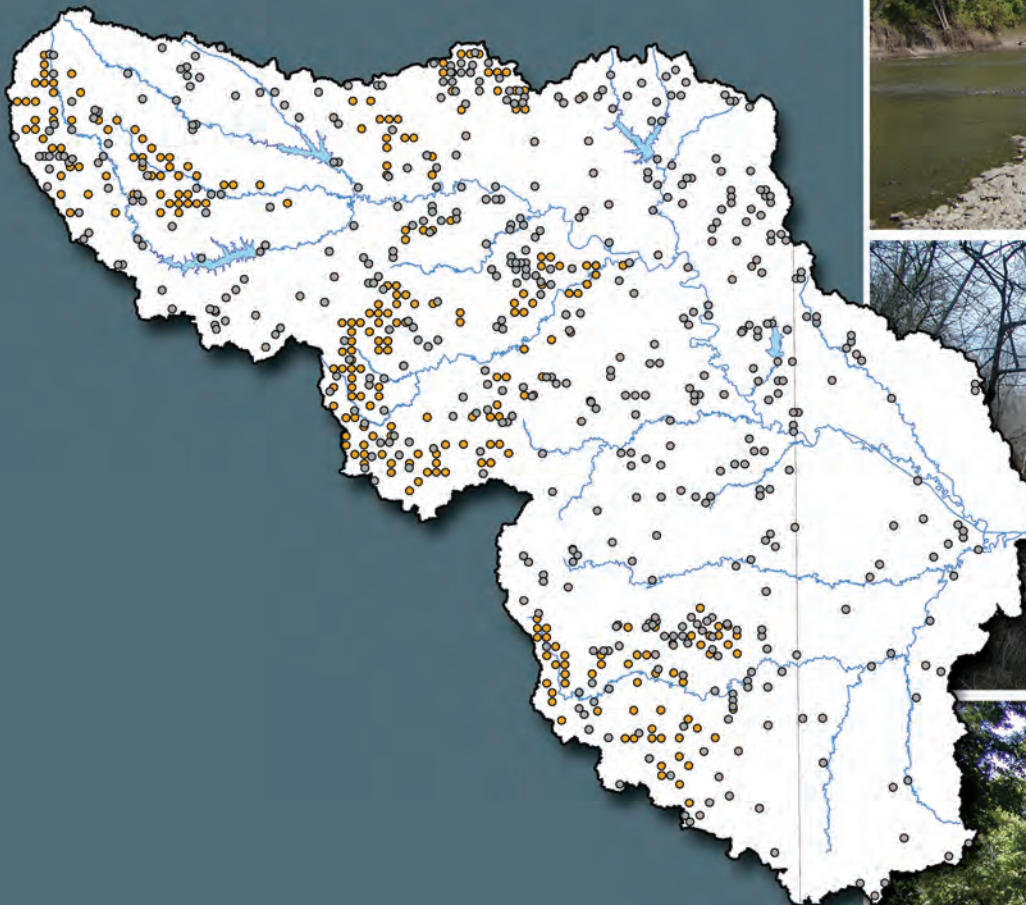


Prepared in cooperation with the Missouri Department of Conservation

Effects of Impoundments and Land-Cover Changes on Streamflows and Selected Fish Habitat in the Upper Osage River Basin, Missouri and Kansas



Scientific Investigations Report 2007–5175

Cover Photographs. Marais des Cygnes River near the Missouri-Kansas state line, Missouri, July 2003 (top). Little Osage River near the junction with the Marmaton River, Missouri, February 2000 (middle). Marmaton River near the Missouri-Kansas state line, Missouri, September 2004 (bottom). Photographs by David C. Heimann, U.S. Geological Survey.

Effects of Impoundments and Land-Cover Changes on Streamflows and Selected Fish Habitat in the Upper Osage River Basin, Missouri and Kansas

By David C. Heimann, Susan S. Licher, and Gregg K. Schalk

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Scientific Investigations Report 2007–5175

U.S. Department of the Interior
U.S. Geological Survey

U.S. Department of the Interior
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Suggested citation:
Heimann, D.C., Licher, S.S., and Schalk, G.K., 2007, Effects of impoundments and land-cover changes on streamflows and selected fish habitat in the upper Osage River Basin, Missouri and Kansas: U.S. Geological Survey Scientific Investigations Report 2007-5175, 96 p. plus CD.

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Conversion Factors

Multiply	By	To obtain
Length		
inch (in)	.3937	centimeters (cm)
foot (ft)	0.3048	meter (m)
mile (mi)	1.609	kilometer (km)
Area		
acre	4,047	square meter (m ²)
acre	0.4047	hectare (ha)
acre	0.004047	square kilometer (km ²)
square foot (ft ²)	929.0	square centimeter (cm ²)
square foot (ft ²)	0.09290	square meter (m ²)
square mile (mi ²)	259.0	hectare (ha)
square mile (mi ²)	2.590	square kilometer (km ²)
Volume		
gallon (gal)	3.785	liter (L)
gallon (gal)	0.003785	cubic meter (m ³)
million gallons (Mgal)	3,785	cubic meter (m ³)
cubic foot (ft ³)	28.32	cubic decimeter (dm ³)
cubic foot (ft ³)	0.02832	cubic meter (m ³)
acre-foot (acre-ft)	1,233	cubic meter (m ³)
acre-foot (acre-ft)	0.001233	cubic hectometer (hm ³)
Flow rate		
acre-foot per day (acre-ft/d)	0.01427	cubic meter per second (m ³ /s)
acre-foot per year (acre-ft/yr)	1,233	cubic meter per year (m ³ /yr)
cubic foot per second (ft ³ /s)	0.02832	cubic meter per second (m ³ /s)
cubic foot per day (ft ³ /d)	0.02832	cubic meter per day (m ³ /d)
gallon per day (gal/d)	0.003785	cubic meter per day (m ³ /d)
million gallons per day (Mgal/d)	0.04381	cubic meter per second (m ³ /s)

Temperature in degrees Fahrenheit (°F) may be converted to degrees Celsius (°C) as follows:

$$^{\circ}\text{C}=(^{\circ}\text{F}-32)/1.8$$

Vertical coordinate information is referenced to the North American Vertical Datum of 1988 (NAVD 88).

Horizontal coordinate information is referenced to the North American Datum of 1983 (NAD 83).

Water year is defined as October 1 through September 30.

GLOSSARY OF ACRONYMS AND ABBREVIATIONS

Hydrologic Simulation Program-FORTRAN parameters and terms (Bicknell and others, 2004)

AGWETP— Fraction of evapotranspiration from active ground water
AGWRC—Base groundwater recession
BASETP— Fraction of evapotranspiration from baseflow
CEPSC— Interception storage capacity
DEEPPFR—Fraction of groundwater inflow transferred to deep recharge
FTABLES—Depth-area, depth-volume, and depth-outflow relation table developed for each model RCHRES
IMPLND—Impervious land area
INFILT—Index to infiltration capacity
INTFW— Interflow inflow parameter
IRC—Interflow recession parameter
KVARY—Variable ground-water recession parameter
LSUR—Length of overland flow path
LZETP—Lower zone evapotranspiration parameter
LZSN—Lower zone nominal soil moisture storage
NSUR—Manning's n (roughness coefficient) for overland flow
PERLND—Pervious land area
RCHRES—HSPF stream channel reach or reservoir
SLSUR—Slope of overland flow plane
UZSN—Upper zone nominal soil moisture storage

Software programs used in development of hydrologic simulations

BASINS—Better assessment science integrating point and non-point sources; used in development of HSPF and stormwater assessment tool (SWAT) models (U.S. Environmental Protection Agency, 2005a).
CGAP—Channel geometry analysis program used to develop depth-area, depth-volume, and depth-outflow tables from stream channel information (U.S. Geological Survey, 2005d)
GenScn—Program used for the generation and analysis of model simulation scenarios (U.S. Geological Survey, 2005d).
GenFTABLE—Program used to convert CGAP output into FTABLES for use in HSPF (U.S. Geological Survey, 2005d).
HSPF—Hydrologic Simulation Program-FORTRAN (Bicknell and others, 2004).
HSPFEXP—Expert system for calibration of HSPF (U.S. Geological Survey, 2005d).
WdmUtil—Utility program used to import hydrologic and meteorologic time series data into watershed data management (wdm) file format (U.S. Environmental Protection Agency, 2005b).
WinHSPF—Microsoft Windows version of HSPF (packaged with BASINS and available at U.S. Environmental Protection Agency, 2005a).

Hydrologic Simulation Program-FORTRAN simulation scenarios developed in study

Pre+infiltr—The simulated daily or hourly pre-settlement streamflow time series using an INFILTR parameter double that used in the Current scenario.

Pre+infiltr_{adj}—The adjusted (based on the ratio of observed to simulated values) daily or hourly pre-settlement streamflow time series using an INFILTR parameter double that used in the Current scenario.

Current—The simulated daily or hourly streamflow time series using 1995 land cover and existing (2005) impoundments.

Observed—The observed daily or hourly streamflow time series that is only available at those reporting locations corresponding to a U.S. Geological Survey streamflow gaging station.

Prop-sel50—Simulated proposed streamflow time series using selected proposed impoundments and median values as estimates for any missing impoundment design criteria.

Prop-sel10—Simulated proposed streamflow time series using selected proposed impoundments and 10th-percentile values as estimates for any missing impoundment design criteria.

Prop-sel90—Simulated proposed streamflow time series using selected proposed impoundments and 90th-percentile values as estimates for any missing impoundment design criteria.

Prop-sel50_{adj}—Adjusted (based on the ratio of observed to simulated values) proposed streamflow time series for selected proposed impoundments using median estimates for any missing impoundment design criteria.

Prop-sel10_{adj}—Adjusted (based on the ratio of observed to simulated values) proposed streamflow time series for selected proposed impoundments using 10th-percentile estimates for any missing impoundment design criteria.

Prop-sel90_{adj}—Adjusted (based on the ratio of observed to simulated values) proposed streamflow time series for selected proposed impoundments using 90th-percentile estimates for any missing impoundment design criteria.

Prop-all50—Simulated proposed streamflow time series using all proposed impoundments and median estimates for any missing impoundment design criteria.

Prop-all10—Simulated proposed streamflow time series using all proposed impoundments and 10th-percentile estimates for any missing impoundment design criteria.

Prop-all90—Simulated proposed streamflow time series using all proposed impoundments and 90th-percentile estimates for any missing impoundment design criteria.

Prop-all50_{adj}—Adjusted (based on the ratio of observed to simulated values) proposed streamflow time series using all proposed impoundments and median estimates for any missing impoundment design criteria.

Prop-all10_{adj}—Adjusted (based on the ratio of observed to simulated values) proposed streamflow time series using all proposed impoundments and 10th-percentile values as estimates for any missing impoundment design criteria.

Prop-all90_{adj}—Adjusted (based on the ratio of observed to simulated values) proposed streamflow time series using all proposed impoundments and 90th-percentile values as estimates for any missing impoundment design criteria.

Effects of Impoundments and Land-Cover Changes on Streamflows and Selected Fish Habitat in the Upper Osage River Basin, Missouri and Kansas

By David C. Heimann, Susan S. Licher, and Gregg K. Schalk

Abstract

A study was conducted by the U.S. Geological Survey in cooperation with the Missouri Department of Conservation to estimate the effects of existing and proposed impoundments, land-cover changes, and reported water uses on streamflows in the 5,410-square mile upper Osage River Basin. The hydrologic model Hydrologic Simulation Program-FORTRAN (HSPF) was calibrated and validated to current (1995–2004 water years) regulation and water-use conditions, and scenarios were developed to evaluate differences for the same 10-years of record under pre-settlement, and proposed impoundment conditions. Analyses included quantification of changes in the magnitude, frequency, timing, and duration of streamflows under each simulation scenario. Streamflows from the simulations were used in conjunction with known streamflow-fish habitat relations to quantify effects of altered flows on fish-habitat area at selected Marais des Cygnes and Marmaton River locations.

The cumulative effects of impoundments and land-cover changes were determined to substantially alter streamflows in the upper Osage River Basin model simulations spanning pre-settlement to proposed future conditions. The degree of streamflow alteration varied between major subbasins. Streamflows in the Marais des Cygnes River Basin were altered between pre-settlement and current conditions, primarily by major impoundments, with smaller changes expected with proposed regulation. Streamflows in the Little Osage River Basin were relatively unchanged between pre-settlement and current conditions with land-cover changes (primarily the conversion of native prairies to cultivated land) affecting flows more than the few current impoundments in this basin. The current peak flows in the Marmaton River Basin generally were higher than pre-settlement or proposed scenario peak flows. Of the three major subbasins, the Marmaton River Basin is likely to be the most affected by proposed impoundments.

Declines in monthly minimum streamflows under a proposed impoundment scenario at the Marais des Cygnes River near the Kansas-Missouri state line, Kansas, were greatest for the lowest 10 percent of corresponding observed flows

and during the driest years (2000, 2001 water years); that is, the greatest percent declines in flows under proposed conditions generally occurred during the lowest current/observed flow periods. In a small headwater basin in the Marmaton River Basin, simulated declines in minimum flows were small (generally less than 6 cubic feet per second and less than 1 cubic foot per second for 1- and 3-day scenarios), but resulted in 10 to 18 additional zero flow days for the 10-year simulation for the proposed scenarios relative to current simulated conditions. Reductions in minimum monthly flows as a result of additional impoundments generally were less than 5 cubic feet per second at the Marmaton River near Marmaton, Kansas, and resulted in 6 additional zero flow days. The greatest declines between proposed and current flows at the Marmaton River near the Kansas-Missouri state line, Missouri, generally occurred in the lower 50 percentile of the distribution of current simulated flows and during the drier simulation years (2001–2003). Proposed conditions resulted in declines in the 0-10 percentile flow values for the 1-, 3-, and 7-day durations. July, August, and October had the largest declines in proposed low flows relative to current simulated low flows for the 10-year simulation at this site.

The flood frequency for the Marais des Cygnes River near the Kansas-Missouri state line was unchanged between observed and proposed conditions for the 10-year simulation, but was 450 percent greater under the pre-settlement scenarios compared to observed conditions. Flood frequency generally was greatest for the current condition scenarios in the Marmaton River Basin and least for the proposed conditions, although the effects of regulation on flood frequency decreased downstream from the Kansas-Missouri state line with substantial streamflow contributions from less regulated tributaries. The flood frequencies of the proposed scenarios were 54 to 60 percent less than current conditions at the Marmaton River near the Kansas-Missouri state line, whereas at a downstream location the flood frequencies under proposed conditions were 39 to 45 percent less than current conditions.

Minimum annual 1-, 7-, or 14-day selected fish habitat availability under a proposed scenario declined at three Marais des Cygnes River reporting sites by more than 10 percent

2 Effects of Impoundments and Land-Cover Changes on Streamflows and Fish Habitat

compared with observed conditions, for one or more years, for each of nine seasonal fish habitat categories. Declines in minimum habitat availability under proposed conditions were at or near 100 percent for one or more years for summer flathead catfish, fall flathead catfish, fall channel catfish, and fall stonecat habitat categories at one Marais des Cygnes River location, and for summer flathead catfish, summer channel catfish, and fall flathead catfish at another Marais des Cygnes River location. Declines in annual 1-, 7-, or 14-day minimum habitat also were greater than 10 percent for one or more years for all categories at both Marmaton River reporting locations except for spring paddlefish habitat, which generally remained unchanged between current and proposed scenarios at one location. Declines in 1-, 7-, or 14-day proposed minimum habitat availability were at or near 100 percent for one or more years for slenderhead darter, summer flathead catfish, fall channel catfish, and fall stonecat habitat categories at one Marmaton River location and for spring suckermouth minnow, spring slenderhead darter, summer channel catfish, summer stonecat, and fall flathead catfish habitat at another.

Simulations of the Marmaton River Basin (1995 through 2004) indicated that the effects of a conversion of cultivated row crops back to pre-settlement native prairie soils, simulated using an increase in the infiltration model parameter, accounted for a greater difference in total runoff between pre-settlement and current/proposed scenarios than other changes in land cover or from impoundments. The simulated increase in soil infiltration capacity under native prairie conditions also resulted in lower peak flows for the pre-settlement model scenario compared with the current or proposed scenarios. Evaporative water losses increased with the addition of impoundments, and while these increases did not have a substantial effect on the total runoff from the basin they could account for simulated declines in low flows. The greater detention associated with proposed impoundments resulted in longer hydrograph recessions and lowered peak flows; this varied with modifications in simulated impoundment design.

Introduction

Artificial impoundments (ponds, lakes, reservoirs) can alter natural stream channel and riparian habitats by changing the timing and quantities of streamflow, stream/ground-water interactions, trapping of sediment, and the hydroperiods of riparian wetlands. Impoundments increase the storage and surface area of water at a local scale, but this may result in a net loss of water at a larger basin scale through increased evaporative losses. Another ecological consequence of the increase in localized water storage, wetted area, and sediment trapping may be an alteration in the maintenance and function of downstream riparian habitat and wetlands. Smith and others (2002) determined the frequency distribution of impoundments in the conterminous United States to be dominated by water bodies in the small (less than 25 acres) size class. Although the regu-

lated area and the detention capability of any small, individual, impoundment may be insignificant in comparison to larger water bodies, the sheer numbers of these features can result in substantial cumulative effects at a larger basin scale (Smith and others, 2002; Renwick and others, 2005).

The greatest concentration of impoundments in the conterminous United States is in agricultural areas, especially in the eastern part of the Great Plains (Smith and others, 2002) including the upper Osage River Basin in eastern Kansas and west-central Missouri. Flood-control and recreation impoundments are common features in this basin. These impoundments are designed and permitted individually; a single impoundment may have little effect on streamflows, but the cumulative effects of hundreds of current and proposed impoundments on the natural streamflow regime of this basin are unknown. Multiple watershed districts within the Marais des Cygnes, Little Osage, and Marmaton River Basin, which compose the upper Osage River Basin, have watershed plans that include the construction of an estimated 283 new impoundments (Kansas Water Office, 2004). These proposed impoundments, in conjunction with an estimated 539 existing permitted impoundments (U.S. Army Corps of Engineers, 2005; Joe File, Kansas Department of Agriculture, written commun., 2005) within multiple watershed Districts, and three large U.S. Army Corps of Engineers reservoirs, can further regulate flows in the basin. The proposed impoundments would range from 2 to 215 acres at the primary spillway water level, with most impoundments less than 30 acres.

Dry-weather flows in the upper Osage River Basin may be altered substantially by additional impoundments and changes in land cover. During runoff events following dry periods, when lake levels have decreased because of infiltration and evaporation, little runoff will reach the channel immediately downstream from the dam until the lake comes up to the spillway elevation. Streamflow demands for irrigation, wetland management, drinking-water supplies, and power generation have continued to increase, and these increases in river withdrawals and withdrawals from flood plain wells in Kansas and Missouri also are concerns for sustaining dry-weather flows in the upper Osage River Basin (Dent and others, 1997). Land-cover changes also may have affected dry-weather flows and the frequency of floods in the basin. Originally, more than 80 percent of the upper Osage River Basin was covered by tall grass prairie (Schroeder, 1983). Only a few small, scattered tracts of native prairie remain; most have been converted to agriculture. Fuentes and others (2004) determined soil hydraulic conductivities for cultivated lands to be about an order of magnitude less than those for native prairies in Washington State, and these differences persisted after 27 years of continuous no-till practices. Such alterations in soil hydraulic properties, when applied over extensive areas, also can alter the hydrologic characteristics of a basin.

The West Osage River Watershed Inventory and Management Plan (Dent and others, 1997) provides information about the current hydrology, land cover, water quality, aquatic habitat, and aquatic biota in the upper Osage River

Basin. Although most wetlands (Dahl, 1990) and much of the bottomland forests (Nelson, 1985) of Missouri have been removed for agriculture, numerous remnant wetlands and tracts of bottomland timber remain in the upper Osage River Basin, particularly in the Marmaton River Basin. Many of these areas are in public land holdings (Dent and others, 1997). Because streams and channelized rivers are deeply incised and alluvial deposits underlying wetlands consist of fine-grained material of low permeability, ground water generally is not a substantial source of water to riparian wetlands in the basin. Consequently, riparian wetlands in the upper Osage River Basin are recharged primarily through floods and precipitation (Heimann and Mettler-Cherry, 2004).

Stream channel and riparian habitats may be affected by changes in the natural streamflow regime of the upper Osage River Basin. Naturally-variable flood flows create and maintain habitats that are essential to aquatic and riparian species (Poff and others, 1997). Channel-forming flows that create and shape stream habitats and affect channel stability typically are floods that have a recurrence interval of 1–2.5 years (Leopold, 1994). Changes in frequency and duration of larger floods could substantially affect the hydroperiod (period, timing, and amount of water retention) of numerous remnant and managed wetlands. Native riparian wetland communities evolved under pre-development hydroperiods; therefore, changes to the frequency and duration of wetland inundation may jeopardize the survival of native communities and ongoing efforts to restore native riparian vegetation. Alterations in the magnitude, timing, or flow duration also can limit the operation and maintenance of publicly and privately managed riparian wetlands.

In addition to the ecological concerns of altered flows in the upper Osage River Basin, low-flow depletions may have undesirable effects on drinking water supplies as surface water is the primary source of drinking water for residents of the basin. In Missouri alone more than 25,000 people use more than 2.5 million gallons per day (Mgal/day) from surface-water sources (Missouri Department of Natural Resources, 2000) in the basin.

Objectives

A study was conducted by the U.S. Geological Survey (USGS) in cooperation with the Missouri Department of Conservation to estimate the effects of impoundments, land-cover changes, and point-source withdrawals and discharges on streamflows in the upper Osage River Basin. More information was needed to better understand the specific effects of flow regulation on riverine habitats and the quantity and timing of water supplies to aid managers and policy makers in determining and maintaining appropriate streamflow regimes for the upper Osage River Basin. The specific objectives of the study are as follows:

1. Simulate and predict the effects of proposed impoundments on low-flow sustainability and the

frequency and duration of flood events with calibrated and validated hydrologic models.

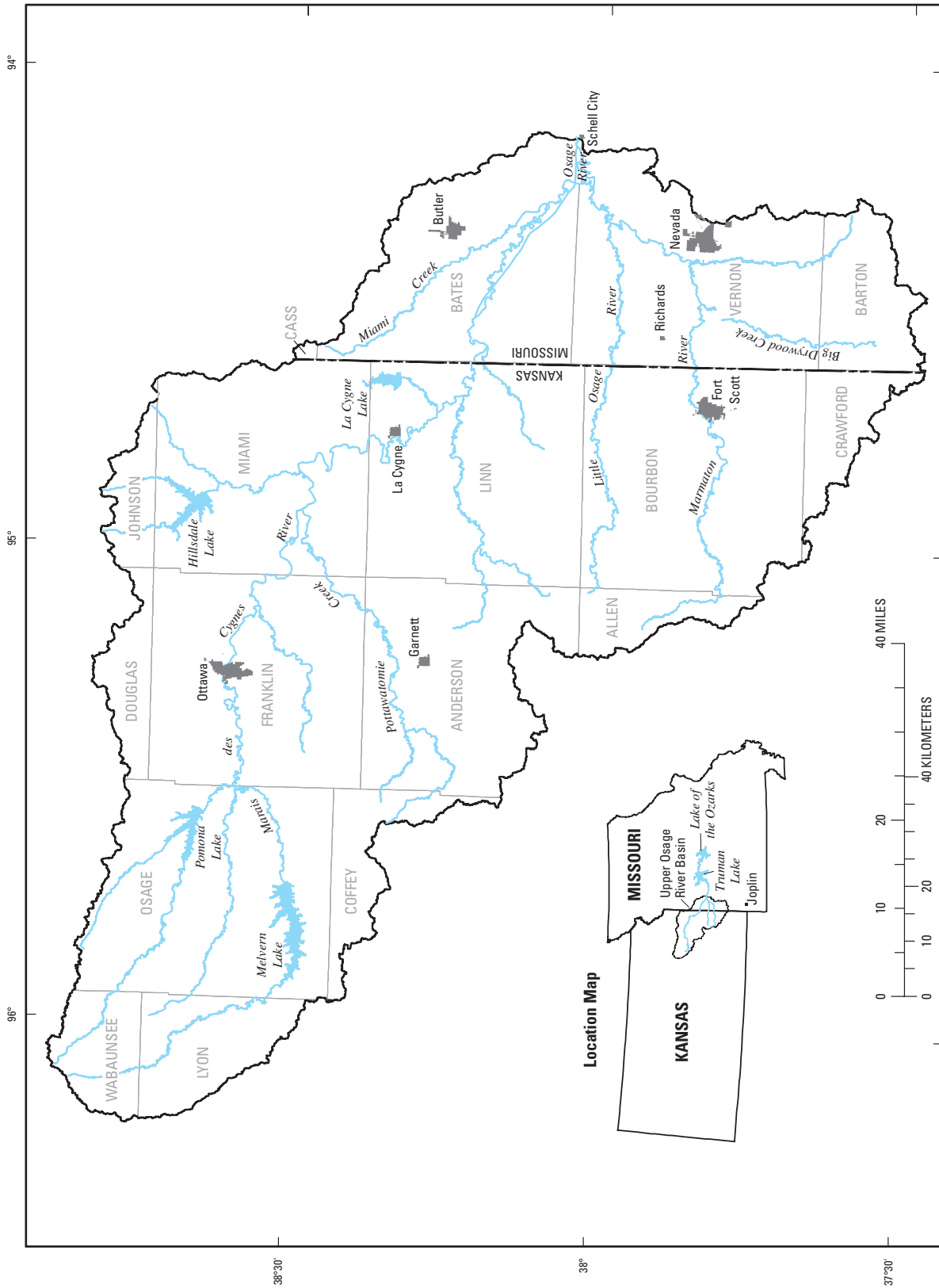
2. To the extent that land-cover and hydrologic characteristics of pre-settlement conditions were available, use the hydrologic models to qualitatively characterize the historical streamflow regime under which native plant and animal species adapted.
3. Utilize calibrated models to identify and evaluate selected water-management strategies that have the potential to minimize any adverse effects from upstream impoundments.
4. Determine the contributions and depletions of streamflow by municipal, industrial, recreational, and irrigation sources in the Marais des Cygnes, Little Osage, and Marmaton River Basins.

Purpose and Scope

The purpose of this report is to present the results of numerical hydrologic simulations developed to estimate the effects of impoundments, land-cover changes, and reported water uses on streamflows in the upper Osage River Basin. Hydrologic models were calibrated and validated to current (1995–2004 water years) regulation and water-use conditions. The models were then modified to simulate the relative differences in flow for the same period of climatological record for pre-settlement and proposed-regulation conditions. Simulated streamflow time series for pre-settlement, current, and proposed conditions were compared at selected locations on the Marais des Cygnes, Little Osage, and Marmaton Rivers, along with changes in streamflow as a result of differences in impoundment outflow design and reported point-source withdrawals and discharges. Analyses included quantification of changes in the magnitude, frequency, duration, and timing of streamflows for each scenario. Output from simulation scenarios were used in conjunction with known streamflow-fish habitat relations at select locations in the upper Osage River Basin to quantify the effects of altered flow on fish habitat.

Description of Study Area

The upper Osage River Basin encompasses 5,410 square miles (mi²) upstream from the Osage River at Schell City, Missouri, which represents the downstream extent of the study area (fig. 1). The basin lies in the Osage Plains physiographic region (Nelson, 1985) of Missouri and Osage Cuestas physiographic region of Kansas (Kansas Geological Survey, 1997). About 4,050 mi² of the basin lies in Kansas and 1,360 mi² in Missouri, and includes all or part of 13 Kansas and 4 Missouri counties. The largest population centers include Nevada, Missouri (fig. 1; population 8,607); Fort Scott, Kansas (population 8,297); and Butler, Missouri (population 4,209; U.S. Census Bureau, 2006).



Base from U.S. Geological Survey digital data, 1:24,000, 1989
 Albers Equal Area Conic Projection

Figure 1. Location of the upper Osage River Basin.

The 1949–2005 mean annual precipitation for the upper Osage River Basin was about 40 inches (in.); Butler, Missouri, (fig. 1) received an average of 40.2 in., and an average of 39.9 in. was received at Garnett, Kansas (fig. 1). June (5.38 in.) and May (4.90 in.) had the highest long-term average monthly rainfall at Butler, Missouri, and February (1.76 in.) had the lowest (High Plains Regional Climatic Center, 2007). The annual distribution of rainfall during the 10-year (1995 through 2004) study period fell within the central distribution of the 1949–2005 rainfall distribution at Butler, Missouri, whereas the 10-year distribution of precipitation at Garnett, Kansas, mirrored the 1949–2005 distribution at this site (fig. 2). The annual precipitation for the 10-years (1995 through 2004) of analysis at both locations included 4 years of above-average, 4 years of below-average, and 2 years of near-average precipitation (fig. 2).

The average annual 1949–2005 snowfall at Butler, Missouri, was 11.2 in. (equivalent to about 0.9 in. of rainfall). The 1949 through 2005 monthly mean temperatures at Butler, Missouri, ranged from 30.4 °F in January to 79.0 °F in July, with a mean annual temperature of 56.1 °F (High Plains Regional Climatic Center, 2007).

The upper Osage River Basin is comprised of the Marais des Cygnes, Little Osage, and Marmaton River subbasins (fig.1), which originate in the Osage plains in Kansas and flow east into Missouri. The Marais des Cygnes River Basin has a

drainage area of about 3,260 mi² upstream from the Missouri-Kansas state line (table 1), and is the only Osage River tributary that is regulated by U.S. Army Corps of Engineers reservoirs (Pomona, Melvern, and Hillsdale Lakes; fig. 1). The 1959–2005 mean annual flow for the Marais des Cygnes River at the Kansas-Missouri state line was 2,180 cubic feet per second (ft³/s) (U.S. Geological Survey, 2007). The current (2005) percentage of the area regulated in the Marais des Cygnes River Basin upstream from the Kansas-Missouri state line is about 34 percent, with about 27 percent attributable to the U.S. Army Corps of Engineers and LaCygne Lake impoundments, and the remainder attributable to smaller (each regulating less than 8 mi²) impoundments (table 1). This value is expected to increase to about 42 percent with additional proposed impoundments, many of which are located in the Pottawatomie Creek Basin (fig. 3), and all of which are located in the upper one-half of the Marais des Cygnes River Basin. The Little Osage River has a drainage area of about 363 mi² upstream from the Kansas-Missouri state line, a 1949–2005 mean annual flow of 232 ft³/s (U.S. Geological Survey, 2007), and is the least regulated tributary in the basin. Less than about 3 percent of the drainage area was regulated by impoundments in 2005, and no changes in the regulated area are proposed (table 1). The drainage area of the Marmaton River at the Kansas-Missouri state line is about 424 mi², and the current regulated area upstream from this point is about 25 percent.

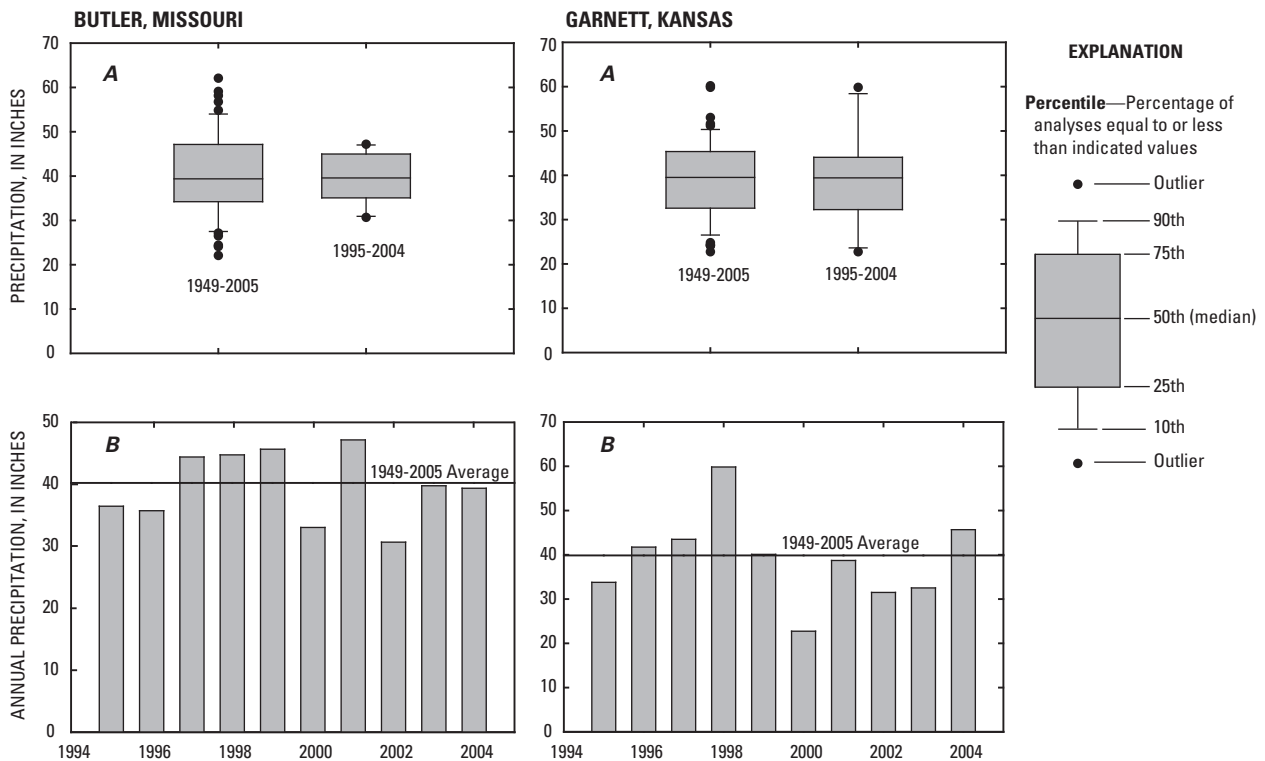


Figure 2. Comparisons of the (A) annual precipitation distribution for 1949 through 2005 to the 1995 through 2004 study period, at Butler, Missouri, and Garnett, Kansas, and (B) 1949 through 2005 mean precipitation to the annual 1995 through 2004 precipitation at Butler, Missouri, and Garnett, Kansas.

6 Effects of Impoundments and Land-Cover Changes on Streamflows and Fish Habitat

Table 1. Current and proposed regulated drainage area by upper Osage River subbasin.

Basin (fig. 1)	Total drainage area, in square miles	Current regulated drainage area, in square miles ^a	Additional proposed regulated drainage area, in square mile ^b	Cumulative total proposed regulated drainage area, in square miles	Current regulated drainage area, in percent	Additional proposed regulated drainage area, in percent	Total proposed regulated drainage area, in percent
Marais des Cygnes River at Kansas-Missouri State Line	3,260	1,107 (873)	250	1,357	34.0 (26.8)	7.70	41.6
Marais des Cygnes River Basin	3,790	1,175 (873)	250	1,366	29.5(23.0)	6.60	36.1
Little Osage River at Kansas-Missouri State Line	363	11.1	0	11.1	3.10	0	3.10
Little Osage River Basin (upstream of confluence with Marmaton River)	502	13.5	0	13.5	2.70	0	2.70
Marmaton River at Kansas- Missouri State Line	424	106	186	292	25.1	41.7	66.8
Marmaton River Basin	1,150	134	186	320	11.7	16.2	27.9
Osage River near Schell City, Missouri	5,410	1,264	436	1,700	23.3	8.10	31.4

^aValues in parenthesis represent only Pomona, Melvern, Hillsdale, and LaCygne Lakes (see figure 1 for locations).

^bProposed drainage area determined from data provided by the Kansas Department of Agriculture, Division of Water Resources.

This value is expected to increase to about 67 percent (table 1) if all proposed impoundments are completed. The Marmaton River near Marmaton, Kansas (drainage area of 295 mi²), had a 1972–2005 mean annual flow of 284 ft³/s (U.S. Geological Survey, 2007).

Historically, the Osage River Basin was dominated by tall grass prairies with narrow oak-hickory forests along stream channels (fig. 4; Schroeder, 1983). Settlement of the basin in Missouri primarily began in the 1830's, after statehood, along primary transportation routes in the basin including the Osage, Marais des Cygnes, and Marmaton Rivers (McDermott, 1940; U.S. Department of Agriculture, 1970). The population of the basin declined during the civil war but recovered between 1866 and 1885 with the introduction of the railroad and the discovery of coal in the region. The population in Missouri counties of the basin steadily declined from the early 1900's through the 1960's, but has since remained steady with a shift from rural to urban population centers (U.S. Department of Agriculture, 1970; Hall and Orazem, 2005). With the advent of the steel plow and the quelling of wildfires, most of the native tallgrass prairies in the basin have been replaced by agricultural land covers. The historical land cover of the basin is represented in vegetation maps of Kansas (Küchler, 1974) and Missouri (Schroeder, 1983). Major land-cover types identified included prairie, and forests (table 2, fig. 4). Although not specified on the Küchler map, there also was a substantial amount of wetlands in the basin. Wetlands in the basin have been reduced through channelization, levees, and conversion

to agriculture (Dent and others, 1997). The 1995 wetland land-cover information was used as an estimate for the historical wetland area for this study. The lower 44 miles of the Marais des Cygnes River was channelized in the early 1900's to decrease flooding and increase agricultural land area, resulting in the loss of about 10 mi of the original stream channel (Dent and others, 1997).

Land-cover information for the upper Osage River Basin was derived from the National Land Cover Database (USGS, 2005a) and based on 30-meter satellite thematic mapper data from 1992–1995. The 1995 land cover in the basin primarily was row crops (soybeans, wheat, corn, sorghum) and cool-season pasture (fig. 5; table 2). The trend in agricultural land cover in the Marmaton River Basin has been a shift from row crops to pasture (Bill Schoenberger, U.S. Department of Agriculture, Natural Resource Conservation Service, oral commun., 2005). The basin retains oak-hickory forests primarily along stream and river channels. Despite initial forest losses as a result of settlement, there was an 11 percent increase in forested area in the basin between 1959 and 1989 (Hahn and Spencer, 1991) because of wildfire suppression. The Marmaton River is considered one of the highest quality remaining prairie streams in Missouri, and its riparian corridor contains numerous remnant forested and constructed wetlands managed by public and private entities (Dent and others, 1997).

Surface bedrock in the upper Osage River Basin consists primarily of Pennsylvanian shales and sandstones with smaller

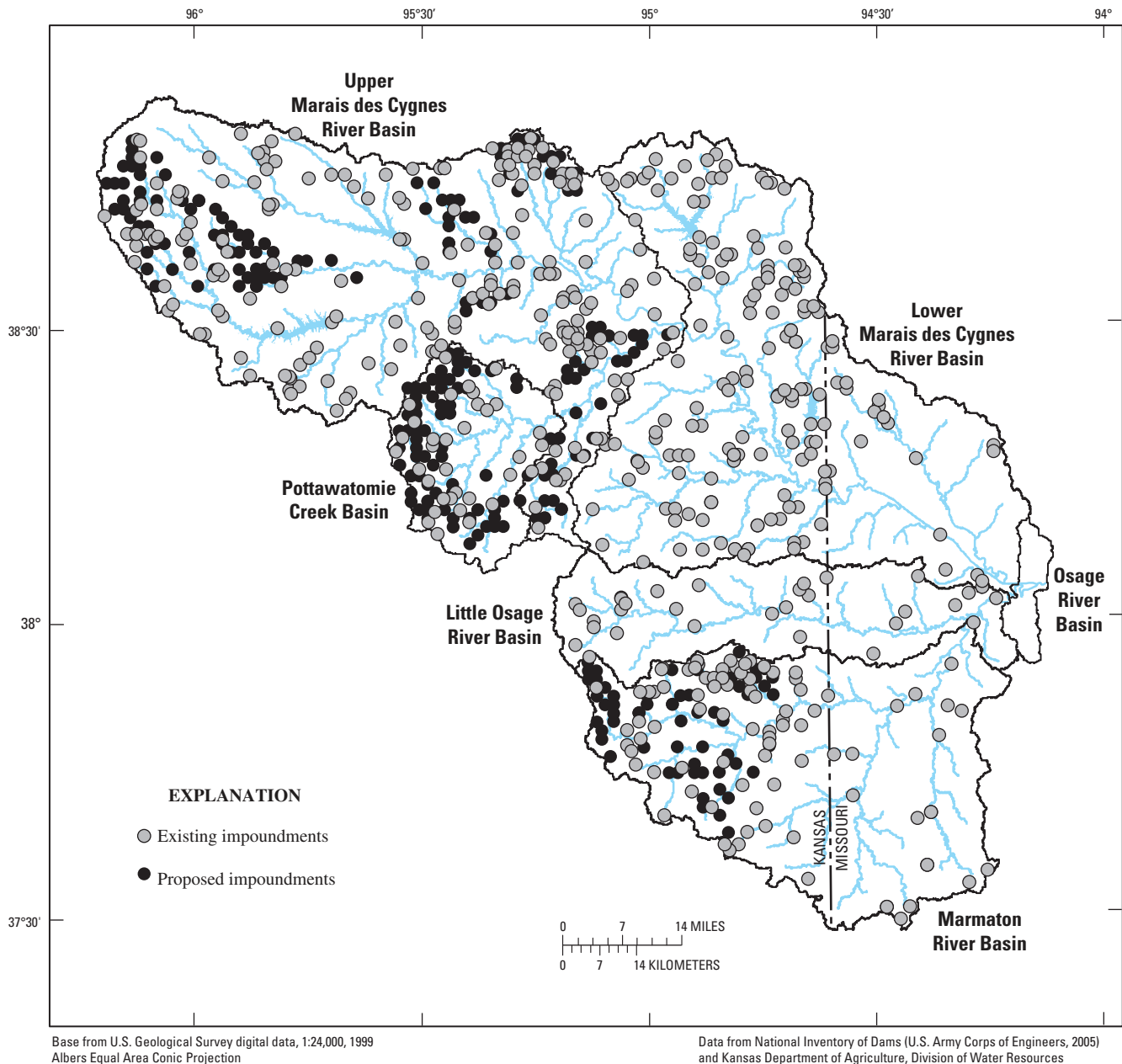


Figure 3. Existing and proposed impoundments in the upper Osage River Basin.

amounts of limestone and coal (Bevans and others, 1984). These geologic strata are of low permeability; consequently, streamflows are sustained primarily by surface runoff. Baseflows are not well sustained in the basin (Skelton, 1976), and ground-water usage is limited (Bevans and others, 1984).

The upper Osage River Basin is located in the Cherokee Prairies Major Land Resource Area (U.S. Department of Agriculture, 1981) and soils mostly are Aqualfs and Udolls that are shallow to deep, medium textured, and poorly to well drained. The geologic parent material is shale, sandstone, and limestone on gently sloping uplands. Soils developed from limestone and sandstone tend to be loamy and shallow on

ridges, whereas soils developed from shale usually are deep, claypan soils. Four major soil hydrologic groups are specified in the STATSGO data base (U.S. Department of Agriculture, 1991) including high infiltration rates (class A)—soils are deep, well drained to excessively drained sands and gravels; moderate infiltration rates (class B)—soils are deep and moderately deep, moderately well and well drained with moderately coarse textures; slow infiltration rates (class C)—soils with layers impeding downward movement of water or soils with moderately fine or fine textures; very slow infiltration rates (class D)—soils are clayey, have a high water table, or are shallow with an impervious layer. The upper Osage River

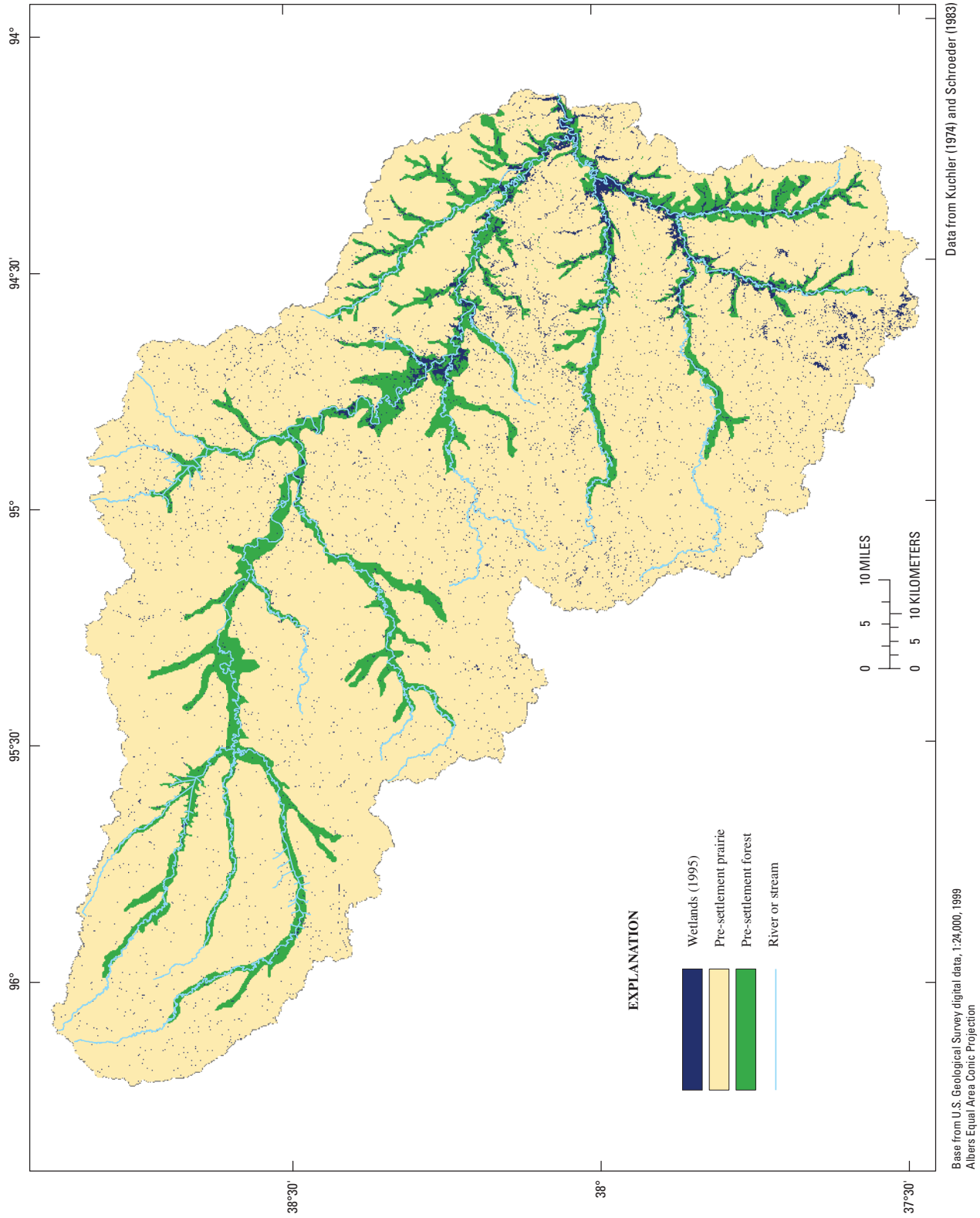


Figure 4. Historical land cover in the upper Osage River Basin.

Table 2. Land-cover information by major sub-basin within the upper Osage River Basin.

[--, no data]

Historical land cover	Percent of				Percent of			
	Little Osage River Basin	Marmaton River Basin	Upper Marais des Cygnes River Basin	Lower Marais des Cygnes River Basin	Upper Marais des Cygnes River Basin	Lower Marais des Cygnes River Basin	Percent of local Osage River Basin	Percent of local Osage River Basin
Prairie	85.2	84.6	87.6	82.0	87.6	82.0	77.4	77.4
Forest	10.0	9.7	11.3	13.9	11.3	13.9	11.8	11.8
Wetland ^a	4.80	5.70	1.1	4.1	1.1	4.1	10.8	10.8

1995 land cover	Percent effective imperviousness		Percent of Upper Marais des Cygnes River Basin		Percent of Lower Marais des Cygnes River Basin		Percent of local Osage River Basin	
	Little Osage River Basin	Marmaton River Basin	Upper Marais des Cygnes River Basin	Lower Marais des Cygnes River Basin	Upper Marais des Cygnes River Basin	Lower Marais des Cygnes River Basin	Percent of local Osage River Basin	Percent of local Osage River Basin
Agricultural	--	33.8	28.5	31.5	27.0	26.3	26.3	26.3
Forest	--	18.0	16.2	9.40	16.6	13.6	13.6	13.6
Pasture	--	37.8	40.4	24.5	51.4	43.2	43.2	43.2
Rangeland	--	5.45	8.00	32.7	.00	5.87	5.87	5.87
Urban or built-up	50.0	.33	1.20	.80	.90	.23	.23	.23
Wetland	--	4.80	5.70	1.10	4.10	10.8	10.8	10.8

^a1995 wetland area from National Land Cover Database (U.S. Geological Survey, 2005a) used as estimate for historic wetland area.

Basin primarily consists of class C soil types with some class D soils in valley deposits, and class B soils on uplands (fig. 6). The spatial distribution of the soil hydrologic classes was used to distribute infiltration and soil storage parameters in the hydrologic models.

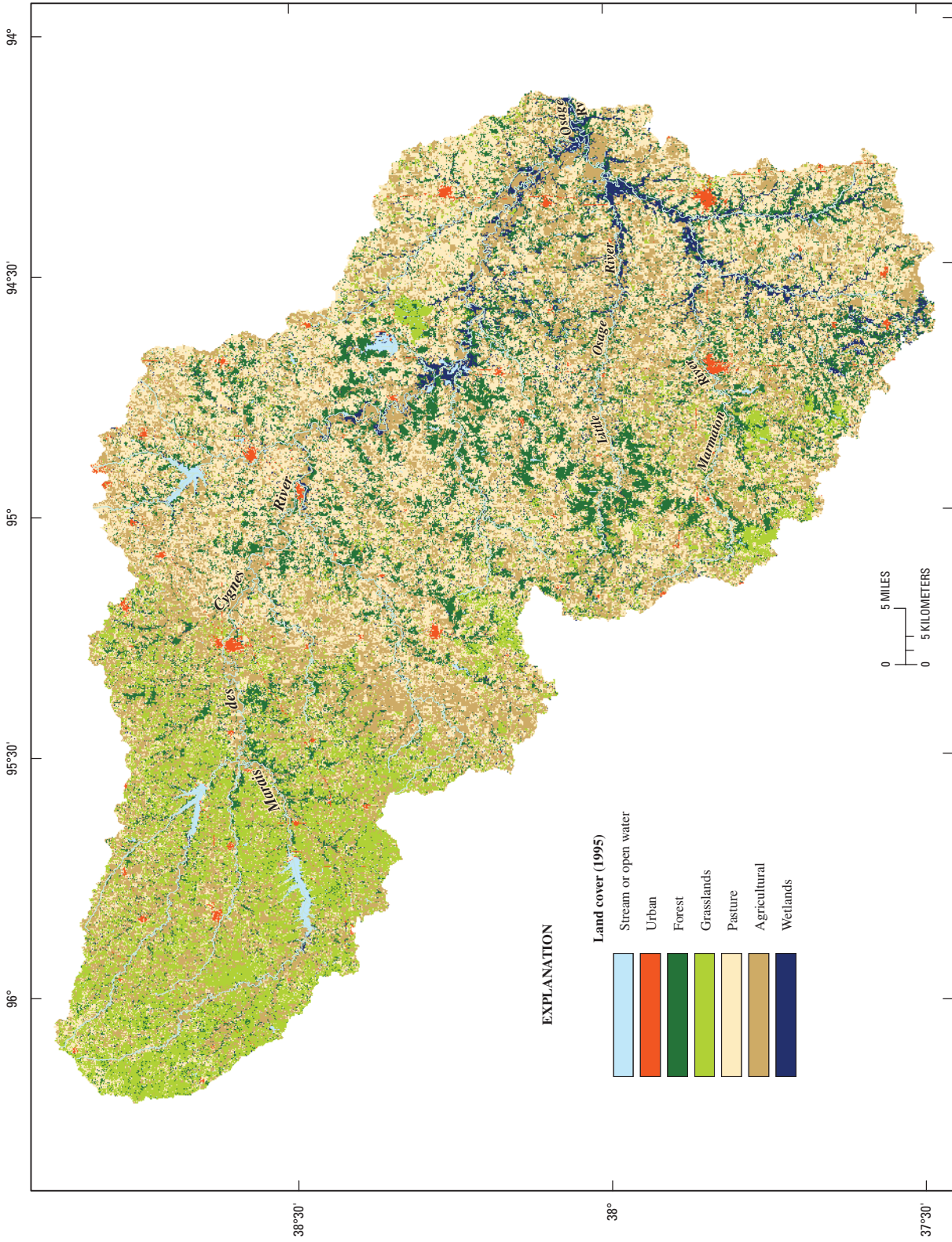
The land surface in the upper Osage River Basin is rolling plains (Bevans and others, 1984) with 900 feet (ft) of relief, ranging from about 710 ft near Schell City, Missouri, to a maximum of about 1,600 ft in the western headwaters of the Marais des Cygnes River Basin (fig. 7). The upper Osage Basin is characterized by gentle slopes with most of the basin having slopes less than 3 percent (fig. 8). The subbasins composing the upper Osage Basin lie primarily in a west-east orientation.

Acknowledgments

The authors acknowledge support provided by the Missouri Department of Conservation and technical assistance and guidance provided by Jason Persinger, Del Lobb, Bill Turner, and Ron Dent of the Missouri Department of Conservation. Technical assistance with the Hydrologic Simulation Program-FORTRAN was provided by Paul Duda and Paul Hummel with Aqua Terra Consultants. The authors acknowledge the assistance in summarizing and generating tables and figures provided Bruce Perkins (U.S. Environmental Protection Agency; USGS Volunteer for Science).

Development of the Upper Osage River Basin Hydrologic Model

The Hydrologic Simulation Program-FORTRAN (HSPF Version 12; Bicknell and others, 2004) was used to simulate hydrologic relations and streamflows at selected locations in the upper Osage River Basin under varying land cover and regulation scenarios. HSPF has been used to simulate hydrologic processes in a variety of geographic locations, spatial scales, and for a number of hydrologic applications (Duncker and others, 1995; Duncker and Melching, 1998; Jones and Winterstein, 2000; Zarriello and Sherwood, 1993; Zarriello, 1996; Coon and Johnson, 2005; Zarriello and Ries, 2000; Zarriello and Bent, 2004; Berris and others, 2001; Dinocola, 1990; Dinocola, 2001). HSPF is supported by the USGS



Base from U.S. Geological Survey digital data, 1:24,000, 1999
 Albers Equal Area Conic Projection

Data from National Land Cover Data Base (U.S. Geological Survey, 2005a)

Figure 5. Land cover in the upper Osage River Basin, 1995.

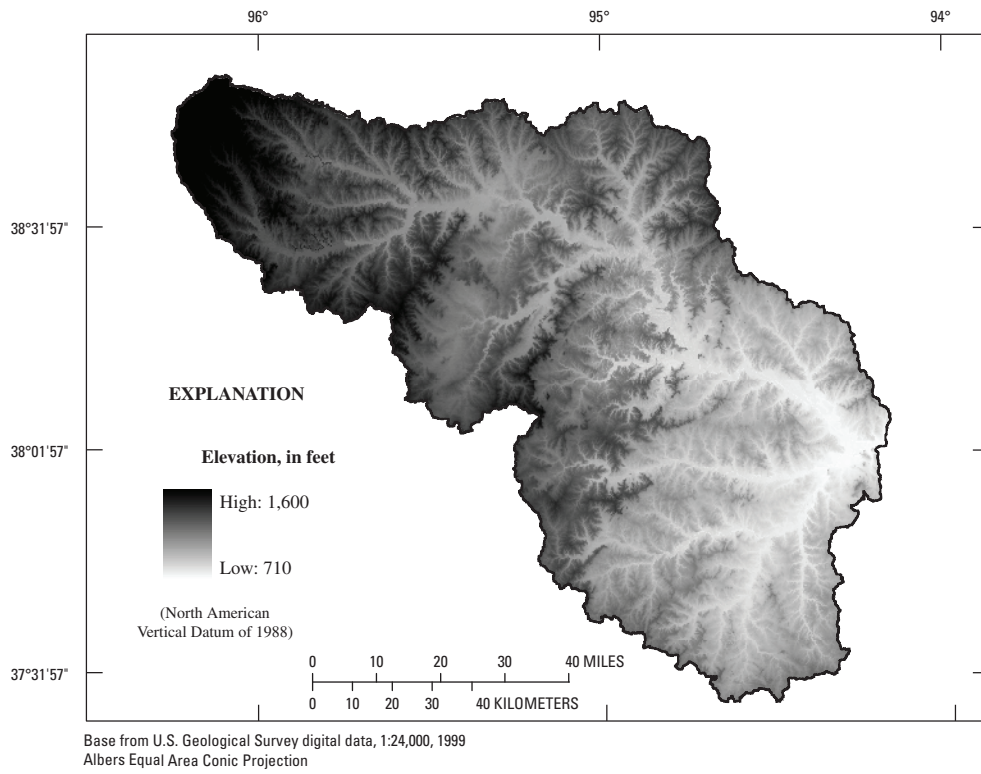


Figure 7. Topographic characteristics of the upper Osage River Basin.

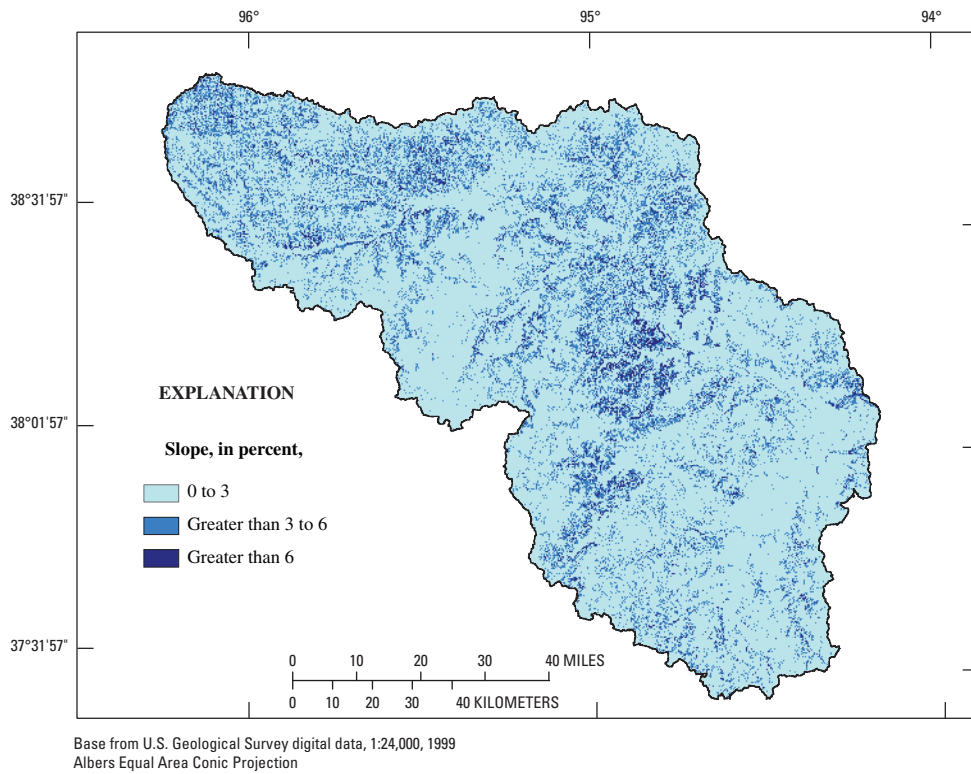
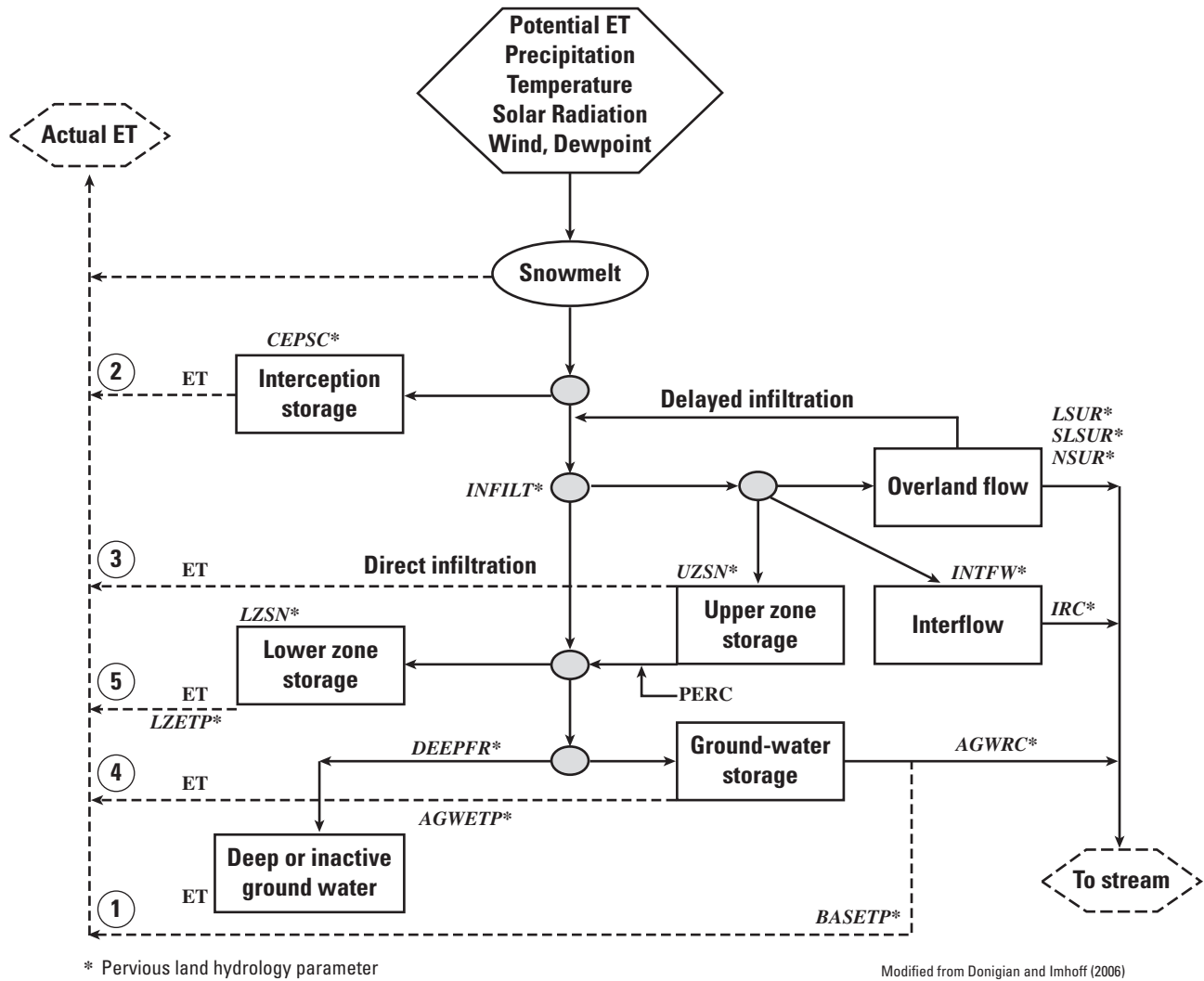


Figure 8. Land-surface slope characteristics of the upper Osage River Basin.



EXPLANATION

<i>AGWETP*</i>	Ground-water evapotranspiration	<i>IRC*</i>	Interflow recession constant		Input
<i>AGWRC*</i>	Ground water recession constant	<i>LSUR*</i>	Length of overland flow path		Process
<i>BASETP*</i>	Baseflow evapotranspiration	<i>LZETP*</i>	Lower zone evapotranspiration		Output
<i>CEPSC*</i>	Interception storage	<i>LZSN*</i>	Lower zone nominal storage		Storage
<i>DEEPPFR*</i>	Inactive ground water	<i>NSUR*</i>	Roughness of overland flow path		Order taken to meet ET demand
ET	Evapotranspiration	PERC	Percolation or delayed infiltration		Decision
<i>INFILT*</i>	Infiltration	<i>SLSUR*</i>	Slope of overland flow path		Path of model process
<i>INTFW*</i>	Interflow-inflow	<i>UZSN*</i>	Upper zone nominal storage		ET source

Figure 9. Schematic of pervious land hydrology module (PERLND) in the Hydrologic Simulation Program-FORTRAN.

ing the hydrologic response for each land-cover category varied for each PERLND area unit and were determined through the model calibration procedure.

The PERLND module of the HSPF model represents the primary inputs and processes used to generate runoff from pervious land segments (fig. 9). Input into HSPF consisted of precipitation, temperature, and potential evapotranspiration (PET) data. Snowfall was so minimal (about 1 in. of water equivalent) compared to rainfall (about 42 in.) that the snowmelt module was not used in any of the upper Osage River Basin simulations; that is, all precipitation was assumed to occur in the form of rainfall. Before infiltration (fig. 9), water can be stored as plant interception/depression storage, evaporate, or runoff to the stream. After infiltration (fig. 9), water can be stored in the upper zone, lower zone, ground-water storage, or deep aquifer/inactive ground-water storage. The upper-zone storage represents water stored and available in the shallow root zone (upper 6–12 in. of soil). The amounts of overland flow and interflow are affected by the upper-zone storage quantity (fig. 9). The lower zone consists of soil and geologic material, and supplies moisture to deep-rooted vegetation. The active ground-water zone contains ground water that provides baseflow to streams. The upper zone, lower zone, and active ground-water storage all may supply water to shallow/deep rooted vegetation, and thus contribute to evapotranspiration losses (fig. 9). The magnitude of the storage components and movement of water between these processes is controlled through the HSPF model parameters provided in figure 9.

Model Segmentation

The upper Osage River Basin was too large and complex to adequately represent it as a single model within HSPF, and, therefore, initially it was sub-divided into the Marais des Cygnes, Little Osage, and Marmaton River Basins based on the eight digit Hydrologic Unit Code (HUC) boundaries. With the addition of proposed impoundments it was necessary to further divide the Marais des Cygnes River Basin into smaller model units based on the scenario simulated (fig. 10). The local model results were routed together (output from one model was used as input to the downstream model) to determine streamflows at the basin outlet near Schell City (fig. 1).

Each model was divided into meteorological segments to capture the spatial variability of precipitation and temperature. These segments (figs. 11–13) corresponded to meteorological segments utilized by the National Weather Service Missouri Basin River Forecast Center (NWS-MBRFC) in an operational hydrologic forecast model of the Missouri River Basin. Hourly mean-area meteorological time-series were computed for each segment. NWS-MBRFC delineated the meteorological segments along subbasin boundaries using the GIS-based Integrated Hydrologic Automatic Basin Boundary System (IHABBS) (NOHRSC, 2005). IHABBS uses USGS 15 arc-second Digital Elevation Model (DEM) data and its derivatives, flow direction, and accumulation to delineate basin

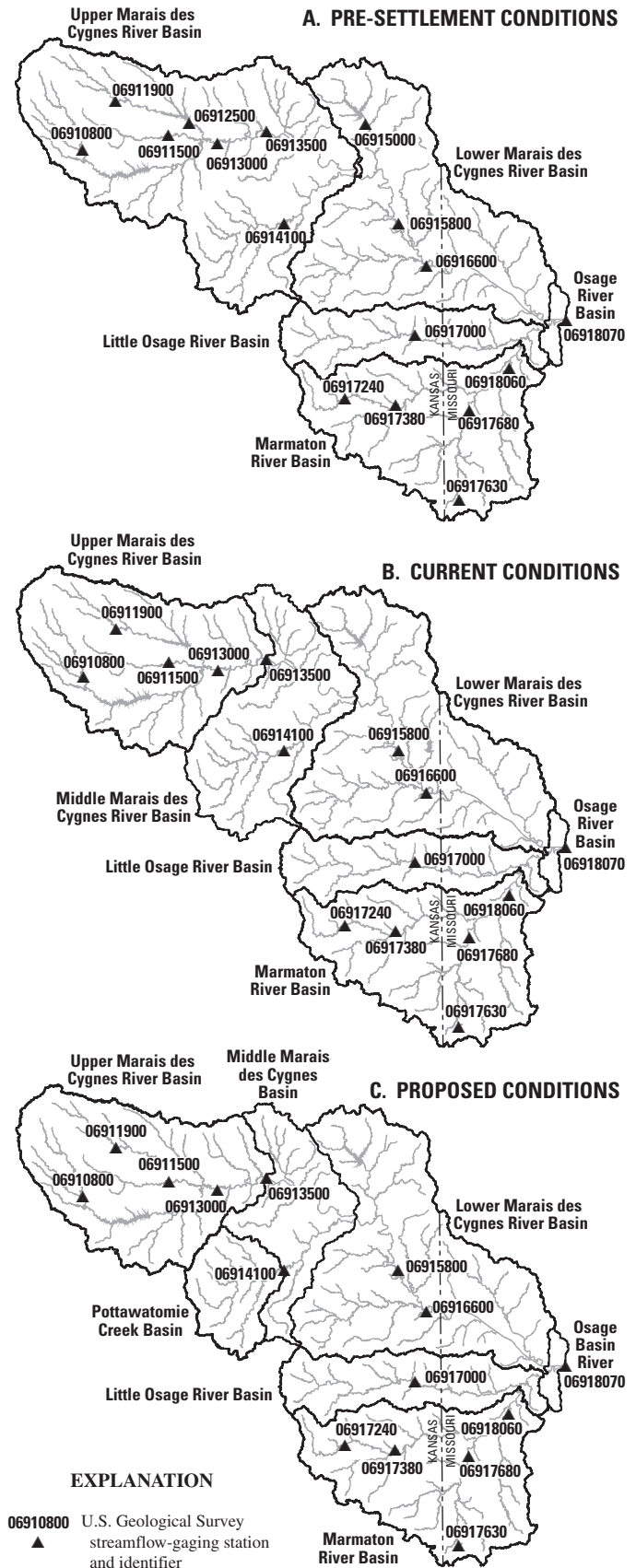
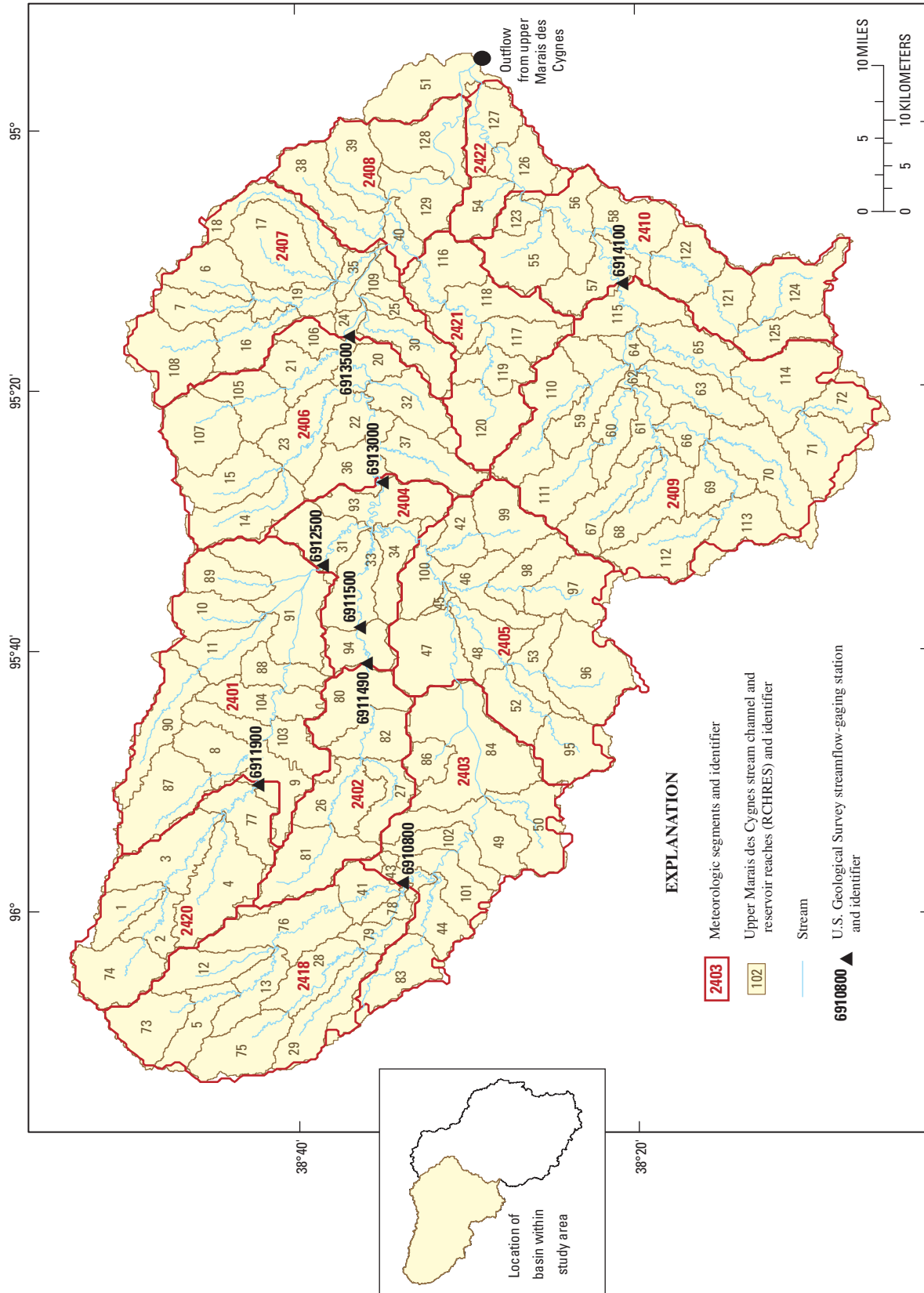


Figure 10. Upper Osage River basins for (A) pre-settlement (B) current, and (C) proposed simulation scenarios.



Base from U.S. Geological Survey digital data, 124,000, 1999
Albers Equal Area Conic Projection

Figure 11. Segmentation of the upper Marais des Cygnes River Basin.

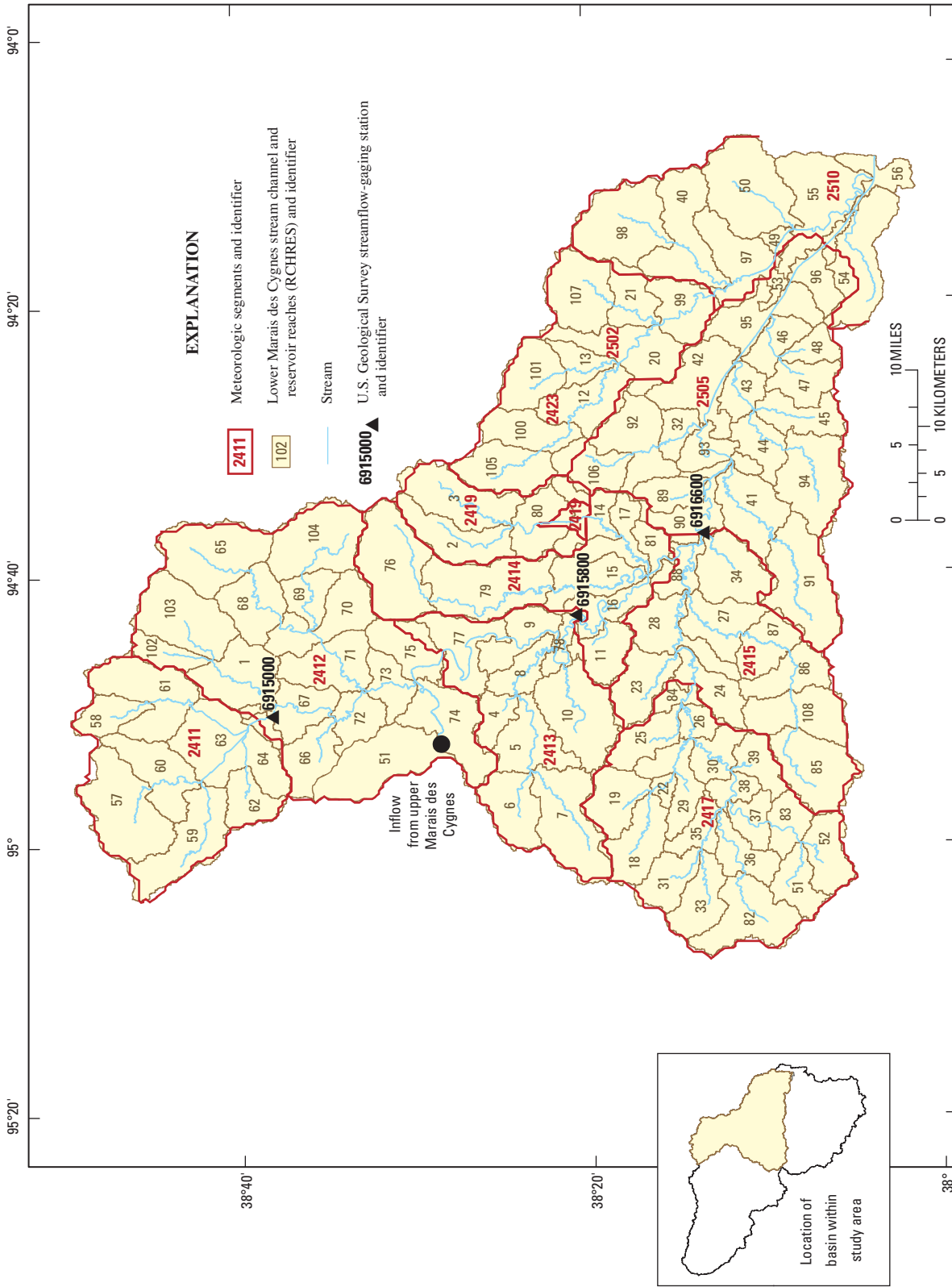


Figure 12. Segmentation of the lower Marais des Cygnes River Basin.

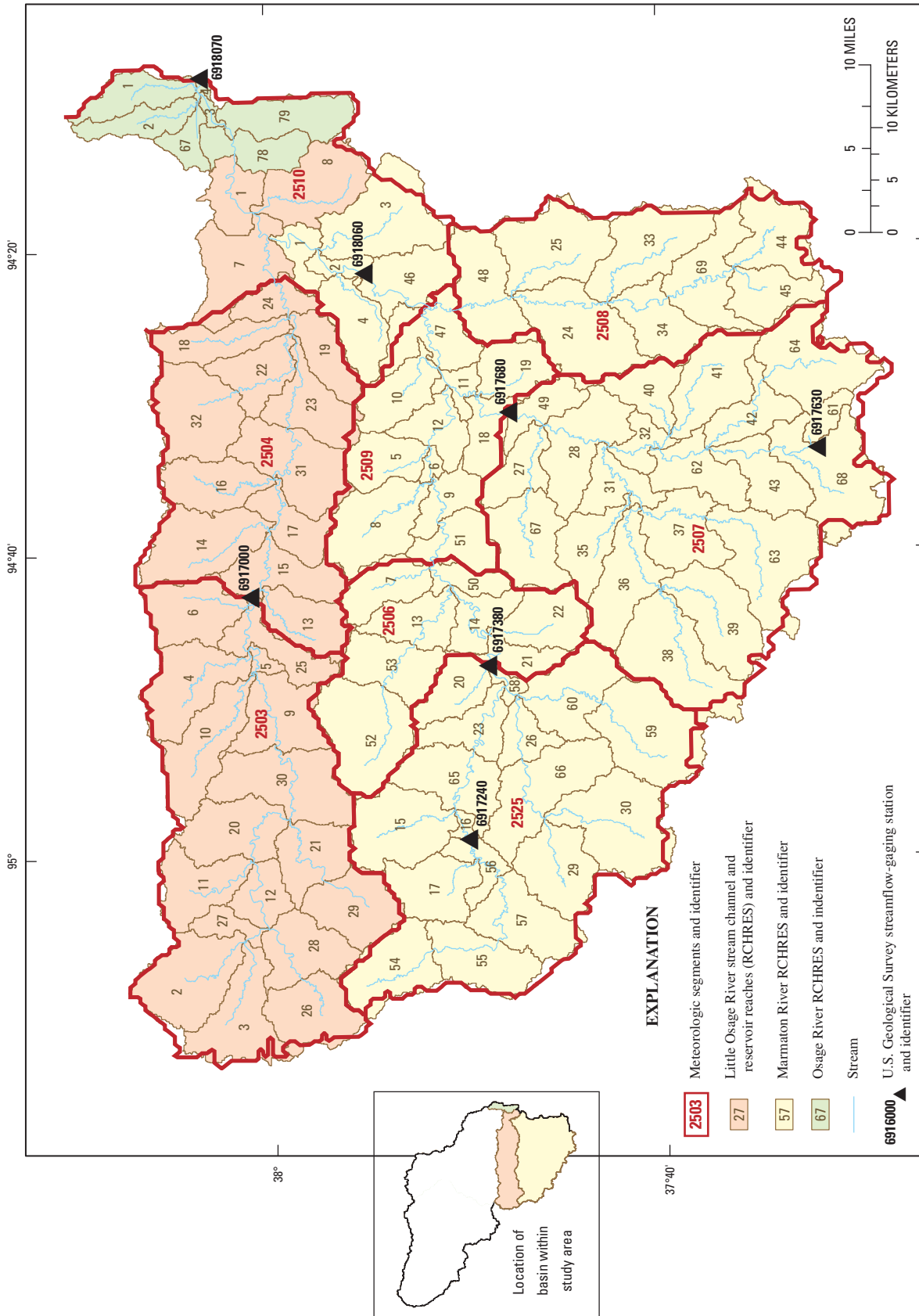


Figure 13. Segmentation of the Little Osage, Marmaton, and Osage River Basins.

boundaries. Each meteorological segment was sized to account for the spatial and temporal availability of meteorological station data, especially precipitation (Love and Donigian, 2004). Most of the meteorological segment outlets were located at USGS streamflow gaging station locations, current NWS segment locations, and large reservoirs. Outlets for model segments 2407, 2412, 2414, 2421, 2422, 2505, and 2508 (figs. 11–13) were chosen for the HSPF application to provide better spatial resolution of the meteorological characteristics.

Stream channel and reservoir reaches (RCHRES) were developed through automatic basin delineation within the pre-processing software Better Assessment Science Integrating Point and non-point Sources (BASINS, version 3.1; U.S. Environmental Protection Agency, 2005a). BASINS is incorporated into Arcview 3.3 (Environmental Systems Research Institute, 2002) and used for the development of the initial HSPF user-control input (UCI) files (examples of UCI files for selected Marmaton River Basin model scenarios are provided in appendixes 1–3, on compact disc, at the back of this report). Digital surfaces used in the development of RCHRESs included land cover, soils, topography, hydrography, meteorological segments, streamflow-gaging stations, and dams. The DEM used in model development was obtained from the National Elevation Dataset (NED) (USGS, 2005b). National Hydrographic Data (NHD) data (USGS, 2005c) were used in the “burn-in” option when auto-delineating the subbasins. A minimum drainage area of 11.6 mi² was used for stream definition. RCHRES outlets were made to correspond with USGS streamflow-gaging station locations to allow model calibration and validation at these gages. RCHRES areas were sized to approximate a 1-hour travel time through the stream reach, which corresponded to the primary simulation-time increment. Digital surfaces of existing and proposed impoundments were used to determine where to disaggregate automatically delineated RCHRES areas that were greater than 30 mi² into more evenly distributed impoundments within RCHRES, and to better approximate the 1-hour travel time distance. The ending mean stream RCHRES size was approximately 17 mi². Each impoundment was added to the model by first assigning it to a stream RCHRES based on its reported location. A new RCHRES was created within the original stream RCHRES to represent each individual impoundment, with the size of the newly created RCHRES corresponding to the contributing drainage area of the impoundment, and the size of the original RCHRES decreased accordingly. The areas of each land-cover category of the impoundment RCHRES were assumed to be proportional to that of the original RCHRES.

Time Series Data

Hourly meteorological and hydrologic time-series data were used in simulating hourly streamflow for the Marais des Cygnes, Little Osage, and Marmaton River models, whereas the Osage River model was run using daily input and output. The most recent 10-year period of processed meteorologic

and hydrologic time-series data available was from October 1, 1994, to September 30, 2004 (the 1995 through 2004 water years); these data were used for calibration, validation, and simulation of predicted model scenarios. Meteorological data consisted of precipitation, temperature, and evapotranspiration time series, whereas hydrologic time series consisted of observed streamflow and point-source surface-water withdrawals and discharges. All time-series data were stored and accessed using binary Watershed Data Management (WDM) files.

Meteorological Characteristics

Precipitation and temperature data collected at individual meteorological stations were spatially weighted, averaged, and uniformly applied over each defined meteorological segment. Potential evapotranspiration time series data were estimated from generated temperature time series using the Hamon method (Hamon, 1961), and also uniformly applied to each meteorological segment.

USGS and National Climatic Data Center (NCDC) precipitation stations located in and around the study area were analyzed to determine the best sites to be used in a precipitation network for the Thiessen polygon analyses. Meteorologic stations were selected based on length and quality of record following the guidelines described in Anderson (2002). Precipitation data were compiled from 74 stations located in and near the study basin (fig. 14; table 3, on compact disc, at the back of this report) including stations used for disaggregating purposes and those used to estimate missing record. Precipitation stations included stations from the NCDC daily cooperative network, NCDC hourly precipitation observations, hourly precipitation data collected at USGS streamflow-gaging stations, and one unofficial National Weather Service (NWS) daily observation site.

The precipitation station data were examined for consistency in addition to general quality and applicability. Station consistency is defined as the relation between observed record for a particular station and other stations within the network during the study period. This check helps ensure that the data are not biased with time, within the calibration period or between the model calibration and validation periods. A double-mass analysis was used to check the station consistency using techniques contained in the National Weather Service River Forecast Center’s (NWSRFS) calibration system Interactive Double Mass Analysis (IDMA) program (National Weather Service Office of Hydrologic Development, 1999; Pan and others, 1998). Guidelines for making resulting adjustments to the data are described by Anderson (2002). The adjustment factor is a multiplier that is applied to each data value in the time-series for the period of inconsistent record. Adjustment factors for this study varied from 0.85 to 3.5 (table 3) and the median adjustment factor was 1.30. Except for three daily reporting precipitation stations, all adjustment factors were applied to hourly reporting stations. Hourly tipping-bucket recording precipitation gages generally undercatch compared

to co-located or nearby daily observation stations (Hanson and others, 1996; Groisman and others, 1999) and, therefore, were used primarily for timing of disaggregation of daily stations. All adjustment factors were applied before the station data were used to compute mean areal precipitation values.

Hourly mean areal precipitation characteristics for each meteorological segment were calculated by the NWS Missouri Basin River Forecast Center using the Thiessen polygon analysis method contained in the NWSRFS's calibration system (National Weather Service Office of Hydrologic Development, 2004). A set of Thiessen polygons was computed for each meteorological segment using a common network of 39 precipitation stations (fig. 14). In using the Thiessen polygon method a separate set of precipitation station weights were computed for each segment (table 3) to represent the spatial affect each precipitation station had within the segment. Daily and hourly precipitation recording sites (fig. 14) were used in the creation of mean areal precipitation time-series.

After the adjustment factors were determined, all the mean areal precipitation time-series were computed using a utility in NWSRFS's calibration system. This utility applied the adjustment factors, computed the Thiessen weights for each meteorological segment (table 3), estimated any missing daily or hourly values, disaggregated the daily into hourly observations, and computed a mean areal precipitation time-series for each meteorological segment (National Weather Service Office of Hydrologic Development, 2004). Only 11 out of 31 meteorological segments used hourly stations in the Thiessen station weight analysis. Hourly precipitation observations were considered to be less consistent than daily observations, and were not used to compute mean areal precipitation except where the spatial coverage of the daily observations was considered inadequate. Because hourly USGS precipitation sites did not exist from October 1994 through September 1995, only NCDC hourly stations were used for this period to disaggregate the daily observations.

The hourly mean areal temperature characteristics for each meteorologic segment were computed from a network of 26 NCDC daily stations reporting daily maximum and minimum temperatures (fig. 15; table 4, on compact disc, at the back of this report). Station selection and quality-control procedures followed the methods outlined for precipitation stations. Missing temperatures were estimated using the nearest station. The areal station weights—a measure of the spatial affect of the temperature station on the meteorological segment—were calculated using the gridded methods in the NWSRFS calibration system for computing mean areal temperature (National Weather Service Office of Hydrologic Development, 2004). The final mean areal temperature daily maximum and minimum values for each meteorological segment were calculated by multiplying the station weights by the daily station temperatures and summing the weighted daily temperatures for each segment. The program WDMUtil (U.S. Environmental Protection Agency, 2005b; available at U.S. Environmental Protection Agency, 2005a) was used to disaggregate the minimum and maximum temperatures into hourly

values using a fixed diurnal temperature pattern because the hours of the daily maximum and minimum temperatures were not known.

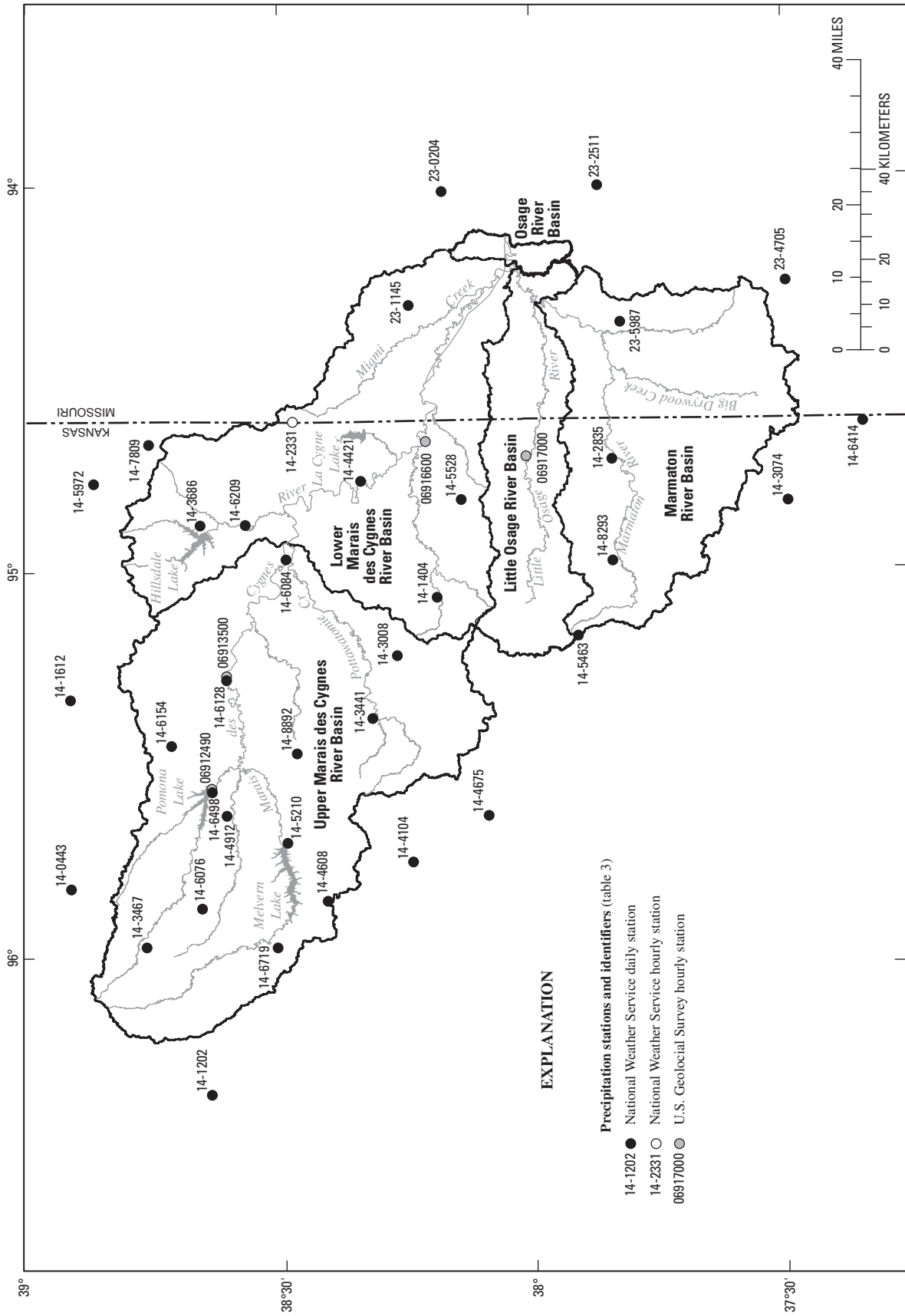
Hourly potential evapotranspiration time series were computed using the Hamon Method (Hamon, 1961) within the USGS program WDMUtil (U.S. Environmental Protection Agency, 2005b; available at U.S. Environmental Protection Agency, 2005a). The Hamon method is a simplified potential evapotranspiration estimate utilizing the latitude, hourly temperature, and monthly constants as input. The default monthly constants within WDMUtil were used for all the Hamon evapotranspiration calculations.

Hydrologic Characteristics

All documented point-source withdrawals and discharges in the upper Osage River study basins were included in the HSPF hydrologic simulations. Observed hourly streamflows were used for calibration and validation of the simulated streamflow conditions at each streamflow gage location, with the exception of the Osage River near Schell City, Missouri, streamflow gage location where mean daily values were used.

Point-source surface-water withdrawals (including those for municipal, industrial, irrigation, and recreation uses) from the Kansas part of the upper Osage River Basin were reported either monthly or annually from the Kansas Department of Agriculture, Division of Water Resources (Joe File, Kansas Department of Agriculture, Division of Water Resources, written commun., 2005). There were 162 permitted withdrawals affecting 42 RCHRESs in the Kansas part of the upper Osage River Basin (fig. 16). Annual estimates from 23 documented withdrawals were available for the Missouri part of the upper Osage River Basin from “major users” as defined by Missouri Department of Natural Resources (Amy Crews, Missouri Department of Natural Resources, Missouri Water Resources Center, written commun., 2005). The annual or monthly municipal and industrial withdrawals were disaggregated into daily values based on communications with the water-use entities. The daily values were then evenly disaggregated into hourly values using the disaggregation function within WDMUtil. Reported annual irrigation withdrawals in Kansas and Missouri were distributed during July and August of each year as this is the primary irrigation period in the basins (Vernon County Missouri Office of the University of Missouri Extension service, oral commun., 2004). Withdrawals in each model basin primarily were from surface- rather than ground-water sources—either directly from a river or a surface-water impoundment (fig. 17). Runoff retained in selected Kansas recreational impoundments was classified as a permitted withdrawal, and such withdrawals were not included in the HSPF models. Only those withdrawals that were obtained directly from a stream channel (as opposed to an impoundment) and used for a purpose other than impoundment storage were included in the HSPF simulations.

Discharges for nine Kansas permitted point-source discharges at 13 locations (fig. 16) were obtained from the



Base from U.S. Geological Survey digital data, 1:24,000, 1989
 Albers Equal Area Conic Projection

Figure 14. Distribution of primary precipitation stations used in the upper Osage River Basin.

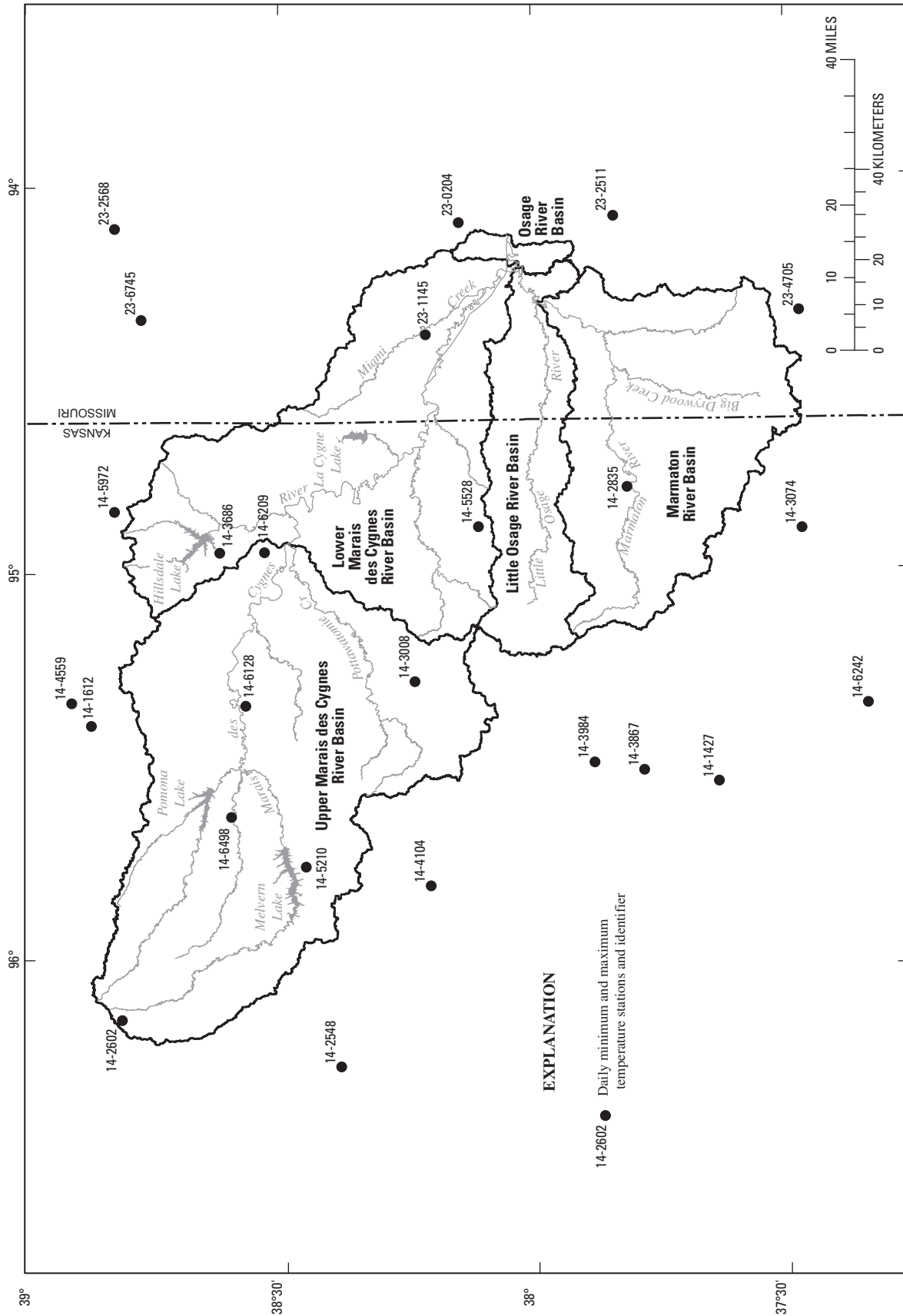


Figure 15. Distribution of temperature stations used in the upper Osage River Basin.

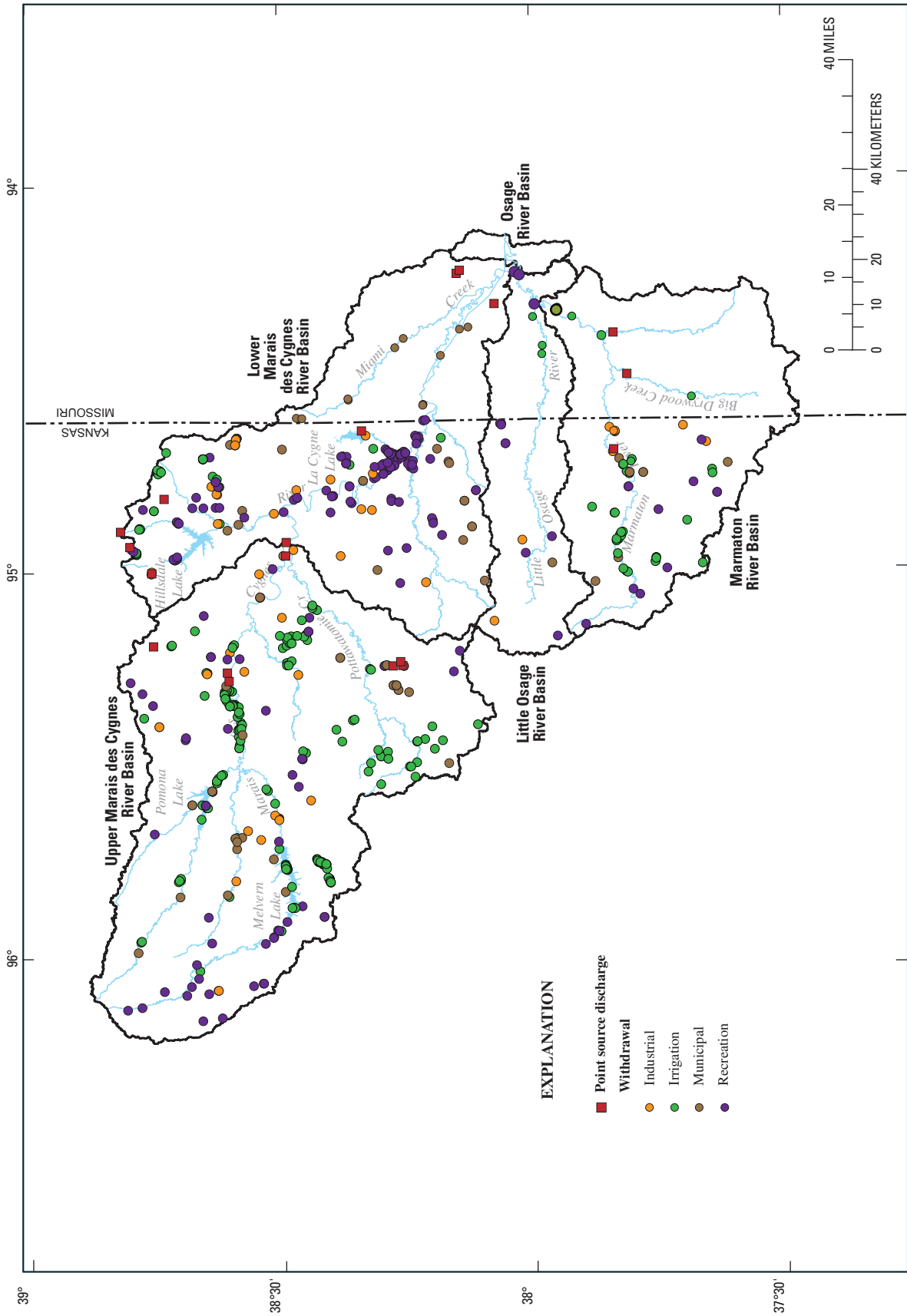


Figure 16. Location of reported point-source discharges and withdrawals in the upper Osage River Basin.

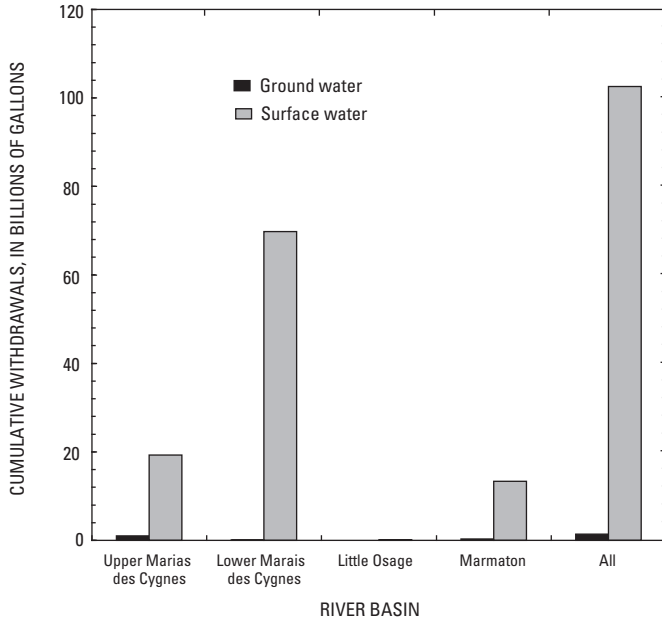


Figure 17. Comparison of cumulative reported ground-water and surface-water withdrawals in the upper Osage River Basin, water years 1995 through 2004.

Kansas Department of Health and Environment (KDHE, Ed Dillingham, Kansas Department of Health and Environment, written commun., 2005) and reported daily or monthly. Point discharges for three Missouri primary water users at five locations (fig. 16; Amy Crews, Missouri Department of Natural Resources, written commun., 2005) also were incorporated into the models. Discharges were primarily associated with municipal (sewage effluent) sources, but also included industrial sources. Monthly values were disaggregated into daily and then hourly values by assuming a constant rate of discharge during the day or month.

Reported point-source withdrawals in the Marais des Cygnes, Little Osage, and Marmaton River Basins generally exceeded point-source discharges (fig. 18). Cumulative reported point-source withdrawals in the Marais des Cygnes Basin generally were less than 40 ft³/s but reached a maximum of about 90 ft³/s in 2002 (fig.18). Cumulative reported discharges in the Marais des Cygnes Basin were less than 10 ft³/s, with net basin loss of about 8 ft³/s during non-irrigation periods and a net loss of about 80 ft³/s during irrigation periods (fig. 18). Reported withdrawals in the Little Osage River Basin generally were less than 2 ft³/s, and there were no reported point-source discharges in this basin. Cumulative reported point-source withdrawals generally were under 3 ft³/s in the Marmaton River Basin, except during summer irrigation (July-August) when estimated withdrawals could exceed 7 ft³/s (fig. 18). Cumulative point-source discharges in the Marmaton River Basin generally were less than 1 ft³/s resulting in a net loss of about 2 ft³/s in flows during non-

irrigation periods, and as much as about 6 ft³/s during irrigation periods.

Most of the reported total annual withdrawal volume in the modeled Marais des Cygnes Basin for 1995–2000 was for municipal purposes (fig. 19); however, the primary use for the total annual withdrawals from 2001 through 2003 was for recreation purposes, whereas in 2004 the primary withdrawal use

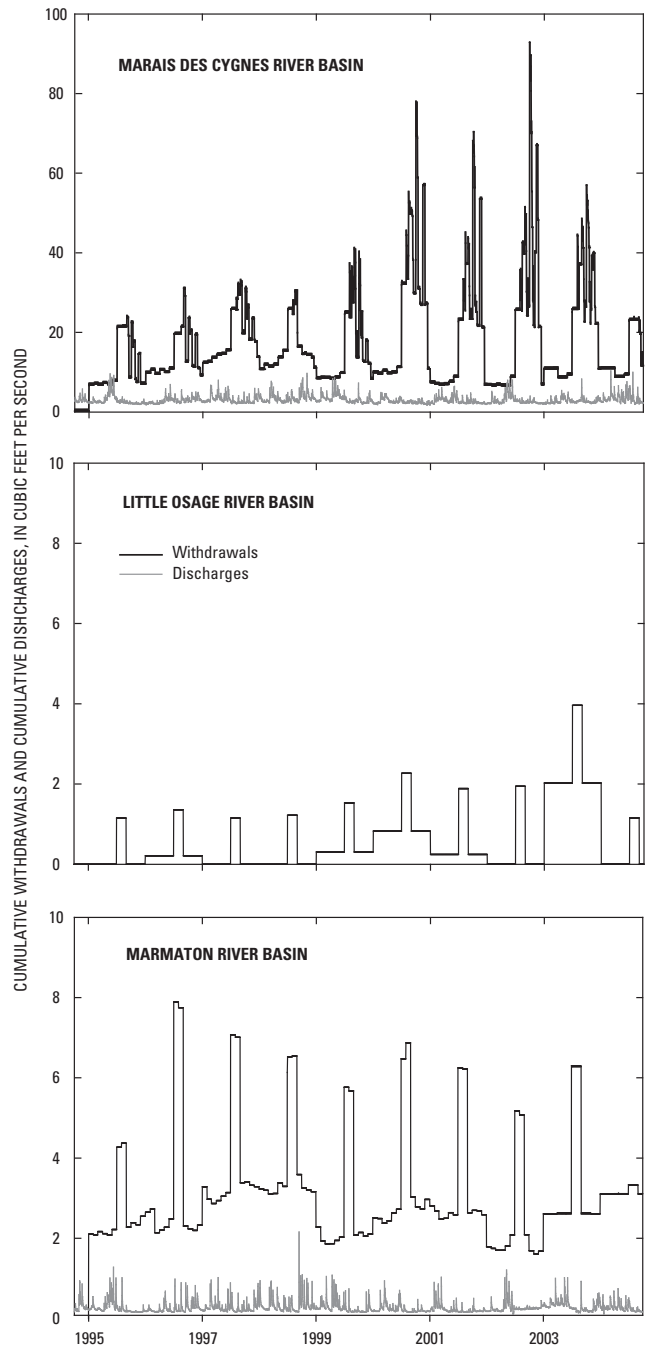


Figure 18. Temporal distribution of point-source withdrawals and discharges in the Marais des Cygnes, Little Osage, and Marmaton River Basins, water years 1995 through 2004.

was again municipal. Irrigation and recreation were the two primary use categories for the reported total annual withdrawals for the Little Osage model basin (fig. 19), whereas municipal and irrigation withdrawals were the primary withdrawal categories for the Marmaton River Basin (fig. 19).

Observed streamflow records from 20 USGS continuous streamflow-gaging stations were used for calibration/validation for all or part of the 1995–2004 study period (table 5, fig. 10). The hourly observed record from the USGS streamflow-gaging stations below the large reservoirs in Kansas (Pomona, Melvern, and Hillsdale reservoirs; fig. 1) were used as input points for the current (2005) and proposed Marais des Cygnes

River simulation scenarios; the observed outflows from the reservoirs were used instead of simulating runoff upstream from, and routing streamflows through, the large reservoirs in the current and proposed models. The outflow from LaCygne Lake (fig. 1) also was used as a model input point rather than simulating runoff into and through this reservoir. The outflow from LaCygne Lake was estimated based on daily lake stage and stage/outflow rating provided by the LaCygne Power plant (Bruce Beckman, Kansas City Power and Light Company, written commun., 2005).

Stage-Area and Stage-Volume Outflow Relations

Relations between water depth and surface area, volume, and outflow were developed for all model channel reaches and impoundments (all RCHRESs). Volume-outflow relation tables (FTABLEs) were developed for each RCHRES by BASINS in the development of the initial UCI file based on channel geometry and slope derived from the input DEM. Whereas the HSPF hydrologic simulations are relatively insensitive to the FTABLEs for streamflow-only applications (Alan Lumb, U.S. Geological Survey, oral commun., 2004; Tony Donigian, Aqua-Terra Consultants, oral commun., 2004), the original FTABLEs were replaced with tables derived manually for both stream channel and impoundment RCHRESs to increase the accuracy of streamflow routing.

Initial hydrologic conditions were specified for each stream and impoundment RCHRES in the models. Flow conditions for each stream RCHRES were set to observed discharges at the beginning of the simulation. The initial volumes for impoundments generally were set at 75 percent of capacity, based in part on quarterly permittee information provided to the Kansas Department of Agriculture (Joe File, Kansas Department of Agriculture, written commun., 2005).

FTABLEs for stream RCHRESs were developed using the USGS programs CGAP and GenFtable (Regan and Schaffranek, 1985; available at USGS, 2005d). Data input consisted of channel cross section, slope, and Mannings “n” roughness coefficients for each computation reach. Channel cross sections were obtained from discharge measurements at USGS streamflow-gaging station locations, measured cross sections within the upper Marmaton Basin (Ed Radatz, Natural Resource Conservation Service, written commun., 2005), or measured cross sections available for the Marmaton and Marais des Cygnes Rivers from a previous study of hydraulic conditions and fish habitat (Heimann and others, 2005). Data from 64 measured cross sections were applied to the remaining 278 unmeasured sites based on similarities in contributing drainage area. Channel slope for each RCHRES was obtained from the BASINS analyses of the input DEM. Manning’s “n” roughness coefficients were estimated from a calibrated one-dimensional model of the upper Marmaton River (Ed Radatz, USDA NRCS, written commun., 2005) and from calibrated two-dimensional models of selected reaches of the Marmaton

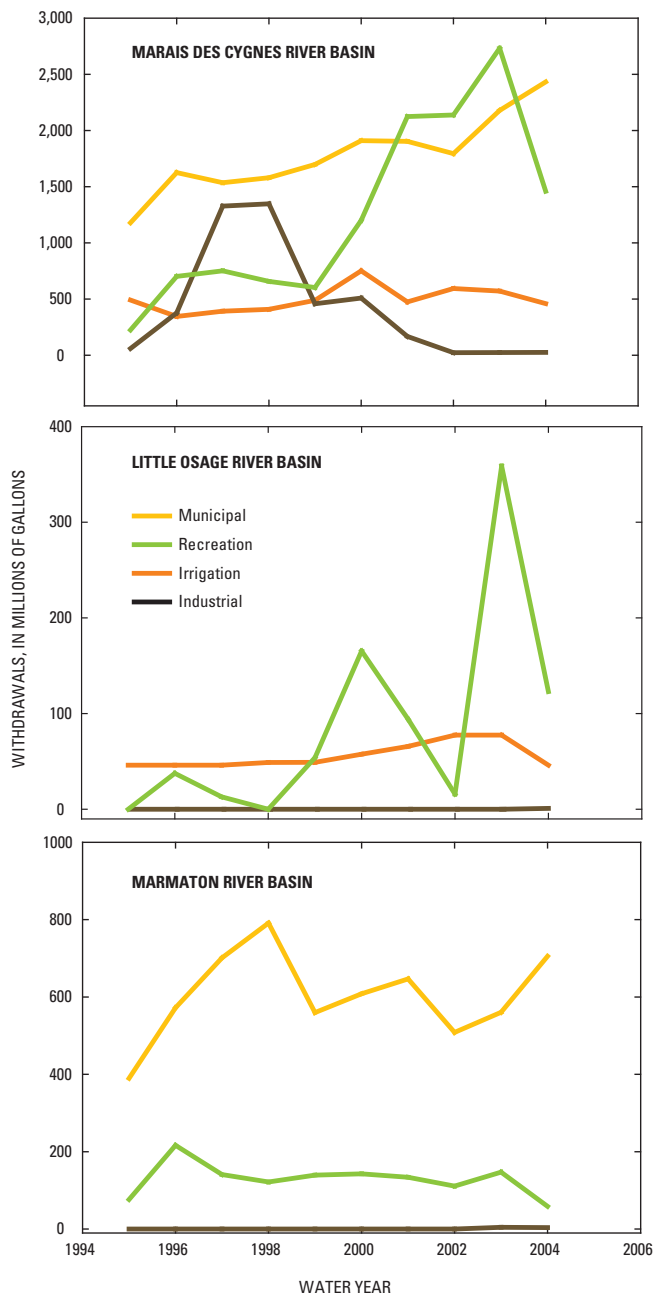


Figure 19. Annual reported surface-water withdrawals, by basin and water-use category, water years 1995 through 2004.

Table 5. U.S. Geological Survey streamflow-gaging stations used in the development and calibration/validation of the upper Osage River Basin models.

[USGS, U.S. Geological Survey; streamflow records for stations in gray were combined for calibration/validation]

USGS streamflow-gaging station number	Station name	Drainage area, in square miles	Period of discharge record
06910800	Marais des Cygnes river near Reading, Kansas	177	1969–2004
06911490	Salt Creek at Lyndon, Kansas	97.8	1999–2003
06911500	Salt Creek near Lyndon, Kansas	111	1940–1999
06911900	Dragoon Creek near Burlingame, Kansas	114	1960–2004
06912500	Hundred and Ten Mile Creek near Quenemo, Kansas	322	1939–2004
06913000	Marais des Cygnes River near Pomona, Kansas	1,040	1922–1938; 1968–2004
06913500	Marais des Cygnes River near Ottawa, Kansas	1,250	1918–2004
06914000	Pottawatomie Creek near Garnett, Kansas	334	1939–2001
06914100	Pottawatomie Creek near Scipio, Kansas	343	2001–2004
06915000	Big Bull Creek near Hillsdale, Kansas	147	1958–2004
06915800	Marais des Cygnes River at La Cygne, Kansas	2,670	1984–2004
06916600	Marais des Cygnes River near Kansas-Missouri state line, Kansas	3,230	1958–2004
06917000	Little Osage River at Fulton, Kansas	295	1948–2004
06917240	Marmaton River at Uniontown, Kansas	84	2001–2004
06917380	Marmaton River near Marmaton, Kansas	292	1971–2004
06917630	East Drywood Creek at Prairie State Park, Missouri	3.92	2001–2004
06917680	Dry Wood Creek near Deerfield, Missouri	358	2001–2004
06918060	Marmaton River near Nevada, Missouri	1,070	2003–2004
06918065	Marmaton River below Nevada, Missouri	1,090	2000–2003
06918070	Osage River above Schell City, Missouri	5,410	1981–2004

and Marais des Cygnes River Basins (Heimann and others, 2005) for 48 cross sections. At the remaining 294 sites for which no roughness information was available, Manning's "n" values were assigned based on roughness values of reaches with similar contributing drainage area.

FTABLES for each of 539 existing (2005) and 283 proposed impoundments were developed in the HSPF simulations using impoundment design specifications obtained from the National Inventory of Dams (U.S. Army Corps of Engineers, 2005); the Kansas Department of Agriculture, Division of Water Resources (Joe File, Kansas Department of Agriculture, Division of Water Resources, written commun., 2005); and stage-discharge relations developed for culvert and weir structures. The number of impoundments simulated in this study is a conservative estimate as an impoundment is included in the National Inventory of Dams only if it's "high" or has a "significant" hazard potential dam, or is a "low" hazard potential dam that exceeds 25 ft in height and 15 acre-ft of storage, or

has a "low" hazard potential dam that exceeds 6 ft in height and 50 acre-ft of storage (U.S. Army Corps of Engineers, 2005). The National Inventory of Dams data were augmented with a listing of all permitted reservoirs within the upper Osage River Basin in Kansas as maintained by the Kansas Department of Agriculture, Department of Water Resources (Joe File, Kansas Department of Agriculture, Division of Water Resources, written commun., 2005). Impoundment specifications in the National Inventory of Dams included contributing drainage area, water volumes, and surface areas at the primary and emergency spillway levels, along with dimensions of the primary and secondary spillway structures. The outflow from the primary spillway culvert was determined using reservoir water depth, culvert diameter, and an outflow relation defined in Ward and Elliot (1995) provided in table 6. These primary spillway outflow values were corrected for culvert length (Ward and Elliot, 1995) using corrections provided in table 7. Culvert length was estimated from a formula computed from known impoundment

characteristics (Joe File, Kansas Department of Agriculture, Division of Water Resources, written commun., 2005) as:

$$L_c = H_d \times 5.5 + 20 \tag{1}$$

where

- L_c is culvert length, in ft; and
- H_d is height of dam.

The emergency spillway water depth-outflow relation was defined using the following equation for a broad crested weir as defined in Hulsing (1967)

$$Q = CbH^{3/2} \tag{2}$$

where

- Q is discharge in ft³/s
- C is a coefficient of discharge (assumed value of 3.1)
- b is width of the weir normal to the flow, in ft
- H is total energy head ($h + V_1^2/2g$ where h = static head and V_1 is the mean velocity at the approach section to the weir, and g = acceleration because of gravity), in feet. For the purpose of this application it was assumed that the total energy head was equal to the depth of water above the weir crest.

Total outflow was calculated for each impoundment as the sum of the corrected primary outflow and the calculated emergency spillway outflows.

For all but 35 impoundments there was at least one missing outflow design specification; these missing values were estimated based on correlations of design specification for reservoirs with existing information. Missing contributing drainage area, surface area at primary and emergency spillway elevations, and storage volumes at the primary and emergency spillway elevations were estimated using a simple

linear regression equation developed from known data within each model basin (table 8). Missing values for those relations of impoundment characteristics that did not follow a simple linear expression were estimated by using categorical breaks in graphical relations between predictive design specifications. The relations between detention volume and dam height (table 9), detention volume and elevation difference between primary and emergency spillways (table 10), impoundment drainage area and primary spillway culvert diameter (table 10), and drainage area and emergency spillway width (table 10) were plotted, and slope breaks were used to define three to five discrete categories. The distribution within these categories was used in defining the relation between variables shown in tables 9–10. As culverts are manufactured in fixed diameters, the relation between drainage area and culvert diameter yielded a discrete culvert size estimate for each of five established contributing drainage area categories (table 10).

To quantify the range of variation that might result from estimating unknown impoundment design variables, three outflow ratings, or FTABLES, were constructed for each impoundment with an estimated outflow design variable (these included primary culvert diameter, spillway width, and/or elevation differential between primary and secondary spillway). These three outflow ratings (10th-, 50th-, and 90th-percentile rating) were constructed using all reported impoundment design information, and an estimate of the corresponding 10th-, 50th-, or 90th-percentile design values (table 10) only for any missing outflow design variables. The 10th-percentile design scenario was representative of the smallest outflow structures, from the distribution of reported specifications, and had the longest detention times, whereas the 90th-percentile design scenario was representative of largest reported outflow structures, and had the shortest detention times. The 10th-percentile rating scenario would, for instance, include FTABLES constructed using the 10th percentile value from distributions of known values for any missing impoundment outflow variable for all impoundments in a particular model basin. In this way the sensitivity of

Table 6. Pipe flow outflow rates, in cubic feet per second, with varying pipe diameter and water depths (from Ward and Elliot, 1995).

Water depth, in feet	Pipe diameter, in inches				
	18	24	30	36	48
1	5.47	11.0	18.8	28.8	55.7
2	7.74	15.6	26.6	40.8	78.8
3	9.48	19.1	32.6	49.9	96.5
4	10.9	22.1	37.6	57.7	111
5	12.2	24.7	42.1	64.5	125
6	13.4	27.0	46.1	70.6	136
8	15.5	31.2	53.2	81.5	158
10	17.3	34.9	59.5	91.2	176
12	19.0	38.2	65.2	99.9	193
15	21.2	42.8	72.8	112	216

Table 7. Correction factors (dimensionless) for spillway culvert pipe lengths (from Ward and Elliot, 1995).

Pipe length, in feet	Pipe diameter, in inches				
	18	24	30	36	48
20	1.42	1.34	1.28	1.24	1.18
30	1.29	1.24	1.21	1.18	1.13
40	1.20	1.17	1.14	1.12	1.10
50	1.12	1.1	1.09	1.08	1.06
60	1.05	1.05	1.04	1.04	1.03
70	1.00	1.00	1.00	1.00	1.00
80	.95	.96	.96	.97	.97
90	.91	.92	.93	.94	.95
100	.88	.89	.90	.91	.93
120	.82	.83	.85	.86	.89
140	.77	.79	.81	.82	.85
160	.73	.75	.77	.79	.82

model results to the estimation of design criteria, representing a range of possible detention characteristics, could be quantified, and the effects of altering outflow structures of proposed impoundments also could be assessed. FTABLEs for existing (2005) impoundments were developed using the 50th-percentile value from the distribution of known values for any estimated outflow variables, and these median condition FTABLES were used for calibration and validation scenarios.

Whereas the number of proposed reservoirs is about one-half that of existing (2005) impoundments, the proposed reservoirs provide significantly higher median detention storage [Mann-Whitney test (Helsel and Hirsch, 1992), $p < 0.001$; significance level = 0.05] and regulated contributing drainage area ($p < 0.001$) when compared to existing impoundments (fig. 20). Median detention storage in proposed impoundments is 400 percent greater than existing median impoundment storage in the Marais des Cygnes River Basin, 428 percent greater in the Marmaton River Basin, and 358 percent greater in the entire upper Osage River Basin.

Model Calibration, Validation, Sources of Error, and Model Uncertainty

Calibration of the HSPF models was conducted by minimizing differences between simulated and observed streamflow data at model locations corresponding to USGS streamflow-gaging stations by adjusting model process-related parameters. Validation was used to test the calibration on an independent data set. In addition to the calibration and validation techniques and results, the possible sources of error in the model representation are discussed, along with uncertainties in the model results and techniques used to minimize these uncertainties.

Calibration and Validation Methods

Calibration is the process by which the mathematical model process-related parameters are adjusted such that simulated results are fitted to observed data. Simulated results were fitted to observed data by varying model parameters through “trial and error” and determining fit using the weight-of-evidence approach as described in Donigian (2002). Because the “trial and error” calibration process was subjective, it represents one of several possible combinations of model parameters that could be used to fit simulated to observed streamflow values. The weight-of-evidence approach utilizes several graphical (qualitative) and quantitative characteristics in optimizing the model fit during the subjective calibration process. Graphical comparisons included arithmetic and logarithmic time series plots of observed and simulated data, flow duration plots of observed and simulated streamflows, and scatter plots of monthly and daily observed and simulated streamflow. Quantitative evaluation of model fit was based on the USGS program HSPF Expert (HSPFEXP; Lumb and others, 1994) output statistics, mean error statistics, along with additional correlation coefficients, including the correlation coefficient, coefficient of determination, and model fit efficiency.

The correlation coefficient, r , was calculated as

$$r = \frac{\sum_{i=1}^N (q_o - \bar{q}_o) \times (q_s - \bar{q}_s)}{\sqrt{\sum_{i=1}^N (q_o - \bar{q}_o)^2 \times \sum_{i=1}^N (q_s - \bar{q}_s)^2}} \quad (3)$$

where

- q_o is observed flow for given time step,
- \bar{q}_o is average observed flow for given time step,
- q_s is simulated flow for given time step,
- \bar{q}_s is average simulated flow for given time step.

Table 8. Linear regression models used in estimating missing outflow design specifications for impoundments in the upper Osage River Basin.

Estimated specification	Regression model	Values in analysis	Coefficient of determination
Upper Marais des Cygnes River Basin			
Contributing drainage area	$y = 0.0037 X$ Volume at emergency spillway, in acre-feet	381	0.76
Surface area at emergency spillway, in acres	$y = 3.41 X$ Primary spillway surface area, in acres	360	.82
Volume at emergency spillway, in acre-feet	$y = 5.45 X$ Primary spillway volume, in acre-feet	333	.71
Surface area at primary spillway, in acres	$y = 5.55 X$ Drainage area, in square miles	292	.89
Volume at primary spillway, in acre-feet	$y = 10.71 X$ Primary spillway surface area, in acres	463	.81
Lower Marais des Cygnes River Basin			
Contributing drainage area	$y = 0.0037 X$ Volume at emergency spillway, in acre-feet	381	0.76
Surface area at emergency spillway, in acres	$y = 3.41 X$ Primary spillway surface area, in acres	360	.82
Volume at emergency spillway, in acre-feet	$y = 1.838 X$ Primary spillway volume, in acre-feet	147	.94
Surface area at primary spillway, in acres	$y = 33.54 X$ Drainage area, in square miles	51	.84
Volume at primary spillway, in acre-feet	$y = 10.71 X$ Primary spillway surface area, in acres	463	.81
Little Osage River Basin			
Contributing drainage area	$y = 0.571 X \ln$ (Volume at emergency spillway, in acre-feet) - 2.17	13	0.88
Surface area at emergency spillway, in acres	$y = 3.41 X$ Primary spillway surface area, in acres	360	.82
Volume at emergency spillway, in acre-feet	$y = 1.22 X$ Primary spillway drainage volume, in acre-feet	18	.97
Surface area at primary spillway, in acres	$y = 36.38 X$ Drainage area, in square miles	6	.86
Volume at primary spillway, in acre-feet	$y = 10.71 X$ Primary spillway surface area, in acres	463	.81
Marmaton River Basin			
Contributing drainage area	$y = 0.0043 X$ (Volume at emergency spillway, in acre-feet)	118	0.93
Surface area at emergency spillway, in acres	$y = 3.41 X$ Primary spillway surface area, in acres	360	.81
Volume at emergency spillway, in acre-feet	$y = 2.18 X$ Primary spillway volume, in acre-feet	142	.68
Surface area at primary spillway, in acres	$y = 7.18 X$ Drainage area, in square miles	108	.62
Volume at primary spillway, in acre-feet	$y = 10.71 X$ Primary spillway surface area, in acres	463	.81

Table 9. Relation between detention volume and median dam height used in estimating missing impoundment design specifications for the upper Osage River Basin models (n = 664).

[<, less than; >, greater than]

Detention volume, in acre feet	Median dam height, in feet
<200	23
200–500	29
500–1,000	33
>1,000	42

The coefficient of determination (r^2) is calculated simply as the square of the correlation coefficient (r). The coefficient of efficiency, E , (Nash and Sutcliffe, 1970) has been widely used to evaluate the performance of hydrologic models and is defined as

$$E = 1 - \frac{\sum_{t=1}^T (q_o^t - \bar{q}_s)^2}{\sum_{t=1}^T (q_o^t - \bar{q}_o)^2} \quad (4)$$

Where q_o and q_s are defined as above for each time step t .

The coefficient ranges from minus infinity (poor model) to 1.0 (perfect model). Legates and McCabe (1999) state that “The coefficient of efficiency represents an improvement over the Coefficient of Determination for model evaluation purposes because it is sensitive to differences in observed and model simulated means and variances...”. Donigian (2002) characterize model “goodness of fit” based on the percent difference in observed and simulated total runoff volume and the coefficient of determination values computed from daily and monthly observed and simulated streamflow values (table 11).

Calibration and validation also were conducted at two sites (Marais des Cygnes River near Reading, Kansas and Dragoon Creek near Burlingame, Kansas) upstream from U.S. Army Corps of Engineers reservoirs, although these areas were not used in the current (2005) and proposed model scenarios because observed reservoir outflows were available and used instead. These two basins were calibrated and validated for the development of parameters in the pre-settlement scenario and to confirm spatial trends in model parameters; that is, to confirm differences in parameter values with the size and location of the simulated subbasins.

Coon and Johnson (2005) and Laroche and others (1996) provide a list of parameters that were found to most strongly affect the hydrologic response of a HSPF model; that is, those parameters to which the model was most sensitive. These include the ground-water recession constant (AGWRC), which controls the rate at which ground water drains from the land); infiltration equation exponent (INFEXP), which controls the rate of infiltration decrease as a function of increasing soil moisture; the ratio of maximum to mean infiltration capacities (INFILD); index to mean soil-infiltration rate (INFILT); interflow-inflow parameter (INTFW), which controls the amount of infiltrated water that becomes shallow subsurface

flow; interflow recession constant (IRC); ground-water flow parameter used to describe non-linear groundwater recession rate (KVARY); lower zone evapotranspiration (LZETP), which represents the density of deep rooted vegetation that conveys water from the unsaturated zone to the atmosphere; lower-zone nominal soil moisture storage (LZSN), which is an index to the soil-moisture holding capacity of the unsaturated zone); and upper-zone nominal storage (UZSN), which is an index to the soil-moisture holding capacity of depressions and the surface soil layer). For this study, AGWRC, INFILT, INTFW, LZSN, LZETP, and UZSN were the parameters for which the simulations were observed to be most sensitive. Parameters of lesser sensitivity included IRC and KVARY.

The period of streamflow record used for calibration varied for the 15 streamflow-gaging station locations used in the calibration of the upper Osage River Basin models (eight gages used in the Marais des Cygnes, one in the Little Osage, five in the Marmaton, and one in the Osage River Basin; table 12). Whereas 10 years (1995–2004 water years) of precipitation and temperature data were available for calibration and validation, the actual period of calibration and validation for a particular streamflow-gaging station location varied and was limited by the available observed streamflow record. Whereas years 1995 through 2003 generally were used for calibration of the Marais des Cygnes and Little Osage River models, the calibration period at the Marmaton River streamflow-gaging stations only was about 4 years (generally from 2001 through 2004; table 12) because of limited streamflow record.

Calibration was begun at headwater streamflow-gaging station locations, and then conducted at downstream gages within each model basin. Initial parameter values for the headwater streamflow-gaging station sites were obtained from guidance provided from the U.S. Environmental Protection Agency (2000). This publication provides guidelines for both the typical and maximum working ranges of HSPF model parameters. The parameters generated as a result of these calibrations subsequently were used as the initial estimate of parameter values at downstream locations and adjusted as necessary within the specified working ranges.

Version 2.4 of the USGS software HSPFEXP (Lumb and others, 1994; available at USGS, 2005d) was used for calibrating gaged Marais des Cygnes River headwater basins. Further calibration of these and remaining basins was conducted using WinHSPF (version 2.3; packaged with BASINS 3.1, U.S. Environmental Protection Agency, 2005a) and calculated summary statistics. The use of HSPFEXP was limited to select headwater basins as it utilizes a version of HSPF with a 200 operations limit that was incapable of running the larger models, whereas WinHSPF has a 500 operations limit with greater model size capability. The number of operations in a model is determined by the number of RCHRESs, along with the number of pervious and impervious land-cover types represented in the model. The use of WinHSPF also allowed for a recommended model initialization or “start-up” period (Gutierrez-Magness, 2005) before determining the summary statistics.

Table 10. Impoundment outflow characteristics used for estimating the elevation difference between primary and emergency spillway, primary spillway culvert diameter, and emergency spillway width in the 10-, 50-, and 90th-percentile design scenarios.

[<, less than; >, greater than]

Detention volume, in acre-feet	Elevation difference between primary and emergency spillway, in feet		
	10 th Percentile design	50 th Percentile design	90 th Percentile design
< 100	2	6	8
100–500	2	8	12
> 500	3	10	15

Contributing drainage area, in square miles	Primary spillway culvert diameter, in inches		
	10 th Percentile design	50 th Percentile design	90 th Percentile design
<2	18	18	24
2–4	18	24	30
>4–8	24	30	36
>8–12	30	36	48
>12	36	48	48

Contributing drainage area, in square miles	Emergency spillway width, in feet		
	10 th Percentile design	50 th percentile design	90 th percentile design
<0.65	30	55	105
0.65–5.0	40	70	120
>5.0	275	275	275

During the HSPF calibration procedure it became evident that the largest discrepancies between observed and simulated streamflows occurred with extreme flows for storm peak flows and low flows (figs. 21, 22). Observed hourly flows in the upper Osage River Basin varied from zero to nearly 63,000 ft³/s during water years 1995–2004. The study objectives were such that simulating the full range of flows was of interest and, therefore, no more or less emphasis was placed on calibrating one range of flows at the expense of another. Logarithmic plots of daily observed and simulated streamflows (fig. 21) indicated that timing of simulated runoff generally matched well with observed flows, and that no consistent biases were observed with runoff event size or timing.

The most obvious differences between observed and simulated streamflows in the flow duration plots was an overestimation of low flows at select stations (fig. 22). When attempting to calibrate to streamflows that may reach zero flow, even absolute errors of less than 1 ft³/s can result in substantial percent differences. One possible cause for the overestimation of low flows was the possible erroneous simulation of impoundment outflows. Minor absolute outflow errors of less than 1 ft³/s, when compounded over 10's or 100's of impoundments in a basin, may account for the over-

estimated flows. A consistent and cumulative bias such as this would become obvious when comparing low-flow statistics for basins with (Marais des Cygnes River and Marmaton River Basins) and without (Little Osage River Basin) a substantial numbers of impoundments. The calibration results (table 12), however, did not indicate that substantial differences in the 50-percent lowest flows existed between basins. A comparison of the error in the 50-percent lowest flows at the Marmaton River near Nevada, Missouri, for a current, no impoundment (180 percent) and current, with impoundment scenario (203 percent) indicated similarities in errors despite the addition of impoundments. Adjustment of the parameters controlling low-flow characteristics within recommended ranges (U.S. Environmental Protection Agency, 2000) failed to substantially reduce the low-flow errors, and any minor benefits gained for this specific model fit statistic came at the expense of overall model fit statistics.

The total volume errors in the calibration summary statistics were within recommended targets, but other statistics confirmed discrepancies between observed and simulated extreme flows. Using a target of 10 percent for the error between total observed and simulated runoff volume, all calibrated scenarios were within the limits of a “very good” calibra-

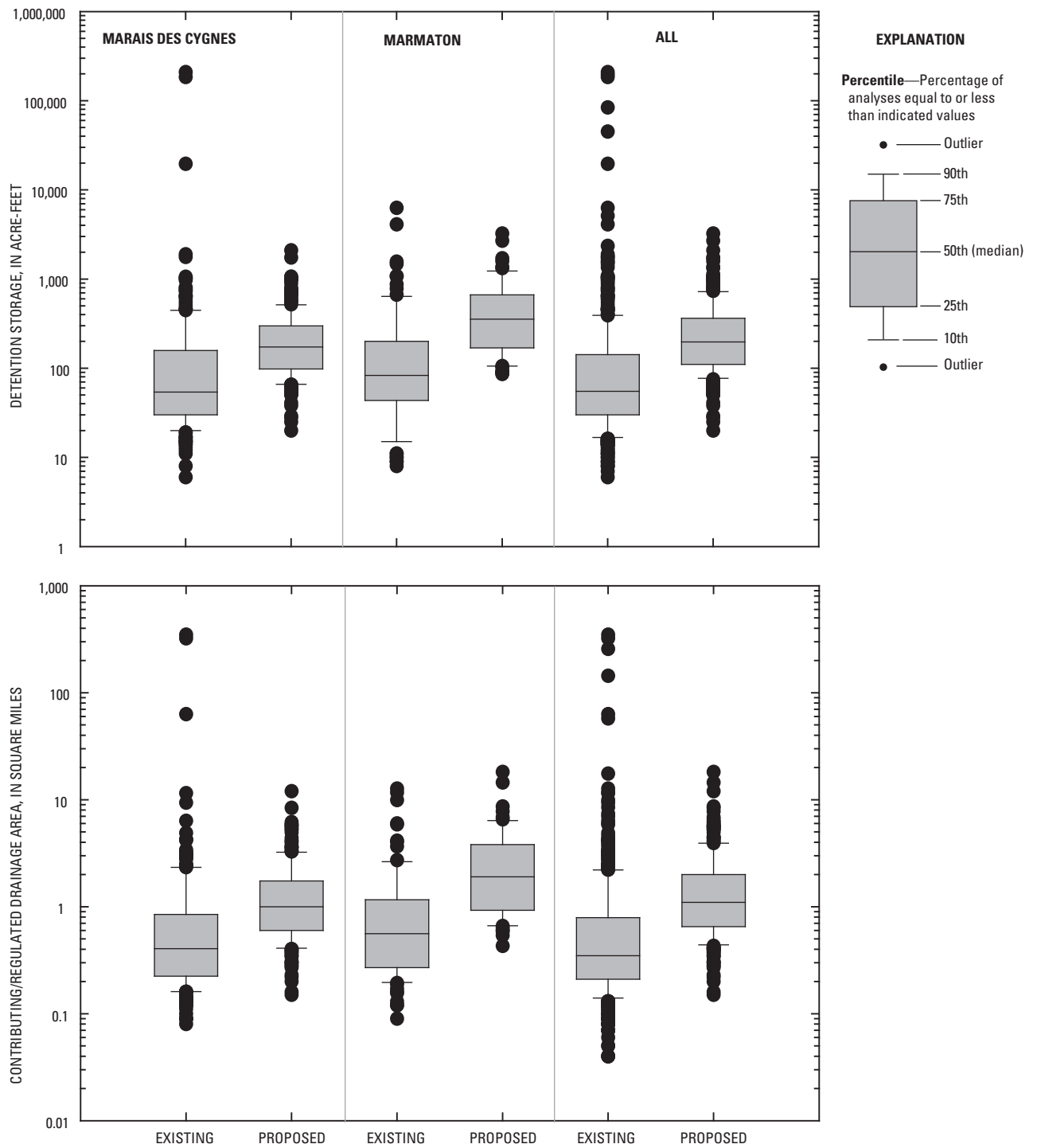


Figure 20. Existing and proposed impoundment detention volume and contributing drainage area for the Marais des Cygnes, Marmaton, and entire upper Osage River basins.

Table 11. Criteria for determining Hydrologic Simulation Program-FORTRAN calibration and validation goodness of fit (modified from Donigan, 2002).

[<, less than; >, greater than]

	Very Good	Good	Fair
Percent difference between observed and simulated monthly or annual values			
Total runoff volume	<10	10–15	15–25
Coefficient of determination			
Daily streamflow	>0.80	0.70–0.80	0.60–0.70
Monthly streamflow	>.85	.75–.85	.65–.75

tion (tables 11, 12). As indicated in the flow-duration curves, however, the error in “50-percent lowest flows” exceeded the error target of 15 percent at all sites. The storm peaks volume error also exceeded the target of 15 percent at 11 of 15 sites (table 12). The exceptions were at streamflow-gaging station sites near the Kansas-Missouri state line including the Marais des Cygnes River at Lacygne, Kansas (station 06915800; fig. 12); Marais des Cygnes River near Kansas-Missouri state line, Kansas (station 06916600; fig. 12); Little Osage River at Fulton, Kansas (station 06917000; fig. 13); and the Marmaton River near Nevada, Missouri (station 06918060; fig. 13). In some cases, the magnitude of the peak was similar, but the timing was such that the daily volume differences exceeded the threshold. Whereas the storm peaks volume typically was underestimated, the storm volume and summer storm volume typically were within the 15 percent error target (table 12).

Using the overall model fit summary statistics including the coefficient of determination (r^2), and model fit efficiency (E), the daily flow values were rated “very good” at five sites, “good” at four sites, “fair” at four sites, and “poor” at two locations (Osage River above Schell City, Missouri (USGS streamflow gaging station 06918070); Little Osage River at Fulton, Kansas (USGS streamflow gaging station 06917000; tables 11, 13). Monthly coefficient of determination values were rated “very good” for 14 of the 15 sites and “good” for the remaining site location [East Drywood Creek at Prairie State Park, Missouri (USGS streamflow gaging station 06917630)]. Model fit efficiency, E , correlated well with the coefficient of determination, although generally was 0.01–0.02 units lower. The daily calibration values of model fit efficiency of 0.38 to 0.88, and monthly calibration values of 0.73 to 0.98 were within published limits of other HSPF studies (Coon and others, 2005).

Every attempt was made during calibration to not only maximize the accuracy of simulated high and low flows, but also to maximize the target hydrograph components (those shown in table 12) and the overall model fit statistics (r , r^2 , E ; table 13). During calibration, additional improvements in the hydrograph summary statistics (for example the “lowest 50 percent flows” or “storm peak volumes”) were minimal,

and any additional improvement in individual hydrograph statistics were at the expense of the overall model fit statistics. The calibration results represent a balance between accurately simulating flow extremes and maximizing the overall model fit statistics.

Most final model parameters (table 14) fell within the typical specified (U.S. Environmental Protection Agency, 2000) working ranges, and all fell within the specified possible ranges. Monthly values were specified for parameters CEPSC (interception storage), NSUR (Manning’s n values for overland flow), and LZETP (lower zone evapotranspiration) to simulate changes in vegetation growth and density (table 14). There was an increase in the values of parameters LZSN, INFILT, INTFW, IRC and AGWRC with increasing drainage area. This can be explained by physical changes in the basin associated with increasing drainage area, including more alluvial deposits, along with lower channel gradients.

Validation is the process of obtaining assurance that the calibrated model provides a reasonable representation of the modeled system. This was accomplished by comparing the simulation results to an independent observed data set outside the calibration period. The qualitative and quantitative tools used to assess the quality of the calibration also were used to determine the merit of the validation results. For those streamflow-gaging stations with greater than 5 years of record available during the 1995–2004 water years, 1 year of record was selected for validation. For the calibrated model to reasonably simulate the validation period, it was necessary that the streamflow conditions in the validation period fell within the range of streamflow conditions occurring during the calibration period. Generally, a single “normal” year of record, either at the beginning or end of 1995–2004, was used for validation. Four stations [Marmaton River near Uniontown, Kansas (USGS streamflow gaging station 06917240); East Drywood at Prairie State Park, Missouri; Drywood Creek near Deerfield, Missouri (USGS streamflow gaging station 06917680); and Marmaton River below Nevada, Missouri] had less than 5 years of total record and, therefore, all of the available streamflow record was used in the calibration process.

Five of the 11 sites used in the validation analyses had total volume errors of less than 10 percent (table 12) corresponding to a “very good” classification (table 11). Of the six remaining sites, four were considered “fair” or better and two [Dragoon Creek near Burlingame, Kansas (USGS streamflow gaging station 06911900) and Pottawatomie Creek near Garnett, Kansas (USGS streamflow gaging station 06914000)] were rated “poor”. The model-fit parameters at 9 of the 11 sites had a daily r^2 of 0.6 or better, placing them in the “fair” or better category, whereas the Little Osage River at Fulton, Kansas, and Osage River above Schell City, Missouri, sites were in the “poor” category (tables 11, 13). The monthly r^2 values were rated “very good” at 10 of 11 validation sites, and the remaining site (Dragoon Creek near Burlingame, Kansas) was rated “good”. Validation model fit efficiency values varied from 0.13 to 0.89 for daily values and from 0.67 to 1.00 using monthly flow values (table 13).

Table 12. Summary statistics for the upper Osage River Basin model calibration and validation.

[-, no data]

Calibration criterion	Marais des Cygnes River near Reading, Kansas ^a (06910800)			Salt Creek near Lyndon, Kansas (06911500)			Dragoon Creek near Burlingame, Kansas ^a (06911900)			Marais des Cygnes River near Pomona, Kansas ^a (06913000)		
	Target value	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 10/95-9/99	Validation period 10/94-9/95	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 10/94-9/03	Validation period 10/03-9/04	
	Error in total volume (percent)	10.0	4.48	16.7	8.64	19.1	-4.91	40.1	3.44	8.26		
Error in low-flow recession (dimensionless)	.01	-.001	.008	.034	.046	.002	.024	0	.013			
Error in 50-percent lowest flows (percent)	10.0	159	137	58.5	271	61.0	112	29.5	48.8			
Error in 10-percent highest flows (percent)	15.0	-8.02	-3.52	-1.35	2.37	-14.5	27.4	-7.93	-2.08			
Error in storm volume (percent)	15.0	-13.1	-12.0	1.39	-10.3	-23.0	6.74	2.54	-1.10			
Seasonal volume error (percent)	10.0	5.17	23.7	37.7	582	25.1	34.2	1.19	50.7			
Summer storm volume error (percent)	15.0	-8.68	-1.95	56.2	-6.12	-6.23	12.7	-9.49	.21			
Storm peaks volume error (percent)	15.0	-57.7	-61.8	-50.3	-50.2	-63.2	-46.1	-37.6	-23.4			
Calibration criterion	Marais des Cygnes River near Ottawa, Kansas ^a (06913500)			Pottawatomie Creek near Garnett, Kansas ^a (06914000)			Marais des Cygnes River at La Cygne, Kansas ^a (06915800)			Marais des Cygnes near Kansas-Missouri State Line, Kansas ^a (06916600)		
	Target value	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 10/94-9/03	Validation period 10/03-9/04	
	Error in total volume (percent)	10.0	7.18	18.0	7.56	53.3	-2.06	8.55	6.52	8.98		
Error in low-flow recession (dimensionless)	.01	.003	.015	-.046	-.012	.005	.006	0	.008			
Error in 50-percent lowest flows (percent)	10.0	43.5	81.2	958	533	65.7	88.5	86.4	94.2			
Error in 10-percent highest flows (percent)	15.0	-5.84	1.37	-14.5	8.11	-12.5	-8.30	-5.63	-10.1			
Error in storm volume (percent)	15.0	5.72	8.96	-10.7	13.2	-6.96	3.58	13.2	-1.16			
Seasonal volume error (percent)	10.0	1.78	62.1	3.67	221	.215	77.1	5.07	65.6			
Summer storm volume error (percent)	15.0	-1.51	2.86	-8.99	-16.0	.023	-6.21	15.3	-10.5			
Storm peaks volume error (percent)	15.0	-21.7	-24.4	-39.1	-30.1	-4.17	2.06	12.0	18.6			

Table 12. Summary statistics for the upper Osage River Basin model calibration and validation.—Continued

[--, no data]

Calibration criterion	Little Osage River at Fulton, Kansas (06917000)				Marmaton River near Uniontown, Kansas (06917240)		Marmaton River near Marmaton, Kansas (06917380)		East Drywood Creek at Prairie State Park, Missouri (06917630)	
	Target value	Calibration period 10/94–9/03	Validation period 10/03–8/04	Calibration period 4/01–10/03	Validation period	Calibration period 10/94–9/03	Validation period 10/03–10/04	Calibration period 10/02–10/04	Validation period	
Error in total volume (percent)	10.0	5.32	20.7	3.42	--	9.21	-3.45	6.37	--	
Error in low-flow recession (dimensionless)	.01	.002	.005	.020	--	0	.010	0	--	
Error in 50-percent lowest flows (percent)	10.0	144	67.5	77.4	--	199	75.9	655	--	
Error in 10-percent highest flows (percent)	15.0	2.01	18.5	7.00	--	-3.47	-23.8	-5.78	--	
Error in storm volume (percent)	15.0	-7.53	12.3	6.52	--	-1.20	-12.0	-22.1	--	
Seasonal volume error (percent)	10.0	6.34	69.3	88.0	--	22.0	3.94	275	--	
Summer storm volume error (percent)	15.0	14.5	4.18	-3.63	--	-2.05	6.61	157	--	
Storm peaks volume error (percent)	15.0	-4.39	8.78	-27.2	--	-42.4	-61.9	-72.3	--	
Calibration criterion	Drywood Creek near Deerfield, Missouri (06917680)				Marmaton River below Nevada, Missouri (06918060)		Osage River above Schell City, Missouri (06918070)			
	Target value	Calibration period 10/01–10/04	Validation period	Calibration period	Validation period	Calibration period 10/94–9/03	Validation period 10/03–9/04			
Error in total volume (percent)	10.0	4.26	--	5.42	--	1.58	3.26			
Error in low-flow recession (dimensionless)	.01	0	--	0	--	.009	.005			
Error in 50-percent lowest flows (percent)	10.0	174	--	203	--	73.9	42.8			
Error in 10-percent highest flows (percent)	15.0	-22.8	--	-15.9	--	-9.45	-8.40			
Error in storm volume (percent)	15.0	-12.8	--	-5.44	--	-3.50	-3.63			
Seasonal volume error (percent)	10.0	1.69	--	4.65	--	6.62	26.1			
Summer storm volume error (percent)	15.0	1.83	--	-5.98	--	-1.41	-8.35			
Storm peaks volume error (percent)	15.0	-51.0	--	-4.59	--	23.4	46.2			

^aValidation start up 06/03–09/03.

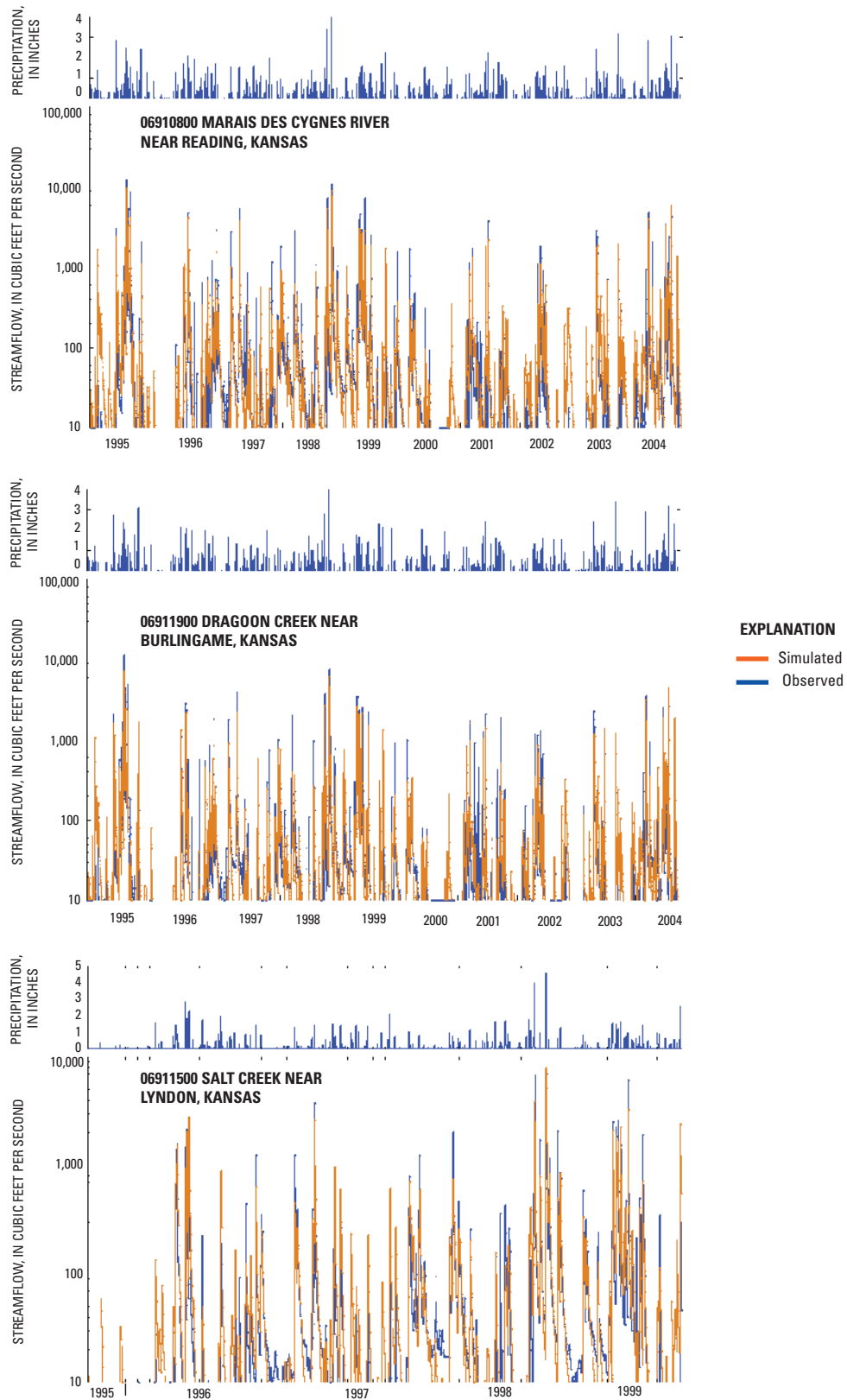


Figure 21. Comparison of simulated and observed daily streamflow at selected streamflow gaging locations in the upper Osage River Basin.

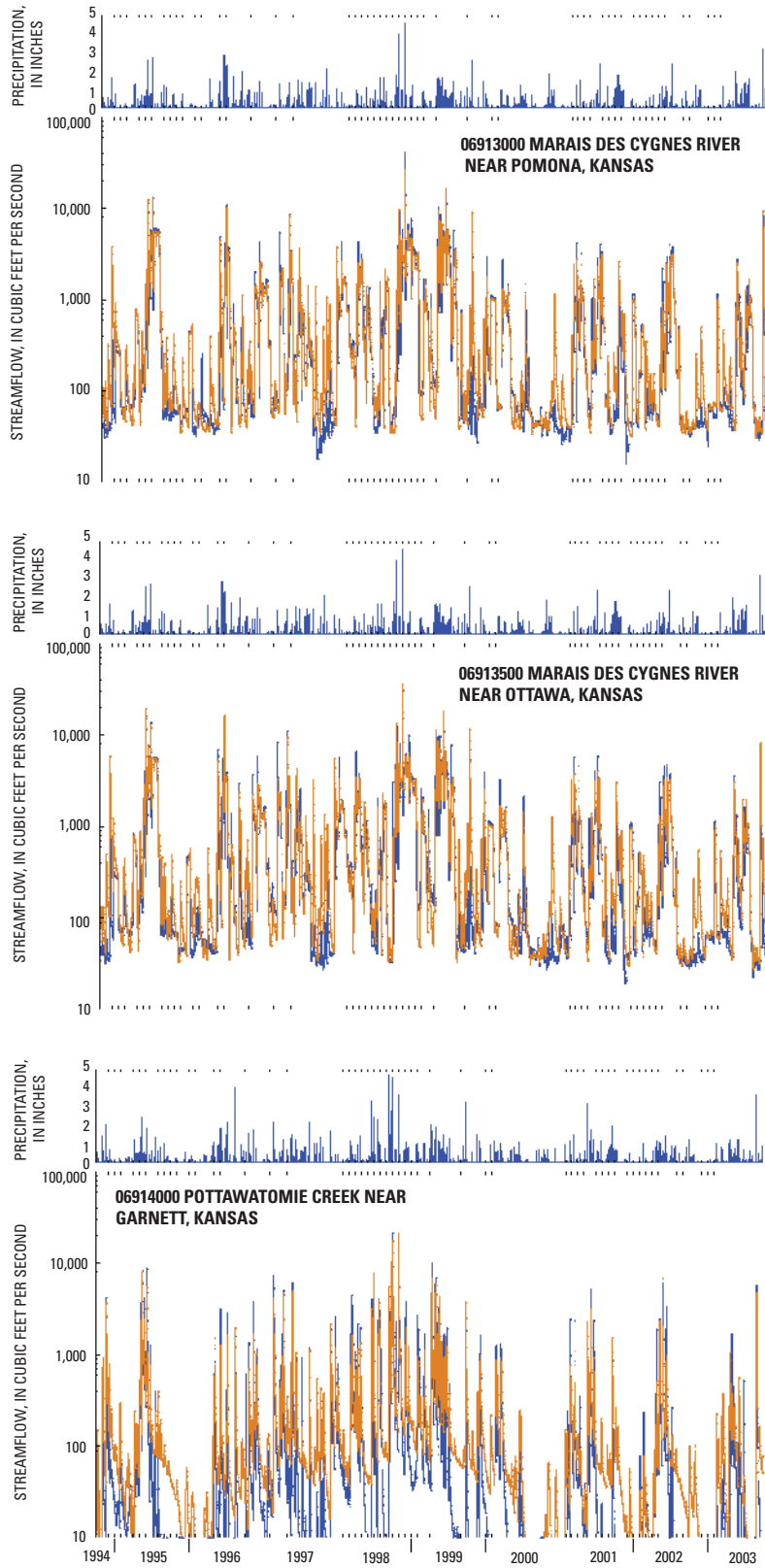


Figure 21. Comparison of simulated and observed daily streamflow at selected streamflow gaging locations in the upper Osage River Basin.—Continued

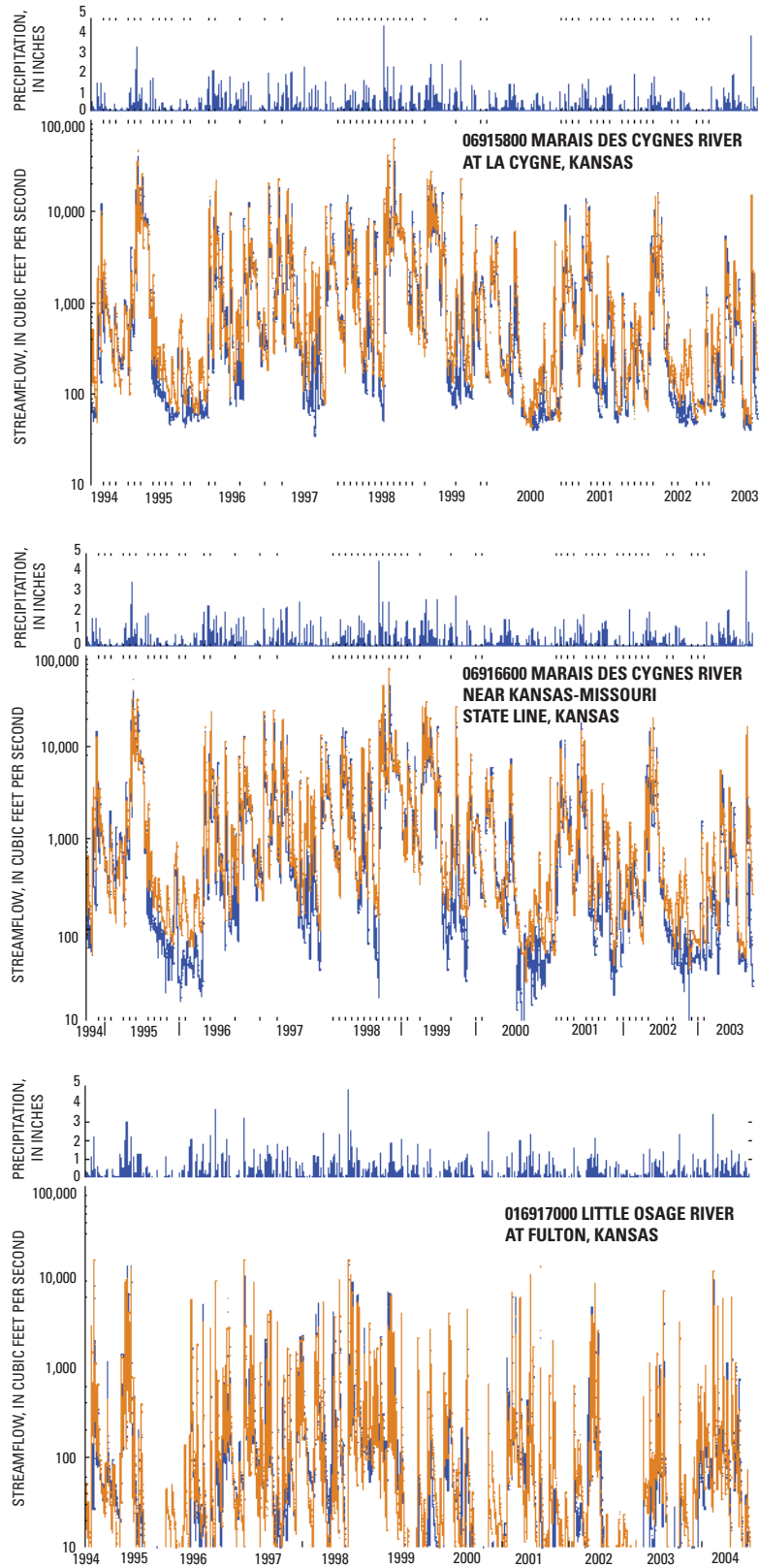


Figure 21. Comparison of simulated and observed daily streamflow at selected streamflow gaging locations in the upper Osage River Basin.—Continued

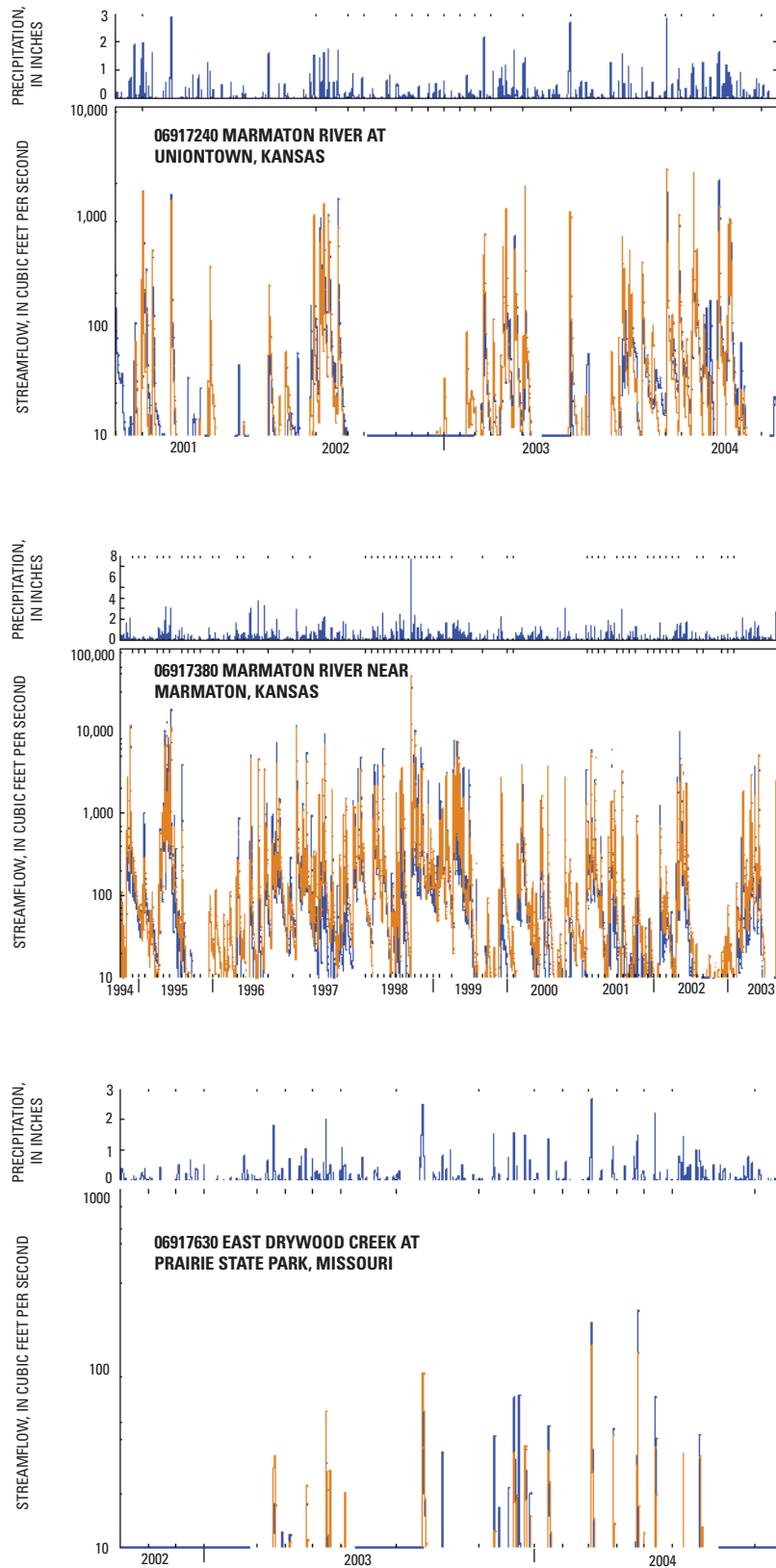


Figure 21. Comparison of simulated and observed daily streamflow at selected streamflow gaging locations in the upper Osage River Basin.—Continued

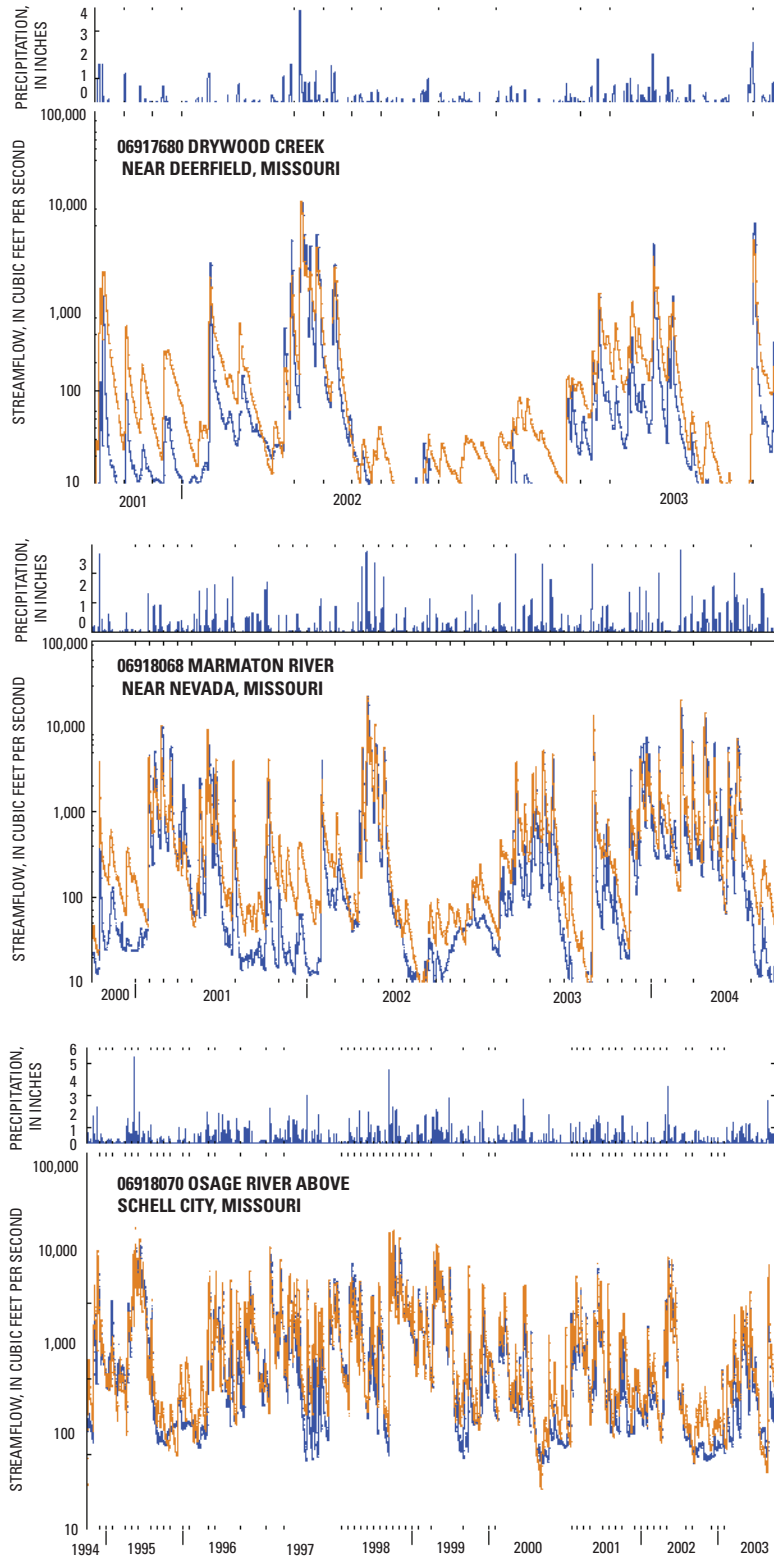


Figure 21. Comparison of simulated and observed daily streamflow at selected streamflow gaging locations in the upper Osage River Basin.—Continued

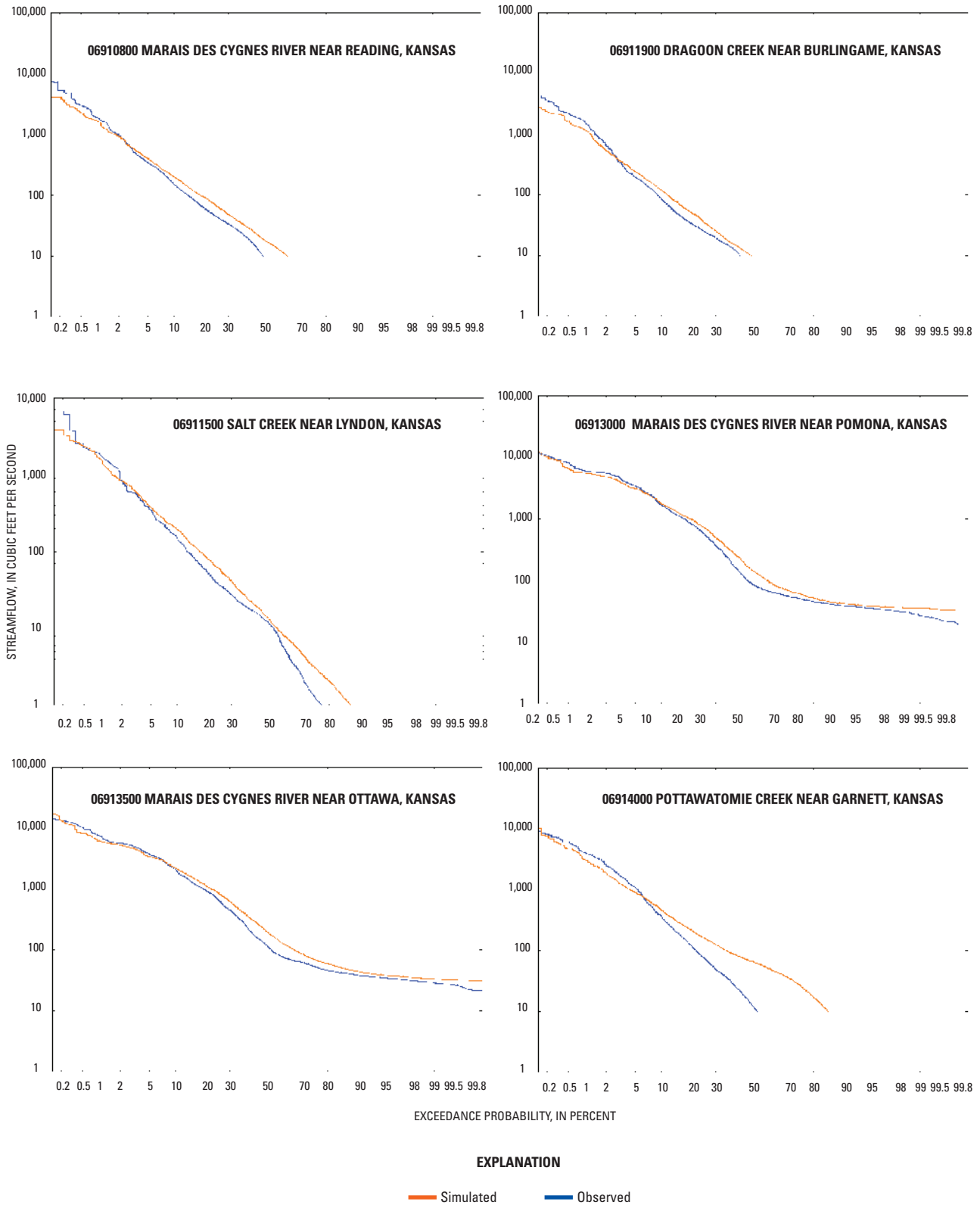


Figure 22. Streamflow-duration distributions for simulated and observed daily streamflows in the upper Osage River Basin.

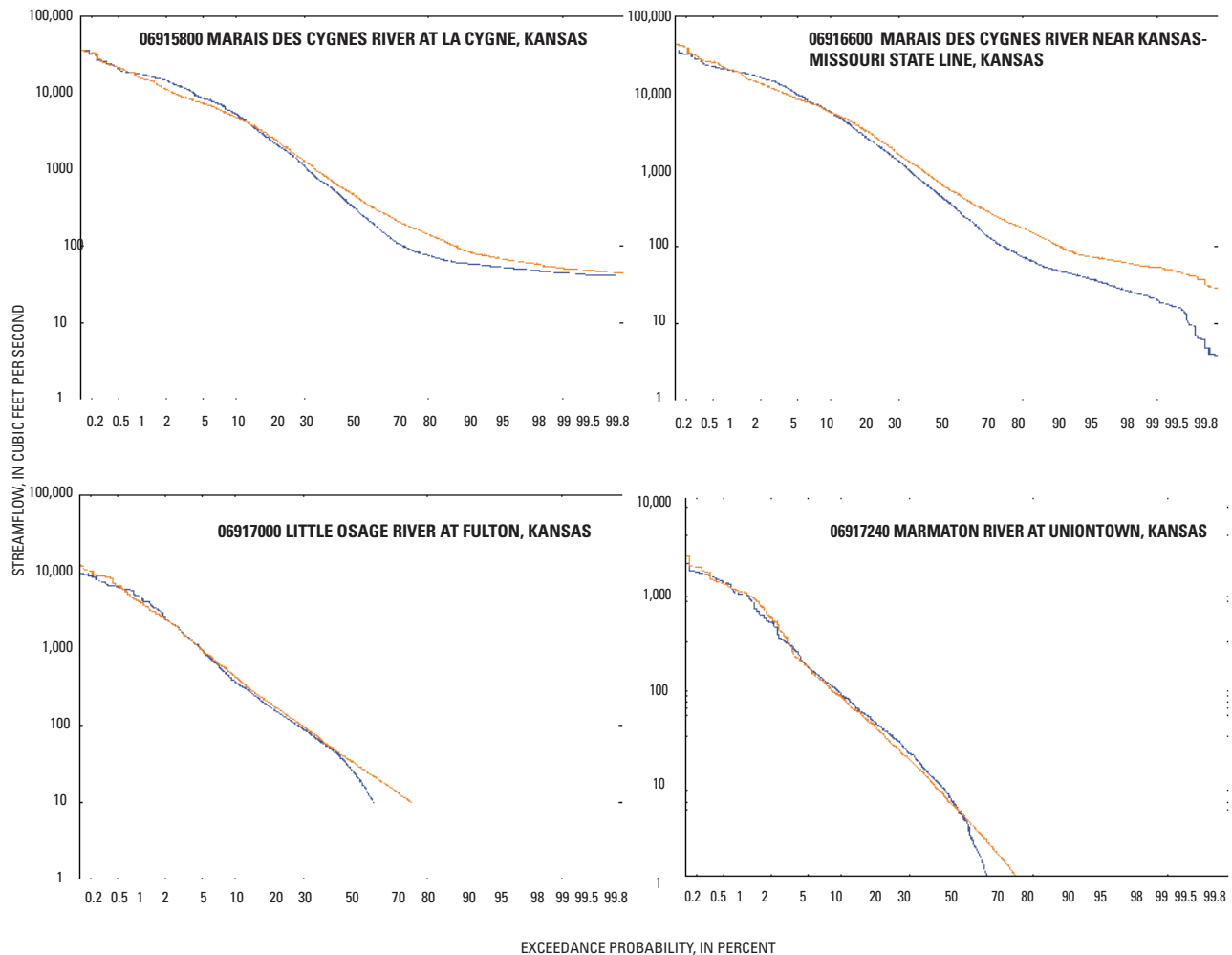


Figure 22. Streamflow-duration distributions for simulated and observed daily streamflows in the upper Osage River Basin.—Continued

Sources of Error and Model Uncertainty

HSPF is used as a simplified numerical representation of a complex and constantly changing natural system. The majority of calibration targets were within specified target criterion, but any model results will include errors as a result of approximations and simplifications. It is important, therefore, to document potential sources of errors, model uncertainties, and any compensatory measures taken to limit these errors in reported results.

Sources of error in the HSPF application included possible non-reported withdrawals; limited spatial definition of input data time series (precipitation, temperature, evapotranspiration, streamflow), lack of temporal changes in land-cover information, initial estimates of hydrologic quantities, and estimates of impoundment design specifications and operation. All reported withdrawals in each of the model basins were included in the simulations; however, the likelihood of non-reported withdrawals remains a potential source of the discrepancy detected between simulated and observed low

flows. During base-flow periods, even small (less than 1 ft³/s) errors in a non-reported withdrawal could cause simulated flows to be substantially different than observed flows, and at these low flows small absolute differences can result in substantial percent differences in the reported calibration results. In such cases, large discrepancies can result between simulated and observed values because of inaccurate input data, despite a reasonable representation of hydrologic processes by the model.

Precipitation likely is the most important input data source in the hydrologic model and limitations in representing the spatial and temporal variability of precipitation may be a primary source of error. Hourly precipitation time series were developed for meteorological segments covering tens or hundreds of square miles, and these time series were developed using data from some precipitation stations that are located outside the segment and were derived from disaggregated daily precipitation data. Gutiérrez-Magness and McCuen (2004) and Gutiérrez-Magness (2005) evaluated several methods used to disaggregate daily precipitation to

Table 13. Daily and monthly model fit statistics for the upper Osage River Basin calibration and validation sites.

[-, no data]

	Marais des Cygnes River near Reading, Kansas (06910800)		Salt Creek near Lyndon, Kansas (06911500)		Dragoon Creek near Burlingame, Kansas (06911900)		Marais des Cygnes River near Pomona, Kansas (06913000)	
	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 10/95-9/99	Validation period 10/94-9/95	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 10/94-9/03	Validation period 10/03-9/04
	Daily							
Coefficient of determination(r^2)	0.80	0.70	0.80	0.92	0.71	0.60	0.87	0.89
Root mean square error	245	279	197	180	206	225	544	361
Model fit efficiency (E)	.78	.69	.79	.89	.69	.49	.87	.89
	Monthly							
Coefficient of determination(r^2)	.92	.95	.94	.93	.92	.82	.99	.99
Root mean square error	61.0	49.6	37.7	53.8	44.4	53.1	157	78.5
Model fit efficiency (E)	.91	.93	.94	.92	.90	.67	.98	.99
	Marais des Cygnes River near Ottawa, Kansas (06913500)		Pottawatomie Creek near Garnett, Kansas (06914000)		Marais des Cygnes River at La Cygne, Kansas (06915800)		Marais des Cygnes near Kansas-Missouri State Line, Kansas (06916600)	
	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 10/94-9/03	Validation period 10/03-04	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 10/94-9/03	Validation period 10/03-9/04
	Daily							
Coefficient of determination(r^2)	0.86	0.89	0.81	0.61	0.68	0.64	0.68	0.62
Root mean square error	651	443	396	433	2,190	2,160	2,420	2,500
Model fit efficiency (E)	.85	.88	.81	.40	.66	.60	.63	.58
	Monthly							
Coefficient of determination(r^2)	.98	.99	.95	.95	.98	.99	.98	.99
Root mean square error	177	138	93.1	106	432	256	433	320
Model fit efficiency (E)	.98	.97	.94	.78	.97	.98	.98	.98

Table 13. Daily and monthly model fit statistics for the upper Osage River Basin calibration and validation sites.—Continued

[--, no data]

	Little Osage River at Fulton, Kansas (06917000)		Marmaton River near Uniontown, Kansas (06917240)		Marmaton River near Marmaton, Kansas (06917380)		East Drywood Creek at Prairie State Park, Missouri (06917630)	
	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 4/01-9/04	Validation period --	Calibration period 10/94-9/03	Validation period 10/03-9/04	Calibration period 10/00-9/04	Validation period --
Daily								
Coefficient of determination(r^2)	0.51	0.47	0.68	--	0.75	0.86	0.71	--
Root mean square error	677	617	101	--	558	347	6.76	--
Model fit efficiency (E)	.38	.13	.60	--	.71	.79	.70	--
Monthly								
Coefficient of determination(r^2)	.86	.97	.87	--	.86	.95	.75	--
Root mean square error	137	55.0	22.5	--	146	62.6	1.81	--
Model fit efficiency (E)	.86	.94	.87	--	.84	.93	.73	--
Daily								
	Drywood Creek near Deerfield, Missouri (06917680)		Marmaton River below Nevada, Missouri (06918060)		Osage River above Schell City, Missouri (06918070)			
	Calibration period 10/01-9/04	Validation period 10/03-9/04	Calibration period 10/00-9/04	Validation period --	Calibration period 10/94-9/03	Validation period 10/03-9/04		
Daily								
Coefficient of determination(r^2)	0.78	--	0.64	--	0.59	0.58		
Root mean square error	494	--	1,060	--	5,280	4,520		
Model fit efficiency (E)	.74	--	.63	--	.53	.50		
Monthly								
Coefficient of determination(r^2)	.93	--	.94	--	.95	.98		
Root mean square error	133	--	259	--	1,450	755		
Model fit efficiency (E)	.90	--	.93	--	.94	.97		

Table 14. Parameters used in the upper Osage River Basin Hydrologic Simulation Program-FORTRAN models.

[LZSN, lower zone nominal soil moisture storage; INFILT, index to infiltration capacity; LSUR, length of overland flow; SLSUR, slope of the overland flow plane; KVARY, variable groundwater recession; AGWRC, base groundwater recession; DEEPFR, fraction of groundwater inflow to deep recharge; BASETP, fraction of remaining evapotranspiration; UZSN, upper zone nominal soil moisture storage; INTFW, interflow inflow parameter; IRC, interflow recession parameter; MON-INTERCEP, monthly interception storage capacity; MON-MANNING, monthly Manning's n (roughness) for overland flow; MON-LZETPARM, monthly lower zone evapotranspiration parameter]

	Marais des Cygnes River near Reading, Kansas (06910800)	Salt Creek near Lyndon, Kansas (06911500)	Dragoon Creek near Burlingame, Kansas (06911900)	Marais des Cygnes River near Pomona, Kansas (06913000)	Marais des Cygnes River near Ottawa, Kansas (06913500)
HSPF Parameters					
LZSN (in.)	7.00	7.00	7.00	7.00	7.00
INFILT (in./hr)	.017	.018	.017	.03	.03
LSUR (ft)	300	300	300	300	300
SLSUR (ft/ft)	.038	.038	.038	.038	.038
KVARY (1/in.)	3.00	3.00	3.00	3.00	3.00
AGWRC (dimensionless)	.95-.98	.89-.98	.93-.98	.90-.98	.90-.98
DEEPFR (dimensionless)	.10	.10	.10	.10	.10
BASETP (dimensionless)	.15	.15	.15	.15	.15
UZSN (in.)	1.10	.90	1.09	.90	1.00
INTFW (dimensionless)	3.00	3.00	3.00	3.00	4.00
IRC (dimensionless)	.30	.30	.30	.50	.60
MON-INTERCEP (in.)	.10-.30	.10-.30	.10-.30	.10-.30	.10-.30
MON-MANNING (dimensionless)	.20-.35	.20-.35	.20-.35	.20-.35	.20-.35
MON-LZETPARM (dimensionless)	.20-.90	.20-.90	.20-.90	.20-.90	.20-.90
HSPF Parameters					
LZSN (in.)	7.00	7.00-9.00	9.00	9.00	10.0
INFILT (in./hr)	.02	.04	.04	.04	.04
LSUR (ft)	300	300	300	300	300
SLSUR (ft/ft)	.038	.038	.038	.034	.034
KVARY (1/in.)	3.00	3.00	3.00	3.00	3.00
AGWRC (dimensionless)	.85-.98	.90-.98	.90-.98	.91-.98	.97-.98
DEEPFR (dimensionless)	.10	.10	.14	.14	.18
BASETP (dimensionless)	.15	.15	.15	.15	.15
UZSN (in.)	1.20	1.20	1.20	1.20	1.20
INTFW (dimensionless)	4.00	5.00	5.00	5.00	5.00
IRC (dimensionless)	.30	.70	.70	.70	.70
MON-INTERCEP (in.)	.10-.30	.10-.30	.10-.30	.10-.30	.10-.30
MON-MANNING (dimensionless)	.20-.35	.20-.35	.20-.35	.20-.35	.20-.35
MON-LZETPARM (dimensionless)	.20-.90	.20-.90	.20-.90	.20-.90	.20-.90
				Marais des Cygnes River near Kansas-Missouri State Line, Kansas (06916600)	Interim Area (between 06916600 and 06918070)

hourly values, and determined limitations with all methods to accurately determine the distribution and intensity of rainfall events. Although several hourly precipitation gages were located in the basin, these primarily were used for disaggregation of daily precipitation values. Maximum and minimum daily temperature information also was estimated for meteorological segments based on stations that may not be located within these segments. The subsequent hourly temperature distributions were estimated based on a sinusoidal relation fit to these two daily extremes. As hourly potential evapotranspiration data were determined from temperature data, these input time series also were subject to the same sources of error as temperature. The USGS hourly streamflow time series used in calibration and validation were estimated using a stage-discharge relation developed from instantaneous measurements. Typically, instantaneous streamflow measurements have associated errors of 5 to 8 percent; this error can account for some of the differences computed between observed and simulated hourly streamflow.

The land-cover information in the current scenarios was developed from 1992–1995 data, and may not fully represent conditions during the entire study period (1995 through 2004). The land-cover information also was assumed to be static during the 10-year simulation period. Land-cover information for the pre-development scenarios was estimated using simplified constructed maps of historical land-cover information with corresponding parameter characteristics obtained from current calibrations. The historical wetland area was estimated using the 1995 wetland area and, therefore, the historical wetland area may be under-represented. The current channel geometry was used in developing and comparing the pre-settlement, current, and proposed hydrologic conditions. Anthropogenic alterations of the Little Osage and Marmaton River channels have been minimal; however, there have been extensive modifications to the lower Marais des Cygnes River channel that could affect such comparisons.

Errors in model scenarios can result from erroneous initial condition estimates. Whereas certain parameters may reach equilibration within a few model iterations (hours), others may take weeks or months to equilibrate. Two methods were employed to minimize the effects of erroneous initial conditions in the study simulations. The first was to incorporate a model “start-up” period before the simulation period of interest. This requires the availability of additional meteorological data outside the simulated period of interest. Gutierrez-Magness (2005) determined that by incorporating a start-up period of “about a year” before the calibration or validation period that associated errors were minimized. Available streamflow and meteorologic data allowed for an approximate 3-month start-up period in the validation record at select gaging stations. A second approach used in minimizing errors attributable to initial parameter conditions was to use computed ending values for the specified initial conditions as a starting point for subsequent simulation runs in an iterative estimate approach.

The number, distribution, and design specifications of impoundments were additional potential sources of error in the numerical simulations. The estimate of reservoir numbers in the basin provided by the National Inventory of Dams (U.S. Army Corps of Engineers, 2005) augmented with state-permitted impoundments (Joe File, Kansas Department of Agriculture, Division of Water Resources, written commun., 2005) may underestimate the total number of impoundments in the basin. Smith and others (2002) determined that the National Inventory of Dams underestimated the number of impoundments in selected basins when compared to Geographic Information System analyses of the areas. Therefore, simulations of the possible effects of impoundments used in this study are likely conservative estimates of the actual effects of impoundments on streamflows in the upper Osage River Basin. The impoundment location information was accurate enough to assign an impoundment to a RCHRES, but it was not possible to determine the actual overlap of contributing drainage area, or “chaining,” between impoundments. All impoundments were assumed to have independent contributing drainage areas unless the total drainage area associated with the impoundments in a RCHRES exceeded the total drainage area of that RCHRES. In this case, the reservoirs were “chained” until the amount of regulated area was within the total RCHRES area. Whereas reservoir storage was assumed to be static, in actuality, the elevation-storage and elevation-area relations of impoundments are constantly changing as a result of sedimentation. The effects of possible errors resulting from erroneous design specification estimates was quantified using multiple scenario runs under a range of varying design specifications (10th-, 50th-, and 90th-percentiles in input design characteristics representing varying levels of reservoir detention as described in the “Stage-area and Stage-volume outflow relations” section of this report) from the distribution of possible design variables including spillway width, culvert diameter, and primary-secondary spillway elevation differences.

Uncertainty in the model results differs with the time interval, hydrologic characteristic of interest, and reporting location. Results for longer duration times will be more accurate than for shorter times, and, therefore, the 10-year, annual, or monthly simulation results are more accurate than daily or hourly values as various sources of errors, rather than compounding, would tend to “cancel out” over longer times. The calibration results indicated that the peak streamflow simulations were most accurate near the Kansas-Missouri state line (table 12) in each primary river basin and, therefore, summarizations of high flows are limited to these reporting locations.

Techniques used to limit the uncertainty in model results, particularly in simulated low flows, included adjusting simulated time series with observed flows at gaged reporting locations, and presenting results in terms of relative differences between current-simulated (or observed, if streamflow data available) and proposed scenarios. To provide more accurate summarization of streamflow results at gaged reporting locations, the simulated pre-settlement and proposed streamflow results were adjusted with observed values to better deter-

mine the relative differences between simulated scenarios. This adjustment procedure is discussed further in the “Model Scenarios” section of this report. Potential errors in simulated flow values also are limited by presenting results in terms of relative differences. Relative differences in proposed results isolate differences as a result of the addition of impoundments, and the range of uncertainties in impoundment design are quantified using the 10th-, 50th- and 90th-percentile of estimated input design information.

Despite errors and uncertainty in all numerical simulations as a result of simplifications, HSPF and the associated calibration and validation techniques represent the “state of the science” in determining basin-scale hydrologic processes—including the possible effects of impoundments and land-cover changes on streamflows. Poor calibration results for low flows limit the direct use of non-adjusted results; however, the general trends (did flows increase or decrease as a result of this simulated change?), and relative quantitative differences between the pre-settlement or proposed simulations and current conditions are reasonable and useful products from the model.

Model Scenarios

Pre-settlement, current, and proposed impoundment scenarios were developed for the Marais des Cygnes, Little Osage, and Marmaton River model areas (fig. 10). The same 1995–2004 meteorologic inputs were used for all scenarios, but land cover, impoundments, and point-source withdrawals and discharges were varied (table 15). The pre-settlement scenario represented historical land cover (fig. 4), and did not include impoundments or point-source withdrawals and discharges. The hydrologic properties of native prairie soils have been determined to have an order of magnitude greater hydraulic conductivity compared with cultivated soils (Fuentes and others, 2004). The INFILT value for prairie/rangeland land cover from the pre-settlement scenario was, therefore, doubled (compared with agriculture, pasture, and rangeland INFILT values from the current and proposed scenarios) in an attempt to estimate the effects of the probable greater hydraulic conductivity under pre-settlement conditions. Current scenarios included 1992–1995 land cover, existing impoundments (2005), and 1995–2004 reported point-source withdrawals and discharges. A proposed scenario, (Prop-all), was developed for the Marais des Cygnes River that included all existing and proposed impoundments, 1992–1995 land cover, and current point-source withdrawals and discharges. Two proposed scenarios were developed for the Marmaton River Basin; one containing a select group of four soon-to-be-constructed impoundments in addition to all existing impoundments (Prop-sel), and another that included existing and all proposed impoundments (Prop-all). Both proposed Marmaton River Basin scenarios (Prop-sel, Prop-all) included 1992–1995 land cover and current point-source withdrawals and discharges. No

proposed scenario was developed for the Little Osage River as no additional impoundments were proposed for this basin.

The Prop-sel simulation for the Marmaton and the Prop-all simulations for the Marais des Cygnes and Marmaton River Basins each included three different scenarios accounting for possible variability in impoundment design. As described previously in the “Stage-area and Stage-volume outflow relations” section of this report, three outflow ratings (10th-, 50th-, and 90th-percentile outflow characteristics) were constructed for each proposed impoundment with any missing design variables. In this way the variability in streamflow as a result of design uncertainty could be quantified along with the applicability of using proposed impoundment design as a means of controlling the cumulative effects of impoundments.

To obtain the most accurate simulated streamflow possible, at locations where observed streamflow data were available, each value in the hourly or daily simulated-scenario time series was multiplied by the corresponding ratio of hourly or daily observed streamflow to simulated current flows to adjust for possible biases associated with the simulations. The relative differences between the observed and simulated scenarios, as a result of modifications for proposed or pre-settlement conditions, were quantified and isolated, and these differences were applied to the simulated scenario (pre-settlement or proposed) flows. For example, the differences between pre-settlement streamflow and current simulation results were considered proportional to the differences between a new pre-settlement adjusted time series and the observed streamflows as follows:

$$\frac{Pre + infilt}{Current} = \frac{Pre + infilt_{adj}}{Observed} \quad (5)$$

or alternatively,

$$Pre + infilt_{adj} = \frac{(Pre + infilt \times Observed)}{Current} \quad (6)$$

where

- $Pre + infilt_{adj}$ is the adjusted daily or hourly pre-settlement streamflow time series with an INFILT parameter value double that used in the Current scenario,
- $Pre + infilt$ is the simulated daily or hourly pre-settlement streamflow time series with an INFILT parameter double that used in the Current scenario,
- $Current$ is the simulated daily or hourly streamflow values for the 1995 land-cover and current (2005) impoundment conditions, and
- $Observed$ is the observed daily or hourly streamflow values at USGS streamflow-gaging stations.

At those locations where observed values were available, the adjusted simulation scenario values are presented, other-

Table 15. Upper Osage River Basin model scenario characteristics.

[ft, feet ; <, less than; >, greater than; --, not applicable]

Scenario ^a	Land cover	Impoundments	Detention volume, in acre-ft	Withdrawals	Discharges	Spillway width estimate, in ft	Primary-secondary spillway differential, in ft		
							detention<100 acre-ft	detention 100-500 acre-ft	detention>500 acre-ft
Marais des Cygnes River Basin									
Pre+Infiltr	Historical	0	0	None	None	--	--	--	--
Current	Current	407	^b 24,620	Yes	Yes	--	--	--	--
Prop-all									
50 percentile	Current	626	^b 69,428	Yes	Yes	70	6	8	10
10 percentile	Current	626	^b 69,428	Yes	Yes	40	2	2	3
90 percentile	Current	626	^b 69,428	Yes	Yes	120	8	12	15
Little Osage River Basin									
Pre+Infiltr	Historical	0	0	None	None	--	--	--	--
Current	Current	29	3,425	Yes	Yes	--	--	--	--
Marmaton River Basin									
Pre+Infiltr	Historical	0	0	None	None	--	--	--	--
Current	Current	103	30,268	Yes	Yes	--	--	--	--
Prop-sel									
50 percentile	Current	107	37,288	Yes	Yes	70	6	8	10
10 percentile	Current	107	37,288	Yes	Yes	40	2	2	3
90 percentile	Current	107	37,288	Yes	Yes	120	8	12	15
Prop-all									
50 percentile	Current	167	65,365	Yes	Yes	70	6	8	10
10 percentile	Current	167	65,365	Yes	Yes	40	2	2	3
90 percentile	Current	167	65,365	Yes	Yes	120	8	12	15

^aSee Glossary at front of report for abbreviations used in report.^bCurrent simulation scenario excludes area above Pomona, Melvern, Hillsdale, and La Cygne Lakes with detention volume of 862,600 acre-ft.

wise simulated values are presented. Values were adjusted for six of the nine reporting locations used in this report (table 16). Other simulated or adjusted streamflow-time series generated in this study included:

<i>Prop-sel50</i>	Simulated proposed streamflow time series using selected proposed impoundments and median values as estimates for any missing impoundment design criteria.
<i>Prop-sel10</i>	Simulated proposed streamflow time series using selected proposed impoundments and 10 th -percentile values as estimates for any missing impoundment design criteria.
<i>Prop-sel90</i>	Simulated proposed streamflow time series using selected proposed impoundments and 90 th -percentile values as estimates for any missing impoundment design criteria.
<i>Prop-sel50_{adj}</i>	Adjusted proposed streamflow time series for selected proposed impoundments using median estimates for any missing impoundment design criteria.
<i>Prop-sel10_{adj}</i>	Adjusted proposed streamflow time series for selected proposed impoundments using 10 th -percentile estimates for any missing impoundment design criteria.
<i>Prop-sel90_{adj}</i>	Adjusted proposed streamflow time series for selected proposed impoundments using 90 th -percentile estimates for any missing impoundment design criteria.
<i>Prop-all50</i>	Simulated proposed streamflow time series using all proposed impoundments and median estimates for any missing impoundment design criteria.
<i>Prop-all10</i>	Simulated proposed streamflow time series using all proposed impoundments and 10 th -percentile estimates for any missing impoundment design criteria.
<i>Prop-all90</i>	Simulated proposed streamflow time series using all proposed impoundments and 90 th -percentile estimates for any missing impoundment design criteria.
<i>Prop-all50_{adj}</i>	Adjusted proposed streamflow time series using all proposed impoundments using median estimates for any missing impoundment design criteria.
<i>Prop-all10_{adj}</i>	Adjusted proposed streamflow time series using all proposed impoundments using 10 th -percentile values as estimates for any missing impoundment design criteria.
<i>Prop-all90_{adj}</i>	Adjusted proposed streamflow time series using all proposed impoundments using 90 th -percentile values as estimates for any missing impoundment design criteria.

Comparisons of the magnitude, frequency, duration, and timing of the various streamflow time series were conducted using the Indicators of Hydrologic Alteration analysis software (The Nature Conservancy, 2005).

Computation of Fish Habitat Area under Simulated Streamflow Scenarios

The streamflow time series resulting from the various hydrologic simulations were used in conjunction with existing developed streamflow-habitat area time series to compare and quantify the effects of altered streamflow on fish-habitat area availability. Streamflow-fish habitat area relations previously were developed for 26 species/life stage categories (Heimann and others, 2005) at three near-state-line locations on the Marais des Cygnes River (RCHRES 90, RCHRES 93, RCHRES 95; fig 12), and two near-state-line locations on the Marmaton River (RCHRES 6, RCHRES 11; fig. 13. Daily streamflow time series were used to develop selected daily fish-habitat area time series for 9 of the 26 categories (table 17) for this study. These nine categories were selected to represent a variety of seasonal conditions, and because of the associated level of vulnerability with these categories as a result of spawning or juvenile development concerns (Heimann and others, 2005).

Effects of Impoundments and Land-Cover Changes on Streamflows

Hydrologic simulations were developed for the upper Osage River Basin for the 1995 through 2004 water years using land cover, impoundment, and withdrawal/discharge information representing pre-settlement, current, and proposed conditions. Comparisons of streamflow conditions were conducted using a generalized water balance along with selected ecological flow characteristics (magnitude, frequency, duration, and timing) for the different simulation scenarios. The simulated streamflow time series also were used to compute fish habitat area at selected Marais des Cygnes and Marmaton River locations near the Kansas-Missouri state line for selected fish species/life stages to quantify the effects of simulated flow alterations on stream channel habitat. Comparisons are made between simulated proposed conditions and observed record when streamflow record is available, but if streamflow record is not available at a reporting location the comparisons are made to simulated current conditions.

Water Balance

Runoff volume and actual evapotranspiration statistics for the Marmaton River Basin (1995 through 2004 water years), provide an indication of the effects of land-cover and regulation changes on the water balance in the study basins (table 18). The addition of impoundments under proposed conditions had no substantial effect on the 10-year total runoff as current, no-impoundment; current, with-impoundment; and proposed impoundment scenarios all had 7,770 thousand acre-feet of total runoff during the 10-year simulation period (table 18).

Table 16. Model scenarios for reporting locations used in study.

	Model Scenario ^a						Simulated results adjusted with observed record			
	Pre+Infiltr	Current	Observed	Prop-sel10	Prop-sel50	Prop-sel90	Prop-all10	Prop-all50	Prop-all90	
Marais des Cygnes River										
Marais des Cygnes River near the Kansas-Missouri state line, Kansas (RCHRES 88) ^b	Yes	Yes	Yes	No	No	No	Yes	Yes	Yes	Yes
Marais des Cygnes River near the Kansas-Missouri state line, Kansas (RCHRES 90) ^b	Yes	Yes	Yes ^c	No	No	No	Yes	Yes	Yes	Yes
Marais des Cygnes River near the Kansas-Missouri state line, Kansas (RCHRES 93) ^b	Yes	Yes	Yes ^c	No	No	No	Yes	Yes	Yes	Yes
Marais des Cygnes River near the Kansas-Missouri state line, Kansas (RCHRES 95) ^b	Yes	Yes	Yes ^c	No	No	No	Yes	Yes	Yes	Yes
Little Osage River										
Little Osage River at Fulton, Kansas (RCHRES 25) ^d	Yes	Yes	Yes	No	No	No	No	No	No	Yes
Marmaton River										
Marmaton River at RCHRES 54, Kansas ^d	Yes	Yes	No	No	No	No	Yes	Yes	Yes	No
Marmaton River at Marmaton, Kansas (RCHRES 58) ^d	Yes	Yes	Yes	Yes	Yes	Yes	Yes	Yes	Yes	Yes
Marmaton River near the Kansas-Missouri state line, Missouri (RCHRES 6) ^d	Yes	Yes	No	Yes	Yes	Yes	Yes	Yes	Yes	No
Marmaton River at RCHRES 11, Missouri ^d	Yes	Yes	No	Yes	Yes	Yes	Yes	Yes	Yes	No

^aSee Glossary at front of report for description of model scenario abbreviations.^bSee figure 12 for location.^cUsed observed record from the state line gaging station (station number 06916600).^dSee figure 13 for location.

Table 17. Selected fish habitat categories used in the determination of habitat area for Marais des Cygnes and Marmaton River study sites (modified from Heimann and others, 2005).

[ID, identification; >, greater than]

Category ID	Species	Life stage	Selected depth, in meters	Selected velocity, in meters per second	Selected substrate	Vulnerability to habitat alterations
				Spring		
1	Suckermouth minnow (<i>Phenacobius mirabilis</i>)	Spawning	0.05–0.30	0.20–0.80	Fines, gravel, pebble, cobble, boulder, bedrock	High
4	Slenderhead darter (<i>Percina phoxocephala</i>)	Spawning	.20–.80	.6–1.0	Cobble	High
7	Paddlefish (<i>Polyodon spathula</i>)	Spawning	>.5	>1.4	Gravel, pebble, cobble	High
				Summer night		
14	Flathead catfish (<i>Pylodictis olivaris</i>)	Juvenile	0.10–0.40	0.20–0.80	Cobble, boulder	High
16	Channel catfish (<i>Ictalurus punctatus</i>)	Juvenile	.30–.50	.80–1.0	Fines, gravel, pebble, cobble, boulder, bedrock	High
19	Stoner cat (<i>Noturus flavus</i>)	Juvenile	.05–.20	.20–1.0	Cobble	High
				Fall		
20	Flathead catfish (<i>Pylodictis olivaris</i>)	Juvenile	0.10–0.40	0.20–0.60	Cobble, boulder	High
22	Channel catfish (<i>Ictalurus punctatus</i>)	Juvenile	.20–.40	.20–.60	Fines, gravel	High
24	Stoner cat (<i>Noturus flavus</i>)	Juvenile	.05–.30	.20–1.0	Fines, gravel, pebble, cobble, boulder	High

The change from historical land cover (pre-settlement using current INFILT) to current land cover (Current, no-impoundments scenario with water use) resulted in a decline of 120 thousand acre-feet of total runoff (table 18) indicating that the combined land-cover changes (100 thousand acre-feet) and net water use losses (20 thousand acre-feet; table 18) may have a greater effect on total runoff than impoundments. The sensitivity of the model to the INFILT parameter was evidenced by the 190 thousand acre-feet decline in total runoff between the Pre-settlement (current INFILT) and Pre+infilt scenarios (table 18); the largest runoff difference between any two scenarios.

Surface runoff was 70 to 90 thousand acre-feet less under the pre-settlement (current INFILT) scenario compared with the current and proposed scenarios, indicating differences resulting from land cover and water use (table 18). The doubling of the INFILT model parameter (index to mean soil infiltration rate) between the Pre-settlement (current INFILT) and Pre+Infilt scenarios resulted in a 640 thousand acre feet difference in surface runoff. Runoff volume for the 50 percent lowest streamflows was the least for the Current, no-impoundment and Pre-settlement (current INFILT) scenarios and greatest for the Prop-all50 and Pre+infilt scenarios. Conversely, streamflow for the 10 percent highest flows was greatest for the Pre-settlement (current INFILT) and Current, no-impoundment scenarios and least for the Prop-all50 and Pre+infilt scenarios (table 18). These differences can be attributed to the effects of impoundments and changes in infiltration on the streamflow hydrograph, rather than to other changes in land cover and water use.

Impoundments decreased hydrograph peaks and extended the recession limb of hydrographs in the simulations (fig. 23), whereas increased infiltration had similar effects. Potential impoundment retention volume (water stored below the primary spillway outflow level), and particularly the impoundment detention volume (water temporarily stored between the primary and emergency spillway outflow levels), increased substantially under proposed scenarios (tables 15, 18; fig. 24). The temporal variability in outflow characteristics for a single impoundment demonstrated the variability associated with proposed impoundment design scenarios and the potential effects on streamflow magnitude (fig. 25). Impoundment outflows for a single, isolated, impoundment under “normal”, “wet”, and “dry” precipitation years indicate that both low flows and high flows may be affected by design characteristics. The 10th-percentile design scenario, representative of smaller outflow structures and longer detention time, resulted in the longest hydrograph recessions and lowest peak flows (fig. 25). The 90th-percentile design scenario, representative of larger possible outflow structures and shorter detention time, resulted in the fastest hydrograph recessions and highest peak flows of the three proposed impoundment design scenarios (fig. 25).

Evapotranspiration losses from the land and open water surfaces varied little between current and proposed scenarios, but could account for the differences in total runoff between pre-settlement (current INFILT) and the current and proposed

Table 18. Water balance summary information for hydrologic simulations of the Marmaton River Basin, water years 1995 through 2004.

[INFILT, infiltration parameter used in Hydrologic Simulation Program--FORTRAN to define infiltration rate]

	Model Scenario					
	Pre-settlement ^a	Pre+INFILT ^b	Current (no impoundments)	Current (with impoundments)	Prop-sel50	Prop-all50
Total runoff, in thousands of acre-feet	7,890	7,700	7,770	7,770	7,770	7,770
Surface runoff, in thousands of acre-feet	2,000	1,360	2,090	2,070	2,070	2,070
Runoff from 50 percent lowest flows, in thousands of acre-feet	427	482	423	465	475	485
Runoff from 10 percent highest flows, in thousands of acre-feet	4,810	4,450	4,740	4,570	4,500	4,300
Actual Evapotranspiration (land surface), in thousands of acre-feet	17,900	17,800	18,000	18,000	18,000	18,000
Free surface evaporation (water surfaces) in thousands of acre-feet	55	57	54	101	108	140
Total retention volume, in thousands of acre-feet	0	0	0	25	26	31
Total detention volume, in thousands of acre-feet	0	0	0	30	37	65
Net Water use losses, in thousands of acre-feet	0	0	20	20	20	20

^aUses INFILT value from Current model scenario.

^bUses 2 x INFILT value from Current model scenario to simulate pre-cultivation conditions.

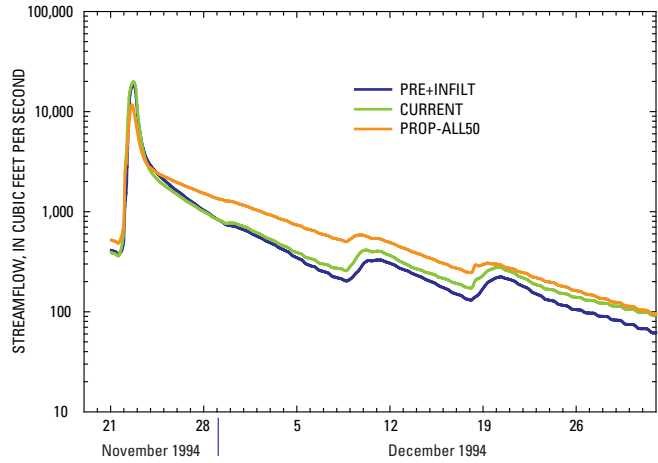


Figure 23. Comparison of streamflow hydrographs, by simulation scenario, at the Marmaton River near the Kansas-Missouri state line, Missouri (RCHRES 6), for November 19, 1994, through January 3, 1995.

scenarios. Evapotranspiration losses from the land surface were 100 thousand acre-feet less under the pre-settlement (current INFILT) scenario than the current and proposed scenarios, corresponding to the difference in total runoff between these scenarios. Whereas evapotranspiration losses were 200 thousand acre-feet less for the Pre+ infilt scenario compared with the current and proposed scenarios, the total runoff was similar and the reduction in evapotranspiration losses could be offset by greater deep aquifer storage (losses) associated with the greater infiltration rates. Evaporative water losses increased with the addition of impoundments, and while these increases did not have a substantial effect on the total runoff (table 18) they could have an effect on flows during dry periods.

Streamflow Magnitude

The magnitude of low flows and high flows in the upper Osage River Basin were compared for the various simulation scenarios to quantify possible differences in these streamflow extremes as a result of proposed and historical conditions. Summary annual and monthly streamflow statistics for selected Marais des Cygnes, Little Osage, and Marmaton River reporting sites are provided in tables 19–28, on compact disc, at the back of this report. Differences in low flows indicated the possible amount of alterations of in-channel habitat, whereas differences in high flows indicated the likelihood of possible ecological consequences of hydrologic regulation on riparian systems in the upper Osage River Basin.

Low Flows

The primary concerns of resource managers regarding alterations in low flows are the possible effects that additional impoundments may have on sustaining low flows and

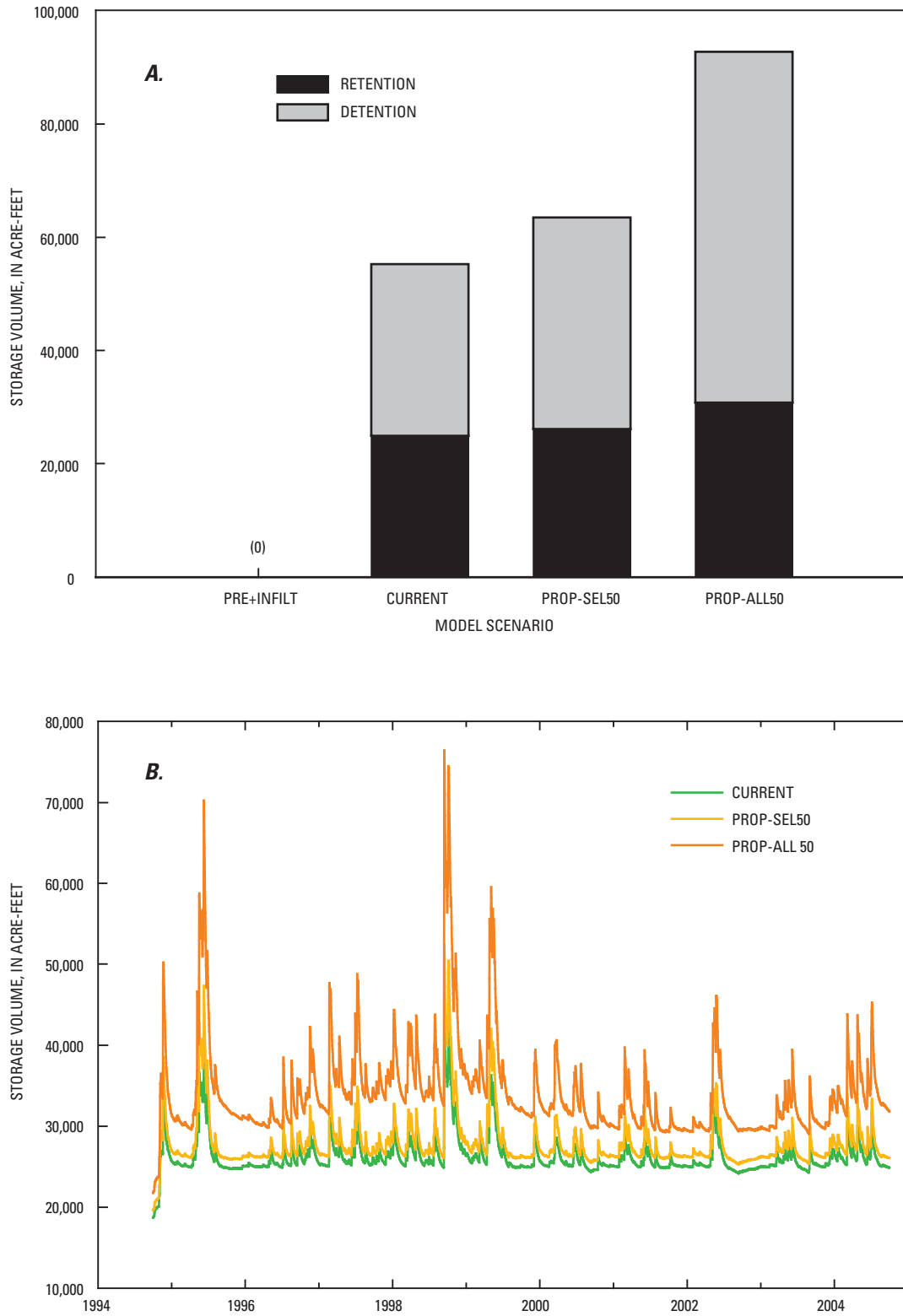


Figure 24. (A) Total potential impoundment retention and detention volume and (B) temporal variability in simulated impoundment storage in the Mamaton River Basin by simulation scenario.

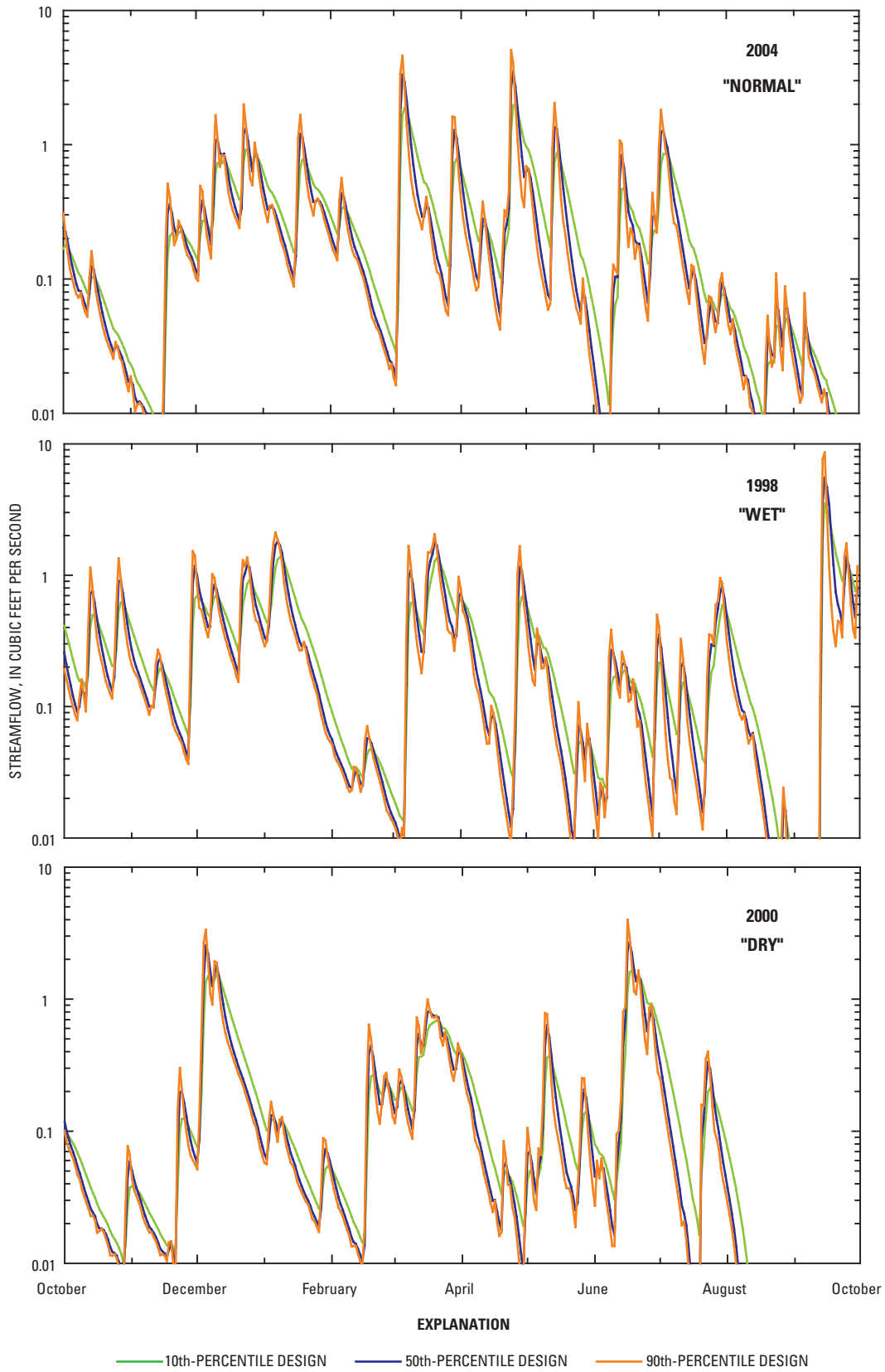


Figure 25. Comparison of outflows from a single Marmaton River impoundment by outflow design and range of selected annual climatic conditions.

corresponding in-stream habitat in the upper Osage River Basin. Low flow reductions would exacerbate minimum flow conditions that already result in periods of limiting habitat or “bottlenecks”. Differences in monthly minimum flow characteristics were compared by site, scenario, duration (1-, 3-, 7-, 30-days), and time (monthly, annual).

Differences in monthly proposed minimum 1-, 3-, 7-, and 30-day low flows from observed flows at the Marais des Cygnes River near the Kansas-Missouri state line, Kansas (RCHRES 88; fig. 12), for the 10-year simulation period, indicated that decreases in minimum flows were common [appendix 4 (figures 4-1 to 4-120), on compact disc, at the back of this report; tables 19, 24]. The Prop-all10_{adj} scenario (greatest detention characteristics) resulted in the least decreases in low flows relative to Observed flows; the Prop-all90_{adj} scenario (least detention characteristics) resulted in the greatest occurrence of low-flow declines (table 24). The maximum duration of declines in low flows was 4 months for the Prop-all50_{adj} and 5 months for the Prop-all90_{adj} with the longest period of extended declines occurring in the summer, 2001 and 2002. October had the greatest magnitude of declines in Prop-all50_{adj} monthly minimum 1-, 3-, 7-, and 30-day low flows compared with Observed flows at the Marais des Cygnes near the Kansas-Missouri state line, Kansas, but declines were present in April through November throughout the 10-year simulation period (fig. 26). The greatest declines under Prop-all50_{adj}

conditions were for the lowest 10 percentile of Observed flows (fig. 27) and during the driest years (2000, 2001 water years; fig. 28). One obvious biologically important change in low flows that could occur would be an increase in zero flow days, but no proposed scenario resulted in a greater number of zero flow days at this site (table 24).

The most apparent difference in minimum flows at the Marais des Cygnes River near the Kansas-Missouri state line, Kansas, was between Observed and Pre+infil_{adj} scenario results (appendix 4; tables 19, 24) indicating that the magnitude of changes that have occurred between pre-settlement and current conditions in this basin may be greater than what would be expected under proposed conditions. The 1-, 3-, 7-, and 30-day minimum flows were substantially greater for Observed flows than Pre+infil_{adj} (tables 19, 24) as a result of controlled releases from large impoundments in this basin. Monthly minimum 1-, 3-, and 7-day Pre-infil_{adj} flows were as much as 6,600 ft³/s less than the corresponding observed minimum flows (table 24). Observed flows also were less than Pre+infil_{adj} scenario flows for several months with 30-day Pre+infil_{adj} flows up to maximum of 2,800 ft³/s greater than Observed, but differences generally were less than 300 ft³/s.

There were no additional impoundments proposed for the Little Osage River Basin and, therefore, proposed scenarios were not conducted for this basin. In contrast to differences detected in observed and pre-settlement low flows in the

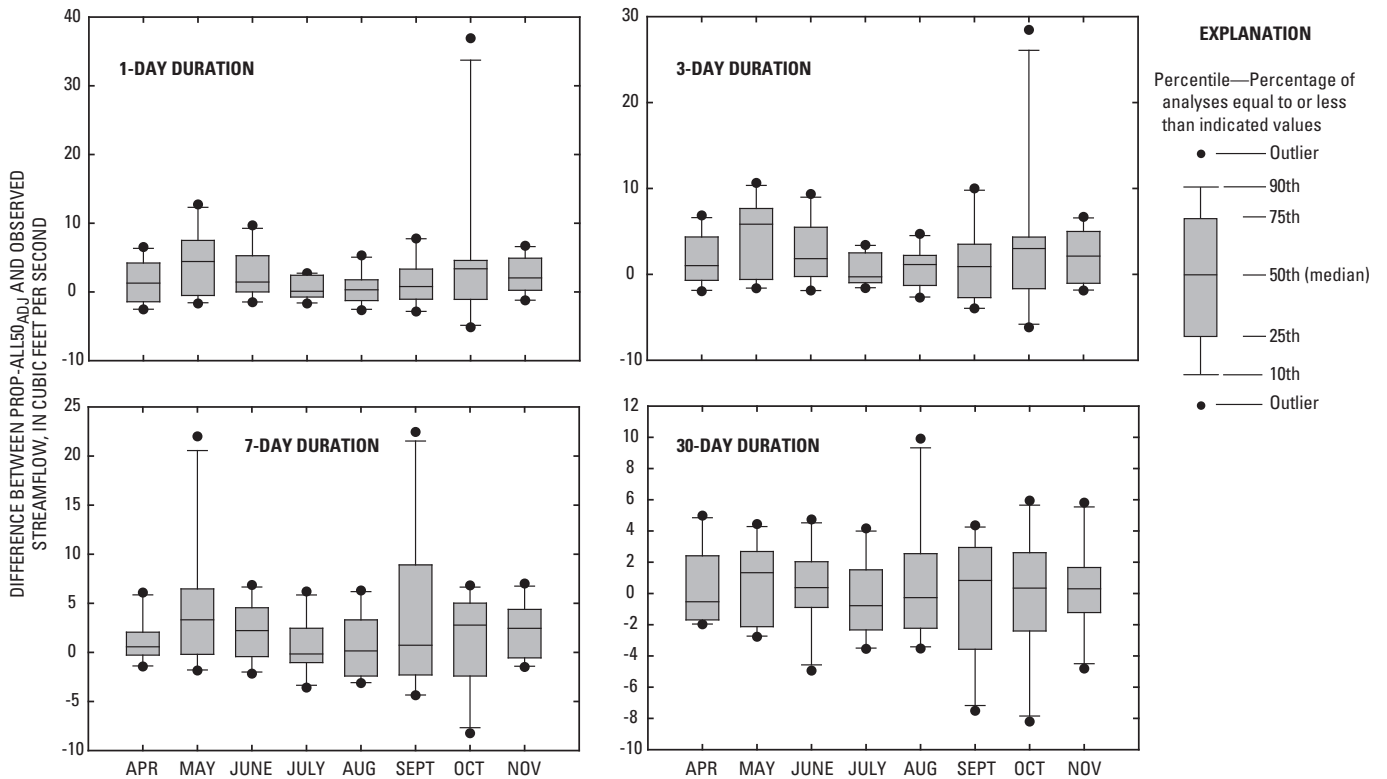


Figure 26. Distribution of differences between monthly minimum 1-, 3-, 7-, and 30-day Prop-all50_{adj} and Observed streamflows, by selected months, for the Marais des Cygnes River near the Kansas-Missouri state line, Kansas (RCHRES 88), water years 1995 through 2004.

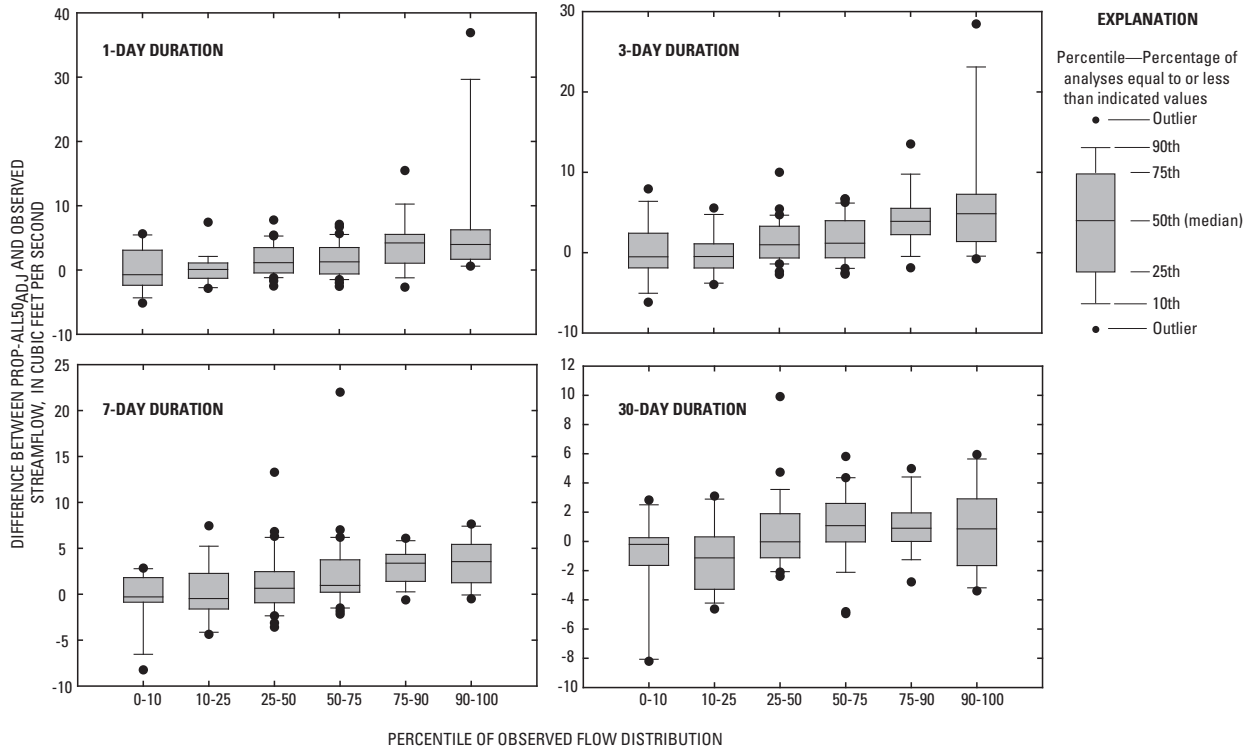


Figure 27. Distribution of differences between monthly 1-, 3-, 7-, and 30-day minimum Prop-all50_{adj} and Observed streamflows, by distribution of Observed flows, for the Marais des Cygnes River near the Kansas-Missouri state line, Kansas (RCHRES 88), water years 1995 through 2004.

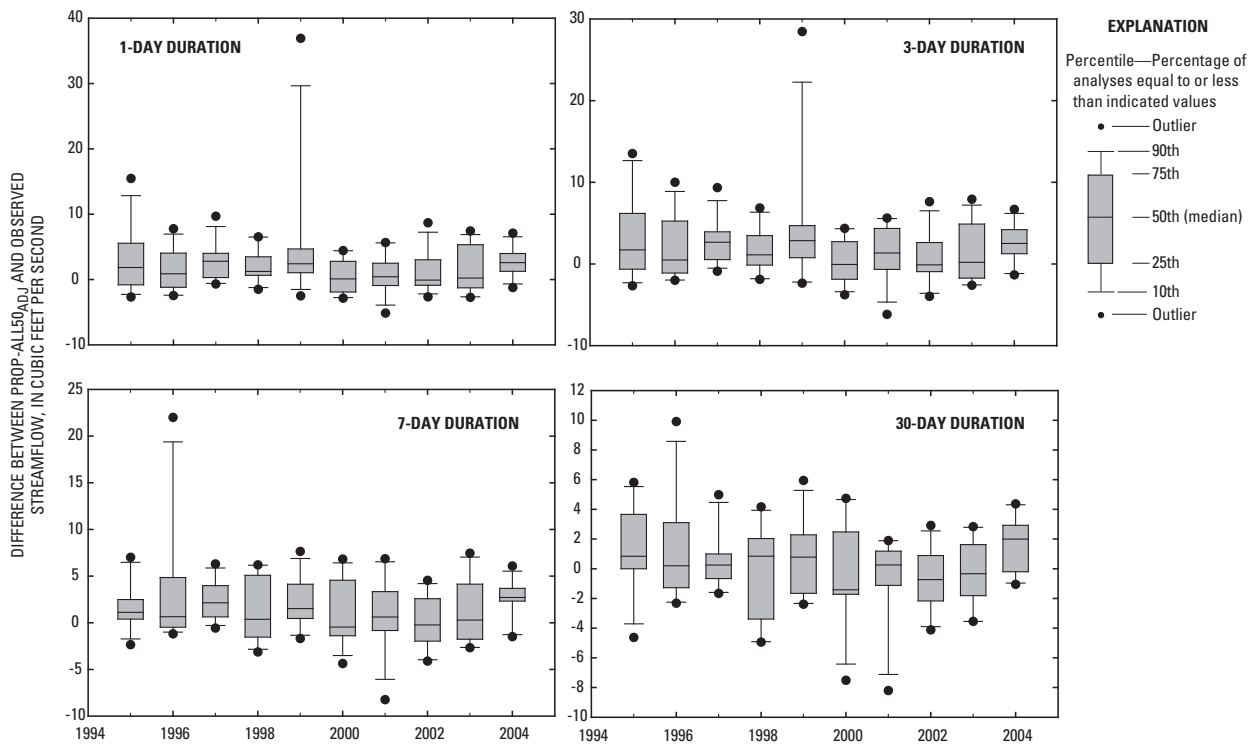


Figure 28. Distribution of differences between monthly minimum 1-, 3-, 7-, and 30-day Prop-all50_{adj} and Observed streamflows, by year, for the Marais des Cygnes River near the Kansas-Missouri state line, Kansas (RCHRES 88), water years 1995 through 2004.

Marais des Cygnes River Basin, the Pre+infiltration_{adj} low flows generally were greater than Observed flows at the Little Osage River near Fulton, Kansas (RCHRES 25; fig. 13; table 20, 25). The increased minimum low flows under simulated historical conditions compared with Observed flows resulted from greater INFILT parameter values representing greater infiltration and hydraulic conductivity characteristics of historical uncultivated prairie soils. The reduction in the extent of native prairies was the primary difference between current and historical conditions in this basin as effects from impoundments were minimal. Pre+infiltration_{adj} low flows seemed to be particularly better sustained than Observed flows during dry years (water years 2000–2003), and with increasing flow duration (from 1-day to 30-day periods; table 20, 25). The magnitude of the maximum differences between Pre+infiltration_{adj} and Observed flows generally was less than 100 ft³/s (table 25).

Differences in low-flow characteristics for simulated and Observed (or Current simulated scenarios at ungaged sites) are presented for multiple locations in the Marmaton River Basin to present the spatial variability in potential effects from impoundments and land-cover changes. In a small headwater basin in the upper Marmaton River Basin (RCHRES 54; fig. 13) simulated declines in minimum flows were small (generally less than 6 ft³/s and less than 1 ft³/s for 1- and 3-day scenarios), but resulted in 10 to 18 additional zero flow days for the Prop-all50 and Prop-all90 scenarios relative to Current conditions (table 26). Declines primarily occurred during the summer months when minimum flows already were near zero flow and the largest duration of consecutive declines from existing conditions was 6 months for the Prop-all90 30-day duration for the summer and fall of 2002. Simulated Pre+infiltration_{adj} minimum flows generally were greater than Current simulated flow conditions for all flow duration periods. Maximum monthly differences generally were less than 10 ft³/s (table 26).

Differences between simulated and observed minimum monthly flows were similar at a downstream Marmaton River near Marmaton, Kansas (RCHRES 58; fig. 13), site to those at the headwater site. Reductions in minimum monthly flows as a result of additional impoundments generally were less than 5 ft³/s (tables 22, 27) but resulted in additional zero flow days for all flow duration periods (table 27). The maximum duration of consecutive monthly declines from Observed conditions was 5 months for the Prop-all90_{adj} simulation in the summer of 1995 (table 27). The number of additional zero flow days at the Marmaton River near Marmaton, Kansas, location were fewer (6 days; table 27) than at the headwater site (18 days; table 26). Simulated pre-settlement (Pre+infiltration_{adj}) monthly minimum flows at the Marmaton River near Marmaton, Kansas, generally were greater than Observed flows (tables 22, 27). Maximum monthly reductions in minimum flows between Pre+infiltration_{adj} and Observed flows, as a result of increased INFILT parameter representing land-cover changes, were 36 to 71 ft³/s (table 27).

Similar to the headwater and Marmaton, Kansas, locations, the reductions in flows at the Marmaton River near the

Kansas-Missouri state line, Missouri (RCHRES 6; fig. 13), as a result of proposed impoundments, were most frequent for the Prop-all90 (least detention) scenario and generally occurred in the summer months [table 28; appendix 5 (figures 5-1 through 5-120), on compact disc, at the back of this report]. The maximum duration of declines in proposed scenario low flows relative to current flows was 6 months for the Prop-all90 scenario in the summers of 2000 and 2002 for multiple flow duration periods. July, August, and October had the largest declines in Prop-all50 low flows relative to Current scenario low flows during the 10-year simulation period (fig. 29). The greatest declines between Prop-all50 and Current flows generally occurred in the lower 50 percentile of Current scenario flows (fig. 30) and during the drier years of 2001–2003 (fig. 31). Proposed conditions resulted in declines in the 0–10 percentile flow values for the 1-, 3-, and 7-day duration periods. Simulated reductions for proposed scenarios did not result in an increase in zero flow days at this location. Pre+infiltration_{adj} minimum flows were, again, generally greater than Current scenario flows for all flow duration periods (appendix 5; table 28) with maximum declines in monthly low flows of 150 to 165 ft³/s between pre-settlement and current scenarios (table 28).

High Flows

Proposed scenario monthly high flows were reduced as a result of the addition of impoundments in all basins and at all reporting locations (fig. 32; tables 19–23). Observed 1-day, monthly maximum streamflows at the Marais des Cygnes near the Kansas-Missouri state line, Kansas, were, on average, 5 to 17 ft³/s greater than proposed maximum monthly flows (fig. 32; table 19). Given the existing amount of controlled regulation in the basin, the proposed scenarios represent a small potential change and little variability resulting from impoundment design. Pre-settlement monthly maximum flows were, on average, 540 to 776 ft³/s greater than Observed 1-day monthly maximum flows, indicating that similar to low flows, the characteristics of high flows in this basin also have undergone greater changes between pre-settlement to current conditions. These changes between pre-settlement and current conditions in this basin were, again, the result of several large impoundments with managed detention and outflow characteristics. Observed 1-day monthly maximum flows at the Little Osage River near Fulton, Kansas, were similar to pre-settlement monthly maximums as Observed flows only were 0.8 ft³/s greater, on average, than Pre+infiltration_{adj} monthly maximum flows (fig. 32; table 20).

The Observed or Current scenario 1-day monthly maximum flows were, on average, greater than pre-settlement or proposed flow conditions in the Marmaton River Basin. The possible increased infiltration under historical native prairies compared with current cultivated land and the addition of impoundments under proposed conditions resulted in reduced streamflow peaks. The 1-day monthly maximum Current scenario flows were, on average, 38 to 63 ft³/s greater

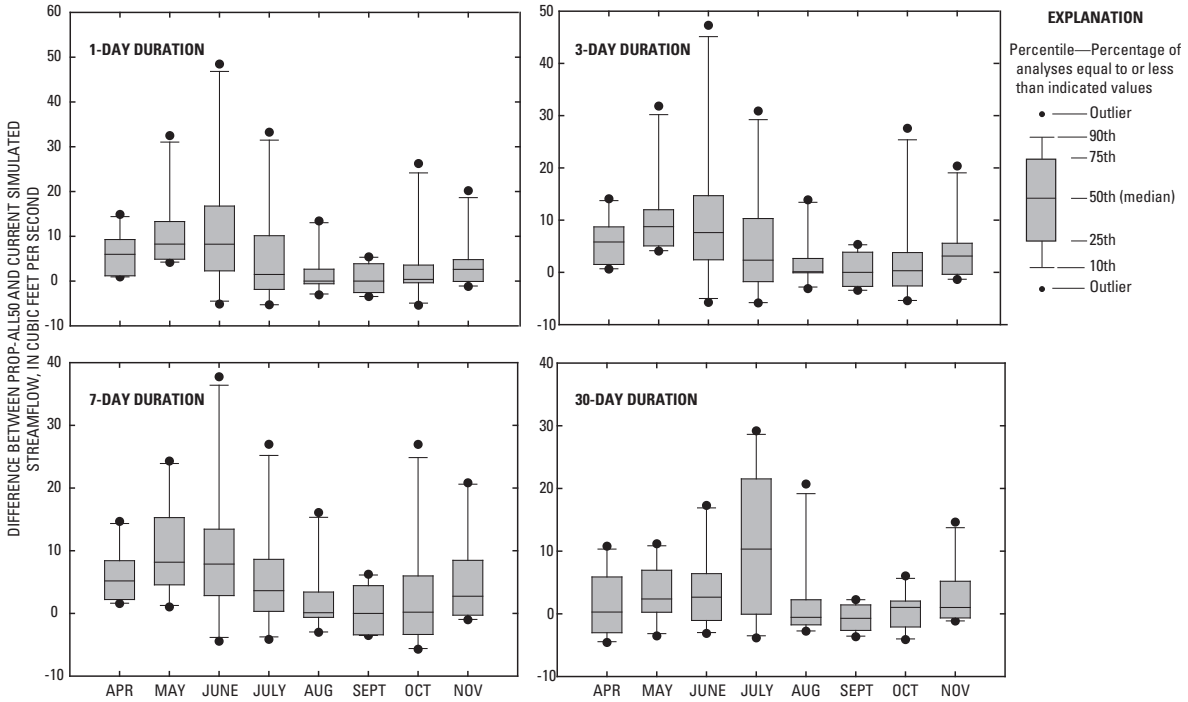


Figure 29. Distribution of differences between monthly minimum 1-, 3-, 7-, and 30 -day Prop-all50 and Current simulated streamflows, by months, for the Marmaton River near the Kansas-Missouri state line, Missouri (RCHRES 6), water years 1995 through 2004.

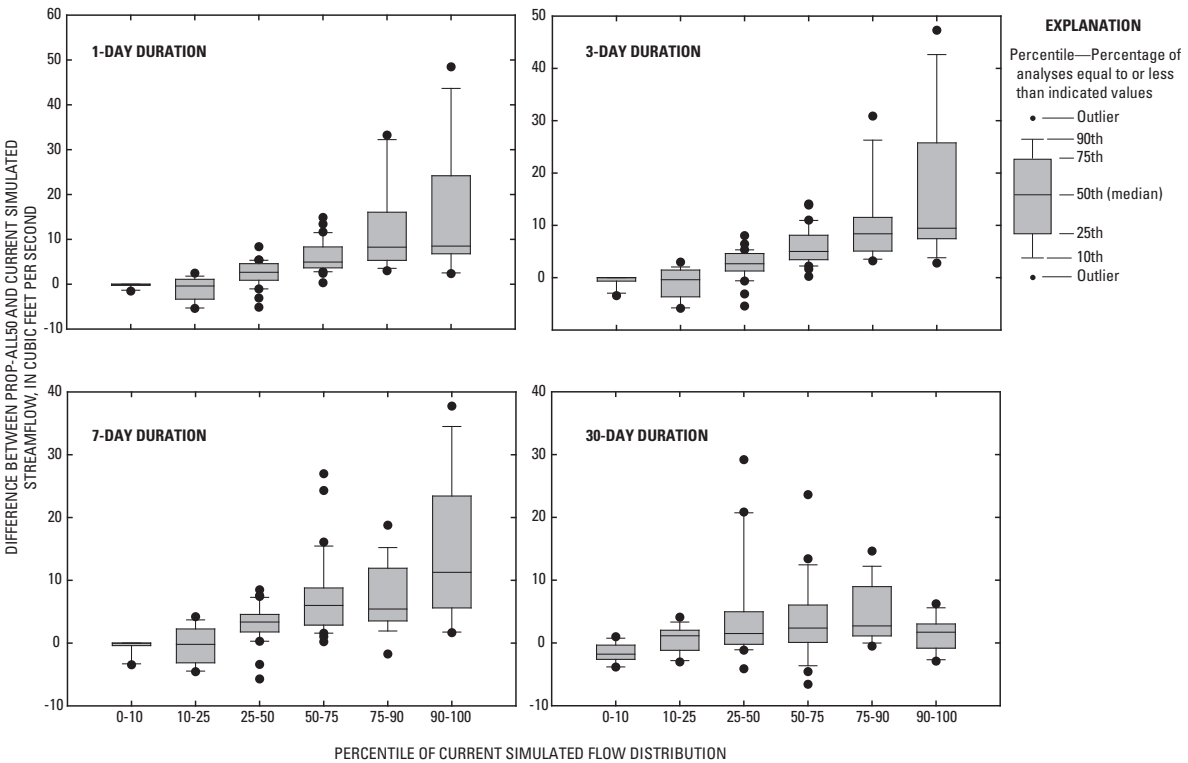


Figure 30. Distribution of differences between monthly minimum 1-, 3-, 7-, and 30 -day Prop-all50 and Current simulated streamflows, by distribution of Current simulated flows, for the Marmaton River near the Kansas-Missouri state line, Missouri (RCHRES 6), water years 1995 through 2004.

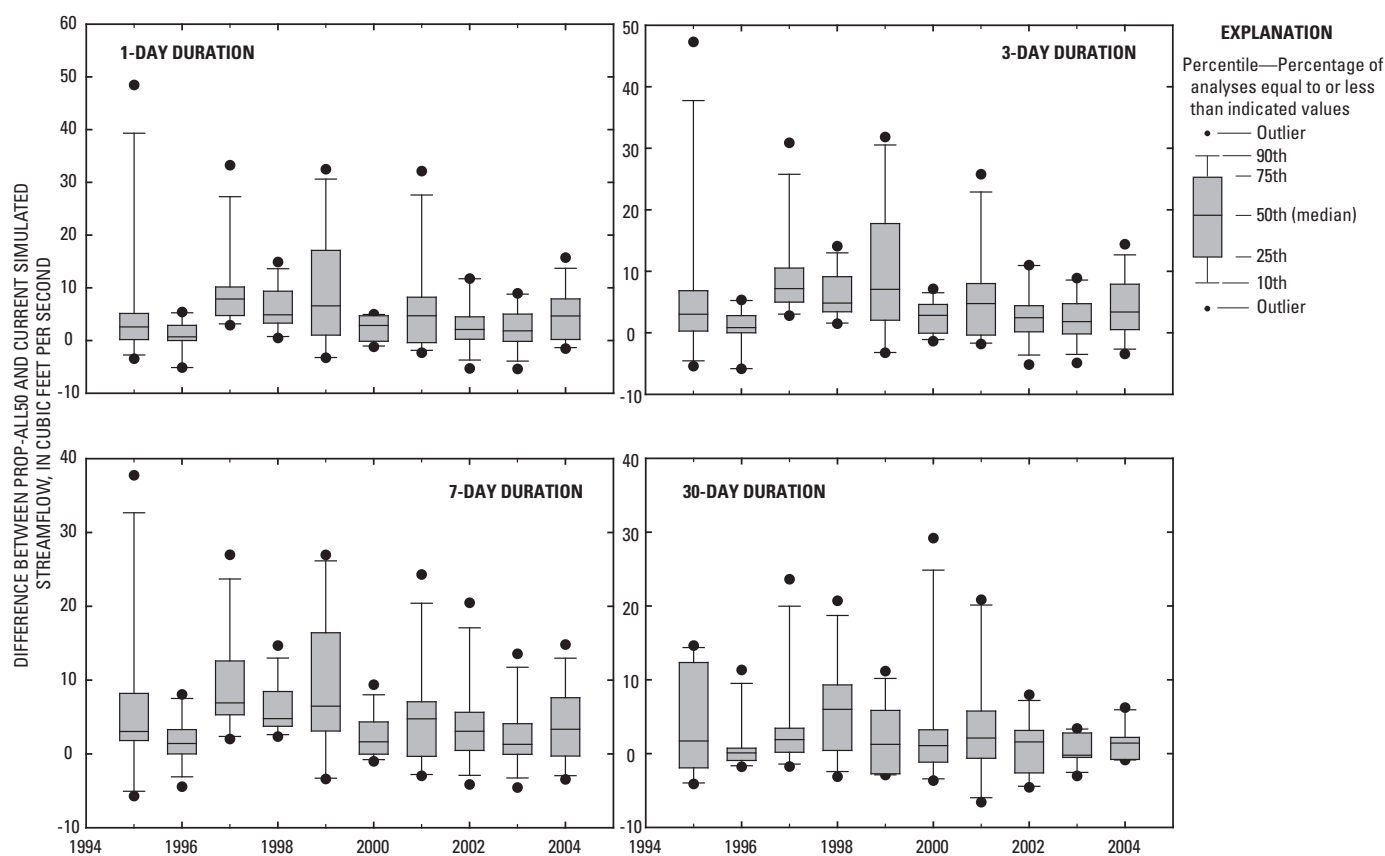


Figure 31. Distribution of differences between monthly minimum 1-, 3-, 7-, and 30-day Prop-all50 and Current simulated streamflows, by year, for the Marmaton River near the Kansas-Missouri state line, Missouri (RCHRES 6), water years 1995 through 2004.

than proposed scenario maximum flows at the Marmaton River headwater location (RCHRES 54; fig. 32; table 21). The maximum median difference ($63 \text{ ft}^3/\text{s}$) was between the Current and Prop-all10 design scenario (greatest detention) and the least difference ($38 \text{ ft}^3/\text{s}$) was between the Current and Prop-all90 design scenario (least detention). Current scenario maximum monthly flows were about $24 \text{ ft}^3/\text{s}$ greater, on average, than Pre+infiltration scenario flows, again as a result of greater INFILT parameter values representing native prairie soils.

At the Marmaton River near Marmaton, Kansas, the Prop-sel_{adj} monthly maximum flows were 46 to $61 \text{ ft}^3/\text{s}$ less than Observed flows, on average (fig. 32; table 22), whereas the Prop-all_{adj} monthly maximum flows were 100 to $200 \text{ ft}^3/\text{s}$ less than Observed flows, on average, depending on outflow design characteristics. Despite adding only four additional impoundments under the Prop-sel scenarios, there was a sizeable reduction in 1-day maximum flows. Additional impoundments under the Prop-all scenarios resulted in greater reductions in monthly maximum flows compared with Observed flows. In both proposed sets of scenarios, the 10th- or 50th- percentile scenario resulted in the greatest reduction and the 90th- percentile scenario the least. Pre+infiltration_{adj} monthly maximum flows were, on average, $51 \text{ ft}^3/\text{s}$ greater than Observed flows (fig. 32; table 22).

Current monthly maximum flows in the Marmaton River near the Kansas-Missouri state line, Missouri, were 46 to $64 \text{ ft}^3/\text{s}$ greater, on average, than Prop-sel_{adj} flows, and 173 to $300 \text{ ft}^3/\text{s}$ greater than Prop-all_{adj} monthly maximum flows, depending on impoundment outflow design (fig. 32; table 23). While the Prop-sel scenario differences were similar to those at the upstream Marmaton, Kansas location, the Prop-all scenario differences were greater, reflecting the additional regulation from proposed impoundments in the interim area between these locations. The Prop-all10 (greatest detention) design scenario again resulted in the greatest average declines in maximum flows ($300 \text{ ft}^3/\text{s}$) from Current conditions. Current scenario 1-day monthly maximum flows at this site were $109 \text{ ft}^3/\text{s}$ greater than Pre+infiltration_{adj} flows, on average (fig. 32; table 23).

Flood Frequency

An ecological consequence of a reduction in streamflow magnitude is a reduction in flood frequency. Natural flooding supplies a primary source of water replenishment to riparian wetlands in the upper Osage River Basin (Heimann and Mettler-Cherry, 2004) providing the physical mechanism for channel and flood plain formation, a primary factor in vegeta-

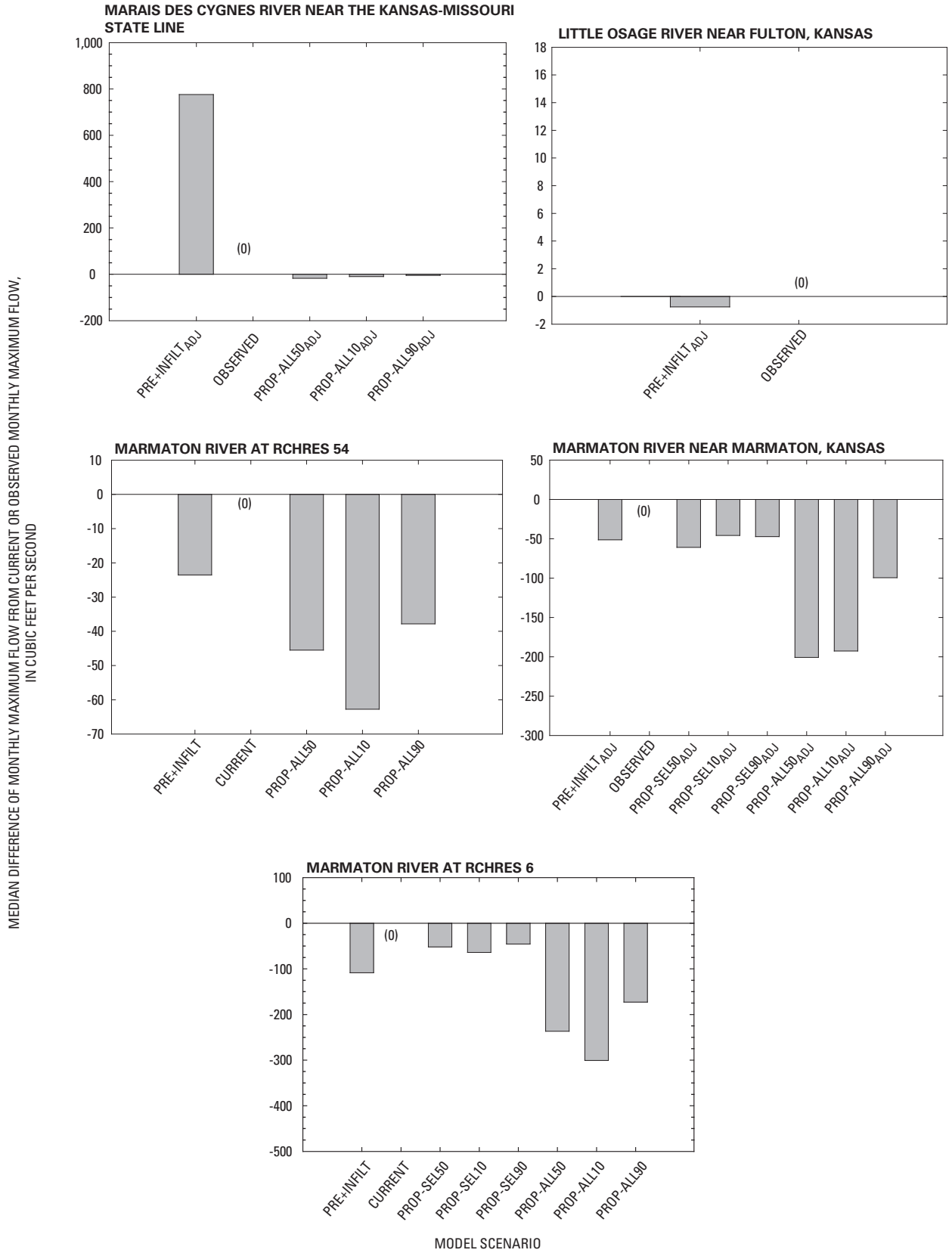


Figure 32. Median difference in 1-day monthly maximum flows from Current or Observed monthly maximum flow, by model basin and scenario, water years 1995 through 2004.

tion composition, distribution, and maintenance as well as habitat for a variety of wildlife. Analyses of hourly simulated and adjusted time series indicated that flood frequency generally was reduced between Current and proposed scenarios in each model basin, although the degree of reduction varied with location in the basin.

The flood frequency for the Marais des Cygnes River near the Kansas-Missouri state line, Kansas, was substantially reduced between pre-settlement and observed conditions, but minimal change was simulated between Observed and proposed scenarios. Flood frequency under the pre-settlement scenario was 50 percent greater than the Observed scenario (fig. 33) for the 10-year simulation period. This reduction in flooding can be attributed to the substantial amount of current controlled regulation in the Marais des Cygnes River Basin. Flood frequencies remained unchanged between the Observed and proposed scenarios at this site as the additional regulation from the small, uncontrolled impoundments did not result in substantial reductions in flood peaks.

There were no differences between Pre+infil_{adj} and Observed scenario flood frequency (fig. 33) at the Little Osage River near Fulton, Kansas. Despite land-cover changes, flooding characteristics remained similar in this basin and the effects from current impoundments was not substantial enough to result in changes in flood frequencies. There are no proposed impoundments in this basin, so flood characteristics should remain unchanged for the foreseeable future.

Flood frequency generally was greatest for the Current or Observed scenarios in the Marmaton River Basin and least for the proposed conditions and the effects of regulation on flood frequency decreases downstream from the Kansas-Missouri state line. The number of floods at the Marmaton River near Marmaton, Kansas, under the Prop-all_{adj} scenario was 28 to 52 percent less than the Observed (fig. 33) scenario; the largest declines were associated with the Prop-all10_{adj} scenario, and the least were from the Prop-all90_{adj} scenario. The flood frequency at the Marmaton River near Marmaton, Kansas, under the Pre+infil_{adj} scenario was about 10 percent less than Observed scenario flood frequency conditions. The flood frequency of the Prop-all scenarios were 54 to 60 percent less (fig. 33) than Current scenario at the Marmaton River near the Kansas-Missouri state line, Missouri (RCHRES 6). The flood frequency of the Pre+infil scenario was 15 percent less than Current simulated conditions. At the downstream Marmaton River RCHRES 11 site the total flood frequency reduction from Current simulated scenario to Prop-all conditions was less (21 to 26 percent) than the upstream RCHRES 6 site (fig. 33). The Pre+infil scenario flood frequency was 9 percent less than the Current simulated scenario. The RCHRES 11 site is downstream from a primary Marmaton River tributary (Big Drywood Creek, fig. 1) with little local regulation and, therefore, the effects of proposed upstream regulation were reduced.

Hydrograph Duration and Flooding Period

The simulated effects of regulation on “large” (as reported in the “Indicators of Hydrologic Alteration” software results; tables 19–23) flood total hydrograph durations varied between model basins and with degree of regulation. Increased regulation increased the total hydrograph duration (duration from pre-event baseflow to post-peak baseflow) of floods. High flows were detained in the impoundments at levels above the primary spillway and released slowly until water levels again reached the level of the primary outflow structure, which extended the streamflow hydrograph duration. The average duration of large flood hydrographs was similar for Observed and proposed scenarios at the Marais des Cygnes River near Kansas-Missouri state line, Kansas (fig. 34). Pre-settlement scenario flood hydrograph durations were, on average, about 80 percent less than Observed. The average large flood hydrograph duration (fig. 34) at the Little Osage River near Fulton, Kansas, for the Pre+infil_{adj} scenario, using assumed increased infiltration estimates under historical land cover, was about 120 percent greater than Observed scenario durations. At the Marmaton River RCHRES 54 and Marmaton River near Marmaton, Kansas, locations the average large flood hydrograph durations under proposed and pre-settlement scenarios were about 50 percent greater than Observed and Current scenarios (fig. 34). At the Marmaton River near the Kansas-Missouri state line, Missouri, the average proposed scenario flood hydrograph durations were 2 to 70 percent greater than Current simulated conditions depending on scenario, whereas flood hydrograph durations were similar under pre-settlement and Current simulated conditions.

Whereas the total hydrograph duration increased with regulation, the actual duration of streamflows above estimated flood levels decreased with proposed regulation in all model basins. The total duration of streamflows above flood levels (flooding period) at the Marais des Cygnes River near the Kansas-Missouri state line, Kansas, decreased only about 2 to 7 percent from Observed to Prop-all_{adj} duration levels (fig. 35). The largest difference in flooding period at this Marais des Cygnes River site occurred between pre-settlement and Observed scenarios, in which Pre+infil_{adj} average flooding period was more than 200 percent greater than Observed scenario flooding period. Pre-settlement flooding periods in the Little Osage River near Fulton, Kansas were similar (within 4 percent) to Observed scenario conditions (fig. 35). At the Marmaton River near Marmaton, Kansas, and near the Kansas-Missouri state line, Missouri, the average flooding period under Prop-all scenarios indicated a 50 to 60 percent reduction from Observed and Current scenario conditions (fig. 35). The Prop-all flooding periods were only 21 to 26 percent less than Current simulated conditions, on average, at the downstream Marmaton River near RCHRES 11 site (fig. 35).

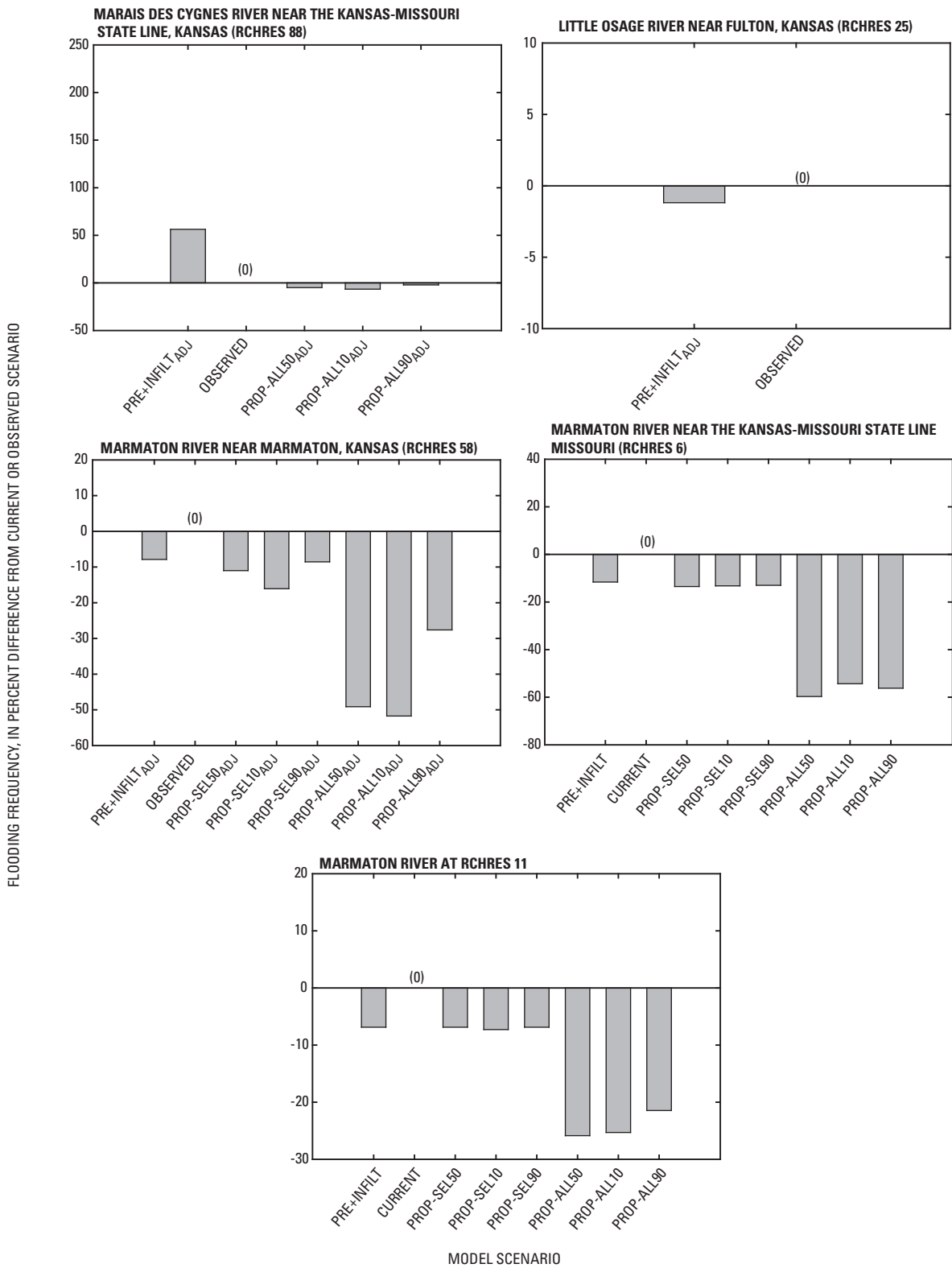


Figure 33. Differences in simulated flood frequency characteristics from Current simulated or Observed frequency, by model basin and scenario, water years 1995 through 2004.

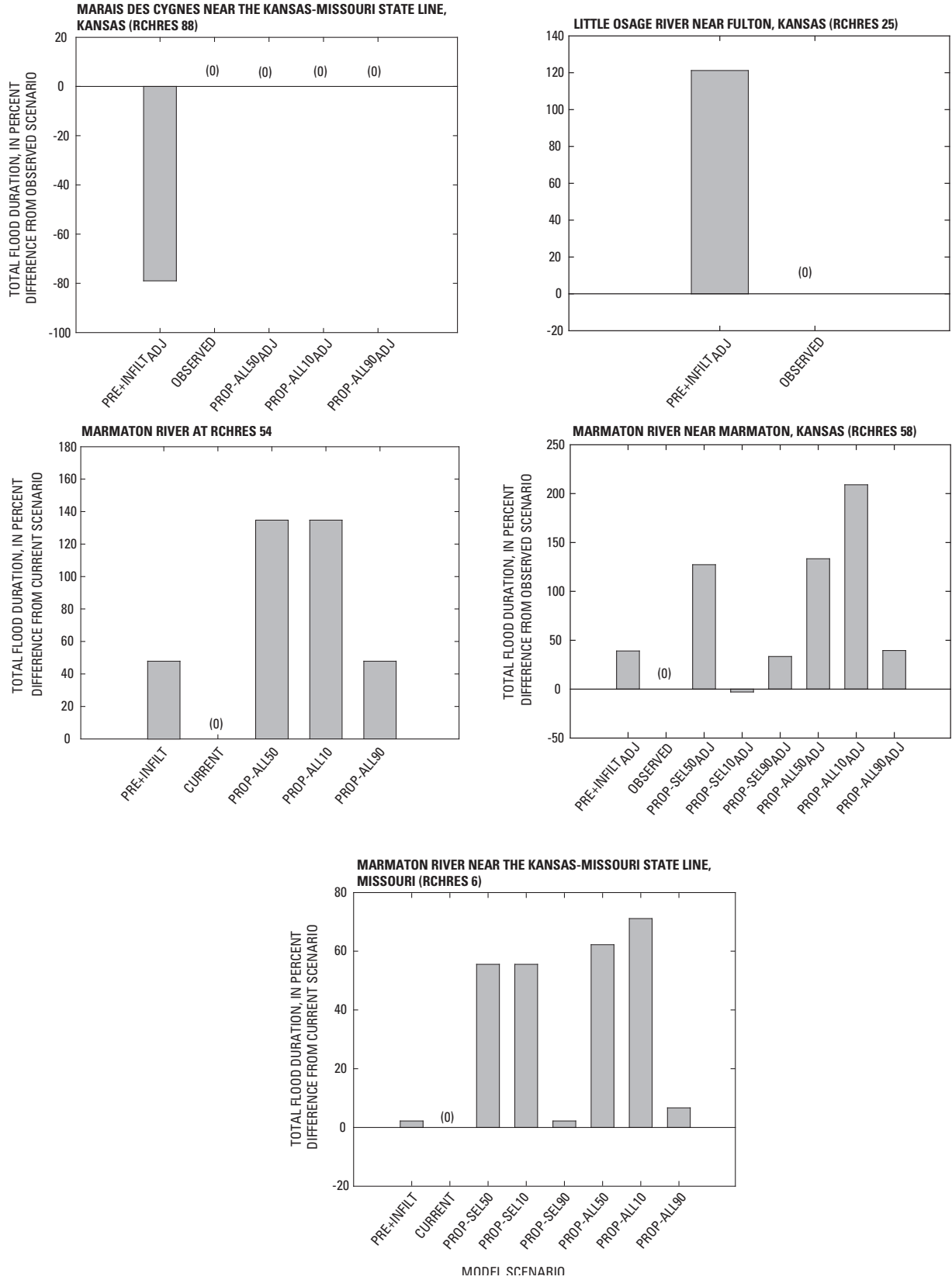


Figure 34. Differences in total large-flood hydrograph duration from Current simulated or Observed conditions, by model basin, reporting site, and simulation scenario, water years 1995 through 2004.

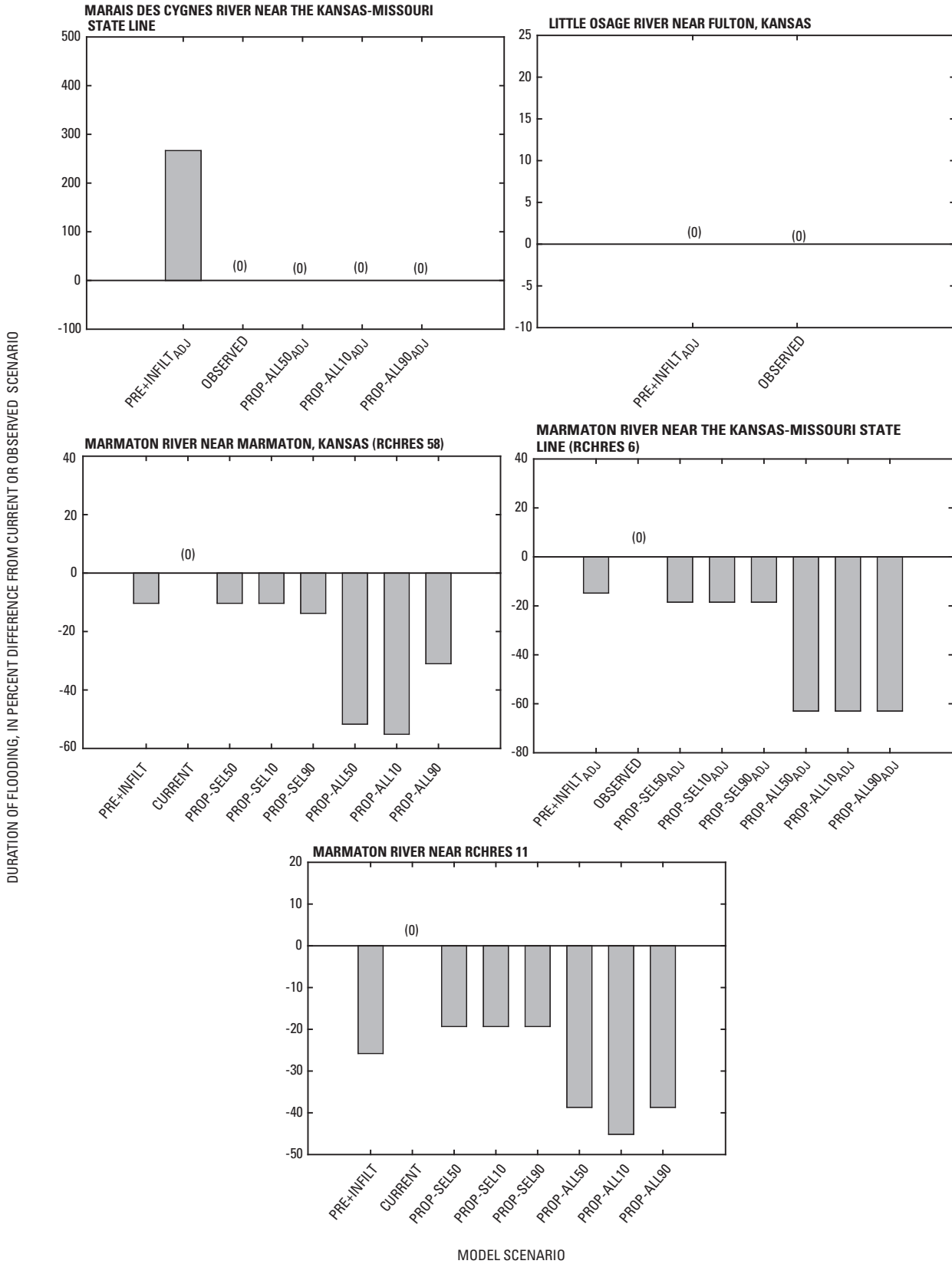


Figure 35. Differences in simulated flooding duration characteristics from Current simulated or Observed conditions, by model basin and scenario, water years 1995 through 2004.

Timing of Extreme Flows

The timing of low and high flows also can be affected by impoundments, particularly those with controlled or managed releases. An alteration in the timing of low flows or peak flows can greatly affect stream channel and riparian biological processes including spawning of fishes, seed transport, and the germination of herbaceous and woody plants. The timing of floods during the 10-year simulation period in the model basins was not substantially affected by regulation, although the timing of low flow extremes was altered in some basins. At the Marais des Cygnes River near the Kansas-Missouri state line, Kansas, the timing of low flows along with small and large floods was similar between scenarios (table 19). The timing of Little Osage River extreme low flows was similar between pre-settlement and Observed flow scenarios, but Pre+infil_{adj} large flood timing was in June rather than September, as for Observed flows (table 20). The timing of low flows and large floods at the Marmaton River near RCHRES 54 were similar between scenarios (table 21). At the Marmaton River near Marmaton, Kansas the timing of peak flows was similar between scenarios, but extreme low flows occurring in August for Pre+infil_{adj} conditions and November for Observed and most proposed scenarios (table 22). The timing of low and high flows was similar under all scenarios at the Marmaton River near the Kansas-Missouri state line, Missouri (table 23).

Effects of Impoundments and Land-Cover Changes on Selected Fish Habitat

The effects of impoundments and land-cover changes within a basin can be quantified with changes in streamflow; however, another means of assessing the ecological consequences of altered conditions is through changes in in-stream habitat. The streamflow changes derived from the hydrologic models were used in conjunction with developed streamflow-fish habitat relations (relations defining the quantity of fish habitat available for a given streamflow; Heimann and others, 2005) to provide a means of quantifying selected ecological effects of streamflow alteration on in-channel habitat in the Marais des Cygnes and Marmaton Rivers. Fish habitat area was not directly proportional to streamflow for most categories (Heimann and others, 2005) so habitat availability can be adversely affected by either declines or increases in streamflows.

Comparisons of 10-year daily habitat area distributions for nine selected fish species/life stage categories (table 17) under the varying streamflow simulation scenarios indicated that the effect of flow alteration on fish habitat varied by basin, scenario, time distribution, and fish species/life stage category. Of particular concern in comparing differ-

ences in habitat by scenario were any possible declines in the minimum habitat availability that could lead to greater limiting conditions or “bottlenecks” in fish habitat availability and possible habitat declines of extended (7–14 days) durations.

Generally, the overall 10-year median fish habitat area in the Marais des Cygnes River for the selected habitat categories was similar between Observed and proposed scenarios at RCHRES 90, 93, and 95 (figs. 36–38; see fig. 12 for site locations; tables 29–31, on compact disc, at the back of this report). Exceptions included median suckermouth minnow habitat at sites RCHRES 90 and 95 that declined 3 (RCHRES 90) to 50 (RCHRES 95) square meters per 100 meters ($m^2/100\text{ m}$) of stream channel under proposed conditions, slenderhead darter habitat at sites RCHRES 90 and 95 that declined about 2 to 10 $m^2/100\text{ m}$ under proposed conditions, and paddlefish habitat, which increased under proposed conditions relative to Observed conditions at all sites as much as 150 $m^2/100\text{ m}$.

Overall distributions of fish habitat area may be similar between simulation scenarios, but the annual distributions and comparisons during selected minimum availability periods provide a better indication of the habitat variability possible with time and between scenarios. The quantity of habitat availability and relative differences under Observed and Prop-all50_{adj} scenarios varied by water year. For example, habitat for paddlefish, whose habitat area is directly proportional to streamflow, varied substantially between wet (1999) and dry (2000) water years at RCHRES90; however, the relative differences in habitat availability between the two scenarios within a particular year also varied substantially (fig. 39). Of particular concern to managers are the habitat bottleneck periods and how such conditions are affected by flow alterations. Minimum annual 1-, 7-, or 14-day habitat availability for the prop-all50_{adj} simulation declined at the Marais des Cygnes River by more than 10 percent compared with Observed conditions for 1 or more years for each of the nine seasonal fish habitat categories at each of the three Marais des Cygnes sites (figs. 40–42) indicating that habitat bottlenecks may be greater under proposed conditions for some categories for some seasons. Declines in minimum habitat availability were at or near 100 percent for 1 or more years for summer flathead catfish, fall flathead catfish, fall channel catfish, and fall stonecat habitat categories at RCHRES 93 (fig. 41) and for summer flathead catfish, summer channel catfish, and summer stonecat at RCHRES 95 (fig. 42). This indicates that habitat for these categories may be eliminated for 1 to 14 days during some part of some years under proposed conditions.

Overall, median habitat area for paddlefish increased substantially (120 to 600 $m^2/100\text{ m}$) between pre-settlement and Observed conditions at the Marais des Cygnes RCHRES 90, 93, and 95 sites, whereas summer and fall flathead catfish, summer channel catfish, and summer and fall stonecat habitat generally increased (1 to 60 $m^2/100\text{ m}$) with slower recession hydrographs under Observed conditions (fig. 36–38; tables

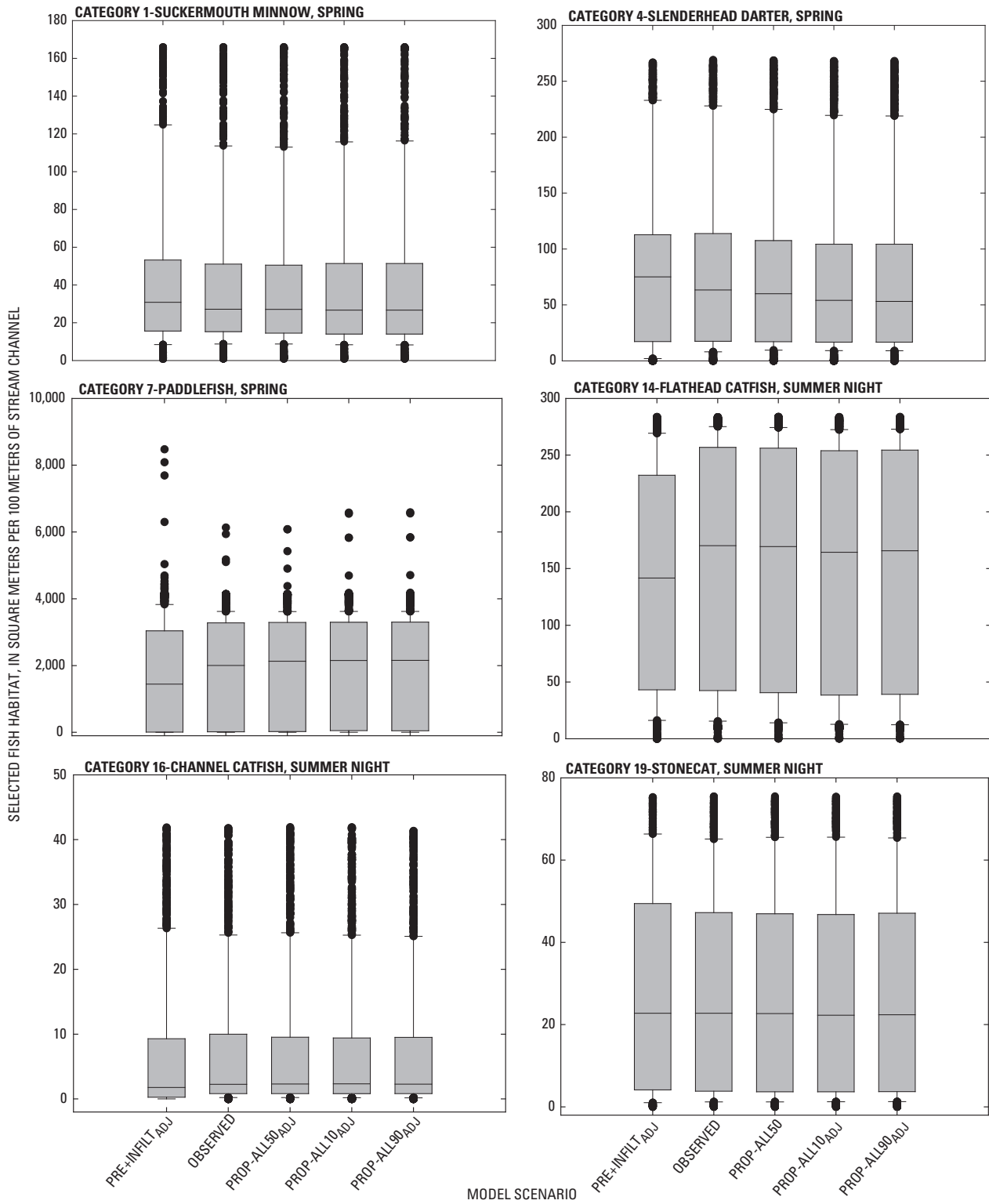


Figure 36. Distribution of daily fish habitat for select categories at the Marais des Cygnes River RCHRES 90, water years 1995 through 2004.

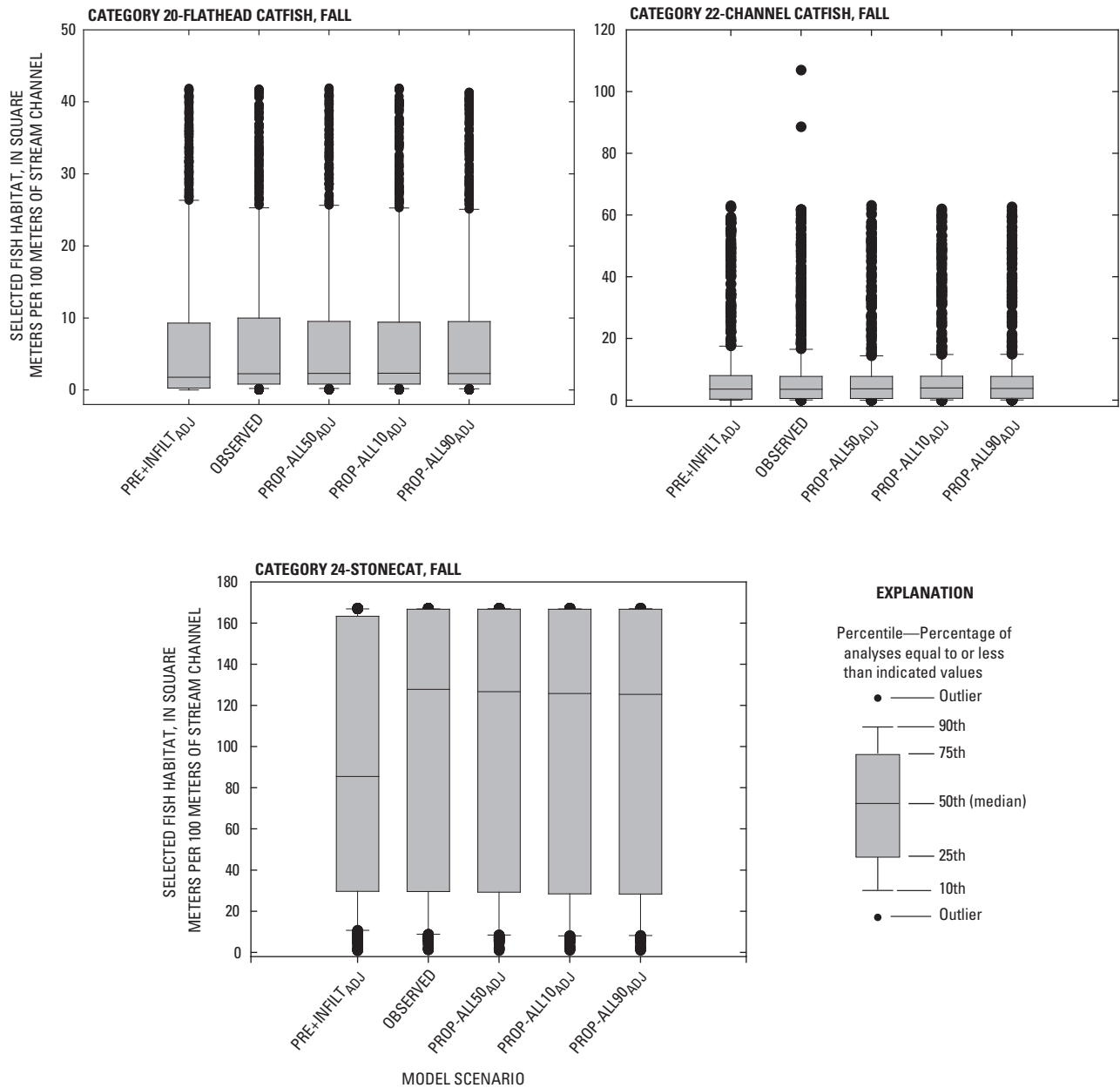


Figure 36. Distribution of daily fish habitat for select categories at the Marais des Cygnes River RCHRES 90, water years 1995 through 2004.—Continued

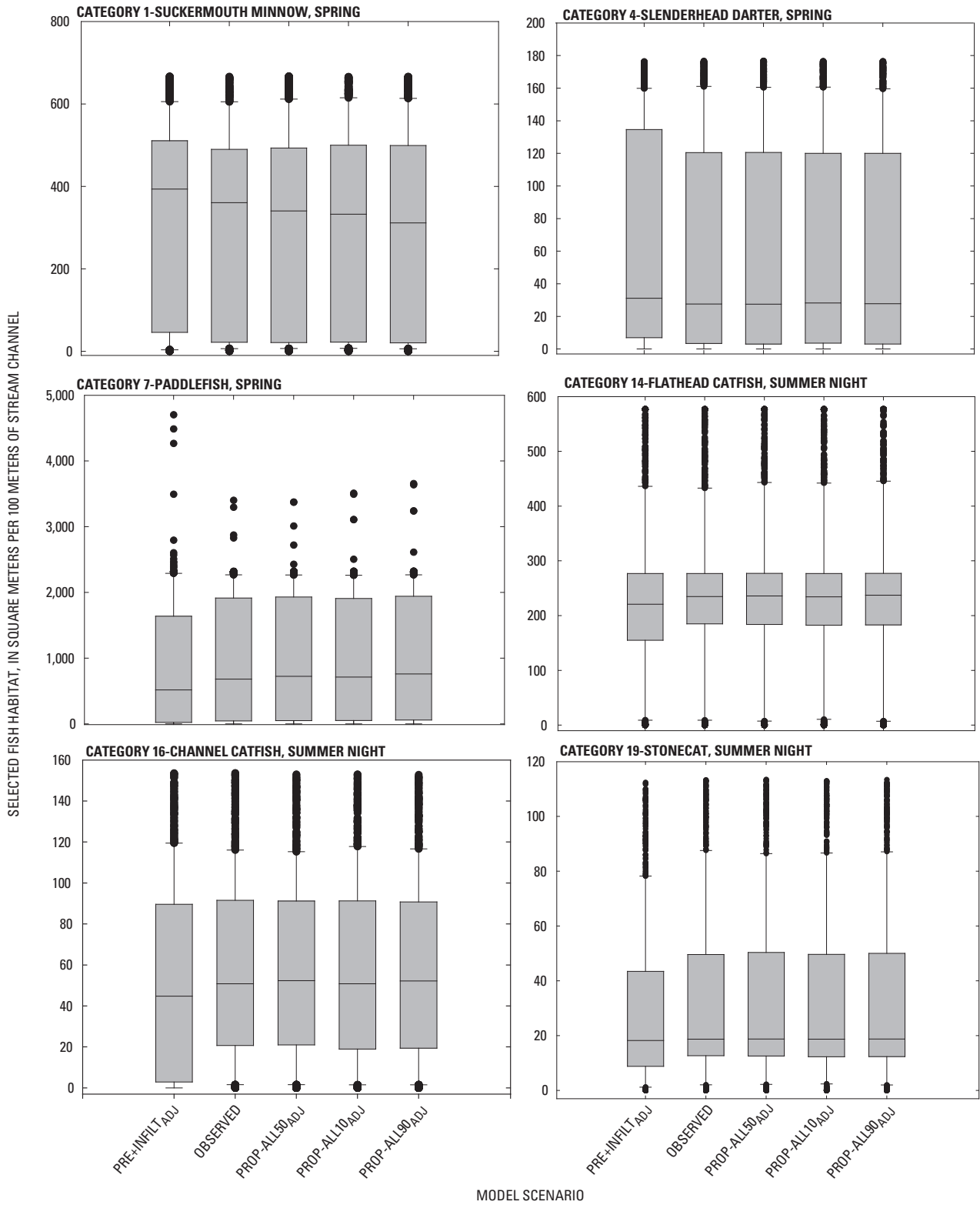


Figure 37. Distribution of daily fish habitat for select categories at the Marais des Cygnes River RCHRES 93, water years 1995 through 2004.

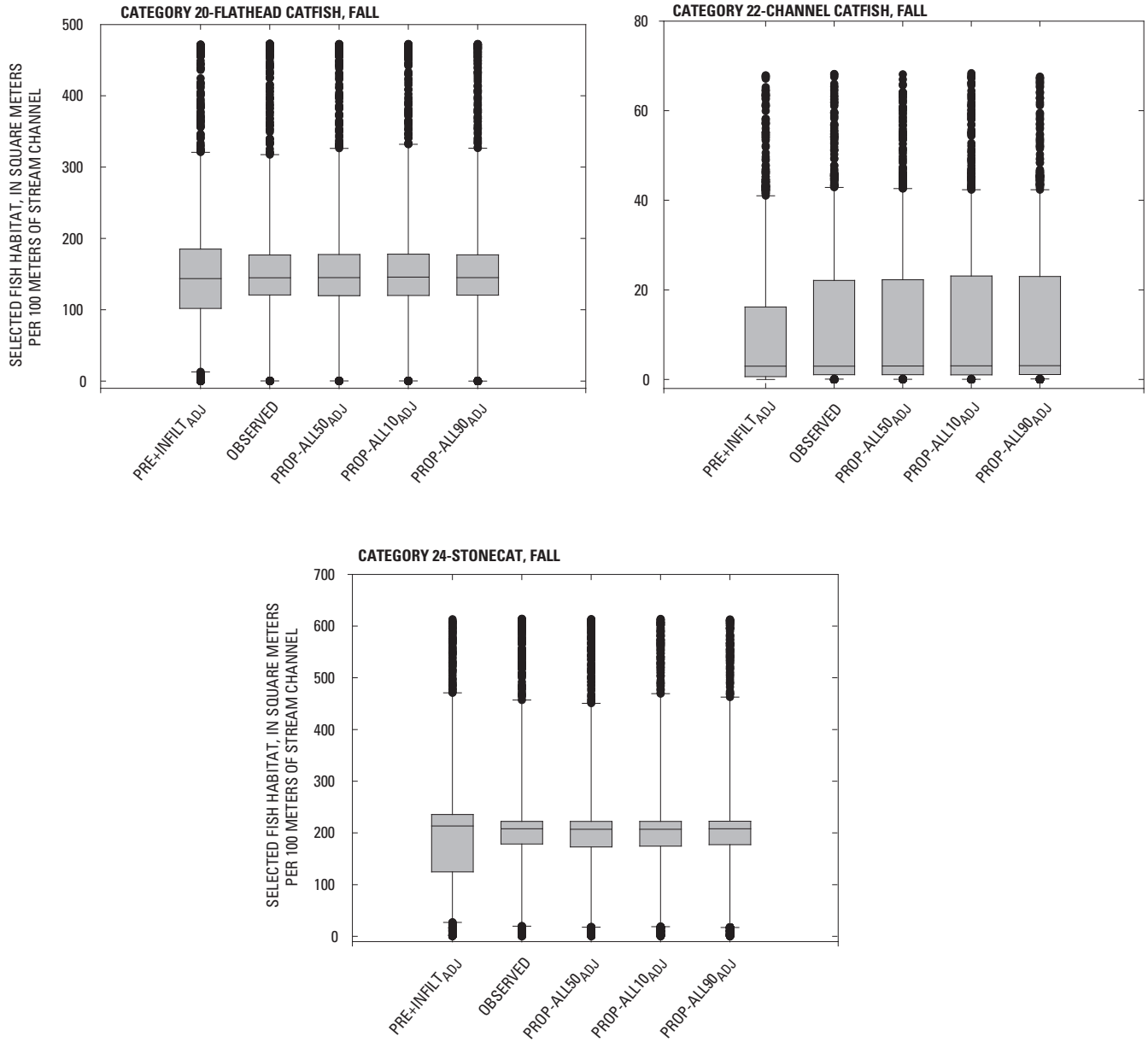


Figure 37. Distribution of daily fish habitat for select categories at the Marais des Cygnes River RCHRES 93, water years 1995 through 2004.—Continued

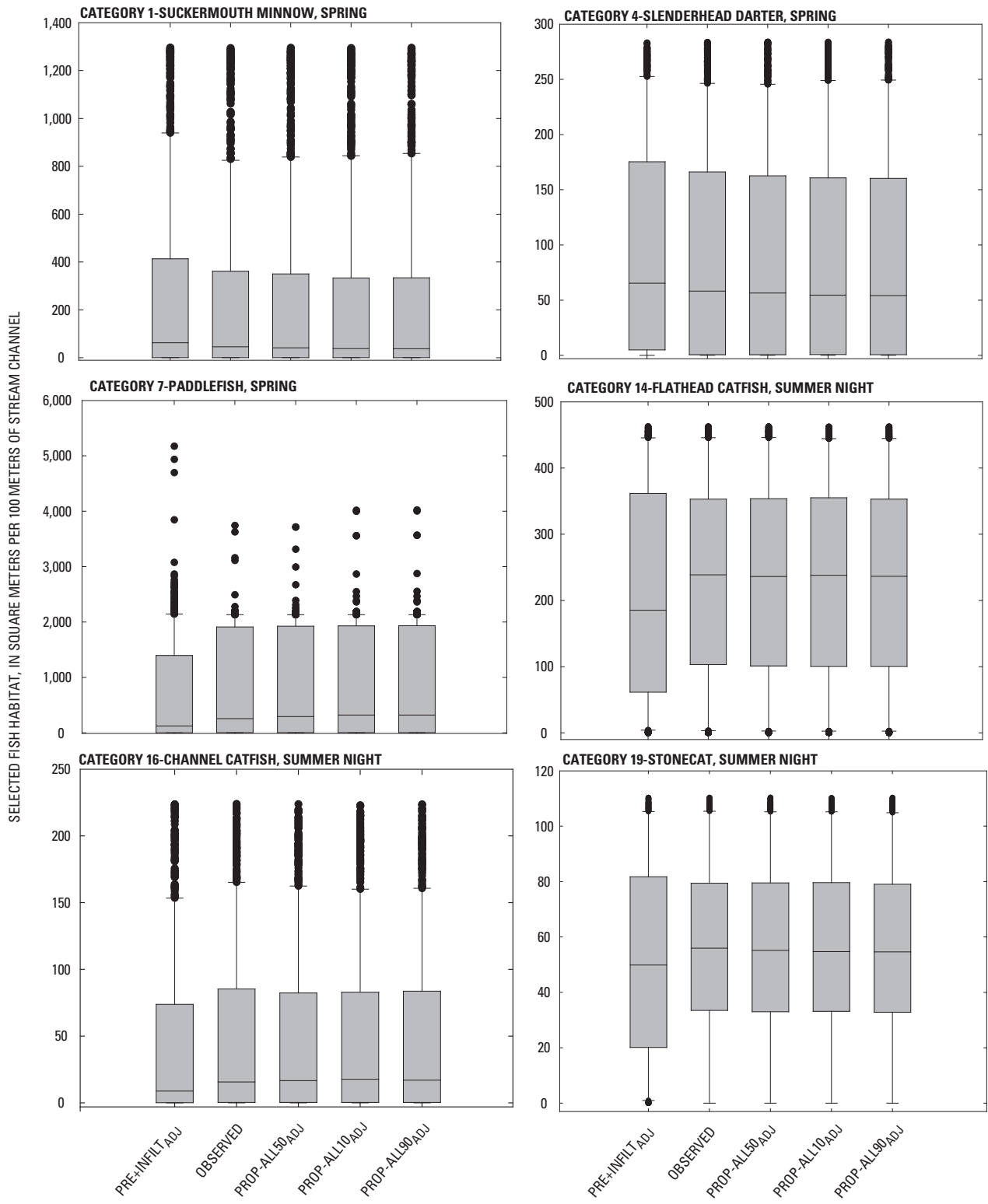


Figure 38. Distribution of daily fish habitat for select categories at the Marais des Cygnes River RCHRES 95, water years 1995 through 2004.

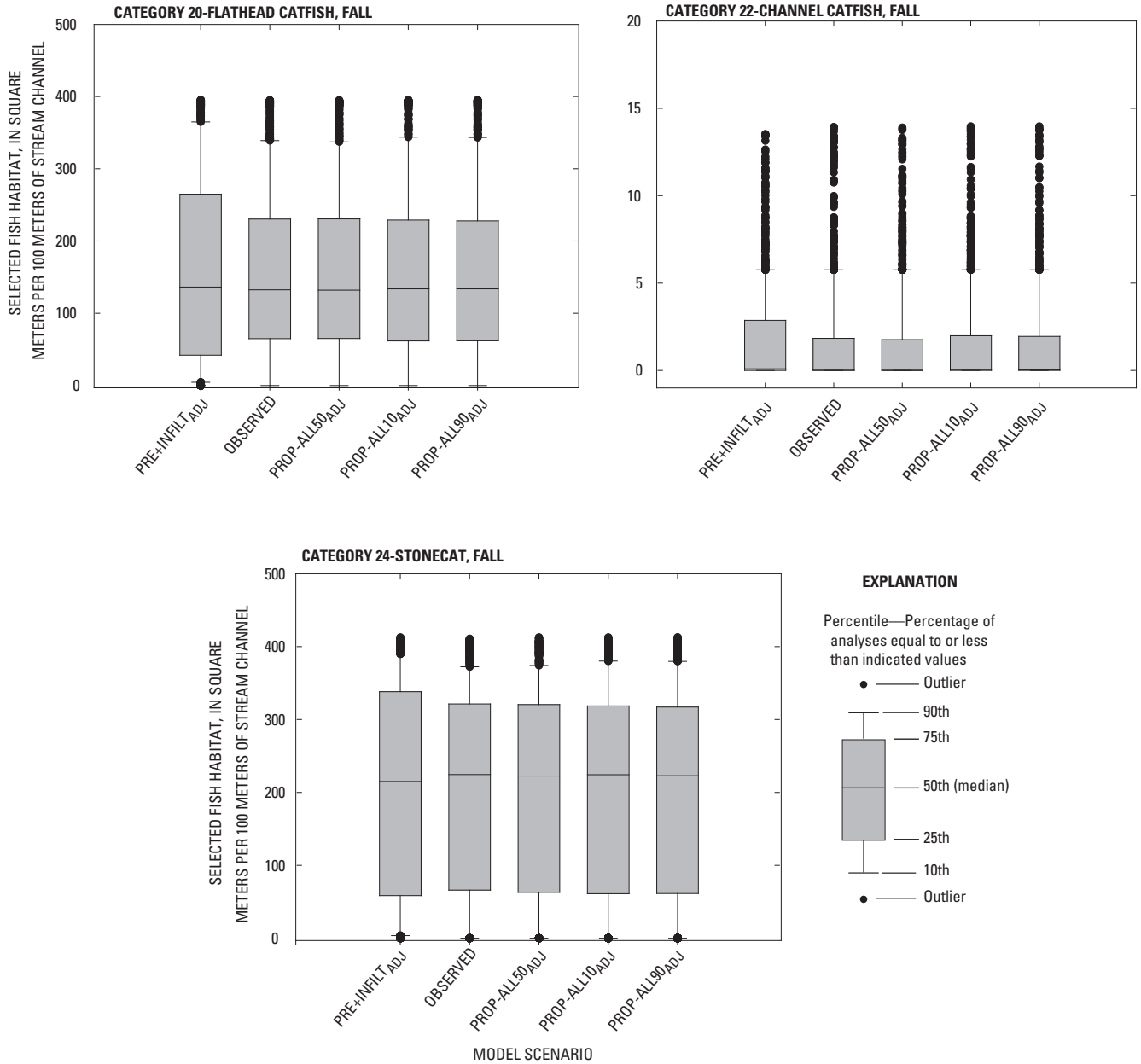


Figure 38. Distribution of daily fish habitat for select categories at the Marais des Cygnes River RCHRES 95, water years 1995 through 2004.—Continued

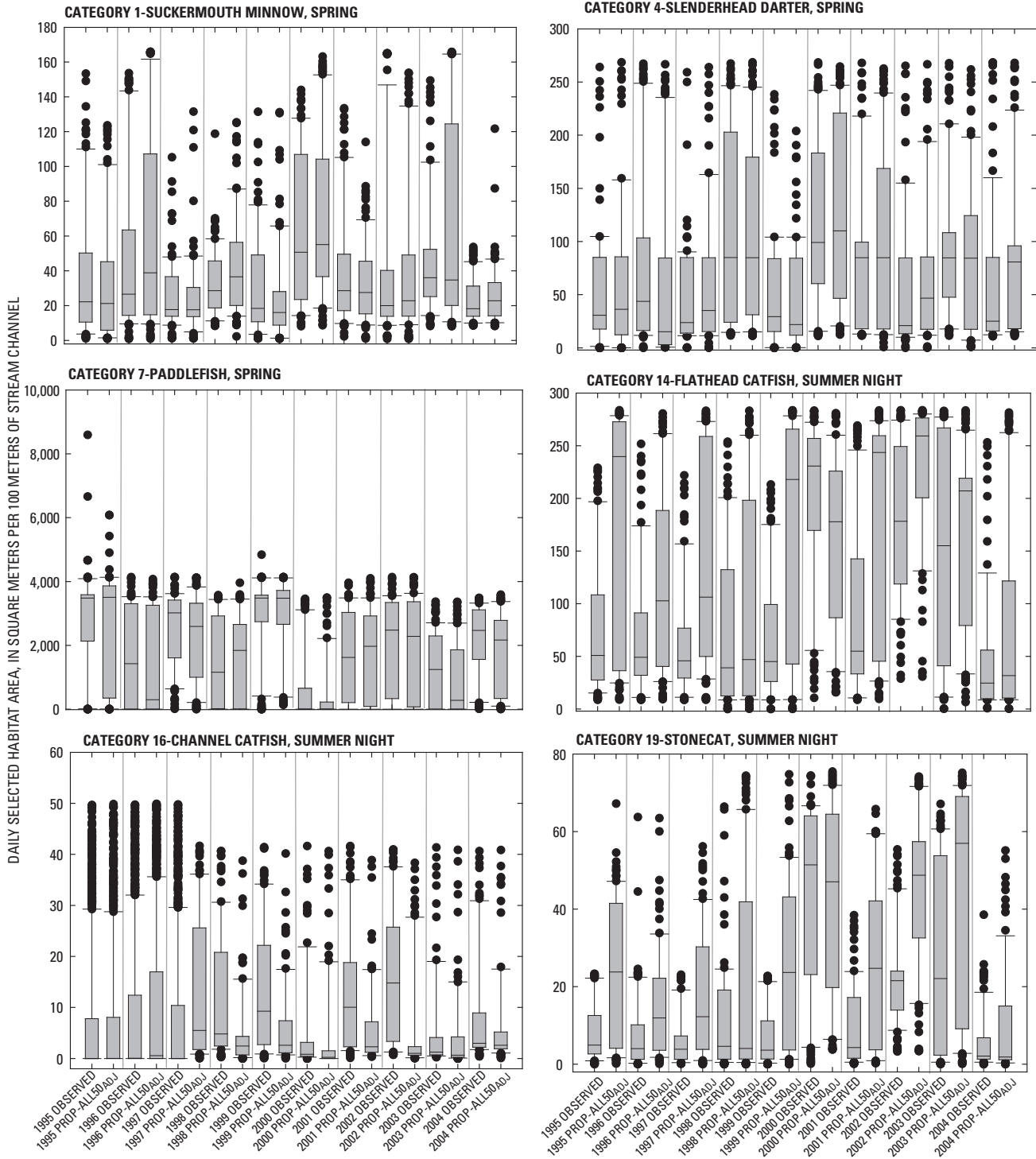


Figure 39. Comparison of distributions of daily fish habitat for Observed and Prop-all50_{adj} streamflow scenarios, by year and select species/life stage categories, at the Marais des Cygnes River RCHRES 90, water years 1995 through 2004.

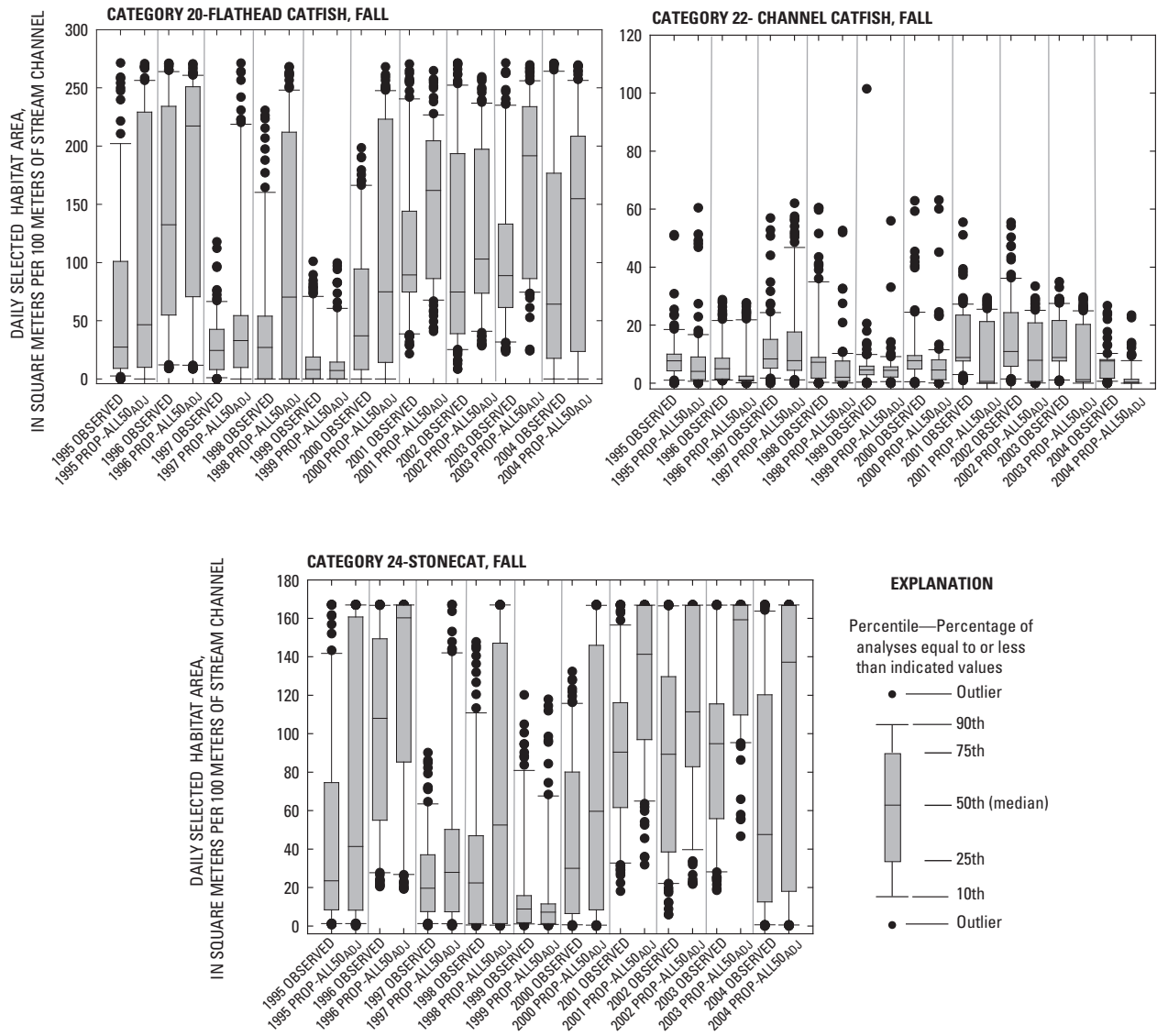


Figure 39. Comparison of distributions of daily fish habitat for Observed and Prop-all50_{adj} streamflow scenarios, by year and select species/life stage categories, at the Marais des Cygnes River RCHRES 90, water years 1995 through 2004.—Continued

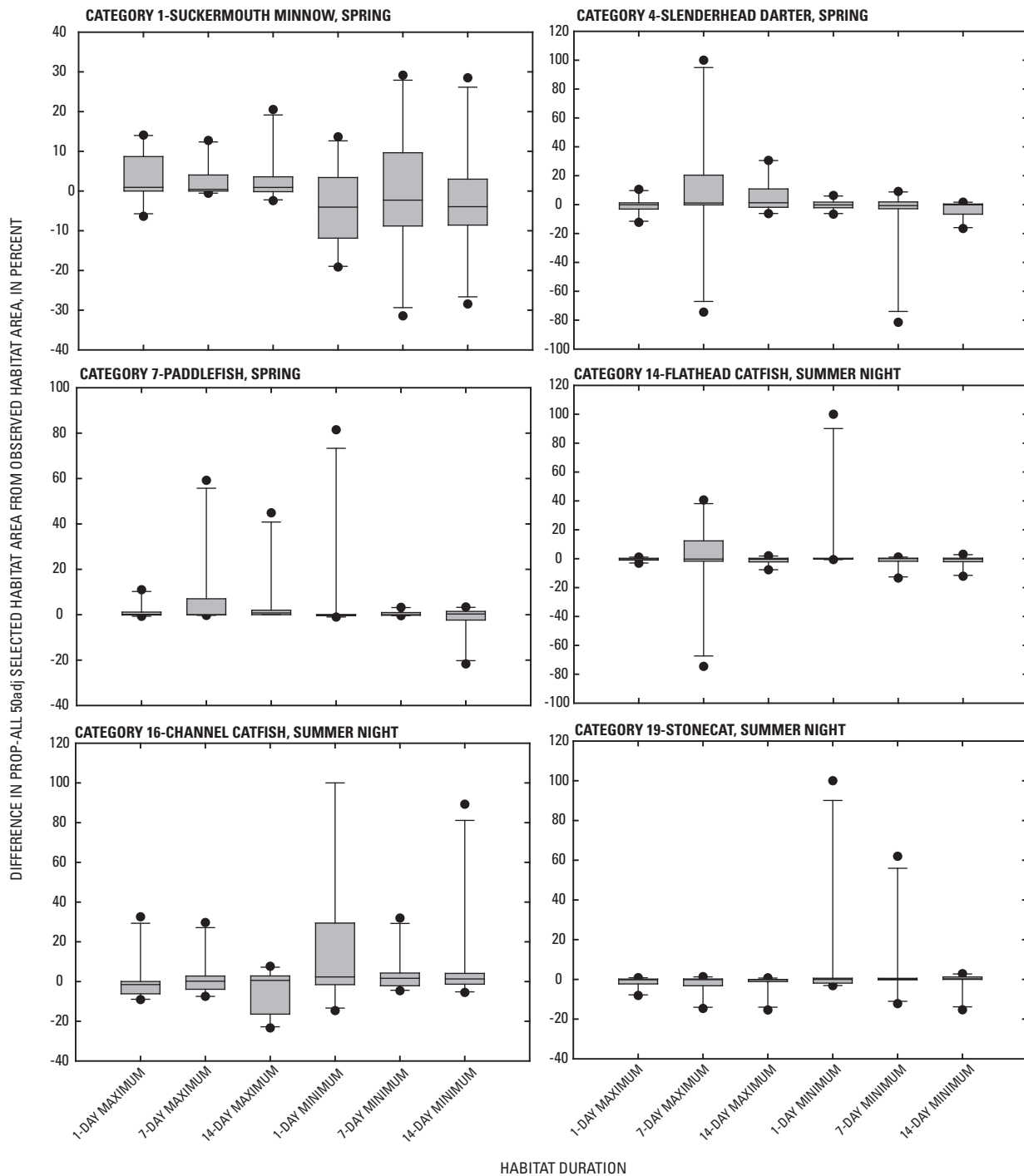


Figure 40. Distribution of percent differences between Prop-all50_{adj} and Observed fish habitat conditions for 1-, 7-, and 14-day maximum and minimum durations, by selected species/life stage categories, at the Marais des Cygnes River RCHRES 90, water years 1995 through 2004.

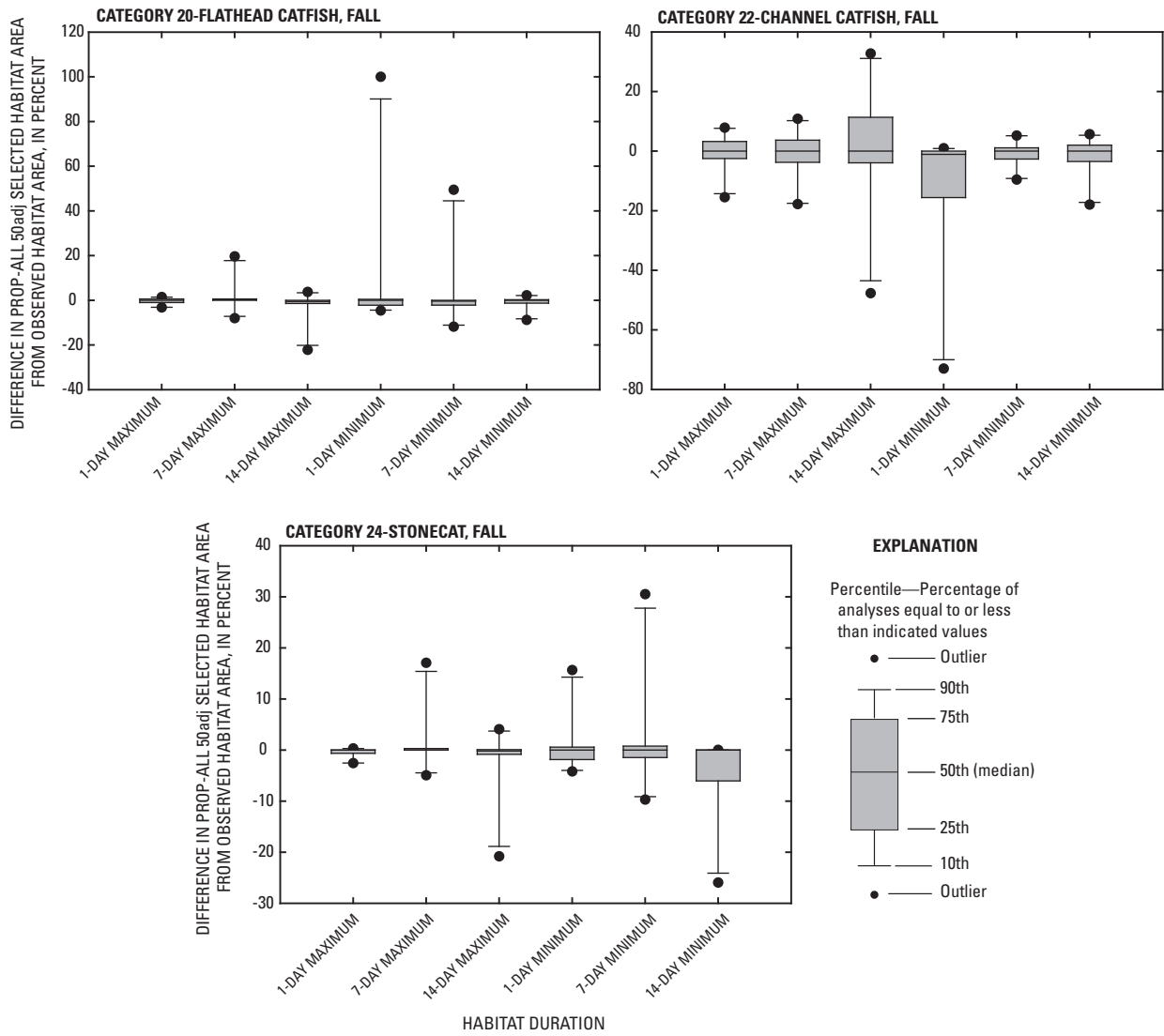


Figure 40. Distribution of percent differences between Prop-all50_{adj} and Observed fish habitat conditions for 1-, 7-, and 14-day maximum and minimum durations, by selected species/life stage categories, at the Marais des Cygnes River RCHRES 90, water years 1995 through 2004.—Continued

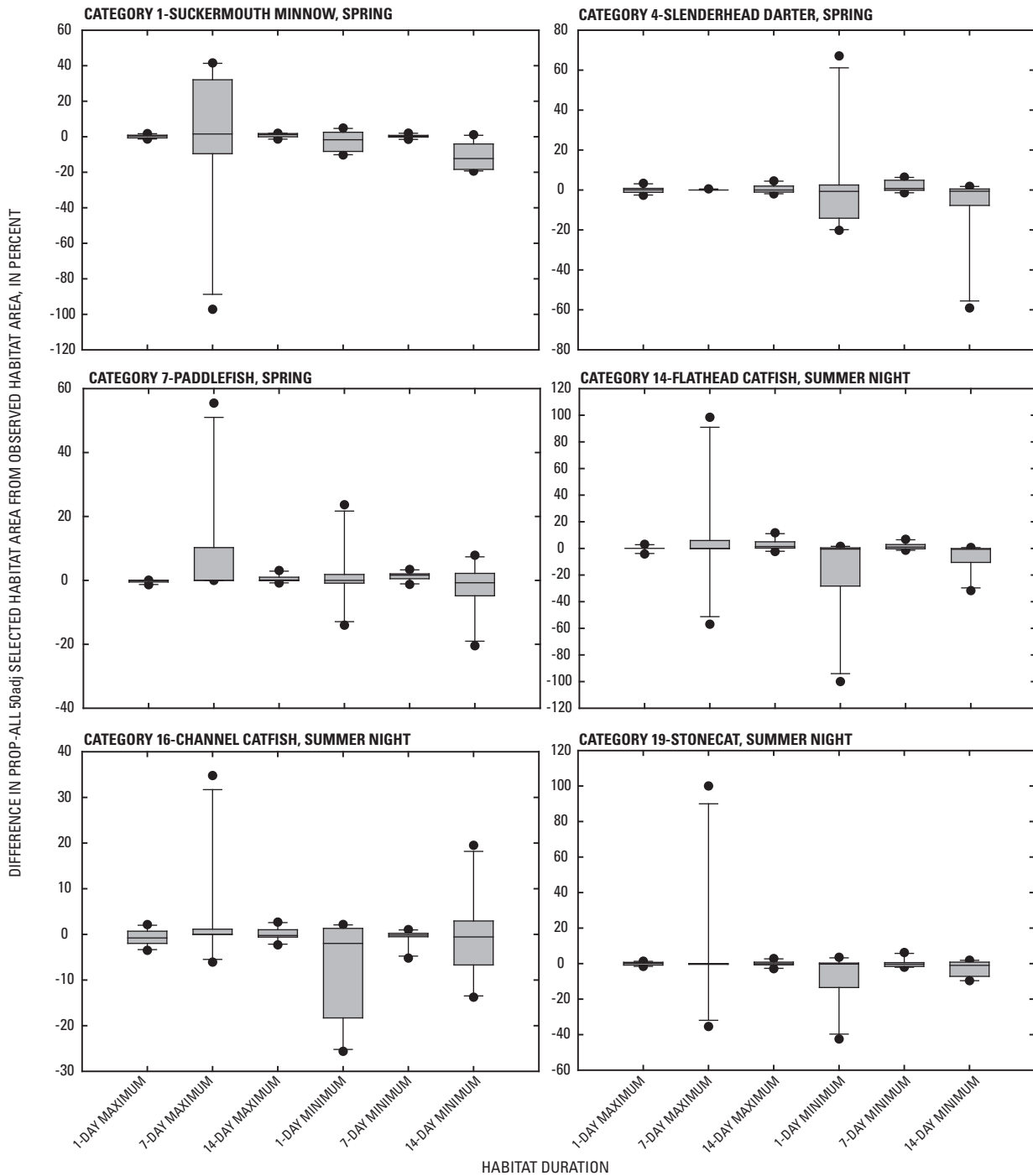


Figure 41. Distribution of percent differences between Prop-all50_{adj} and Observed fish habitat conditions for 1-, 7-, and 14-day maximum and minimum durations, by selected species/life stage categories, at the Marais des Cygnes River RCHRES 93, water years 1995 through 2004.

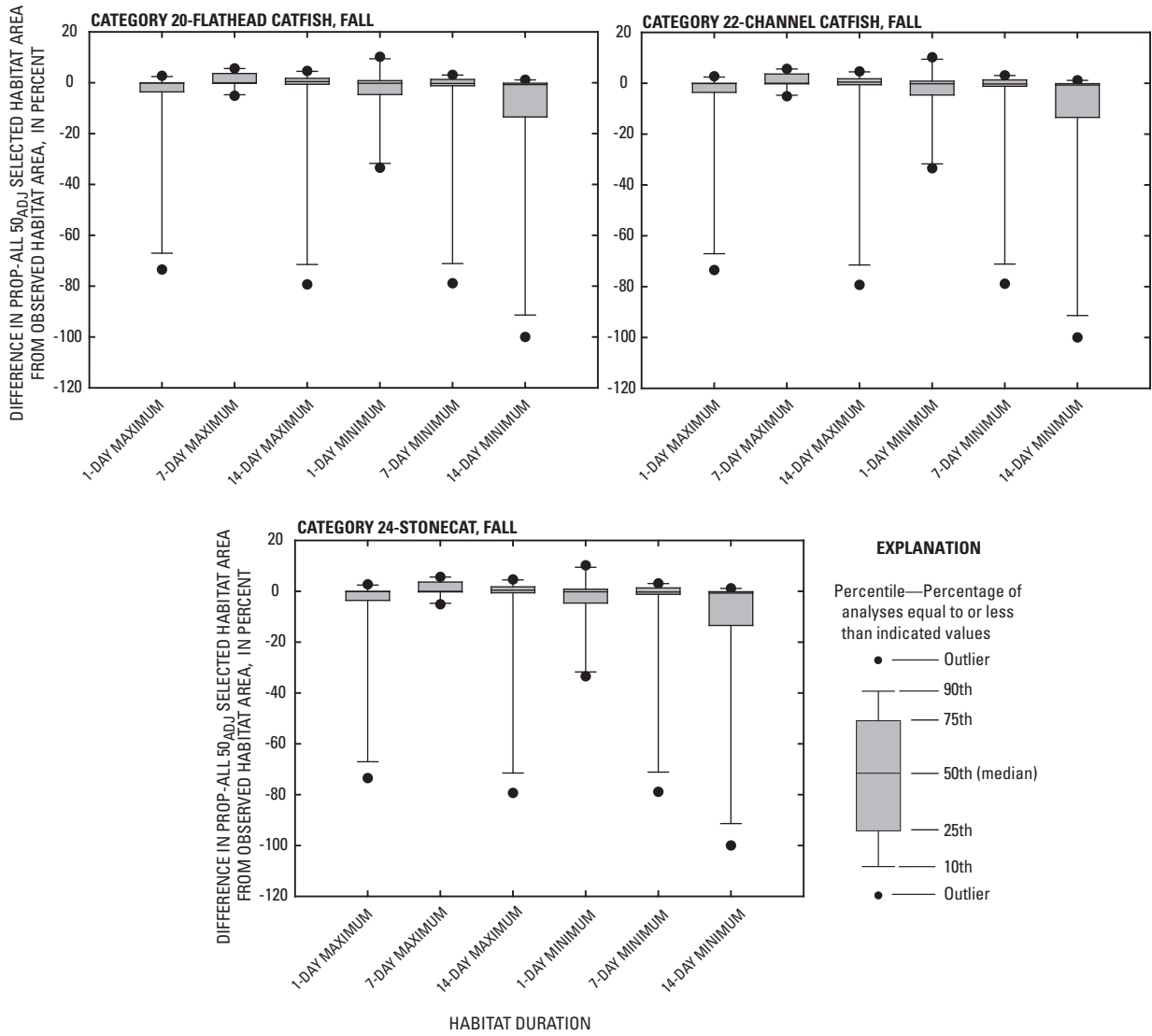


Figure 41. Distribution of percent differences between Prop-all50_{adj} and Observed fish habitat conditions for 1-, 7-, and 14-day maximum and minimum durations, by selected species/life stage categories, at the Marais des Cygnes River RCHRES 93, water years 1995 through 2004.—Continued

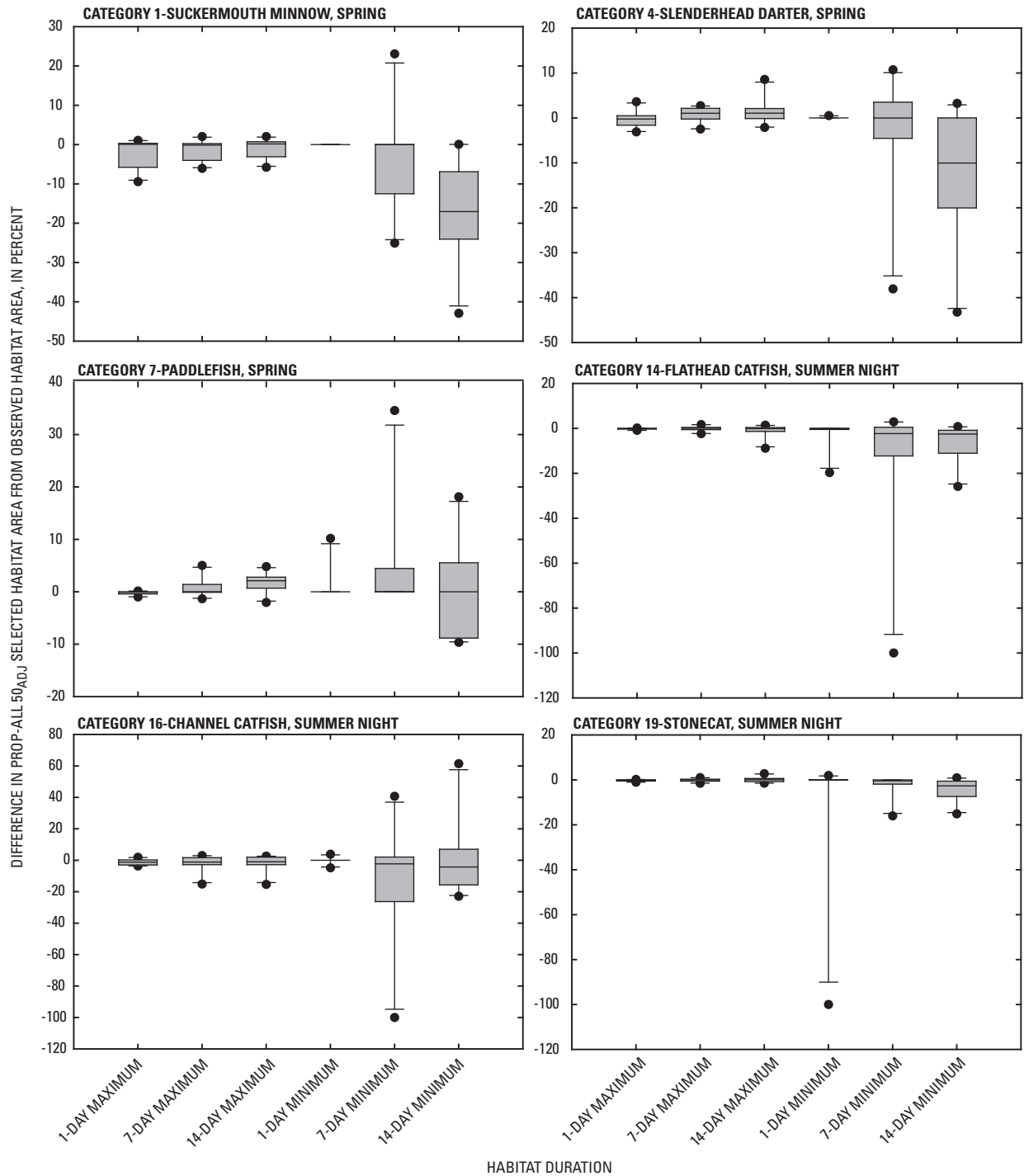


Figure 42. Distribution of percent differences between Prop-all50_{adj} and Observed fish habitat conditions for 1-, 7-, and 14-day maximum and minimum durations, by selected species/life stage categories, at the Marais des Cygnes River RCHRES 95, water years 1995 through 2004.

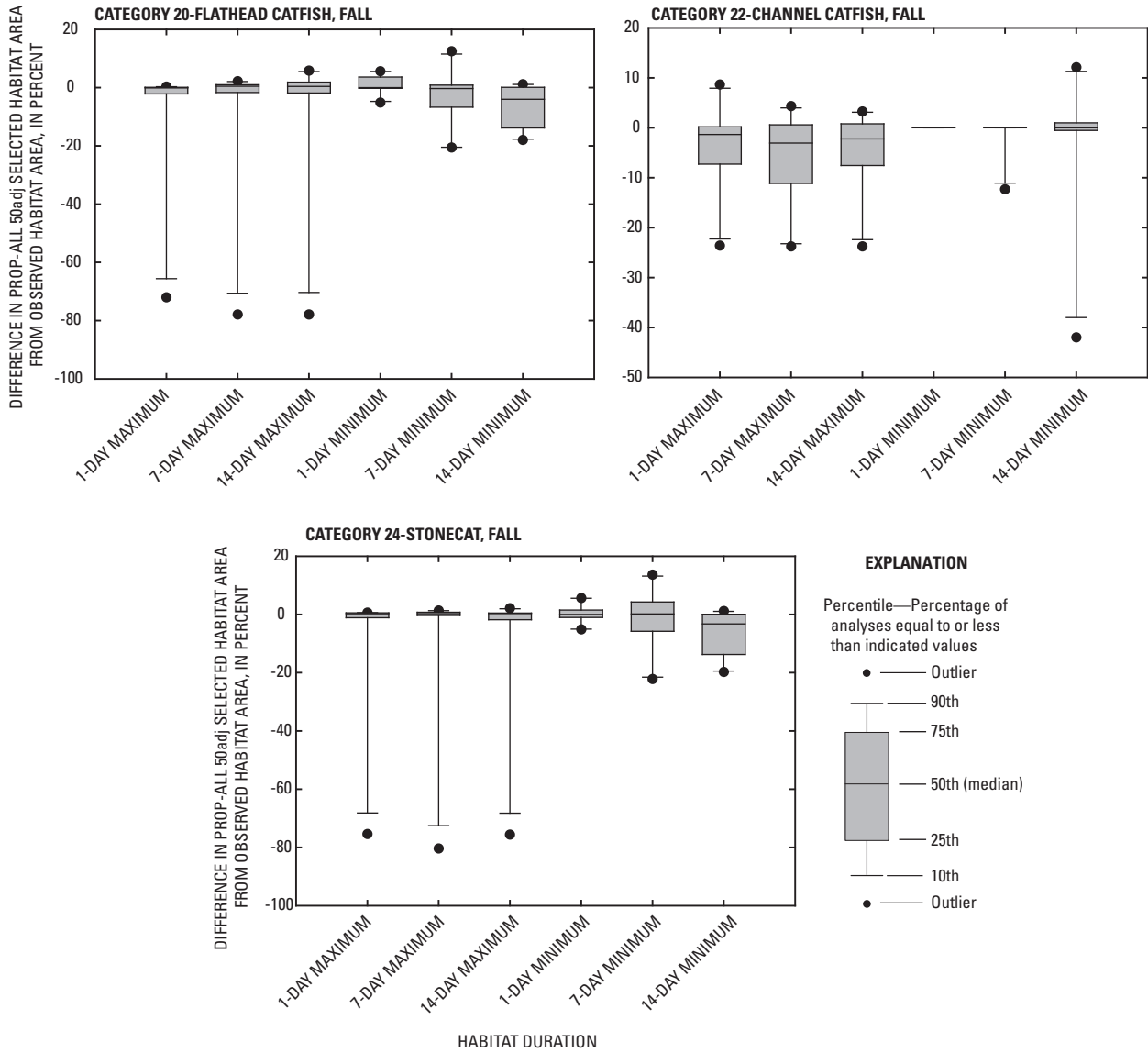


Figure 42. Distribution of percent differences between Prop-all50_{adj} and Observed fish habitat conditions for 1-, 7-, and 14-day maximum and minimum durations, by selected species/life stage categories, at the Marais des Cygnes River RCHRES 95, water years 1995 through 2004.—Continued

29–31). Median paddlefish habitat increased substantially between pre-settlement and Observed scenarios, but maximum paddlefish habitat availability was greatest under unregulated pre-settlement conditions at all three locations (figs. 36–38; tables 29–31) as unregulated peak flows were greater. Overall median suckermouth minnow habitat declined 3 to 30 m²/100 m, and slenderhead darter habitat declined between 3 and 7 m²/100 m from pre-settlement to Observed conditions, depending on location.

Overall, median paddlefish habitat area increased (0 to 170 m²/100 m) between current and proposed conditions at the Marmaton River at the Kansas-Missouri state line, Missouri (RCHRES 6), and downstream Marmaton River RCHRES 11 sites; however, median habitat generally decreased for suckermouth minnow (2 to 13 m²/100m), slenderhead darter (0 to 20 m²/100m), summer flathead catfish (0 to 44 m²/100 m), and summer stonecat (0 to 10 m²/100 m) categories (figs. 43, 44; tables 32 and 33, on compact disc, at the back of this report). Habitat for the remaining categories was similar between Current and proposed conditions.

The annual quantity of fish habitat varied by water year at the Marmaton River near state line, Missouri, but the relative differences between Current and Prop-all50 scenarios were less variable than the annual changes, and differences were more consistent than those at the Marais des Cygnes RCHRES 90 site. Habitat generally declined under the Prop-all50 scenario compared with current conditions (fig. 45); this relation was more consistent between years and categories than at the Marais des Cygnes River RCHRES 90 site (fig. 39). One possible explanation for the greater habitat variability between scenarios at the Marais des Cygnes site than at the Marmaton River site is the greater complexity in channel geometry resulting from secondary levels of in-channel substrate deposits leading to a greater variability in habitat-streamflow relations (Heimann and others, 2005). Declines in annual 1-, 7-, or 14-day minimum habitat were greater than 10 percent for 1 or more years for all categories at both Marmaton River locations, except for spring paddlefish habitat at RCHRES 6, which generally remained unchanged between Current and proposed scenarios (figs. 46–47). Declines in 1-, 7-, or 14-day proposed minimum habitat availability were at or near 100 percent for 1 or more years for slenderhead darter, summer flathead catfish, fall channel catfish, and fall stonecat habitat categories at RCHRES 6 and for spring suckermouth minnow, spring slenderhead darter, summer channel catfish, summer stonecat, and fall flathead catfish habitat at RCHRES 11 (figs. 46–47).

Generally, overall median habitat area for suckermouth minnow (2 to 7 m²/100m), summer and fall flathead catfish (summer 4 to 28 m²/100m, fall 0 to 9 m²/100 m), fall channel catfish (0 to 3 m²/100 m) and fall stonecat (0 to 13 m²/100 m) habitat declined from pre-settlement to Current conditions at the two Marmaton River reporting locations (figs. 43, 44; tables 32, 33). Median paddlefish habitat increased 1 to 10 m²/100 m from pre-settlement to Current conditions at the RCHRES 11 location, but remained unchanged at the

RCHRES 6 location. Maximum paddlefish habitat area was greater for Current simulated conditions at both Marmaton River locations compared with pre-settlement conditions (figs. 43, 44; tables 32, 33).

Simulations indicate that alterations in streamflow have resulted in substantial changes in the availability of habitat in the Marais des Cygnes and Marmaton Rivers between pre-settlement and present-day conditions and through possible proposed conditions. These changes in habitat availability under various simulated flow conditions can be quantified, and are shown to vary with fish species, life stage, season, and year-to-year flow variability.

Summary and Conclusions

This report summarizes results from a study to estimate the effects of impoundments, land-cover changes, and reported point-source withdrawals and discharges on streamflows in the 5,410-square mile upper Osage River Basin in Missouri and Kansas. Hydrologic models developed using the Hydrologic Simulation Program-FORTRAN were calibrated and validated to current (1995–2004 water years) regulation and water-use conditions, and modified to simulate changes in regulation and water-use conditions under pre-settlement and proposed-regulation conditions for the same period of record. Analyses included quantification of changes in the magnitude, frequency, duration, and timing of streamflows under each scenario. Output from simulation scenarios were used in conjunction with known streamflow-fish habitat relations to quantify effects of altered flow on fish habitat area.

Analyses of simulated runoff and evapotranspiration for the Marmaton River Basin (1995 through 2004 water years) provided an indication of the effects of land-cover and regulation changes on the water balance in the study basins. The change from historical land cover to current land cover resulted in a decline of 120 thousand acre-feet of total runoff, indicating that the combined land-cover changes (100 thousand acre-feet) and net water use losses (20-thousand acre feet) might have a greater effect on total runoff than impoundments. The effects of a conversion of cultivated row crops back to pre-settlement native prairie soils were simulated using an increase in the infiltration model parameter for the pre-settlement prairie/rangeland land cover. This parameter modification accounted for a greater difference in total runoff between pre-settlement and current/proposed scenarios than other changes in land cover or from impoundments. The simulated increase in soil infiltration capacity under native prairie conditions also resulted in lower peak flows for the pre-settlement model scenario compared with current/proposed scenarios. Impoundments decreased hydrograph peaks and extended the recession limb of the hydrographs in the simulations; increased infiltration had similar effects. Evapotranspiration from the land and open-water surfaces varied little between current and proposed scenarios, but could account for differ-

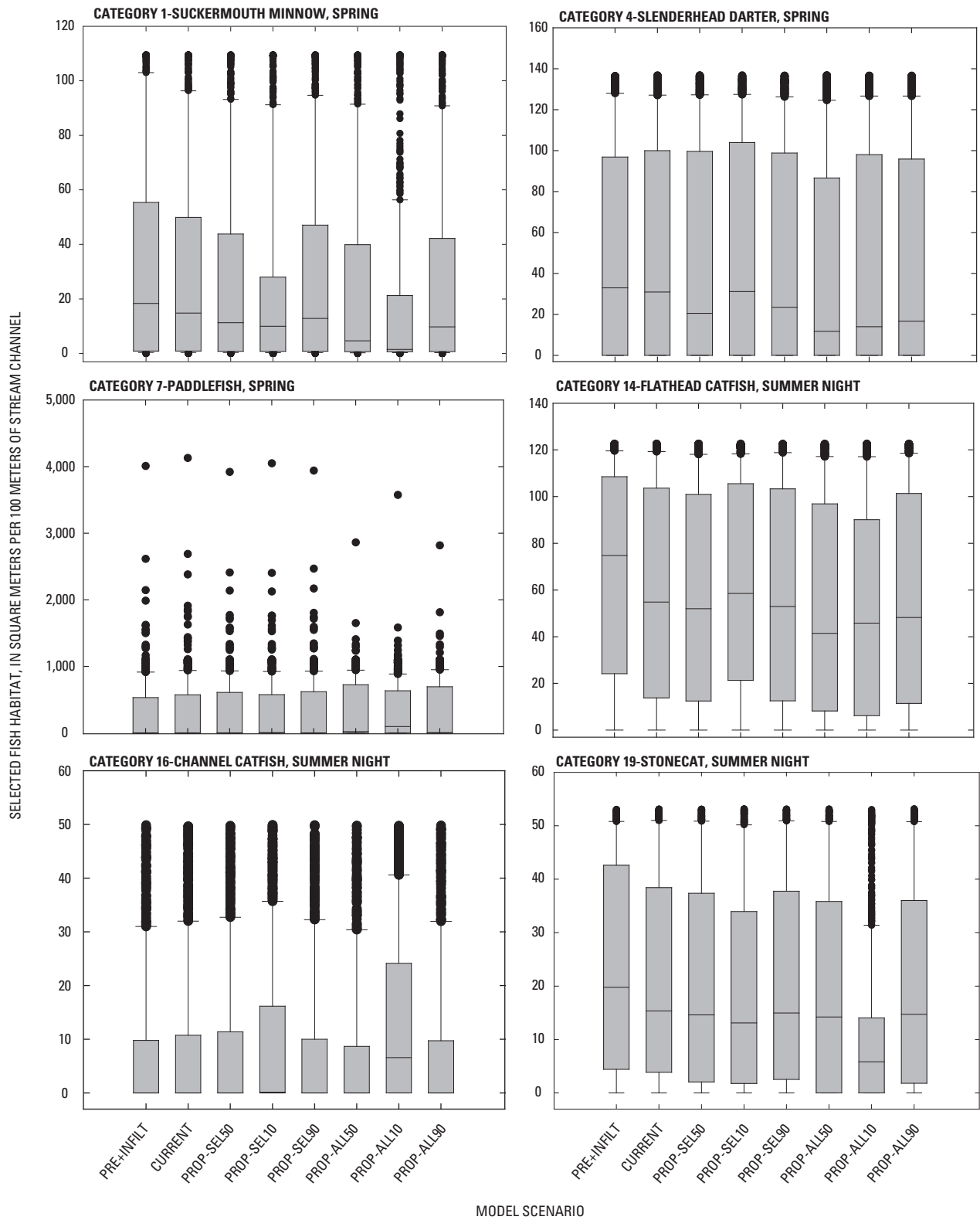


Figure 43. Distribution of daily fish habitat, by streamflow scenario and selected species/life stage categories, at the the Marmaton River near the Kansas-Missouri state line, Missouri (RCHRES 6), water years 1995 through 2004.

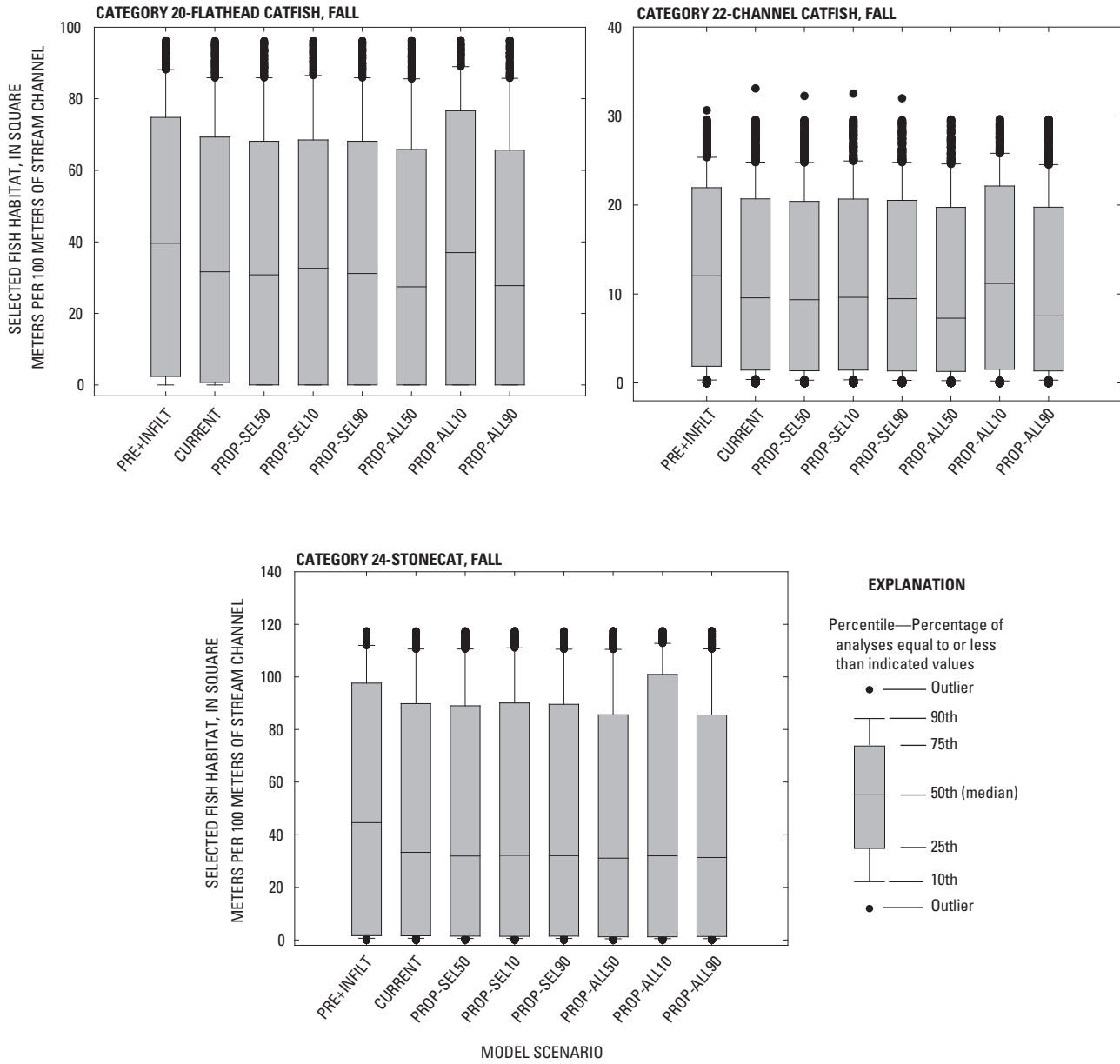


Figure 43. Distribution of daily fish habitat, by streamflow scenario and selected species/life stage categories, at the the Marmaton River near the Kansas-Missouri state line, Missouri (RCHRES 6), water years 1995 through 2004.—Continued

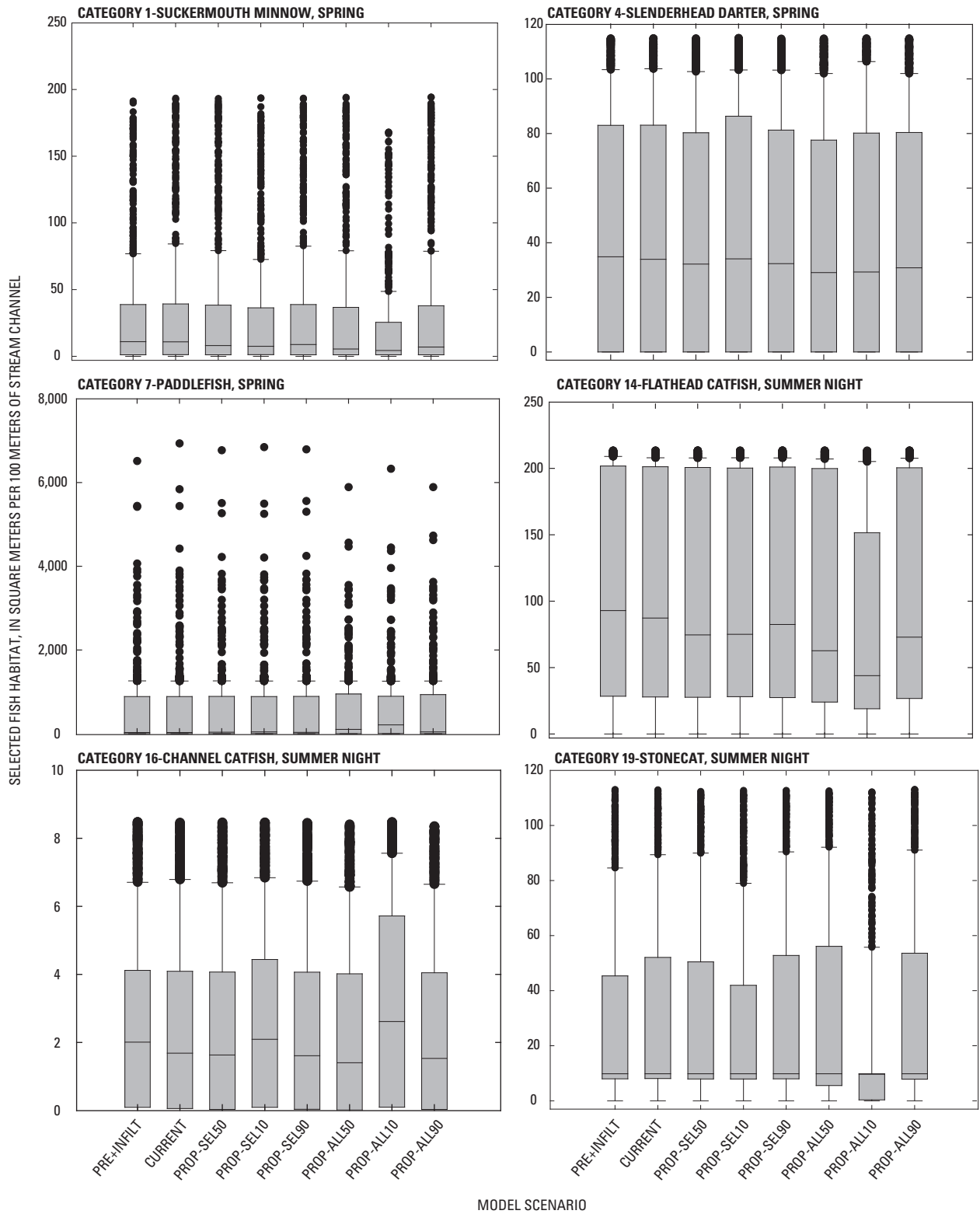


Figure 44. Distribution of daily fish habitat, by streamflow scenario and selected species/life stage categories, at the Marmaton River RCHRES 11, water years 1995 through 2004.

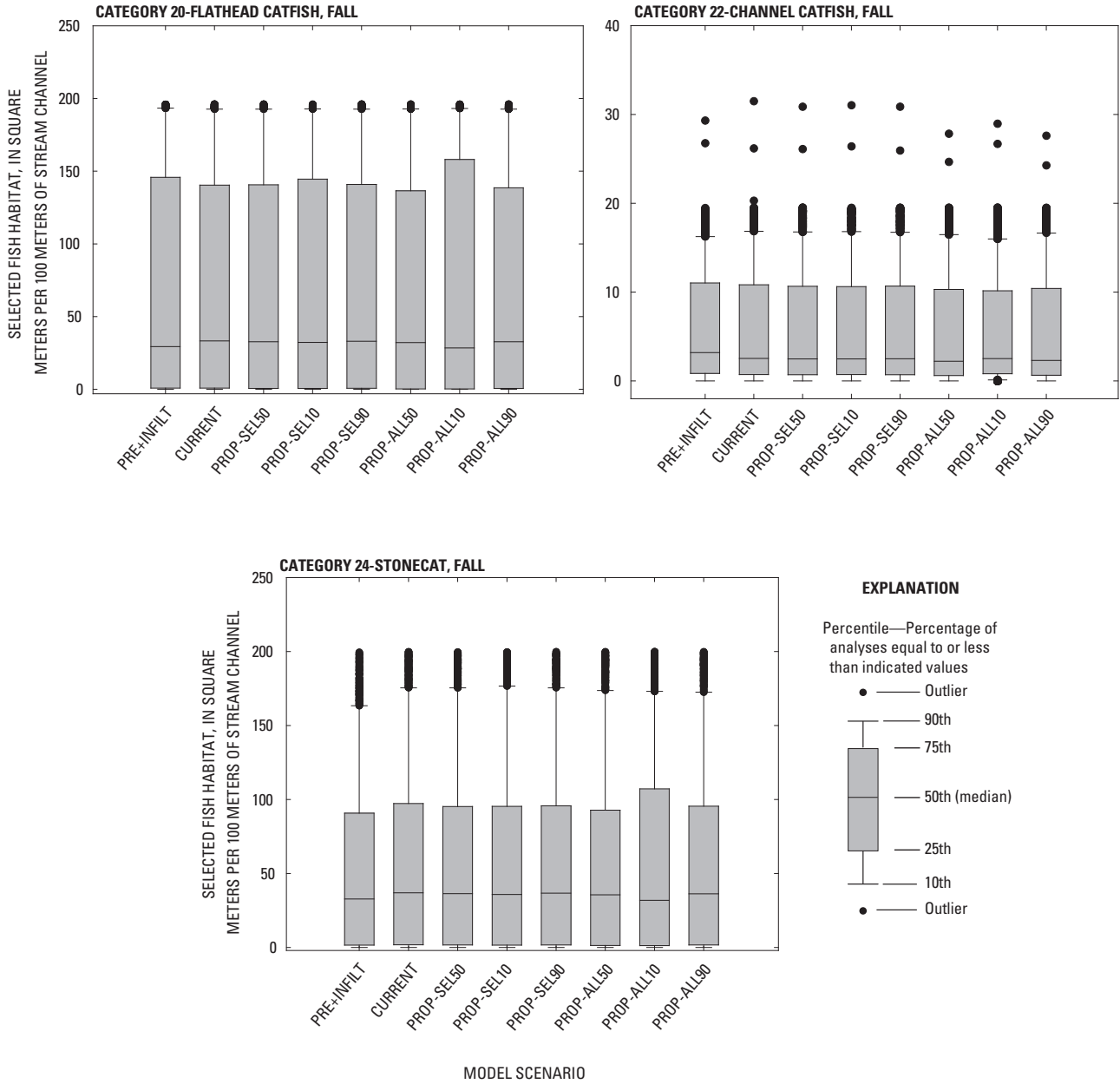


Figure 44. Distribution of daily fish habitat, by streamflow scenario and selected species/life stage categories, at the Marmaton River RCHRES 11, water years 1995 through 2004.—Continued

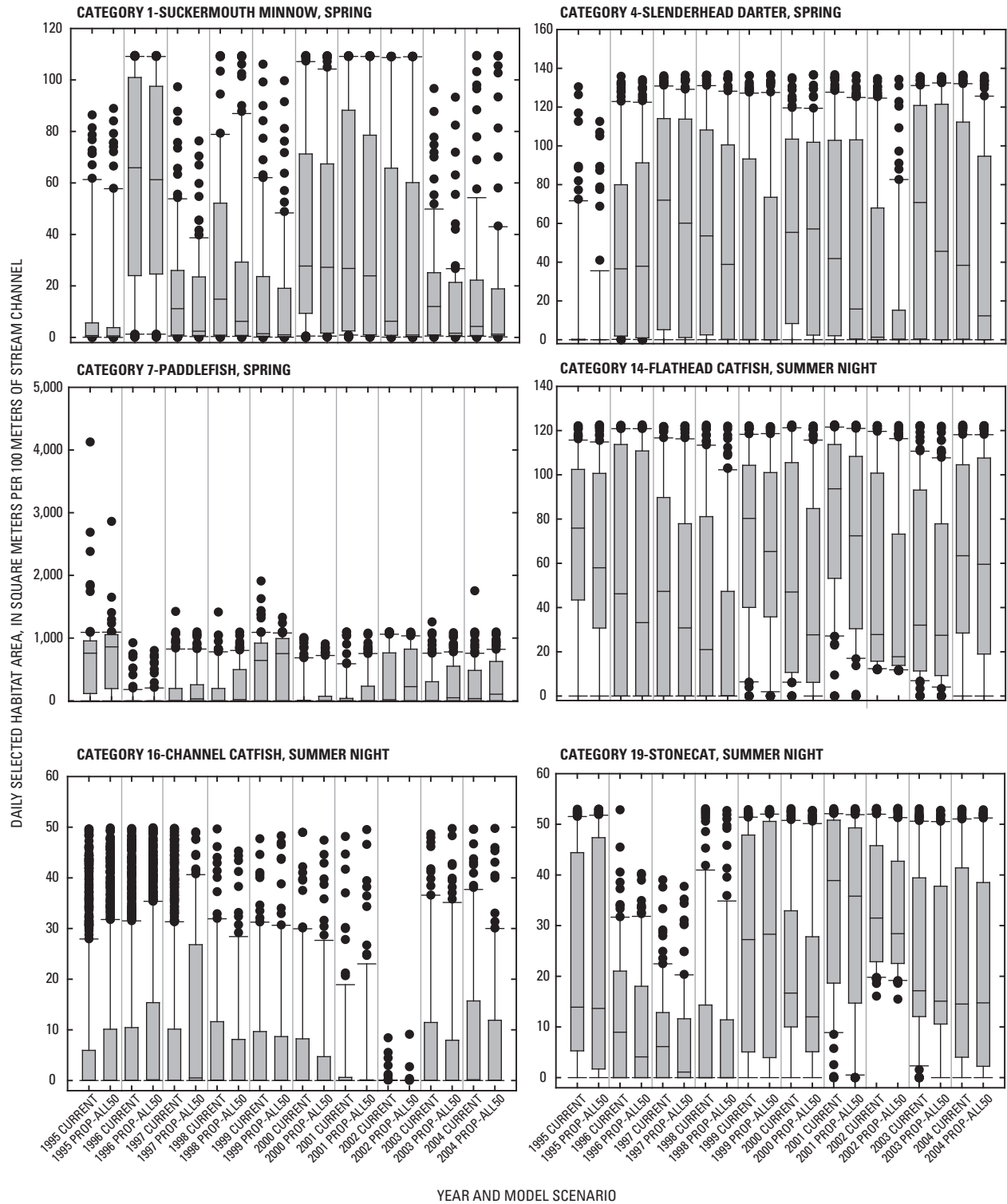


Figure 45. Comparison of distributions of daily fish habitat for Current simulated and Prop-all50 streamflow scenarios, by year and select species/life stage categories, at the Marmaton River near the Kansas-Missouri state line, Missouri (RCHRES 6), water years 1995 through 2004.

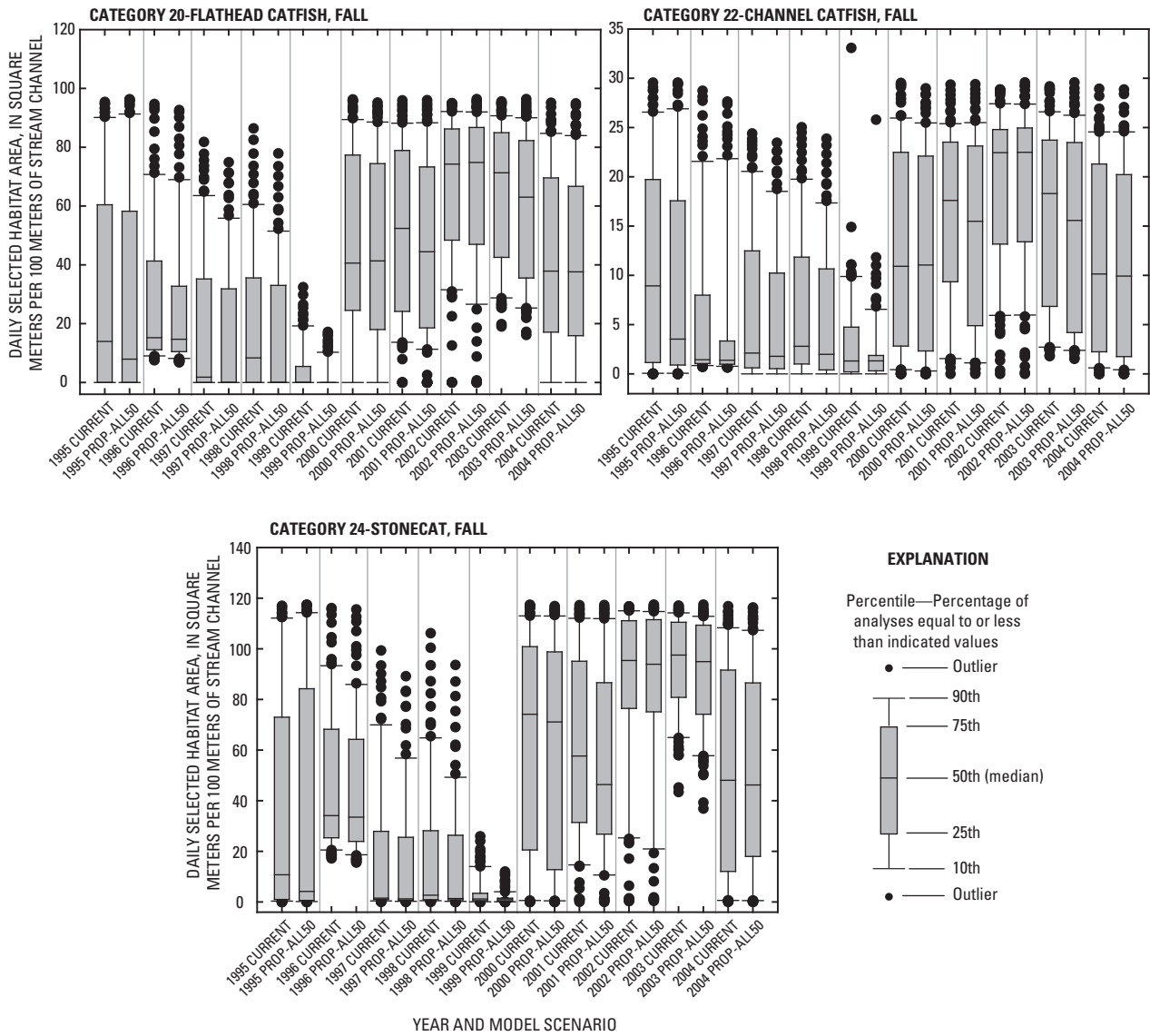


Figure 45. Comparison of distributions of daily fish habitat for Current simulated and Prop-all50 streamflow scenarios, by year and select species/life stage categories, at the Marmaton River near the Kansas-Missouri state line, Missouri (RCHRES 6), water years 1995 through 2004.—Continued

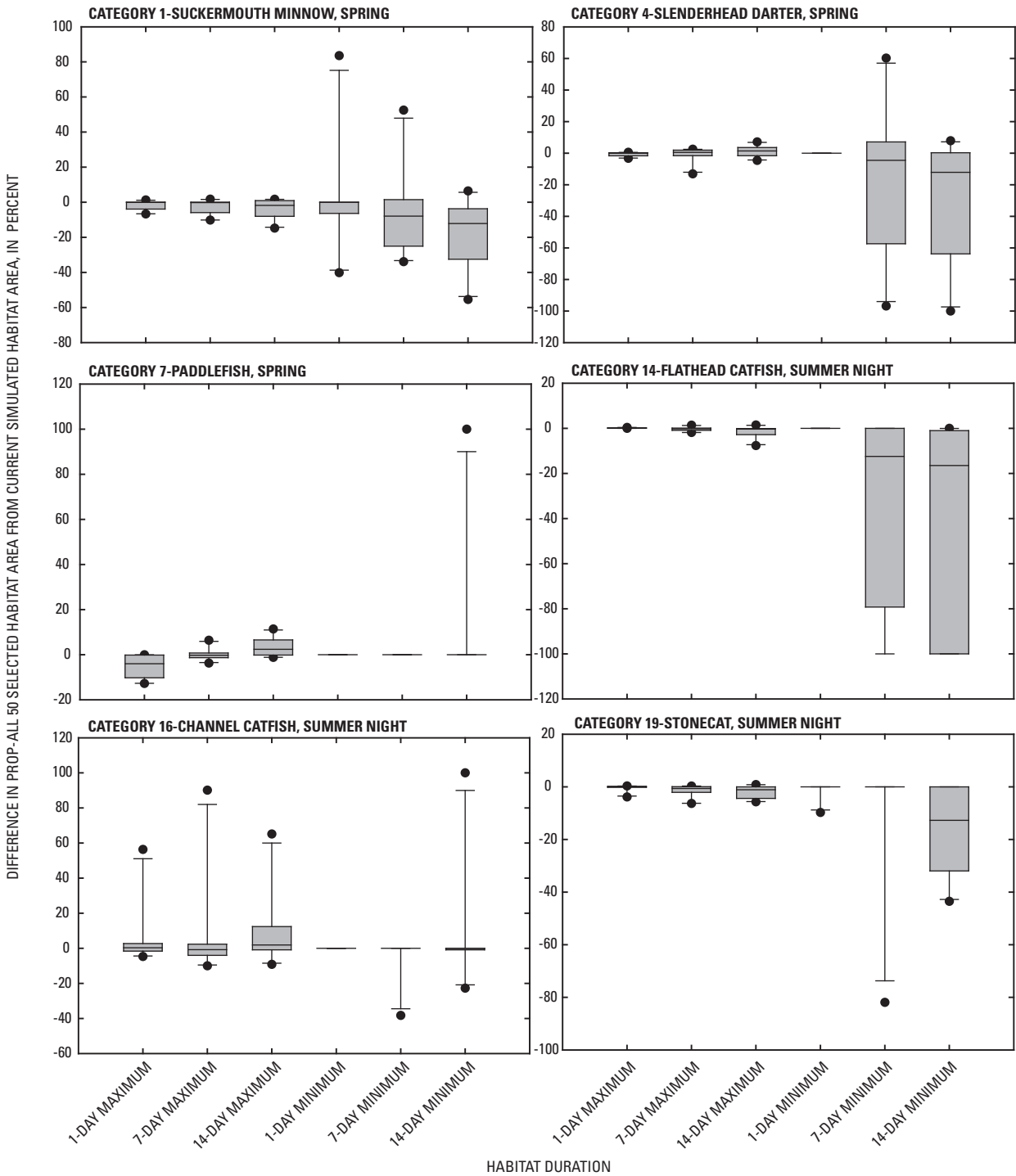


Figure 46. Distribution of percent differences between Prop-all50 and Current simulated fish habitat for 1-, 7-, and 14-day maximum and minimum durations, by selected species/life stage categories, at the Marmaton River near the Kansas-Missouri state line, Missouri (RCHRES 6), water years 1995 through 2004.

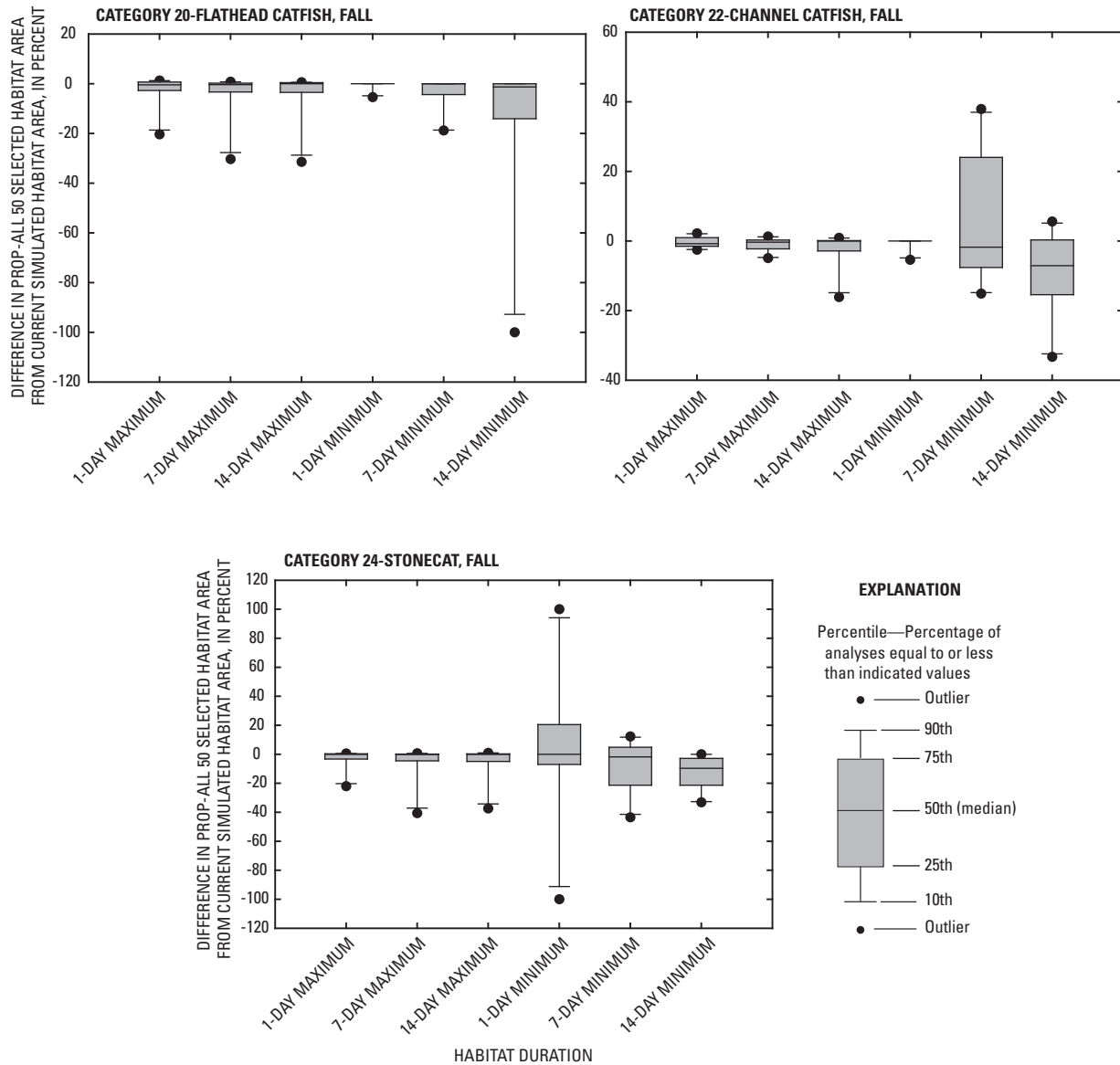


Figure 46. Distribution of percent differences between Prop-all50 and Current simulated fish habitat for 1-, 7-, and 14-day maximum and minimum durations, by selected species/life stage categories, at the Marmaton River near the Kansas-Missouri state line, Missouri (RCHRES 6), water years 1995 through 2004.—Continued

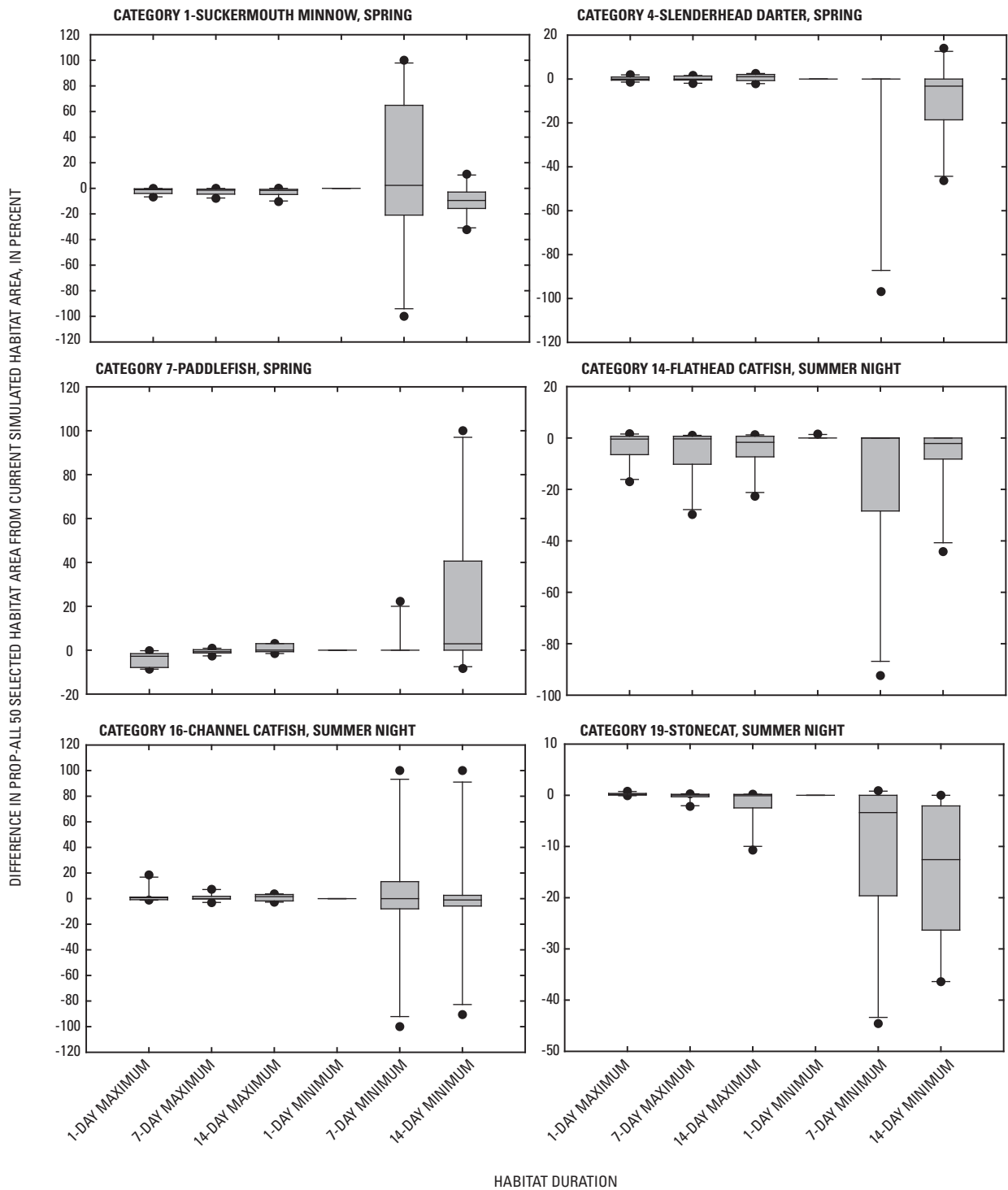


Figure 47. Distribution of percent differences between Prop-all50 and Current simulated fish habitat for 1-, 7-, and 14-day maximum and minimum durations, by selected species/life stage categories, at the Marmaton River RCHRES 11, water years 1995 through 2004.

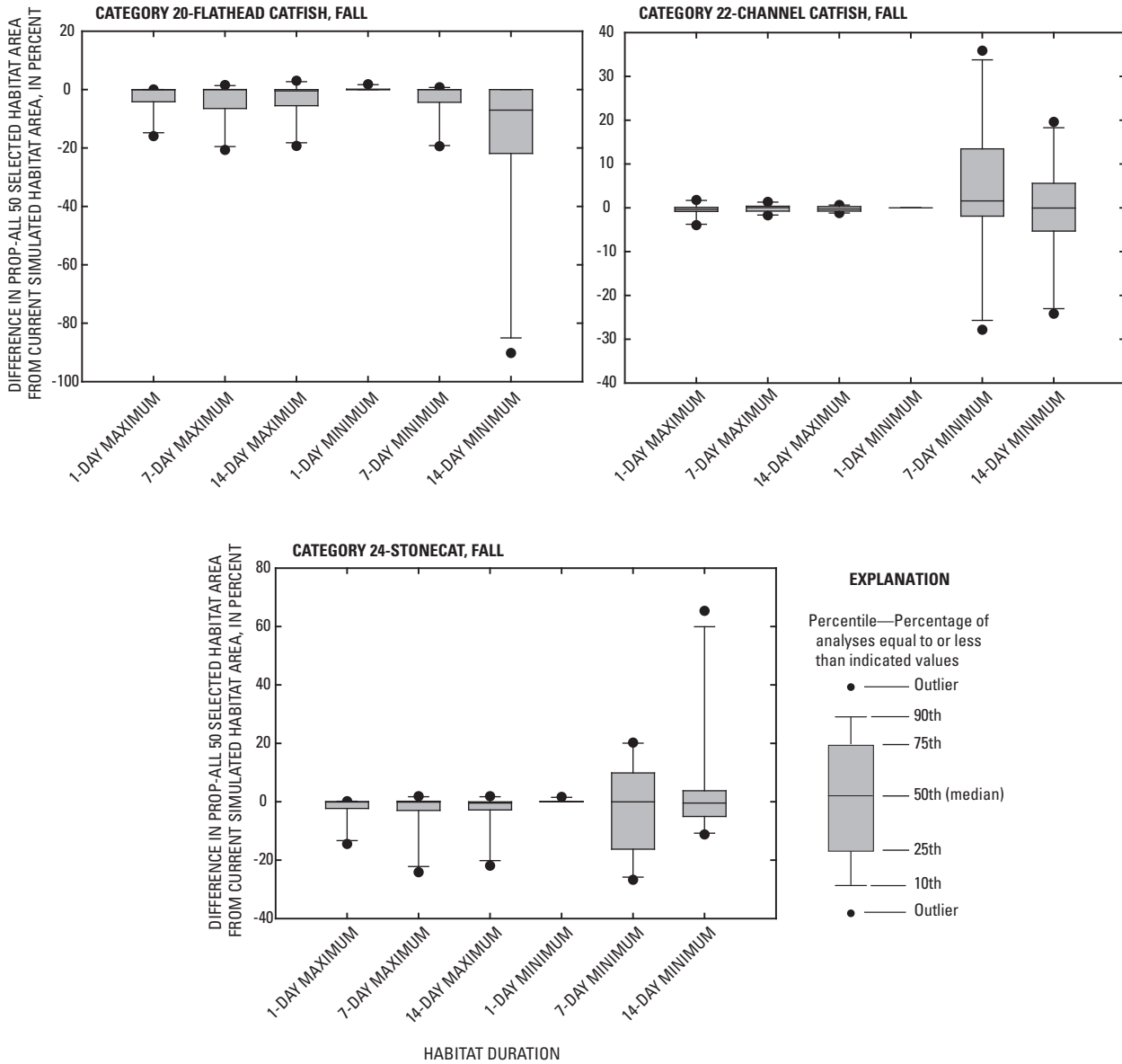


Figure 47. Distribution of percent differences between Prop-all50 and Current simulated fish habitat for 1-, 7-, and 14-day maximum and minimum durations, by selected species/life stage categories, at the Marmaton River RCHRES 11, water years 1995 through 2004.—Continued

ences in total runoff between pre-settlement and the current and proposed scenarios. Evaporative water losses increased with the addition of impoundments, and while these increases did not have a substantial effect on the total runoff they could have an effect on flows during dry periods. Greater detention associated with impoundments resulted in longer hydrograph recessions and lowered peak flows. Outflows for a single, isolated impoundment indicated that low and high flows may be affected by design characteristics. The 10th-percentile design scenario, representative of smaller outflow structures and longer detention time, resulted in the longest hydrograph recessions and lowest peak flows. The 90th-percentile design scenario, representative of larger possible outflow structures and shorter detention time, resulted in the fastest hydrograph recessions and highest peak flows of the three proposed impoundment design scenarios.

Differences in simulated monthly minimum 1-, 3-, 7-, and 30-day low flows from observed flows at the Marais des Cygnes River near the Kansas-Missouri state line, Kansas, for the 10-year simulation period, indicated that the proposed design scenario with the greatest detention characteristics resulted in the least decreases in low flows relative to observed flows, whereas the proposed scenario with the least detention characteristics resulted in the greatest occurrence of low flow declines. The longest period of extended declines (5 months) occurred in the summer months of 2001 and 2002. October had the greatest magnitude of declines in proposed monthly minimum 1-, 3-, 7-, and 30-day low flows compared with observed flows. The greatest declines for proposed conditions were for the lowest 10 percentile of observed flows and during the driest years (2000, 2001 water years). The most apparent difference in minimum flows for the simulation scenarios at the Marais des Cygnes River near the Kansas-Missouri state line, Kansas, were between observed and pre-settlement scenarios.

There were no additional impoundments proposed for the Little Osage River Basin and, therefore, proposed scenarios were not simulated for this basin, but the pre-settlement low flows generally were greater than observed flows at the Little Osage River near Fulton, Kansas. Pre-settlement low flows seemed to be particularly better sustained than observed flows during dry years (2000–2003 water years), and with increasing flow duration period (from 1-day to 30-day periods).

Low flows in the Marmaton River Basin generally were lower for the current simulated or observed scenarios than any other simulated conditions. In a small headwater basin in the upper Marmaton River Basin, simulated declines in minimum flows were small [generally less than 6 cubic feet per second (ft³/s)] and less than 1 ft³/s for 1- and 3-day scenarios), but resulted in 10 to 18 additional zero flow days for the proposed scenarios relative to current simulated conditions. Declines occurred primarily during the summer months. Simulated pre-settlement minimum flows generally were greater than current simulated flows for all flow duration periods, with maximum monthly differences generally less than 10 ft³/s. Differences between simulated and observed minimum monthly flows

were similar at the Marmaton River near Marmaton, Kansas, site to those at the headwater site. Reductions in minimum monthly flows as a result of additional impoundments generally were less than 5 ft³/s and resulted in 6 additional zero flow days. Simulated pre-settlement monthly minimum flows generally were greater than observed flows, with maximum monthly reductions in minimum flows between pre-settlement and observed flows of 36 to 71 ft³/s. Similar to the headwater and Marmaton, Kansas, locations, the reductions in flows at the Marmaton River near the Kansas-Missouri state line, Missouri, as a result of proposed impoundments, were most frequent for the scenario representing the least impoundment detention and generally occurred in the summer months. The greatest declines between proposed and current flows generally occurred in the lower 50 percentile of current simulated flows and during the drier water years of 2001–2003. Proposed conditions resulted in declines in the 0–10 percentile flow values for the 1-, 3-, and 7-day duration periods. Pre-settlement minimum flows generally were greater than current simulated scenario flows for all flow duration periods, with maximum declines in monthly low flows of 150 to 165 ft³/s between pre-settlement and current scenarios.

Simulated proposed monthly high flows were reduced as a result of the addition of impoundments in all basins and at all reporting locations. Observed 1-day, monthly maximum streamflows at the Marais des Cygnes near the Kansas-Missouri state line, Kansas, were, on average, 5 to 17 ft³/s greater than proposed maximum monthly flows. Pre-settlement 1-day monthly maximum flows were, on average, 540 to 776 ft³/s greater than observed 1-day monthly maximum flows, indicating that, similar to low flows, the characteristics of high flows in this basin also have undergone greater changes between pre-settlement to current conditions than would be expected from current to proposed conditions. These changes between pre-settlement and current conditions in this basin were the result of several large impoundments with managed detention and outflow characteristics. Observed monthly maximum flows at the Little Osage River near Fulton, Kansas, were similar to pre-settlement monthly maximums. The observed or current simulated maximum monthly flows were, on average, greater than historical or proposed flow conditions in the Marmaton River Basin. Current monthly maximum flows in the Marmaton River near the Kansas-Missouri state line, Missouri, were 46 to 300 ft³/s greater than proposed monthly maximum flows, depending on impoundment outflow design. The simulations with the greatest detention capabilities of impoundments resulted in the largest differences in high flows. Observed 1-day maximum monthly flows at this site were, on average, 109 ft³/s greater than pre-settlement scenario flows.

An ecological consequence of a reduction in streamflow magnitude is a decrease in flood frequency and flood-plain inundation. The flood frequency for the Marais des Cygnes River near the Kansas-Missouri state line, Kansas, was substantially decreased between pre-settlement and observed conditions, but observed and proposed conditions were similar. This decrease in flooding can be attributed to the substantial

amount of current, controlled regulation in the Marais des Cygnes River Basin. There were no differences between pre-settlement and observed flood frequency at the Little Osage River near Fulton, Kansas. Flood frequency generally was greatest for the current simulated or observed scenarios in the Marmaton River Basin and least for the proposed conditions. The effects of regulation on flood frequency decreased downstream from the Kansas-Missouri state line. The flood frequency of the proposed scenarios were 54 to 60 percent less, depending on outflow and detention characteristics, than current simulated conditions at the Marmaton River near the Kansas-Missouri state line, Missouri. The reduction in flood frequency was greatest in the scenario incorporating the greatest estimated impoundment detention characteristics. At a downstream site, the total reduction in flood frequency from current to Prop-all conditions was less (21 to 28 percent) than the state line site. The downstream site is below a primary Marmaton River tributary with little local regulation and, therefore, the effects of proposed upstream regulation were reduced.

The simulated effects of regulation on large flood durations varied between model basins and with degree of regulation. Increased regulation increased the total hydrograph duration (duration from pre-event base flow to post-peak base flow) of floods. The average duration of large flood hydrographs was similar for observed and proposed scenarios at the Marais des Cygnes River near Kansas-Missouri state line, Kansas. Pre-settlement scenario flood hydrograph durations were, on average, about 80 percent less than observed. The average large flood hydrograph duration at the Little Osage River near Fulton, Kansas, for the pre-settlement scenario, using assumed increased infiltration estimates under historical land cover, was about 120 percent greater than observed conditions. At the Marmaton River headwater and Marmaton River near Marmaton, Kansas, locations, the average large flood hydrograph durations under proposed and pre-settlement scenarios were about 50 percent greater than observed and current simulated scenarios. At the Marmaton River near the Kansas-Missouri state line, Missouri, the average proposed scenario total flood hydrograph durations were 2 to 70 percent greater than current simulated conditions depending on scenario, whereas flood hydrograph durations were similar under pre-settlement and current simulated conditions.

Although the total hydrograph duration increased with regulation, the actual duration of streamflows above estimated flood levels (flooding period) decreased with proposed regulation in all model basins. The flooding period at the Marais des Cygnes River near the Kansas-Missouri state line, Kansas, decreased about 2 to 7 percent from observed to proposed conditions. The largest difference in flooding periods at this Marais des Cygnes River site occurred between pre-settlement and observed scenarios in which pre-settlement average flooding periods were 200 percent greater than observed conditions. Pre-settlement flooding periods in the Little Osage River near Fulton, Kansas, were similar (within 4 percent) to observed conditions. At the Marmaton River near the Kansas-Missouri

state line, Missouri, the average flooding period under proposed scenarios showed a 60 percent reduction from current simulated conditions.

The timing of floods during the 10-year simulation period in the model basins was not substantially affected by regulation, although the timing of low flow extremes was altered in some basins. The timing of Little Osage River extreme low flows were similar between pre-settlement and observed flow scenarios, but the pre-settlement large flood timing was in June rather than September, as for observed flows. At the Marmaton River near Marmaton, Kansas, the timing of peak flows was similar between scenarios, but the timing of extreme low flows was August for pre-settlement conditions and November for observed and most proposed scenarios. The timing of low and high flows was similar under all scenarios at the Marmaton River near the Kansas-Missouri state line, Missouri.

The ecological consequences of altered conditions also were assessed through changes in in-stream habitat. Comparisons of 10-year daily habitat area distributions for nine selected fish species/life stage categories under the varying streamflow simulation scenarios indicated that the effect of flow alteration on fish habitat varied by basin, scenario, time distribution, and fish species/life stage category. Of particular concern in comparing differences in habitat by scenario were any possible declines in the minimum habitat availability that could lead to greater limiting conditions or “bottlenecks” in fish habitat availability and possible habitat declines of extended (7–14 days) durations. Minimum annual 1-, 7-, or 14-day habitat availability for the prop-all50_{adj} simulation declined at each of three Marais des Cygnes River sites by more than 10 percent compared with observed conditions for 1 or more years for each of the nine seasonal fish habitat categories, indicating that habitat bottlenecks may be greater during proposed conditions for some categories under some seasons. Declines in minimum habitat availability were at or near 100 percent for 1 or more years for summer flathead catfish, fall flathead catfish, fall channel catfish, and fall stonecat habitat categories at one Marais des Cygnes River location and for summer flathead catfish, summer channel catfish, and summary stonecat at another Marais des Cygnes River location. This indicates that habitat for these categories may be eliminated for 1 to 14 days during some period of some years under proposed conditions. Habitat generally declined at the Marmaton River near the Kansas-Missouri state line, Missouri, under the proposed scenario compared with current conditions. Declines in annual 1-, 7-, or 14-day minimum habitat were greater than 10 percent for 1 or more years for all categories at both Marmaton River reporting locations, except for spring paddlefish habitat, which generally remained unchanged between current and proposed scenarios at one location. Declines in 1-, 7-, or 14-day proposed minimum habitat availability were at or near 100 percent for 1 or more years for slenderhead darter, summer flathead catfish, fall channel catfish, and fall stonecat habitat categories at one Marmaton River location and for spring suckermouth minnow, spring

slenderhead darter, summer channel catfish, summer stonecat, and fall flathead catfish habitat at another.

The cumulative effects of impoundments and land-cover changes were determined to substantially alter streamflows in the upper Osage River Basin in simulations spanning pre-settlement to proposed future conditions. The impoundments in these basins were designed, located, and constructed to address local considerations and the hydrologic simulations provided a means of quantifying the cumulative effects of these local projects. The degree of streamflow alteration varied between major subbasins. Streamflows in the Marais des Cygnes River Basin were altered between pre-settlement and current conditions, primarily by major impoundments, with smaller changes expected with proposed regulation. Streamflows in the Little Osage River Basin were relatively unchanged between pre-settlement and current conditions with land-cover changes (primarily the conversion of native prairies to cultivated land) affecting flows more than the few current impoundments in this basin. The current peak flows in the Marmaton River Basin were higher than pre-settlement or proposed scenario peak flows. Of the three major subbasins, this basin is likely to be the most affected by proposed impoundments.

Simulations for differing impoundment outflow designs indicated that outflow design considerations can make a substantial difference in low and high flows, particularly in the Marmaton River Basin. The primary effects of impoundments on flows were to detain event peaks and slowly release detained flows, thereby extending recessions, but the retention of flow also can lead to additional zero flow days compared with current conditions. Unknown potential changes in water-use conditions also may be a consideration under future streamflow conditions, particularly during extreme low-flow conditions. Proposed conditions will result in alterations in streamflows in the upper Osage River Basin; therefore, the ecological effects of these flow alterations also were of interest. Streamflow alterations resulted in quantified changes in in-stream fish habitat with the potential of magnifying habitat bottlenecks for some species, locations, and years providing managers with another means of assessing the effects of proposed alterations. Whereas flood storage and peak flow reduction is a primary purpose for the construction of many of the impoundments, the ecological consequences of reduced flood frequency and magnitude under proposed conditions on the maintenance and function of riparian systems, particularly in the Marmaton River Basin, are unknown.

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Publishing support provided by:
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