

# Methylmercury Production in the Coastal Zone: An Important Source of Methylmercury to Marine Fish?

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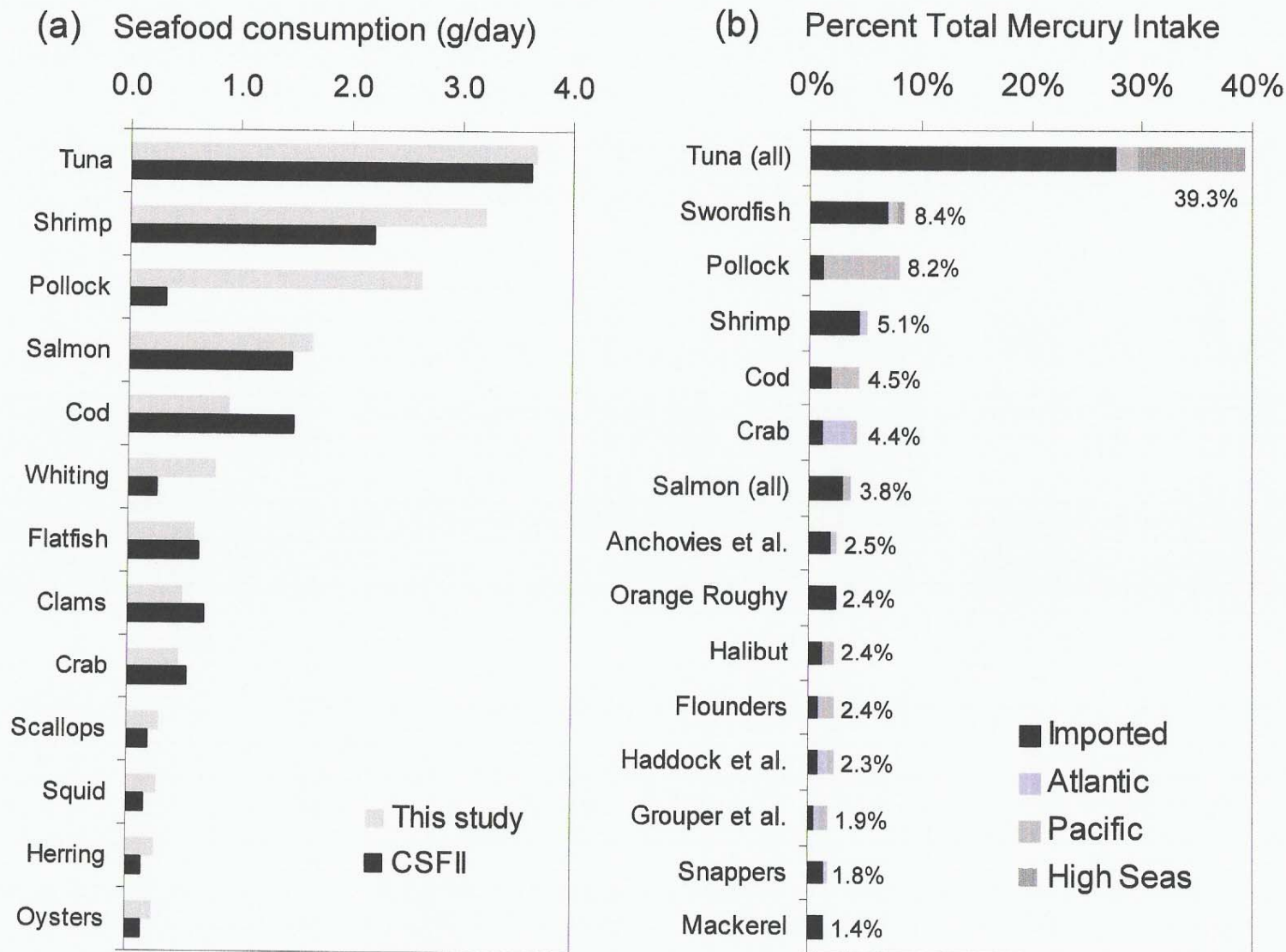
# Outline

- Brief background on Mercury in the Environment
- Methylmercury in Ocean Fish, and its Sources
- Mercury Methylation in Coastal Waters
- Importance of Extreme Events in Mercury Cycling and Fate
- Is the Coastal Zone Important?

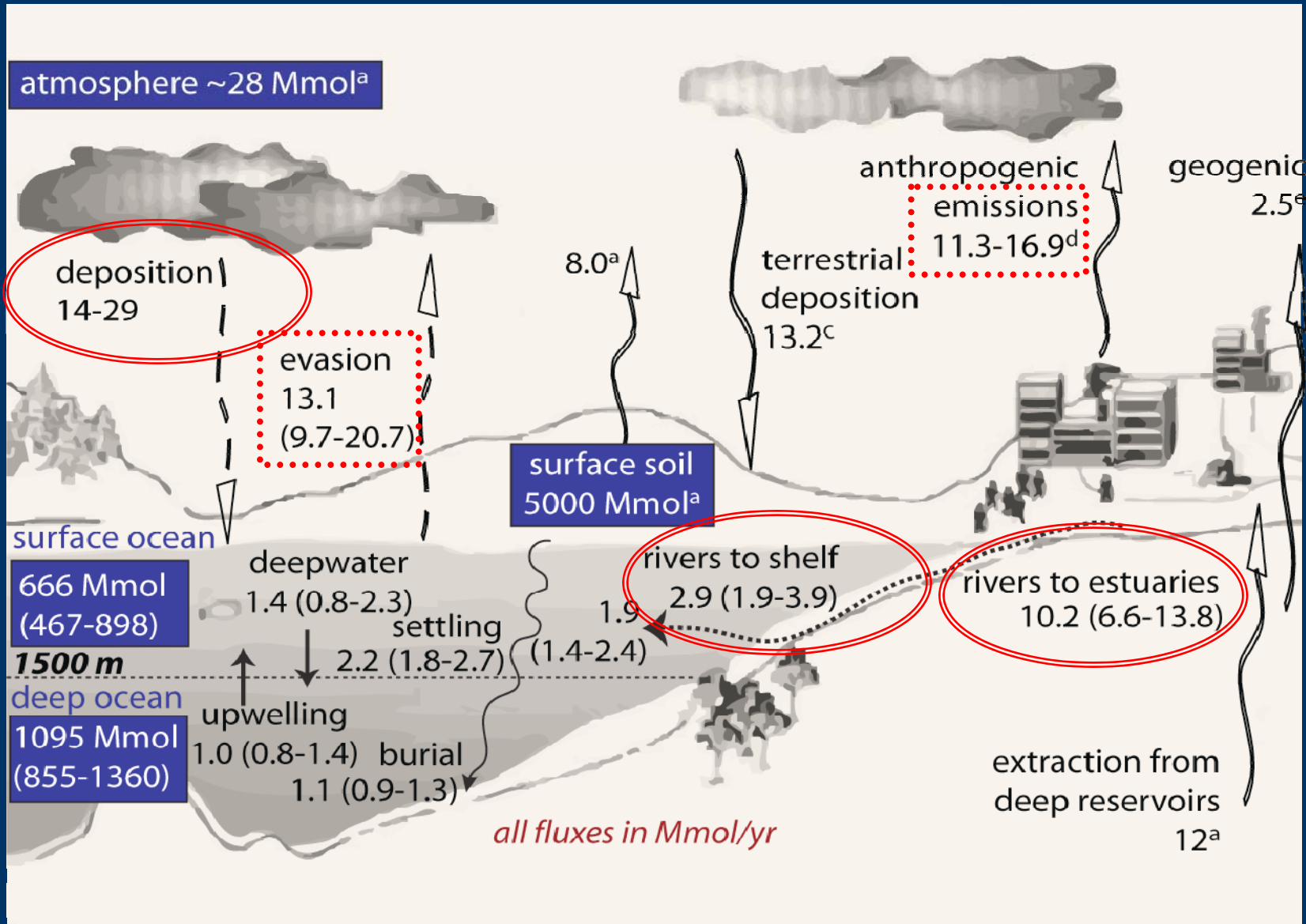


# Consumption Patterns and Percent of Population-Wide Mercury Intake by Fish Species & Geographic Source Region for the “Average” Consumer

Sunderland (2007)



**EPA’s RFD exceeded by 1% of the “average consumers”**



Source: Sunderland & Mason, 2007

We know about the inputs to the ocean of inorganic Hg (mostly atmospheric). But, where does the methylmercury come from?

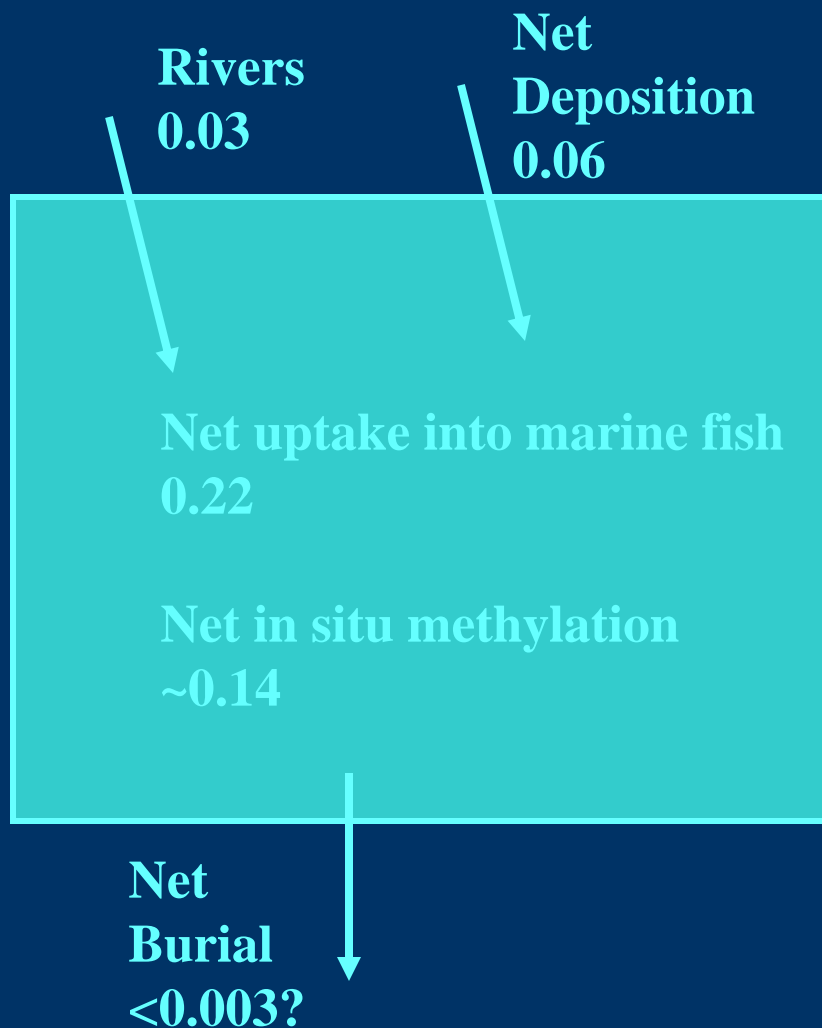
Ratio: Pluvial/Fluvial

Hg	~15
MeHg	~2

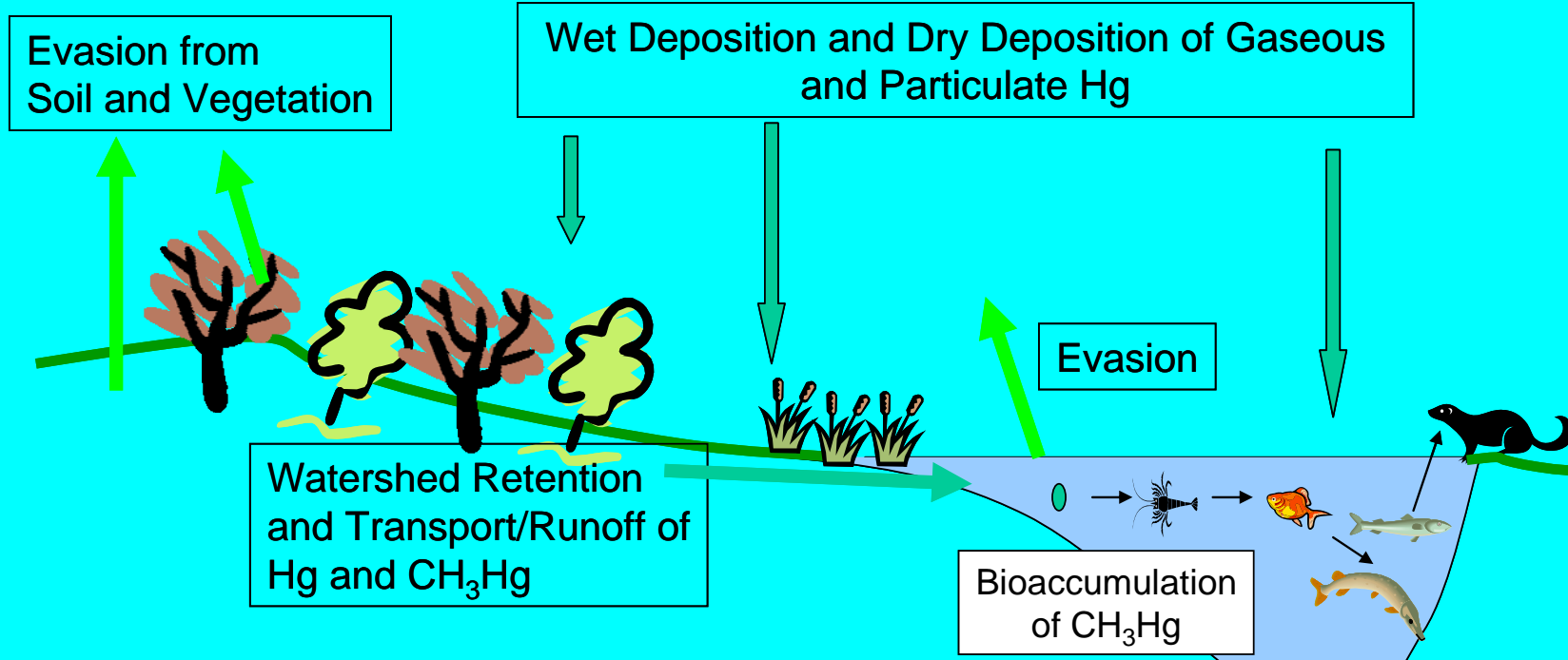
Major source of MeHg is *in situ* production

But where?

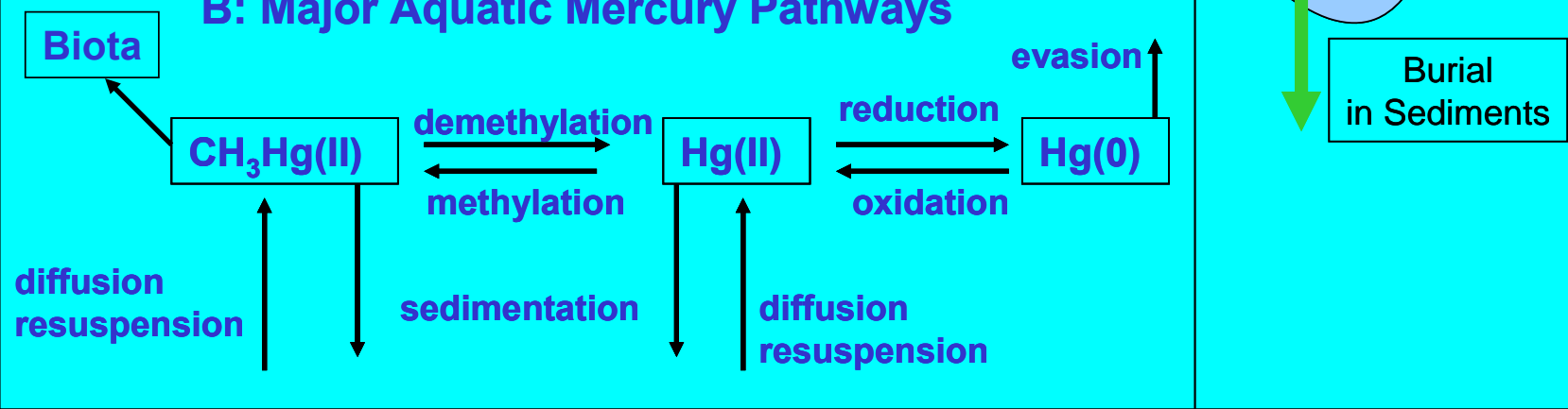
1. Deep ocean sediments (ala Kraepiel et al., 2003)
2. Water column in low oxygen environments (ala Mason et al)
3. Shelf and slope sediments (ala Mason and Fitzgerald et al)



## A: Major Ecosystem Inputs and Outputs of Mercury

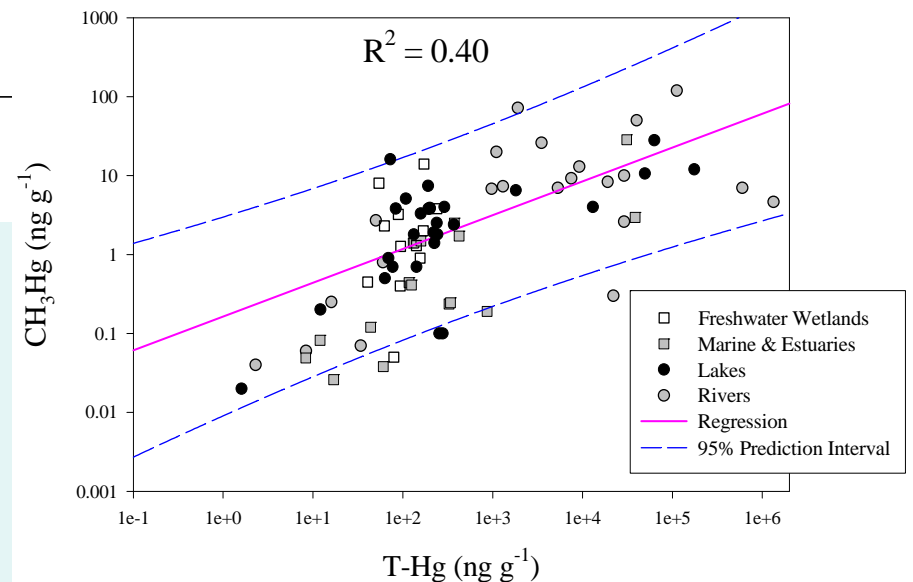
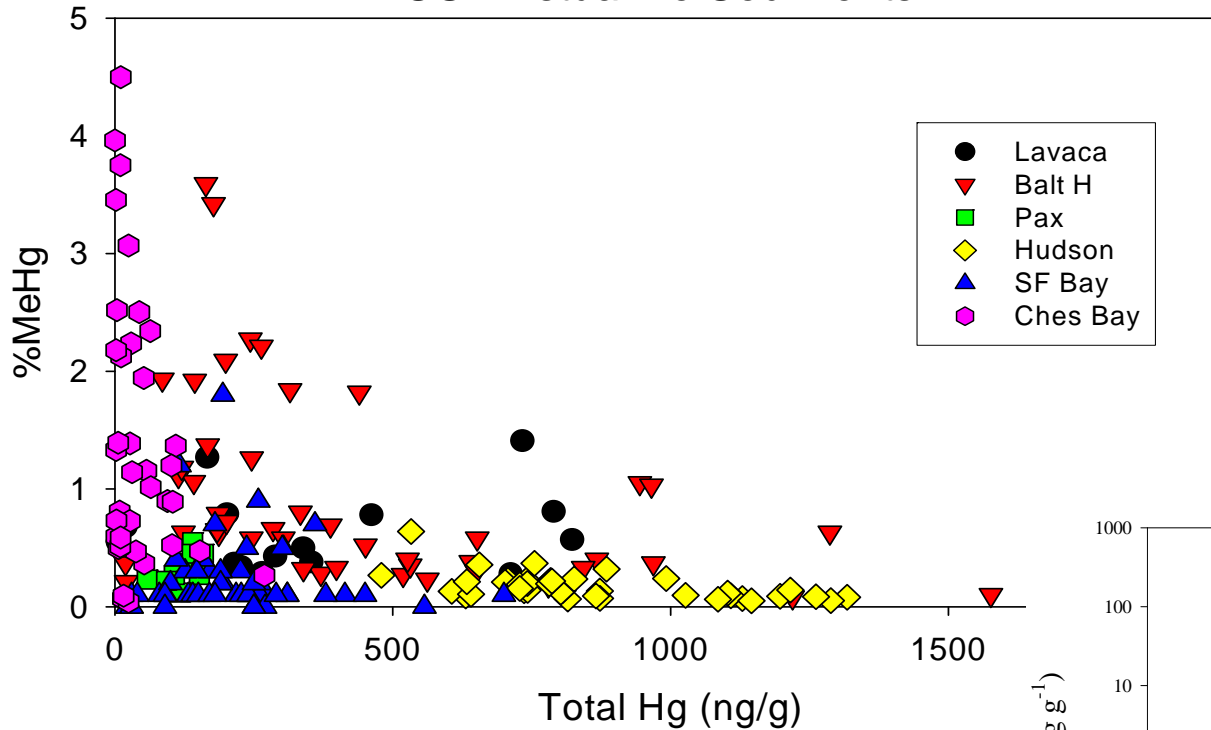


## B: Major Aquatic Mercury Pathways

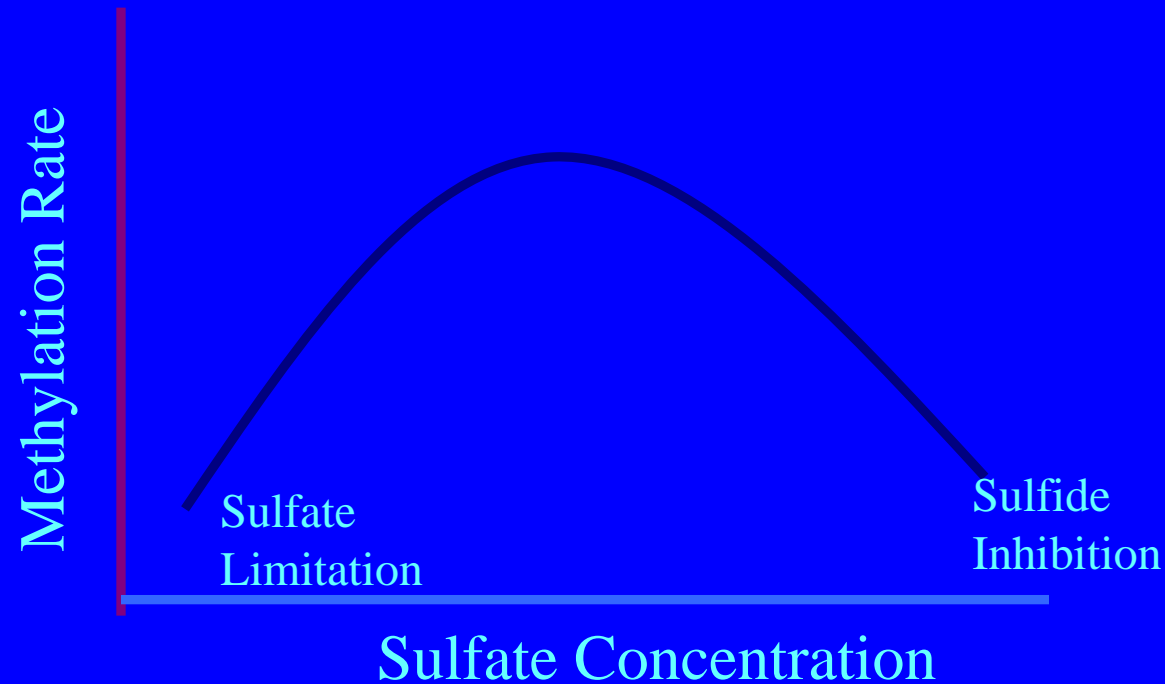


The fraction of total mercury that is methylmercury appears to decrease with increasing mercury concentration in estuaries. Also, it is different for different types of ecosystems

Total Hg versus %MeHg in USA Estuarine Sediments



## An Often-Observed Empirical Relationship Between Mercury Methylation and Sulfate/Sulfide levels



At low sulfate, microbes are limited by sulfate concentration and methylation is lower.

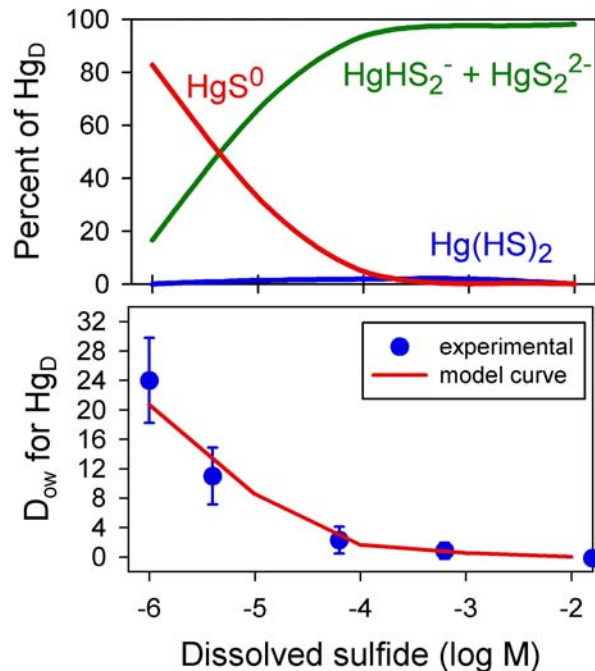


At high sulfate, inhibition of methylation appears to coincide with high sulfide. However precipitation of insoluble  $\text{HgS}$  does not appear to be controlling factor

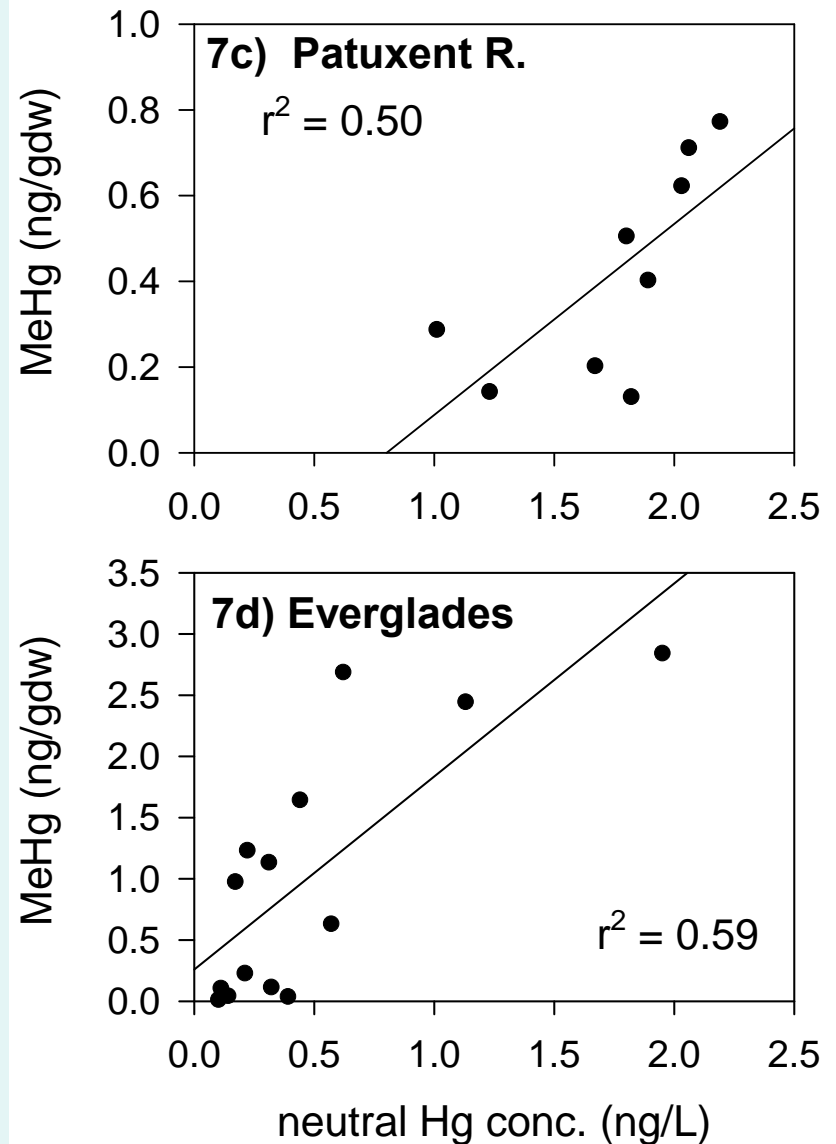


# The Role of Mercury Speciation

Hg speciation and octanol-water partitioning at pH 7

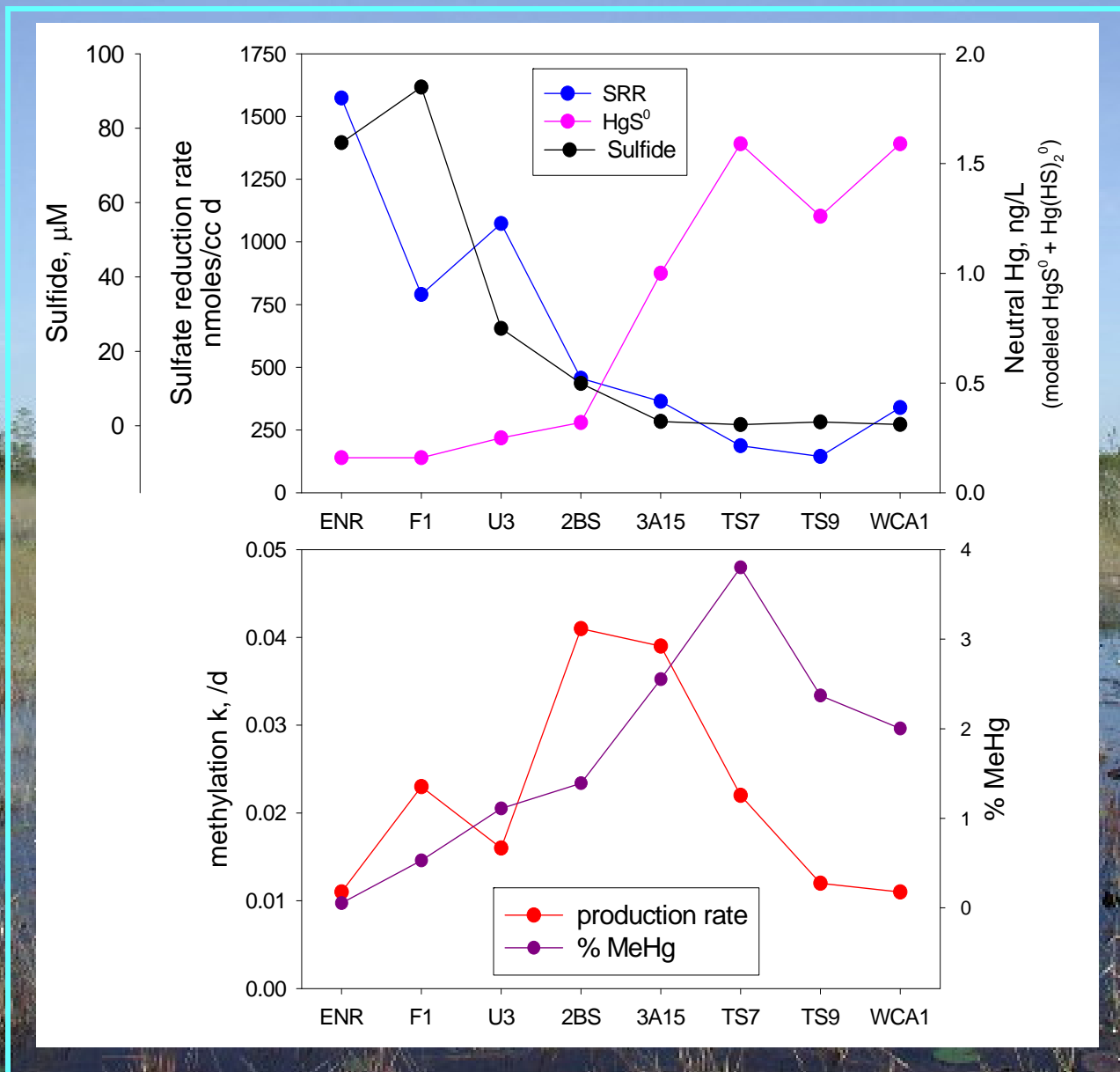


**Hypothesis:** Neutral Hg-sulfide species concentration controls methylation rate as this is the form of the Hg taken up by bacteria. The fraction as the neutral complexes is a function of sulfide conc. Thus, methylation rate, or in-situ MeHg conc. is a function of dissolved speciation



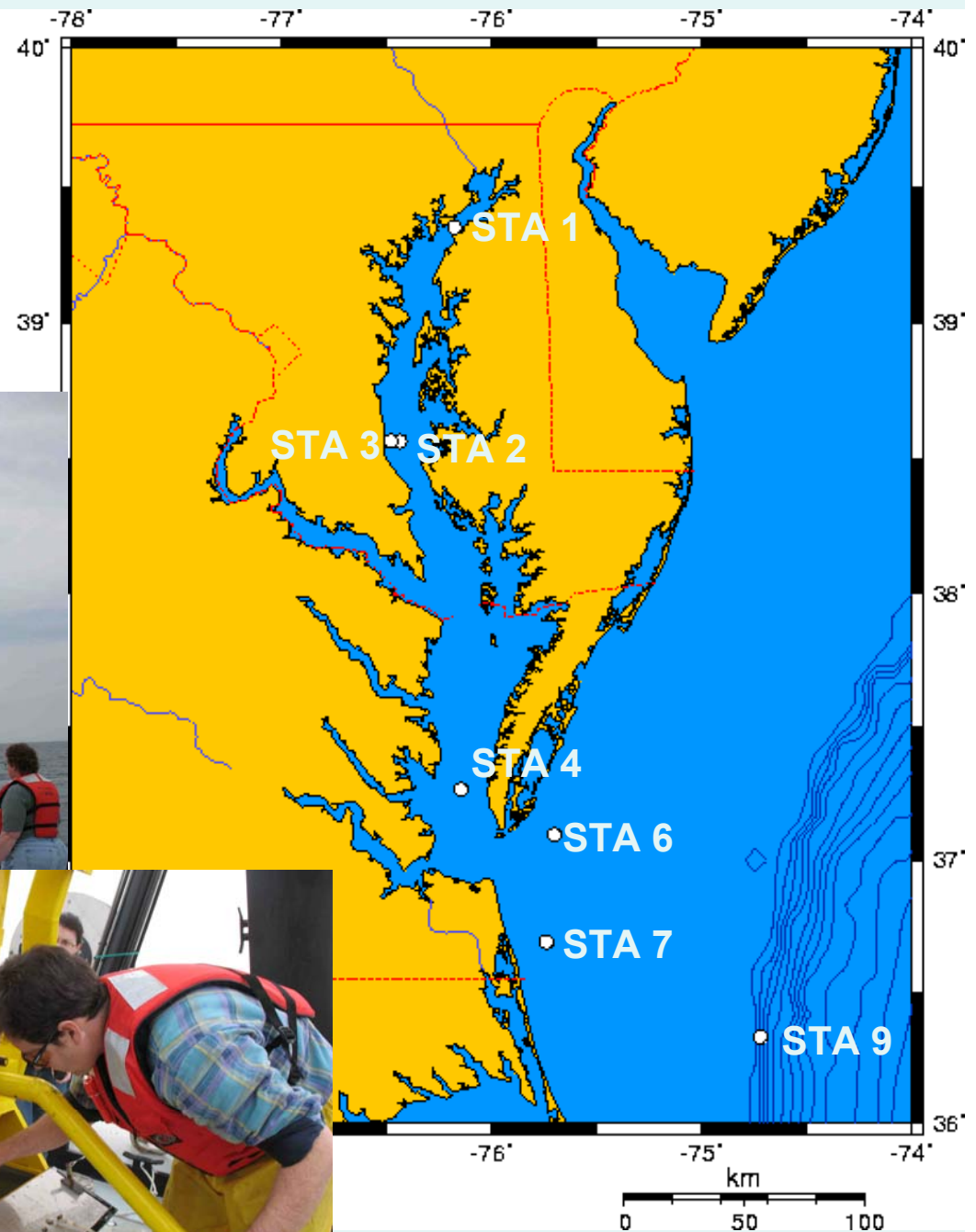
# Everglades – Changes across stations in ~ N-S direction

## Illustrates the importance of chemistry and Hg speciation



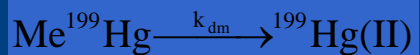
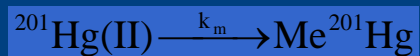
# Experimental Method

Undisturbed sediment was collected from stations in the Chesapeake Bay and mid-Atlantic shelf and slope



# Experimental Method

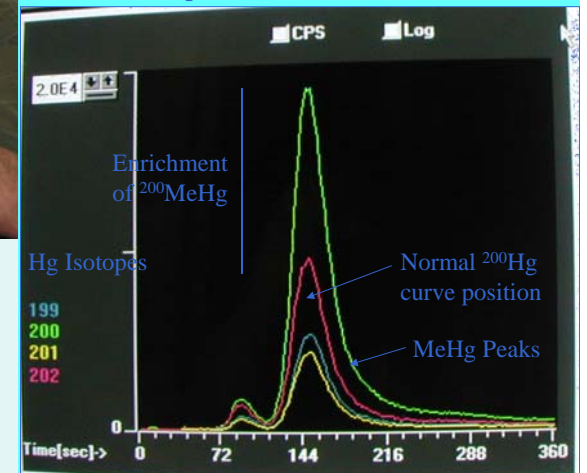
- Solid phase
  - Hg<sub>T</sub>
  - MeHg
  - Ancillary - C, N, S, Fe(II), Fe(III)
- Pore water
  - Hg<sub>T</sub>
  - MeHg
  - Ancillary - SO<sub>4</sub><sup>2-</sup>, HS<sup>-</sup>, Cl<sup>-</sup>, Mn, Fe
- Bacterial activity
  - CO<sub>2</sub> and CH<sub>4</sub> production
  - SO<sub>4</sub><sup>2-</sup> reduction
- Hg methylation and MeHg demethylation rates
  - Stable mercury isotope incubations



$$k_m = \frac{[\text{Me}^{201}\text{Hg}]}{[^{201}\text{Hg}]t} \quad k_{dm} = \frac{1}{t} \times \ln \frac{[\text{Me}^{199}\text{Hg}]}{[\text{Me}^{199}\text{Hg}]_0}$$

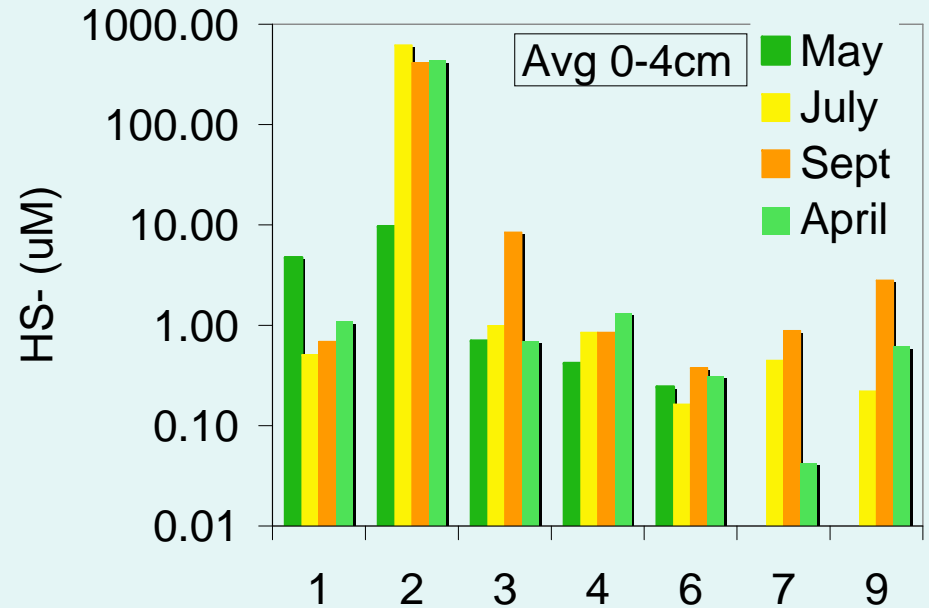
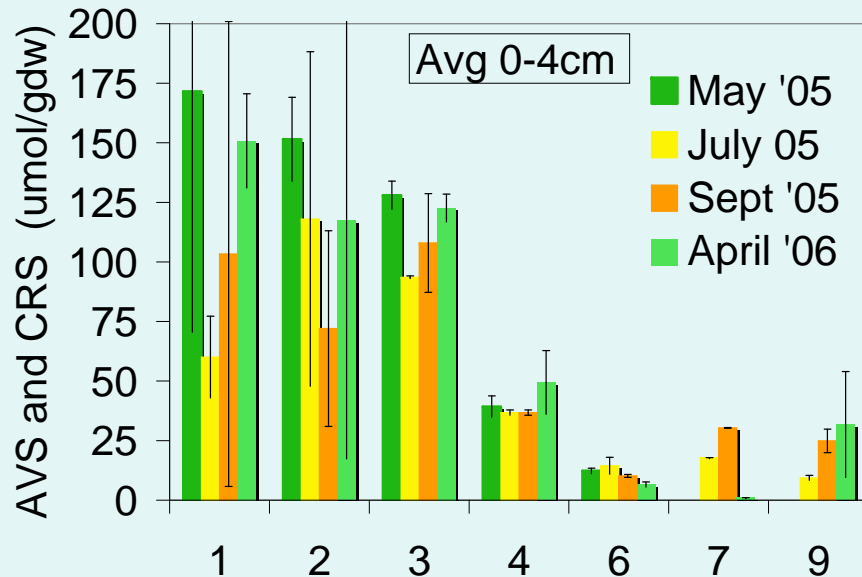
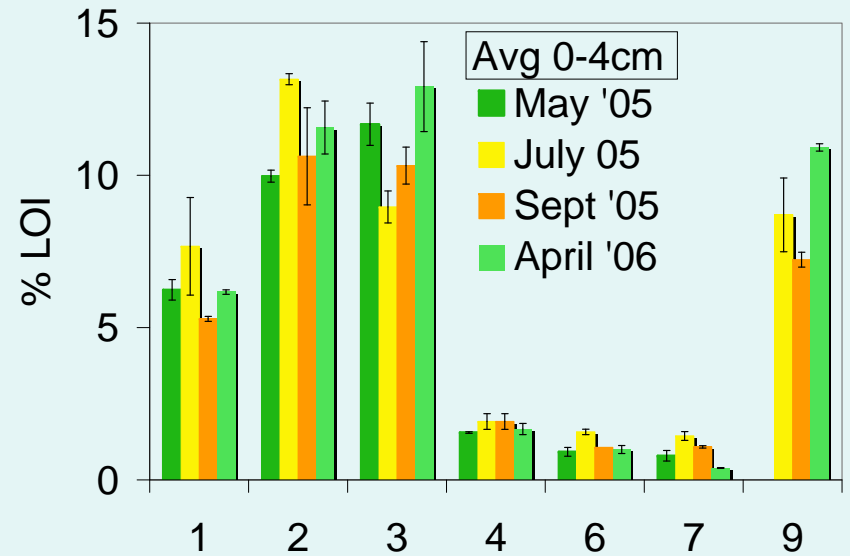
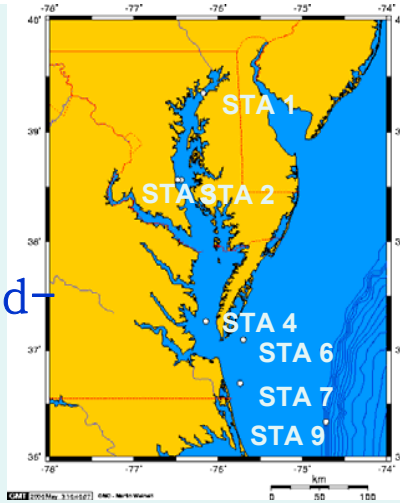


MeHg analysis by aqueous ethylation  
GC separation and ICP-MS detection

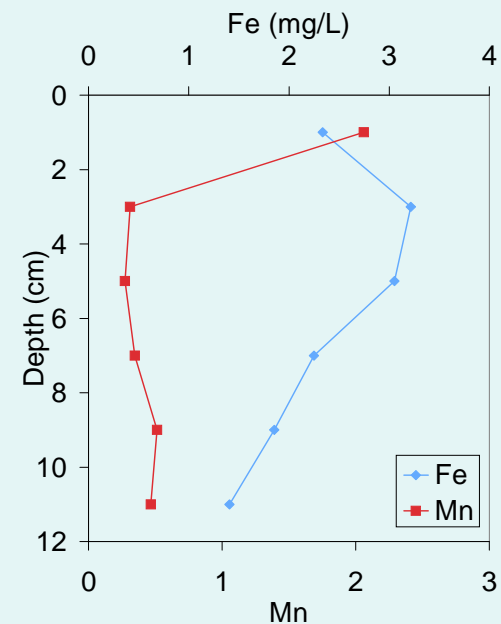
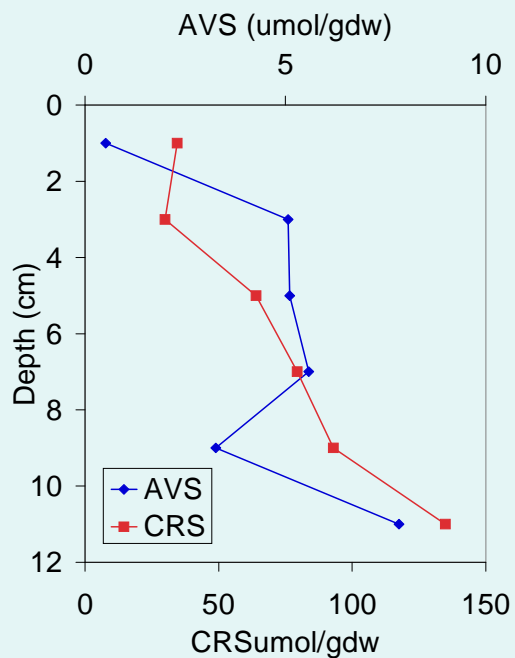
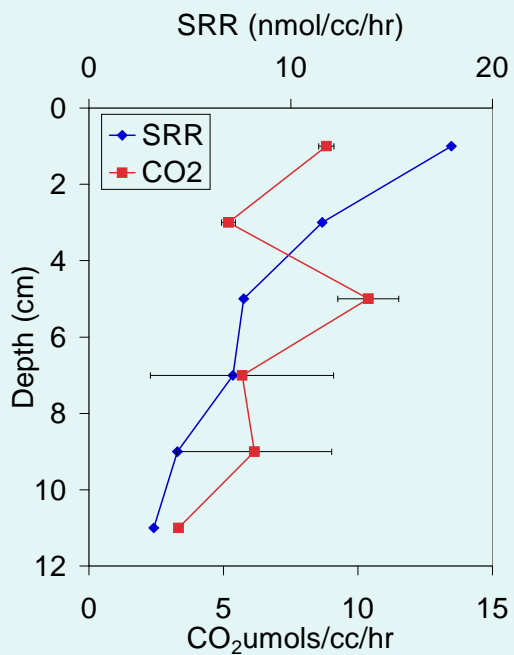
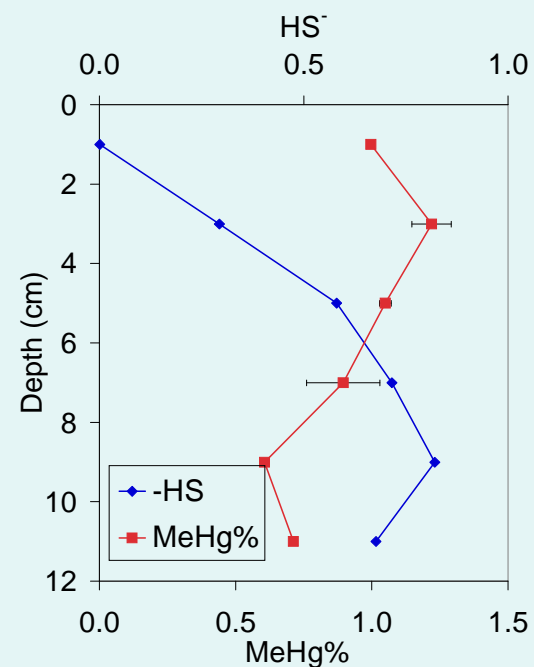
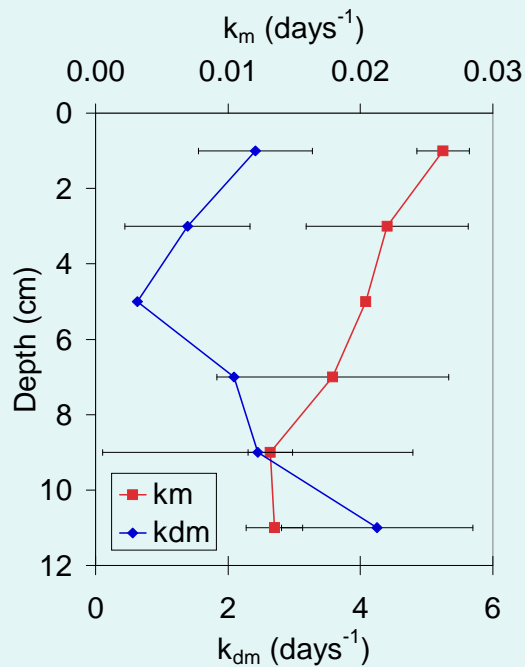
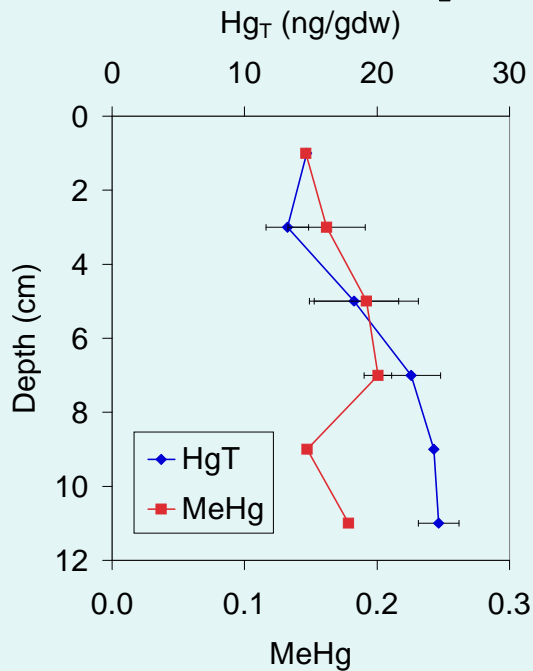


# Results – Ancillary Data

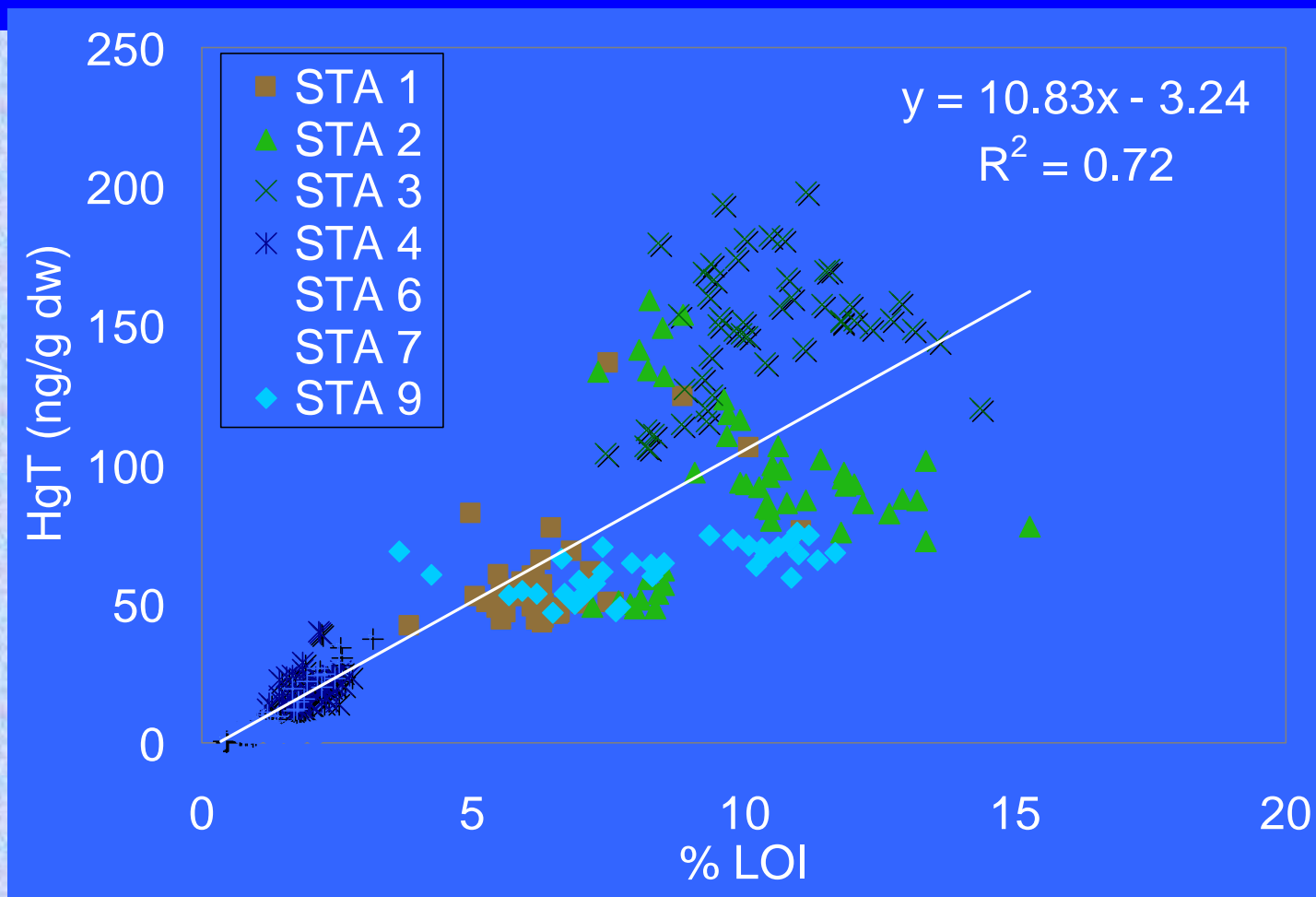
Sediment organic content (as %LOI) is high in the mid-Bay and low on the shelf. The slope (~600 m) site has higher OC than Sts. 4-7. Sulfide varies from very low offshore values to high values at the mid-Bay site. Shows a strong seasonal cycle in the Bay related to sulfate reduction



# Vertical profiles for the sediments at St. 4

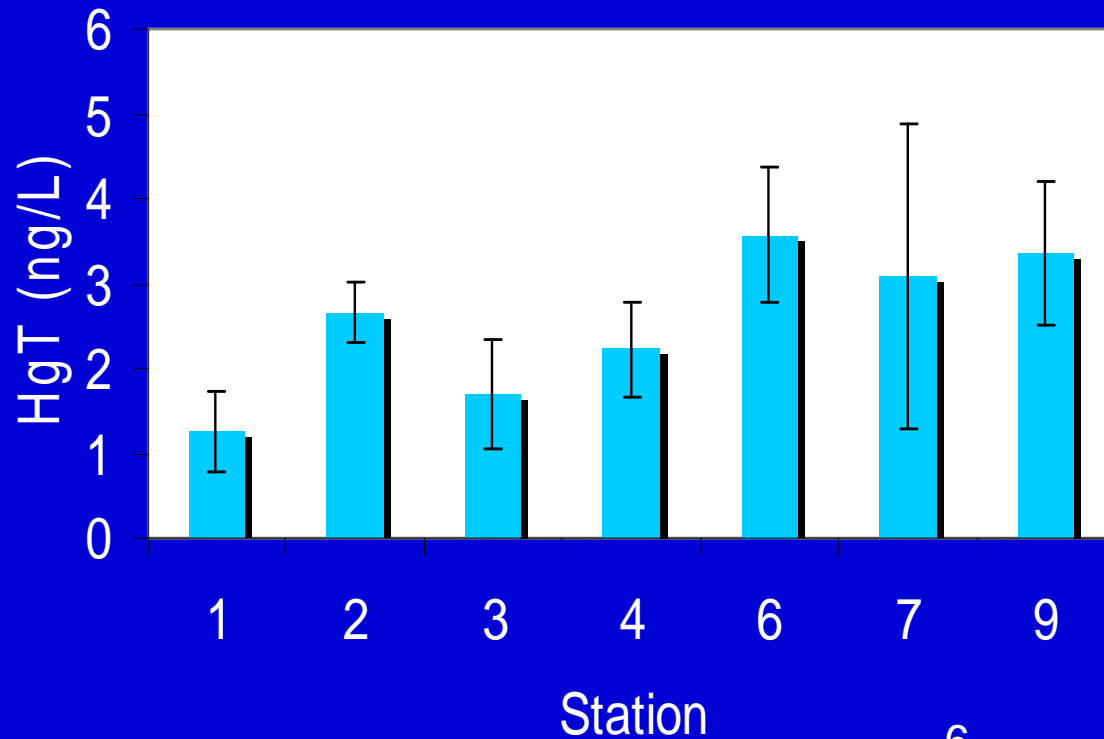


# Results - Controls of Hg<sub>T</sub> speciation



	Adj $r^2$	y-intercept	% LOI	AVS/CRS	HS-
1	0.724	-5.16	11.0		
2	0.769	-10.97	8.91	0.180	
3	0.810	-15.00	9.85	0.210	-0.0444

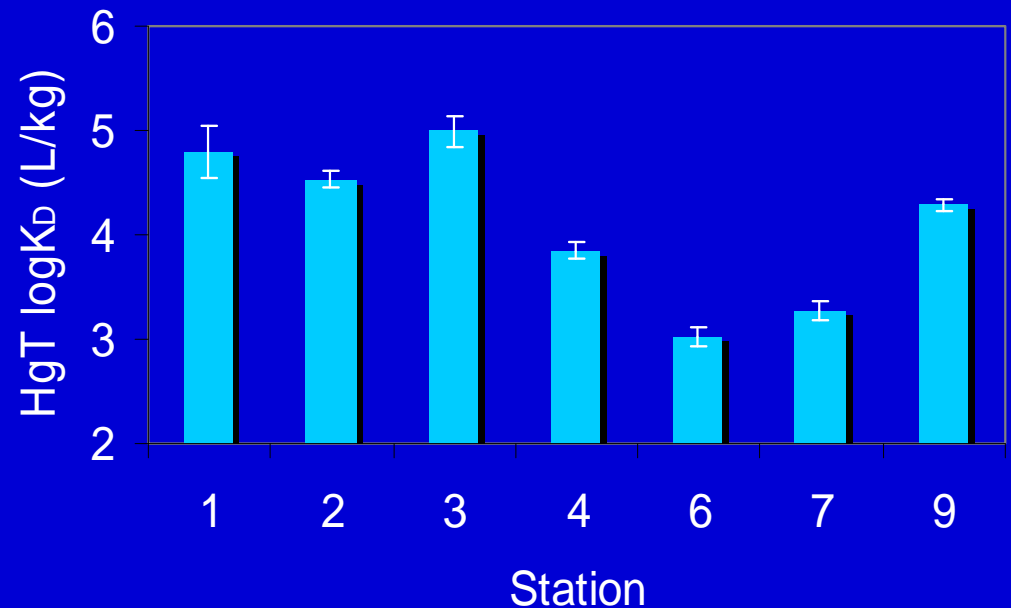
Fe(II) and Fe(III) not significant.



Porewater concentrations are of the same order for the offshore sandy sites compared to the organic rich Bay sites, suggesting differences in the dissolved-solid partitioning between these locations

$$K_D = \frac{[Hg_{solid}]}{[Hg_{dissolved}]}$$

In the upper sediments, therefore, the  $K_D$  differences are largely driven by differences in the solid phase concentration



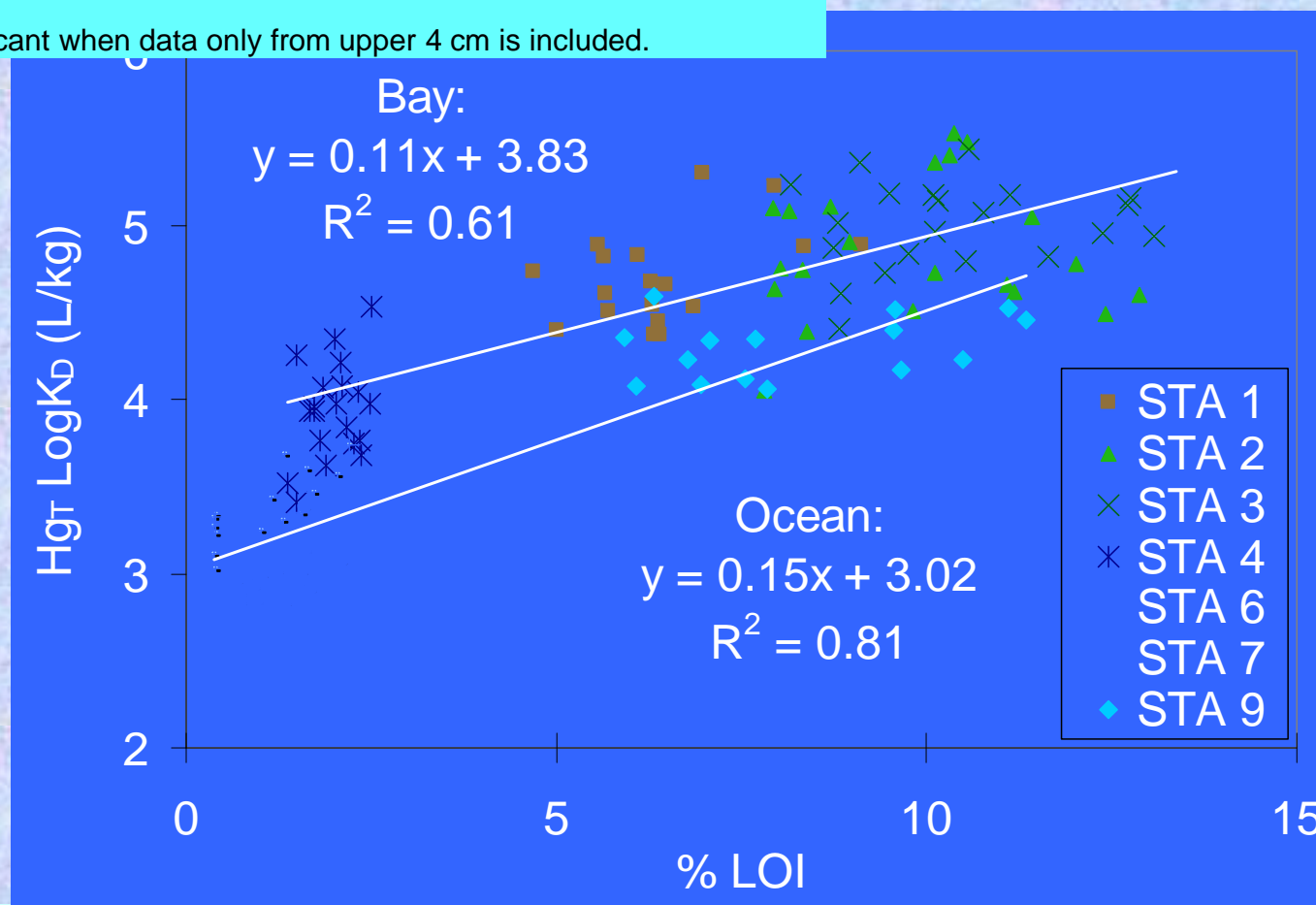


Stepwise Multiple Linear Regression for Hg<sub>T</sub> LogK<sub>D</sub>

	Adj. r <sup>2</sup>	y-intercept	% LOI	AVS/CRS
1	0.726	3.27	0.161	
2	0.810	3.18	0.113	0.00375

Not significant: HS<sup>-</sup>, Fe(II), Fe(III)

HS<sup>-</sup> becomes significant when data only from upper 4 cm is included.

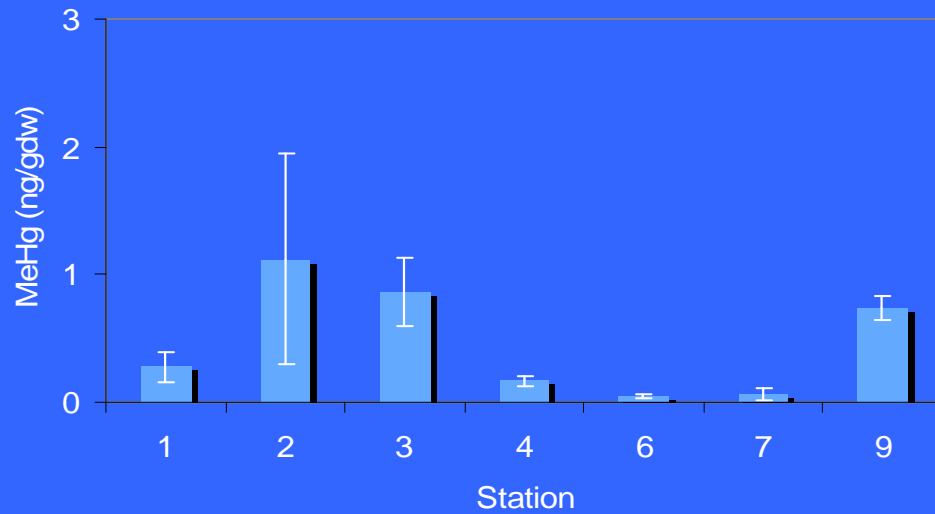


Hammerschmidt and Fitzgerald (2006): LIS and Shelf

$\text{Log}(K_D) = 0.15[\%OM] + 3.13; r^2 = 0.75, p < 0.01$

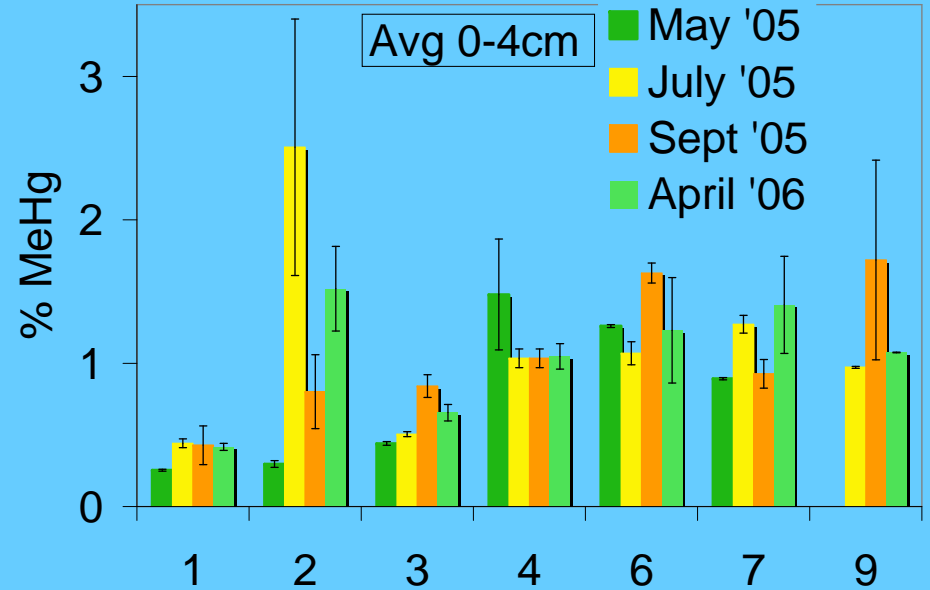
Senn et al (in review) Gulf of Mexico; Much lower %OM - <3%

$\text{Log}(K_D) = 0.59[\%OM] + 3.2; r^2 = 0.65, p < 0.01$

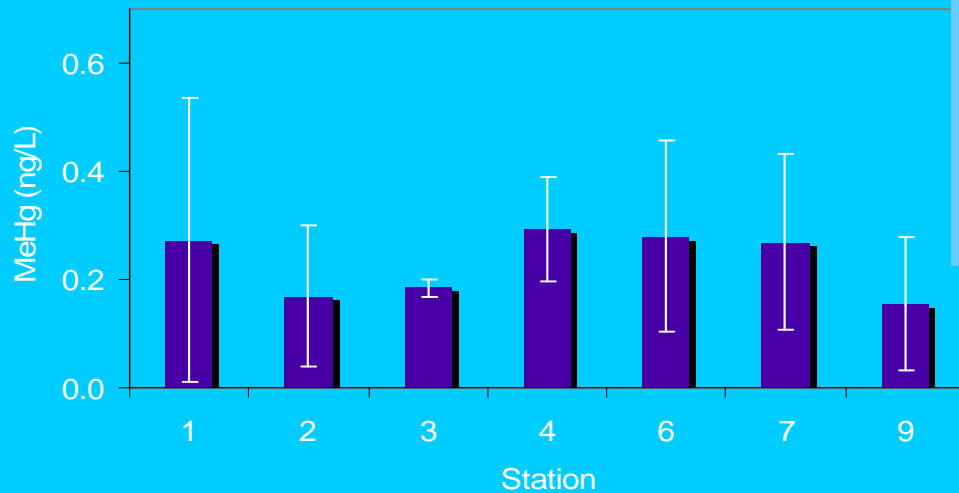


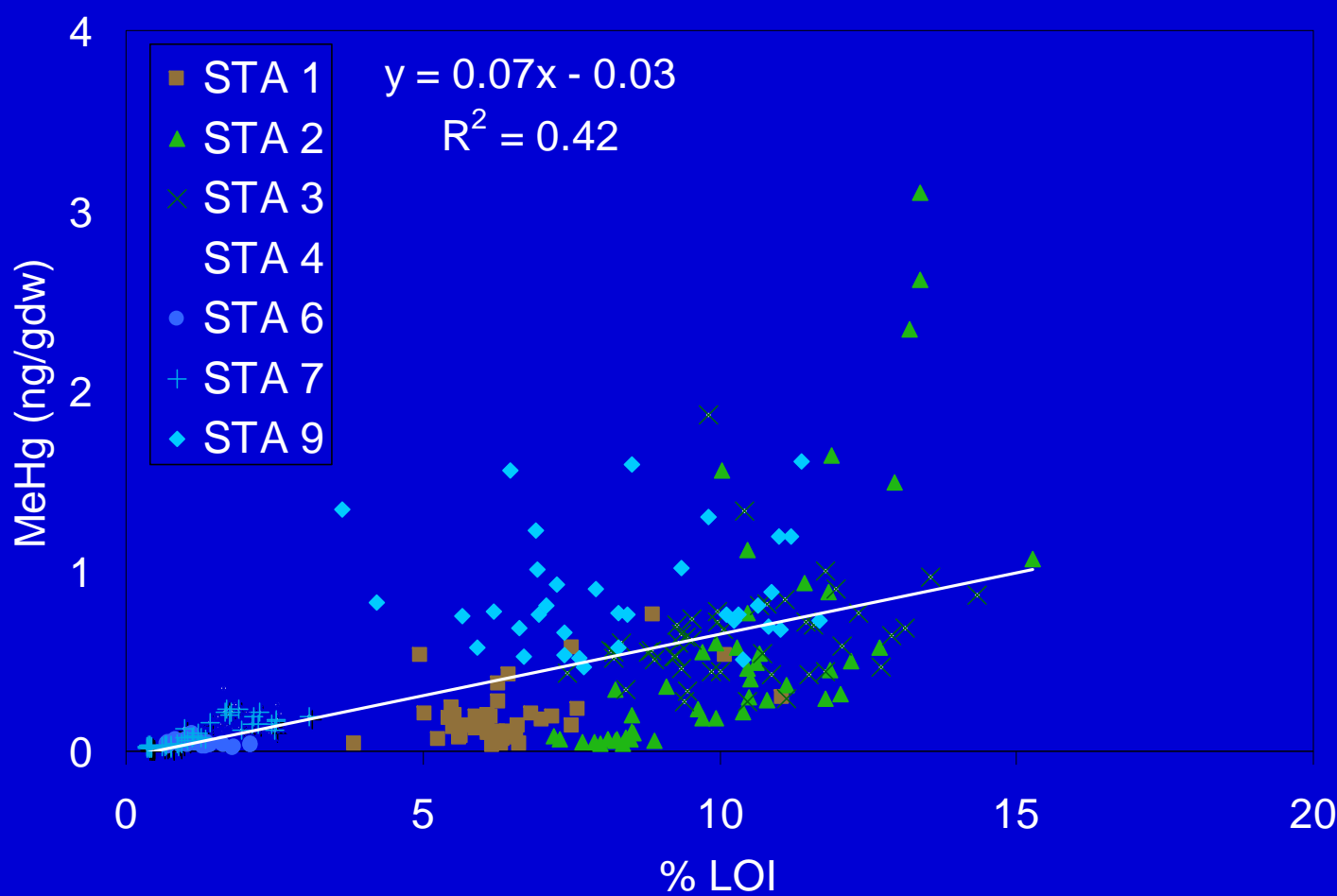
While MeHg concentrations are higher in the bulk phase for the high OC Bay sites, there is much less difference in terms of porewater concentration. Overall, the fraction of Hg as MeHg is higher for the offshore sites

Station 2 has high variability in all parameters; may reflect the fact that this site has seasonal water column anoxia; water column methylation occurs at St. 2



The relationship between total Hg and MeHg is not strong, as found elsewhere. %MeHg is relatively high cf. other coastal systems.





Hammerschmidt  
 and Fitzgerald  
 (2006): LIS/Shelf  
 $y = 0.13x + 1.55$   
 $r^2 = 0.77$

Stepwise Multiple Linear Regression for bulk-phase MeHg concentration

	Adj. $r^2$	y-intercept	% LOI	AVS/CRS	$k_m$
1	0.419	-0.0353	0.0682		
2	0.605	0.0650	0.101	-0.00288	
3	0.630	-0.0440	0.0966	-0.00236	3.84

Not significant:  $Hg_{T(pw)}$ ,  $Fe_{dis}$ ,  $SO_4$ ,  $k_m * Hg_{T(pw)}$ ,  $HS^-$

Sites sampled in July 06., in estuary, on the shelf and slope, including a transect at Station 9

Sta #	Dpth (m)	Type
STA 2	16	mud
STA 3	6	mud
STA 4	16	sandy mud
STA 6	16	dry clayey sand
STA 7	15	sand
STA 9	646	mud
STA 9A	85	coarse, worms
STA 9B	107	sandy mud
STA 10	48	coarse sand/shell
STA 11	227	clayey sand
STA 12	600	organic, mud
STA 13	30	sand
STA 14	38	sand
STA 15	50	coarse grain

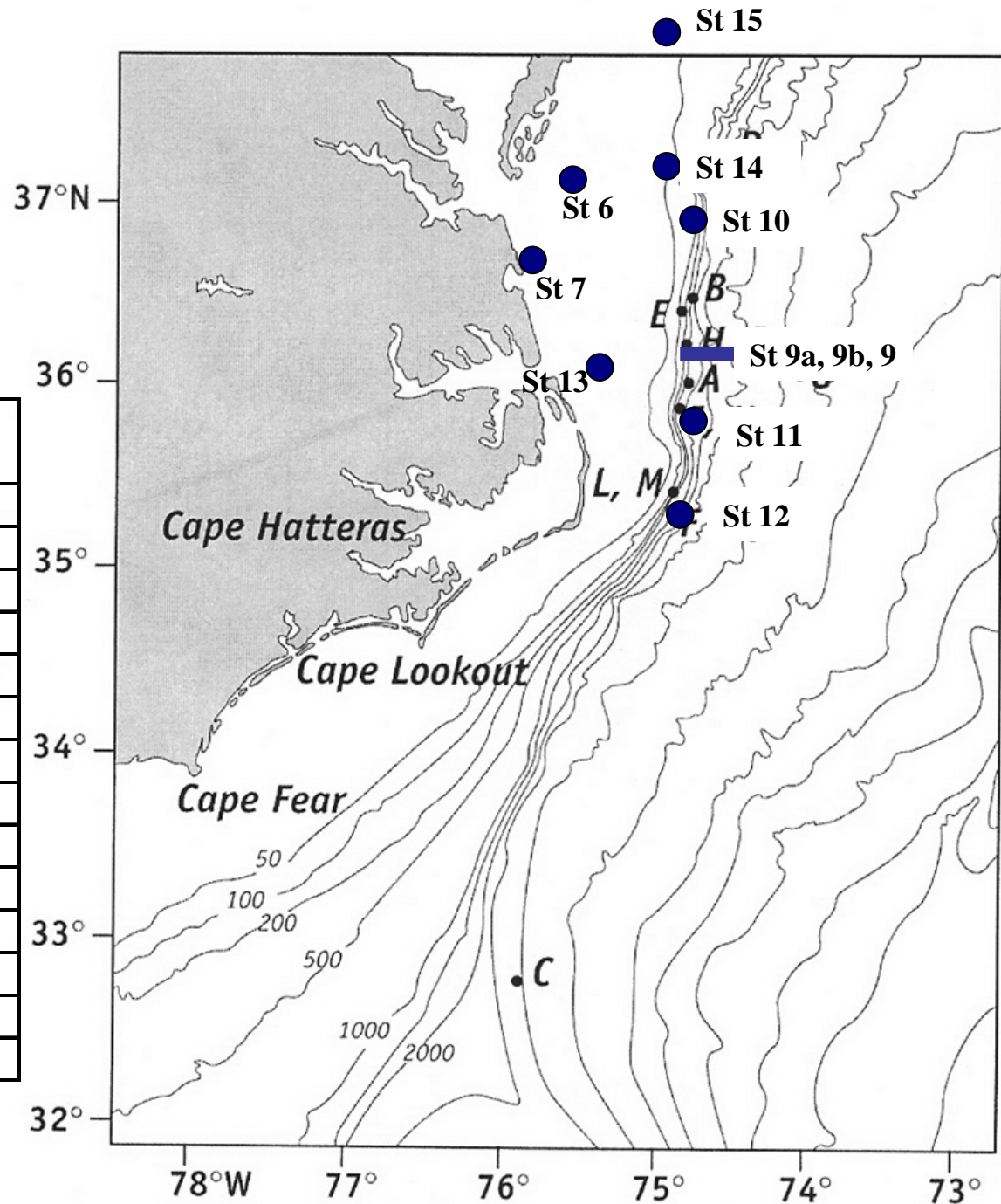
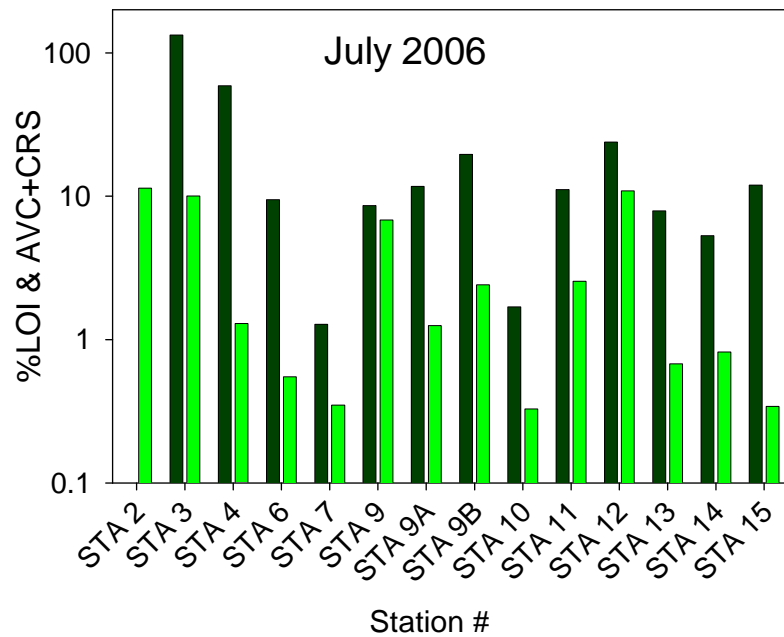
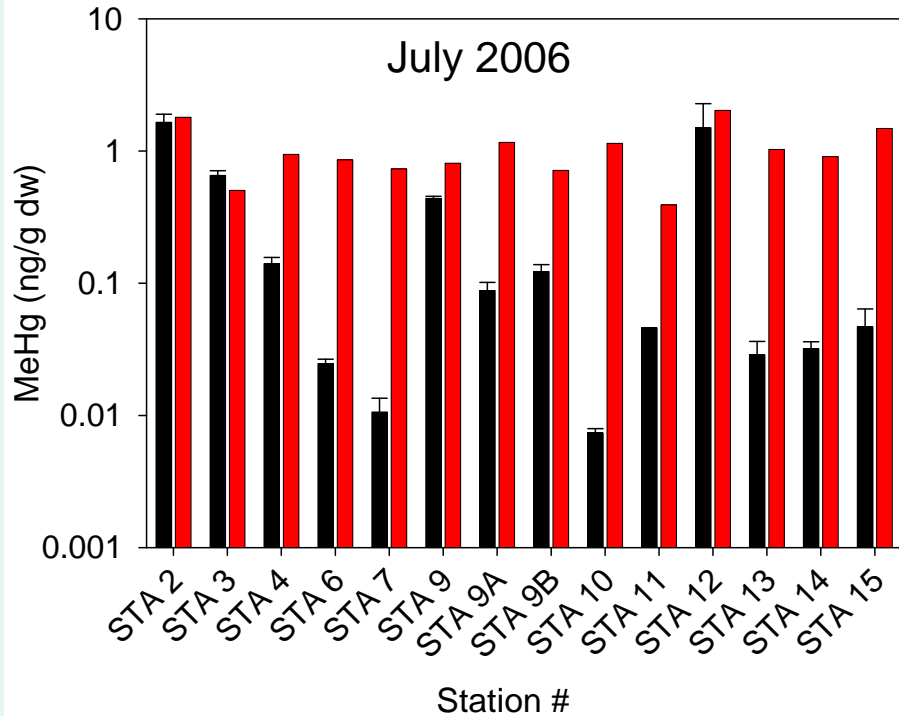
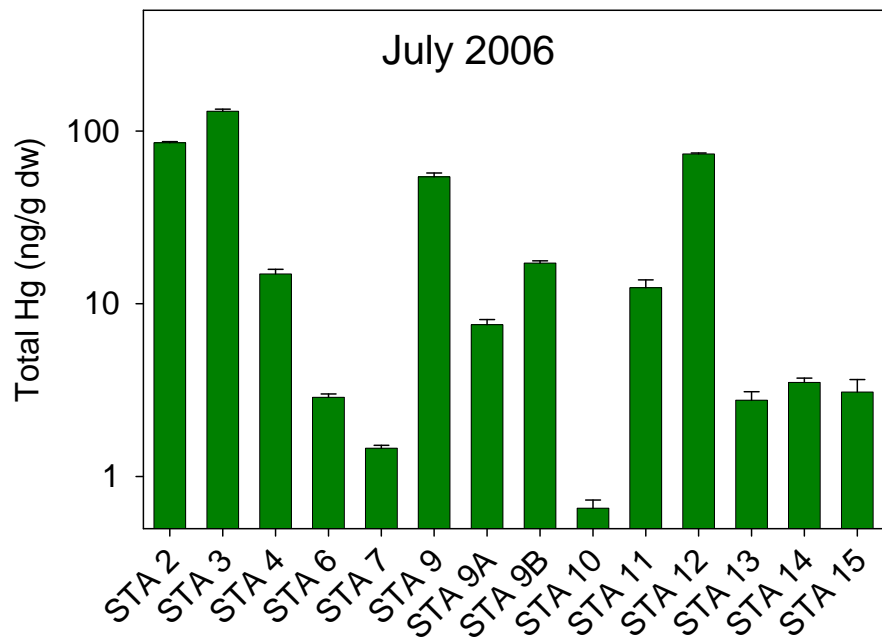
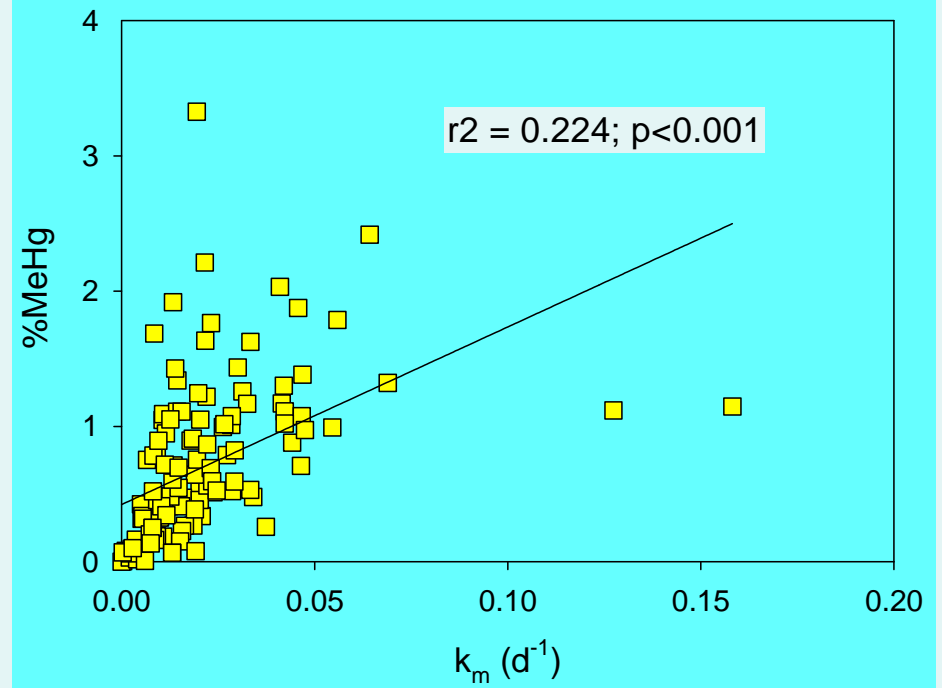
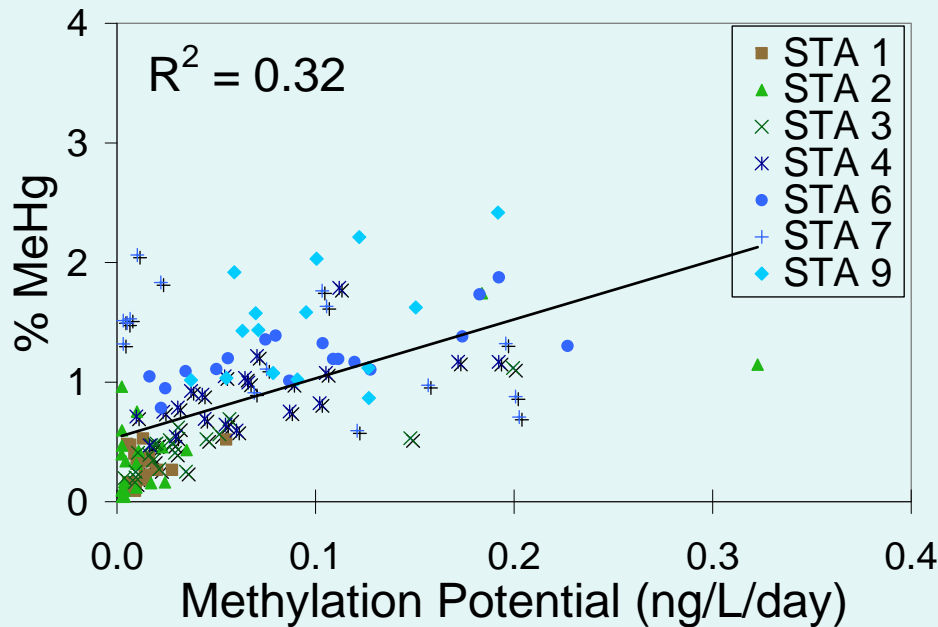
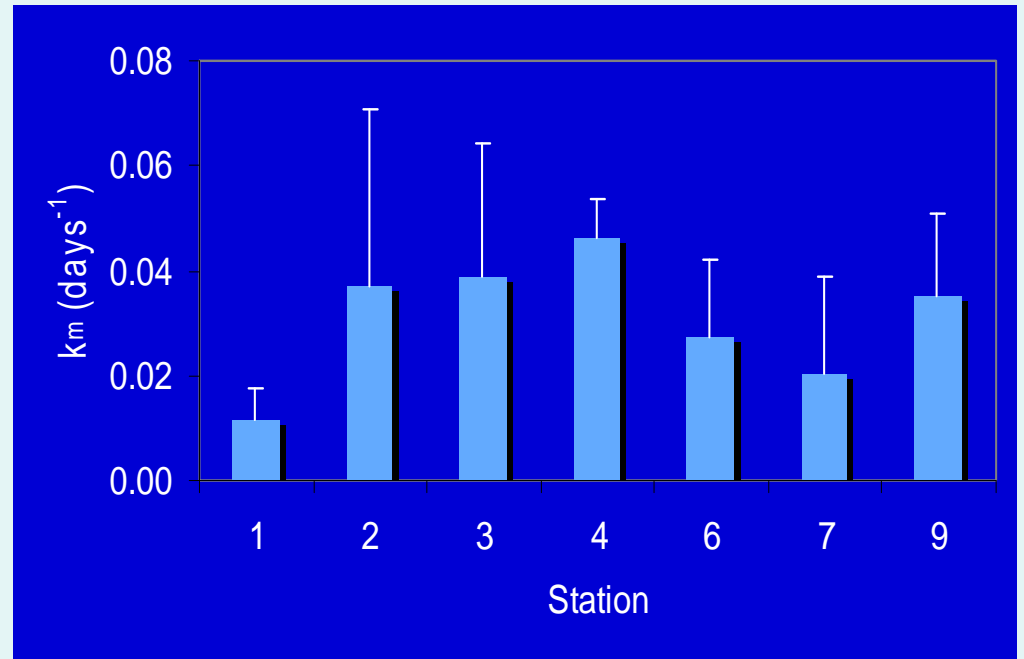


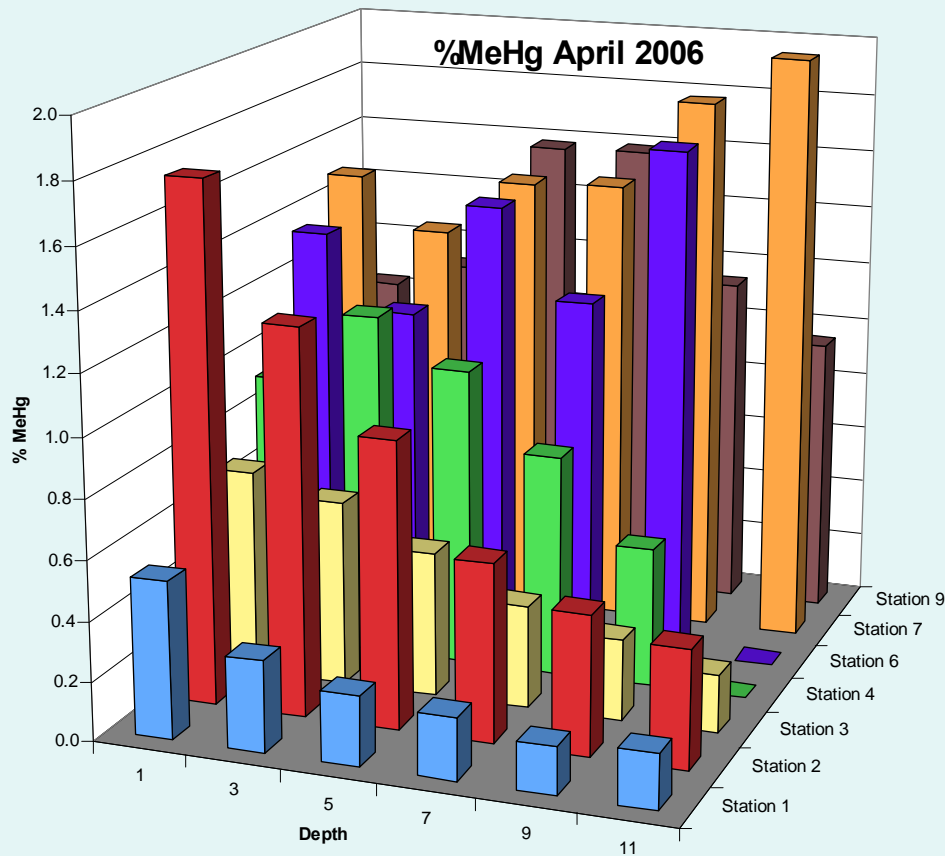
Fig. 1. Locations of in situ benthic flux chamber deployment sites. Depth contours are in meters.



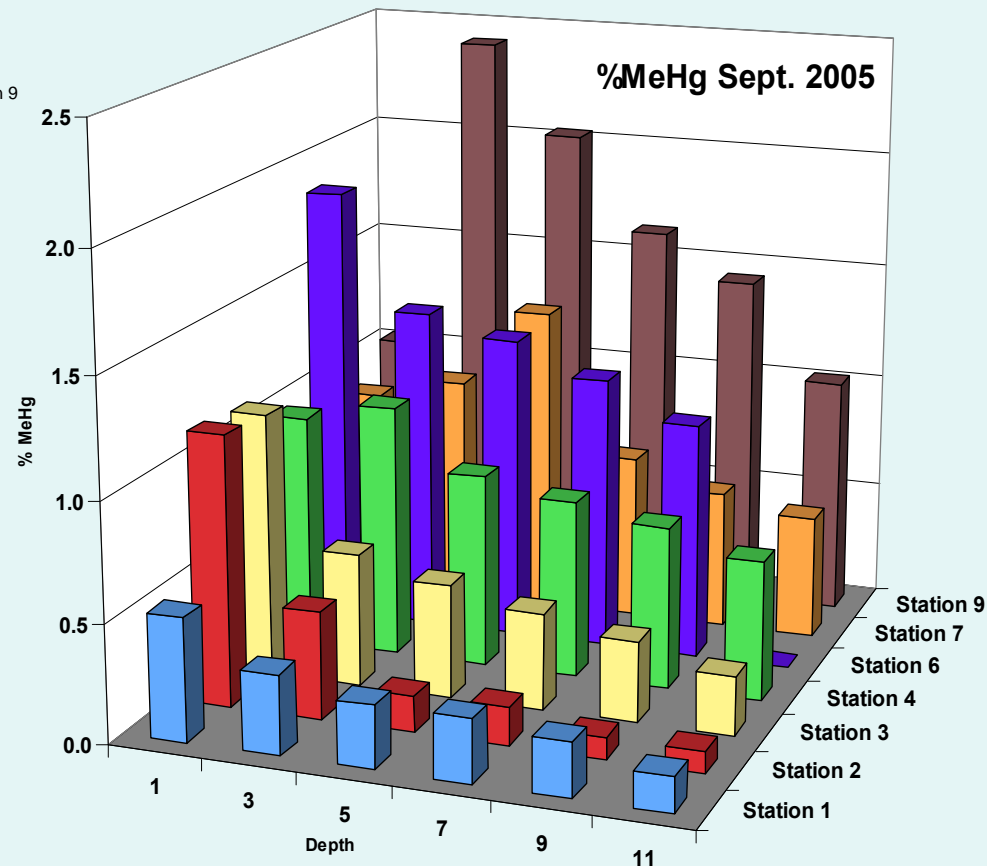
# Mercury Methylation

The estimated methylation rate, and the potential rate ( $k[\text{Hg}_T]$ ), correlate with the in situ %MeHg. Such a relationship is valid if methylation/demethylation are pseudo reversible first order reactions and demeth rates are relatively constant across sites



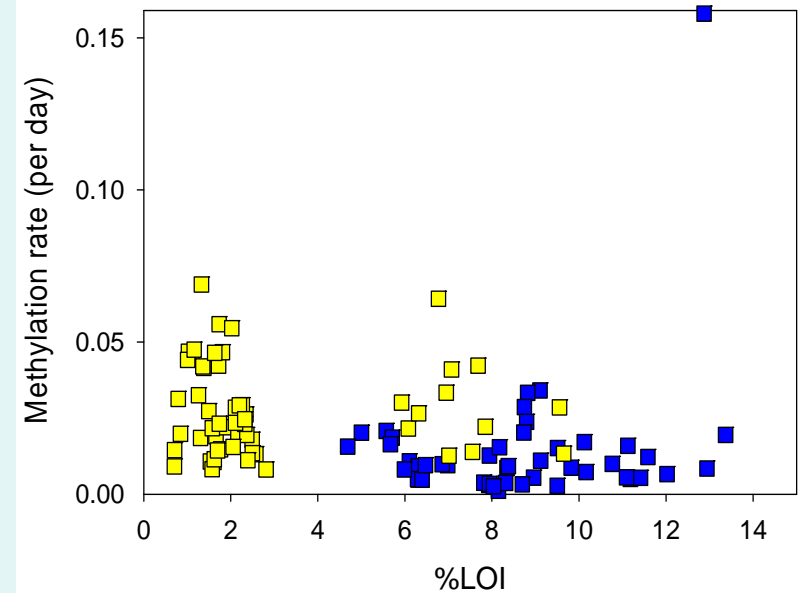
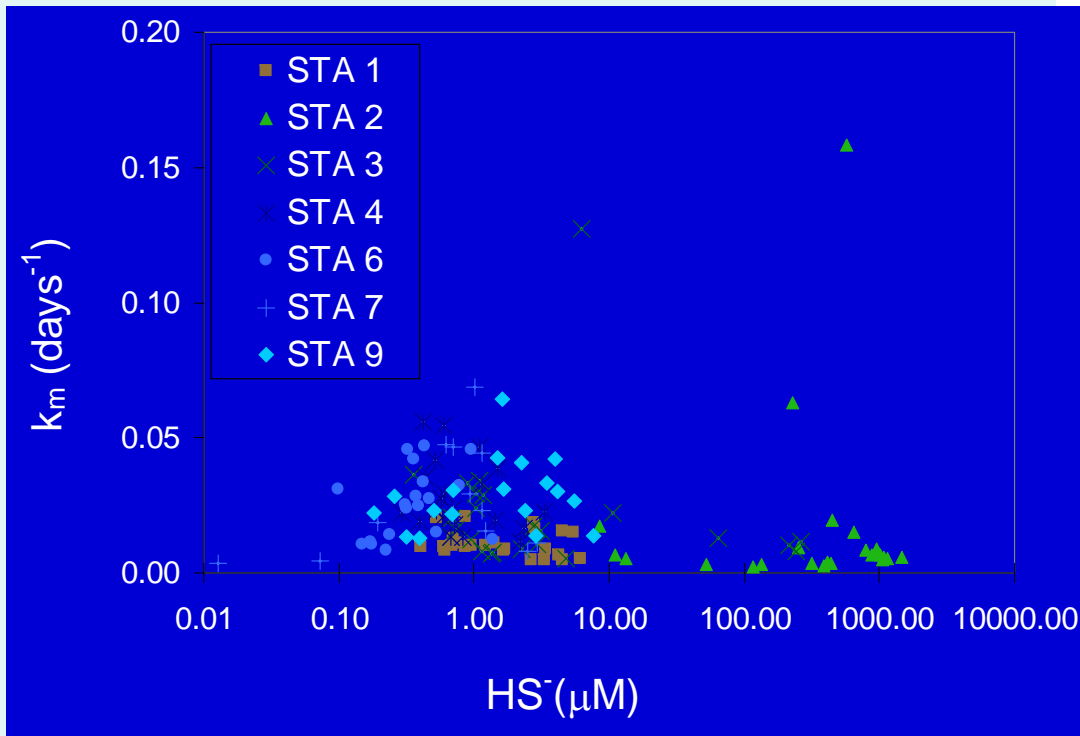
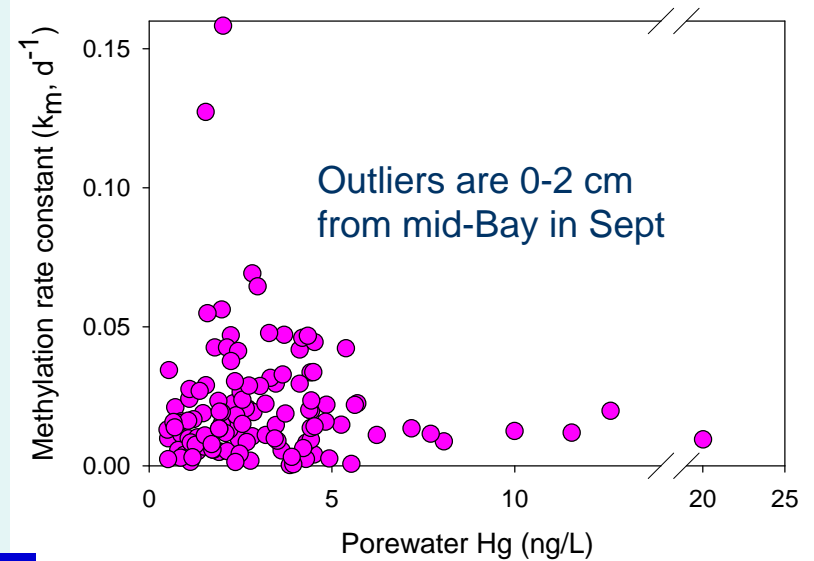


3D Plot of Station #, sediment depth and %MeHg for two of the sampling periods (early spring and early fall) for the Chesapeake Bay/shelf

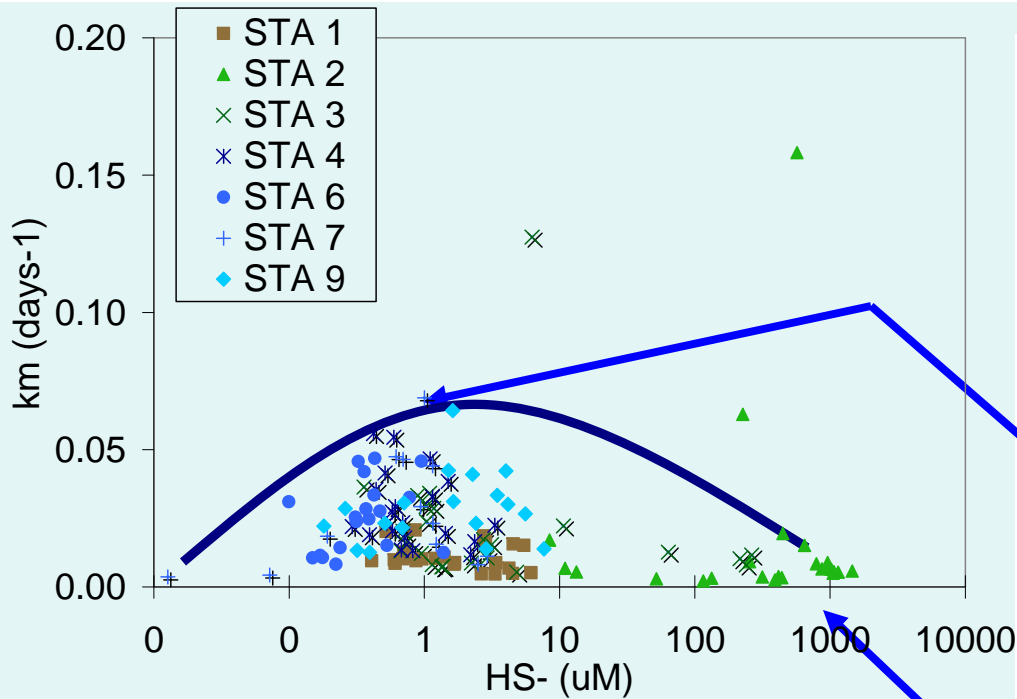


The %MeHg is highest in the surface sediments for the estuarine sites and decreases markedly with depth. For the shelf sites, there is a larger depth interval over which the %MeHg is relatively high. Overall the integrated signal is higher for the shelf sediments

- For the Chesapeake Bay/shelf, the methylation rate constant does not correlate well with porewater Hg
- Does not appear to be strongly related to sediment organic carbon
- Looks to be lower at high sulfide and can be high at low sulfide
- **What's going on?**



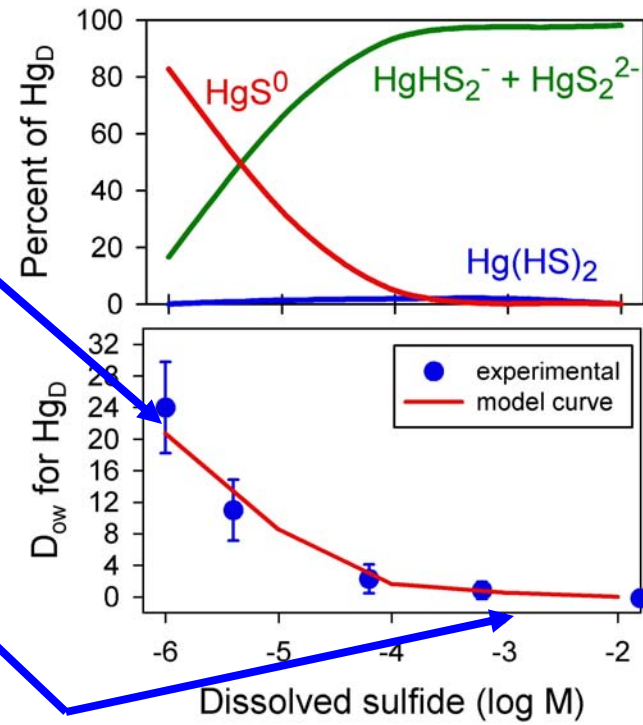




At least, qualitatively, the results fit with the predicted impact of sulfide on Hg bioavailability to the methylating organisms.

But what about the variability at intermediate sulfide? There appears to be other factors that are as important

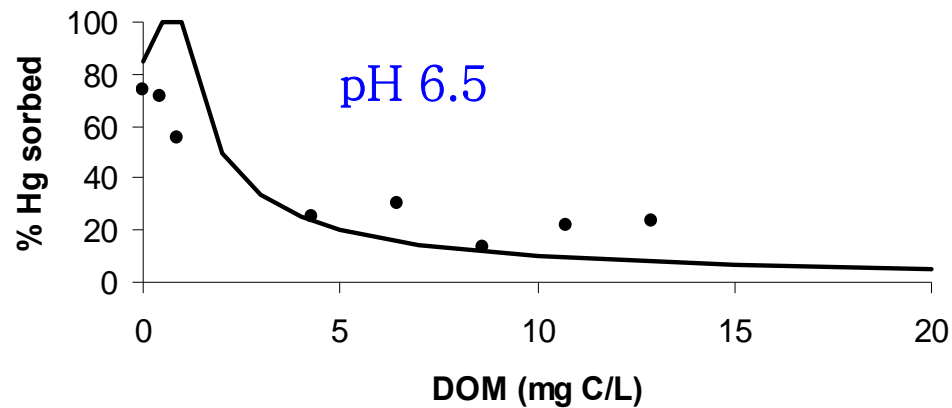
### Hg speciation and octanol-water partitioning at pH 7



Besides sulfide, other factors that could also impact methylation are other complexing agents such as DOC, also pH, and clearly the bacterial community structure

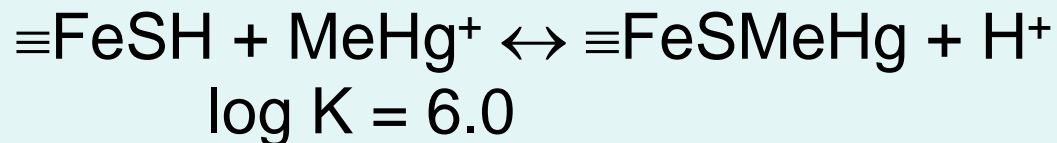
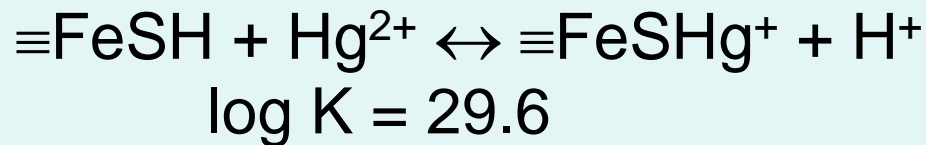
This is all dandy, but what other factors should be considered (DOM, sulfide, solid sulfide etc)?

➤ For hydrous Fe-Oxide, Hg bound strongly in the absence of DOM. Binding was reduced as DOM increased.



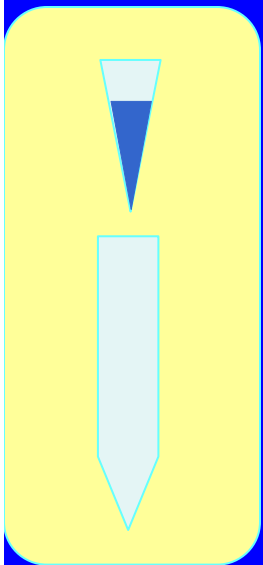
*Hg, MeHg Sorption to FeS?*

- Hg, MeHg strongly adsorbed (>99.9% in lab exps). DOM no effect. Complexation constants were determined.
- In oxic sediments, Hg bound to POC
- Organic matter and FeS likely responsible for sorption under anoxic conditions



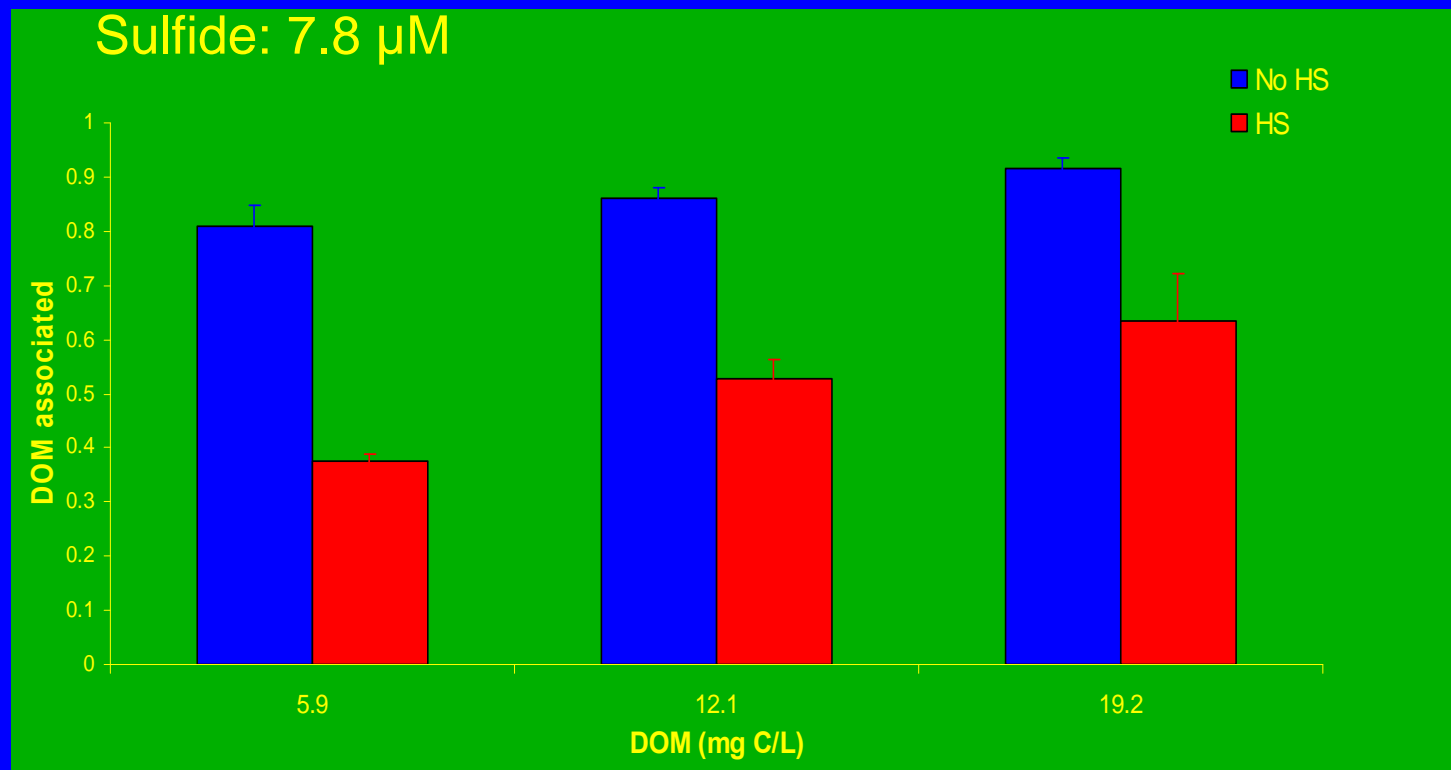
# The interaction between Hg, Sulfide & DOM was investigated using ultrafiltration experiments

- Amicon Ultra 15 mL centrifuge filters (5000 Da)
- DOM > 5000 Da in all experiments
  - Hg associated with DOM remained in extract
- Inorganic Hg-sulfide complexes < 5000 Da
  - Pass through or sorbed onto filter
- Stable isotope used for all experiments



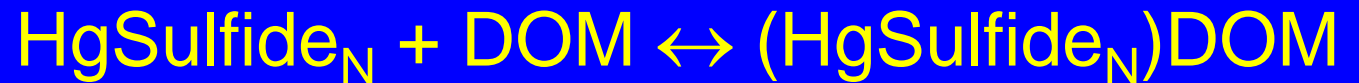
At these sulfide levels, thermodynamic models predict all the Hg should be associated with sulfide

Miller et al., 2007

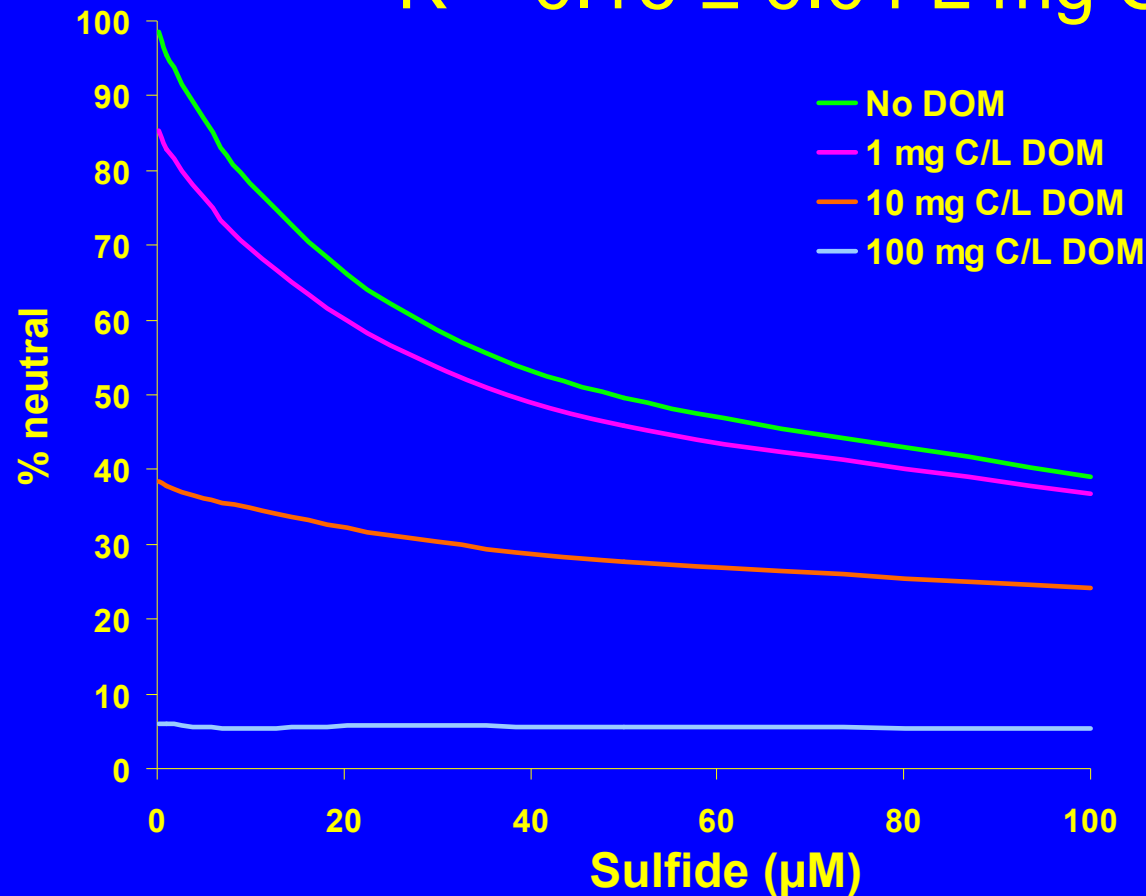


Conclusion: Neutral Hg-S complexes interact with DOM thereby reducing their concentration in solution and the bioavailability of Hg to bacteria

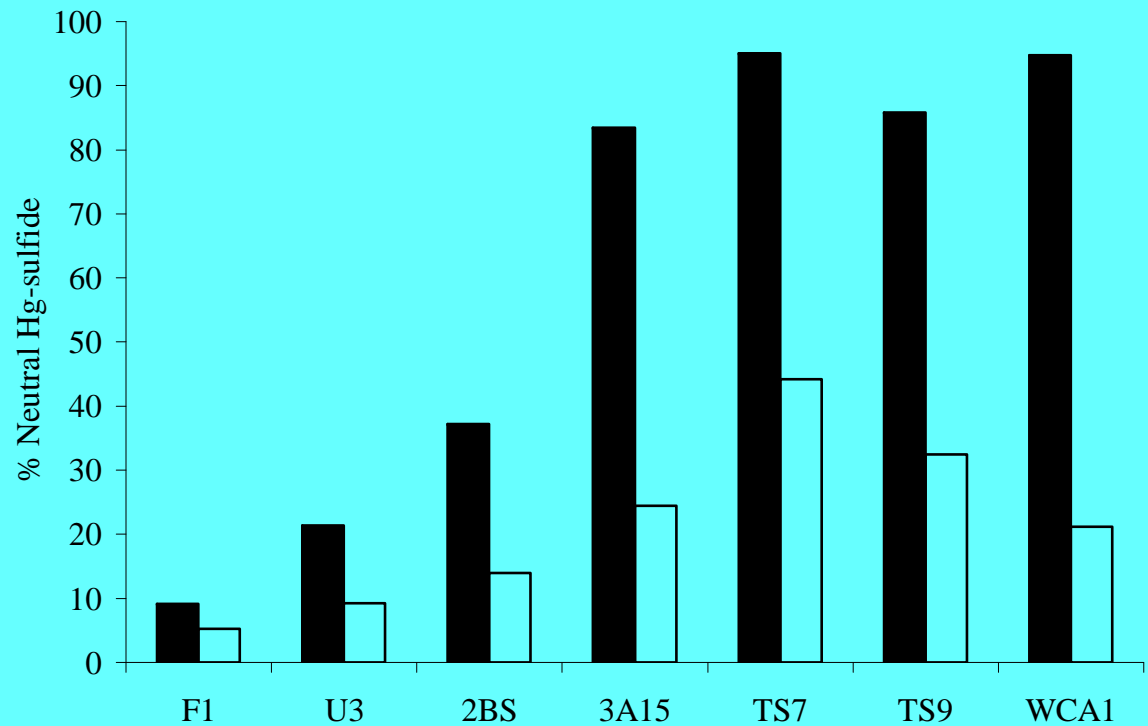
Stability constant for the interaction



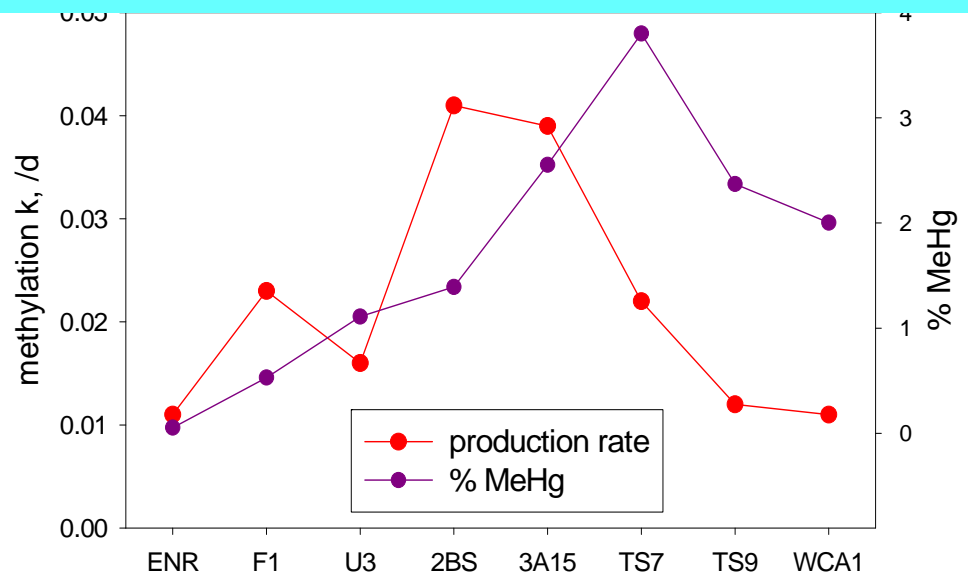
$$K = 0.16 \pm 0.04 \text{ L mg C}^{-1}$$

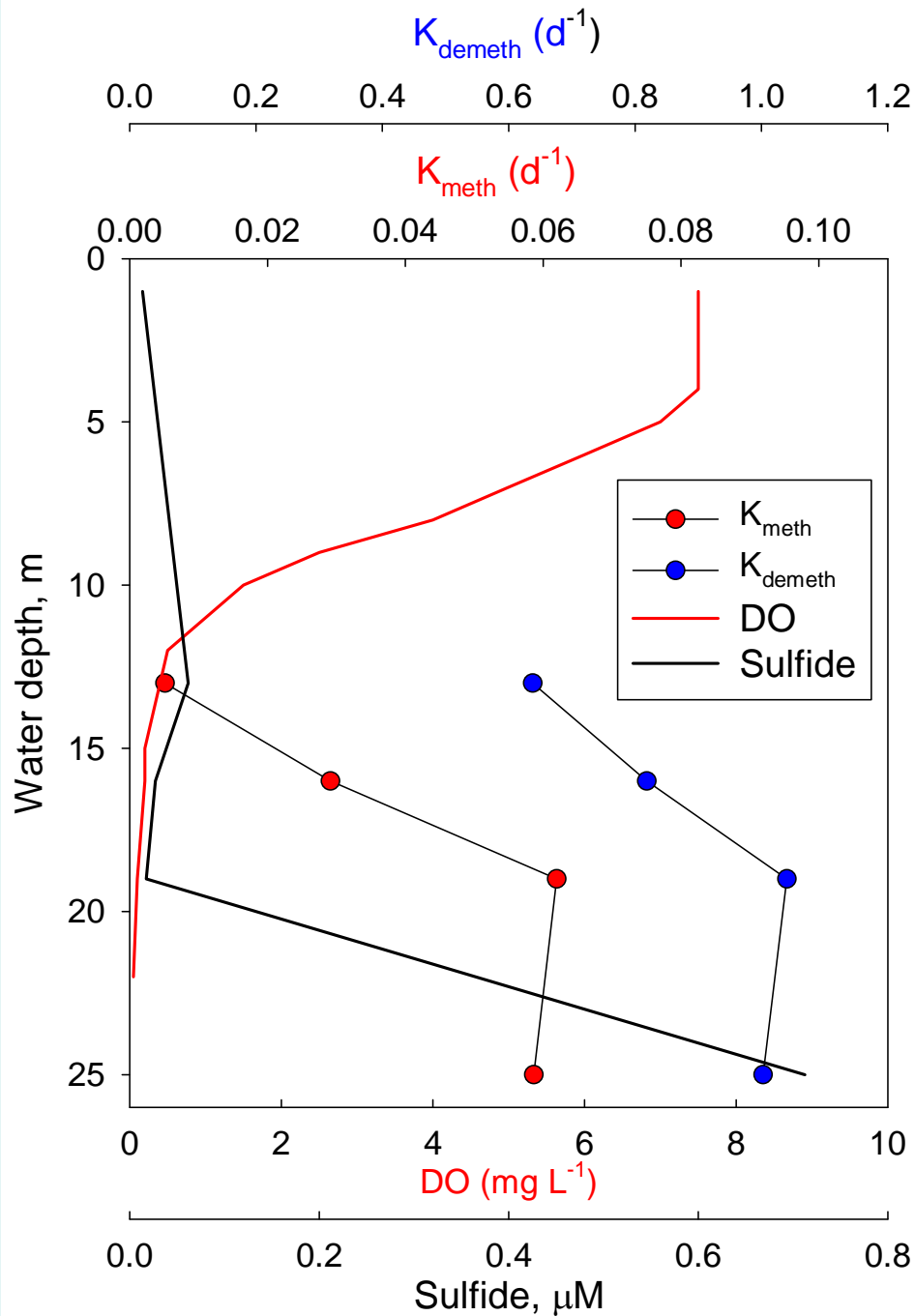


Calculated abundance of neutral Hg-sulfide complexes in surficial soil pore waters for seven sites across the Florida Everglades without (filled bars) and with (open bars) the interaction involving HgS-DOM included in the model. Overall while the magnitude changes, the overall trends are similar. The calculated concentrations with the interaction match the %MeHg more closely overall



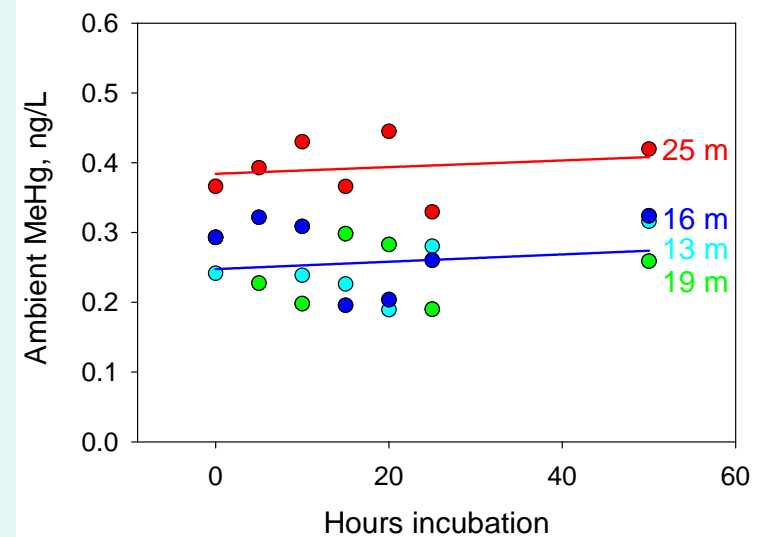
Miller et al., submitted





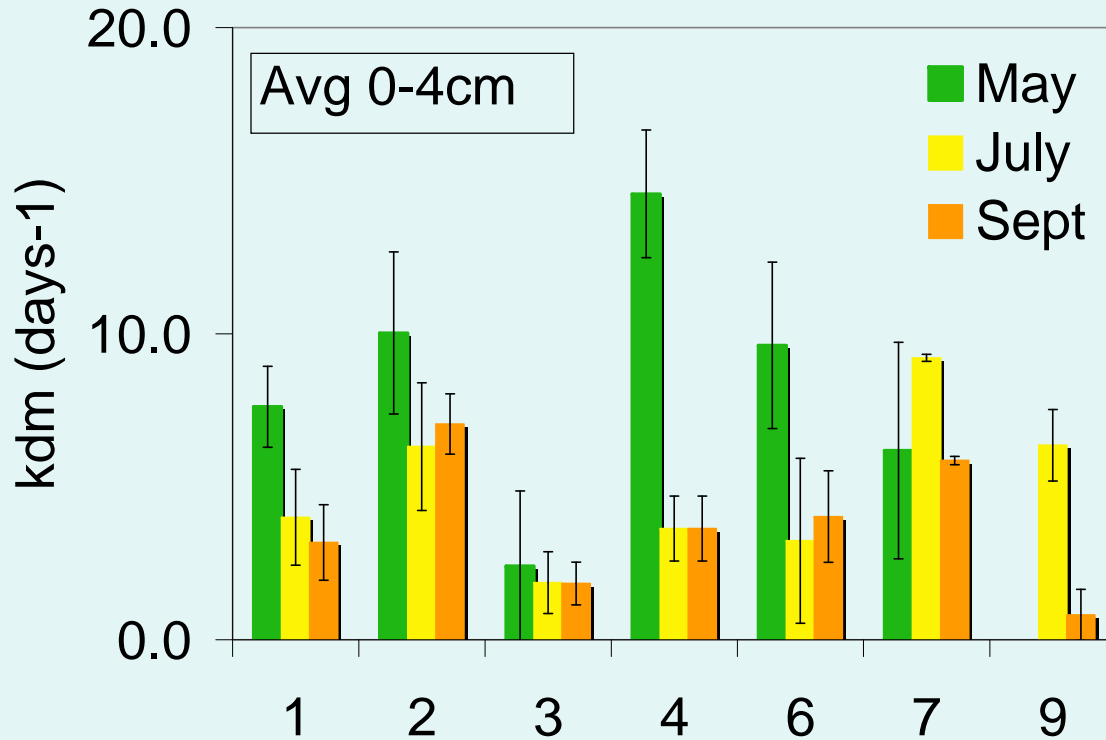
## Station 2 – the importance of water column methylation

- Sulfide levels in the bottles were low throughout the incubations (<0.2 μM) although there was a suggestion of higher levels overall at the final timepoint.**
- pH increased over time of the incubations from initial values of 7.3-7.4 to 7.8-8.0 at the end of the incubations**



And finally, we should not ignore the reverse reaction – demethylation

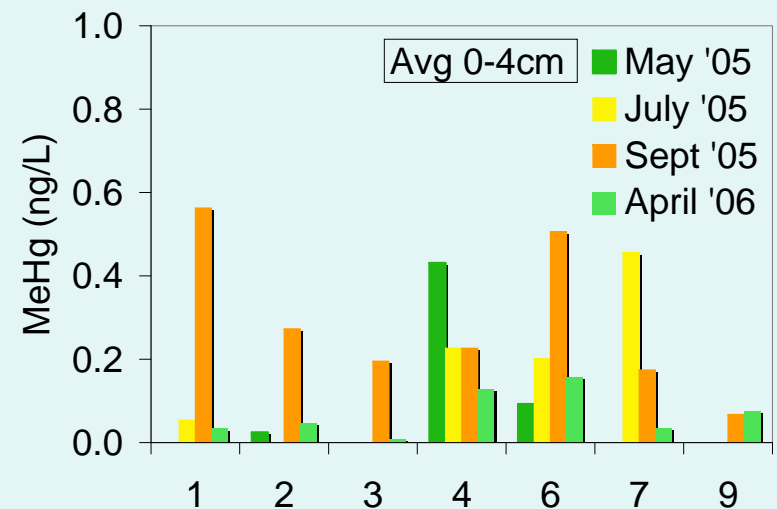
## Results - $k_{dm}$

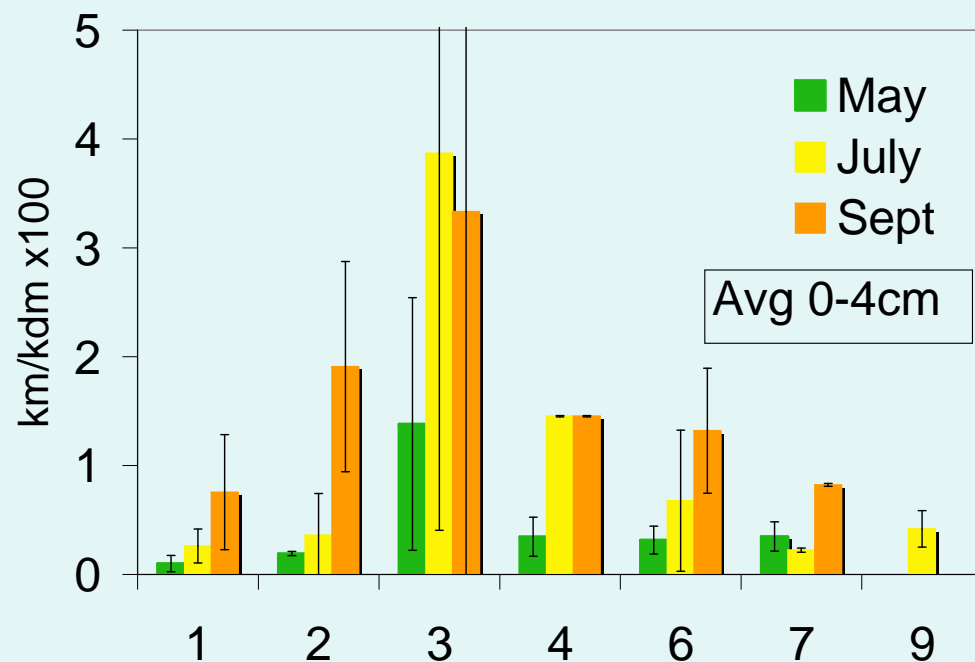


Rates are relatively variable across sites  
Demethylation rate constants DO NOT correlate with:  
%LOI, Porewater sulfide, Sediment MeHg, AVS  
Sulfate reduction rate

Questions:

1. What form of MeHg is taken up?
2. What microbes are responsible?
3. What is the relationship with depth?





The ratio of the rate constants for the Chesapeake Bay/shelf data is comparable to that of other estuarine and coastal systems, which are not highly contaminated. Also, as found elsewhere, the ratio of the rate constants is comparable to the range in %MeHg found in the sediments

Location Type and Method <sup>a</sup>	$k_1$ ( $\times 10^{-2}$ ) ( $d^{-1}$ )	$k_2$ ( $d^{-1}$ )	$k_1/k_2$ ( $\times 10^{-3}$ )	$[MeHg]_{tot}$ / $[Hg]_{tot}$ ( $\times 10^{-3}$ )	Refs
Hudson River (S)	0.2	15	0.1	2.0	Heyes et al., 2005
Bay of Fundy (S)	4.4	3.6	12	3.3	Sunderland et al., 2004
Mesocosm Studies (S)	3.1	16	2.3	2.8	Kim et al., 2005 <sup>b</sup>
San Pablo Bay (R)	1.4	0.3	56	18	Marvin- DiP et al., 03
Berry's Creek, NY (R)	1-18	0.06-0.12	10-150	4-8	Cardona-Marek, PhD
Delaware River, DE (S)	0.7-3	3-15	~2	~2	Cardona-Marek, PhD

a: Method – S = Stable isotopes used; R = Radioisotopes used.

b: average values for both R and NR systems in the top sediment layer (0-0.5 cm).



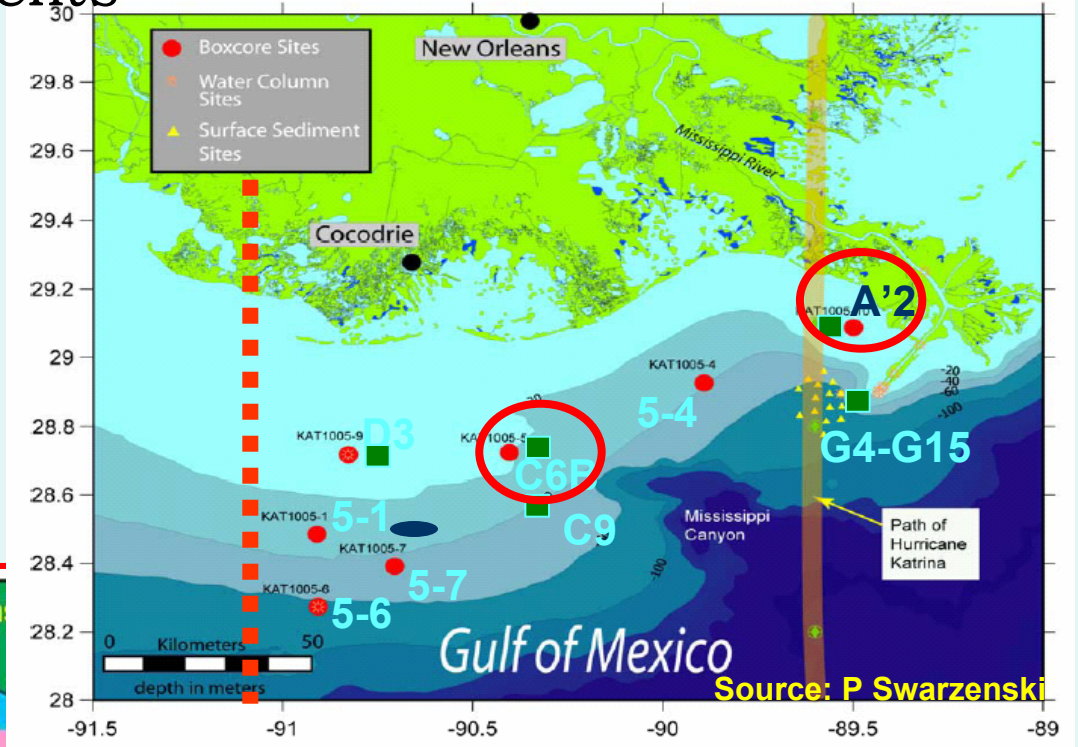
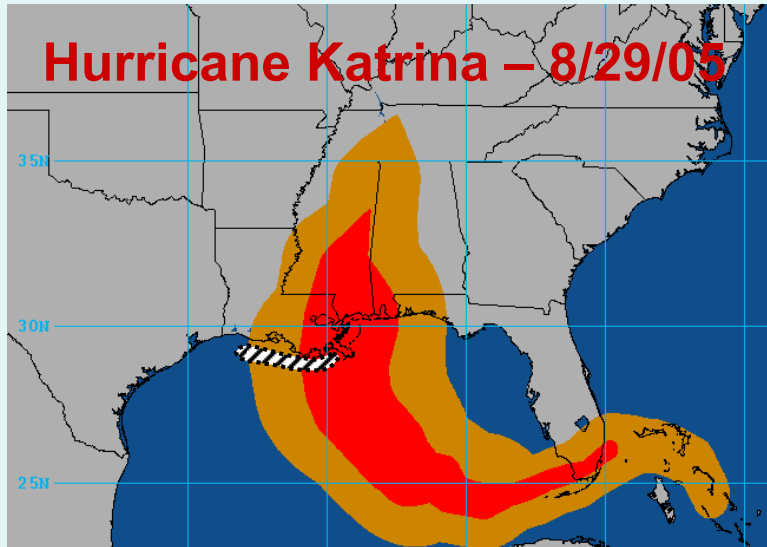
The focus has been on chemical factors,  
but what about?

- **Biological:** Differences in community structure and the ability to methylate Hg. Not all SRB's methylate. Some Fe reducers do. What about the impact of C supply? Other limitations (sulfate etc)
- **Physical:** Disturbance is important in “resetting” the system. Tidal resuspension can enhance methylation. Wetting/drying leads to sediment oxidation/enhance methylation
- What about **extreme events**?

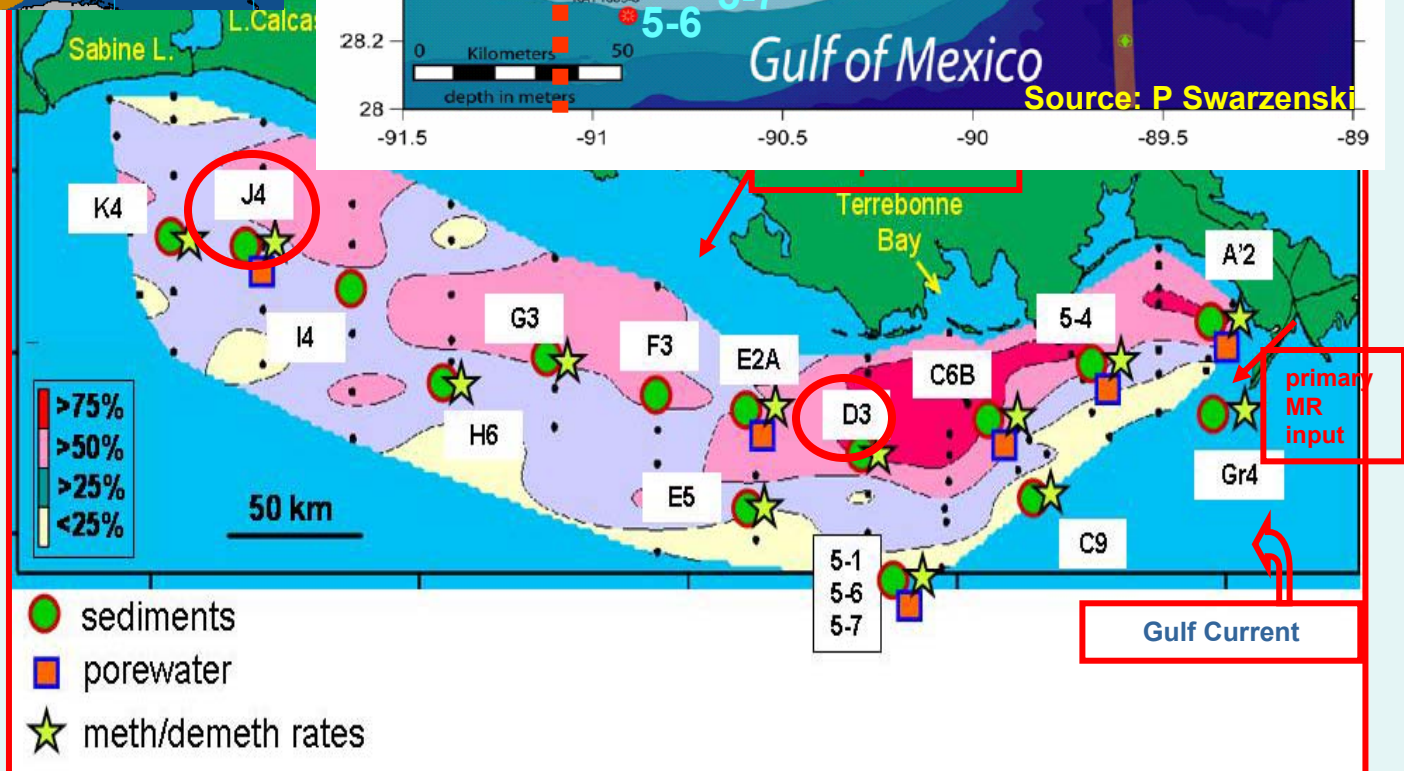


# The Impact of Extreme Events

, Jul & Oct 05, Mar & Jul 06



Stations A'2 was directly impacted by Katrina while J4 was less directly impacted by Katrina but was impacted by Rita; D3 was on the outskirts of both storms

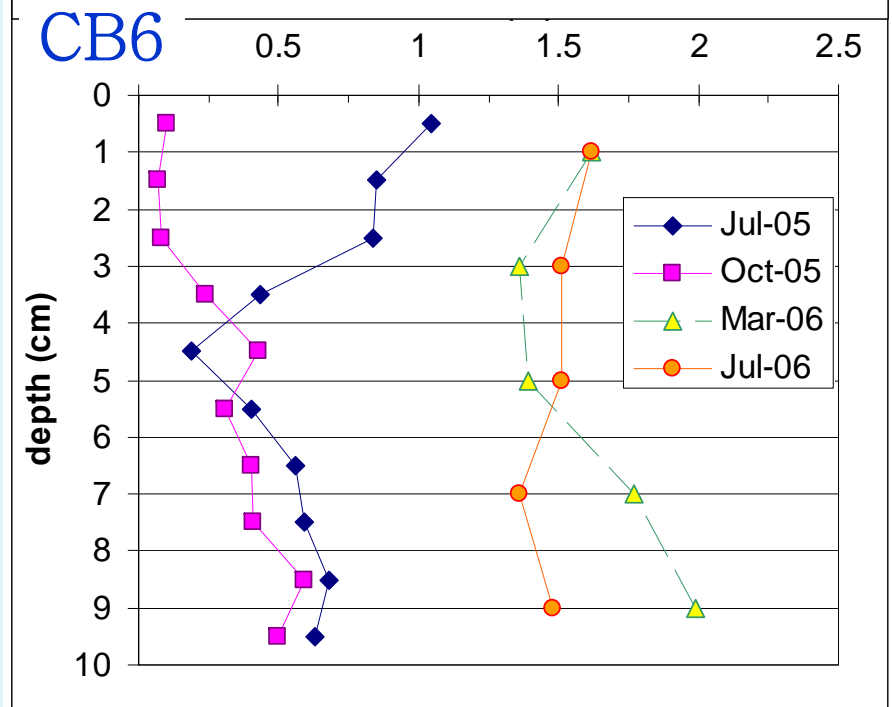
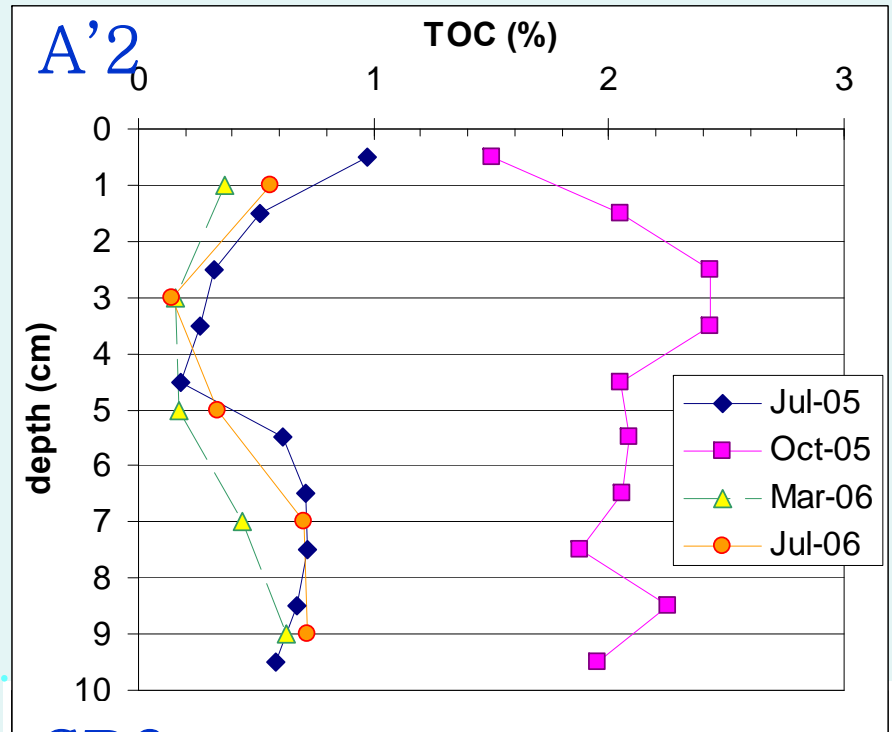
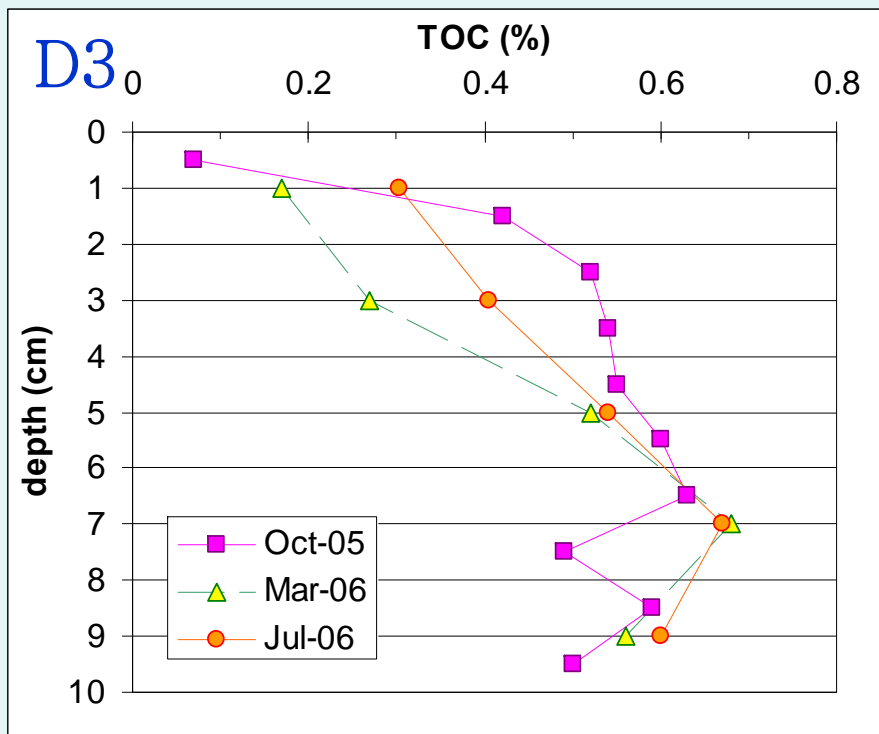


At station D3, there is little change in %TOC over the various sampling periods.

At A'2, there is a very different sediment %TOC in Oct 05 compared to the other times

At CB6, there appears to be a complex signal which shows a decrease in %TOC from Jul 05 to Oct 05 for the upper 4 cm but a very different %TOC after that at all depths

Liu et al., in rev.



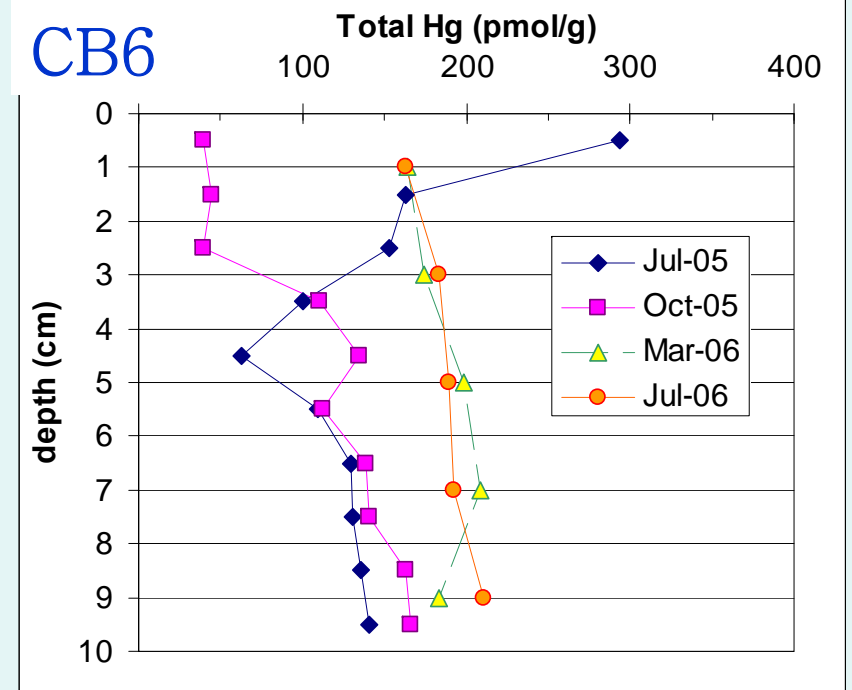
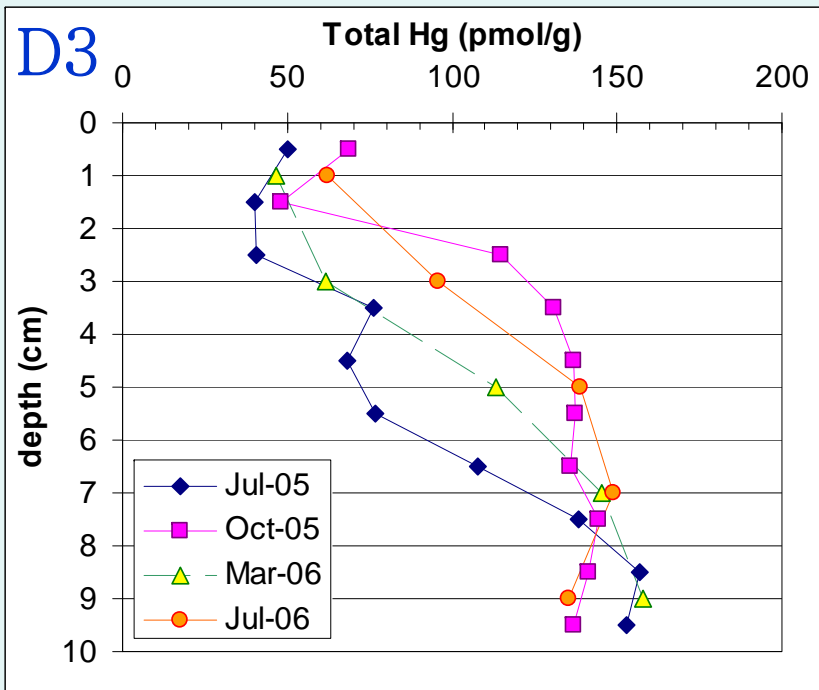
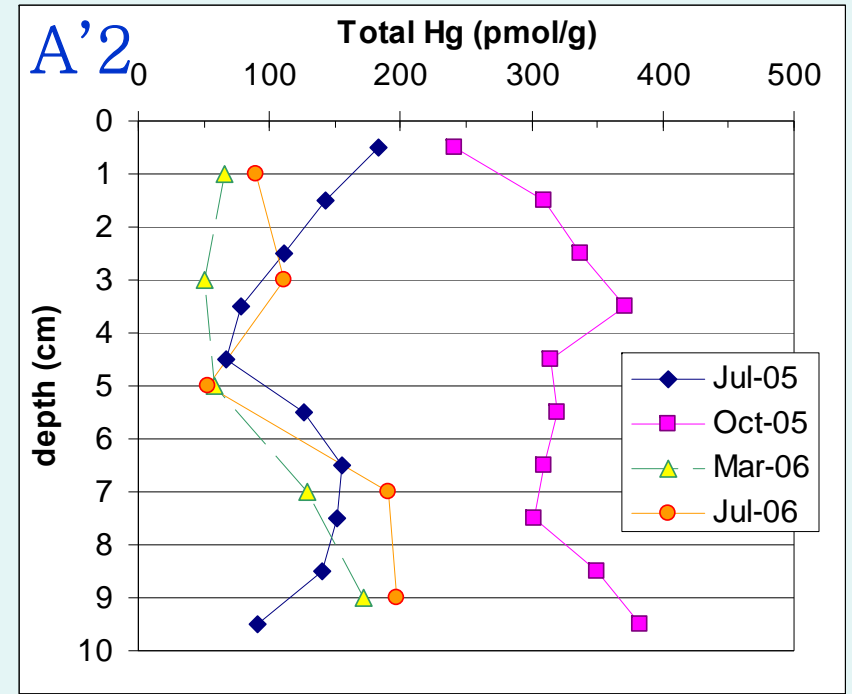
The total Hg concentrations mimic those of %TOC. It appears that both are reflecting similar changes in the sediment profile

For station D3, little change over time

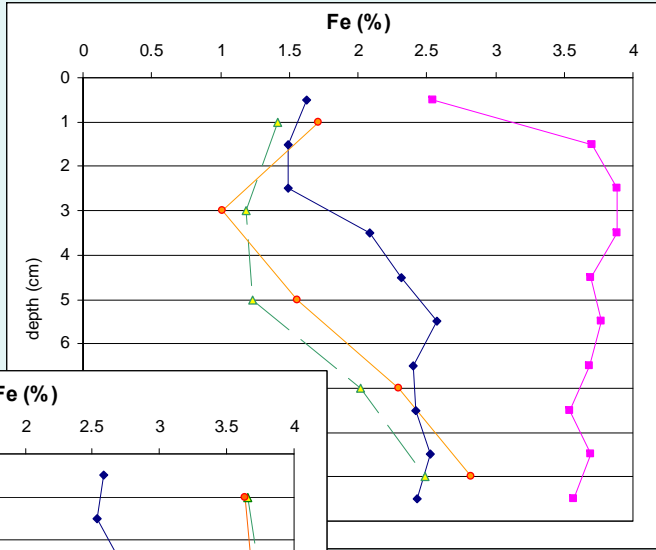
For A'2, much higher in Oct 05, rest similar

For CB6, decrease in upper sediment in Oct 05, then increase throughout

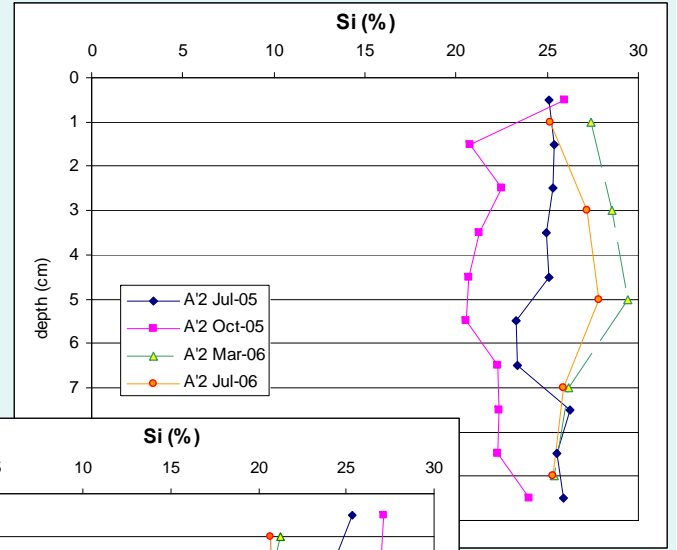
Liu et al., in rev.



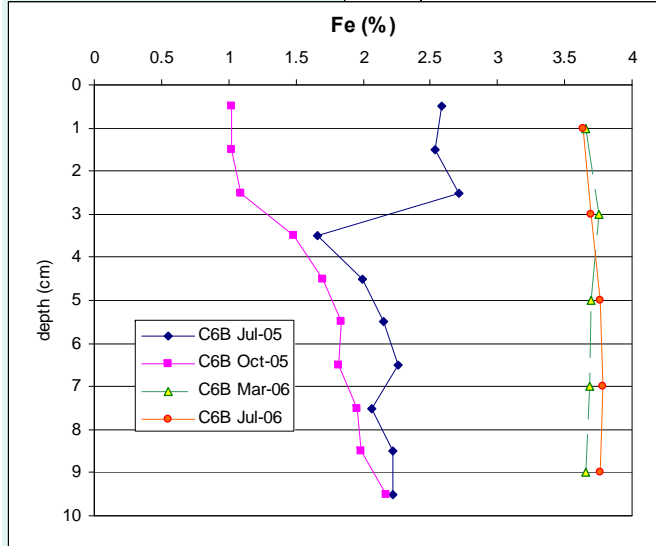
A'2 Fe



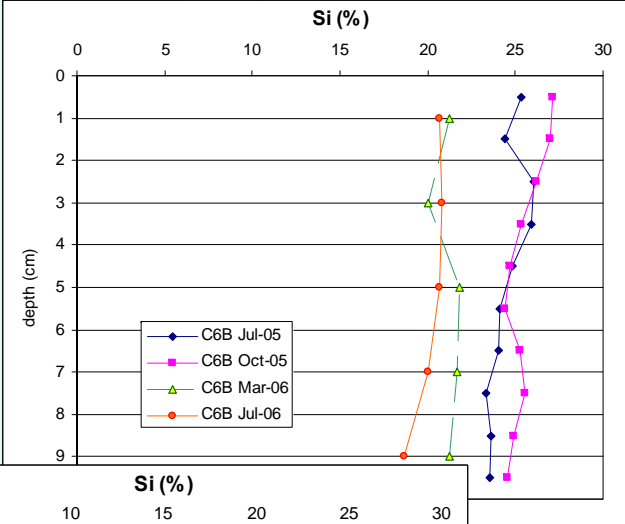
A'2 Si



CB6 Fe

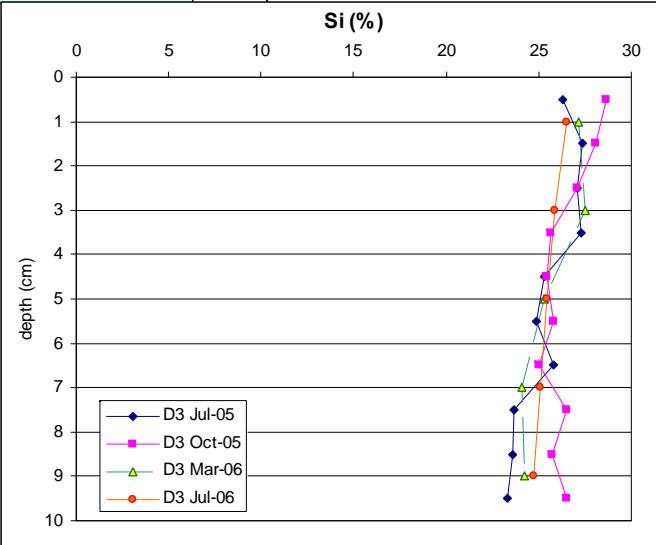
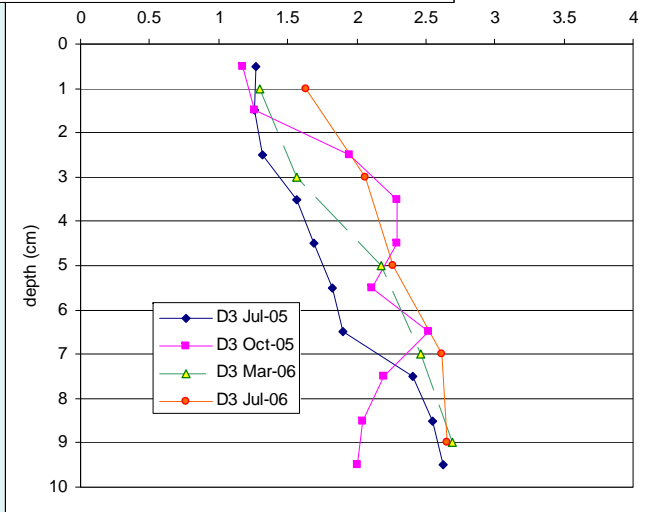


Overall, Fe trends similar to Hg/TOC. Si opposite



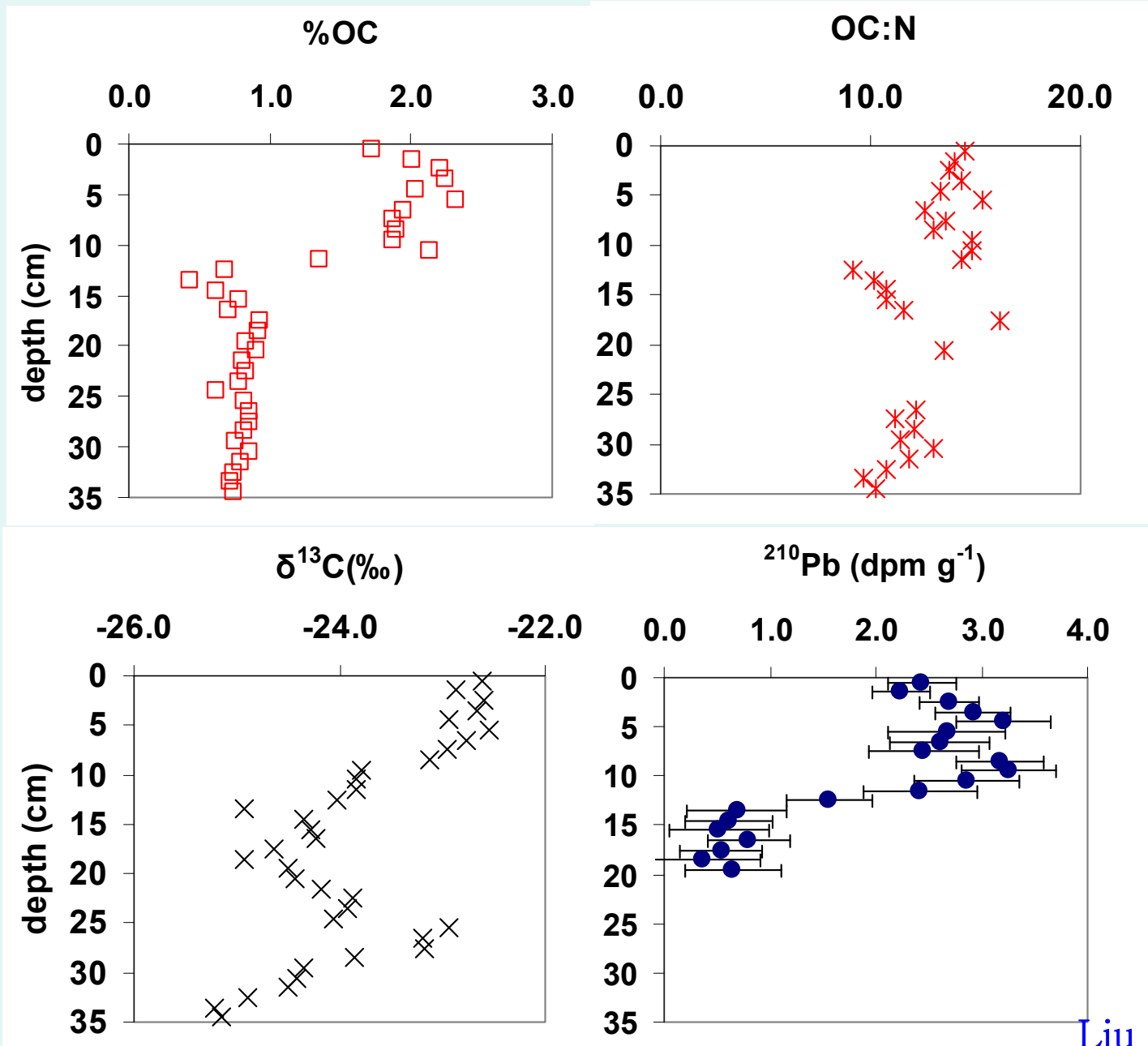
CB6 Si

D3 Fe



D3 Si

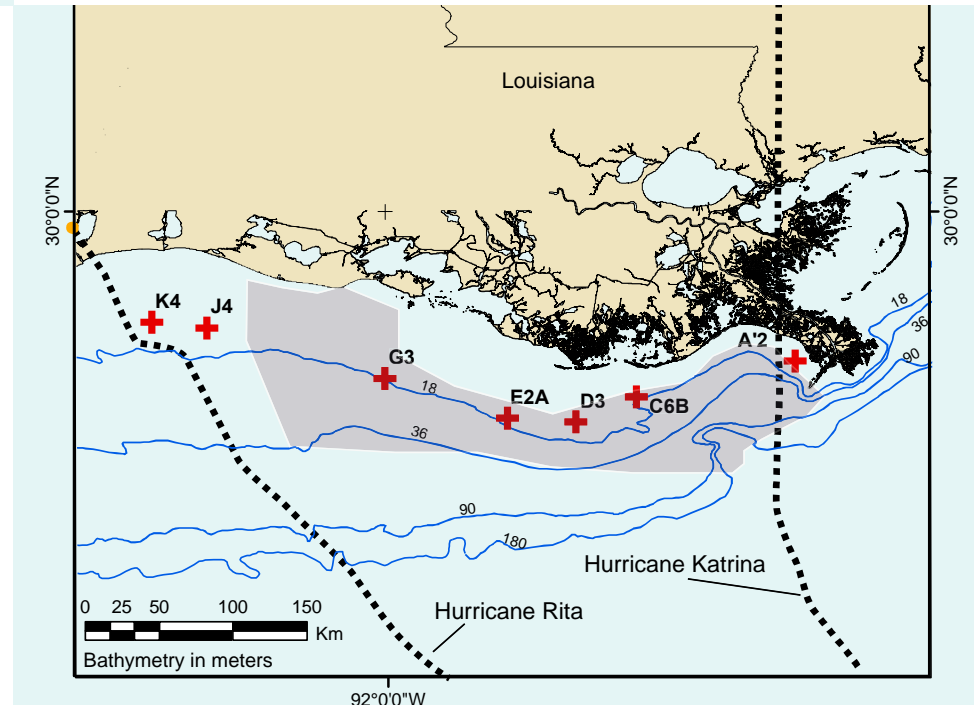
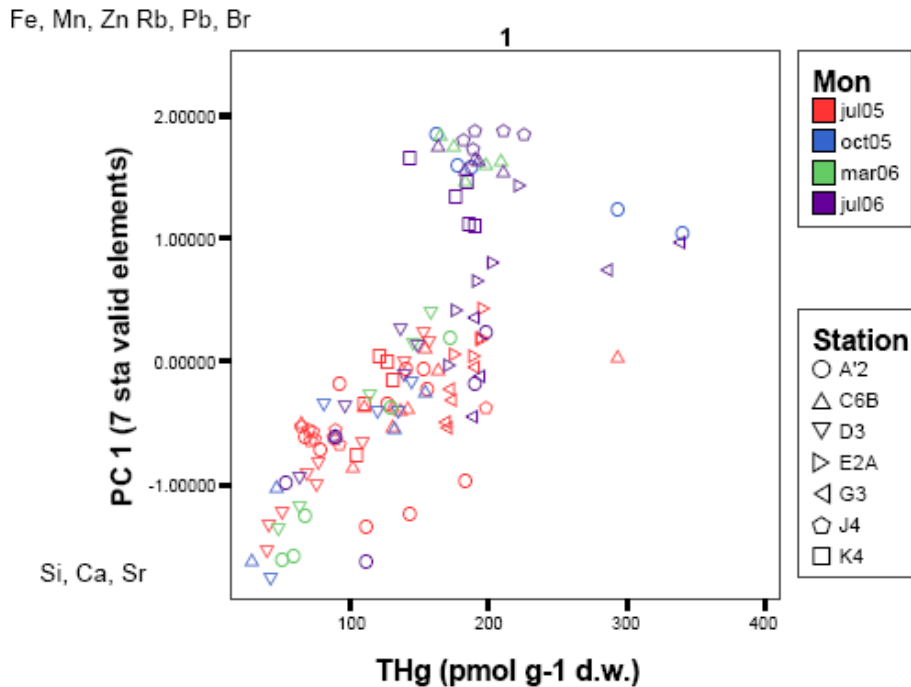
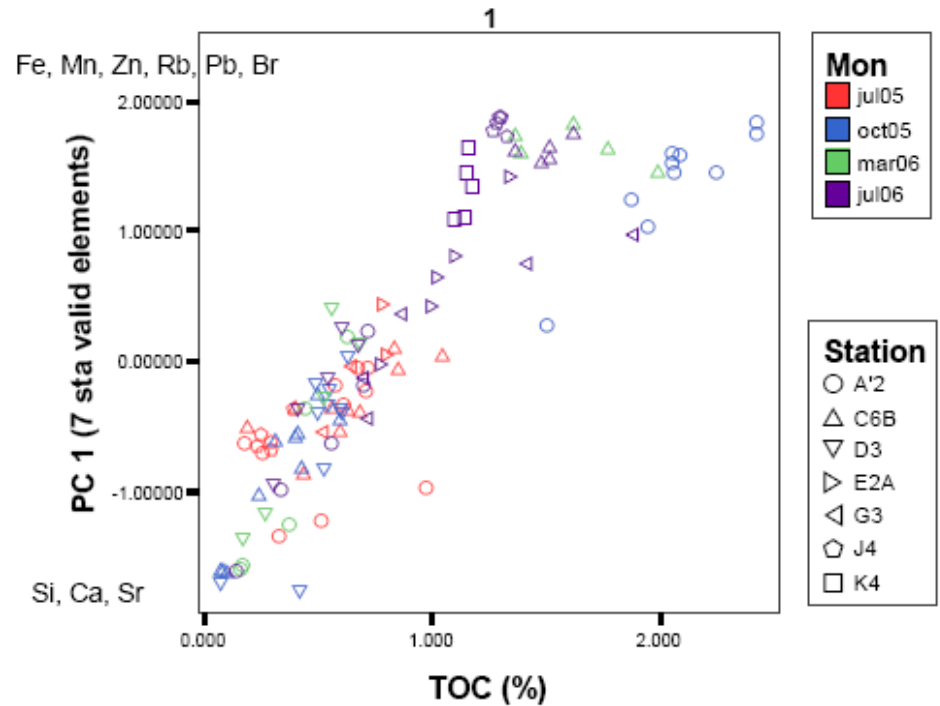
Other parameters also indicate a substantial disturbance at Station A'2



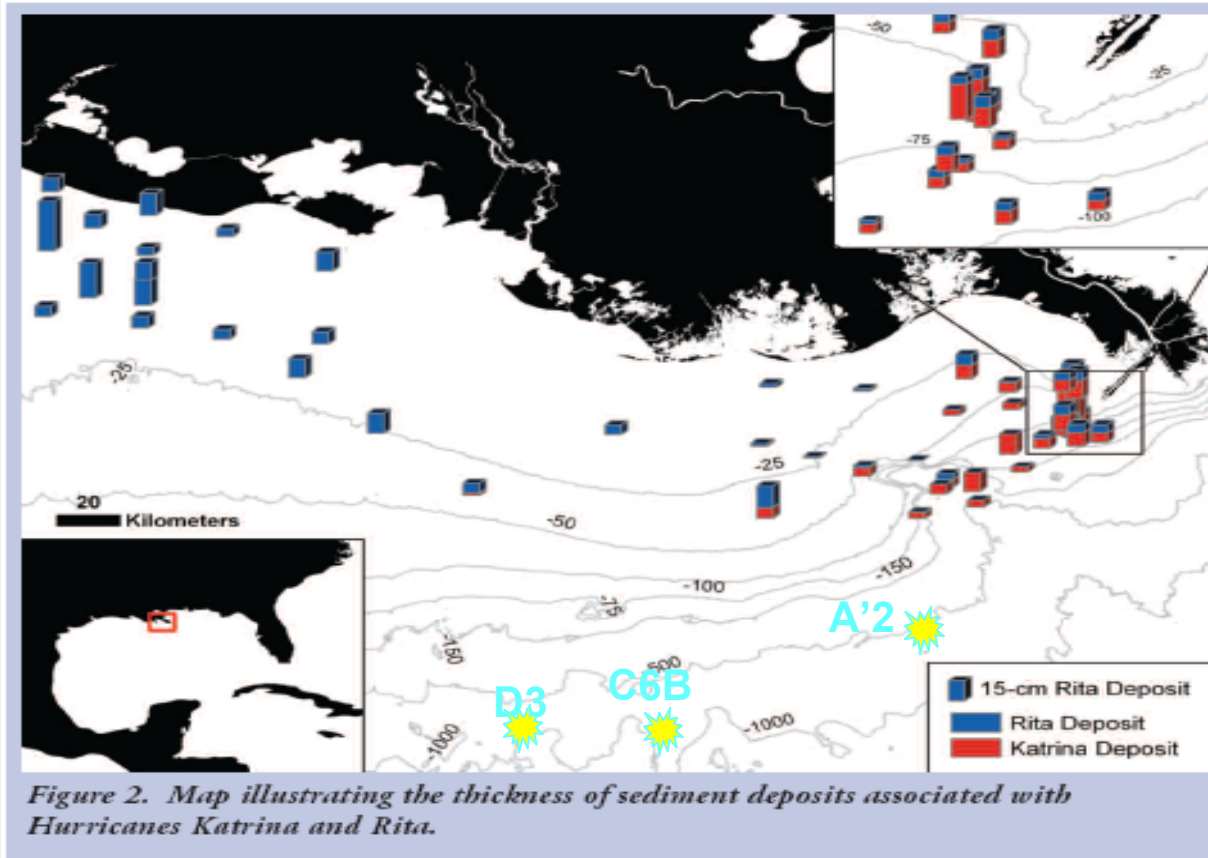
Results from PCA suggest that the sediments from A'2 & K4 in Oct 05, from C6B in Mar 06 and Jul 06 may differ from the rest.

This suggests there was sediment movement both during and in the months after the hurricanes

Liu et al., in rev.



Goni et al. (07) calculate that  $1.2 \times 10^{15}$  g of sediment and  $1.4 \times 10^{13}$  g org-C were redistributed during the hurricanes, as indicated on the map.



Goni report March 2007. *The Sedimentary Record* 5(1).

Based on these org-C estimates and the THg-TOC correlation, we estimate that ~50 tons of Hg were redistributed by the hurricanes. This is 3-5 times the annual river Hg input (10 ton/year) and atmosphere (2.5 ton/year) combined.

Liu et al., in rev.



# Finally, to answer the initial question, and returning to the Chesapeake/shelf.....

We find....

- High  $Hg_T$  and MeHg in solid phase of organic-rich bay and slope sediment, low in sandy shelf
- Comparable  $Hg_T$  and MeHg in porewater of all sediment
- %MeHg relatively high compared to other systems
- Seasonal trends consistent with biotic production
- Environmental factors that control Hg bioavailability elucidated - Particulate organic matter impacts  $K_D$ ; Sulfide effects  $K_D$  and  $Hg_{pw}$  and both impact bioavailability

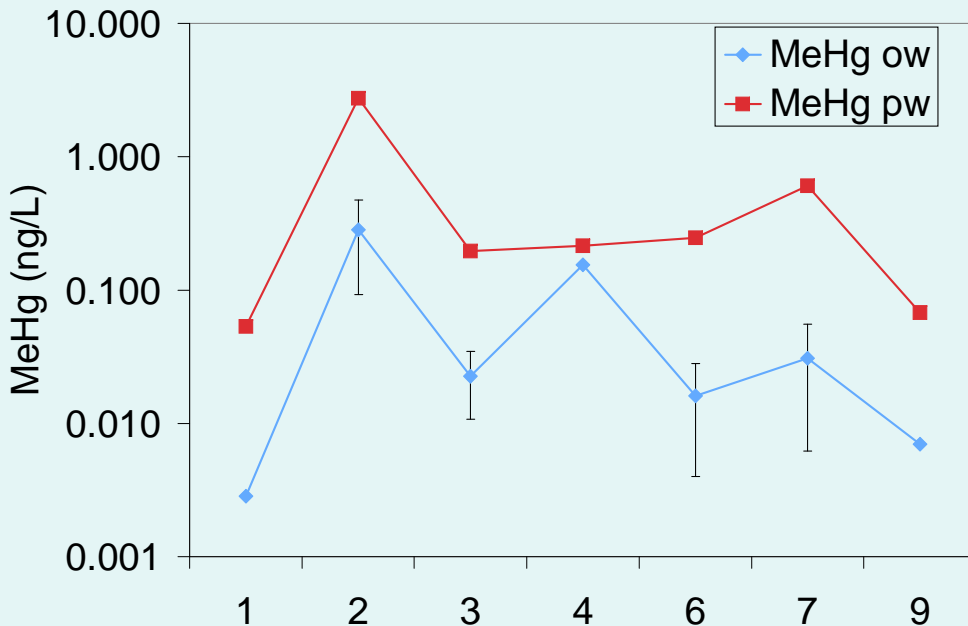
## 1. Is methylation in coastal zone important?

- Net MeHg production substantial at all sites
- But is the flux out of the sediments important?

# Results - Sediment-water MeHg Diffusive Flux

Estimated Diffusive Flux, F (Gill et al., 1999)

- $D_w \text{ MeHg-MOM} = 2 \times 10^{-6} \text{ cm}^2 \text{ s}^{-1}$ ;  $D_w \text{ MeHgSH} = 1.3 \times 10^{-5} \text{ cm}^2 \text{ s}^{-1}$



Studies in other systems suggest that the actual flux is higher than that estimated assuming simple diffusion and also MeHg flux appears to be enhanced under low oxygen conditions

Results suggest that sediments could be an important source of MeHg to coastal waters. Also, shelf and slope sediments have abundant macrofauna and therefore there is substantial potential for bioaccumulation through the benthos

Assuming MeHgSH is species diffusing:

Site	Depth (m)	F (pg m <sup>-2</sup> hr <sup>-1</sup> )	Time (yr)
Bay	10	52.3 ± 40.4	0.217 ± 0.11
Shelf	16	163.6 ± 91.9	0.339 ± 0.16
Slope	620	28.3	17.5