

The RE-Cycle System for Hog Waste Management

J. B. Koger, R.P. Burnette, T. van Kempen,

The authors are **Jeanne B. Koger**, and R. Preston Burnette, Department of Animal Science, North Carolina State University, Raleigh, North Carolina, 27695; T.A. van Kempen, formally Professor of Animal Science, North Carolina State University; currently, Director of Swine Research, Provimi R & T Centre, 2 Lenneke Marelaan, Sint Stevens Woluwe, B-1932, Belgium.

Corresponding author: Jeanne B. Koger, Department of Animal Science, Box 7626, North Carolina State University, Raleigh, NC 27695-7626; (919) 515-3319; FAX: (919) 515-7780; jeanne_koger@ncsu.edu.

ABSTRACT. *The RE-Cycle system is a combination of technologies for improving the environmental sustainability of hog production. The principal components are belt-based manure harvesting and gasification of the solids for recovery of energy and ash. The belt limits emissions of NH_3 and CH_4 to 1 kg/pig/y. Manure is collected in separate liquid and solid streams. Solids are harvested at 50% DM. The energy content of the solids (19.7 MJ/kg DM) is similar to wood making them suitable for thermochemical processing. A fixed-bed, batch-fed gasifier was chosen for its ease of operation, processing temperature ($>800^\circ C$), and ability to accept a wide array of feedstocks. Its only products are a product gas, a sterile mineral ash, and waste heat. Combustion of the product gas eliminates tars and provides heat to sustain the process and generate power. NO_x and SO_2 emissions are low. Swine waste solids, chicken litter, and chicken, turkey, and pig mortalities have each been processed in this unit. Ash mineral content is dependent upon the composition of the gasifier feedstock, but is generally high in P, K, and Ca. Ash minerals are bioavailable and can be used as a mineral supplement in animal feeds. Feeding trials have demonstrated the bioavailability of its mineral constituents. Mass balance data indicate good nutrient accountability.*

Keywords. *swine housing, animal manure, gasification, waste nutrients, waste management, sustainable agriculture.*

INTRODUCTION

The United States currently imports over 60% of its petroleum supply (EIA, 2004). The push toward energy independence and environmental sustainability has focused attention on biomass since it is renewable, widely available, and, unlike fossil fuels, makes no net contribution to greenhouse gas emissions (Gielen et al, 1998). It has been estimated that biomass could supply 56% of US gasoline needs (Riley, 2001), essentially replacing the imported oil. Animal wastes are a valuable source of

30 bioenergy and offer advantages such as continuous availability without long-term storage, energy
31 content comparable to other traditional sources such as wood, and low intrinsic value due to present
32 disposal costs.

33 Swine waste solids are 1 to 10 % dry matter (DM) under present harvesting methods (Smith et
34 al., 2000). Attempts to reduce the moisture content have focused on post-collection technologies such
35 as centrifuges, settling basins, and screen separators that generally have low solids recovery, leave a
36 problematic residual aqueous stream, and do not address the barn emissions that are responsible for
37 over 50% of the NH_3 emitted from hog farms (Doorn et al., 2002). Moreover these technologies suffer
38 from high investment costs for equipment, maintenance, and skills on the farm level (Westerman and
39 Bicudo, 2000). In order to capture energy from the solids, they must be sufficiently dry (50 to 70%
40 DM) to permit cost effective transport and processing.

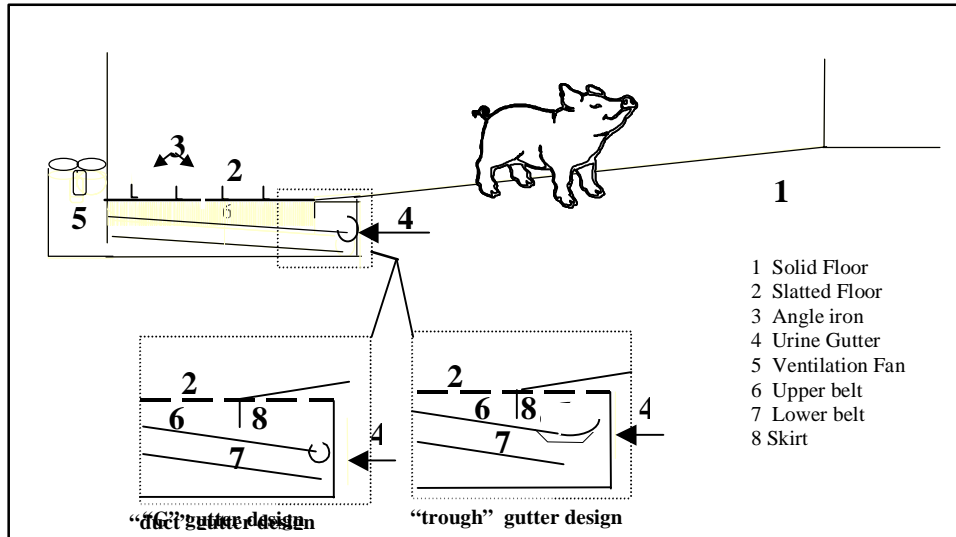
41 One promising method of capturing the energy of biomass is through gasification. This
42 oxygen-deprived, thermal decomposition technology converts biomass to low molecular weight
43 combustible gases and retains mineral nutrients in a sterile ash byproduct rich in P and Ca. In the
44 reducing atmosphere of the gasifier, N is converted to NH_3 that can be easily scrubbed from the
45 product gas stream. Given the low sulfur content of biomass, SO_x emissions are far less than in coal-
46 fired processes (Graham et al., 1996). The resulting product gas can be combusted to sustain the
47 gasification process and waste heat can be captured to heat buildings, generate steam, or provide power
48 through a Sterling engine. The non-odorous ash by-product can be cost-effectively transported for use
49 as a mineral supplement in animal feeds, as a fertilizer ingredient, or as a building materials
50 component.

51 The objective of the research presented here was to demonstrate a management strategy that
52 would avoid the negative environmental consequences of hog production while capturing the energy
53 and nutrients of waste in value-added products. The belt-based housing system is designed for
54 separate collection of urine and feces and for improving the air quality in and emissions from hog
55 buildings. The harvested solids have been processed through gasification to dispose of the waste and

56 recover valuable nutrients. Ash digestibility has been evaluated in vitro and in vivo. This approach
57 also allows for energy capture in the form of heat and steam.

58 **EXPERIMENTAL PROCEDURES**

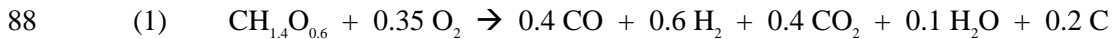
59 The belt-based housing was modeled after a conventional grow-finish facility at 75% of full scale
60 (Figure 1). Pens were two-thirds solid floor, one-third slatted floor (tri-bar, Nooyen, Mt. Sterling,
61 KY). A high-density polyethylene (1 mm thick) belt conveyor system (Big Dutchman, Vechta,
62 Germany) was installed beneath the tri-bar. There were 5 pens with 16 pigs per pen (25-55 kg) at a
63 stocking density of $0.63 \text{ m}^2 \text{ pig}^{-1}$. This is similar to the commercial stocking density of 0.4 to 0.7 m^2
64 pig^{-1} for pigs of 18 to 72 kg body weight (Meyer et al., 1991). The solid flooring had an 8% slope
65 allowing liquids to run off into the belt gutter. The 1° lengthwise slope of the belt, designed to mimic
66 the slope of a commercial building, insured that urine flowed continuously into an enclosed liquid
67 collection vessel at the end of the belt. Two different gutter designs, the “duct” design and the
68 “trough” design, were evaluated for odor and emissions reduction and for ease of cleaning. Animals
69 were fed a standard commercial diet and monitored for weight gain, feed and water intake, and waste
70 production. Ammonia and CH_4 concentrations in the exhaust air from the belt housing facility were
71 determined by Fourier Transform Infrared Spectroscopy (FTIR) as described in van Kempen (2001).
72 Emissions were calculated, after subtraction of background levels in the input air, as the detected
73 differential concentration times the ventilation rate at the time of analysis. The detected differential
74 concentration was multiplied by the ventilation rate at the time of analysis to determine emission levels
75 over time. Materials harvested from the belt were characterized by proximate and ultimate analysis
76 (Hazen Research Laboratories), and mineral composition was determined in house by inductively
77 coupled plasma spectroscopy (ICP). Energy content was established by bomb calorimetry and dry
78 matter determinations were calculated from weights before and after drying samples in a 60°C oven
79 till weight change was less than 1% in a 24 h period.



80 **Figure 1. Schematic diagram of the pen and belt layout as tested. The belt is positioned beneath the slats so that**
 81 **animals defecate most on the highest portion of the belt. Urine is allowed to run off into a gutter that continuously**
 82 **carries it out of the room, by gravity flow, into a closed container. The “duct” and “trough” gutter designs are shown**
 83 **in the expanded views of the boxed pen area.**

84 Gasification employs heat and sub-stoichiometric amounts of oxygen to break down, or crack, the
 85 feedstock into carbon monoxide and hydrogen gases, as well as other low molecular weight gases. The
 86 idealized, general equation (Reed and Gaur, 1999) is:

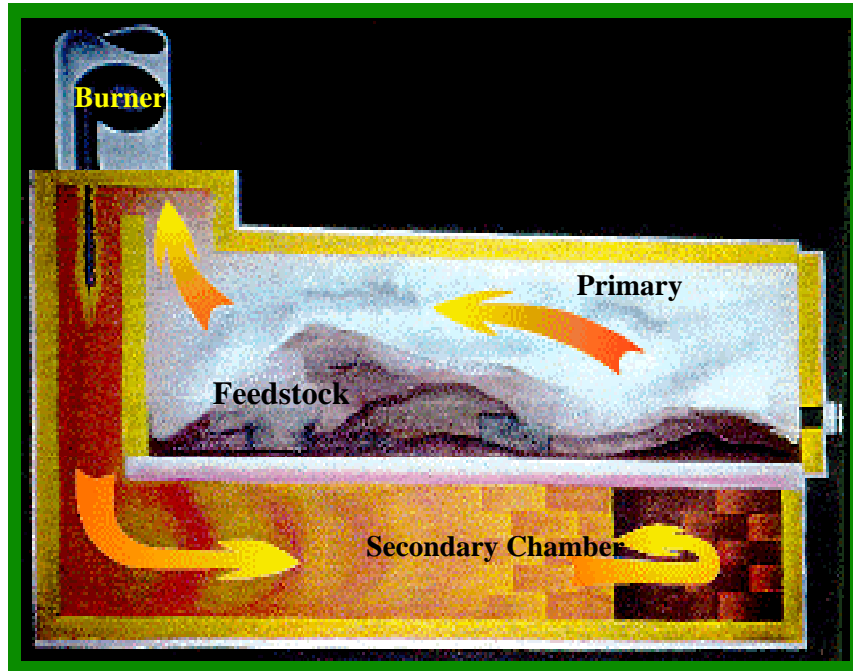
87



89

90 For biomass, temperatures are generally held below 1000 °C to avoid agglomeration within the ash
 91 bed. Gasification studies were conducted in a Brookes Gasification Process (BGP) gasifier. This is a
 92 batch-fed, indirectly heated, staged-combustion gasifier selected for its simplicity and ease of operation
 93 (Figure 2). At start-up, the burner in the secondary chamber combusts propane to heat the firebrick
 94 and, indirectly, the primary chamber containing the feedstock. At temperatures of 600 to 900 °C,
 95 biomass feedstock is cracked into low molecular weight gases that are then drawn into the secondary
 96 chamber where they are combusted to sustain the process. Ash is removed after completion of the
 97 feedstock processing. Combustion gases exit the flue along with waste heat. Gases in the two

98 chambers and the flue are monitored at 15 minute intervals to evaluate temperature, NO_x, SO₂, CO,
99 CO₂, and O₂ levels using a combustion meter (Kane May Instruments). Temperature and oxygen
100 availability can be controlled during the reaction. This control and the static nature of the system result
101 in a clean process that meets both European and Californian emission standards without requiring gas
102 cleanup.



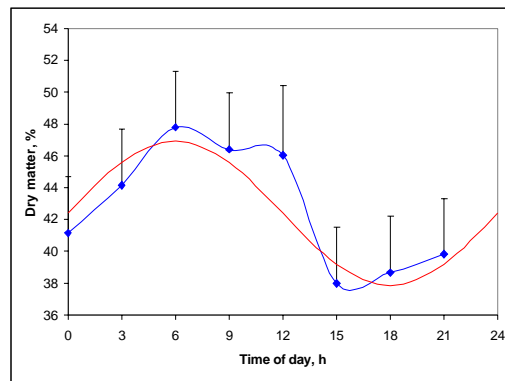
103 **Figure 2. Schematic of the Brookes Gasification Process gasifier. Heat from the burner in the secondary**
104 **chamber heats the primary chamber containing the feedstock. Gases from the feedstock are drawn into the**
105 **combustion chamber to sustain the reaction.**

106 **RESULTS AND DISCUSSION**

107 **BELT OPERATION, PIG PERFORMANCE, AND EXCRETA PROPERTIES**

108 Belt operating variables were evaluated in the first two experiments in order to optimize
109 performance. Prior to data collection, it was hypothesized that the DM content would increase
110 with belt residence time; this was evaluated in Exp. 1. DM content of the recovered solids
111 was found to decrease after 30 h of belt residence time (data not shown). Apparently, when

112 solids remained on the belt for greater than 30 h, the increased fecal load trapped more of the
 113 urine thereby reducing the DM content. Collection time-of-day was examined in Exp. 2 by harvesting
 114 waste solids at a 27-hour interval with 3 replicates of each time-of-day. A sigmoidal DM pattern was
 115 observed (Figure 3). Collections at 06:00 resulted in a $9.8 \pm 5.0\%$ DM increase over collections at
 116 15:00. The diurnal activity pattern of the pigs probably produced this result since animals eliminate
 117 less at night when they are resting and more when they first eat and become active in the morning. In
 118 all subsequent experiments, the belt contents were harvested every 24 hours at 06:00 resulting in an
 119 average DM content of $49 \pm 4\%$ (TABLE 1). At this DM content, swine feces typically have a brittle,
 120 dry appearance externally with a moist core present in larger pieces. It does not clump or pack during
 121 handling and it has minimal odor.



122 **Figure 3. Effect of collection time-of-day on feces DM content. Belt residence time was constant at 27**
 123 **h and each time-of-day was evaluated in triplicate.**

124 **TABLE 1. Experimental parameters and animal performance for five trials designed to evaluate a housing system**
 125 **employing a belt for harvesting urine and feces separately. Data are expressed on a per animal basis, except as**
 126 **indicated.**

Trial date and Number	Aug 01 1	Jan 02 2	Mar 02 3	Oct 02 4	Apr 03 5	Mar 02 3 Ref.	Mean ± SE
N, number of animals	100	80	80	80	80	80	
Avg. T, °C	28	30	26	22	25		26 ± 3
Avg. humidity, %	66	54	52	64	64		60 ± 5
Gutter design	duct	duct	duct	trough	trough	none	
Ave. Wts, in - out, kg	23-57	27-55	23-51	30-57	32-56	24-53	
ADG, kg-d-1	0.79	0.76	0.82	0.97	0.83	0.83	0.83 ± 0.07
ADFI, kg-d-1	1.54	1.52	1.72	1.96	1.73	1.84	1.69 ± 0.16
Gain/Feed	0.51	0.50	0.48a	0.48	0.51	0.45b	0.50 ± 0.02
Fecal output, kg DM-d-1	0.23	0.26	0.22	0.34	0.26	ND	0.26 ± 0.04
Fecal DM as coll'd, %	§	52	54	43	46	ND	49 ± 4
Urine output, L-d-1	ND	1.05	1.54	1.27	1.42	ND	1.32 ± 0.19
NH3 emission, kg-y-1,c	1.08	0.80	0.84	1.20	1.24	ND	1.03 ± 0.18
CH4 emission, kg-y-1	1.27	1.06	0.75	1.39	0.76	ND	1.05 ± 0.26

127 Collected feces indicate a fecal DM production of $0.26 \pm 0.05 \text{ kg}\cdot\text{pig}^{-1}\cdot\text{day}^{-1}$ (TABLE 1), similar
128 to the 0.27 kg reported by Smith et al. (2000) for the grower pig. Absolute levels of NH_3 emissions
129 averaged $1.0 \pm 0.2 \text{ kg}\cdot\text{pig}^{-1}\cdot\text{y}^{-1}$ over the five experiments (TABLE 1). This suggests a 73%
130 reduction from the literature value of $3.7 \text{ kg}\cdot\text{pig}^{-1}\cdot\text{y}^{-1}$ for conventional barns (Doorn et al., 2002).
131 The low levels of NH_3 are attributed to the minimal contact time between urea and the fecal microbes
132 that metabolize it to NH_3 and CO_2 (Rom, 1995) and to the rapid sequestering of the urine in closed
133 containers. Emissions of CH_4 , a potent greenhouse gas, were only $1.05 \pm 0.26 \text{ kg}\cdot\text{pig}^{-1}\cdot\text{y}^{-1}$.

134 The gutter design was altered between Exp. 3 and 4 and this may have impacted NH_3
135 emissions and solids DM. NH_3 emissions and DM content were $1.22 \text{ kg}\cdot\text{pig}^{-1}\cdot\text{y}^{-1}$ and 45% DM
136 with the trough design, but $0.91 \text{ kg}\cdot\text{pig}^{-1}\cdot\text{y}^{-1}$ and 53% DM with the duct design. This could be
137 attributable to the gutter design or to the higher average temperatures and lower average humidity in
138 Exp. 2 and 3 (duct gutter) relative to Exp. 4 and 5 (trough gutter). The trough design does offer some
139 unique advantages, namely: it occupies less space and so facilitates retrofits, and it is cleaned each time
140 the belt is operated. Addition of a cover to the trough design could limit airflow across the urine and
141 might help to further reduce NH_3 emissions from the trough gutter, if indeed the gutter is responsible
142 for the changes noted.

143 Animal performance was evaluated in each of the trials (TABLE 1). The average daily feed
144 intake (ADFI) was $1.69 \pm 0.18 \text{ kg}\cdot\text{pig}^{-1}\cdot\text{d}^{-1}$ and average daily gain (ADG) was $0.83 \pm 0.08 \text{ kg}\cdot\text{pig}^{-1}\cdot\text{d}^{-1}$.
145 The gain to feed ratios (G/F) averaged 0.50 ± 0.02 across the five belt-based housing trials. In Exp. 3,

146 where performance in the belt-based housing was compared to that in conventional flush system barns,
147 there was a significant improvement in the G/F ratio of 5.5% ($P = 0.01$) for the belt-based animals
148 despite the fact that housing density, diet, and animals were matched. Experiments 1, 2, 4 and 5
149 combined yielded an 11% improvement in feed efficiency when compared to the Exp. 3-Reference or
150 to other trials carried out in the conventional housing system. The reference farm had a lower group
151 size (7 animals·pen⁻¹ compared to 16 animals·pen⁻¹ in the belt-based facility) necessitated by the desire
152 to match housing density despite different pen sizes. Group size has been shown not to impact the G/F
153 ratio, although larger groups may reduce both FI and growth rate (Hyun and Ellis, 2001; Wolter et al.,
154 2001). Housing environment, however, has been shown to impact pig health and productivity
155 (Donham, 1991). The source of the G/F improvement requires further investigation, but may be due to
156 improvements in air quality resulting from the separate collection of urine and feces. If these
157 improvements are indeed attributable to the housing design, then the belt-based system could result in
158 substantial feed cost savings to producers.

159 The nutrient loads for various inputs and outputs of the RE-Cycle system were calculated based
160 on consumption and production data for feed and wastes as well as nutrient concentration data for all
161 streams under investigation (TABLE 2). Data for accretion calculations were from Mahan and Shields
162 (1998) for the weight pigs used in our studies. The feed nutrients recovered in the various waste
163 streams plus those calculated as accreted by the animal, show good closure with the nutrients present in
164 the feed consumed. For N, P, K, and Ca, the percent closure is close to or exceeds 90%. For Mg, the
165 lack of accretion data limits conclusions, but the fact that there is 75% accountability in the absence of
166 accretion data is very promising. For Cu and Zn, however, the situation is just the reverse; it seems
167 that more of these nutrients have been recovered than were provided to the animals. Perhaps some
168 copper could have been added through contact of the drinking water with copper pipes. Both of these
169 elements, though, are in very low abundance making it more difficult to segregate the mineral
170 analytical response from the background noise. It is thought that the results shown for Cu and Zn in
171 feces are an overestimation of the amount present and this is probably due to analytical error.

172 **TABLE 2. Nutrient closure from the belt-based housing system.**

Component	Intake/Output g DM/pig/d	Energy kJ/pig/d	Mineral Ash %	Nutrients per pig per day								
				units	N	P	K	Ca	Mg	Cu	Zn	
IN												
Feed	1690	29925		g/pig/d	49.1	10.8	13.4	12	2.7	0.038	0.259	
OUT												
Feces	260	5319	12 – 15	g/pig/d	10.5	5.1	5.1	4.9	2	0.037	0.540	
Urine	1320			g/pig/d	9.2	0.2	6.1	0.1	0.0	0.01	0.02	
NH ₃ Emissions				g/pig/d	2.4							
Accretion				g/pig/d	21.04	3.99	0.632	6.79	UNK	0.001	0.01	
OUT / IN				%	87.9	86.0	88.3	98.3	74.7	126.4	220.3	

173 **GASIFICATION OF SWINE MANURE SOLIDS**

174 Prior to gasification, swine waste solids were analyzed for DM, energy, mineral content, and ash
 175 fusion temperature in an effort to understand the probable operating conditions required by this
 176 feedstock. At 19.7 MJ/kg DM, swine waste compares favorably with wood (20.2 MJ/kg DM) from an
 177 energy standpoint. Waste solids were 45% C, 0.4 % S, and 0.3% Cl⁻¹. Fixed C was 13%; volatile
 178 matter, 75%; and ash, 12.2%. The lowest temperature for ash fluidization, seen under oxidizing
 179 conditions, was 1218 °C so processing temperatures less than 1000 °C were not expected to cause ash
 180 agglomeration problems.

181 The static system of the BGP gasifier makes it ideal for a wide variety of feedstocks since
 182 virtually anything can be loaded into an open cavity for a batch feed reaction. In addition to the swine
 183 waste feedstock, the gasifier has thus far processed nursery pig mortalities, poultry litter, chicken and
 184 turkey mortalities. Ashes from swine waste solids and poultry litter were compared to the ashes
 185 obtained by incinerating pig mortalities (Prestage Farms, grower/finisher, G/F) and by gasifying turkey
 186 litter (Energy Products of Idaho), TABLE 3. The data were corrected to an ash only basis for purposes
 187 of comparison. High levels of Ca are found in mortality ash in keeping with the bone content of this
 188 feedstock. The elevated Cu levels in turkey litter ash may result from dietary levels of this mineral. In
 189 short, differences in the ash mineral compositions appear to result from differences in the feedstock
 190 rather than from differences in the processing technologies.

191

192 TABLE 3. **Comparison of ashes from different feedstocks and from various thermochemical processing technologies.**

Alternative Feedstocks		Corrected to 100% ash—expressed on “ash only” basis					
		%				ppm	
Type	Source	P	K	Ca	Mg	Zn	Cu
Hog waste	Ash, gasifier NCSU-BGP	14.34	11.34	16.05	4.97	9044.5	619.4
Chicken litter	Ash, gasifier NCSU-BGP	4.7	4.0	23.3	1.1	0.1	0.02
swine mortalities	Ash, incinerator Prestage G/F	17.45	5.49	27.40	1.22	945.2	480.1
Turkey litter	Ash, gasifier EPI	6.15	3.38	9.76	1.98	586.5	1443.1

193
 194 Experiments with a pilot-scale gasifier (~70 kg per batch), on loan from BGP, have confirmed its
 195 ease of operation. One load can be processed in 3 to 4 h providing a 20 kg/h through-put with this
 196 demonstration unit. In order to determine the optimal operating temperature, two trials were performed
 197 each day at each temperature ranging from 700°C to 900°C (Table 4). The second trial of each day is
 198 reported here in order to compare “steady-state” conditions. From these trials, 800°C was selected as
 199 the standard operating temperature since it balanced residence time with propane usage and provided
 200 sufficient temperature to guarantee sterile ash and minimal dioxin formation. Since the BGP gasifier is
 201 designed to consume its product gases, the only available energy output is heat from the flue stack. We
 202 do not currently have the ability to use this energy, but this heat could be used for steam generation,
 203 hot water, heating animal barns, or any other process that needs a high heat source.

204 **TABLE 4. Effect of operating temperature on feedstock processing time and propane use.**

Set Temperature	700	750	800	850	900
Feedstock: expired pelleted feed	70 kg at 73% DM	70 kg at 73% DM	70 kg at 73% DM	70 kg at 73% DM	70 kg at 73% DM
Total time, h	3.5	3.25	3.25	3.25	3.75
Total Propane used, cf	85.5	81.9	87.6	96.6	118.3
PC, max, °C	1003	1033	1035	1047	1028

205
 206 There are several factors that influence the efficiency of the the BGP unit. The feedstock type,
 207 DM, feedstock load, and whether the unit was in steady state operation. Data from processing pig
 208 mortalities indicate that propane use in a cold start experiment is 3 times that when the unit has been
 209 preheated (28.4 cf propane /kg 100% DM feedstock when cold and 9.5 cf propane when preheated).

210 Recovery of fecal nutrients in BGP gasifier ash has also been evaluated with good results.
 211 Gasifying concentrates these nutrients into a mass that is only 17% of the initial feedstock mass (from
 212 260 g manure solids to 44 g pig⁻¹•d⁻¹. The reduced mass has obvious advantages for transport of these
 213 nutrients to areas deficient in them. With the exception of Cu and Zn, mineral recoveries were
 214 generally in the 80 to 110% range. The Cu and Zn results again suggest that the fecal values obtained
 215 for these minerals are spurious. Artificially high fecal values would result in greater than 100%
 216 recovery when feces are compared to feed, and substantially lower than expected recovery when ash is
 217 compared to feces.

218

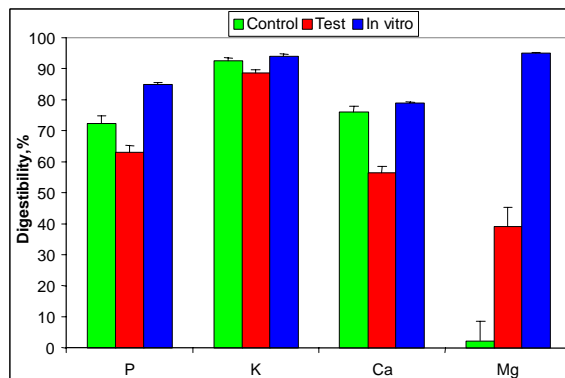
219 **TABLE 3. Nutrient balance from gasification of swine waste solids.**

Component	Intake/Output	Energy	Mineral Ash	Nutrients per pig per day							
				units	N	P	K	Ca	Mg	Cu	Zn
Feedstock IN	g DM/pig/d	kJ/pig/d	%								
Feces	260	5319	12 – 15	g/pig/d	10.5	5.1	5.1	4.9	2	0.03	0.54
PROCESSED											
Ash ^a	44	612	85	g/pig/d	ND	4.8	4.0	5.3	1.7	0.02	0.29
ASH / FEEDSTOCK				% Closure		94.5	78.3	108.1	84.2	65.6	54.3

220

221 Prior to generating sufficient gasifier ash in house, preliminary digestibility trials were conducted
 222 with gasified swine waste solids from the EPI process. In vitro solubilities serve to predict
 223 bioavailability and are determined by the extent to which ash minerals are soluble in an aqueous
 224 solution of HCl at pH 2 (Figure 5). P, K, Ca, and Mg were monitored and all these minerals were very
 225 soluble under the conditions described. In vivo, digestibility was determined with 7 pigs in a crossover
 226 design. Animals were fed 80g/kg^{0.75} per day of a highly purified (low mineral) cornstarch-casein diet
 227 supplemented with either ash or a mineral mix based on monocalcium phosphate, limestone, KOH,
 228 magnesium acetate, HCl, and salt. After 5 days of adaptation, feces were collected quantitatively for 3
 229 days before the crossover and repeat. Digestibility was determined as the difference between total
 230 intake and total fecal excretion for each nutrient studied. While palatability was a problem, the
 231 absorption data indicate that ash minerals are bioavailable and digestibility (Figure 5). Two problems
 232 occurred in this preliminary trial: phosphorus was inadvertently over-formulated in the test diet (due to

233 inaccurate analytical results) and an inappropriate, non-digestible source ($MgSO_4$) was used in
234 the control diet. A full scale trial is planned to better confirm digestibility data, pig performance, and
235 pig health when fed ash as a mineral supplement throughout the grow/finish period.



236 **Figure 5. Mineral availability from gasifier ash.**

237
238 NO_x levels were monitored in the secondary chamber and the flue stack to determine the extent of
239 these emissions to the environment. Flue stack gases are most relevant to environmental impact, but
240 vary with damper opening so that measurements from the secondary chamber were used to follow
241 reaction progress. NO_x levels in the flue ranged from 2 to 20 ppm. NO_x is elevated initially, but
242 decreases after 3 hours of processing as feedstock is reduced to carbon char and mineral ash.
243 Preliminary tests have been conducted to determine the effect of injecting liquid waste into the BGP
244 gasifier secondary chamber to reduce NO_x levels in flue gas as predicted by Pulkrabek (1997). Data
245 obtained were very encouraging since NO_x levels were reduced from over 340 ppm to less than 80
246 ppm. However, no measure of flue ammonia levels was available during this process, and further
247 testing is required before the exact impact of the procedure can be described. For this reason, the fate
248 of nitrogen has not yet been well characterized throughout the system.

249 **CONCLUSION AND FUTURE DIRECTIONS**

250 Belt-based housing holds promise for reducing emissions from hog buildings, improving pig
251 performance, and facilitating waste nutrient transport to areas of need. Improved feed efficiency may
252 help to improve producers' profitability. The solid material recovered from the belt is ~50% DM and

253 suitable for gasification. A commercial-scale belt installation is a vital component for verifying the
254 benefits observed with the demonstration unit and for supplying sufficient waste to operate the gasifier
255 over extended periods. The BGP gasifier is easy to operate, completely disposes of bioactive
256 molecules and pathogens, and offers the potential of recovering energy from waste heat. The residual
257 ash retains valuable nutrients, such as P, K, and Mg, that can replace mined mineral supplements in
258 feed or be part of a fertilizer blend. More work must be done to optimize the gasification process and
259 reduce secondary fuel use. A continuous feed gasifier is currently under development; it will be
260 operated under steady state conditions which is expected to dramatically enhance efficiency. Addition
261 of steam reforming and product gas capture to the new design will make it possible to modify the
262 product gas composition and recover it for synthesis of biofuels.

263 **ACKNOWLEDGEMENTS**

264 The authors wish to thank the North Carolina Department of Administration, State Energy Office,
265 The Animal and Poultry Waste Management Center at NCSU for providing initial funding for the
266 reported work.

267 **REFERENCES**

- 268 Donham, K.J. 1991. Association of environmental air contaminants with disease and productivity in
269 swine. *Am. J. Vet. Res.* 52:1723-1730.
- 270 Doorn, M.R.J., D.F. Natschke, S.A. Thorneloe, and J. Southerland. 2002. Development of an
271 emission factor for ammonia emissions from US swine farms based on field tests and
272 application of a mass balance method. *Atmospheric Environ.* 36:5619-5625.
- 273 Energy Information Administration (EIA). 2004. Country Analysis Briefs: United States of America.
274 Available at <http://www.eia.doe.gov/emeu/cabs/usa.html>. Accessed November 12, 2004.

275 Gielen, D.J., A.J.M. Bos, and T. Gerlagh. 1998. Biomass for Greenhouse Gas Emission Reduction
276 (BRED). Paper prepared for the Conaccount meeting: Ecologizing Societal Metabolism;
277 Designing Scenarios for Sustainable Materials Management, 21 November, Amsterdam.

278 Graham, R.L., M. Downing, and M.E.Walsh. 1996. A Framework to Assess Regional Environmental
279 Impacts of Dedicated Energy Crop Production. *Environmental Management* 20, 475-485.

280 Hyun, Y. and M. Ellis. 2001. Effect of group size and feeder type on growth performance and feeding
281 patterns in growing pigs. *J. Anim. Sci.* 79:803-810.

282 Koger, J., P. Burnette, A. Wossink, B. Kaspers, D. Ali, J. Spivey, and T. van Kempen. 2004. RE-
283 Cycle: the Production of Liquid Fuels from Swine Waste. Final report submitted to the NC
284 Department of Administration, State Energy Office. August 27.

285 Mahan, D.C. and R.G. Shields, Jr. 1998. Macro- and micromineral composition of pigs from birth to
286 145 kilograms of body weight. *J. Anim. Sci.* 76:506-512.

287 Meyer, V. M., L. B. Driggers, K. Ernest, and D. Ernest. 1991. Swine Growing-Finishing Units. in
288 *Pork Industry Handbook*, PIH-11. North Carolina Cooperative Extension Service, Raleigh,
289 NC.

290 Pulkrabek, W.W. 1997. *In* *Engineering Fundamentals of the Internal Combustion Engine*. Prentice
291 Hall, NJ. pp 303-304.

292 Reed, T.B., and S. Gaur. 1999. A Survey of Biomass Gasification 2000, Gasifier Projects and
293 Manufacturers Around the World. National Renewable Energy Laboratory and The Biomass
294 Energy Foundation, Inc., Golden, CO.

295 Riley, C. (US DOE/NREL, Golden, CO), "Bioethanol: A Renewable Transportation Fuel from
296 Biomass", presented at AIChE Spring Meeting, March, 21, 2001; available at:
297 http://www.ott.doe.gov/biofuels/recent_posters_abstracts.html#bioethanol .

298 Rom, H.B. 1995. Ammonia emission from pig confinement buildings. System analysis and
299 measuring methods. Ph.D. Thesis. Danish Institute of Animal Science, Bygholm.

300 Smith, K.A., D.R. Charles, and D. Moorhouse. 2000. Nitrogen excretion by farm livestock with
301 respect to land spreading requirements and controlling nitrogen losses to ground and surface
302 waters. Part 2: pigs and poultry. *Biores. Tech.* 71:183-194.

303 van Kempen, T., B. Kaspers, P. Burnette, M. van Kempen, and J. Koger. 2003. Swine Housing with a
304 Belt for Separating Urine and Feces; Key to Flexibility? Second International Swine Housing
305 Conference, Sponsored by ASAE, Durham, NC, 12-15 October.

306 van Kempen, T. 2001. Dietary adipic acid reduces ammonia emission from swine excreta. *J.*
307 *Anim. Sci.* 79:2412-2417.

308 Westerman, P. W., and J. R. Bicudo. 2000. Tangential flow separation and chemical enhancement to
309 recover swine manure solids, nutrients and metals. *Biores.Technol.* 73:1-11.

310 Wolter, B.F., M. Ellis, S.E. Curtis, N.R. Augspurgen, D.N. Hamilton, E.N. Parr, and D.M. Webel.
311 2001. Effect of group size on pig performance in a wean-to-finish production system. *J.*
312 *Anim. Sci.* 79:1067-1073.