

Diesel Exhaust Particulates*

Reasonably anticipated to be a human carcinogen

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Carcinogenicity

Exposure to diesel exhaust particulates is *reasonably anticipated to be a human carcinogen*, based on limited evidence of carcinogenicity from studies in humans, which indicates elevated lung cancer rates in occupational groups exposed to diesel exhaust (IARC 1989, Cohen and Higgins 1995, Bhatia *et al.* 1998) and supporting animal and mechanistic studies. An increased risk of lung cancer is found in the majority of human studies. The overall relative risk is approximately 1.3, and higher risks are found in more heavily exposed subgroups in some studies. The increased risk is not readily explained by confounding by either smoking or asbestos exposure. However, the increased risk cannot always be clearly ascribed to diesel exhaust exposure. Although some studies employed semiquantitative estimates of diesel exhaust exposure (Steenland *et al.* 1998), most studies used inadequate measures of exposure.

Studies of the carcinogenicity of diesel exhaust particulates in animals have shown a consistent lung tumor response in rats but not in the mouse or hamster. The response in rats appears to be due to the particulate component of exhaust, as the filtered vapor phase of exhaust has been shown not to be tumorigenic. Solvent extracts of diesel exhaust particles are carcinogenic when applied to the skin or administered by intratracheal instillation or intrapulmonary implantation to rats, mice, or hamsters.

Additional Information Relevant to Carcinogenicity

Diesel exhaust is a complex mixture of combustion products of diesel fuel, with the exact composition depending on the type of engine, the speed and load at which it is run, and the composition of the fuel used. Diesel exhaust contains identified mutagens and carcinogens both in the vapor phase and associated with respirable particles. Diesel exhaust particles are considered likely to account for the human lung cancer findings because they are almost all of a size small enough to penetrate to the alveolar region. Mutagenic and carcinogenic chemicals, including polyaromatic hydrocarbons and nitroarenes, have been extracted from these particles with organic solvents, or with a lipid component of mammalian lung surfactant.

While diesel particulate exposures produce lung cancer in rats, the relevance of this result in predicting the human response has been questioned. Diesel exhaust particulate exposure produces a spectrum of inflammatory and neoplastic pulmonary responses in the rat that are characteristic of responses also seen with other inhaled particles of varying toxicity. These responses are apparently little influenced by the chemical constituents of the particles. Although the precise bioavailability of chemical mutagens and carcinogens from inhaled diesel particulates is not known, DNA adducts have been found in the lung of rats exposed to diesel exhaust particulates. Similarly, more lymphocyte-DNA adducts were found in groups occupationally exposed to diesel exhaust than in groups with lower or ambient exposures. However, diesel exhaust exposure was not quantified in these studies, and exposure to used motor oil may have contributed to the adducts observed in one study.

Properties

Diesel exhaust is a complex mixture of combustion products of diesel fuel, with the exact composition dependent upon the type of engine, operating conditions, lubricating oil, additives, emission control system, and the composition of the fuel used (Ullman 1989, Obert 1973). Diesel engines are typically separated according to their service requirements, light-duty or heavy-duty. The operating conditions for light-duty and heavy-duty diesel engines differ in terms of engine speed,

expected load, fuel composition, and engine emission controls. Typically light-duty vehicles such as automobiles and light trucks operate at higher speeds than heavy-duty vehicles such as trucks. The total particulate emission concentration from light-duty diesel engines is much smaller than from heavy-duty diesel engines. In general, newer heavy-duty trucks emit diesel particulates at a rate 20 times greater than catalyst-equipped gasoline-fueled vehicles (WHO 1996). However, depending on operating conditions, fuel composition, and engine control technology, light-duty diesel engines and heavy-duty diesel engines can emit 50 to 80 times and 100 to 200 times, respectively, more particulate mass than typical catalytically equipped gasoline engines (McClellan 1986).

The particle size distribution of diesel exhaust is bi-modal with a nuclei mode (0.0075 to 0.042 μm in diameter) and an accumulation mode (0.042 to 1.0 μm in diameter) (Baumgard and Johnson 1996), most of which occur in aerodynamic diameters ranging from 0.1 to 0.25 μm in diameter (Groblicki and Begeman 1979, Dolan *et al.* 1980, NRC 1982, Williams 1982). Approximately 98% of the particles emitted from diesel engines are less than 10 microns in diameter, 94% less than 2.5 microns in diameter, and 92% less than 1.0 microns in diameter (ARB 1997).

Engines running under low load typically produce fewer particles with a higher proportion of organic compounds associated with the available particle mass. Conversely, engines under high load typically produce more particulate matter with a lower proportion of organic compounds associated with available particles. Kishi *et al.* (1992) found that exhaust gas temperatures affect particle composition. Low exhaust gas temperatures produce particulate matter with more adsorbed soluble organics than particulate matter produced in a high exhaust gas temperature environment.

The emissions consist of a nonpolar fraction (57%), a moderately polar fraction (9%), and a polar fraction (32%) (Schuetzle 1983, Schuetzle *et al.* 1985), with the remainder unrecoverable. Diesel engines operate with excess air (~25-30 parts air to 1 part fuel) (Lassiter and Milby 1978). The gas phase fraction is composed primarily of typical combustion gases such as nitrogen, oxygen, carbon dioxide, and water vapor. As a result of incomplete combustion, the gaseous fraction also contains pollutants such as carbon monoxide, sulfur oxides, nitrogen oxides, volatile hydrocarbons, and low molecular weight polyaromatic hydrocarbons (PAHs) and their derivatives.

The inorganic fraction of the particulate phase of diesel fuel combustion emissions primarily consists of small elemental carbon particles ranging from 0.01 to 0.08 μm in diameter. The organic and elemental carbon accounts for approximately 80% of the total particulate matter mass. The remaining 20% is composed of sulfate (mainly sulfuric acid) (Pierson and Brachaczek 1983) and some inorganic additives and components of fuel and motor oil.

In general, the organic compounds identified in diesel exhaust emissions contain hydrocarbons, hydrocarbon derivatives, PAHs, PAH derivatives, multifunctional derivatives of PAHs, heterocyclic compounds, heterocyclic derivatives, and multifunctional derivatives of heterocyclic compounds (Schuetzle 1988). The organic fractions consist of soluble organic compounds such as aldehydes, alkanes, alkenes, and high molecular weight PAHs and PAH-derivatives.

Because of their high surface area, diesel exhaust particulates are capable of adsorbing relatively large amounts of organic material. The adsorbed elements come from unburned fuel, lubricating oil, and pyrosynthesis during fuel combustion. A variety of mutagens and carcinogens such as PAH and nitro-PAH (NRC 1982, Tokiwa and Ohnishi 1986, WHO 1996) are adsorbed by the particulates. There is sufficient evidence for the carcinogenicity for 15 PAHs (a number of these PAHs are found in diesel exhaust particulate emissions) in experimental animals. The nitroarenes (five listed) meet the established criteria for listing as *reasonably anticipated to be a human*

carcinogen based on carcinogenicity experiments with laboratory animals. The organic extractable fraction of diesel exhaust particulates is typically in the 20% to 30% range but it may be as high as 90% (Williams *et al.* 1989), depending upon vehicle type and operating conditions. In general, incomplete combustion in diesel engines operating under low load conditions produces relatively low particle concentrations and a higher proportion of organic-associated particles (Dutcher *et al.* 1984).

Use

No known uses of diesel exhaust particulates exist.

Production

Internal combustion engines have been used in cars, trucks, locomotives, and other motorized machinery for approximately 100 years (IARC 1989). The combustion of diesel fuel in a compression ignition engine produces diesel exhaust. Engine exhaust contains thousands of gaseous and particulate substances. There are three major groups of diesel exhaust sources: mobile sources (on-road vehicles and other mobile sources), stationary area sources (oil and gas production facilities, stationary engines, repair yards, shipyards etc.), and stationary point sources (chemical manufacturing, electric utilities, etc.). The composition and quantity of the emissions from an engine depend mainly on the type and condition of the engine, fuel composition and additives, operating conditions, and emission control devices.

Exposure

Various employee groups have been studied to determine their occupational exposures to diesel exhaust particulates. They include railroad workers, mine workers (who use diesel-powered equipment), bus garage workers, trucking company workers, fork-lift truck operators, firefighters, lumberjacks, toll-booth and parking garage attendants, and many professions servicing or handling automobiles (car mechanics, professional drivers, *etc.*). The National Institute for Occupational Safety and Health (NIOSH) has estimated that approximately 1.35 million workers are occupationally exposed to diesel exhaust particulates in approximately 80,000 workplaces in the United States (MMWR 1989).

Railroad workers' potential for exposure has increased since 1959, when almost all the U.S. railroad system (95%) was converted to diesel engines. Varying degrees of exposure to diesel exhaust particulates (from 17 $\mu\text{g}/\text{m}^3$ for clerks to 134 $\mu\text{g}/\text{m}^3$ for locomotive shop workers) has been reported based upon job groups (Woskie *et al.* 1988).

Diesel engines have been, and continue to be, commonly used in U.S. mines since their first introduction in the early 1950s. Exposure occurs from activities that use diesel-fueled heavy machinery, such as blasting. Holland (1978) conducted an exposure study in 24 U.S. coal mines and showed that, while certain PAHs expected from diesel exhaust particulates were found (anthracene and phenanthrene), other PAHs (benz[*a*]anthracene, benz[*a*]pyrene, benz[*e*]pyrene, chrysene, and pyrene) were not found in measurable quantities. Cornwell (1982) found other PAHs not found in Holland's study in diesel emissions from a molybdenum mine in Colorado. This mine was equipped with drills and various load haul-dumps.

Bus repair facilities were studied to determine diesel exhaust emissions (Apol 1983). The highest exposure levels were observed when buses were started at peak times of dispatch and return. Pryor (1983) studied area levels of diesel exhaust particulates at a bus garage. Elevated levels of diesel exhaust emissions were observed during peak hours of bus activity. Levels rapidly returned to normal 10 to 15 minutes after the passing of peak times. Using elemental carbon, Zaebst *et al.* (1991) found an effective way of determining truck drivers' exposure to diesel exhaust particulates. Temperature was an important factor with higher exposures occurring at higher

temperatures. This study showed no discernible difference between the exposure levels of truckers (3.8 $\mu\text{g}/\text{m}^3$) and highway background concentrations (2.5 $\mu\text{g}/\text{m}^3$). These results indicate that the highway environment, rather than the truck itself, is the cause of the truck driver's exposure (Zaebst *et al.* 1991).

Firefighters in New York, Boston, and Los Angeles were studied to determine exposure to diesel exhaust particulates (Froines *et al.* 1987). Total exposure to airborne particles was measured using personal air samplers. Approximately 86% of the study population were nonsmokers. Smokers were exposed to higher average airborne particle concentrations than were nonsmokers. Sampling was performed only when firefighters were in the fire stations. For the three cities, total airborne particulate exposure had a time-weighted average (TWA) ranging from below 100 $\mu\text{g}/\text{m}^3$ to 480 $\mu\text{g}/\text{m}^3$. An average particulate exposure of 300 $\mu\text{g}/\text{m}^3$ was measured in Boston and New York. After adjustment for background levels and smoking (75 $\mu\text{g}/\text{m}^3$), firefighters in these two cities were exposed to a total diesel exhaust particulates level of 225 $\mu\text{g}/\text{m}^3$. Los Angeles had the highest exposure levels of the three cities. For a "worst-case" scenario, the mean concentration levels were as high as 748 $\mu\text{g}/\text{m}^3$. The authors noted that these were busy fire stations located in large metropolitan areas. Other factors such as smoking, building design, age and maintenance of vehicles, activities of the firefighters, and timing of runs also affected results.

Three studies reviewed by the IARC (1989) found that toll-booth workers had elevated levels of exposure to carbon monoxide (although this was decreased with ventilation systems) and diesel exhaust particulates. carbon monoxide exposure also was elevated among border-station and motor vehicle inspectors and parking garage attendants (IARC 1989). In many of these studies, however, it was difficult to differentiate between gasoline exhaust and diesel exhaust. Numerous studies have combined the two types of exhausts together. Therefore, exact determinations of diesel exhaust particulates exposure is difficult in these cases.

Regulations

EPA

Clean Air Act

Control of Emissions of Hazardous Air Pollutants from Mobile Sources: Listed as a Mobile Source Air Toxic for which regulations are to be developed

Guidelines

NIOSH

Diesel exhaust listed as a potential occupational carcinogen

*No separate CAS registry number assigned to diesel exhaust particulates.

REFERENCES

- Apol, A. G. 1983. Health Hazard Evaluation, Regional Transportation District, Denver CO. HETA-82-137-1264. Cincinnati, OH: National Institute for Occupational Safety and Health.
- ARB. 1997. Emission Inventory, 1995. Air Resource Board, Technical Support Division.
- Baumgard, K. J. and J. H. Johnson. 1996. The Effect of Fuel and Engine Design on Diesel Exhaust Particle Size Distributions. SAE Technical Paper Series, #960131.
- Bhatia, R., P. Lopipero and A. H. Smith. 1998. Diesel exhaust exposure and lung cancer. *Epidemiology* 9(1): 84-91.
- Cohen, A. J. and M. W. P. Higgins. 1995. Diesel Exhaust. A Critical Analysis of Emissions, Exposure, and Health Effects. Health Effects Institute.
- Cornwell, R. J. 1982. Health Hazard Evaluation Determination: Climax Molybdenum Company, Climax, CO. Report No. MHETA-81-108-9004, PB84-14850 1. Morgantown, WV.
- Dolan, D. F., D. B. Kittelson and D. Y. H. Pui. 1980. Diesel Exhaust Particle Size Distribution Measurement Techniques. SAE Technical paper # 870254.
- Dutcher, J. S., J. D. Sun, J. A. Lopez, I. Wolf, R. K. Wolff and R. O. McClellan. 1984. Generation and characterization of radiolabeled diesel exhaust. *Am Ind Hyg Assoc J* 45(7): 491-8.
- Froines, J. R., W. C. Hinds, R. M. Duffy, E. J. Lafuente and W. C. Liu. 1987. Exposure of firefighters to diesel emissions in fire stations. *Am Ind Hyg Assoc J* 48(3): 202-7.
- Groblicki, P. J. and C. R. Begeman. 1979. Particle size variation in diesel car exhaust. SAE Technical paper #790421. Society of Automotive Engineers.
- Holland, W. D. 1978. Determination of Breathing Zone Concentration of Contaminants from Emission from Diesel Powered Vehicles in Underground Mines. BuMines OFR 24-80; U.S. NTIS PB80-150776. Washington, D.C.: U.S. Department of the Interior, Bureau of Mines.
- IARC. 1989. Diesel and Gasoline Engine Exhausts and Some Nitroarenes. IARC Monographs on the Evaluation of Carcinogenic Risk of Chemicals to Humans, vol. 46. Lyon, France: International Agency for Research on Cancer. 458 pp.
- Kishi, Y., H. Tohno and M. Ara. 1992. Characteristics and Combustibility of Particulate Matter. Reducing Emissions from Diesel Combustion. Warrendale, PA: Society of Automotive Engineers. p. 139-146.

- Lassiter, D. V. and T. H. Milby. 1978. Health Effects of Diesel Exhaust Emission: A Comprehensive Literature Review, Evaluation and Research Gaps Analysis. U.S. NTIS PB-282-795. Washington, D.C.: American Mining Congress.
- McClellan, R. O. 1986. 1985 Stokinger lecture. Health effects of diesel exhaust: a case study in risk assessment. *Am Ind Hyg Assoc J* 47(1): 1-13.
- MMWR. 1989. Publication of NIOSH Current Intelligence Bulletin on Carcinogenic Effects of Diesel Exhaust. *Morbidity and Mortality Weekly Report* 38: 76-78.
- NCR. 1982. Diesel cars: benefits, risks, and public policy, final report of the Diesel Impacts Study Committee. Washington, D.C.: National Academy Press.
- Obert, E. F. 1973. *Internal Combustion Engines and Air Pollution*. Harper and Row.
- Pierson, W. R. and W. W. Brachaczek. 1983. Particulate matter associated with vehicles on the road II. *Aerosol. Sci Tech* 2: 1-40.
- Pryor, P. 1983. Trailways Bus System, Denver CO. Health Hazard Evaluation Report No. HETA 81-416-1334. Cincinnati, OH: National Institute of Occupational Safety and Health.
- Schuetzle, D. 1983. Sampling of vehicle emissions for Chemical Analysis and biological testing. *Environ Health Perspect* 47: 65-80.
- Schuetzle, D. 1988. Organic tracers for source apportionment of combustion emissions to ambient air particulates. In *Proceedings of the Sixth Symposium on Environmental Analytical Chemistry and Source Apportionment in Ambient and Indoor Atmospheres*.
- Schuetzle, D., T. E. Jensen and J. C. Ball. 1985. Polar PAH derivatives in extracts of particulates: Biological characterization and techniques for chemical analysis. *Environ Int* 11: 169-181.
- Steenland, K., J. Deddens and L. Stayner. 1998. Diesel exhaust and lung cancer in the trucking industry: exposure-response analyses and risk assessment. *Am J Ind Med* 34(3): 220-8.
- Tokiwa, H. and Y. Ohnishi. 1986. Mutagenicity and carcinogenicity of nitroarenes and their sources in the environment. *Crit Rev Toxicol* 17(1): 23-60.
- Ullman, T. L. 1989. Investigation of the Effects of Fuel Composition on Heavy Duty Diesel Engine Emissions. SAE Technical Paper No. 892072. Society of Automotive Engineers.
- WHO. 1996. Diesel Fuel and Exhaust Emissions. Geneva: World Health Organization.
- Williams, D. J., J. W. Milne and D. B. K. Roberts, M.C. 1989. Particulate emission from "in use" motor vehicles - II Diesel vehicles. *Atmos Environ* 23: 2647-2661.
- Williams, R. L. 1982. Diesel particulate emissions: composition, concentration, and control. *Dev Toxicol Environ Sci* 10: 15-32.
- Woskie, S. R., T. J. Smith, S. K. Hammond, M. B. Schenker, E. Garshick and F. E. Speizer. 1988. Estimation of the diesel exhaust exposures of railroad workers: I. Current exposures. *Am J Ind Med* 13(3): 381-94.
- Zaebst, D. D., D. E. Clapp, L. M. Blade, D. A. Marlow, K. Steenland, R. W. Hornung, D. Scheutzle and J. Butler. 1991. Quantitative determination of trucking industry workers' exposures to diesel exhaust particles. *Am Ind Hyg Assoc J* 52(12): 529-41.
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