## M. Gulf of Maine-Georges Bank Acadian Redfish by R.K. Mayo and L. Col

### 1.0 Background

The most recent stock assessment of Acadian redfish in Subarea 5 was completed in 2001 (Mayo et al. 2002), and the results were reviewed at the $33^{\text {rd }}$ Northeast Regional Stock Assessment Workshop in June, 2001 (NEFSC 2001a, 2001b). The assessment was based on several analyses including trends in catch/survey biomass exploitation ratios; a yield and biomass per recruit analysis; an age-structured dynamics model which incorporates information on the age composition of the landings, size and age composition of the population, and trends in relative abundance derived from commercial CPUE and research vessel survey biomass indices; and an age-aggregated biomass dynamics model. Surplus production estimates were derived from the age-structured dynamics model, and information on current biomass and fishing mortality relative to MSY-based reference points were also provided by the biomass dynamics model.

At that time, the NEFSC autumn survey biomass index had increased substantially during the mid-1990s and had remained relatively high through 2000. The rapid increase in abundance and biomass was attributed to recruitment and growth of the 1992 and other early-1990s year classes. The assessment conducted in 2001 provided no basis with which to evaluate the state of the stock relative to the control rule as determined by the Overfishing Definition Review Panel (Anon. 1998).

### 2.0 The Fishery

During the early development phase of the Gulf of Maine redfish fishery, USA landings increased rapidly to a peak level of about $56,000 \mathrm{mt}$ in 1942 followed by a steep decline through the early 1950s (Table M1; Figure M1). Nominal catches then declined at a more gradual rate to less than $10,000 \mathrm{mt}$ during the 1960s. During the 1970s, USA landings increased again, peaking at $16,000 \mathrm{mt}$ in 1971 and again at $15,000 \mathrm{mt}$ in 1979. During the 1970 s , additional catches by Canadian and distant water fleets increased the total redfish catch to a maximum of about 17,000 to $20,000 \mathrm{mt}$ per year from 1970 through 1973; catches of redfish by these fleets declined to negligible levels after 1976. Landings of redfish declined steadily throughout the 1980s, remaining below $1,000 \mathrm{mt}$ per year since 1989, and at less than 500 mt per year since 1994. Total redfish landings in 2001 were 360 mt compared to 319 mt in 2000.

### 3.0 Research Survey Indices

Indices of relative biomass, derived from NEFSC autumn research vessel bottom trawl surveys, although variable, exhibited a steady decline between 1963 and 1982 (Table M2, Figure M2). On average, the biomass index appears to have declined by about $90 \%$ over a 20 year period. During this time, only 2 year classes of any significance were produced, 1971 and 1978. Between 1983 and 1993, the biomass index approximately doubled, reflecting the relatively low rate of removals by the fishery and the very slow growth rate of the species. No substantial year classes were detected by research vessel surveys in the inshore survey strata traditionally used to
monitor recruitment until autumn 1995 when a substantial number of fish in the $15-19 \mathrm{~cm}$ range were noted, suggesting the possibility of above average reproduction in 1990 and/or 1991 . This was followed by a very large increase in the index in the offshore strata in the autumn of 1996. The autumn biomass index has fluctuated between 20 and 30 kg per tow since then, a magnitude comparable to the period between 1963 and the mid-1970s.

During the earlier periods, however, redfish were generally first detected in the inshore strata at relatively small sizes ( $\sim 10 \mathrm{~cm}$ or less, age 1 or 2 ), only to appear in the offshore strata after about 5 or six years (Mayo 1993). During the 1990s recruitment event, the year class was not detected until fish were close to 20 cm , or about ages 4 or 5 , and the numbers appeared to be present in both inshore and offshore strata. The autumn biomass index increased 4-5 fold between the early 1990s and the mid-1990s, a rate that is inconsistent with the dynamics of this species. The spring index, however, suggests only a very modest change in biomass since the mid-1990s.

### 4.0 Assessment Results

Since the assessment reviewed at SAW 33 was completed, no additional aging data have become available to allow an assessment update. Landings remained very low in 2001 and the 2001 NEFSC autumn survey biomass index remained similar to that of 2000, indicating no appreciable change in the exploitation rate since 2000. Therefore, the results from the 2001 assessment serve as the basis for the present assessment report.

Exploitation ratios (catch/survey biomass) suggest that fishing mortality has been very low since the mid-1980s compared to previous periods (Table M3; Figure M3). Estimates of fishing mortality derived from the age-structured dynamics model and the age-aggregated biomass model were similar (Mayo et al. 2002), both indicating that current fishing mortality is low relative to past decades and less than $5 \%$ of $\mathrm{F}_{\text {MSY }}$. Spawning stock biomass has increased since the mid-1990s, and was estimated to be 119,600 mt in 2000 (Mayo et al. 2002) due, in large part, to strong recruitment from the early 1990s. When measured against the estimates of $\mathrm{F}_{\mathrm{MSY}}$ and $\mathrm{SSB}_{\text {MSY }}$ provided in NEFSC (2002), the stock is not overfished, and overfishing is not occurring.

Given the continued extremely low landings of redfish relative to the recent increase in biomass, exploitation is now extremely low compared to the 1960s and 1970s (Table M3; Figure M3). However, in contrast to this earlier period, where a substantial proportion of the stock persisted in the $30-40 \mathrm{~cm}$ range (Mayo, 1993), during the 1990s, almost all of the redfish were less than 25 cm , and almost none are greater than 30 cm . This suggests that, given the present demographics of the stock, only a small fraction of the biomass would be considered exploitable.

### 5.0 Biological Reference Points

Estimates of recruitment obtained from the age-structured biomass dynamics model reviewed at the $33^{\text {rd }}$ SAW were used to imply the probable recruitment that could be produced by a rebuilt stock as described in NEFSC (2002). Recruitment estimates derived by the model from the

1952-1999 yearclasses served as the basis for evaluating trends and patterns in recruitment. The stock-recruitment data suggest an increase in the frequency of larger year classes ( $>50$ million fish) at higher biomass levels. Therefore, recruitment estimates corresponding to the upper quartile of the SSB range served as the basis for deriving mean and median recruitment estimates. In accordance with the recommendation of the Stock Assessment Review Committee of the $33^{\text {rd }}$ SAW, the estimate of $\mathrm{F}_{50 \%}(0.04)$ is taken as a proxy for $\mathrm{F}_{\mathrm{MSY}}$. This fishing mortality rate produces 4.1073 kg of spawning stock biomass per recruit and 0.1429 kg of yield per recruit. The resulting mean recruitment of 57.63 million fish results in an $\mathrm{SSB}_{\text {MSY }}$ estimate of 236, 700 mt when multiplied by the SSB per recruit, and an MSY estimate of $8,235 \mathrm{mt}$ when multiplied by the yield per recruit.

Reference points derived from the non parametric approach are:

| MSY | $8,235 \mathrm{mt}$ |
| :--- | :--- |
| $\mathrm{B}_{\text {MSY }}$ | $236,700 \mathrm{mt}$ |
| $\mathrm{F}_{\text {MSY }}$ | $0.04=\mathrm{F}_{50 \%}$ MSP |

It was determined (NEFSC 2002) that the stock could not be rebuilt to $\mathrm{B}_{\text {MSY }}$ by 2009 even at $\mathrm{F}=0.0$. Therefore, the rebuilding scenario invoked a 20 year plus 1 mean generation time ( 31 years for Acadian redfish) to achieve rebuilding. This results in an $\mathrm{F}_{\text {rebuild }}=0.01$.

### 6.0 GARM Panel Comments

A question was raised as to why the catches have not followed the increase in the survey biomass. The current mesh size is too large for the size of the fish which make up the bulk of the biomass. The fishery for redfish from the 1950s to the 1980s used a smaller mesh size for redfish trips (3"). Some fishers claim to be discarding but there do not appear to be any large discarding events in the data. There is no evidence of targeting at present. The market was lost when the stock declined.

The change in mesh size used in the fishery was a concern in the interpretation of exploitation ratios. Ratios of catch to total biomass indices may not be comparable under different mesh regimes because the change in the amount of exploitable biomass would produce different q's. This is probably not a direct concern because exploitation ratios are not the basis for the assessment and the overall conclusion would not change. For species in which larger fish make up the major portion of the catch, this may not be a problem, but it may be for smaller-sized species such as redfish.

There was a question as to whether the year classes from the 1990s may have been inshore of the survey at younger ages. This had not been the case in the past for other large year classes. The Massachusetts survey does occasionally catch small redfish.

## Recommendations

- Compute survey biomass indices of exploitable biomass and utilize these for calculating exploitation ratios.
- Perform a more systematic analysis of the data to determine discard rates.


### 7.0 Sources of Uncertainty

- The sharp increase in the survey biomass index in 1996 is inconsistent with the life history characteristics of this species.
- Given the pelagic diurnal movement and general distribution of redfish, swept area estimates of stock biomass derived from bottom trawl survey data will tend to underestimate absolute stock size.


### 8.0 References

Anon. 1998. Evaluation of existing overfishing definitions and recommendations for new overfishing definitions to comply with the Sustainable Fisheries Act. Final Report. Overfishing Definition Review Panel. June 17, 1998.

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NEFSC 2002. Working Group on Re-Evaluation of Biological Reference Points for New England Groundfish, . NMFS/NEFSC, Reference Document 02-04, 254p.

Table M1 Nominal redfish catches (metric tons), actual and standardized catch per unit effort, and calculated standardized USA and total effort for the Gulf of Maine-Georges Bank redfish fishery.

|  | Nominal | Catch | etric tons) | USA Cat Effort | ch per Unit (tons/day) | Calcula <br> Effort | Standard ys fished) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | USA | Others | Total | Actual | Standard | USA | Total |
| 1934 | 519 |  | 519 |  |  |  |  |
| 1935 | 7549 |  | 7549 |  |  |  |  |
| 1936 | 23162 |  | 23162 |  |  |  |  |
| 1937 | 14823 |  | 14823 |  |  |  |  |
| 1938 | 20640 |  | 20640 |  |  |  |  |
| 1939 | 25406 |  | 25406 |  |  |  |  |
| 1940 | 26762 |  | 26762 |  |  |  |  |
| 1941 | 50796 |  | 50796 |  |  |  |  |
| 1942 | 55892 |  | 55892 | 6.9 | 6.9 | 8100 | 8100 |
| 1943 | 48348 |  | 48348 | 6.7 | 6.7 | 7216 | 7216 |
| 1944 | 50439 |  | 50439 | 5.4 | 5.4 | 9341 | 9341 |
| 1945 | 37912 |  | 37912 | 4.5 | 4.5 | 8425 | 8425 |
| 1946 | 42423 |  | 42423 | 4.7 | 4.7 | 9026 | 9026 |
| 1947 | 40160 |  | 40160 | 4.9 | 4.9 | 8196 | 8196 |
| 1948 | 43631 |  | 43631 | 5.4 | 5.4 | 8080 | 8080 |
| 1949 | 30743 |  | 30743 | 3.3 | 3.3 | 9316 | 9316 |
| 1950 | 34307 |  | 34307 | 4.1 | 4.1 | 8368 | 8368 |
| 1951 | 30077 |  | 30077 | 4.1 | 4.1 | 7336 | 7336 |
| 1952 | 21377 |  | 21377 | 3.5 | 3.4 | 6287 | 6287 |
| 1953 | 16791 |  | 16791 | 3.8 | 3.6 | 4664 | 4664 |
| 1954 | 12988 |  | 12988 | 3.4 | 3.1 | 4190 | 4190 |
| 1955 | 13914 |  | 13914 | 4.5 | 4.0 | 3479 | 3479 |
| 1956 | 14388 |  | 14388 | 4.4 | 3.8 | 3786 | 3786 |
| 1957 | 18490 |  | 18490 | 4.3 | 3.6 | 5136 | 5136 |
| 1958 | 16043 | 4 | 16047 | 4.4 | 3.6 | 4456 | 4458 |
| 1959 | 15521 |  | 15521 | 4.3 | 3.5 | 4435 | 4435 |
| 1960 | 11373 | 2 | 11375 | 3.8 | 3.0 | 3791 | 3792 |
| 1961 | 14040 | 61 | 14101 | 4.6 | 3.5 | 4011 | 4029 |
| 1962 | 12541 | 1593 | 14134 | 5.4 | 4.0 | 3135 | 3534 |
| 1963 | 8871 | 1175 | 10046 | 4.1 | 3.0 | 2957 | 3349 |
| 1964 | 7812 | 501 | 8313 | 4.3 | 2.9 | 2694 | 2867 |
| 1965 | 6986 | 1071 | 8057 | 7.0 | 4.4 | 1588 | 1831 |
| 1966 | 7204 | 1365 | 8569 | 11.7 | 6.4 | 1126 | 1339 |
| 1967 | 10442 | 422 | 10864 | 12.4 | 5.6 | 1865 | 1940 |
| 1968 | 6578 | 199 | 6777 | 14.7 | 6.1 | 1078 | 1111 |
| 1969 | 12041 | 414 | 12455 | 11.4 | 4.9 | 2457 | 2542 |
| 1970 | 15534 | 1207 | 16741 | 9.0 | 4.0 | 3884 | 4185 |
| 1971 | 16267 | 3767 | 20034 | 7.0 | 3.2 | 5083 | 6261 |
| 1972 | 13157 | 5938 | 19095 | 5.7 | 2.9 | 4537 | 6584 |
| 1973 | 11954 | 5406 | 17360 | 5.3 | 2.9 | 4122 | 5986 |
| 1974 | 8677 | 1794 | 10471 | 5.0 | 2.6 | 3337 | 4027 |
| 1975 | 9075 | 1497 | 10572 | 4.0 | 2.2 | 4125 | 4805 |
| 1976 | 10131 | 565 | 10696 | 4.6 | 2.3 | 4405 | 4650 |
| 1977 | 13012 | 211 | 13223 | 4.9 | 2.5 | 5205 | 5289 |
| 1978 | 13991 | 92 | 14083 | 4.8 | 2.4 | 5830 | 5868 |
| 1979 | 14722 | 33 | 14755 | 3.6 | 1.9 | 7748 | 7766 |
| 1980 | 10085 | 98 | 10183 | 3.2 | 1.6 | 6303 | 6364 |
| 1981 | 7896 | 19 | 7915 | 2.7 | 1.4 | 5640 | 5654 |
| 1982 | 6735 | 168 | 6903 | 2.7 | 1.5 | 4490 | 4602 |
| 1983 | 5215 | 113 | 5328 | 2.1 | 1.2 | 4346 | 4440 |
| 1984 | 4722 | 71 | 4793 | 1.9 | 1.1 | 4293 | 4357 |
| 1985 | 4164 | 118 | 4282 | 1.4 | 0.9 | 4627 | 4758 |
| 1986 | 2790 | 139 | 2929 | 1.0 | 0.6 | 4650 | 4882 |
| 1987 | 1859 | 35 | 1894 | 1.1 | 0.7 | 2656 | 2706 |
| 1988 | 1076 | 101 | 1177 | 0.9 | 0.5 | 2152 | 2354 |
| 1989 | 628 | 9 | 637 | 1.1 | 0.6 | 1047 | 1062 |
| 1990 | 588 | 13 | 601 |  |  |  |  |
| 1991 | 525 |  | 525 |  |  |  |  |
| 1992 | 849 |  | 849 |  |  |  |  |
| 19934* | 800 |  | 800 |  |  |  |  |
| 1994** | 440 |  | 440 |  |  |  |  |
| 1998* | 320 |  | 320 |  |  |  |  |
| 1999* | 353 |  | 353 |  |  |  |  |
| 2000* | 319 |  | 319 |  |  |  |  |
| 2001* | 360 |  | 360 |  |  |  |  |

* Preliminary

CPUE and effort not calculated after 1989 due to sharp reduction in directed redfish trips

|  | INSHORE 1 |  |  |  | OFFSHORE 2 |  |  |  | COMBINED 3 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | $\begin{aligned} & \text { Stratif } \\ & \text { Catch pe } \\ & \text { Number } \\ & \hline \end{aligned}$ | ed Mean Tow kg | Avg. Wgt. (kg) | Avg. Length (cm) | $\begin{array}{r} \text { Stratifi } \\ \text { Catch p } \\ \text { Number } \\ \hline \end{array}$ | $\begin{aligned} & \text { d Mean } \\ & \text { Tow } \\ & \text { (kg) } \\ & \hline \end{aligned}$ | Avg. Wgt. (kg) | Avg. Length (cm) | Stratifi Catch Number | $\begin{aligned} & \text { Mean } \\ & \text { Tow } \\ & \text { kg } \\ & \hline \end{aligned}$ |
| 1963 | 86.3 | 7.6 | 0.088 | 17.4 | 87.5 | 27.0 | 0.309 | 26.4 | 87.3 | 24.1 |
| 1964 | 81.3 | 13.5 | 0.166 | 20.2 | 122.3 | 61.8 | 0.505 | 30.8 | 116.3 | 54.6 |
| 1965 | 189.5 | 22.3 | 0.118 | 17.7 | 33.9 | 11.5 | 0.339 | 25.3 | 57.0 | 13.1 |
| 1966 | 172.8 | 17.0 | 0.098 | 16.2 | 77.8 | 31.2 | 0.401 | 27.4 | 91.9 | 29.1 |
| 1967 | 62.9 | 5.3 | 0.084 | 17.7 | 107.1 | 27.6 | 0.258 | 23.6 | 100.5 | 24.3 |
| 1968 | 41.1 | 4.7 | 0.114 | 18.3 | 161.3 | 46.6 | 0.289 | 25.1 | 143.4 | 40.4 |
| 1969 | 105.9 | 16.0 | 0.151 | 20.7 | 65.2 | 24.8 | 0.380 | 27.4 | 71.2 | 23.5 |
| 1970 | 18.2 | 2.8 | 0.154 | 20.3 | 107.2 | 38.2 | 0.356 | 26.3 | 94.0 | 32.9 |
| 1971 | 20.7 | 4.7 | 0.227 | 21.8 | 52.8 | 26.7 | 0.506 | 29.7 | 48.0 | 23.4 |
| 1972 | 36.4 | 6.6 | 0.181 | 20.8 | 58.9 | 27.8 | 0.472 | 29.2 | 55.6 | 24.6 |
| 1973 | 26.2 | 2.1 | 0.080 | 15.6 | 41.4 | 19.7 | 0.476 | 29.7 | 39.2 | 17.0 |
| 1974 | 44.4 | 4.7 | 0.106 | 18.0 | 49.0 | 27.6 | 0.563 | 30.1 | 48.3 | 24.2 |
| 1975 | 45.7 | 6.0 | 0.131 | 19.6 | 79.9 | 45.9 | 0.574 | 30.6 | 74.8 | 39.9 |
| 1976 | 11.6 | 2.5 | 0.216 | 22.6 | 31.9 | 17.5 | 0.549 | 30.2 | 28.9 | 15.3 |
| 1977 | 54.6 | 12.3 | 0.225 | 23.4 | 37.9 | 18.1 | 0.478 | 28.5 | 40.4 | 17.3 |
| 1978 | 20.4 | 5.5 | 0.270 | 24.6 | 49.5 | 23.4 | 0.473 | 29.0 | 45.2 | 20.7 |
| 1979 | 6.2 | 2.1 | 0.339 | 26.5 | 32.8 | 18.4 | 0.561 | 30.5 | 28.9 | 16.0 |
| 1980 | 20.6 | 6.2 | 0.301 | 24.6 | 20.6 | 13.8 | 0.670 | 31.8 | 20.6 | 12.6 |
| 1981 | 6.8 | 1.9 | 0.279 | 24.9 | 22.7 | 14.0 | 0.617 | 31.8 | 20.4 | 12.2 |
| 1982 | 28.2 | 4.6 | 0.163 | 21.2 | 5.6 | 3.2 | 0.571 | 31.5 | 9.0 | 3.4 |
| 1983 | 30.2 | 8.7 | 0.288 | 24.8 | 6.5 | 3.3 | 0.508 | 29.1 | 10.0 | 4.1 |
| 1984 | 7.7 | 3.2 | 0.416 | 27.9 | 7.8 | 4.1 | 0.526 | 29.0 | 7.8 | 3.9 |
| 1985 | 7.2 | 2.1 | 0.292 | 24.8 | 14.0 | 6.3 | 0.450 | 28.0 | 13.0 | 5.7 |
| 1986 | 67.6 | 15.3 | 0.226 | 23.3 | 18.8 | 6.7 | 0.356 | 26.1 | 26.1 | 8.0 |
| 1987 | 26.5 | 4.8 | 0.181 | 21.9 | 11.5 | 5.6 | 0.487 | 29.2 | 13.7 | 5.5 |
| 1988 | 18.5 | 5.1 | 0.276 | 21.9 | 11.4 | 6.5 | 0.570 | 29.1 | 12.4 | 6.3 |
| 1989 | 14.0 | 2.9 | 0.207 | 22.6 | 21.3 | 7.5 | 0.352 | 25.9 | 20.3 | 6.8 |
| 1990 | 57.6 | 14.5 | 0.252 | 23.8 | 31.7 | 11.7 | 0.369 | 26.7 | 35.5 | 12.2 |
| 1991 | 7.2 | 1.1 | 0.153 | 20.4 | 21.1 | 9.6 | 0.455 | 28.5 | 19.1 | 8.4 |
| 1992 | 7.8 | 1.2 | 0.147 | 20.0 | 24.9 | 9.3 | 0.374 | 27.3 | 22.4 | 8.1 |
| 1993 | 53.7 | 7.4 | 0.137 | 20.0 | 32.5 | 11.9 | 0.366 | 26.3 | 35.6 | 11.2 |
| 1994 | 31.5 | 5.4 | 0.171 | 21.7 | 19.0 | 6.0 | 0.317 | 25.0 | 20.9 | 5.9 |
| 1995 | 109.7 | 11.1 | 0.102 | 18.5 | 19.9 | 3.5 | 0.177 | 21.3 | 33.2 | 4.7 |
| 1996 | 53.8 | 9.1 | 0.169 | 21.5 | 189.9 | 34.4 | 0.181 | 21.9 | 169.6 | 30.6 |
| 1997 | 105.6 | 15.7 | 0.149 | 20.3 | 57.9 | 19.5 | 0.337 | 26.0 | 65.0 | 18.9 |
| 1998 | 48.7 | 10.7 | 0.219 | 20.4 | 128.9 | 35.4 | 0.275 | 23.6 | 117.0 | 31.7 |
| 1999 | 164.2 | 35.1 | 0.214 | 23.2 | 68.2 | 20.7 | 0.304 | 25.6 | 82.5 | 22.9 |
| 2000 | 133.3 | 21.8 | 0.164 | 21.6 | 99.4 | 26.9 | 0.271 | 24.8 | 104.4 | 26.2 |
| 2001 | 144.4 | 28.9 | 0.200 | 22.8 | 80.2 | 28.0 | 0.349 | 27.3 | 89.8 | 28.2 |

1. Strata Set: 26, 27, 39, 40
2. Strata Set: 24, 28-30, 36-38

Table M3. Commercial landings (mt), NEFSC autumn survey biomass index (kg/tow), and index of exploitation for Gulf of Maine redfish.

| Year | $\begin{gathered} \text { Commercia1 } \\ \text { 1andings } \\ (\mathrm{mt}) \end{gathered}$ | Biomass Index | Exploitation Ratio |
| :---: | :---: | :---: | :---: |
| 1963 | 10046 | 24.1 | 0.4168 |
| 1964 | 8313 | 54.6 | 0.1523 |
| 1965 | 8057 | 13.1 | 0.6150 |
| 1966 | 8569 | 29.1 | 0.2945 |
| 1967 | 10864 | 24.3 | 0.4471 |
| 1968 | 6777 | 40.4 | 0.1677 |
| 1969 | 12455 | 23.5 | 0.5300 |
| 1970 | 16741 | 32.9 | 0.5088 |
| 1971 | 20034 | 23.4 | 0.8562 |
| 1972 | 19095 | 24.6 | 0.7762 |
| 1973 | 17360 | 17.0 | 1.0212 |
| 1974 | 10471 | 24.2 | 0.4327 |
| 1975 | 10572 | 39.9 | 0.2650 |
| 1976 | 10696 | 15.3 | 0.6991 |
| 1977 | 13223 | 17.3 | 0.7643 |
| 1978 | 14083 | 20.7 | 0.6803 |
| 1979 | 14755 | 16.0 | 0.9222 |
| 1980 | 10183 | 12.6 | 0.8082 |
| 1981 | 7915 | 12.2 | 0.6488 |
| 1982 | 6903 | 3.4 | 2.0303 |
| 1983 | 5328 | 4.1 | 1.2995 |
| 1984 | 4793 | 3.9 | 1.2290 |
| 1985 | 4282 | 5.7 | 0.7512 |
| 1986 | 2929 | 8.0 | 0.3661 |
| 1987 | 1894 | 5.5 | 0.3444 |
| 1988 | 1177 | 6.3 | 0.1868 |
| 1989 | 637 | 6.8 | 0.0937 |
| 1990 | 601 | 12.2 | 0.0493 |
| 1991 | 525 | 8.4 | 0.0625 |
| 1992 | 849 | 8.1 | 0.1049 |
| 1993 | 800 | 11.2 | 0.0714 |
| 1994 | 440 | 5.9 | 0.0741 |
| 1995 | 440 | 4.7 | 0.0946 |
| 1996 | 322 | 30.6 | 0.0105 |
| 1997 | 251 | 18.9 | 0.0133 |
| 1998 | 320 | 31.7 | 0.0101 |
| 1999 | 353 | 22.9 | 0.0154 |
| 2000 | 319 | 26.2 | 0.0122 |
| $\underline{2001}$ | 360 | 28.2 | 0.0128 |

## Gulf of Maine-Georges Bank Redfish

Commercial Landings


Figure M1. Total commercial landings of Acadian redfish from the Gulf of Maine-Georges Bank region, 1934-2001

## Subarea 5 Acadian Redfish

Landings and Biomass Index


Figure M2. Commercial landings and biomass index derived from NEFSC autumn survey biomass indices for Acadian redfish, 1963-2001.

## Subarea 5 Acadian Redfish

Landings and Exploitation Ratio


Figure M3. Commercial landings and exploitation ratios derived from NEFSC autumn survey biomass indices for Acadian redfish.

