## ANALYSIS OF SPMD SAMPLES FROM THE

# APRIL 2004 DEPLOYMENT IN BLUESTONE RIVER, VA <br> FOR PCBs AS BIOAVAILABLE ORGANIC CONTAMINANTS 

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## EXECUTIVE SUMMARY

This report summarizes findings from a study using lipid-containing semipermeable membrane devices (SPMDs) to monitor Polychlorinated Biphenyls (PCBs) as bioavailable organic contaminants in the waters of the Bluestone River watershed near Bluefield, Virginia. The work was conducted as part of a continuing collaborative effort between the U.S. Geological Survey’s (USGS) Columbia Environmental Research Center (CERC) and the Virginia Department of Environmental Quality (DEQ) to assess waterborne hydrophobic contaminants in the rivers and streams within in the Commonwealth of Virginia. CERC scientists fabricated the SPMDs, and DEQ personnel deployed them for 30 days (March 31, 2004 to April 30, 2004) at six sites in the aquatic environments of the Bluestone River watershed, Virginia.

## "Summary of Deployment Locations"

## DEQ CEDS ID DEQ Site Description

9-BST021.26 YARDS: Norfolk Southern rail yard @ state line; in mainstem Bluestone River, just downstream of the railroad bridge, on RHS looking upstream.

9-BFK000.02 BRUSH FORK: In Brush Creak tributary, 40 yards above confluence with Bluestone River, just upstream of Rt 717 bridge, on RHS looking upstream.

VA0025054-001 BLUEFIELD WESTSIDE STP: In last chlorine contact tank (in sewage treatment plant chlorine contact channel, just before exiting into main flow) discharges into mainstem of Bluestone River, on LHS looking upstream.

9-BPB000.02 BEAVERPOND CREEK: In Beaverpond Creek tributary, 3 yards above confluence with Bluestone River.

9-WVC000.05 WRIGHT'S VALLEY CREEK: In Wright's Valley Creek tributary, 10 yards above confluence with Bluestone River, on LHS looking upstream.

9-BST028.90 BELOW WTP: In mainstem of Bluestone River, behind First Church of God parsonage, at Fincastle Country Club water withdrawal location, on RHS looking upstream.

The SPMDs were subsequently shipped to the CERC where they were processed and analyzed for polychlorinated biphenyls (PCBs). Concentrations of total PCB contamination were found to be present above the gas chromatographic (GC) method quantitation limits (MQL) at all sites as given in "Summary Table I," Previously developed models were employed to estimate the water concentrations of PCBs within the aquatic environments of the Bluestone River watershed as given in " Summary Table I."

## "Summary Table I"

Results from the PCB Analysis of SPMDs Deployed During April of 2004 in the Bluefield River (Virginia) Watershed

| DEQ CEDS ID | Total PCBs Sequestered <br> by SPMDs (ng/SPMD) | SPMD derived water <br> concentration of Total <br> PCBs (pg / Liter) |
| :---: | :---: | :---: |
| 9-BST021.26 | 180 | 1300 |
| 9-BFK000.02 | 167 | 1200 |
| VA0025054-001 | 89 | 640 |
| 9-BPB000.02 | 505 | 3700 |
| 9-WVC000.05 | 170 | 1200 |
| 9-BST028.90 | 32 | 230 |

Determinations for total PCBs sequestered in deployed SPMDs were measured as the sums of the individual analytical responses from 140 individual PCB congeners as
detected during GC-ECD analysis. Not all targeted congeners were detected in individual study samples. And, not all detected congeners gave response signals which were quantifiable. Whereas all detected signals were summed to derive the total PCB levels reported, summations of quantifiable congeners do add up to the total PCB values reported. The analyses results from these six study sites, along with the analytical method detection limits (MDLs) and method quantitation limits (MQLs) for total PCBs and individual PCB congeners are reported herein.

## TABLE OF CONTENTS

EXECUTIVE SUMMARY ..... 2
TABLE OF CONTENTS ..... 5
INTRODUCTION ..... 6
EXPERIMENTAL .....  8
RESULTS AND DISCUSSIONS ..... 17
ACKNOWLEDGEMENTS ..... 24
LITERATURE CITED ..... 25
TABLES ..... 29
FIGURES ..... 56

## INTRODUCTION

Increasingly, human activities adversely affect the quality of our water resources. Anthropogenic contaminants are often released as a result of these activities, which may have deleterious effects on aquatic life and humans at ultra-trace levels. Thus, it is critical to determine the presence of these chemicals and trends in the concentrations to aid in designing effective mitigation strategies.

Unfortunately, there is a paucity of water concentration data for toxicologically significant waterborne contaminants such as PCBs. The levels of these very hydrophobic compounds in waters are usually below the detection limits of commonly employed low volume (i.e., $\leq 5 \mathrm{~L}$ ) sampling and analytical methods. Even when high volume solid phase extraction systems are used for the analysis of environmental waters, concerns exist for sample contamination, analyte losses, and procedurally mediated changes in the aqueous distribution of target compounds due to the collection, filtration, and extraction of large volumes of water. Also, episodic changes in environmental contaminant concentrations may not be detected, because sampling with conventional methods provide data only for a single point or small window in time.

To address most of these analytical methods issues and others related to organism exposure, scientists at the U.S. Geological Survey's Columbia Environmental Research Center (CERC) have developed a semipermeable membrane device (SPMD) for in situ passive integrative monitoring of aquatic contaminants (3-5). The SPMD consists of layflat, low-density polyethylene (LDPE) tubing containing a thin film of high-molecular
weight (~800 Da), high-purity triolein. The nonporous polymeric membrane used in the SPMD sampler functions by allowing the readily bioavailable contaminant molecules to pass through transient membrane cavities approaching $10 \AA$ in cross sectional diameter. The molecular size limitation of the LDPE used for SPMD membrane suggests that only dissolved chemicals will be sampled by SPMDs, which has been confirmed by Ellis et al. (6). Note that for compounds with $\log$ octanol-water partition coefficients $\left(\mathrm{K}_{\mathrm{ow}} \mathrm{s}\right) \leq 6.0$, the dissolved phase is the major source of residues accumulated in aquatic organism tissues (7). Also, residue transfer through the polymeric cavities in the LDPE membrane of SPMDs appears to mimic the transport of contaminants through biomembranes (8), which leads to bioconcentration of recalcitrant hydrophobic contaminants.

Lipid containing SPMDs are applicable to all neutral organic compounds with $K_{o w} s \geq 3.0$, which includes PCBs, and in general all persistent organic pollutants (POPs). In the case of very hydrophobic analytes with $\log \mathrm{K}_{\mathrm{ow}} \mathrm{S}>6.0$, where a large portion of the total waterborne chemical is associated with particulate or dissolved organic carbon phases, SPMDs readily concentrate ultra-trace residues of these compounds present in the dissolved phase to quantifiable levels (3, 9). Depending on environmental conditions such as velocity-turbulence at the membrane surface, temperature, and biofouling level, a "standard" 1-mL triolein SPMD (5) will passively extract hydrophobic contaminant residues from about 1 to 10 L of water per day. Sampling of hydrophobic compounds will generally remain integrative (i.e., no significant losses of residues accumulated in the device, even when ambient concentrations fall) throughout a 30 day (d) sampling period. Thus, more than 100 L of water is typically extracted of hydrophobic organic
contaminant residues by one-standard SPMD during a 30 d exposure period. Also, we have developed an approach (10) that permits determination of site-specific SPMD sampling rates. Details of this method are given in the experimental section. The purpose of this project was to assess the levels of total PCBs in the watershed of the Bluestone River, VA for the Virginia Department of Environmental Quality (DEQ), and to provide previously unattainable water column concentrations of PCBs to evaluate any impairment of water quality. A secondary objective of the work was to provide DEQ with an assessment of the distribution of individual sequestered PCB congeners within the SPMDs from each study deployment site. The congener specific information was intended as an aid to DEQ personnel in the identification of potential sources of environmental contamination and as a comparison to historic PCB congener specific analyses data from sediments of the Bluestone River.

## EXPERIMENTAL

Materials: LDPE layflat tubing was purchased from Environmental Sampling Technologies, St. Joseph, MO. The tubing was a 2.54 cm wide, No. 940, untreated (pure PE; no slip additives, antioxidants, etc.) clear tubing. The wall thickness of this lot ranged from 84 to $89 \mu \mathrm{~m}$. Triolein (1,2,3-tri-[cis-9-octadecenoyl]glycerol) was obtained from Nu-Check Prep Inc. Elysian, MN, this 99\% triolein, Lot T-235-05-L and was further purified at CERC (11) prior to use in SPMD preparation. Florisil ${ }^{\circledR}$ (60-100 mesh) was obtained from Fisher Scientific Company, Pittsburgh, PA. The Florisil ${ }^{\circledR}$ was first heated at $475^{\circ} \mathrm{C}$ for 8 hours, then equilibrated at $130^{\circ} \mathrm{C}$ for 48 hours, and subsequently
stored at room temperature over $\mathrm{P}_{2} \mathrm{O}_{5}$ as desiccant. Silica gel (SG-60, 70-230 mesh) was obtained from Thomas Scientific, Swedesboro, NJ. The silica gel was washed with 40:60 methyl tert-butyl ether:hexane (V:V) followed by 100\% hexane. The silica gel was activated at $130^{\circ} \mathrm{C}$ for a minimum of 72 hours before use and stored at room temperature over $\mathrm{P}_{2} \mathrm{O}_{5}$ as a desiccant. All organic solvents were Optima grade from Fisher Scientific.

SPMD Preparation: The "standard" (6) SPMDs used in this project consisted of a 90 cm length of 2.5 cm wide layflat LDPE tubing containing 1 mL ( $\geq 99 \%$ purity) triolein. The resulting SA-V (membrane surface area to total SPMD volume) ratio is $\approx 90 \mathrm{~cm}^{2} / \mathrm{mL}$ or $\approx$ $460 \mathrm{~cm}^{2} / \mathrm{mL}$ of triolein, where the SPMD consisted of $\approx 20 \%$ triolein. The SPMDs weighed 4.4 to 4.6 g with the active exchanging or sampling surface of $\approx 440 \mathrm{~cm}^{2}$.

Four of the five SPMDs for each deployment site and one of the two SPMD Field Blanks for each deployment site were spiked with $14 \mu \mathrm{~g}$ of each of five perdeuterated compounds 1) Acenaphthylene- $d_{10}$, 2) Acenaphthene- $d_{10}$, 3) Fluorene- $d_{10}, 4$ ) Phenanthrene- $d_{10}$, and 5) Pyrene- $d_{10}$, which are permeability/performance reference compounds (PRCs) A discussion of PRCs follows. The SPMDs were then placed into labeled, solvent rinsed, gas-tight cans. Afterwards, the cans were flushed with argon and sealed. These cans were placed in coolers and then shipped overnight to the Virginia DEQ for deployment.

Permeability/Performance Reference Compounds: PRCs are analytically non-interfering organic compounds with moderate to high fugacity from SPMDs that are added to the
lipid prior to membrane enclosure and field deployment (5). The functional basis of the PRC approach has been previously described in detail by Huckins et al. (10). In this study Phenanthrene- $d_{10}$ was used as the PRC. Of the five PRCs added to the study SPMDs, only Phenanthrene- $d_{10}$ was found to dissipate during deployment at greater than $20 \%$ and less than $80 \%$. The rate constants for Phenanthrene- $d_{10}$ dissipation from SPMDs at each sample site were determined and compared to PRC dissipation rate constants measured during laboratory calibration studies. The values so determined were similar for the six sites and a mean value was used in the calculation of an exposure adjustment factor (EAF). Using the EAF ratios, calibration data (i.e., SPMD uptake rate constants for analytes of interest) were adjusted to more accurately reflect actual in situ sampling rates. As suggested earlier, the effects of exposure conditions on SPMD uptake and dissipation rates are largely a function of 1) exposure medium temperature, 2) facial velocity-turbulence at the membrane surface, which is affected by the design of the deployment apparatus (i.e., baffling of media flow-turbulence), and 3) membrane biofouling. Based on our PRC research (10), the use of EAFs should permit the estimation of analyte water concentrations within $\pm 75 \%$ of the actual time weighted average values.

SPMD Deployment and Retrieval: Study samples, as previously identified, were deployed by personnel of the Virginia DEQ on 3/31/2004. The SPMD deployment containers used were provided by the Virginia DEQ. All sites were visited weekly to ensure that the deployment canisters were fully submersed. All SPMDs were collected on 4/30/2004 and shipped to CERC.

SPMD Storage and Custody: Following receipt of the samples at CERC and prior to processing, the SPMDs were stored in a laboratory freezer at $-15^{\circ} \mathrm{C}$ as described in CERC SOP: P. 453 entitled, "Documentation of Sample Receipt and Storage by Chemical Fate and Dynamics Branch."

Sample Processing and Residue Enrichment : Sample processing was similar to published procedures (5) as described in CERC SOP: P002 entitled, "Procedure for Cleanup and Fractionation of SPMD Dialysates and Extracts of Other Environmental Samples for Chlorinated Pesticides, PAHs, PCBs and Other Targeted Contaminants," with abbreviated details noted in the following sections.

SPMD Membrane Cleaning: The exterior membrane surfaces of all SPMDs were cleaned prior to dialysis. Sealed cans with SPMDs were opened and the SPMDs were removed and momentarily immersed (<30 sec.) in 100 mL of hexane. Then the SPMDs were washed thoroughly to remove all remaining surface adhering material. Any SPMD tether loops outside the lipid containment seals were cut away and discarded. The SPMDs were then immersed in a glass tank containing $1-\mathrm{N} \mathrm{HCl}$ for approximately 30 seconds. Subsequently, they were rinsed with tap water to remove the acid. All surface water was removed from individual SPMDs by successive rinses of acetone followed by isopropanol.

SPMD Dialysis (i.e., Recovery of Analytes): Each SPMD was submersed in 175 mL of hexane in glass jars and was dialyzed individually at $18{ }^{\circ} \mathrm{C}$ for 18 hours. The hexane was removed and transferred into an evaporation flask. A second volume of 175 mL of hexane was added to each sample jar and the SPMDs were dialyzed for an additional 6 hours at $18{ }^{\circ} \mathrm{C}$. The second dialysate was transferred into the flask containing the first dialysate. The dialysates from three SPMDs from each deployment site were combined into a composite. These were reduced to 3 to 5 mL by rotary evaporation, filtered, and quantitatively transferred into test tubes. The solvent volume was then reduced to about 1.0 mL , using high purity nitrogen.

SEC Cleanup: The size exclusion chromatography (SEC) cleanup of study samples followed CERC SOP p. 588 entitled, "The Use of Size Exclusion Chromatography (SEC) in the Cleanup of Dialysates From SPMDs and Other Extracts." A Perkin-Elmer Series 410 HPLC (Perkin-Elmer, Inc., Norwalk, CT) was employed as the solvent delivery system. This HPLC unit was equipped with a Thermo Finnigan AS3000 variable-volume injection auto sampler (Thermo Finnigan, Inc. San Jose, CA). The SEC column was a $300-\mathrm{mm} \times 21.2-\mathrm{mm}$ I.D. (10- $\mu \mathrm{m}$ particle size, 100 Å pore size) Phenogel column (Phenomenex, Inc., Torrance, CA), equipped with a $50-\mathrm{mm} \times 7.5-\mathrm{mm}$ I.D. Phenogel guard column. A DFW-20 series fixed wavelength UV absorption detector (D-Star Instruments, Inc. Manasses, VA) operating at 254 nm, a Hewlett-Packard Co, HP 3396 Series II Integrator (Hewlett Packard, Inc., Palo Alto, CA), and an ISCO Foxy ${ }^{\circledR} 200$ (ISCO, Inc., Lincoln, NE) fraction collector completed the SEC system. The isocratic
mobile phase was 98:2 (V:V) dichloromethane:methanol (DCM:MeOH) with a flow rate of 4.0 mL per minute.

In keeping with CERC SOP p.588, the SEC system was calibrated on a daily basis using compounds of environmental interest and potentially interfering materials. These compounds, in sequence of elution, were diethylhexylphthalate (DEHP; a common plasticizer with lipid-like chromatographic behavior), biphenyl and naphthalene (small aromatic analytes), coronene (a large PAH later eluting than any anticipated analyte), and elemental sulfur (a problematic interference frequently encountered in environmental samples). SEC cleanup was accomplished using a collect fraction (i.e., window in which target analytes elute) determined by the calibration of the system on the day of operation. The fractions collected were amended with approximately 2 mL of isooctane, reduced to a volume of about 1 mL on a rotary evaporation system, and quantitatively transferred with hexane into appropriately labeled test tubes.

Post-SEC Column Chromatography: After SEC cleanup, samples were further processed using open column chromatography. The fractions ( 1 mL volume) were applied to Florisil ${ }^{\circledR}$ columns ( 5 g ) and subsequently eluted with 60 mL of 75:25 (V:V) methyl tertbutyl ether:hexane. Following volume reduction to 0.5 mL , each sample was applied to a silica gel (SG) column (5 g). Two fractions were collected; fraction SG-1 (46 mL of hexane) and fraction SG-2 (55 mL of 40:60 [V:V] methyl tert-butyl ether:hexane). This enrichment procedure provided fractions for the analysis of PCBs (in SG-1) and PRCs (in SG-2).

Gas Chromatographic Analysis of PRCs: Analyses of SG-2 Fractions for PRCs was conducted on a Hewlett Packard 5890 series gas chromatograph (GC) equipped with a Hewlett Packard 7673A autosampler (Hewlett Packard, Inc., Palo Alto, CA). In all analyses, $1.0 \mu \mathrm{~L}$ of sample extract was injected using the "cool-on-column" technique with hydrogen as the carrier gas. Analysis of the SG-2 fractions was performed using a DB-5 (30 m x 0.25 mm i.d x $0.25 \mu \mathrm{~m}$ film thickness.) capillary column (J\&W Scientific, Folsom, CA) with the following temperature program: injection at $60^{\circ} \mathrm{C}$ with a 2 minute hold at $60^{\circ} \mathrm{C}$, then $10^{\circ} \mathrm{C} / \mathrm{min}$ to $110^{\circ} \mathrm{C}$ and held for 5 minutes at $110^{\circ} \mathrm{C}$, followed by $3^{\circ} \mathrm{C} / \mathrm{min}$ to $200^{\circ} \mathrm{C}$ and held for 10 minutes at $200^{\circ} \mathrm{C}$, then $4^{\circ} \mathrm{C} / \mathrm{min}$ to $310^{\circ} \mathrm{C}$ and held at $310^{\circ} \mathrm{C}$ for 3 minutes. Detection was performed using a Hewlett Packard flame ionization detector (FID) operating at $330^{\circ} \mathrm{C}$. Quantitation was accomplished using an eight point calibration curve spanning a 50 -fold range of concentration from 0.2 to10 $\mu \mathrm{g} / \mathrm{mL}$ for each of the five perdeuterated PRCs. 4-Terphenyl- $d_{14}$ was used as the instrumental internal standard (IIS). Figures 1,2, and 3 show representative chromatograms of standards, field blank SPMDs, and deployed SPMDs analyzed for PRCs.

Gas Chromatographic Analysis of PCB Congeners: Gas chromatographic analyses of SG1 fractions for PCB congeners were conducted using a Hewlett Packard 5890 series gas chromatograph (GC) equipped with two Hewlett Packard 7673A autosamplers (Hewlett Packard, Inc., Palo Alto, CA) to allow for dual column injection. In all analyses, $1.0 \mu \mathrm{~L}$ of sample extract was injected using the "cool-on-column" technique with hydrogen as
the carrier gas (25 psig). Analyses were performed using both a DB-5 (60 m x 0.25 mm i.d. x $0.25 \mu \mathrm{~m}$ film thickness) capillary column (J\&W Scientific, Folsom, CA) and a DB17 ( 60 m x 0.25 mm i.d. x $0.25 \mu \mathrm{~m}$ film thickness) capillary column (J\&W Scientific, Folsom, CA) with the following temperature program: injection at $60^{\circ} \mathrm{C}$; then ramped at $15{ }^{\circ} \mathrm{C} / \mathrm{min}$ to $150{ }^{\circ} \mathrm{C}$; followed by $1^{\circ} \mathrm{C} / \mathrm{min}$ ramp to $260{ }^{\circ} \mathrm{C}$; and finally ramped at 10 ${ }^{\circ} \mathrm{C} / \mathrm{min}$ to $300{ }^{\circ} \mathrm{C}$ with a final hold of 15 minutes at $300{ }^{\circ} \mathrm{C}$. The electron capture detectors (ECDs, Hewlett Packard, Inc., Palo Alto, CA) were maintained at $330{ }^{\circ} \mathrm{C}$ with nitrogen at 55-60 mL/minute as make-up gas. Individual PCB congeners were identified on one or both GC capillary columns based upon known retention times for each congener in the calibration standards. The best resolved peak was picked from the column giving the best resolution, with some congeners being analyzed on both columns for confirmatory analysis. A 1:1:1:1 mixture of Aroclor ${ }^{\circledR} 1242,1248,1254$, and 1260 was used to produce the PCB congener calibration standards. These standards were quantified based on pure primary PCB standards (Accustandard, New Haven, CT) and were used as secondary standards (12) at five calibration levels to quantify up to 140 congeners and combined congener peaks in the study samples by an internal standard method. The levels of the PCB standards spanned a 160-fold concentration range from 50 to 8,000 total ng/mL. PCB congener I-30 was employed as the instrumental internal standard, and congeners I-30 and I-207 were used as retention time references.

Determinations for total PCBs sequestered in deployed SPMDs were measured as the sums of the individual analytical responses from 140 individual PCB congeners as detected during GC-ECD analysis. Figures 4,5, and 6 show representative
chromatograms of SG-1 fractions for standards, field blank SPMDs, and deployed SPMDs analyzed for PCBs.

Quality Control: Field blank SPMDs accompanied the deployment SPMDs during transportation to the field, deployment, retrieval, and transportation to CERC. These field blanks were processed and analyzed exactly as the deployed SPMDs. The field blank samples exhibited no coincident GC peaks at levels significantly higher than those associated with the laboratory control SPMDs and were indicative of successful deployments and retrievals.

The method detection limit (MDL) and method quantitation limit (MQL) for analysis of SPMD samples were determined for each analyte by measuring the values of coincident GC-ECD or GC-FID peaks for each compound in all blank samples processed with this study. The MDL was defined as the mean plus three standard deviations of values so determined (13). The MQL was defined as the mean plus 10 standard deviations of values so determined (13). For individual analytes having no coincident GC peak, an assumed value equal to the low sample reject for the GC method was used to calculate the mean. In the cases where the MQLs were below the level of the calibration curve employed in the GC-analysis, the MQLs were set at the value of the lowest level of the calibration curve employed in quantifying the analyte levels. The background, MDLs and MQLs for analysis of the study samples for PCBs and PRCs are presented in Table I.

QC checks were employed to demonstrate an acceptable outcome of sample analyses. These checks included (but were not limited to) monitoring the recovery of PCBs through
the entire dialysis, fractionation and enrichment procedures by spiking a blank SPMD with a known quantity of PCBs (nominal amount of total PCBs was 4000 ng ) and measuring recovery following dialysis, SEC , Florisil ${ }^{\circledR}$, and Silica Gel chromatographic cleanup. Recoveries of individual PCB congeners and total PCBs are given in Table II. These values are consistent with those typically obtained at CERC.

## RESULTS AND DISCUSSIONS

Observations and Findings: The deployed SPMDs from this study were processed concurrently with the above referenced quality control samples. Therefore, the results obtained from processing and analyses conducted on riverine exposed SPMDs were taken to be similar to the observed results for the quality control samples described.

The reported values of PCB levels are from SG-1 fraction analyses only. The low recoveries observed (Table II) for 12 of the 140 individual congeners is an artifact of the silica gel fractionation. These congeners elute from silica gel, in part or in whole, in the SG-2 fraction. A more rigorous analysis, including determination of PCBs in the SG-2 fraction, was beyond the scope of this project and was considered to be an excessive expenditure of time and effort for a minimal increase of useful information. The SG-2 fraction was not included in the analyses for total PCBs because historic CERC data shows that carry-over of these congeners into SG-2 (using the materials and methods described in this report) represents less than $5 \%$ of total PBCs. This level of potential processing loss was not considered to be significant in the determinations of total PCB levels wherein reported results were not corrected for recovery and were reported to only two significant figures.

During the gas chromatographic analysis of study samples for PCB congeners, two GC columns were used to obtain sufficient resolution for quantitation of the targeted PCB congeners. Quantitation for each congener was done using the column giving the better resolution where the second column then became the conformational analysis (Table III).

The results of the gas chromatographic analyses of the deployed SPMDs are given for all PCB congeners in Tables IV through IX. The results of the gas chromatographic analyses of the deployment SPMDs are given for all PRCs in Table X. PCB congener analyses results are corrected for background, but are not corrected for recovery and therefore represent minimum levels of PCBs within the aquatic environments of the Bluestone River watershed. Although the backgrounds, MDLs, and MQLs (Table I) are given in units of ng/3-SPMD composite (because this is representative of the sample size and matrix as injected), SPMD analysis results (Tables IV through IX) are given in units of ng/SPMD. Raw results (after subtracting out background) were divided by three to give these values. These units were used because they are applicable to deriving water concentrations from modeling equations.

## Derivation of Water Concentrations from SPMD Levels (Modeling): SPMD uptake

 kinetic data are required to accurately estimate aquatic concentrations of environmental contaminants. Using models previously developed (2, 5, 10, 14), data from the analysis of the PRC levels, and from calibration studies, the bioavailable (i.e., dissolved phase) concentrations of PCBs in SPMDs exposed to water in the Bluefield River Watershed canbe estimated. For compounds with $\log K_{\text {ow }}$ values $\geq 5.0$, sampling is integrative and reported water concentration values represent a time weighted average of residue concentration (i.e., residues were accumulated in an additive manner throughout the exposure with no significant losses of sequestered PCBs) during the 30 day exposure period. For compounds with $\log \mathrm{K}_{\mathrm{ow}} \mathrm{s}<5.0$, equilibrium is reached in less than 30 days. Therefore, sampling of these compounds is not integrative throughout the whole exposure period and reported water concentrations represent a smaller window of time than the 30 day exposure. Regardless of these considerations, water concentrations can be determined by using different assumptions and models (5).

An example of the overall modeling procedure for compounds with $\log \mathrm{K}_{\text {ows }} \geq 5.0$ is as follows. Note that Huckins et al. (5) have also described modeling procedures for water concentration estimates of compounds with $\log \mathrm{K}_{\mathrm{ow}} \mathrm{S}<5.0$. The analyte sampling rate ( $\mathrm{R}_{\text {sw }}$ ) is determined from laboratory exposures conducted under some of the same conditions (i.e., water temperature and exposure duration) as the current study. The linear or integrative uptake of OCs by SPMDs as applicable to PCBs has been described by Huckins, et al. (5) as follows:

$$
\begin{equation*}
\mathrm{C}_{\mathrm{SPMD}}=\mathrm{C}_{\mathrm{w}} \mathrm{k}_{0} \mathrm{~K}_{\mathrm{mw}} \mathrm{At} / \mathrm{V}_{\text {SPMD }} \tag{1}
\end{equation*}
$$

substituting $\mathrm{R}_{\mathrm{sw}}$ for $\mathrm{k}_{0} \mathrm{~K}_{\mathrm{mw}} \mathrm{A}$ in equation 1 gives

$$
\begin{equation*}
\mathrm{C}_{\text {SPMD }}=\mathrm{C}_{\mathrm{w}} \mathrm{R}_{\mathrm{sw}} \mathrm{t} / \mathrm{V}_{\text {SPMD }} \tag{2}
\end{equation*}
$$

where $\mathrm{C}_{\text {SPMD }}$ is the concentration of the PCB in the whole SPMD (i.e., the membrane + lipid), $\mathrm{C}_{\mathrm{w}}$ is the concentration of the PCB in the water, t is the exposure time in days, and $\mathrm{V}_{\text {SPMD }}$ is the volume of the SPMD. Rearranging equation 2 results in

$$
\begin{equation*}
\mathrm{C}_{\mathrm{w}}=\mathrm{C}_{\mathrm{SPMD}} \mathrm{~V}_{\mathrm{SPMD}} / \mathrm{R}_{\mathrm{sw}} \mathrm{t} \tag{3}
\end{equation*}
$$

In the present case we use the uptake rate constant ( $\mathrm{k}_{\mathrm{uw}}$ ) defined as L/d g (liter/day.gram) of SPMD (membrane + lipid).

$$
\begin{equation*}
\mathrm{C}_{\mathrm{w}}=\mathrm{C}_{\text {SPMD }} /\left(\mathrm{R}_{\mathrm{sw}} / \mathrm{M}_{\mathrm{SPMD}}\right) \mathrm{t}=\mathrm{C}_{\mathrm{w}}=\mathrm{C}_{\text {SPMD }} / \mathrm{k}_{\mathrm{uw}} \mathrm{t} \tag{4}
\end{equation*}
$$

where $\mathrm{M}_{\text {SPMD }}$ is the mass of the SPMD and is substituted for $\mathrm{V}_{\text {SPMD }}$.

SPMD sampling rates can change due to changes in temperature, turbulence/facial velocity of water at the membrane surface, and buildup of periphyton on the membrane surface. To account for changes in these variables from the laboratory calibration studies, PRCs are used to allow estimation of actual in situ sampling rates. The PRC concept was described earlier. However, models to enable the use of a PRC were not. Measuring the loss of a PRC over a study exposure period provides in situ $k_{e}$ values which when compared to the calibration $\mathrm{k}_{\mathrm{e}}$ values can serve as an indicator of differences in the environmental conditions. If large differences exist between the $\mathrm{k}_{\mathrm{e}}$ calibration and exposure values, adjustments must be made to the laboratory calibration data (Table X). The $\mathrm{k}_{\text {eprc }}$ values are derived as follows

$$
\begin{align*}
& \mathrm{C}_{\text {SPMD }}=\mathrm{C}_{\text {SPMDo }} \exp \left(-\mathrm{k}_{\mathrm{eprc}} \mathrm{t}\right)  \tag{5}\\
& \mathrm{k}_{\mathrm{eprc}}=\ln \left(\mathrm{C}_{\text {SPMDo }} / \mathrm{C}_{\text {SPMD }}\right) / \mathrm{t} \tag{6}
\end{align*}
$$

where C $_{\text {SPMDo }}$ is the day 0 level of the PRC and C $_{\text {SPMD }}$ is the concentration of PRC remaining in the SPMD following exposure. Comparison of the $\mathrm{k}_{\text {eprc }}$ values derived from the field-exposed SPMDs (Equations 7 or 8), to the $k_{e}$ values of the PRCs measured in SPMD calibration exposures (i.e., $\mathrm{k}_{\text {eprc }} / \mathrm{k}_{\text {ec }}$ ), provides an estimate of the relative effect of
environmental variables on SPMD sampling. Laboratory $\mathrm{k}_{\mathrm{ec}}$ values of PRCs are determined by direct measurement or by

$$
\begin{equation*}
\mathrm{k}_{\mathrm{ec}}=\mathrm{R}_{\mathrm{s}} / \mathrm{K}_{\text {SPMD }} \mathrm{V}_{\text {SPMD }} \mathrm{d}_{\text {SPMD }} \tag{7}
\end{equation*}
$$

where K $_{\text {SPMD }}$ is the equilibrium SPMD-water partition coefficient and $\mathrm{d}_{\text {SPMD }}$ is the SPMD density ( $\mathrm{g} / \mathrm{mL}$ ), which is 0.91 . Estimates of in situ $\mathrm{R}_{\mathrm{s}}$ values from the $\mathrm{k}_{\mathrm{ec}}$ of PRCs can be made with the following relationship

$$
\begin{equation*}
\mathrm{R}_{\mathrm{sf}}=\left(\mathrm{k}_{\mathrm{eprc}} / \mathrm{k}_{\mathrm{ec}}\right) \mathrm{R}_{\mathrm{sc}} \tag{8}
\end{equation*}
$$

These models and assumptions as well as others have been incorporated into an Excel spreadsheet (5) for rapid estimation of water concentrations and the Excel calculator was used for water concentration estimates in this work (Tables XI through XVI). Note that an average temperature of $11.8^{\circ} \mathrm{C}$ was assumed for SPMD exposure at all sample sites because this temperature is the most representative water temperature for which calibration data exists (14). The $\mathrm{k}_{\text {eprc }}$ values (Equation \# 6 above) incorporated into the Excel calculator for water concentration estimations was the average obtained from the PRC analyses of the SG-2 fractions from four of the six study sites. The SG-2 fractions from the other two sites contained chromatographic interferences which prohibited the determinations of PRC compound levels in these samples. Therefore, the values obtained for the other four sites (being within a range of $20 \%$ ) were averaged and mean was applied uniformly to all six sites (Table X).

The analytical sampling rate $\left(\mathrm{R}_{\mathrm{sw}}\right)$ value for total PCBs which was incorporated into the Excel calculator for water concentration estimations (Table XI through XVI) from total

PCB levels, $(5,14)$ is a weighted average of $\mathrm{R}_{\mathrm{sw}}$ values from the eighty individual PCB congeners for which calibration data exists (14). This weighted average does not reflect the distribution of PCB congeners seen at each study site, does not reflect the relative amounts of each congener seen at each study site, and is not representative of any specific Aroclor ${ }^{\circledR}$ mixture.

When congener distributions are known, as in this study where analytical data was obtained for 140 congeners, an alternate method of estimation of water concentrations would involve the inclusion of the remaining 60 congeners (for which calibration data does not exists) into the Excel calculator for water concentration estimations. In this case, average $\mathrm{R}_{\mathrm{sw}}$ values for these 60 additional congeners can be employed to estimate the water concentration of these congeners. The summation of the individual water concentrations of the 140 total congeners would then be taken as an estimation of the total PCB water concentration. The overall accuracy of this technique has not been fully evaluated using these 60 estimated $\mathrm{R}_{\mathrm{sw}}$ values. The results so obtained are within the two-fold margin of error associated in determinations of water concentrations from SPMDs $(5,14)$. For example, the estimated water concentration of total PCB from site 9BPB000.02 using this method would be $2900 \mathrm{ng} / \mathrm{L}$ as compared to $3700 \mathrm{ng} / \mathrm{L}$ (Table XVI) as reported using the more traditional modeling methods as described above.

In addition to the inherent limitations of estimation of water concentrations by application of SPMD technology (i.e. the two-fold margin of error), the utility of this technology is further limited by three other factors, 1) Sample size, i.e. number of SPMDs used, 2)

Length of deployment, and 3) sensitivity of the analytical method used to determine the levels of contaminants sequestered. Insertion of the MQL values (Table I) for individual congeners and total PCBs specific to this project with adjustment of units to ng/SPMD into the Excel calculator for water concentration estimations gives minimum values at which water concentrations can be estimated (Table XVII) under the specific conditions as they existed during the deployment phase of this study.

Conclusions: Levels of PCBs within the Bluestone River watershed differ significantly from site to site both in quantity and profile. The observed levels of total PCBs sequestered are given below in order of highest to lowest.

## DEQ CEDS ID Total PCBs Sequestered by SPMDs (ng/SPMD)

9-BPB000.02 ..... 505
9-BST021.26 ..... 180
9-WVC000.05 ..... 170
9-BFK000.02 ..... 167
VA0025054-001 ..... 89
9-BST028.90 ..... 32

The GC-ECD profiles of Aroclors ${ }^{\circledR} 1242,1248,1254$, and 1260 along with those observed from the six study sites are presented in Figures 7 through 16. These graphic representations are intended to assist, by pattern recognition, in the identification of
potential sources of introduction of PCBs into the Bluestone River watershed. Figures are given at the same scale to give a visual representation of relative quantity of sequestered PCBs.

The high recovery (Table X) of pyridine- $d_{10}$ (a PRC) indicates that photolysis of PRC did not occur during deployment and that the EAFs used in calculating water concentrations are accurate. The elucidation of the potential biological effects from exposure to these levels of PCBs will require further research. While the chronic effects on aquatic organisms present in the Bluestone River watershed from exposure to the PCBs found in the SPMD samples is unknown, the water concentrations of the PCBs found in these samples may be of concern.

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Table I
Background, MDL, \& MQL Values for PCB Congeners

| Congener | Background ng / 3SPMD | MDL ng / 3SPMD | MQL ng / 3SPMD |
| :---: | :---: | :---: | :---: |
| Identification | Composite | Composite | Composite |
| 001 | 0.15 | 0.52 | 5.33 |
| 003 | 0.01 | 0.07 | 0.73 |
| 004 | 0.03 | 0.16 | 1.69 |
| 005 | 0.00 | 0.01 | 0.10 |
| 006 | 0.05 | 0.20 | 2.09 |
| 007 | 0.06 | 0.10 | 1.01 |
| 008 | 0.22 | 0.54 | 5.49 |
| 009 | 0.02 | 0.09 | 0.88 |
| 010 | 0.00 | 0.01 | 0.12 |
| 011 | 0.00 | 0.01 | 0.10 |
| 015 | 0.04 | 0.13 | 1.30 |
| 016 | 0.04 | 0.08 | 0.77 |
| 017 | 0.12 | 0.23 | 2.30 |
| 018 | 0.31 | 0.54 | 5.49 |
| 019 | 0.07 | 0.34 | 3.45 |
| 020 | 0.00 | 0.02 | 0.17 |
| 022 | 0.04 | 0.08 | 0.87 |
| 024 | 0.00 | 0.01 | 0.10 |
| 025 | 0.01 | 0.02 | 0.20 |
| 026 | 0.04 | 0.08 | 0.82 |
| 027 | 0.01 | 0.03 | 0.35 |
| 028 | 0.14 | 0.30 | 3.01 |
| 031 | 0.22 | 0.44 | 4.44 |
| 032 | 0.05 | 0.12 | 1.24 |
| 033 | 0.11 | 0.23 | 2.31 |
| 034 | 0.00 | 0.01 | 0.06 |
| 035 | 0.02 | 0.07 | 0.67 |
| 037,059 | 0.03 | 0.04 | 0.42 |
| 040 | 0.03 | 0.07 | 0.74 |
| 041 | 0.01 | 0.03 | 0.33 |
| 042 | 0.10 | 0.19 | 1.92 |
| 043 | 0.01 | 0.04 | 0.39 |
| 044 | 1.57 | 3.24 | 32.9 |
| 045 | 0.06 | 0.11 | 1.14 |
| 046 | 0.02 | 0.05 | 0.55 |
| 047 | 0.09 | 0.21 | 2.10 |

Table I (continued)
Background, MDL, \& MQL Values for PCB Congeners

| Congener | Background ng / 3SPMD | MDL ng / 3SPMD | MQL ng / 3SPMD |
| :---: | :---: | :---: | :---: |
| Identification | Composite | Composite | Composite |
| 048 | 0.05 | 0.14 | 1.41 |
| 049 | 1.07 | 2.19 | 22.3 |
| 051 | 0.02 | 0.04 | 0.38 |
| 052 | 5.45 | 10.7 | 109 |
| 053 | 0.31 | 0.53 | 5.42 |
| 054 | 0.01 | 0.04 | 0.39 |
| 055 | 0.00 | 0.01 | 0.12 |
| 056,060 | 0.10 | 0.19 | 1.97 |
| 057 | 0.01 | 0.04 | 0.44 |
| 058 | 0.00 | 0.01 | 0.10 |
| 063 | 0.01 | 0.02 | 0.21 |
| 064 | 0.30 | 0.62 | 6.29 |
| 066 | 0.15 | 0.29 | 2.99 |
| 067 | 0.00 | 0.02 | 0.17 |
| 069 | 0.01 | 0.02 | 0.22 |
| 070 | 0.55 | 1.15 | 11.8 |
| 071 | 0.12 | 0.21 | 2.18 |
| 072 | 0.01 | 0.06 | 0.60 |
| 074 | 0.11 | 0.21 | 2.17 |
| 075 | 0.02 | 0.05 | 0.52 |
| 082 | 0.11 | 0.23 | 2.37 |
| 083 | 0.02 | 0.05 | 0.49 |
| 084 | 0.78 | 1.65 | 16.8 |
| 085 | 0.00 | 0.01 | 0.10 |
| 086 | 0.01 | 0.03 | 0.28 |
| 087 | 0.53 | 1.13 | 11.5 |
| 090 | 0.02 | 0.05 | 0.47 |
| 091 | 0.35 | 0.75 | 7.65 |
| 092 | 0.33 | 0.70 | 7.10 |
| 095 | 3.07 | 6.24 | 63.4 |
| 096 | 0.43 | 2.52 | 25.9 |
| 097 | 0.40 | 0.88 | 8.96 |
| 099 | 0.46 | 0.98 | 9.98 |
| 101 | 1.42 | 3.02 | 30.7 |
| 102 | 0.06 | 0.12 | 1.23 |
| 105 | 0.18 | 0.42 | 4.24 |

Table I (continued)
Background, MDL, \& MQL Values for PCB Congeners

| Congener Identification | Background ng / 3SPMD | MDL <br> ng / 3SPMD | MQL <br> ng / 3- <br> SPMD |
| :---: | :---: | :---: | :---: |
| Identification | Composite | Composite | Composite |
| 109 | 0.05 | 0.12 | 1.18 |
| 110 | 1.06 | 2.31 | 23.6 |
| 112 | 0.00 | 0.01 | 0.10 |
| 113 | 0.00 | 0.02 | 0.20 |
| 114 | 0.00 | 0.02 | 0.17 |
| 115 | 0.02 | 0.07 | 0.67 |
| 117 | 0.04 | 0.06 | 0.64 |
| 118 | 0.64 | 1.43 | 14.6 |
| 119 | 0.01 | 0.04 | 0.37 |
| 122 | 0.00 | 0.01 | 0.06 |
| 123 | 0.01 | 0.02 | 0.23 |
| 128 | 0.08 | 0.18 | 1.79 |
| 129 | 0.03 | 0.06 | 0.64 |
| 130 | 0.02 | 0.07 | 0.69 |
| 131 | 0.01 | 0.05 | 0.47 |
| 132 | 0.25 | 0.61 | 6.25 |
| 133 | 0.01 | 0.02 | 0.22 |
| 134 | 0.06 | 0.12 | 1.22 |
| 136 | 0.24 | 0.53 | 5.36 |
| 137 | 0.02 | 0.06 | 0.57 |
| 138 | 0.42 | 0.88 | 8.97 |
| 139 | 0.01 | 0.02 | 0.22 |
| 141 | 0.10 | 0.22 | 2.28 |
| 144 | 0.05 | 0.10 | 0.97 |
| 146 | 0.07 | 0.15 | 1.57 |
| 147 | 0.01 | 0.02 | 0.21 |
| 149 | 0.61 | 1.31 | 13.3 |
| 151 | 0.16 | 0.31 | 3.17 |
| 153 | 0.56 | 1.12 | 11.4 |
| 156 | 0.05 | 0.12 | 1.18 |
| 157 | 0.01 | 0.03 | 0.34 |
| 158 | 0.07 | 0.15 | 1.53 |
| 163 | 0.10 | 0.18 | 1.79 |
| 164 | 0.04 | 0.09 | 0.88 |
| 166 | 0.00 | 0.01 | 0.10 |
| 167 | 0.01 | 0.01 | 0.10 |

Table I (continued)
Background, MDL, \& MQL Values for PCB Congeners

| Congener | Background ng / 3SPMD | MDL ng / 3SPMD | MQL ng / 3SPMD |
| :---: | :---: | :---: | :---: |
| Identification | Composite | Composite | Composite |
| 170 | 0.12 | 0.29 | 2.94 |
| 171 | 0.01 | 0.04 | 0.39 |
| 172 | 0.00 | 0.01 | 0.13 |
| 173 | 0.00 | 0.01 | 0.10 |
| 174 | 0.05 | 0.12 | 1.20 |
| 175 | 0.00 | 0.01 | 0.10 |
| 176 | 0.00 | 0.01 | 0.13 |
| 177 | 0.02 | 0.04 | 0.45 |
| 178 | 0.01 | 0.02 | 0.22 |
| 179 | 0.03 | 0.07 | 0.76 |
| 180 | 0.10 | 0.18 | 1.86 |
| 183 | 0.03 | 0.07 | 0.72 |
| 185 | 0.00 | 0.01 | 0.10 |
| 187 | 0.06 | 0.11 | 1.14 |
| 189 | 0.00 | 0.01 | 0.10 |
| 190 | 0.01 | 0.05 | 0.48 |
| 191 | 0.00 | 0.01 | 0.10 |
| 193 | 0.00 | 0.01 | 0.10 |
| 194 | 0.01 | 0.02 | 0.22 |
| 195 | 0.00 | 0.03 | 0.32 |
| 196 | 0.02 | 0.03 | 0.34 |
| 197 | 0.02 | 0.03 | 0.32 |
| 198 | 0.00 | 0.01 | 0.10 |
| 199 | 0.02 | 0.05 | 0.53 |
| 200 | 0.03 | 0.11 | 1.14 |
| 201 | 0.00 | 0.02 | 0.17 |
| 202 | 0.00 | 0.02 | 0.18 |
| 203 | 0.01 | 0.03 | 0.35 |
| 205 | 0.00 | 0.01 | 0.10 |
| 206 | 0.01 | 0.04 | 0.41 |
| 208 | 0.11 | 0.17 | 1.67 |
| 209 | 0.46 | 0.73 | 7.35 |
| Total PCBs* | 26.1 | 10.0* | 50.0* |

* The values reported for MDLs and MQLs (of Total PCBs) are representative of residues following substraction of summed values for individual congener signals (i.e.values above total background). Units of ng/3-SPMD composite are representative of sample sizes and matrix as injected

Table II
Recovery of PCBs From Spiked SPMDs

| Congener <br> Identification | Spiking <br> Level <br> ng | SPMD <br> Spike <br> ng | Recovery <br> $\%$ |
| :---: | :---: | :---: | :---: |
| 001 | 11.6 | 0.09 | $0.8^{*}$ |
| 003 | 3.85 | 0.13 | $3.4^{*}$ |
| 004 | 40.7 | 0.78 | $1.9^{*}$ |
| 006 | 20.5 | 12.9 | 62.9 |
| 007 | 3.38 | 1.07 | $31.7^{*}$ |
| 008 | 86.8 | 54.2 | 62.4 |
| 009 | 6.18 | 2.80 | $45.3^{*}$ |
| 010 | 2.26 | 0.18 | $8.0^{*}$ |
| 015 | 26.9 | 21.0 | 78.1 |
| 016 | 67.2 | 15.0 | $22.3^{*}$ |
| 017 | 60.5 | 43.7 | 72.2 |
| 018 | 156 | 104 | 66.9 |
| 019 | 13.9 | 0.22 | $1.6^{*}$ |
| 020 | 6.53 | 5.04 | 77.2 |
| 022 | 51.2 | 39.9 | 77.9 |
| 024 | 1.71 | 0.01 | $0.8^{*}$ |
| 025 | 8.69 | 6.98 | 80.3 |
| 026 | 22.3 | 17.9 | 80.5 |
| 027 | 6.94 | 4.87 | 70.2 |
| 028 | 121 | 98.7 | 81.8 |
| 031 | 125 | 101 | 80.4 |
| 032 | 36.3 | 27.5 | 75.8 |
| 033 | 87.3 | 71.0 | 81.3 |
| 034 | 0.42 | 0.37 | 88.1 |
| 035 | 0.91 | 0.93 | 102 |
| 037,059 | 11.9 | 10.8 | 90.2 |
| 040 | 25.6 | 12.6 | 49.2 |
| 041 | 18.3 | 11.8 | 64.2 |
| 042 | 37.8 | 30.4 | 80.5 |
| 043 | 5.55 | 4.33 | 78.0 |
| 044 | 134 | 106 | 79.0 |
| 045 | 30.5 | 13.8 | 45.4 |
| 046 | 13.0 | 4.98 | 38.2 |
| 047 | 25.5 | 20.6 | 80.9 |

Table II (continued)
Recovery of PCBs From Spiked SPMDs

| Congener <br> Identification | Spiking <br> Level <br> ng | SPMD <br> Spike <br> ng | Recovery <br> $\%$ |
| :---: | :---: | :---: | :---: |
| 048 | 37.0 | 27.4 | 74.2 |
| 049 | 103 | 82.5 | 80.4 |
| 051 | 6.20 | 4.85 | 78.2 |
| 052 | 177 | 143 | 81.0 |
| 053 | 27.7 | 20.8 | 75.1 |
| 054 | 0.16 | 0.03 | $18.8^{*}$ |
| 055 | 1.88 | 1.56 | 83.0 |
| 056,060 | 84.4 | 70.8 | 83.9 |
| 057 | 0.75 | 0.64 | 85.3 |
| 058 | 0.31 | 0.19 | 61.3 |
| 063 | 4.41 | 3.67 | 83.2 |
| 064 | 64.4 | 52.7 | 81.8 |
| 066 | 96.9 | 82.6 | 85.3 |
| 067 | 3.28 | 2.71 | 82.6 |
| 069 | 0.20 | 0.20 | 100 |
| 070 | 161 | 134 | 83.2 |
| 071 | 34.6 | 28.1 | 81.2 |
| 072 | 1.07 | 0.45 | $42.1^{*}$ |
| 074 | 56.1 | 47.8 | 85.2 |
| 075 | 2.32 | 1.45 | 62.5 |
| 082 | 20.8 | 16.4 | 78.9 |
| 083 | 2.35 | 1.86 | 79.1 |
| 084 | 49.0 | 37.4 | 76.3 |
| 086 | 3.74 | 1.10 | 29.4 |
| 087 | 68.4 | 57.1 | 83.4 |
| 090 | 3.66 | 2.92 | 79.8 |
| 091 | 22.6 | 18.3 | 80.9 |
| 092 | 30.2 | 24.4 | 81.0 |
| 095 | 140 | 111 | 79.6 |
| 096 | 1.95 | 1.28 | 65.6 |
| 097 | 55.9 | 42.5 | 76.0 |
| 099 | 52.9 | 44.0 | 83.3 |
| 101 | 139 | 114 | 82.1 |
| 102 | 3.78 | 3.03 | 80.2 |
| 105 | 48.3 | 41.5 | 86.0 |
|  |  |  |  |

Table II (continued)
Recovery of PCBs From Spiked SPMDs

| Congener <br> Identification | Spiking <br> Level <br> ng | SPMD <br> Spike <br> ng | Recovery <br> $\%$ |
| :---: | :---: | :---: | :---: |
| 109 | 10.7 | 9.03 | 84.5 |
| 110 | 145 | 118 | 81.3 |
| 112 | 0.64 | 0.63 | 98.4 |
| 113 | 0.52 | 0.48 | 92.3 |
| 114 | 4.88 | 3.92 | 80.3 |
| 115 | 4.54 | 3.89 | 85.7 |
| 117 | 8.02 | 4.77 | 59.5 |
| 118 | 110 | 92.8 | 84.5 |
| 119 | 2.43 | 2.02 | 83.1 |
| 122 | 1.50 | 1.26 | 84.0 |
| 123 | 2.19 | 1.28 | 58.6 |
| 128 | 26.4 | 21.2 | 80.2 |
| 129 | 8.29 | 6.67 | 80.5 |
| 130 | 8.50 | 6.84 | 80.5 |
| 131 | 2.95 | 2.40 | 81.4 |
| 132 | 54.5 | 41.8 | 76.7 |
| 133 | 2.34 | 1.86 | 79.5 |
| 134 | 11.5 | 9.22 | 80.5 |
| 136 | 37.7 | 26.0 | 68.9 |
| 137 | 7.07 | 5.64 | 79.8 |
| 138 | 124 | 101 | 81.5 |
| 139 | 3.59 | 2.25 | 62.7 |
| 141 | 49.6 | 36.7 | 74.0 |
| 144 | 15.1 | 11.0 | 73.2 |
| 146 | 24.4 | 19.4 | 79.4 |
| 147 | 0.65 | 0.53 | 81.5 |
| 149 | 156 | 122 | 77.9 |
| 151 | 53.9 | 42.1 | 78.2 |
| 153 | 160 | 130 | 81.3 |
| 156 | 17.5 | 15.5 | 88.7 |
| 157 | 3.36 | 2.72 | 81.0 |
| 158 | 21.7 | 17.4 | 80.3 |
| 163 | 38.5 | 31.5 | 81.8 |
| 164 | 12.2 | 9.58 | 78.5 |
| 166 | 0.59 | 0.49 | 83.1 |
| 167 | 4.68 | 3.83 | 81.8 |
|  |  |  |  |

Table II (continued)
Recovery of PCBs From Spiked SPMDs

| Congener <br> Identification | Spiking <br> Level <br> ng | SPMD <br> Spike <br> ng | Recovery <br> $\%$ |
| :---: | :---: | :---: | :---: |
| 170 | 87.6 | 68.8 | 78.5 |
| 171 | 18.8 | 14.8 | 79.1 |
| 172 | 10.1 | 7.72 | 76.4 |
| 173 | 1.30 | 1.01 | 77.7 |
| 174 | 73.1 | 56.5 | 77.2 |
| 175 | 3.39 | 2.58 | 76.1 |
| 176 | 6.02 | 4.45 | 73.9 |
| 177 | 32.2 | 24.9 | 77.2 |
| 178 | 12.6 | 9.63 | 76.4 |
| 179 | 25.7 | 23.5 | 91.4 |
| 180 | 113 | 87.0 | 76.9 |
| 183 | 39.4 | 30.6 | 77.7 |
| 185 | 7.73 | 6.00 | 77.6 |
| 187 | 65.5 | 50.1 | 76.6 |
| 189 | 2.25 | 1.77 | 78.7 |
| 190 | 12.9 | 10.9 | 84.4 |
| 191 | 2.44 | 1.90 | 77.9 |
| 193 | 6.35 | 4.88 | 76.9 |
| 194 | 22.7 | 16.8 | 73.6 |
| 195 | 12.7 | 9.84 | 77.3 |
| 196 | 8.31 | 7.37 | 88.7 |
| 197 | 1.31 | 0.97 | 74.0 |
| 198 | 1.60 | 1.20 | 75.0 |
| 199 | 23.3 | 17.3 | 74.3 |
| 200 | 3.92 | 2.96 | 75.5 |
| 201 | 3.43 | 2.53 | 73.8 |
| 202 | 4.59 | 3.51 | 76.5 |
| 203 | 18.6 | 14.4 | 77.6 |
| 205 | 1.41 | 1.03 | 73.0 |
| 206 | 6.22 | 4.66 | 74.9 |
| 208 | 1.12 | 0.88 | 78.6 |
| 209 | 0.58 | 0.58 | 99.1 |
| Total PCBs | 4540 | 3475 | 76.5 |
|  |  |  |  |
|  |  |  |  |

* The reported values are from SG-1 analysis only. Low recoveries for these congeners result from carry-over into SG-2 during silica gel fractionation. Historic CERC data shows that carry-over of these congeners into SG-2 (using the materials and methods described in this report) represents less than $5 \%$ of total PBCs.

Table III

GC-ECD Congener Specific Analysis for PCBs
Column Selection by Congener
Congener Analytical Congener Analytical Congener Analytical Congener Analytical Identification Column Identification Column Identification Column Identification Column

| 001 | DB-17 | 048 | DB-17 | 109 | DB-5 | 170 | DB-5\&17 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 003 | DB-17 | 049 | DB-5 | 110 | DB-5 | 171 | DB-17 |
| 004 | DB-17 | 051 | DB-17 | 112 | DB-5 | 172 | DB-5 |
| 005 | DB-17 | 052 | DB-5 | 113 | DB-5 | 173 | DB-5 |
| 006 | DB-17 | 053 | DB-5 | 114 | DB-17 | 174 | DB-17 |
| 007 | DB-17 | 054 | DB-17 | 115 | DB-17 | 175 | DB-5 |
| 008 | DB-17 | 055 | DB-17 | 117 | DB-5 | 176 | DB-17 |
| 009 | DB-17 | 056,060 | DB-5 | 118 | DB-5 | 177 | DB-5 |
| 010 | DB-17 | 057 | DB-5 | 119 | DB-17 | 178 | DB-5 |
| 011 | DB-17 | 058 | DB-5 | 122 | DB-17 | 179 | DB-5 |
| 015 | DB-17 | 063 | DB-17 | 123 | DB-5 | 180 | DB-5 |
| 016 | DB-5\&17 | 064 | DB-17 | 128 | DB-5 | 183 | DB-17 |
| 017 | DB-5\&17 | 066 | DB-17 | 129 | DB-5 | 185 | DB-17 |
| 018 | DB-5 | 067 | DB-17 | 130 | DB-17 | 187 | DB-5 |
| 019 | DB-5 | 069 | DB-17 | 131 | DB-17 | 189 | DB-5 |
| 020 | DB-17 | 070 | DB-5 | 132 | DB-17 | 190 | DB-5\&17 |
| 022 | DB-5 | 071 | DB-5 | 133 | DB-5 | 191 | DB-5 |
| 024 | DB-5\&17 | 072 | DB-5 | 134 | DB-5 | 193 | DB-5 |
| 025 | DB-5 | 074 | DB-5 | 136 | DB-5 | 194 | DB-5 |
| 026 | DB-17 | 075 | DB-5 | 137 | DB-5 | 195 | DB-17 |
| 027 | DB-17 | 082 | DB-5 | 138 | DB-17 | 196 | DB-5 |
| 028 | DB-5\&17 | 083 | DB-17 | 139 | DB-17 | 197 | DB-5 |
| 031 | DB-17 | 084 | DB-5 | 141 | DB-5 | 198 | DB-5 |
| 032 | DB-17 | 085 | DB-5 | 144 | DB-17 | 199 | DB-5 |
| 033 | DB-17 | 086 | DB-5 | 146 | DB-5 | 200 | DB-5 |
| 034 | DB-5 | 087 | DB-17 | 147 | DB-5 | 201 | DB-5 |
| 035 | DB-5 | 090 | DB-17 | 149 | DB-5 | 202 | DB-17 |
| 037,059 | DB-5 | 091 | DB-5 | 151 | DB-5 | 203 | DB-17 |
| 040 | DB-17 | 092 | DB-5 | 153 | DB-5\&17 | 205 | DB-5 |
| 041 | DB-17 | 095 | DB-5\&17 | 156 | DB-5\&17 | 206 | DB-17 |
| 042 | DB-17 | 096 | DB-5 | 157 | DB-5 | 208 | DB-17 |
| 043 | DB-5 | 097 | DB-5 | 158 | DB-5 | 209 | DB-5\&17 |
| 044 | DB-5 | 099 | DB-17 | 163 | DB-17 |  |  |
| 045 | DB-5 | 101 | DB-17 | 164 | DB-17 |  |  |
| 046 | DB-5 | 102 | DB-5 | 166 | DB-5 |  |  |
| 047 | DB-5\&17 | 105 | DB-17 | 167 | DB-5 |  |  |

## Table IV

## Congener Specific Analyses Results for Site 9-BST021.26 SPMDs

(Corrected for Background)*

| Congener <br> Identification | ng/SPMD | Congener <br> Identification | ng/SPMD |
| :---: | :---: | :---: | :---: |
| 001 | $<$ MDL | 048 |  |
| 003 | 0.37 | 049 | $<$ MQL |
| 004 | 0.61 | 051 | 0.77 |
| 005 | $<$ MDL | 052 | $<$ MQL |
| 006 | $<$ MQL | 053 | 3.76 |
| 007 | $<$ MQL | 054 | $<$ MQL |
| 008 | $<$ MQL | 055 | 0.12 |
| 009 | $<$ MQL | 056,060 | 2.82 |
| 010 | 0.11 | 057 | 0.29 |
| 011 | $<$ MDL | 058 | $<$ MDL |
| 015 | 2.20 | 063 | 0.21 |
| 016 | 3.30 | 064 | 3.78 |
| 017 | 3.45 | 066 | 3.73 |
| 018 | 5.02 | 067 | 0.33 |
| 019 | $<$ MQL | 069 | 0.96 |
| 020 | 0.25 | 070 | $<$ MQL |
| 022 | 1.93 | 071 | 2.48 |
| 024 | 0.09 | 072 | 0.38 |
| 025 | 0.35 | 074 | 1.66 |
| 026 | 1.73 | 075 | $<$ MQL |
| 027 | 1.46 | 082 | $<$ MQL |
| 028 | 5.87 | 083 | $<$ MQL |
| 031 | 2.95 | 084 | $<$ MQL |
| 032 | 6.16 | 085 | $<$ MDL |
| 033 | 1.48 | 086 | 0.21 |
| 034 | 0.31 | 087 | $<$ MQL |
| 035 | 3.69 | 090 | $<$ MQL |
| 037,059 | 1.06 | 091 | $<$ MQL |
| 040 | 1.19 | 092 | $<$ MQL |
| 041 | 0.54 | 095 | $<$ MQL |
| 042 | 2.51 | 096 | $<$ MDL |
| 043 | 5.95 | 097 | $<$ MQL |
| 044 | 2.47 | 109 | $<$ MQL |
| 045 | 1.30 | $<M Q 5$ | $<$ MQL |
| 046 | 2.77 | $<$ MQL |  |
| 047 |  |  |  |

Table IV (continued)
Congener Specific Analyses Results for Site 9-BST021.26 SPMDs
(Corrected for Background)*

| Congener Identification | ng/SPMD | Congener Identification | ng/SPMD |
| :---: | :---: | :---: | :---: |
| 109 | <MQL | 170 | 1.05 |
| 110 | <MQL | 171 | 0.26 |
| 112 | <MQL | 172 | 0.14 |
| 113 | 0.10 | 173 | <MQL |
| 114 | 0.16 | 174 | 1.27 |
| 115 | <MQL | 175 | 0.05 |
| 117 | 0.23 | 176 | 0.06 |
| 118 | <MQL | 177 | 0.62 |
| 119 | 0.13 | 178 | 0.26 |
| 122 | 0.04 | 179 | 0.69 |
| 123 | 0.15 | 180 | 1.52 |
| 128 | 0.88 | 183 | 0.76 |
| 129 | <MQL | 185 | 0.12 |
| 130 | 0.31 | 187 | 1.53 |
| 131 | <MQL | 189 | <MQL |
| 132 | <MQL | 190 | <MQL |
| 133 | 0.08 | 191 | 0.05 |
| 134 | 0.41 | 193 | 0.10 |
| 136 | <MQL | 194 | 0.20 |
| 137 | 0.28 | 195 | 0.13 |
| 138 | 3.36 | 196 | 0.16 |
| 139 | 1.08 | 197 | <MDL |
| 141 | 0.99 | 198 | <MQL |
| 144 | <MQL | 199 | 0.36 |
| 146 | 0.85 | 200 | <MQL |
| 147 | <MQL | 201 | <MQL |
| 149 | 5.05 | 202 | 0.11 |
| 151 | 1.83 | 203 | 0.27 |
| 153 | 4.90 | 205 | <MQL |
| 156 | 0.18 | 206 | <MQL |
| 157 | <MQL | 208 | <MDL |
| 158 | <MQL | 209 | <MDL |
| 163 | 1.04 | Total PCBs** | 180 |
| 164 | 0.52 |  |  |
| 166 | 0.08 | * (raw analysis results - background)/3 = ng/SPMD |  |
| 167 | 0.13 | ** Total PCBs not calculated as s congeners | quantifiable |

Table V
Congener Specific Analyses Results for Site 9-BFK000.02 SPMDs
(Corrected for Background)*

| Congener Identification | ng/SPMD | Congener Identification | ng/SPMD |
| :---: | :---: | :---: | :---: |
| 001 | <MDL | 048 | 2.32 |
| 003 | <MQL | 049 | <MQL |
| 004 | 1.10 | 051 | 0.70 |
| 005 | <MDL | 052 | <MQL |
| 006 | <MQL | 053 | 3.68 |
| 007 | <MQL | 054 | <MQL |
| 008 | 3.80 | 055 | 0.15 |
| 009 | <MQL | 056,060 | 4.66 |
| 010 | 0.78 | 057 | $<\mathrm{MQL}$ |
| 011 | <MDL | 058 | <MDL |
| 015 | 2.64 | 063 | 0.23 |
| 016 | 4.65 | 064 | 4.89 |
| 017 | 8.23 | 066 | 6.03 |
| 018 | 9.52 | 067 | 0.25 |
| 019 | <MQL | 069 | 1.11 |
| 020 | 0.38 | 070 | 4.79 |
| 022 | 3.99 | 071 | 3.14 |
| 024 | <MDL | 072 | <MDL |
| 025 | 0.82 | 074 | 3.12 |
| 026 | 2.02 | 075 | 0.48 |
| 027 | 1.83 | 082 | <MQL |
| 028 | 18.2 | 083 | <MQL |
| 031 | 7.51 | 084 | <MDL |
| 032 | 9.42 | 085 | <MDL |
| 033 | 1.88 | 086 | <MQL |
| 034 | 0.07 | 087 | <MQL |
| 035 | 1.15 | 090 | <MQL |
| 037,059 | 0.93 | 091 | <MDL |
| 040 | 1.33 | 092 | <MQL |
| 041 | 0.96 | 095 | <MDL |
| 042 | 3.06 | 096 | <MQL |
| 043 | 1.73 | 097 | <MQL |
| 044 | <MQL | 099 | <MQL |
| 045 | 2.06 | 101 | <MQL |
| 046 | 0.81 | 102 | <MDL |
| 047 | 3.03 | 105 | <MQL |

Table V (continued)
Congener Specific Analyses Results for Site 9-BFK000.02 SPMDs
(Corrected for Background)*

| Congener Identification | ng/SPMD | Congener Identification | /S |
| :---: | :---: | :---: | :---: |
| 109 | <MQL | 170 | <MQL |
| 110 | <MQL | 171 | <MQL |
| 112 | <MQL | 172 | $<\mathrm{MQL}$ |
| 113 | <MQL | 173 | <MDL |
| 114 | <MQL | 174 | <MQL |
| 115 | <MQL | 175 | <MQL |
| 117 | <MQL | 176 | <MDL |
| 118 | <MQL | 177 | <MQL |
| 119 | <MQL | 178 | <MQL |
| 122 | <MQL | 179 | <MQL |
| 123 | <MQL | 180 | <MQL |
| 128 | <MQL | 183 | <MQL |
| 129 | <MQL | 185 | <MQL |
| 130 | <MQL | 187 | <MQL |
| 131 | <MQL | 189 | <MDL |
| 132 | <MQL | 190 | <MQL |
| 133 | <MQL | 191 | <MQL |
| 134 | <MQL | 193 | <MQL |
| 136 | <MDL | 194 | <MQL |
| 137 | <MDL | 195 | <MQL |
| 138 | <MQL | 196 | <MQL |
| 139 | 0.17 | 197 | <MDL |
| 141 | <MQL | 198 | <MDL |
| 144 | <MQL | 199 | <MQL |
| 146 | <MQL | 200 | <MDL |
| 147 | <MDL | 201 | <MDL |
| 149 | <MQL | 202 | <MQL |
| 151 | <MQL | 203 | <MQL |
| 153 | <MQL | 205 | <MDL |
| 156 | <MQL | 206 | <MQL |
| 157 | <MQL | 208 | <MDL |
| 158 | <MQL | 209 | <MDL |
| 163 | <MQL | Total PCBs** | 167 |
| 164 | <MQL |  |  |
| 166 | <MQL | * (raw analysis results - background)/3 = ng/SPMD |  |
| 167 | <MQL | ** Total PCBs not calculated as s congeners | quantifiable |

## Table VI

Congener Specific Analyses Results for Site VA0025054-001 SPMDs
(Corrected for Background)*

| Congener Identification | ng/SPMD | Congener Identification | /SPMD |
| :---: | :---: | :---: | :---: |
| 001 | 3.04 | 048 | <MQL |
| 003 | 6.09 | 049 | <MQL |
| 004 | 7.72 | 051 | <MDL |
| 005 | <MDL | 052 | $<\mathrm{MDL}$ |
| 006 | 3.53 | 053 | <MQL |
| 007 | 0.50 | 054 | <MQL |
| 008 | <MQL | 055 | $<\mathrm{MQL}$ |
| 009 | 0.43 | 056,060 | 0.94 |
| 010 | 0.09 | 057 | <MDL |
| 011 | $<\mathrm{MDL}$ | 058 | $<\mathrm{MDL}$ |
| 015 | 3.23 | 063 | <MQL |
| 016 | 0.99 | 064 | <MQL |
| 017 | 4.24 | 066 | 1.04 |
| 018 | 6.57 | 067 | 0.17 |
| 019 | <MQL | 069 | 0.38 |
| 020 | 0.17 | 070 | <MQL |
| 022 | 1.12 | 071 | <MQL |
| 024 | <MDL | 072 | <MDL |
| 025 | $<\mathrm{MDL}$ | 074 | 1.06 |
| 026 | 6.43 | 075 | 0.34 |
| 027 | 0.68 | 082 | 0.88 |
| 028 | 2.45 | 083 | <MDL |
| 031 | 2.43 | 084 | <MQL |
| 032 | 1.55 | 085 | $<$ MDL |
| 033 | 1.80 | 086 | $<$ MDL |
| 034 | 1.04 | 087 | $<\mathrm{MDL}$ |
| 035 | 5.41 | 090 | <MQL |
| 037,059 | 0.76 | 091 | <MDL |
| 040 | <MQL | 092 | <MQL |
| 041 | $<\mathrm{MDL}$ | 095 | $<$ MDL |
| 042 | 1.07 | 096 | <MDL |
| 043 | 0.38 | 097 | $<$ MQL |
| 044 | <MQL | 099 | $<$ MDL |
| 045 | 0.39 | 101 | $<\mathrm{MDL}$ |
| 046 | 0.59 | 102 | <MQL |
| 047 | 3.29 | 105 | <MQL |

## Table VI (continued)

Congener Specific Analyses Results for Site VA0025054-001 SPMDs
(Corrected for Background)*

| Congener |  | Congener |  |
| :---: | :---: | :---: | :---: |
| Identification | ng/SPMD | Identification | ng/SPMD |
| 109 | <MQL | 170 | $<\mathrm{MDL}$ |
| 110 | <MQL | 171 | <MQL |
| 112 | <MQL | 172 | <MDL |
| 113 | 0.10 | 173 | <MQL |
| 114 | <MQL | 174 | <MQL |
| 115 | $<$ MDL | 175 | <MQL |
| 117 | $<\mathrm{MDL}$ | 176 | <MDL |
| 118 | <MQL | 177 | <MQL |
| 119 | 0.17 | 178 | <MQL |
| 122 | <MQL | 179 | <MQL |
| 123 | 0.39 | 180 | <MDL |
| 128 | <MQL | 183 | 0.68 |
| 129 | <MQL | 185 | <MQL |
| 130 | <MQL | 187 | <MQL |
| 131 | <MQL | 189 | <MQL |
| 132 | $<\mathrm{MDL}$ | 190 | <MDL |
| 133 | $<$ MDL | 191 | 0.11 |
| 134 | <MDL | 193 | <MQL |
| 136 | <MQL | 194 | <MQL |
| 137 | <MQL | 195 | $<\mathrm{MDL}$ |
| 138 | <MDL | 196 | $<\mathrm{MDL}$ |
| 139 | 0.21 | 197 | <MDL |
| 141 | <MQL | 198 | 0.04 |
| 144 | <MQL | 199 | <MDL |
| 146 | $<\mathrm{MQL}$ | 200 | <MQL |
| 147 | <MDL | 201 | <MQL |
| 149 | $<\mathrm{MQL}$ | 202 | <MQL |
| 151 | <MQL | 203 | $<$ MDL |
| 153 | $<\mathrm{MQL}$ | 205 | $<$ MDL |
| 156 | $<\mathrm{MDL}$ | 206 | $<\mathrm{MDL}$ |
| 157 | <MQL | 208 | <MDL |
| 158 | $<\mathrm{MQL}$ | 209 | $<\mathrm{MDL}$ |
| 163 | $<\mathrm{MDL}$ | Total PCBs** | 89 |
| 164 | $<\mathrm{MDL}$ |  |  |
| 166 | 0.04 | * (raw analysis results - backgrou | = ng/SPMD |
| 167 | <MQL | ** Total PCBs not calculated as s congeners | quantifiable |

Table VII
Congener Specific Analyses Results for Site 9-BPB000.02 SPMDs
(Corrected for Background)*

| Congener <br> Identification | ng/SPMD | Congener <br> Identification | ng/SPMD |
| :---: | :---: | :---: | :---: |
| 001 | $<$ MDL | 048 |  |
| 003 | $<$ MDL | 049 | 3.44 |
| 004 | 8.16 | 051 | 32.7 |
| 005 | $<$ MDL | 052 | 3.09 |
| 006 | 2.08 | 053 | 36.5 |
| 007 | 0.85 | 054 | 18.0 |
| 008 | 1.97 | 055 | $<$ MDL |
| 009 | 0.69 | 056,060 | 0.39 |
| 010 | 1.39 | 057 | 12.1 |
| 011 | $<$ MDL | 058 | 0.49 |
| 015 | 5.94 | 063 | 0.11 |
| 016 | 22.3 | 064 | 0.92 |
| 017 | 5.19 | 066 | 19.3 |
| 018 | 8.43 | 067 | 15.7 |
| 019 | 1.20 | 069 | 1.26 |
| 020 | 0.70 | 070 | 2.69 |
| 022 | 4.94 | 071 | 7.51 |
| 024 | 0.37 | 072 | 11.7 |
| 025 | 0.66 | 074 | 0.45 |
| 026 | 2.70 | 075 | 5.69 |
| 027 | 8.14 | 082 | 1.77 |
| 028 | 10.7 | 083 | 1.58 |
| 031 | 5.50 | 084 | 0.16 |
| 032 | 34.6 | 085 | $<$ MQL |
| 033 | 2.26 | 086 | $<$ MDL |
| 034 | 0.40 | 087 | 0.36 |
| 035 | 10.4 | 090 | $<$ MQL |
| 037,059 | 5.13 | 091 | 0.22 |
| 040 | 5.46 | 092 | $<$ MQL |
| 041 | 1.98 | 095 | $<$ MQL |
| 042 | 11.8 | 096 | $<$ MQL |
| 043 | 5.77 | 097 | 20.70 |
| 044 | 40.3 | 099 | $<$ MQL |
| 045 | 12.8 | 101 | $<$ MQL |
| 046 | 5.43 | 102 | 0.59 |
| 047 | 7.03 |  | 1.57 |
|  |  |  |  |

## Table VII (continued)

Congener Specific Analyses Results for Site 9-BPB000.02 SPMDs
(Corrected for Background)*

| Congener Identification | ng/SPMD | Congener Identification | /S |
| :---: | :---: | :---: | :---: |
| 109 | 0.46 | 170 | <MQL |
| 110 | 9.13 | 171 | 0.14 |
| 112 | 0.23 | 172 | 0.07 |
| 113 | 0.18 | 173 | <MQL |
| 114 | 0.16 | 174 | 0.72 |
| 115 | 0.35 | 175 | 0.05 |
| 117 | 0.45 | 176 | <MQL |
| 118 | <MQL | 177 | 0.34 |
| 119 | 0.29 | 178 | 0.15 |
| 122 | 0.05 | 179 | 0.48 |
| 123 | <MQL | 180 | 0.93 |
| 128 | <MQL | 183 | 0.53 |
| 129 | <MQL | 185 | 0.08 |
| 130 | <MQL | 187 | 0.97 |
| 131 | <MQL | 189 | <MQL |
| 132 | <MQL | 190 | <MQL |
| 133 | <MQL | 191 | 0.07 |
| 134 | <MQL | 193 | 0.05 |
| 136 | <MQL | 194 | 0.11 |
| 137 | <MQL | 195 | <MQL |
| 138 | <MQL | 196 | <MQL |
| 139 | 0.65 | 197 | <MDL |
| 141 | <MQL | 198 | <MQL |
| 144 | <MQL | 199 | 0.22 |
| 146 | 0.56 | 200 | <MQL |
| 147 | <MQL | 201 | <MQL |
| 149 | <MQL | 202 | 0.08 |
| 151 | 1.26 | 203 | 0.19 |
| 153 | <MQL | 205 | 0.04 |
| 156 | <MQL | 206 | $<\mathrm{MQL}$ |
| 157 | 0.13 | 208 | <MDL |
| 158 | <MQL | 209 | <MDL |
| 163 | 0.71 | Total PCBs** | 505 |
| 164 | 0.41 |  |  |
| 166 | 0.11 | * (raw analysis results - background)/3 = ng/SPMD |  |
| 167 | 0.12 | ** Total PCBs not calculated as s congeners | quantifiable |

Table VIII
Congener Specific Analyses Results for Site 9-WVC000.05 SPMDs
(Corrected for Background)*

| Congener Identification | ng/SPMD | Congener Identification | ng/SPMD |
| :---: | :---: | :---: | :---: |
| 001 | <MDL | 048 | 1.03 |
| 003 | <MDL | 049 | <MQL |
| 004 | 1.19 | 051 | 0.86 |
| 005 | <MDL | 052 | <MDL |
| 006 | <MQL | 053 | 2.07 |
| 007 | <MQL | 054 | <MQL |
| 008 | <MQL | 055 | 0.08 |
| 009 | <MDL | 056,060 | 1.35 |
| 010 | 0.14 | 057 | 0.22 |
| 011 | <MDL | 058 | <MDL |
| 015 | 1.03 | 063 | 0.17 |
| 016 | 1.29 | 064 | 2.23 |
| 017 | 3.04 | 066 | 2.45 |
| 018 | 3.53 | 067 | 0.26 |
| 019 | <MQL | 069 | 1.46 |
| 020 | 0.19 | 070 | <MQL |
| 022 | 1.82 | 071 | 1.73 |
| 024 | <MQL | 072 | <MQL |
| 025 | 0.20 | 074 | 0.97 |
| 026 | 1.17 | 075 | 0.63 |
| 027 | 0.47 | 082 | <MQL |
| 028 | 4.72 | 083 | <MQL |
| 031 | 2.34 | 084 | <MQL |
| 032 | 2.05 | 085 | <MDL |
| 033 | 1.27 | 086 | <MQL |
| 034 | 0.09 | 087 | <MQL |
| 035 | 3.02 | 090 | 0.23 |
| 037,059 | 0.50 | 091 | <MQL |
| 040 | 0.75 | 092 | <MQL |
| 041 | 0.50 | 095 | <MQL |
| 042 | 1.53 | 096 | <MDL |
| 043 | 2.69 | 097 | <MQL |
| 044 | <MQL | 099 | <MQL |
| 045 | 1.09 | 101 | <MQL |
| 046 | 0.57 | 102 | <MQL |
| 047 | 3.17 | 105 | <MQL |

Table VIII (continued)
Congener Specific Analyses Results for Site 9-WVC000.05 SPMDs
(Corrected for Background)*

| Congener Identification | ng/SPMD | Congener Identification | ng/SPMD |
| :---: | :---: | :---: | :---: |
| 109 | <MQL | 170 | 2.27 |
| 110 | <MQL | 171 | 0.63 |
| 112 | <MQL | 172 | 0.41 |
| 113 | <MQL | 173 | 0.05 |
| 114 | 0.17 | 174 | 3.83 |
| 115 | <MQL | 175 | 0.13 |
| 117 | 0.22 | 176 | 0.29 |
| 118 | <MQL | 177 | 1.73 |
| 119 | 0.14 | 178 | 0.95 |
| 122 | <MQL | 179 | 2.33 |
| 123 | 0.09 | 180 | 4.22 |
| 128 | 0.73 | 183 | 1.70 |
| 129 | <MQL | 185 | 0.38 |
| 130 | 0.29 | 187 | 4.62 |
| 131 | <MQL | 189 | 0.04 |
| 132 | 2.50 | 190 | 0.36 |
| 133 | 0.09 | 191 | 0.07 |
| 134 | 0.53 | 193 | 0.32 |
| 136 | 2.18 | 194 | 0.49 |
| 137 | <MQL | 195 | 0.27 |
| 138 | 4.73 | 196 | 0.34 |
| 139 | 0.62 | 197 | <MQL |
| 141 | 2.12 | 198 | 0.05 |
| 144 | 0.61 | 199 | 0.84 |
| 146 | 1.59 | 200 | <MQL |
| 147 | 0.08 | 201 | 0.10 |
| 149 | 10.75 | 202 | 0.19 |
| 151 | 5.36 | 203 | 0.55 |
| 153 | 9.97 | 205 | 0.05 |
| 156 | <MQL | 206 | $<\mathrm{MQL}$ |
| 157 | <MQL | 208 | $<$ MDL |
| 158 | 0.60 | 209 | <MDL |
| 163 | 2.33 | Total PCBs** | 170 |
| 164 | 0.55 |  |  |
| 166 | <MQL | * (raw analysis results - background)/3 = ng/SPMD |  |
| 167 | 0.12 | ** Total PCBs not calculated as s congeners | quantifiable |

Table IX
Congener Specific Analyses Results for Site 9-BST028.90 SPMDs
(Corrected for Background)*

| Congener Identification | ng/SPMD | Congener Identification | ng/SPMD |
| :---: | :---: | :---: | :---: |
| 001 | <MDL | 048 | <MQL |
| 003 | 0.31 | 049 | <MDL |
| 004 | 1.17 | 051 | $<\mathrm{MQL}$ |
| 005 | <MDL | 052 | <MDL |
| 006 | $<\mathrm{MQL}$ | 053 | $<$ MDL |
| 007 | <MDL | 054 | $<$ MDL |
| 008 | <MQL | 055 | <MDL |
| 009 | $<\mathrm{MDL}$ | 056,060 | $<\mathrm{MQL}$ |
| 010 | 0.26 | 057 | <MDL |
| 011 | <MDL | 058 | <MDL |
| 015 | <MQL | 063 | <MQL |
| 016 | <MQL | 064 | <MDL |
| 017 | 1.93 | 066 | <MQL |
| 018 | <MQL | 067 | 0.13 |
| 019 | <MDL | 069 | 2.09 |
| 020 | <MQL | 070 | <MDL |
| 022 | <MQL | 071 | <MQL |
| 024 | <MDL | 072 | 2.57 |
| 025 | <MDL | 074 | <MQL |
| 026 | 1.42 | 075 | 0.26 |
| 027 | <MQL | 082 | <MQL |
| 028 | <MQL | 083 | <MDL |
| 031 | <MQL | 084 | $<$ MDL |
| 032 | <MQL | 085 | $<\mathrm{MDL}$ |
| 033 | <MQL | 086 | <MDL |
| 034 | 0.08 | 087 | <MQL |
| 035 | 3.02 | 090 | <MQL |
| 037,059 | 0.15 | 091 | <MDL |
| 040 | <MQL | 092 | <MQL |
| 041 | <MQL | 095 | <MDL |
| 042 | <MQL | 096 | <MDL |
| 043 | 4.29 | 097 | <MDL |
| 044 | <MDL | 099 | <MQL |
| 045 | <MQL | 101 | <MQL |
| 046 | 0.19 | 102 | <MDL |
| 047 | 0.94 | 105 | <MQL |

Table IX (continued)
Congener Specific Analyses Results for Site 9-BST028.90 SPMDs
(Corrected for Background)*

| Congener Identification | ng/SPMD | Congener Identification | ng/SPMD |
| :---: | :---: | :---: | :---: |
| 109 | <MQL | 170 | $<\mathrm{MQL}$ |
| 110 | <MQL | 171 | $<\mathrm{MQL}$ |
| 112 | <MDL | 172 | <MQL |
| 113 | <MDL | 173 | <MDL |
| 114 | <MQL | 174 | $<\mathrm{MQL}$ |
| 115 | <MQL | 175 | <MQL |
| 117 | <MDL | 176 | <MQL |
| 118 | <MQL | 177 | <MQL |
| 119 | <MQL | 178 | <MQL |
| 122 | <MQL | 179 | <MQL |
| 123 | <MQL | 180 | <MQL |
| 128 | <MQL | 183 | <MQL |
| 129 | <MQL | 185 | <MQL |
| 130 | <MQL | 187 | <MQL |
| 131 | <MDL | 189 | $<\mathrm{MDL}$ |
| 132 | <MQL | 190 | <MDL |
| 133 | <MQL | 191 | <MQL |
| 134 | <MQL | 193 | <MQL |
| 136 | <MQL | 194 | <MQL |
| 137 | <MQL | 195 | <MQL |
| 138 | <MQL | 196 | <MQL |
| 139 | 0.12 | 197 | <MDL |
| 141 | <MQL | 198 | <MDL |
| 144 | <MDL | 199 | $<\mathrm{MQL}$ |
| 146 | <MQL | 200 | $<$ MDL |
| 147 | <MDL | 201 | <MDL |
| 149 | <MQL | 202 | <MQL |
| 151 | <MQL | 203 | <MQL |
| 153 | <MQL | 205 | <MQL |
| 156 | <MDL | 206 | <MQL |
| 157 | <MQL | 208 | <MDL |
| 158 | <MQL | 209 | <MDL |
| 163 | <MQL | Total PCBs** | 32 |
| 164 | $<\mathrm{MQL}$ |  |  |
| 166 | <MQL | * (raw analysis results - background)/3 = ng/SPMD |  |
| 167 | <MQL | ** Total PCBs not calculated as s congeners | quantifiable |

## Table X

## Recovery of Performance Reference Compounds and Water Concentration

Mean Recovery From Fabrication Blanks, Field Blanks, and Procedural Blanks ( $\mathrm{n}=8$ )

## PRC Compounds

Acenaphthylene- $d_{10}$
Acenaphthene- $d_{10}$ Fluorene- $d_{10}$
Phenanthrene- $d_{10}$
Pyrene- $d_{10}$
$\mu \mathrm{g} /$ SPMD by GC-FID
$6.18 \pm 1.07$
$6.55 \pm 1.01$
$7.18 * \pm 0.98$
$11.1 \pm 1.18$
$11.2 \pm 1.17$

Mean Recovery From Deployment Sites (n=4)

## PRC Compounds

Acenaphthylene- $d_{10}$
$\mu \mathrm{g} / \mathrm{SPMD}$ by GC-FID

| Acenaphthylene $-d_{10}$ | $0.74 \pm 0.25$ |
| :---: | :---: |
| Acenaphthene- $d_{10}$ | $1.66 \pm 0.72$ |
| Fluorene $-d_{10}$ | $2.65 * * \pm 1.03$ |
| Phenanthrene $-d_{10}$ | $10.3^{* * *} \pm 3.95$ |
| Pyrene $-d_{10}$ | $11.0^{* * *} \pm 0.85$ |

$0.74 \pm 0.25$
Fluorene- $d_{10}$
2.65 ** $\pm 1.03$

Pyrene- $d_{10}$
$11.0^{* * *} \pm 0.85$

Estimation of Water Concentration Using PRC Measurements

$$
\text { PRC Compound } \quad \mathbf{k}_{\text {eprc }}\left(\mathbf{d}^{-1}\right)
$$

Acenaphthylene- $d_{10}$
0.0708

Acenaphthene- $d_{10}$
Fluorene- $d_{10}$
Phenanthrene- $d_{10}$
0.0458
0.0332 ****
$\mathrm{k}_{\text {eprc }}=\frac{\ln \left(\text { CSPMDo } / C_{\text {SPMD }}\right)}{t}$

Pyrene- $d_{10}$
0.0025
0.0006

[^0]*** High recovery indicates no photodegradation
**** Value used in calculation of water concentrations

Table XI
SPMD Derived Water Concentrations at Site 9-BST021.26

| Congener Identification | pg / L | Congener Identification | pg / L | Congener Identification | pg / L |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | UtD | 77 | UtD | 134 | 3.0 |
| 18 | 8.8 | 78 | UtD | 136 | UtD |
| 19 | UtD | 79 | UtD | 137 | 2.8 |
| 22 | 12 | 81 | UtD | 138 | 24 |
| 25 | 2.1 | 82 | UtD | 141 | 7.1 |
| 26 | 11 | 83 | UtD | 146 | 6.2 |
| 28 | 24 | 84 | UtD | 149 | 31 |
| 31 | 15 | 85 | UtD | 151 | 12 |
| 40 | 6.2 | 87 | UtD | 153 | 53 |
| 41 | 3.1 | 90 | UtD | 156 | 2.5 |
| 42 | 14 | 91 | UtD | 157 | UtD |
| 43 | 33 | 92 | UtD | 158 | UtD |
| 44 | UtD | 95 | UtD | 169 | UtD |
| 45 | 11 | 97 | UtD | 172 | 3.9 |
| 46 | 10 | 99 | UtD | 174 | 14 |
| 47 | 13 | 101 | UtD | 176 | 1.0 |
| 48 | 13 | 105 | UtD | 178 | 3.0 |
| 49 | UtD | 107 | 1.9 | 179 | 11 |
| 51 | 5.6 | 110 | UtD | 180 | 20 |
| 52 | UtD | 114 | 1.3 | 183 | 8.6 |
| 53 | 27 | 118 | UtD | 187 | 1.2 |
| 63 | 1.4 | 119 | 1.0 | 194 | 5.4 |
| 64 | 18 | 126 | 0.62 | 199 | 6.9 |
| 66 | 24 | 127 | 3.2 | 201 | UtD |
| 67 | 2.2 | 128 | 6.9 | 207 | UtD |
| 70 | UtD | 129 | UtD | Total PCB* | 1300 |
| 74 | 9.3 | 130 | 2.7 |  |  |

UtD $=$ Unable to Derive (i.e. individual congeners not present at > MQL)

* Total PCB value calculated directly from Total PCB (ng/SPMD) and not as sum of those individual congeners present at > MQL.

Table XII
SPMD Derived Water Concentrations at Site 9-BFK000.02

| Congener <br> Identification | $\mathrm{pg} / \mathrm{L}$ |
| :---: | :--- | :---: | :---: | :---: | :--- | | Congener |
| :--- |
| Identification |$\quad \mathrm{pg} / \mathrm{L}$| Congener |
| :--- |
| Identification |$\quad \mathrm{pg} / \mathrm{L}$

UtD $=$ Unable to Derive (i.e. individual congeners not present at > MQL)

* Total PCB value calculated directly from Total PCB (ng/SPMD) and not as sum of those individual congeners present at > MQL.

Table XIII
SPMD Derived Water Concentrations at Site VA0025054-001

| Congener Identification | pg / L | Congener Identification | pg / L | Congener Identification | pg / L |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | 9.6 | 77 | UtD | 134 | UtD |
| 18 | 12 | 78 | UtD | 136 | UtD |
| 19 | UtD | 79 | UtD | 137 | UtD |
| 22 | 6.8 | 81 | UtD | 138 | UtD |
| 25 | UtD | 82 | 6.9 | 141 | UtD |
| 26 | 39 | 83 | UtD | 146 | UtD |
| 28 | 10 | 84 | UtD | 149 | UtD |
| 31 | 12 | 85 | UtD | 151 | UtD |
| 40 | UtD | 87 | UtD | 153 | UtD |
| 41 | UtD | 90 | UtD | 156 | UtD |
| 42 | 6.0 | 91 | UtD | 157 | UtD |
| 43 | 2.1 | 92 | UtD | 158 | UtD |
| 44 | UtD | 95 | UtD | 169 | UtD |
| 45 | 1.7 | 97 | UtD | 172 | UtD |
| 46 | 4.7 | 99 | UtD | 174 | UtD |
| 47 | 15 | 101 | UtD | 176 | UtD |
| 48 | UtD | 105 | UtD | 178 | UtD |
| 49 | UtD | 107 | UtD | 179 | UtD |
| 51 | UtD | 110 | UtD | 180 | UtD |
| 52 | UtD | 114 | UtD | 183 | 7.6 |
| 53 | UtD | 118 | UtD | 187 | UtD |
| 63 | UtD | 119 | 1.3 | 194 | UtD |
| 64 | UtD | 126 | 0.17 | 199 | UtD |
| 66 | 6.9 | 127 | 8.5 | 201 | UtD |
| 67 | 1.1 | 128 | UtD | 207 | UtD |
| 70 | UtD | 129 | UtD | Total PCB* | 640 |
| 74 | 6.0 | 130 | UtD |  |  |
| UtD = Unable to Derive (i.e. individual congeners not present at > MQL) |  |  |  |  |  |

Table XIV
SPMD Derived Water Concentrations at Site 9-BPB000.02

| Congener Identification | pg / L | Congener Identification | pg / L | Congener Identification | pg / L |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | 5.7 | 77 | UtD | 134 | UtD |
| 18 | 15 | 78 | UtD | 136 | UtD |
| 19 | 7.9 | 79 | UtD | 137 | UtD |
| 22 | 30 | 81 | UtD | 138 | UtD |
| 25 | 4.0 | 82 | 12 | 141 | UtD |
| 26 | 16 | 83 | 1.2 | 146 | 4.1 |
| 28 | 44 | 84 | UtD | 149 | UtD |
| 31 | 27 | 85 | UtD | 151 | 8.2 |
| 40 | 29 | 87 | UtD | 153 | UtD |
| 41 | 11 | 90 | 1.2 | 156 | UtD |
| 42 | 66 | 91 | UtD | 157 | 1.7 |
| 43 | 32 | 92 | UtD | 158 | UtD |
| 44 | 190 | 95 | UtD | 169 | UtD |
| 45 | 56 | 97 | 30 | 172 | 2.0 |
| 46 | 43 | 99 | UtD | 174 | 8.1 |
| 47 | 33 | 101 | UtD | 176 | UtD |
| 48 | 34 | 105 | 14 | 178 | 1.6 |
| 49 | 210 | 107 | 3.0 | 179 | 7.6 |
| 51 | 22 | 110 | 56 | 180 | 12 |
| 52 | 200 | 114 | 1.3 | 183 | 5.9 |
| 53 | 130 | 118 | UtD | 187 | 0.76 |
| 63 | 6.0 | 119 | 2.3 | 194 | 3.0 |
| 64 | 89 | 126 | 0.83 | 199 | 4.3 |
| 66 | 100 | 127 | 0.92 | 201 | UtD |
| 67 | 8.3 | 128 | UtD | 207 | UtD |
| 70 | 37 | 129 | UtD | Total PCB* | 3700 |
| 74 | 32 | 130 | UtD |  |  |

UtD = Unable to Derive (i.e. individual congeners not present at > MQL)

* Total PCB value calculated directly from Total PCB (ng/SPMD) and not as sum of those individual congeners present at > MQL.

Table XV
SPMD Derived Water Concentrations at Site 9-WVC000.05

| Congener <br> Identification | $\mathrm{pg} / \mathrm{L}$ | Congener <br> Identification | $\mathrm{pg} / \mathrm{L}$ |
| :---: | :--- | :---: | :--- | :---: | :--- | | Congener |
| :---: |
| Identification |$\quad \mathrm{pg} / \mathrm{L}$

UtD $=$ Unable to Derive (i.e. individual congeners not present at > MQL)

* Total PCB value calculated directly from Total PCB (ng/SPMD) and not as sum of those individual congeners present at > MQL.

Table XVI
SPMD Derived Water Concentrations at Site 9-BST028.90

| Congener Identification | pg / L | Congener Identification | pg / L | Congener Identification | pg / L |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | UtD | 77 | UtD | 134 | UtD |
| 18 | UtD | 78 | UtD | 136 | UtD |
| 19 | UtD | 79 | UtD | 137 | UtD |
| 22 | UtD | 81 | UtD | 138 | UtD |
| 25 | 0.00 | 82 | UtD | 141 | UtD |
| 26 | 8.6 | 83 | UtD | 146 | UtD |
| 28 | UtD | 84 | UtD | 149 | UtD |
| 31 | UtD | 85 | UtD | 151 | UtD |
| 40 | UtD | 87 | UtD | 153 | UtD |
| 41 | UtD | 90 | UtD | 156 | UtD |
| 42 | UtD | 91 | UtD | 157 | UtD |
| 43 | 24 | 92 | UtD | 158 | UtD |
| 44 | UtD | 95 | UtD | 169 | UtD |
| 45 | UtD | 97 | UtD | 172 | UtD |
| 46 | 1.5 | 99 | UtD | 174 | UtD |
| 47 | 4.4 | 101 | UtD | 176 | UtD |
| 48 | UtD | 105 | UtD | 178 | UtD |
| 49 | UtD | 107 | UtD | 179 | UtD |
| 51 | UtD | 110 | UtD | 180 | UtD |
| 52 | UtD | 114 | UtD | 183 | UtD |
| 53 | UtD | 118 | UtD | 187 | UtD |
| 63 | UtD | 119 | UtD | 194 | UtD |
| 64 | UtD | 126 | UtD | 199 | UtD |
| 66 | UtD | 127 | UtD | 201 | UtD |
| 67 | 0.83 | 128 | UtD | 207 | UtD |
| 70 | UtD | 129 | UtD | Total PCB* | 230 |
| 74 | UtD | 130 | UtD |  |  |

UtD $=$ Unable to Derive (i.e. individual congeners not present at > MQL)

* Total PCB value calculated directly from Total PCB (ng/SPMD) and not as sum of those individual congeners present at > MQL.


## Table XVII

Lower Limits of PCB Water Concentration Estimation* For Sampling Conditions In the Bluestone River Virginia During the April 2004 SPMD Deployment
Congener Estimated Congener Estimated Congener Estimated Congener Estimated Identification $\mathrm{pg} / \mathrm{L}$ Identification $\mathrm{pg} / \mathrm{L}$ Identification $\mathrm{pg} / \mathrm{L}$ Identification $\mathrm{pg} / \mathrm{L}$

| 4 | 0.67 | 51 | 53 | 112 | 57 | 172 | 3.3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 1.5 | 52 | 0.70 | 113 | 0.23 | 173 | 0.60 |
| 6 | 0.10 | 53 | 2.6 | 114 | 0.50 | 174 | 0.37 |
| 7 | 1.9 | 54 | 11 | 115 | 0.43 | 175 | 5.3 |
| 8 | 0.93 | 55 | 0.77 | 117 | 1.6 | 176 | 0.53 |
| 9 | 5.0 | 56 | 0.23 | 118 | 1.5 | 177 | 0.60 |
| 10 | 0.80 | 57 | 4.0 | 119 | 37 | 178 | 1.6 |
| 11 | 0.10 | 58 | 0.87 | 122 | 0.90 | 179 | 1.2 |
| 15 | 0.10 | 63 | 0.20 | 123 | 0.13 | 180 | 3.3 |
| 16 | 2.2 | 64 | 0.33 | 128 | 0.60 | 183 | 7.0 |
| 17 | 1.3 | 66 | 14 | 129 | 6.0 | 185 | 3.2 |
| 18 | 1.4 | 67 | 6.3 | 130 | 1.8 | 187 | 0.33 |
| 19 | 12 | 69 | 0.33 | 131 | 1.9 | 189 | 5.0 |
| 20 | 6.0 | 70 | 0.37 | 132 | 1.3 | 190 | 0.43 |
| 22 | 0.33 | 71 | 23 | 133 | 17 | 191 | 2.13 |
| 24 | 1.5 | 72 | 4.3 | 134 | 0.53 | 193 | 0.43 |
| 25 | 0.20 | 74 | 1.1 | 136 | 2.6 | 194 | 0.87 |
| 26 | 0.40 | 75 | 4.3 | 137 | 17 | 195 | 1.6 |
| 27 | 1.4 | 82 | 1.3 | 138 | 1.4 | 196 | 2.3 |
| 28 | 0.47 | 83 | 5.7 | 139 | 25 | 197 | 2.4 |
| 31 | 5.0 | 84 | 1.3 | 141 | 0.53 | 198 | 2.3 |
| 32 | 7.7 | 85 | 40 | 144 | 6.3 | 199 | 0.70 |
| 33 | 2.1 | 86 | 0.23 | 146 | 2.30 | 201 | 8.0 |
| 34 | 4.0 | 87 | 0.60 | 147 | 4.33 | 202 | 1.2 |
| 35 | 0.10 | 90 | 21 | 149 | 0.43 | 203 | 1.3 |
| 37 | 1.1 | 91 | 1.2 | 151 | 29 | 205 | 2.5 |
| 40 | 0.73 | 92 | 16 | 153 | 11 | 206 | 0.70 |
| 41 | 1.4 | 95 | 13 | 156 | 50 | 208 | 2.9 |
| 42 | 0.60 | 96 | 15 | 157 | 5.3 | Total PCBs | 40 |
| 43 | 3.7 | 97 | 67 | 158 | 1.1 |  |  |
| 44 | 0.60 | 99 | 23 | 163 | 4.3 |  |  |
| 45 | 47 | 101 | 18 | 164 | 5.0 |  |  |
| 46 | 2.9 | 102 | 73 | 166 | 2.4 |  |  |
| 47 | 0.83 | 105 | 3.7 | 167 | 0.27 |  |  |
| 48 | 7.0 | 109 | 10 | 170 | 0.43 |  |  |
| 49 | 3.0 | 110 | 2.4 | 171 | 13 |  |  |

[^1]Figure 1
GC-FID Analysis of PRC Standard


NOTE: This chromatogram is representative of the analytical method for determinations of PRCs (Table X) with peak identifications as follows:

PRC Compounds Retention times (minutes)

| Acenaphthylene- $d_{10}$ | 15.71 |
| :---: | :---: |
| Acenaphthene- $d_{10}$ | 17.00 |
| Fluorene- $d_{10}$ | 20.87 |
| Phenanthrene- $d_{10}$ | 28.17 |
| Pyrene- $d_{10}$ | 39.16 |
| 4 -Terphenyl- $d_{14}$ as IIS | 41.87 |

## Figure 2

## Representative GC-FID Analysis of SG-2 Fraction of SPMD Field Blank



SPMD Field Blank of Site 9-BPB000.02

NOTE: This chromatogram is representative of the determinations of "CspmDDo" values from analyses of SPMD blanks (Table X)

Figure 3
Representative GC-FID Analysis of SG-2 Fraction of SPMD Deployment


Deployed SPMD from Site 9-BPB000.02
NOTE: This chromatogram is representative of the determinations of "CSPMD" values from analyses of SPMD blanks (Table X)

Figure 4
GC-ECD Analyses of PCB Standard

$4000 \mathrm{ng} / \mathrm{mL}$ Total PCB Standard on DB-17

$4000 \mathrm{ng} / \mathrm{mL}$ Total PCB Standard on DB-5

Figure 5
Representative GC-ECD Analysis of SG-1 Fraction of SPMD Field Blank


SPMD Field Blank of Site 9-BPB000.02 on DB-17


SPMD Field Blank of Site 9-BPB000.02 on DB-5

NOTE: Duel column analysis required (Table III) for complete resolution and quantification of individual congeners.

Figure 6
Representative GC-ECD Analysis of SG-1 Fraction of SPMD Deployment


Deployed SPMD from Site 9-BPB000.02 on DB-17


Deployed SPMD from Site 9-BPB000.02 on DB-5

Figure 7
GC-ECD Profile of Aroclor® 1242 on DB-5


NOTE: $1000 \mathrm{ng} / \mathrm{mL}$ Standard

Figure 8
GC-ECD Profile of Aroclor® 1248 on DB-5


NOTE: 1000 ng/mL Standard

Figure 9
GC-ECD Profile of Aroclor® 1254 on DB-5


NOTE: $1000 \mathrm{ng} / \mathrm{mL}$ Standard

Figure 10
GC-ECD Profile of Aroclor® 1260 on DB-5


NOTE: $1000 \mathrm{ng} / \mathrm{mL}$ Standard

Figure 11
GC-ECD Profile of Site 9-BST021.26 on DB-5


NOTE: 3-SPMD composite injected at 1.0 mL in hexane

Figure 12
GC-ECD Profile of Site 9-BFK000.02 on DB-5


NOTE: 3-SPMD composite injected at 1.0 mL in hexane

Figure 13
GC-ECD Profile of Site VA0025054-001 on DB-5


NOTE: 3-SPMD composite injected at 1.0 mL in hexane

Figure 14
GC-ECD Profile of Site 9-BPB000.02 on DB-5


NOTE: 3-SPMD composite injected at 1.0 mL in hexane

Figure 15
GC-ECD Profile of Site 9-WVC000.05 on DB-5


NOTE: 3-SPMD composite injected at 1.0 mL in hexane

Figure 16
GC-ECD Profile of Site 9-BST028.90 on DB-5


NOTE: 3-SPMD composite injected at 1.0 mL in hexane


[^0]:    * CspmDDo
    ** $\mathrm{C}_{\text {SPMD }}$

[^1]:    * NOTE: MQL values applied to Excel calculator for water concentration estimation.

