salmon	16	198	$82\,$	$7.5\,$	100.0	100.0	94.3	99.5
	Spillway ^a	Chinook salmon	$16\,$	180	$82\,$	$2.8\,$	100.0	100.0	96.5	98.3
	Spillway ^a	Chinook salmon	$16\,$	170	$82\,$	$6.0\,$	100.0	98.8	95.3	98.2
	Spillway ^a	Chinook salmon	16	210	$82\,$	$7.5\,$	100.0	99.0	82.3	97.6
	Bypass Pipe	Chinook salmon	16	500	76-80	0.4	99.6	99.8	99.6	$100.0\,$
	Spillway ^{a,b}	Chinook salmon	$5-6$	300	81-82	10.4-12.5	100.0	99.0	97.3	99.0
Bonneville, WA	Spillway	Chinook salmon	$15 - 17$	280	60	12.0	96.1	96.8	96.1	$100.0\,$
	Spillway ^a	Chinook salmon	$15 - 17$	280	60	12.0	96.1	99.3	96.1	$100.0\,$
	Spillway ^a	Chinook salmon	$12 - 14$	130	54-58	$3.2 - 4.8$	100.0	97.7	97.7	97.9
	Spillway ^{a*}	Chinook salmon	$12 - 14$	166	54-58	$3.2 - 6.4$	100.0	95.8	97.7	95.9
	Spillway ^a	Chinook salmon	$12 - 14$	238	50-55	5.1-7.9	95.4	98.3	97.7	98.6
	Spillway ^{a*}	Chinook salmon	$12 - 14$	241	50-55	$7.1 - 9.8$	95.4	97.1	97.7	99.0
	Spillway ^a	Chinook salmon	$20 - 21$	166	60-65	$4.0 - 4.1$	86.9	83.7	82.6	90.5
	Spillway ^{a*}	Chinook salmon	$20 - 21$	175	60-65	$5.0 - 6.0$	86.9	88.1	82.6	88.6
	Spillway ^a	Chinook salmon	$20 - 21$	250	60-64	$5.0 - 6.0$	87.6	87.6	82.6	100.0
	Spillway ^{a*}	Chinook salmon	$20 - 21$	250	60-64	6.9-7.9	87.6	89.6	82.6	100.0

Sample size, recapture and control survival rates, and estimated 48 h survival (direct effects) of juvenile salmonids in passage through non-turbine exit routes at hydroelectric dams on the Columbia River Basin. Estimates based on balloon tag-recapture methodology (Heisey *et al.* **1992). Present study data are shown in italics.**

Continued.

a Spillbay with flow deflector.

a* Spillbay with deep flow deflector.

b Overflow weir or slot to attract surface oriented juvenile salmonids.

c Fish released into head pond vortices upstream of tainter gates.

d Spill directed onto concrete slab; survival is relative to survival at another spillbay.

e Periphery release.

1 Spring tests only.

Summary of fish transports from hatcheries and holding tank data collected during spillbay passage survival experiments at Ice Harbor Dam, April and May (spring) and July (summer) 2003.

* Fish with scale loss or symptoms of disease were culled.

Condition codes assigned to fish and dislodged balloon tags for fish passage survival evaluation.

FISH CODES

- **A** No visible marks on fish
- **B** Flesh tear at tag site(s)
- **C** Minor scale loss, 3 to 20% (%s for entire body in immediate recovery; for detailed injury examination %s are for section only)
- **D** Scale loss, >20% per side
- **E** Laceration(s); tear(s) on body
- **F** Severed body parts
- **G** Hemorrhaging, bruised
- **H** Stressed (lethargic, swimming poorly or sporadically)
- **I** Spasmodic movement of body
- **J** Very weak, barely gilling, died within 60 minutes of recovery
- **K** Failed to enter system
- **L** Fish likely preyed on based on telemetry, and/or circumstances relative to Turb'N recapture
- **M** Substantial bleeding at tag site
- **N** Bulging or missing eye(s)
- **P** Observed predator attack or marks indicative of predator
- **Q** Other information
- **R** Replaced due to entrapment in unrecoverable locations (i.e., in rocks, gate slot; recovery time expired)
- **T** Trapped inside tunnel/gate well
- **V** Fins damaged (ripped, split, torn) or pulled from origin
- **W** Abrasion/scrape
- **X** No recovery information at all; fish remains unrecovered
- **Z** Radio telemetry or other information; fish remains unrecovered

DISSECTION CODES

- **B** Swim bladder ruptured or expanded
- **D** Kidneys damaged (hemorrhaging)
- **E** Broken bones obvious
- **F** Hemorrhaging internally
- **L** Organ displacement
- **N** Heart damage, ruptured, hemorrhaging, etc.
- **O** Liver damage, ruptured, hemorrhaging, etc.
- **R** Necropsied, no obvious injuries
- **S** Necropsied, internal injuries observed
- **W** Head removed, i.e., otolith

TURB'N TAG CODES (not used in database)

- **A** Fully inflated
- **B** Partially inflated
- **C** Pinhole, leaking
- **D** Burst
- **E** Not inflated at all
Summary of tag-recapture data of juvenile chinook salmon released through shallow (7 ft above ogee) and deep (3 ft above ogee) release sites in Spillbay 5 at 50%, 100%, and special spill patterns and controls released downstream of Spillbay 1 or via the juvenile fish facility bypass pipe at Ice Harbor Dam, April and May 2003. Proportions given in parentheses.

	50% Spill		100% Spill		Special Spill		Control		
	Shallow*	Deep	Shallow	Deep	Shallow	Deep	Spillbay 1	Bypass Pipe	Combined
Number released	5	305	77	148	60	60	140	150	290
Number recaptured alive	5(1.000)	300 (0.984)	76 (0.987)	(0.993) 147	60(1.000)	59 (0.983)	139 (0.993)	149 (0.993)	288 (0.993)
Number recaptured dead	0(0.000)	4(0.013)	1(0.013)	0(0.000)	0(0.000)	1(0.017)	0(0.000)	0(0.000)	0(0.000)
Number assigned dead**	0(0.000)	1(0.003)	0(0.000)	0(0.000)	0(0.000)	0(0.000)	0(0.000)	(0.007)	1(0.003)
Unknown	0(0.000)	0(0.000)	0(0.000)	1(0.007)	0(0.000)	0(0.000)	1(0.007)	0(0.000)	1(0.003)
Number held	5	300	76	147	60	59	139	149	288
Number alive at 48 h	5	300	76	146	59	59	139	149	288

* Too few fish released for reliable survival estimation.

** Includes dislodged tags, predation, and stationary signals (see Section 2.6).

Summary of tag-recapture data on subyearling chinook salmon released through shallow (7 ft above ogee) and deep (3 ft above ogee) release sites in Spillbay 5 at bulk and dispersed spill patterns and controls released via the juvenile fish facility bypass pipe at Ice Harbor Dam, July 2003. Proportions given in parentheses. Data used for survival estimation. Juvenile rainbow trout were also released at bulk spill.

* Includes dislodged tags, predation and stationary signals (see Section 2.6).

** High mortality of control and treatment groups precluded reliable survial estimation.

*** Too few fish for reliable survival estimation; no concurrent controls were released.

Estimated 1 h and 48 h direct survival probabilities (τ^ˆ **) of juvenile chinook salmon smolts in passage through shallow (7 ft above ogee) and deep (3 ft above ogee) release sites in Spillbay 5 at Ice Harbor Dam, April and May (spring) and July (summer) 2003. The standard errors (bold and italicized) and 90% confidence intervals are shown in parentheses.**

* Only deep release survival calculated because of limited (only five fish) shallow releases (See Table 3-1).

Types of passage-related visible injuries observed on recaptured juvenile chinook salmon passed through Spillbay 5 at shallow (7 ft above ogee) and deep (3 ft above ogee) release sites at 50%, 100%, and special spill patterns and controls released downstream of Spillbay 1 or via the juvenile fish facility bypass pipe, Ice Harbor Dam, April and May 2003. Some fish had multiple injuries.

* Hemorrhaged, ruptured, bulging, or missing.

** Hemorrhage to gill, tears, scrapes, folds to operculum.

Types of passage-related visible injuries observed on recaptured juvenile chinook salmon passed through Spillbay 5 at shallow (7 ft above ogee) and deep (3 ft above ogee) release sites at bulk and dispersed spill and controls released via the juvenile fish facility bypass pipe, Ice Harbor Dam, July 2003. Juvenile rainbow trout were also released at bulk spill. Some fish had multiple injuries.

* Hemorrhaged, ruptured, bulging, or missing.

** Hemorrhage to gill, tears, scrapes, folds to operculum.

1 Majority of the spill (13.6k cfs) diverted through Spillbay 5, remaining spill concentrated in one or two adjacent spillbays.

2 3.4k cfs through Spillbay 5, remaining spill dispersed nearly equally among the other spillbays.

Summary malady data for juvenile chinook salmon released through shallow (7 ft above ogee) and deep (3 ft above ogee) release sites in Spillbay 5 at 50%, 100%, and special spill patterns and controls **released downstream of Spillbay 1 or via the juvenile fish facility bypass pipe at Ice Harbor Dam, April and May 2003.**

Summary malady rate data for juvenile chinook salmon and rainbow trout released through shallow (7 ft above ogee) and deep (3 ft above ogee) release sites in Spillbay 5 at bulk and dispersed spill and controls released via the juvenile fish facility bypass pipe at Ice Harbor Dam, July 2003.

Estimated clean fish probabilities (CFE) of juvenile chinook salmon smolts in passage through shallow (7 ft above ogee) and deep (3 ft above ogee) release sites in Spillbay 5 at Ice Harbor Dam, April and May (spring) and July (summer) 2003. The standard errors (bold and italicized) and 90% confidence intervals are shown in parentheses. Values given herein are based on the reduced model $(H_0: P_A = P_D)$.

* Only deep release calculated because of limited (only five fish) shallow released (See Table 3-6).

FIGURES

General location and layout of Ice Harbor Dam; showing treatment and release locations of fish during April/May and July, 2003.

Cross section of Spillbay 5, Ice Harbor Dam.

Cross section of spillbay showing release locations for juvenile salmon passed through Spillbay 5 at Ice Harbor Dam, April/May (spring) and July (summer), 2003.

Steel support pipe (16 inches diameter) and release hose (4 inches diameter) deployment to introduce fish approximately 3 ft (deep pipe) and 7 ft (shallow pipe) above the ogee of Spillbay 5, Ice Harbor Dam.

50% spill, 2.5stops, 4.2k cfs (Bay 5), 44k cfs (total) 100%, 5 stops, 8.5k cfs (Bay 5), 77.8k (total)

Special spill, 5stops, 8.5k cfs (Bay 5), 45k cfs (total) Bulk spill, 8stops, 13.6k cfs (Bay 5), 23k cfs (total)

Characteristics of the Spillbay 5 discharge upon interception of the flow deflector during the direct survival/injury tests conducted at Ice Harbor Dam, April/May and July, 2003.

Total length (mm) frequency distribution of treatment and control juvenile chinook Figure $2-4$ salmon released at 50%, 100% and special spills in April (May and of bulk and dispersed spills in July 2003 at Spillbay 5 of Ice Harbor Dam.

Attaching sensor fish package to adult rainbow trout while in stress reduction cylinder (top photo) and a trout (approximately 500 mm) equipped with HI-Z tags and sensor package (bottom photo).

Figure 3-1 Frequency distribution of recapture times (minutes) of treatment and control juvenile chinook salmon released at 50%, 100% and special spills in April/May and of bulk and dispersed spills in July 2003 at Spillbay 5 of Ice Harbor Dam.

TM7 - minor hemorrhage right eye LV7 – missing right eye

UJ3 – minor hemorrhage, ruptured pupil HE5 – major hemorrhage left eye

Figure 3-2

Examples of eye injuries to juvenile salmon passed through Spillbay 5 at Ice Harbor Dam, 2003.

FN7 - torn operculum HR4 - torn left operculum (attachment point)

Figure 3-3

Examples of opercular damage to juvenile salmon passed through Spillbay 5 at Ice Harbor Dam, 2003.

HX1 - damaged snout

Figure 3-4

Examples of bruises and cuts to juvenile salmon passed through Spillbay 5 at Ice Harbor Dam, 2003.

APPENDIX A

HYDRAULIC/PHYSICAL CONDITIONS DURING TESTING AND SUPPLY AND TAGGING TANK INFORMATION

Average forebay, tailwater, head (ft), and spill conditions at Spillbay 5 during conduct of passage survival and condition investigation with juvenile chinook salmon at Ice Harbor Dam, 23 April to 2 May and 13 to 15 July 2003. Bay 5 discharge (kcfs) obtained by multiplying stops by 1.7. Impact velocities to be provided by Duncan Hay.

Continued.

Key to condition number:

1 - 50% Spill

2 - 100% Spill

3 - Special Spill (more water diverted to Spillbay 5, see Table 2-3)

4 - Bulk Spill

5 - Dispersed Spill

6 - Control

 $^{\rm 1}$ Rainbow trout releases only

	Supply		Dead Fish	Number of	Dead Fish		Number	
	Tank	Number	Removed	Fish from	Removed	Fish Culled	Remaining	
	Temperature Transported		from	Supply to	from	from	for	
Date	(C)	to Supply	Supply	Tagging	Tagging	Study	Testing	Remarks/Comments
20 Apr	11.0	408	$\overline{0}$	θ	$\overline{0}$	$\overline{0}$		Steve Hays transported 408 chinook salmon from
								Leavenworth Hatchery to Ice Harbor Dam
22 Apr	11.0		2	100	$\mathbf{0}$	$\boldsymbol{0}$	306	Placed into tagging tank #1
23 Apr	11.0		$\mathbf{0}$	115	$\overline{0}$	$\boldsymbol{0}$	191	Placed into tagging tank #2
24 Apr	11.0		$\mathbf{0}$	120	$\overline{0}$	$\boldsymbol{0}$	71	Placed into tagging tank #1
25 Apr	10.5			70	θ	$\boldsymbol{0}$	$\boldsymbol{0}$	Supply tank cleaned out; Steve Hays will bring more fish on 4-26-03
26 Apr	10.5	800	$\boldsymbol{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	798	Steve Hays transported 800 chinook salmon from Leavenworth Hatchery to Ice Harbor Dam
26 Apr	10.5		2	160	$\mathbf{0}$	$\mathbf{0}$	636	Fish placed into tagging tank #1
26 Apr	10.5		$\boldsymbol{0}$	θ	$\mathbf{0}$	$\boldsymbol{0}$	606	30 fish escaped from supply tank by swimming out the
27 Apr	10.5							Did not work - Allowing fish to acclimate to ambient conditions
28 Apr	10.5		$\mathbf{2}$	100		$\boldsymbol{0}$	503	Fish placed into tagging tank #2
29 Apr	10.5		$\mathbf{1}$	Tank 1 - 135;	$\mathbf{0}$	$\mathbf{0}$	327	135 fish placed in tank #1 and 40 fish placed in tank
				Tank 2 - 40				#2
30 Apr								Remaining 327 fish were released into the river
11 Jul	20.0	$475 \pm$						Chinook salmon came from John Day Fish Facility
12 Jul	20.0			140		$\boldsymbol{0}$	401	160 fish placed in tagging tank #1 for testing tomorrow
13 Jul	20.0		2	180		5	198	Fish placed in tagging tank #2
14 Jul	20.0			150		10	26	Fish placed in tagging tank #1
15 Jul	20.5		θ	$\boldsymbol{0}$	0	10	$\boldsymbol{0}$	
11 Jul	20.0	500	θ	Ω	Ω	$\overline{0}$		500 rainbow trout came from Trout Lodge near Moses
								Lake, Washington
14 Jul	20.0			60 rainbow trout	$\boldsymbol{0}$	$\boldsymbol{0}$	440	Brought up 60 rainbow trout for testing on 7-15-03
15 Jul	20.5		Ω	Ω	Ω	θ	460	Took remaining 20 fish back down to supply tank

Supply and tagging tank fish disposition data for the Ice Harbor Dam spring (April and May) and summer (July) 2003 spillway passage survival study.

APPENDIX B

INDIVIDUAL TRIAL DATA AND FISH INJURY DATA

Incidence of injury, scale loss, and temporary loss of equilibrium (LOE) observed on juvenile chinook salmon passed through shallow (7 ft above ogee) and deep (3 ft above ogee) release sites in Spillbay 5 at 50%, 100%, and special spill patterns and controls released downstream of Spillbay 1 or via the fish bypass pipe, Ice Harbor Dam, April and May 2003.

Incidence of injury, scale loss, and temporary loss of equilibrium (LOE) observed on juvenile chinook salmon and rainbow trout passed through shallow (7 ft above ogee) and deep (3 ft above ogee) release sites in Spillbay 5 under bulk (13.6k cfs) and dispersed (3.4k cfs) patterns, Ice Harbor Dam, July 2003.

Continued.

* Not passage related

Continued.

* Fish assigned dead in survival estimation.

Daily tag-recapture data for juvenile chinook salmon and rainbow trout passed through shallow (7 ft above ogee) and deep (3 ft above ogee) release sites in Spillbay 5 at bulk and dispersed spill patterns and controls released via the juvenile fish facility bypass pipe at Ice Harbor Dam, July 2003. Numbers in parentheses indicate recovered fish with predation marks or unrecovered fish presumed ingested by a predator. \equiv

* Fish assigned dead in survival estimation.

Daily malady data for juvenile chinook salmon passed through shallow (7 ft above ogee, designated "S") and deep (3 ft above ogee, designated "D") release sites in Spillbay 5 at 50%, 100%, and special spill patterns and controls released downstream of Spillbay 1 or via the juvenile fish facility bypass pipe at Ice Harbor Dam, April and May 2003.

Continued.

Daily malady data for juvenile chinook salmon and rainbow trout passed through shallow (7 ft above ogee) and deep (3 ft above ogee) release sites in Spillbay 5 at bulk and dispersed spill patterns and controls released via the juvenile fish facility bypass pipe at Ice Harbor Dam, July 2003.

* Includes fish with maladies attributed to predators and/or tags (*i.e.* , tear at tag site).

APPENDIX C

DERIVATION OF PRECISION, SAMPLE SIZE, AND MAXIMUM LIKELIHOOD PARAMETERS, AND STATISTICAL OUTPUTS

DERIVATION OF PRECISION, SAMPLE SIZE, AND MAXIMUM LIKELIHOOD PARAMETERS

The statistical description below is excerpted from Normandeau Associates and Skalski (2000a). For the sake of brevity, references within the text have been removed. However, interested readers can look up these citations in the report prepared by Normandeau Associates and Skalski (2000a).

The estimation for the likelihood model parameters and sample size requirements discussed in the text are given herein. Additionally, the results of statistical analyses for evaluating homogeneity in recapture and survival probabilities, and in testing hypotheses of equality in parameter estimates under the simplified (H_O:P_A=P_D) versus the most generalized model (H_A:P_{A≠}P_D) are given.

The following terms are defined for the equations and likelihood functions which follow:

The precision of the estimate was defined as:

$$
P(-\varepsilon < \hat{\tau} - \tau < \varepsilon) = 1 - \alpha
$$

or equivalently

$$
P(-\varepsilon \langle \hat{\tau} - \tau | \langle \varepsilon \rangle = 1 - \alpha
$$

where the absolute errors in estimation, *i.e.*, $|\hat{\tau} - \tau|$, is $\leq \varepsilon (1-\alpha)$ 100% of the time, $\hat{\tau}$ is the estimated passage survival, and ϵ is the half-width of a (1- α) 100% confidence interval for $\hat{\tau}$ or 1- $\hat{\tau}$. A precision of $\pm 5\%$, 90% of the time is expressed as P($|\hat{\tau} - \tau| < 0.05$)=0.90.

Using the above precision definition and assuming normality of $\hat{\tau} - \tau$, the required total sample size (R) is as follows:

$$
P\left(\frac{-\varepsilon}{\sqrt{Var(\hat{\tau})}} < Z < \frac{\varepsilon}{\sqrt{Var(\hat{\tau})}}\right) = 1 - \alpha
$$

$$
P\left(Z < \frac{-\varepsilon}{\sqrt{Var(\hat{\tau})}}\right) = \alpha/2
$$
\n
$$
\Phi\left(\frac{-\varepsilon}{\sqrt{Var(\hat{\tau})}}\right) = \alpha/2
$$
\n
$$
\frac{-\varepsilon}{\sqrt{Var(\hat{\tau})}} = Z_{\alpha/2}
$$
\n
$$
Var(\hat{\tau}) = \frac{\varepsilon^2}{Z_{1-\frac{\alpha}{2}}^2}
$$
\n
$$
\frac{\tau}{SP_A} \left[\frac{(1 - S\tau P_A)}{R_T} + \frac{(1 - SP_A)\tau}{R_C} \right] = \frac{\varepsilon^2}{Z_{1-\frac{\alpha}{2}}^2}.
$$

where Z is a standard normal deviate satisfying the relationship $P(Z > Z_{1-\alpha/2}) = \alpha/2$, and Φ is the cumulative distribution function for a standard normal deviate.

If data can be pooled across trials and letting $R_C=R_T=R$, the sample size for each release is

$$
R = \frac{\tau}{SP_A} \Big[1 + \tau - 2S\tau P_A \Big] \frac{Z_{1-\alpha/2}^2}{\varepsilon^2} .
$$

By rearranging, this equation can be solved to predetermine the anticipated precision given the available number of fish for a study. In most previous investigations (Normandeau Associates and Skalski 2000a) this equation has been used to calculate sample sizes because of homogeneity between trials; in the present investigation sample size was predetermined using this equation.

If data cannot be pooled across trials the precision is based on

$$
\sum_{i=1}^{n} (1 - \hat{\tau}_i) / n = 1 - \sum_{i=1}^{n} \hat{\tau}_i / n = 1 - \overline{\hat{\tau}}.
$$

Precision is defined as

$$
P(|\overline{\hat{\tau}} - \overline{\tau}| < \varepsilon) = 1 - \alpha
$$

$$
P(-\varepsilon < \overline{\hat{\tau}} - \overline{\tau} \mid < \varepsilon) = 1 - \alpha
$$

$$
P\left(\frac{-\varepsilon}{\sqrt{Var(\overline{\hat{\tau}})}} < t_{n-1} < \frac{\varepsilon}{\sqrt{Var(\overline{\hat{\tau}})}}\right) = 1 - \alpha
$$

$$
P\left(t_{n-1} < \frac{-\varepsilon}{\sqrt{Var(\overline{\hat{\tau}})}}\right) = \alpha/2
$$
\n
$$
\Phi\left(\frac{-\varepsilon}{Var(\overline{\hat{\tau}})}\right) = \alpha/2
$$
\n
$$
\frac{-\varepsilon}{\sqrt{Var(\overline{\hat{\tau}})}} = t_{\alpha/2,n-1}
$$
\n
$$
Var(\overline{\hat{\tau}}) = \frac{\varepsilon^2}{t_{1-\alpha/2,n-1}^2}
$$
\n
$$
\frac{\sigma_{\tau}^2 + \frac{\tau}{SP_A} \left[\frac{(1 - S\tau P_A)}{R_{\tau}} + \frac{(1 - SP_A)\tau}{R_C} \right]}{n} = \frac{\varepsilon^2}{t_{1-\alpha/2,n-1}^2}
$$

where σ_t^2 =natural variation in passage-related mortality. Now letting $R_T=R_C$

$$
\frac{\sigma_{\tau}^2 + \frac{\tau}{SP_A} \left[\frac{(1 - STP_A)}{R} + \frac{(1 - SP_A)\tau}{R} \right]}{n} = \frac{\varepsilon^2}{t_{1 - \alpha/2, n - 1}^2}
$$

which must be iteratively solved for n given R. Or R given n where

$$
R = \frac{\frac{\tau}{SP_A} \left[(1 - S\tau P_A) + (1 - SP_A)\tau \right]}{\left[\frac{n\epsilon^2}{t_{1-\alpha/2,n-1}^2} - \sigma_\tau^2 \right]}
$$

$$
R = \frac{\tau (1 + \tau)}{\left[\frac{n\epsilon^2}{t_{1-\alpha/2,n-1}^2} - \sigma_\tau^2 \right]}
$$

$$
R = \frac{\tau (1 + \tau)}{SP_A} \left[\frac{t_{1-\alpha/2,n-1}^2}{n\epsilon^2 - \sigma_\tau^2 t_{1-\alpha/2,n-1}^2} \right].
$$

The joint likelihood for the passage-related mortality is:

$$
L (S, \tau, P_A, P_D | R_C, R_T, a_C, a_T, d_C, d_T) =
$$
\n
$$
\left(\begin{matrix} R_C \\ a_c d_C \end{matrix}\right) (SP_A)^{a_C} ((1 - S)P_D)^{d_C} (1 - SP_A - (1 - S)P_D)^{R_C - a_C - d_C}
$$
\n
$$
\times \left(\begin{matrix} R_T \\ a_T d_T \end{matrix}\right) (S \tau P_A)^{a_T} ((1 - S \tau)P_D)^{d_T} (1 - S \tau P_A - (1 - S \tau)P_D)^{R_T - a_T - d_T}.
$$

The likelihood model is based on the following assumptions: (1) fate of each fish is independent, (2) the control and treatment fish come from the same population of inference and share that same survival probability, (3) all alive fish have the same probability, P_A , of recapture, (4) all dead fish have the same probability, P_D , of recapture, and (5) passage survival (τ) and survival (S) to the recapture point are conditionally independent. The likelihood model has four parameters $(P_A, P_D, S, \mathcal{L})$ τ) and four minimum sufficient statistics (a_C , d_C , a_T , d_T).

Because any two treatment releases were made concurrently with a single shared control group we used the likelihood model which took into account dependencies within the study design (Normandeau Associates *et al.* 1995). For any two treatment groups (denoted T_1 and T_2), the likelihood model is as follows:

$$
L(S, \tau_1, \tau_2, P_A, P_D | R_C, R_{T_1}, R_{T_2}, a_C, d_c, a_{T_1}, d_{T_1}, a_{T_2}, d_{T_2}) =
$$
\n
$$
\binom{R_C}{a_{c}d_C} (SP_A)^{a_C} ((1 - S)P_D)^{d_C} (1 - SP_A - (1 - S)P_D)^{R_C - a_C - d_C}
$$
\n
$$
\times \binom{R_{T_1}}{a_{T_1}d_{T_1}} (ST_A P_A)^{a_{T_1}} ((1 - ST_A)P_D)^{d_{T_1}} (1 - ST_A P_A - (1 - ST_A)P_D)^{R_{T_1} - a_{T_1} - d_{T_1}}
$$
\n
$$
\times \binom{R_{T_2}}{a_{T_2}d_{T_2}} (ST_A P_A)^{a_{T_2}} ((1 - ST_A)P_D)^{d_{T_2}} (1 - ST_A P_A - (1 - ST_A)P_D)^{R_{T_2} - a_{T_2} - d_{T_2}}.
$$

This likelihood model has the same assumptions as stated in Normandeau Associates and Skalski (2000a) but has five estimable parameters (S, τ_1 , τ_2 , P_A, and P_D). The survival rate for treatment T₁ is estimated by τ_1 and for treatment T₂, by τ_2 . A likelihood ratio test with 1 degree of freedom was used to test for equality in survival rates between treatments τ_1 and τ_2 based on the hypothesis H_O: $\tau_1 = \tau_2$ versus H_a: $\tau_1 \neq \tau_2$.

Likelihood models are based on the following assumptions: (a) the fate of each fish is independent; (b) the control and treatment fish come from the same population of inference and share the same natural survival probability, S; (c) all alive fish have the same probability, P_A , of recapture; (d) all dead fish have the same probability, P_D , of recapture; and (e) passage survival (τ) and natural survival (S) to the recapture point are conditionally independent.

The estimators associated with the likelihood model are:

$$
\hat{\tau} = \frac{a_r R_c}{R_r a_c}
$$

$$
\hat{S} = \frac{R_r d_c a_c - R_c d_r a_c}{R_c d_c a_r - R_c d_r a_c}
$$
\n
$$
\hat{P}_A = \frac{d_c a_r - d_r a_c}{R_r d_c - R_c d_r}
$$
\n
$$
\hat{P}_D = \frac{d_c a_r - d_r a_c}{R_c a_r - R_r a_c}
$$

The variance (Var) and standard error (SE) of the estimated passage mortality (*1-*τˆ) or survival (τˆ) are:

$$
Var(1-\hat{\tau}) = Var(\hat{\tau}) = \frac{\tau}{SP_A} \left[\frac{(1 - S\tau P_A)}{R_T} + \frac{(1 - SP_A)\tau}{R_C} \right]
$$

$$
SE(1-\hat{\tau}) = SE(\hat{\tau}) = \sqrt{Var(1-\hat{\tau})}.
$$

DERIVATION OF VARIANCE FOR WEIGHTED AVERAGE SURVIVAL ESTIMATE

The variance of a weighted average is estimated by the formula

$$
\hat{\overline{\theta}}_{W} = \frac{\sum_{i=1}^{n} W_{i} \hat{\theta}_{i}}{\sum_{i=1}^{n} W_{i}}
$$

with

$$
\widehat{\text{Var}}\left(\widehat{\overline{\theta}}_w\right) = \frac{\sum_{i=1}^n W_i \left(\widehat{\theta}_i - \widehat{\overline{\theta}}_w\right)^2}{(n-1)\sum_{i=1}^n W_i}
$$

where $\hat{\theta}_w$ = the weighted average,

 $\hat{\theta}_i$ = the parameter estimate for the *i*th replicate,

 W_i = weight.

Chi square tests of homogeneity for the recovery of chinook salmon juveniles released through units at the Ice Harbor Dam, April/May 2003.

---------------------------------- spill=_100 --- The FREQ Procedure Table of release by cond release cond Frequency , Expected ' Cell Chi-Square, alive , dead , unkn , Total deep 100 1 , 20, 0, 0, 20 $\,$, $\,$ 19.822 $\,$, 0.0889 $\,$, 0.0889 $\,$, $\,$ $, 0.0016, 0.0889, 0.0889,$ deep_100 2 , 29, 0, 0, 29 , 28.742 , 0.1289 , 0.1289 , $, 0.0023, 0.1289, 0.1289,$ deep_100 3 , 20, 0, 0, 20 , 19.822 , 0.0889 , 0.0889 , $, 0.0016$, 0.0889 , 0.0889 , $\begin{matrix} \text{deep_100} & 6 & \text{\textit{i}} & 20 \text{\textit{i}} & 0 \text{\textit{i}} & 0 \text{\textit{i}} & 20 \end{matrix}$ $, 19.822, 0.0889, 0.0889,$ $, 0.0016$, 0.0889 , 0.0889 , deep_100 $\,$ 7 $\,$, $\,$ 58 $\,$, 0 $\,$, 1 $\,$, 59 $, 58.476$, 0.2622 , 0.2622 , $, 0.0039, 0.2622, 2.0758,$ $\verb+shallow_1001+, \qquad 28~, \qquad \quad 0~, \qquad \quad 0~, \qquad \quad 28$ $, 27.751$, 0.1244 , 0.1244 , $, 0.0022, 0.1244, 0.1244,$ $\texttt{shallow_1002} \quad , \qquad \texttt{19} \ , \qquad \quad \texttt{0} \ , \qquad \quad \texttt{0} \ , \qquad \quad \texttt{19}$, 18.831 , 0.0844 , 0.0844 , $, 0.0015$, 0.0844 , 0.0844 , shallow 1003 , 29 , 1 , 0 , 30 $\,$, 29.733 , 0.1333 , 0.1333 , $, 0.0181$, 5.6333 , 0.1333 , Total 223 1 1 225 Statistics for Table of release by cond Statistic $\begin{array}{ccccccc} \text{Statistic} & & & \text{DF} & & \text{Value} & & \text{Prob} \\ \text{Chi-Square} & & & 14 & & 9.3464 & & 0.8082 \end{array}$ Chi-Square 14 9.3464

Likelihood Ratio Chi-Square 14 6.7399

Mantel-Haenszel Chi-Square 1 0.6812 Likelihood Ratio Chi-Square 14 6.7399 0.9443 Mantel-Haenszel Chi-Square 1 0.6812 0.4092 Phi Coefficient 0.2038 Contingency Coefficient 0.1997

WARNING: 67% of the cells have expected counts less than 5. Chi-Square may not be a valid test.

Cramer's V 0.1441

 The FREQ Procedure Statistics for Table of release by cond Fisher's Exact Test Table Probability (P) 0.0351 $Pr \le P$ 0.9321 Sample Size = 225

Chi square tests of homogeneity for the recovery of chinook salmon juveniles released through units at the Ice Harbor Dam, April/May 2003.

-------------------------------- spill=_50 -- The FREQ Procedure Table of release by cond release cond Frequency , Expected ' Cell Chi-Square, alive , dead , Total deep 50 1 , 5 , 0 , 5 $, 4.9194$, 0.0806 , , 0.0013 , 0.0806 , deep_50 2 , 97, 3, 100 $, 98.387$, 1.6129 , $, 0.0196, 1.1929,$ deep_50 8 , 100, 0, 100 $(98.387 , 1.6129)$ $\,$, 0.0264 , 1.6129 , deep 50 9 , 98 , 2 , 100 $(98.387 , 1.6129 ,$ $, 0.0015 , 0.0929 ,$ shallow 50 1 $\,$ $\,$ $\,$ 5 $\,$ $\,$ 0 $\,$ $\,$ 5 $\,$ $, 4.9194$, 0.0806 , $\,$, 0.0013 , 0.0806 , 0.0013 , 0.0806 , $\,$ Total 305 5 310 Statistics for Table of release by cond Statistic DF Value Prob

WARNING: 70% of the cells have expected counts less than 5. Chi-Square may not be a valid test.

Fisher's Exact Test

Chi square tests of homogeneity for the recovery of chinook salmon juveniles released through units at the Ice Harbor Dam, April/May 2003.

------------------------ desc=deep_100 --- The FREQ Procedure Table of release by cond release cond Frequency , Expected , Cell Chi-Square, alive , unkn , Total deep 100 1 , 20, 0, 20 $\,$, 19.865 , 0.1351 , $, 0.0009, 0.1351,$ deep_100 2 , 29, 0, 29 , 28.804 , 0.1959 , $, 0.0013, 0.1959,$ deep_100 3 , 20, 0, 20 $, 19.865$ $, 0.1351$ $,$ $, 0.0009, 0.1351,$ deep 100 6 , 20 , 0, 20 $, 19.865 , 0.1351$, $, 0.0009, 0.1351,$ $\begin{tabular}{ccccc} deep_100 & 7 & \text{\textit{1}} & 58 & \text{\textit{1}} & 1 & 59 \end{tabular}$ $, 58.601$, 0.3986 , $, 0.0062, 0.9071,$ Total 147 1 148 Statistics for Table of release by cond

WARNING: 50% of the cells have expected counts less than 5. Chi-Square may not be a valid test.

Fisher's Exact Test

Chi square tests of homogeneity for the recovery of chinook salmon juveniles released through units at the Ice Harbor Dam, April/May 2003.

------------------------ desc=deep_50 --- The FREQ Procedure Table of release by cond release cond Frequency , Expected , Cell Chi-Square, alive , dead , Total deep 50 1 , 5 , 0 , 5 $\,$, $\,$ 4.918 , 0.082 , $(0.0014 \; , \; 0.082 \; , \;$ deep_50 2 , 97, 3, 100 $, 98.361$, 1.6393 , $, 0.0188, 1.1293,$ deep_50 8 , 100, 0, 100 $, 98.361$, 1.6393 , $, 0.0273, 1.6393,$ deep 50 9 , 98 , 2 , 100 $, 98.361, 1.6393$, $, 0.0013, 0.0793,$ Total 300 5 305 Statistics for Table of release by cond

WARNING: 63% of the cells have expected counts less than 5. Chi-Square may not be a valid test.

Fisher's Exact Test

Chi square tests of homogeneity for the recovery of chinook salmon juveniles released through units at the Ice Harbor Dam, April/May 2003.

---------------------- desc=shallow_100 ---

release cond

 The FREQ Procedure Table of release by cond

Total 76 1 77

Statistics for Table of release by cond

WARNING: 50% of the cells have expected counts less than 5. Chi-Square may not be a valid test.

Fisher's Exact Test

Chi square tests of homogeneity for the recovery of controls chinook salmon juveniles released through the by-pass pipe and below Spillbay 1 at the Ice Harbor Dam, April/May 2003.

Chi square tests of homogeneity for the recovery of controls chinook salmon juveniles released through the by-pass pipe and below Spillbay 1 at the Ice Harbor Dam, April/May 2003.

Statistics for Table of release by cond

WARNING: 67% of the cells have expected counts less than 5. Chi-Square may not be a valid test.

The FREQ Procedure

Statistics for Table of release by cond

Fisher's Exact Test

Chi square tests of homogeneity of recovery of juvenile chinook salmon released through different spillbays at the Ice Harbor Dam, July 2003.

WARNING: 33% of the cells have expected counts less than 5. Chi-Square may not be a valid test.

Cramer's V 0.0734

One hour survival rates for juvenile chinook salmon released through Spillbay 5 at deep release sites at 50% and 100% spill patterns at Ice Harbor Dam, April-May 2003. Controls: 290 released, 288 alive, 1 dead. 50% spill test: 305 released, 5 dead. 100% spill test: 148 released, 147 alive, 0 dead. (Values from Table 3-1).

RESULTS FOR REDUCED MODEL (EQUAL LIVE/DEAD RECOVERY)

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -46.0103

Tau = $0.9870(0.0081)$ 50% spill, deep release/Control ratio Tau = $1.0035(0.0035)$ 100% spill, deep release/Control ratio

Z statistic for the equality of equal turbine survivals: 1.8742

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00001193 0.00000000 0.00000000 0.00000000 0.00000000 0.00000361 0.00000000 0.00000000 0.00000000 0.00000000 0.00005287 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000

Confidence intervals:

Likelihood ratio statistic for equality of recovery probabilities: 0.0256

Compare with quantiles of the chi-squared distribution with 1 d.f.:

Forty-eight hour survival rates for juvenile chinook salmon released through Spillbay 5 at deep release sites at 50% and 100% spill patterns at Ice Harbor Dam, April-May 2003. Controls: 290 released, 288 alive, 1 dead. 50% spill test: 305 released , 5 dead. 100% spill test: 148 released, 146 alive, 1 dead. (Values from Table 3-1).

```
RESULTS FOR REDUCED MODEL (EQUAL LIVE/DEAD RECOVERY)
```


* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -51.9973

Tau = $0.9870(0.0081)$ 50% spill, deep release/Control ratio Tau = $0.9966(0.0076)$ 100% spill, deep release/Control ratio

Z statistic for the equality of equal turbine survivals: 0.8672

Compare with quantiles of the normal distribution: 1-tailed 2-tailed

Variance-Covariance matrix for estimated probabilities: 0.00001193 0.00000000 0.00000000 0.00000000 0.00000000 0.00000361 0.00000000 0.00000000 0.00000000 0.00000000 0.00005287 0.00000000 0.00000000 0.00000000 0.00000000 0.00004596

Confidence intervals: 50% spill, deep release Tau 100% spill, deep release Tau
90 percent: (0.9738, 1.0003) (0.9841, 1.0092) 90 percent: (0.9738, 1.0003) 95 percent: (0.9712, 1.0028) (0.9817, 1.0116) 99 percent: (0.9663, 1.0078) (0.9770, 1.0163)

== Likelihood ratio statistic for equality of recovery probabilities: 0.0174

Compare with quantiles of the chi-squared distribution with 1 d.f.:

One hour survival rates for juvenile chinook salmon released through Spillbay 5 at deep release sites at 100% and special spill patterns at Ice Harbor Dam, April-May 2003. Controls: 290 released, 288 alive, 1 dead. Special spill test: 60 released , 59 alive, 1 dead. 100% spill test: 148 released, 147 alive, 0 dead. (Values from Table 3-1).

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -24.7815

Tau = $1.0035(0.0035)$ 100% spill, deep release/Control ratio Tau = $0.9867(0.0169)$ Special spill, deep release/Control ratio

Z statistic for the equality of equal turbine survivals: 0.9675

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00001193 0.00000000 0.00000000 0.00000000 0.00000000 0.00000803 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00027315

Confidence intervals: 100% spill, deep release Tau Special spill, deep release Tau 90 percent: $(0.9978, 1.0092)$

Likelihood ratio statistic for equality of recovery probabilities: 0.0092

Compare with quantiles of the chi-squared distribution with 1 d.f.:

Forty-eight hour survival rates for juvenile chinook salmon released through Spillbay 5 at deep release sites at 100% and special spill patterns at Ice Harbor Dam, April-May 2003. Controls: 290 released, 288 alive, 1 dead. Special spill test: 60 released , 59 alive, 1 dead. Deep release test: 148 released, 146 alive, 1 dead. (Values from Table 3-1).

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -30.7686

Tau = $0.9966 (0.0076) 100\%$ spill, deep release/Control ratio
Tau = $0.9867 (0.0169)$ special spill, deep release/Control ratio 0.9867 (0.0169) special spill, deep release/Control ratio

Z statistic for the equality of equal turbine survivals: 0.5329

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00001193 0.00000000 0.00000000 0.00000000 0.00000000 0.00000803 0.00000000 0.00000000 0.00000000 0.00000000 0.00004596 0.00000000 0.00000000 0.00000000 0.00000000 0.00027315

Confidence intervals:

Likelihood ratio statistic for equality of recovery probabilities: 0.0037

Compare with quantiles of the chi-squared distribution with 1 d.f.:

One hour survival rates for juvenile chinook salmon released through Spillbay 5 at deep and shallow release sites at 100% spill pattern at Ice Harbor Dam, April-May 2003. Controls: 290 released, 288 alive, 1 dead. Shallow test: 77 released, 76 alive, 1 dead. Deep release test: 148 released, 147 alive, 0 dead. (Values from Table 3-1).

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -25.1001

Z statistic for the equality of equal turbine survivals: 0.9417

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00001193 0.00000000 0.00000000 0.00000000 0.00000000 0.00000751 0.00000000 0.00000000 0.00000000 0.00000000 0.00016647 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000

Confidence intervals: 100% spill, shallow release Tau 100% spill, deep release Tau 90 percent: (0.9684, 1.0125) (0.9978, 1.0092) 95 percent: (0.9642, 1.0167) (0.9967, 1.0103) 99 percent: (0.9559, 1.0249) (0.9945, 1.0124) == Likelihood ratio statistic for equality of recovery probabilities: 0.0087

Compare with quantiles of the chi-squared distribution with 1 d.f.:

Forty-eight hour survival rates for juvenile chinook salmon released through Spillbay 5 at shallow and deep release sites at 100% spill pattern at Ice Harbor Dam, April-May 2003. Controls: 290 released, 288 alive, 1 dead. Shallow release test: 77 released , 76 alive, 1 dead. Deep release test: 148 released, 146 alive, 1 dead. (Values from Table 3-1).

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -31.0872

Tau = $0.9904 (0.0134)$ 100% spill, shallow release/Control ratio Tau = $0.9966(0.0076)$ 100% spill, deep release/Control ratio

Z statistic for the equality of equal turbine survivals: 0.4026

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00001193 0.00000000 0.00000000 0.00000000 0.00000000 0.00000751 0.00000000 0.00000000 0.00000000 0.00000000 0.00016647 0.00000000 0.00000000 0.00000000 0.00000000 0.00004596

Confidence intervals:

Likelihood ratio statistic for equality of recovery probabilities: 0.0029

Compare with quantiles of the chi-squared distribution with 1 d.f.:

One hour survival rates for juvenile chinook salmon released through Spillbay 5 at shallow and deep release sites at special spill patterns at Ice Harbor Dam, April-May 2003. Controls: 290 released, 288 alive, 1 dead. Deep release test: 60 released, 59 alive, 1 dead. Shallow test: 60 released, 60 alive, 0 dead. (Values from Table 3-1).

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -18.7656

Tau = 1.0035 (0.0035) Special spill, shallow release/Control ratio
Tau = 0.9867 (0.0169) Special spill, deep release/Control ratio 0.9867 (0.0169) Special spill, deep release/Control ratio

Z statistic for the equality of equal turbine survivals: 0.9675

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00001193 0.00000000 0.00000000 0.00000000 0.00000000 0.00000593 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00027315

Confidence intervals: Special spill, shallow release Tau Special spill, deep release Tau 90 percent: (0.9978, 1.0092) (0.9589, 1.0146) 95 percent: (0.9967, 1.0103) (0.9536, 1.0199) 99 percent: (0.9945, 1.0124) (0.9431, 1.0304) == Likelihood ratio statistic for equality of recovery probabilities: 0.0029

Compare with quantiles of the chi-squared distribution with 1 d.f.:

Forty-eight hour survival rates for juvenile chinook salmon released through Spillbay 5 at shallow and deep release sites at a special spill pattern at Ice Harbor Dam, April-May 2003. Controls: 290 released, 288 alive, 1 dead. Shallow release test: 60 released , 59 alive, 1 dead. Deep release test: 60 released, 59 alive, 1 dead. (Values from Table 3-1).

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -23.8516

Tau = 0.9867 (0.0169) Special spill, shallow release/Control ratio Tau = $0.9867(0.0169)$ Special spill, deep release/Control ratio

Z statistic for the equality of equal turbine survivals: 0.0000

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00001193 0.00000000 0.00000000 0.00000000 0.00000000 0.00000593 0.00000000 0.00000000 0.00000000 0.00000000 0.00027315 0.00000000 0.00000000 0.00000000 0.00000000 0.00027315

Confidence intervals: Special spill, shallow release Tau Special spill, deep release Tau 90 percent: (0.9589, 1.0146) (0.9589, 1.0146) 95 percent: (0.9536, 1.0199) (0.9536, 1.0199) 99 percent: (0.9431, 1.0304) (0.9431, 1.0304) == Likelihood ratio statistic for equality of recovery probabilities: 0.0078

Compare with quantiles of the chi-squared distribution with 1 d.f.:

One hour survival estimates of juvenile chinook salmon smolts passed through Spillbay 5 at shallow and deep release sites and 100% spill at Ice Harbor Dam, April/May, 2003. Controls: 290 released, 288 alive, 1 dead. Test fish: 225 released, 223 alive, 1 dead.

RESULTS FOR FULL MODEL (UNEQUAL LIVE/DEAD RECOVERY)

 estim. std. err. $S = 0.9931 (0.0000)$ Control group survival $Pa =$ 1.0 N/A Live recovery probability* $Pd = 0.5000 (0.2500)$ Dead recovery probability Tau = $0.9980(0.0080)$ Spillbay 5 survival $1-Tau = 0.0020 (0.0080)$ Spillbay 5 mortality

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -26.156135

Variance-Covariance matrix for estimated probabilities: -0.00008 0.00000 -0.00002 0.00000 0.06250 0.00000 0.00008 0.00000 0.00006

Profile likelihood intervals: Spillbay 5 survival Spillbay 5 mortality 90 percent: (0.9821, 1.0000) (0.0000, 0.0179) 95 percent: (0.9782, 1.0000) (0.0000, 0.0218) 99 percent: (0.9697, 1.0000) (0.0000, 0.0303)

RESULTS FOR REDUCED MODEL (EQUAL LIVE/DEAD RECOVERY)

==

 estim. std. err. $S = 0.9965 (0.0035)$ Control group survival $Pa = Pd \quad 0.9961 \quad (0.0027)$ Recovery probability Tau = $0.9990(0.0057)$ Spillbay 5 survival 1-Tau = $0.0010(0.0057)$ Spillbay 5 mortality

log-likelihood : -26.172256

Variance-Covariance matrix for estimated probabilities: 0.00001 0.00000 -0.00001 0.00000 0.00001 -0.00000 -0.00001 -0.00000 0.00003

Profile likelihood intervals: Spillbay 5 survival Spillbay 5 mortality 90 percent: (0.9865, 1.0000) (0.0000, 0.0135) 95 percent: (0.9832, 1.0000) (0.0000, 0.0168) 99 percent: (0.9755, 1.0000) (0.0000, 0.0245) ==

Likelihood ratio statistic for equality of recovery probabilities: 0.032242

Compare with quantiles of the chi-squared distribution with 1 d.f.: For significance level 0.10: 2.706 For significance level 0.05: 3.841

For significance level 0.01: 6.635

Forty-eight hour survival estimates of juvenile chinook salmon smolts passed through Spillbay 5 at shallow and deep release sites and 100% spill at Ice Harbor Dam, April/May, 2003. Controls: 290 released, 288 alive, 1 dead. Test fish: 225 released, 222 alive, 2 dead.

RESULTS FOR FULL MODEL (UNEQUAL LIVE/DEAD RECOVERY) estim. std. err.

log-likelihood : -31.174767

Variance-Covariance matrix for estimated probabilities: 0.00003 -0.00003 0.00308 -0.00000 -0.00003 0.00004 -0.00433 -0.00002 0.00308 -0.00433 0.56368 0.00287 -0.00000 -0.00002 0.00287 0.00008

Profile likelihood intervals:

==

RESULTS FOR REDUCED MODEL (EQUAL LIVE/DEAD RECOVERY)

 estim. std. err. $S = 0.9965 (0.0035)$ Control group survival $Pa = Pd \quad 0.9961 \quad (0.0027)$ Recovery probability Tau = $0.9945(0.0072)$ Spillbay 5 survival 1-Tau = $0.0055(0.0072)$ Spillbay 5 mortality

log-likelihood : -31.190888

Variance-Covariance matrix for estimated probabilities: 0.00001 -0.00000 -0.00001 -0.00000 0.00001 0.00000 -0.00001 0.00000 0.00005

Profile likelihood intervals:

== Likelihood ratio statistic for equality of recovery probabilities: 0.032242

Compare with quantiles of the chi-squared distribution with 1 d.f.:

 For significance level 0.10: 2.706 For significance level 0.05: 3.841

For significance level 0.01: 6.635

One hour survival rates for juvenile chinook salmon released through Spillbay 5 at shallow release sites at 100% and special spill patterns at Ice Harbor Dam, April-May 2003. Controls: 290 released, 288 alive, 1 dead. 100% spill test: 77 released, 76 alive, 1 dead. Special spill test: 60 released, 60 alive, 0 dead. (Values from Table 3-1).

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -19.0576

Tau = 0.9904 (0.0134) 100% spill, shallow release/Control ratio Tau = 1.0035 (0.0035) Special spill, shallow release/Control ratio

Z statistic for the equality of equal turbine survivals: 0.9417

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00001193 0.00000000 0.00000000 0.00000000 0.00000000 0.00000547 0.00000000 0.00000000 0.00000000 0.00000000 0.00016647 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000

Confidence intervals:

Likelihood ratio statistic for equality of recovery probabilities: 0.0025

Compare with quantiles of the chi-squared distribution with 1 d.f.:

Forty-eight hour survival rates for juvenile chinook salmon released through Spillbay 5 at shallow release sites at 100% and special spill patterns at Ice Harbor Dam, April-May 2003. Controls: 290 released, 288 alive, 1 dead. 10% spill test: 77 released, 76 alive, 1 dead. Special spill test: 60 released, 59 alive, 1 dead. (Values from Table 3-1).

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -24.1436

Tau = 0.9904 (0.0134) 100% spill, shallow release/Control ratio Tau = $0.9867(0.0169)$ Special spill, shallow release/Control ratio

Z statistic for the equality of equal turbine survivals: 0.1710

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00001193 0.00000000 0.00000000 0.00000000 0.00000000 0.00000547 0.00000000 0.00000000 0.00000000 0.00000000 0.00016647 0.00000000 0.00000000 0.00000000 0.00000000 0.00027315

Confidence intervals:

Likelihood ratio statistic for equality of recovery probabilities: 0.0072

Compare with quantiles of the chi-squared distribution with 1 d.f.:

One hour survival estimates of juvenile chinook salmon smolts passed through Spillbay 5 at shallow and deep release sites and special spill at Ice Harbor Dam, April/May, 2003. Controls: 290 released, 288 alive, 1 dead. Test fish: 120 released, 119 alive, 1 dead.

RESULTS FOR FULL MODEL (UNEQUAL LIVE/DEAD RECOVERY)

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -19.461509

Variance-Covariance matrix for estimated probabilities: 0.00001 -0.00000 -0.00001 -0.00000 0.00001 0.00000 -0.00001 0.00000 0.00008

Profile likelihood intervals: Spillbay 5 survival Spillbay 5 mortality 90 percent: (0.9729, 1.0000) (0.0000, 0.0271) 95 percent: (0.9667, 1.0000) (0.0000, 0.0333) 99 percent: (0.9528, 1.0000) (0.0000, 0.0472)

RESULTS FOR REDUCED MODEL (EQUAL LIVE/DEAD RECOVERY)

==

 estim. std. err. $S = 0.9965 (0.0035)$ Control group survival $Pa = Pd \quad 0.9976 \quad (0.0024)$ Recovery probability Tau = $0.9951(0.0090)$ Spillbay 5 survival 1-Tau = $0.0049(0.0090)$ Spillbay 5 mortality

log-likelihood : -19.462945

Variance-Covariance matrix for estimated probabilities: 0.00001 0.00000 -0.00001 0.00000 0.00001 -0.00000 -0.00001 -0.00000 0.00008

Profile likelihood intervals: Spillbay 5 survival Spillbay 5 mortality 90 percent: (0.9729, 1.0000) (0.0000, 0.0271) 95 percent: (0.9667, 1.0000) (0.0000, 0.0333) 99 percent: (0.9528, 1.0000) (0.0000, 0.0472) ==

Likelihood ratio statistic for equality of recovery probabilities: 0.002872

Compare with quantiles of the chi-squared distribution with 1 d.f.: For significance level 0.10: 2.706 For significance level 0.05: 3.841

For significance level 0.01: 6.635

Forty-eight hour survival estimates of juvenile chinook salmon smolts passed through Spillbay 5 at shallow and deep release sites and special spill at Ice Harbor Dam, April/May, 2003. Controls: 290 released, 288 alive, 1 dead. Test fish: 120 released, 118 alive, 2 dead.

RESULTS FOR FULL MODEL (UNEQUAL LIVE/DEAD RECOVERY)

 estim. std. err. $S = 0.9966 (0.0034)$ Control group survival $Pa = 0.9975 (0.0025)$ Live recovery probability $Pd = 1.0 \qquad N/A \qquad$ Dead recovery probability* Tau = $0.9867(0.0122)$ Spillbay 5 survival 1-Tau = $0.0133(0.0122)$ Spillbay 5 mortality

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -23.847668

Variance-Covariance matrix for estimated probabilities: 0.00001 -0.00000 -0.00001 -0.00000 0.00001 -0.00000 -0.00001 -0.00000 0.00015

Profile likelihood intervals:

 Spillbay 5 survival Spillbay 5 mortality 90 percent: (0.9594, 1.0000) (0.0000, 0.0406) 95 percent: (0.9523, 1.0000) (0.0000, 0.0477) 99 percent: (0.9369, 1.0000) (0.0000, 0.0631)

== **RESULTS FOR REDUCED MODEL (EQUAL LIVE/DEAD RECOVERY)**

 estim. std. err. $S = 0.9965 (0.0035)$ Control group survival $Pa = Pd \quad 0.9976 \quad (0.0024)$ Recovery probability Tau = $0.9867(0.0122)$ Spillbay 5 survival 1-Tau = $0.0133(0.0122)$ Spillbay 5 mortality

log-likelihood : -23.851560

Variance-Covariance matrix for estimated probabilities: 0.00001 0.00000 -0.00001 0.00000 0.00001 -0.00000 -0.00001 -0.00000 0.00015

Profile likelihood intervals: Spillbay 5 survival Spillbay 5 mortality 90 percent: (0.9594, 1.0000) (0.0000, 0.0406)
95 percent: (0.9523, 1.0000) (0.0000, 0.0477) 95 percent: (0.9523, 1.0000) 99 percent: (0.9369, 1.0000) (0.0000, 0.0631) ==

Likelihood ratio statistic for equality of recovery probabilities: 0.007786

Compare with quantiles of the chi-squared distribution with 1 d.f.: For significance level 0.10: 2.706

 For significance level 0.05: 3.841 For significance level 0.01: 6.635
One hour survival rates for juvenile chinook salmon released through Spillbay 5 at deep release sites at bulk and dispersed spill patterns at Ice Harbor Dam, July 2003. Controls: 125 released, 124 alive, 1 dead. Bulk spill test: 60 released, 54 alive, 5 dead. Dispersed spill test: 100 released, 96 alive, 3 dead. (Values from Table 3-2).

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -48.3019

Tau = 0.9775 (0.0191) Dispersed spill, deep release/Control ratio Tau = $(0.9226)(0.0373)$ Bulk spill, deep release/Control ratio

Z statistic for the equality of equal spill pattern survivals: 1.3104

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00006349 0.00000000 0.00000000 0.00000000 0.00000000 0.00002445 0.00000000 0.00000000 0.00000000 0.00000000 0.00029682 0.00000000 0.00000000 0.00000000 0.00000000 0.00131464

Confidence intervals: Dispersed spill, deep release Tau Bulk spill, deep release Tau

== Likelihood ratio statistic for equality of recovery probabilities: 1.9254

Compare with quantiles of the chi-squared distribution with 1 d.f.:

One hour survival rates for juvenile chinook salmon released through Spillbay 5 at deep and shallow release sites at bulk spill patterns at Ice Harbor Dam, July 2003. Controls: 125 released, 124 alive, 1 dead. Deep release test: 60 released, 54 alive, 5 dead. Shallow release test: 119 released, 104 alive, 14 dead. (Values from Table 3-2).

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -77.9652

Tau = $0.8885(0.0308)$ Bulk spill, shallow release/Control ratio Tau = $0.9226(0.0373)$ Bulk spill, deep release/Control ratio

Z statistic for the equality of equal release sites survivals: 0.7061

Compare with quantiles of the normal distribution:

 1-tailed 2-tailed For significance level 0.10: 1.2816 1.6449 For significance level 0.05: 1.6449 1.9600 For significance level 0.01: 2.3263 2.5758

Variance-Covariance matrix for estimated probabilities:

0.00006349 0.00000000 0.00000000 0.00000000 0.00000000 0.00002150 0.00000000 0.00000000 0.00000000 0.00000000 0.00088617 0.00000000 0.00000000 0.00000000 0.00000000 0.00131464

Confidence intervals:

Likelihood ratio statistic for equality of recovery probabilities: 1.7112

Compare with quantiles of the chi-squared distribution with 1 d.f.:

One hour survival estimates of juvenile chinook salmon smolts passed through Spillbay 5 at shallow and deep release sites and bulk spill at Ice Harbor Dam, July, 2003. Controls: 125 released, 124 alive, 1 dead. Test fish: 179 released, 159 alive, 19 dead.

 estim. std. err. $S = 0.9920 (0.0080)$ Control group survival $Pa =$ 1.0 N/A Live recovery probability* $Pd = 0.9524 (0.0465)$ Dead recovery probability Tau = $0.8954(0.0248)$ Spillbay 5 survival 1-Tau = 0.1046 (0.0248) Spillbay 5 mortality

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -72.516274

Variance-Covariance matrix for estimated probabilities: 0.00006 0.00000 -0.00006 0.00000 0.00216 0.00000 -0.00006 0.00000 0.00062

Profile likelihood intervals: Spillbay 5 survival Spillbay 5 mortality 90 percent: (0.8514, 0.9342) (0.0658, 0.1486) 95 percent: (0.8422, 0.9416) (0.0584, 0.1578) 99 percent: (0.8236, 0.9570) (0.0430, 0.1764)

RESULTS FOR REDUCED MODEL (EQUAL LIVE/DEAD RECOVERY)

==

 estim. std. err. $S = 0.9920(0.0080)$ Control group survival $Pa = Pd \quad 0.9967 \quad (0.0033)$ Recovery probability Tau = $0.9005(0.0244)$ Spillbay 5 survival 1-Tau = $0.0995(0.0244)$ Spillbay 5 mortality

log-likelihood : -72.997047

Variance-Covariance matrix for estimated probabilities: 0.00006 0.00000 -0.00006 0.00000 0.00001 0.00000 -0.00006 0.00000 0.00060

Profile likelihood intervals: Spillbay 5 survival Spillbay 5 mortality 90 percent: (0.8570, 0.9387) (0.0613, 0.1430) 95 percent: (0.8479, 0.9460) (0.0540, 0.1521) 99 percent: (0.8294, 0.9614) (0.0386, 0.1706) ==

Likelihood ratio statistic for equality of recovery probabilities: 0.961546

Compare with quantiles of the chi-squared distribution with 1 d.f.: For significance level 0.10: 2.706 For significance level 0.05: 3.841

For significance level 0.01: 6.635

Clean fish rates for juvenile chinook salmon released through Spillbay 5 at shallow release sites at special and 100% spill patterns at Ice Harbor Dam, April and May 2003. Controls: 288 examined, 0 maladies. Special spill test: 77 examined, 1 fish with maladies. 100% spill test: 60 examined, 5 with maladies. (Values from Table 3- 4.)

== Likelihood ratio statistic for equality of recovery probabilities: -0.0001

Compare with quantiles of the chi-squared distribution with 1 d.f.:

Clean fish rates for juvenile chinook salmon released through Spillbay 5 at deep release sites at 50% and 100% spill patterns at Ice Harbor Dam, April and May 2003. Controls: 288 examined, 0 maladies. 50% spill test: 304 examined, 63 fish with maladies. 100% spill test: 147 examined, 18 with maladies. (Values from Table 3-4.)

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -209.7740

Tau = $0.7928(0.0232)$ 50% spill/Control ratio $Tau = 0.8776 (0.0270) 100\%$ spill/Control ratio

Z statistic for the equality between spill patterns without maladies: 2.3779

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00054043 0.00000000 0.00000000 0.00000000 0.00000000 0.00073099

Likelihood ratio statistic for equality of recovery probabilities: -0.0001

Compare with quantiles of the chi-squared distribution with 1 d.f.:

Clean fish rates for juvenile chinook salmon released through Spillbay 5 at deep release sites at special and 100% spill patterns at Ice Harbor Dam, April and May 2003. Controls: 288 examined, 0 maladies. Special spill test: 60 examined, 13 fish with maladies. 100% spill test: 147 examined, 18 with maladies. (Values from Table 3-4.)

==

Likelihood ratio statistic for equality of recovery probabilities: -0.0002

Compare with quantiles of the chi-squared distribution with 1 d.f.:

Clean fish rates for juvenile chinook salmon released through Spillbay 5 at deep and shallow release sites at 100% spill patterns at Ice Harbor Dam, April and May 2003. Controls: 288 examined, 0 maladies. Shallow release test: 77 examined, 1 fish with maladies. deep release test: 147 examined, 18 with maladies. (Values

Likelihood ratio statistic for equality of recovery probabilities: -0.0002

Compare with quantiles of the chi-squared distribution with 1 d.f.:

Clean fish rates for juvenile chinook salmon released through Spillbay 5 at deep and shallow release sites at special spill patterns at Ice Harbor Dam, April and May 2003. Controls: 288 examined, 0 maladies. Shallow release test: 60 examined, 5 fish with maladies. deep release test: 60 examined, 13 with maladies. (Values from $T = 1.1 - 2.4$

Likelihood ratio statistic for equality of recovery probabilities: -0.0001

Compare with quantiles of the chi-squared distribution with 1 d.f.:

Clean fish rates for juvenile chinook salmon released through Spillbay 5 at deep release sites at bulk and dispersed spill patterns at Ice Harbor Dam, July 2003. Controls: 124 examined, 2 maladies. Dispersed spill test: 97 examined, 21 fish with maladies. Bulk spill test: 55 examined, 11 with maladies. (Values from Table 3-5.)

```
RESULTS FOR REDUCED MODEL (EQUAL LIVE/DEAD RECOVERY)
```
 estim. std. err. $S1 = 0.9839 (0.0113)$ {Control group without maladies} $Pa = Pd$ 1.0 N/A Recovery probability* $S2 = 0.8000 (0.0539)$ bulk spill without maladies $S3 = 0.7835(0.0418)$ dispersed spill without maladies

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -88.4364

Tau = $0.8131(0.0556)$ bulk spill/Control ratio Tau = $0.7963(0.0435)$ dispersed spill/Control ratio

Z statistic for the equality between release depths without maladies: $\qquad 0.2375$

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00012797 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00290909 0.00000000 0.00000000 0.00000000 0.00000000 0.00174871

Confidence intervals: Bulk spill Tau Dispersed spill Tau 90 percent: (0.7216, 0.9046) (0.7248, 0.8679) 95 percent: (0.7041, 0.9221) (0.7111, 0.8816) 99 percent: (0.6699, 0.9563) (0.6844, 0.9083) ==

Likelihood ratio statistic for equality of recovery probabilities: 0.0000

Compare with quantiles of the chi-squared distribution with 1 d.f.:

Clean fish rates for juvenile chinook salmon released through Spillbay 5 at deep and shallow release sites at bulk spill patterns at Ice Harbor Dam, July 2003. Controls: 124 examined, 2 maladies. shallow release test: 108 examined, 7 fish with maladies. deep release test: 55 examined, 11 with maladies. (Values from Table 3-5.)

```
RESULTS FOR REDUCED MODEL (EQUAL LIVE/DEAD RECOVERY)
```
 estim. std. err. $S1 = 0.9839 (0.0113)$ Control group without maladies $Pa = Pd$ 1.0 N/A Recovery probability* $S2 = 0.9352 (0.0237)$ shallow release without maladies $S3 = 0.8000 (0.0539)$ deep release without maladies

* -- Because of constraints in the data set, this probability is assumed equal to 1.0; not estimated.

log-likelihood : -63.6818

Tau = $0.9505(0.0264)$ shallow release/Control ratio Tau = $0.8131(0.0556)$ deep release/Control ratio

Z statistic for the equality between release depths without maladies: 2.2313

Compare with quantiles of the normal distribution:

Variance-Covariance matrix for estimated probabilities: 0.00012797 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00000000 0.00056124 0.00000000 0.00000000 0.00000000 0.00000000 0.00290909

Confidence intervals:

== Likelihood ratio statistic for equality of recovery probabilities: -0.0001

Compare with quantiles of the chi-squared distribution with 1 d.f.:

APPENDIX D

DAILY FISH DISPOSITION DATA

Short-term turbine passage survival data on individual chinook salmon released at Spillbay 5 at the Ice Harbor Dam, April-May 2003. Fish were tagged with Normandeau's HI-Z Turb-N tags. Description of condition codes and details on injured fish are presented in Table 2-7.

