## K. Georges Bank Winter Flounder by Lisa Hendrickson

### 1.0 Background

The Georges Bank winter flounder stock was last assessed in October 2002 during a Groundfish Assessment Review Meeting (GARM) meeting (NEFSC 2002c). The assessment updated the SAW 34 (NEFSC 2002a) formulation of a biomass dynamics model known as ASPIC Prager 1995) that incorporated landings (1964-2001) and biomass indices from the NEFSC autumn (1963-2001) and spring (1968-2002) bottom trawl surveys. Model results indicated a reasonable fit to the input data with no strong retrospective pattern in the fishing mortality or biomass estimates. Fishing mortality rates were at or below $\mathrm{F}_{\text {MSY }}$ during 1995-2001. Average total biomass increased during 1994 through 2001 and was slightly above $\mathrm{B}_{\text {MSY }}$ in 2001. The 2001 estimates of fishing mortality rate $(0.25)$ and total biomass $(9,805 \mathrm{mt})$ indicated that the stock was not overfished and overfishing was not occurring in 2001.

After the 2002 GARM, the biological reference points adopted at SAW34 were re-examined and use of the absolute estimates of $\mathrm{F}_{\text {MSY }}$ and $\mathrm{B}_{\text {MSY }}$, rather than survey-based equivalents, were recommended (NEFSC 2002b). In addition, medium term stochastic projections were generated using ASPIC software for 2002-2008 using bootstrap distributions of stock biomass in 2001 generated from the SAW 34 ASPIC model formulation and assuming $\mathrm{F}_{2002}=\mathrm{F}_{2001}$ and $\mathrm{F}_{2003-2008}=\mathrm{F}_{\text {MSY }}$. Projected biomass was maintained at $\mathrm{B}_{\text {MSY }}$ throughout the projection time series with high probability. Projected catch increased to $3,000 \mathrm{mt}$ and was also maintained throughout the projection time series.

### 2.0 Assessment Results

Stock status was assessed from the results of an updated run of the SAW 34 formulation of an ASPIC model, but version 5.10 (Prager 2004) of the ASPIC software was used rather than version 3.7.2 which was used in the 2002 assessment. The major software changes in ASPIC version 5.10 are estimation of the $K$ parameter instead of $r$, a change in starting biomass parameterization from $B_{1} / B_{\text {MSY }}$ to $B_{1} / K$, and model implementation using a continuous time step.

The new version of the ASPIC software was run using the input data input file from the 2002 GARM to determine the effect of the software changes on the assessment results. Similar results were obtained for the two model runs (Table K1), so the new version of the software was then run with the updated time series data that included the NEFSC survey biomass indices for autumn of 2002-2004 and spring of 2003-2005, as well as total landings for 2002-2004. Biascorrected parameter estimates and $80 \%$ confidence intervals were computed from 500 bootstrap trials. Results from the updated model run showed a decrease in the bias-corrected catchabilities $(q)$ of both the spring and autumn surveys. Bias-corrected estimates were lower for relative biomass ( $\mathrm{B}_{2004} / \mathrm{B}_{\mathrm{MSY}}$ ) and higher for relative fishing mortality ( $\mathrm{F}_{2004} / \mathrm{F}_{\mathrm{MSY}}$ ) in comparison to the 2001 estimates of these parameters.

### 2.1 The Fishery

Total commercial landings of Georges Bank winter flounder are predominately from U.S. fisheries, but also include landings from Canadian fisheries. Prior to 1978, USSR fleets also
landed winter flounder from Georges Bank. After 1993, Canadian landings increased and reached a peak of 500 mt in 2001, comprising $25 \%$ of the total landings. Thereafter, Canadian landings declined to 200 mt in 2004.

Total landings peaked at 4,500 mt in 1972 then declined between 1984 and 1995 from 3,900 mt to 800 mt , respectively (Table K2 and Figure K1). Total landings have been increasing since 1995, and during 1999 to 2004 , increased sharply from 1,000 to $3,100 \mathrm{mt}$, respectively.

Discarding of winter flounder occurs in the U.S. multi-species bottom trawl fishery and the scallop dredge fishery. However, data from the Observer Program (1989-2000) and Vessel Trip Report (1994-2000) databases were insufficient to produce reliable estimates of the magnitude or size and age composition of these discards (NEFSC 2002a).

### 2.2 Research Survey Indices

Relative biomass (stratified mean kg per tow) and abundance (stratified mean number per tow) indices from the NEFSC spring (April, 1968-2005) and autumn (October, 1963-2004) bottom trawl surveys (offshore strata 13-22), as well the Canadian spring bottom trawl surveys (February, 1987-2005), for strata 5Z1-Z4, are presented in Table K3. Biomass indices from all three surveys are presented in Figure K2. Canadian survey indices were not included in the current assessment because some winter flounder habitat on Georges Bank cannot be sampled by the survey gear and the inclusion of these indices in previous ASPIC model runs resulted in poor model fits (NEFSC 2002a). The spring survey indices were lagged back one year and used in the ASPIC model as an end-of-year index.

Despite considerable interannual variability, both series of NEFSC biomass indices indicate a declining trend during the 1980s and an increasing trend during the early 1990s through 2002. Thereafter, the spring biomass indices declined, and in 2004-2005, were at the low levels observed during the early 1990s. The autumn biomass indices also declined, but to a lesser extent. Biomass indices from the Canadian survey showed similar trends.

### 2.3 Biological Reference Points

ASPIC model estimates of relative total biomass ( $\mathrm{B}_{\mathrm{t}} / \mathrm{B}_{\mathrm{MSY}}$ ) and fishing mortality rates ( $\mathrm{F}_{t} / \mathrm{F}_{\mathrm{MSY}}$ ) are more precisely estimated than the absolute values (Prager 1995). Therefore, bias-corrected estimates of annual total biomass (as of Jan. 1) and fishing mortality rates are presented in relative terms. In order to determine stock status, these ratios are compared to a biomass threshold ( $50 \%$ of $\mathrm{B}_{\mathrm{MSY}}$ ) of 0.5 and a fishing mortality rate threshold ( $\mathrm{F}_{\mathrm{MSY}}$ ) of 1.0.

### 2.4 $\quad$ ASPIC Model Results and Stock Status

Relative fishing mortality rates increased rapidly after 1999 and were above $1.0\left(=\mathrm{F}_{\mathrm{MSY}}\right)$ during 2000-2004 (Figure K3). During 2004, the relative fishing mortality rate was 1.86 ( $80 \%$ CL $=$ $1.44,3.23$ ), nearly double the level of $\mathrm{F}_{\mathrm{MSY}}$ (Table K4). Relative total biomass (as of January 1) gradually increased from the lowest level on record, in 1994 (0.23), to just above the $\mathrm{B}_{\mathrm{MSY}}$ threshold in 2003 ( 0.56 ), but then declined during 2004 and 2005 (Figure K3). Relative biomass was 0.52 of $\mathrm{B}_{\mathrm{MSY}}$ in $2004(80 \% \mathrm{CL}=0.28,0.76)$ then fell below the biomass threshold to 0.46 of
$\mathrm{B}_{\mathrm{MSY}}$ in $2005(80 \% \mathrm{CL}=0.22,0.76)$. In 2004, relative biomass was just above the threshold limit and relative fishing mortality was well above $\mathrm{F}_{\text {MSY }}$. Therefore, in 2004, the stock was not overfished but overfishing was occurring.

A comparison of relative total biomass and fishing mortality rates from the updated assessment with those from the 2002 GARM indicated why a change in stock status occurred. The increasing biomass trend observed in the 2002 model results shifted to a declining trend after 2003 (Figure K4A) due to a decline in the survey biomass indices after 2002. After 2002, fishing mortality rates continued to increase, reaching levels well-above $\mathrm{F}_{\text {MSY }}$ (Figure K4B). These changes resulted in a divergence, after 1994, in biomass and fishing mortality rate estimates from the two assessments. Divergence was greatest during 2000-2002 when the $80 \%$ confidence intervals of the biomass estimates from the two assessments did not overlap (Figure K4A). Similarly, the confidence intervals for the 2001 and 2002 fishing mortality rate estimates did not overlap for the two times series. Bias-corrected estimates of absolute fishing mortality rates and January 1 total biomass are presented in Table K5.

Based on a comparison of the weighted mean square errors for each of the two survey time series, the fit for the updated model run was slightly poorer than the fit for the 2002 model run. However, in contrast to the 2002 model run, the updated run showed a retrospective pattern. Model runs for terminal years 1998-2004 suggested an underestimation of absolute fishing mortality rates and an overestimation of absolute average biomass during 2002-2004 (Figure K5).

Projections conducted for Amendment 13 included a constant fishing mortality scenario using the $\mathrm{F}_{\text {MSY }}$ estimate of 0.32 from SAW 34. Absolute total biomass estimates indicate a declining trend during 2002-2005, but are within the $80 \%$ confidence limits of the Amendment 13 projections of total biomass (Figure K6A). Unlike the projection scenario, fishing mortality rates increased during 2002-2004 (Figure K6B). During 2003 and 2004, realized landings were at the levels projected for Amendment 13 (Figure K6C).

### 3.0 Sources of Uncertainty

3.1 Exclusion of the discards from the U.S. otter trawl and scallop dredge fisheries results in an underestimation of fishery removals of the younger age classes (ages 0 to 3).
3.2 Current biomass levels estimated from the ASPIC model may not be reliable because recruitment is implicitly assumed to be a function of stock biomass.
3.3 U.S. landings after 1994 are based on prorations of preliminary logbook data. In 2004, a new method of reporting the landings was implemented whereby dealers rather than NMFS port agents entered the landings directly into the Weighout database.
3.4 There is some uncertainty about the accuracy of the Canadian landings because winter flounder are a bycatch species in the Canadian fisheries and a portion of the landings may be reported as unclassified flounders.

### 4.0 Research Recommendations

4.1 Include discards in future assessments.
4.2 Ensure that the survey indices and catch are composed of the same size fish.

### 5.0 Panel Discussion

The Panel discussed the consequence of the re-estimation of reference points in the updated assessment. The model estimates of $\mathrm{F}_{\text {MSY }}$ and $\mathrm{B}_{\text {MSY }}$ differ from those reported in SAW 34 and currently adopted. The terms of reference of the 2005 GARM do not include re-estimating reference points. Therefore, the Panel discussed whether it was consistent with the terms of reference to accept the ASPIC estimates of current stock status. However, because the estimation of K and MSY is intrinsic to the model fit, it is impossible to update the stock assessment without re-estimating the reference points. The Panel discussed several alternatives such as an alternative ASPIC run in which K and MSY were fixed at the values estimated from the 2002 GARM, comparing the new reference point estimates to the ASPIC model results from the 2002 GARM, and projecting from the 2002 GARM model using the actual catches. The model where MSY and K were fixed had a slightly poorer fit than the model which freely estimated the parameters. In addition, panelists noted that by imposing a constraint on $F_{\text {MSY }}$ and $B_{\text {MSY }}$ the model was not the same as that used in the last benchmark assessment. Accepting the results from constrained model would have conflicted with the terms of reference guidance that required the use of the same assessment method approved in the last benchmark assessment. The other two alternatives were considered scientifically invalid because the first method compares unrelated values while the second does not use all the data available. The Panel accepted the freely estimated model which is more pessimistic than the constrained model but has the same trend. The Panel decided to compare the ASPIC model results from the 2002 GARM with the updated model results by using the relative estimates of F and B instead of the absolute values (see comparison with bootstrap CI).

### 6.0 Literature Cited

NEFSC [Northeast Fisheries Science Center]. 2002a. Report of the $34^{\text {th }}$ Northeast Regional Stock Assessment Workshop ( $34^{\text {th }}$ SAW): Stock Assessment Review Committee (SARC) consensus summary of assessments. Northeast Fish. Sci. Cent. Ref. Doc. 02-06; 346 p.
NEFSC [Northeast Fisheries Science Center]. 2002b. Final report of the working group on reevaluation of biological reference points for New England groundfish. 231 p.
NEFSC [Northeast Fisheries Science Center]. 2002c. Assessment of 20 northeast groundfish stocks through 2001: A report of the Groundfish Assessment Review Meeting (GARM), Northeast Fisheries Science Center, Woods Hole, Massachusetts, October 8-11, 2002. Northeast Fish. Sci. Cent. Ref. Doc. 02-16. 511 p.
Prager, M.H. 2004. User's manual for ASPIC: a stock production model incorporating covariates (ver.5). Beaufort Lab. Doc. BL-2004-01. 27 p.
Prager, M.H. 1995. User's manual for ASPIC: a stock production model incorporating covariates, program version 3.6x. Miami Lab. Doc. MIA-92/93-55. 25 p.
Summary of results from ASPIC biomass dynamics models (SAW 34 configuration) utilized during the 2002 GARM and for an updated assessment of Georges Bank winter flounder. NA indicates these values are not estimated by the software version. All values are bias-correct using the methods defined by Prager (2004).

|  | GARM 2002, ASPIC v. 3.7.2 | Rerun GARM 2002, ASPIC v. 5.10 | Assessment update, ASPIC v. 5.10 |
| :--- | :---: | :---: | :---: |
| U.S. autumn survey, 1964-2001 | U.S. autumn survey, 1964-2001 <br> U.S. spring survey, 1968-2002 <br> Total landings, 1964-2001 | U.S. spring survey, 1968-2002 <br> Total landings, 1964-2001 | U.S. autumn survey, 1964-2004 <br> U.S. spring survey, 1968-2005 <br> Total landings, 1964-2004 |
| Total Objective Function | 1.959 | 1.959 | 2.448 |


| q ( $80 \%$ C.L. $),$ U.S. Autumn Survey | $0.265(0.183,0.330)$ | $0.259(0.184,0.352)$ | $0.225(0.156,0.331)$ |
| :--- | :---: | :---: | :---: |
| q ( $80 \%$ C.L. $),$ U.S. Spring Survey | $0.342(0.246,0.430)$ | $0.329(0.235,0.453)$ | $0.266(0.184,0.388)$ |
| r | 0.65 | 0.67 | 0.55 |
| K (mt) | 18,200 | 17,491 | 20,273 |


| $\mathrm{B}_{2001 \text { or } 2004} / \mathrm{B}_{\text {MSY (as of Jan. 1) }}$ | 1.06 | 1.04 | 0.52 |
| :--- | :---: | :---: | :---: |
| $\mathrm{~F}_{2001 \text { or } 2004} / \mathrm{F}_{\text {MSY }}$ | 0.76 | 0.76 | 1.86 |
| $\mathrm{~B}_{2005} / \mathrm{B}_{\text {MSY }}$ (as of Jan. 1) |  |  | 0.46 |
| $\mathrm{~B}_{\text {MSY }}{ }^{1}(\mathrm{mt})$ | 9,099 | 8,746 | 10,136 |
| $\mathrm{~F}_{\text {MSY }}$ | 0.33 | 0.31 | 0.22 |
| MSY (mt) | 3,008 | 2,950 | 2,785 |

${ }^{1}$ GARM 2002 point estimates of $\mathrm{F}_{\mathrm{MSY}}, \mathrm{B}_{\mathrm{MSY}}$ and MSY were not bias-corrected.

Table K2. Landings (mt) of Georges Bank winter flounder, by statistical area and country, during 1964-2004.

| YEAR | $\begin{gathered} 522-525 \\ 561-562 \\ \text { USA }^{1} \end{gathered}$ | $\begin{gathered} 5 \mathrm{Ze}^{2} \\ (521-526 \text { and } 541-562) \end{gathered}$ |  | $\begin{gathered} \hline 5 Z \\ (521-562) \end{gathered}$ |  | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | CANADA | USSR | CANADA | USSR |  |
| 1964 | 1,371 |  |  | 146 |  | 1,517 |
| 1965 | 1,176 |  |  | 199 | 312 | 1,687 |
| 1966 | 1,877 |  |  | 164 | 156 | 2,197 |
| 1967 | 1,917 |  |  | 83 | 349 | 2,349 |
| 1968 | 1,570 | 57 | 372 |  |  | 1,999 |
| 1969 | 2,167 | 116 | 235 |  |  | 2,518 |
| 1970 | 2,615 | 61 | 40 |  |  | 2,716 |
| 1971 | 3,092 | 62 | 1,029 |  |  | 4,183 |
| 1972 | 2,805 | 8 | 1,699 |  |  | 4,512 |
| 1973 | 2,269 | 14 | 693 |  |  | 2,976 |
| 1974 | 2,124 | 12 | 82 |  |  | 2,218 |
| 1975 | 2,409 | 13 | 515 |  |  | 2,937 |
| 1976 | 1,877 | 15 | 1 |  |  | 1,893 |
| 1977 | 3,572 | 15 | 7 |  |  | 3,594 |
| 1978 | 3,185 | 65 |  |  |  | 3,250 |
| 1979 | 3,045 | 19 |  |  |  | 3,064 |
| 1980 | 3,931 | 44 |  |  |  | 3,975 |
| 1981 | 3,993 | 19 |  |  |  | 4,012 |
| 1982 | 2,961 | 19 |  |  |  | 2,980 |
| 1983 | 3,894 | 14 |  |  |  | 3,908 |
| 1984 | 3,927 | 4 |  |  |  | 3,931 |
| 1985 | 2,151 | 12 |  |  |  | 2,163 |
| 1986 | 1,762 | 25 |  |  |  | 1,787 |
| 1987 | 2,637 | 32 |  |  |  | 2,669 |
| 1988 | 2,804 | 55 |  |  |  | 2,859 |
| 1989 | 1,880 | 11 |  |  |  | 1,891 |
| 1990 | 1,898 | 55 |  |  |  | 1,953 |
| 1991 | 1,814 | 14 |  |  |  | 1,828 |
| 1992 | 1,822 | 27 |  |  |  | 1,849 |
| 1993 | 1,662 | 21 |  |  |  | 1,683 |
| 1994 | 907 | 65 |  |  |  | 972 |
| 1995 | 706 | 54 |  |  |  | 760 |
| 1996 | 1,265 | 71 |  |  |  | 1,336 |


| YEAR | $\begin{gathered} \hline 522-525 \\ 561-562 \\ \text { USA }^{1} \end{gathered}$ | $\begin{gathered} 5 \mathrm{Ze}^{2} \\ (521-526 \text { and } 541-562) \end{gathered}$ |  | $\begin{gathered} \hline 5 \mathrm{Z} \\ (521-562) \end{gathered}$ |  | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | CANADA | USSR | CANADA | USSR |  |
| 1997 | 1,287 | 143 |  |  |  | 1,430 |
| 1998 | 1,243 | 93 |  |  |  | 1,336 |
| 1999 | 938 | 104 |  |  |  | 1,042 |
| 2000 | 1,677 | 161 |  |  |  | 1,838 |
| 2001 | 1,629 | 529 |  |  |  | 2,158 |
| 2002 | 2,110 | 244 |  |  |  | 2,354 |
| 2003 | 2,791 | 310 |  |  |  | 3,101 |
| 2004 | 2,931 | 191 |  |  |  | 3,122 |

${ }^{1}$ USA landings prior to 1985 include those from Statistical Areas 551 and 552 and landings during 1994-2004 were prorated from Vessel Trip Reports based on gear, month, and state.
${ }^{2}$ Includes landings from statistical areas 521 and 526, outside of the Georges Bank winter flounder stock area.

Table K3. Standardized, stratified relative abundance (mean number per tow) and biomass (mean kg per tow) indices for Georges Bank winter flounder caught in the U.S. spring and autumn and Canada spring research vessel bottom trawl surveys. U.S. offshore survey strata 13-22;
Canadian survey strata (5Z1-5Z4). Trawl door standardization coefficients of 1.46 (numbers) and 1.39 (weight) were applied to indices from U.S. survey indices prior to 1985 to account for differences in catchability between different survey doors.

| Year | U.S. Spring Survey |  | U.S. Autumn Survey |  | Canada Spring Survey |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number/tow | kg/tow | Number/tow | kg/tow | Number/tow | kg/tow |
| 1963 |  |  | 1.20 | 1.82 |  |  |
| 1964 |  |  | 1.30 | 1.82 |  |  |
| 1965 |  |  | 2.15 | 2.05 |  |  |
| 1966 |  |  | 5.16 | 5.66 |  |  |
| 1967 | Initiated | 1968 | 1.79 | 2.07 |  |  |
| 1968 | 2.70 | 3.11 | 1.31 | 1.07 |  |  |
| 1969 | 3.14 | 4.29 | 2.37 | 2.39 |  |  |
| 1970 | 1.86 | 2.29 | 5.62 | 6.49 |  |  |
| 1971 | 1.84 | 2.17 | 1.32 | 1.26 |  |  |
| 1972 | 4.95 | 5.32 | 1.26 | 1.58 |  |  |
| 1973 | 2.95 | 3.51 | 1.22 | 1.20 |  |  |
| 1974 | 6.05 | 5.78 | 1.19 | 1.46 |  |  |
| 1975 | 1.96 | 1.41 | 3.79 | 2.06 |  |  |
| 1976 | 4.67 | 3.01 | 5.99 | 3.93 |  |  |
| 1977 | 3.79 | 1.58 | 4.86 | 3.99 |  |  |
| 1978 | 7.07 | 5.06 | 4.06 | 3.10 |  |  |
| 1979 | 1.74 | 2.21 | 5.07 | 3.83 |  |  |
| 1980 | 3.22 | 2.80 | 1.66 | 1.87 |  |  |
| 1981 | 3.73 | 3.75 | 3.83 | 2.43 |  |  |
| 1982 | 2.30 | 1.52 | 5.30 | 2.69 |  |  |
| 1983 | 8.41 | 7.11 | 2.73 | 2.36 |  |  |
| 1984 | 5.53 | 5.60 | 3.93 | 2.45 |  |  |
| 1985 | 3.84 | 2.65 | 1.98 | 1.12 |  |  |
| 1986 | 2.00 | 1.21 | 3.58 | 2.18 | Iinitiated | 1987 |
| 1987 | 2.80 | 1.25 | 0.76 | 0.89 | 1.24 | 1.74 |
| 1988 | 2.93 | 1.65 | 4.08 | 1.27 | 4.31 | 2.75 |
| 1989 | 1.30 | 0.76 | 1.56 | 1.05 | 4.05 | 1.95 |
| 1990 | 2.80 | 1.57 | 0.50 | 0.35 | 4.93 | 2.64 |
| 1991 | 2.40 | 1.32 | 0.27 | 0.14 | 1.98 | 1.38 |
| 1992 | 1.42 | 0.90 | 0.68 | 0.38 | 0.51 | 0.59 |
| 1993 | 1.02 | 0.57 | 1.17 | 0.66 | 3.53 | 1.76 |
| 1994 | 1.29 | 0.58 | 0.87 | 0.58 | 5.10 | 2.01 |
| 1995 | 2.61 | 1.49 | 2.36 | 1.34 | 5.63 | 1.96 |
| 1996 | 2.31 | 1.50 | 1.54 | 1.76 | 4.12 | 2.30 |


| U.S. Spring Survey | U.S. Autumn Survey |  | Canada Spring Survey |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |
| Yumber/tow | $\mathrm{kg} /$ tow | Number/tow | $\mathrm{kg} /$ tow | Number/tow | $\mathrm{kg} / \mathrm{tow}$ |  |
| 1997 | 1.61 | 1.19 | 1.74 | 1.53 | 4.58 | 3.09 |
| 1998 | 0.76 | 0.72 | 1.78 | 1.57 | 1.14 | 1.21 |
| 1999 | 3.83 | 3.48 | 1.54 | 1.76 | 1.25 | 1.89 |
| 2000 | 4.42 | 3.69 | 2.16 | 2.66 | 1.48 | 2.22 |
| 2001 | 1.29 | 1.22 | 2.45 | 2.51 | 2.28 | 2.54 |
| 2002 | 5.05 | 5.16 | 2.58 | 3.15 | 3.17 | 3.85 |
| 2003 | 1.22 | 1.34 | 1.94 | 2.24 | 1.09 | 1.31 |
| 2004 | 0.39 | 0.51 | 2.50 | 2.04 | 2.10 | 1.79 |
| 2005 | 1.09 | 0.87 |  |  | 1.19 | 1.23 |
| Mean |  |  |  |  |  |  |
| All | 2.95 | 2.48 | 2.48 | 2.09 | 2.82 | 2.01 |
| Years |  |  |  |  |  |  |

Table K4. Bias-corrected estimates of relative fishing mortality rates ( $\mathrm{F}_{\mathrm{t}} / \mathrm{F}_{\mathrm{MSY}}$ ) and total biomass $\left(\mathrm{B}_{\mathrm{t}} / \mathrm{B}_{\mathrm{MSY}}\right.$, as of Jan. 1), derived using an ASPIC biomass dynamics model, for Georges Bank winter flounder during 1964-2005.

| Year | $\mathrm{F}_{\mathrm{t}} / \mathrm{F}_{\mathrm{MSY}}$ | $\begin{gathered} \mathrm{B}_{\mathrm{t}} / \mathrm{B}_{\mathrm{MSY}} \\ \text { (as of Jan. 1) } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: |
| 1964 | 0.85 | 0.48 |
| 1965 | 0.83 | 0.56 |
| 1966 | 0.98 | 0.64 |
| 1967 | 0.98 | 0.70 |
| 1968 | 0.77 | 0.75 |
| 1969 | 0.90 | 0.84 |
| 1970 | 0.94 | 0.89 |
| 1971 | 1.49 | 0.92 |
| 1972 | 1.79 | 0.85 |
| 1973 | 1.26 | 0.74 |
| 1974 | 0.91 | 0.74 |
| 1975 | 1.16 | 0.80 |
| 1976 | 0.71 | 0.80 |
| 1977 | 1.31 | 0.89 |
| 1978 | 1.23 | 0.85 |
| 1979 | 1.17 | 0.84 |
| 1980 | 1.59 | 0.84 |
| 1981 | 1.80 | 0.76 |
| 1982 | 1.44 | 0.68 |
| 1983 | 2.08 | 0.66 |
| 1984 | 2.56 | 0.56 |
| 1985 | 1.61 | 0.44 |
| 1986 | 1.30 | 0.44 |
| 1987 | 1.99 | 0.46 |
| 1988 | 2.47 | 0.42 |
| 1989 | 1.84 | 0.34 |
| 1990 | 2.01 | 0.33 |
| 1991 | 2.02 | 0.30 |
| 1992 | 2.24 | 0.28 |
| 1993 | 2.25 | 0.25 |
| 1994 | 1.28 | 0.23 |
| 1995 | 0.86 | 0.25 |
| 1996 | 1.32 | 0.31 |
| 1997 | 1.29 | 0.34 |
| 1998 | 1.09 | 0.37 |
| 1999 | 0.73 | 0.41 |
| 2000 | 1.13 | 0.49 |
| 2001 | 1.25 | 0.53 |
| 2002 | 1.31 | 0.55 |
| 2003 | 1.76 | 0.56 |
| 2004 | 1.86 | 0.52 |
| 2005 |  | 0.46 |

Table K5. Bias-corrected estimates of absolute fishing mortality rates and January 1 total biomass ( 000 's mt), derived using an ASPIC biomass dynamics model, for Georges Bank winter flounder during 1964-2005.

| Year | Fishing Mortality Rate | Jan. 1 Total Biomass (000's mt) |
| :---: | :---: | :---: |
| 1964 | 0.19 | 5.886 |
| 1965 | 0.18 | 6.994 |
| 1966 | 0.22 | 8.147 |
| 1967 | 0.22 | 8.941 |
| 1968 | 0.17 | 9.659 |
| 1969 | 0.21 | 10.816 |
| 1970 | 0.21 | 11.502 |
| 1971 | 0.34 | 11.992 |
| 1972 | 0.41 | 10.972 |
| 1973 | 0.29 | 9.496 |
| 1974 | 0.21 | 9.485 |
| 1975 | 0.26 | 10.267 |
| 1976 | 0.16 | 10.369 |
| 1977 | 0.30 | 11.570 |
| 1978 | 0.28 | 11.075 |
| 1979 | 0.27 | 10.897 |
| 1980 | 0.37 | 10.881 |
| 1981 | 0.41 | 9.885 |
| 1982 | 0.33 | 8.731 |
| 1983 | 0.47 | 8.502 |
| 1984 | 0.58 | 7.198 |
| 1985 | 0.36 | 5.562 |
| 1986 | 0.29 | 5.500 |
| 1987 | 0.45 | 5.846 |
| 1988 | 0.56 | 5.270 |
| 1989 | 0.41 | 4.286 |
| 1990 | 0.45 | 4.107 |
| 1991 | 0.45 | 3.789 |
| 1992 | 0.49 | 3.498 |
| 1993 | 0.49 | 3.075 |
| 1994 | 0.27 | 2.694 |
| 1995 | 0.19 | 3.033 |
| 1996 | 0.29 | 3.772 |
| 1997 | 0.28 | 4.119 |
| 1998 | 0.24 | 4.490 |
| 1999 | 0.16 | 5.105 |
| 2000 | 0.26 | 6.244 |
| 2001 | 0.28 | 6.791 |
| 2002 | 0.30 | 7.114 |
| 2003 | 0.40 | 7.297 |
| 2004 | 0.42 | 6.692 |
| 2005 |  | 6.385 |



Figure K1. Total commercial landings of Georges Bank winter flounder during 1964-2004.


Figure K2. Relative biomass indices (stratified mean kg per tow) of Georges Bank winter flounder from NEFSC spring (1968-2005, lagged back one year) and autumn (1963-2004) bottom trawl surveys and the Canadian spring (1987-2005) bottom trawl survey.


Figure K3. Trends in bias-corrected estimates of relative total biomass ( $\mathrm{B}_{\mathrm{t}} / \mathrm{B}_{\mathrm{MSY}}$ on Jan. 1) and relative fishing mortality rates ( $\mathrm{F}_{\mathrm{t}} / \mathrm{F}_{\mathrm{MSY}}$ ), derived using an ASPIC biomass dynamics model, for Georges Bank winter flounder. Error bars represent bias-corrected $80 \%$ confidence intervals.


Figure K4. Bias-corrected estimates of (A) relative total biomass ( $\mathrm{B}_{\mathrm{t}} / \mathrm{B}_{\mathrm{MSY}}$ on Jan. 1), during 19642005, and (B) relative fishing mortality rates ( $\mathrm{F}_{\mathrm{t}} / \mathrm{F}_{\mathrm{MSY}}$ ), during 1964-2004, for the 2002 and 2005 ASPIC model runs for Georges Bank winter flounder. Error bars represent bias-corrected 80\% confidence intervals.


Figure K5. Retrospective patterns in absolute estimates of (A) fishing mortality rates and (B) average biomass, during terminal years 1995-2000, for an updated ASPIC biomass dynamics model for Georges Bank winter flounder, 1964-2004. Estimates of fishing mortality and stock biomass are not bias-corrected.


Figure K6. ASPIC-derived absolute estimates of (A) total biomass (as of Jan. 1) and (B) fishing mortality rates, and (C) actual landings, for Georges Bank winter flounder during 2002-2004, versus projected estimates, for Amendment 13, based on a constant F ( $=0.32$ ) scenario. ASPIC-derived point estimates and $80 \%$ confidence limits are bias-corrected.

